

1 Disclaimer

2 'This is a near final copy of the accepted paper as submitted for publication. Readers are
3 advised to refer to the final version of the paper which can be found at

4
5 <https://www.sciencedirect.com/science/article/abs/pii/S0048969720339346?via%3Di>

6 [hub](#)

7

8

9 The efficiency of devices intended to reduce microfibre release during clothes

10 washing

11

12 Imogen E. Napper, Aaron C. Barrett, Richard C. Thompson

13

14 Abstract:

15

16 The washing of synthetic clothes is considered to be a substantial source of
17 microplastic to the environment. Therefore, various devices have been designed to
18 capture microfibrils released from clothing during the washing cycle. In this study, we
19 compared 6 different devices which varied from prototypes to commercially available
20 products. These were designed to either be placed inside the drum during the
21 washing cycle or fitted externally to filter the effluent wastewater discharge. The aim
22 of this study was to examine the efficacy of these devices at mitigating microfibre
23 release from clothing during washing or capturing any microfibrils released in the
24 effluent. When compared to the amount of microfibrils entering the wastewater
25 without any device (control), the XFiltra filter was the most successful device. This
26 captured microfibrils, reducing their release to wastewater by around 78%. The
27 Guppyfriend bag was the second most successful device, reducing microfibre
28 release to wastewater by around 54%; it appeared to mainly work by reducing
29 microfibre shedding from the clothing during the washing cycle. Despite some
30 potentially promising results it is important to recognise that fibres are also released
31 when garments are worn in everyday use. Researchers and industry need continue
32 to collaborate to better understand the best intervention points to reduce microfibre
33 shedding, by considering both product design and fibre capture.

34

35 **Keywords:** Microfibres; Washing Machines; Plastic Pollution; Microplastics;
36 Solutions

37
38
39
40
41
42 1.0 Introduction

43
44 Textiles have a wide range of applications, including clothing, upholstery and
45 carpeting, with global textile fibre production exceeding 106 million tons in 2018 (The
46 Fiber Year, 2019); approximately 63% of textile fibres produced are synthetic (e.g.,
47 polyester, nylon) (The Fiber Year, 2019). Other textile fibre materials include natural
48 (e.g., cotton, wool) and semi-synthetic or regenerated fibres (e.g., rayon, acetate).
49 While these types of fibres are produced from natural materials, such as wood pulp
50 or cotton, natural and semi-synthetic fibres can be heavily modified with chemical
51 treatments and additives (e.g., colourants, flame retardants) (Lacasse and Baumann,
52 2004; Xue et al., 2017). In this paper the term microfibre will refer exclusively to
53 fibres (synthetic, semi-synthetic and natural) that are typically < 5 mm.

54
55 It has been suggested that a large proportion of the microfibres found in the marine
56 environment are released from textiles; with a key source being washing clothes
57 (Belzagui et al., 2019; Cesa et al., 2020; De Falco et al., 2018; Napper and
58 Thompson, 2016). On a global scale, Boucher and Friot, (2017) estimated that of all
59 primary microplastics in the world's oceans, 35% arise from laundry of synthetic
60 textiles; an estimated 2 - 13 million tons per year globally (Boucher and Friot, 2017;
61 Mishra et al., 2019). However, due to the lack of research on the release of natural
62 and semi-synthetic fibres, this value is likely substantially underestimated.

63 Microfibres can be released from clothing by mechanical stresses that fabrics
64 undergo during the washing process in a washing machine (Belzagui et al., 2019;
65 Cesa et al., 2020; De Falco et al., 2018; Napper and Thompson, 2016).

66
67 The first paper to highlight the importance of microfiber release from clothing was
68 that of Browne et al 2011 More recently, Napper and Thompson, (2016) estimated

69 that a typical wash (6 kg) could produce over 700,000 microfibrils. Since then, there
70 has been further research focussing on microplastics from washing clothes using
71 filters with fine mesh to capture the microfibrils released (5 µm mesh pore size in De
72 Falco et al., (2018) compared to 25 µm in Napper and Thompson, (2016)). As a
73 consequence, it has recently been estimated that over 6,000,000 microfibrils could
74 be released from an average 6 kg wash (De Falco et al., 2018).

75

76 In addition to the pore size used to capture microfibrils, release estimates can be
77 influenced by differences in materials tested (whole garments vs. textile swatches;
78 textile construction; material composition), load composition (mix loads; full loads;
79 single garments), laundering conditions (temperature; detergent use; cycle time;
80 water volume) and laundering methods (simulated laundering vs. household
81 appliances; model; fibre enumeration and characterization) (Belzagui et al., 2019;
82 Cesa et al., 2020; De Falco et al., 2018; Napper and Thompson, 2016). Currently,
83 there is little scientific consensus on factors influencing release or release estimates
84 across the field.

85

86 Microfibrils released as a result of washing clothes, exit the washing machine via the
87 waste effluent. Depending on the place of use, this effluent either passes directly into
88 the environment or is sent to municipal wastewater treatment plants (WWTPs). In a
89 WWTP, microplastic removal from water can be up to 96% (Carr et al., 2016; Murphy
90 et al., 2016) prior to the water being released to the environment.

91

92 During intense rainfall events, influent to the WWTP can exceed the treatment
93 facilities' handling capacity resulting in the direct discharge of untreated wastewater
94 into rivers, lakes or coastal areas. These events, even if occasional, may have a
95 substantial impact on the total amount of microfibrils released to natural
96 environments (Galafassi et al., 2019). Even if microfibrils are intercepted during
97 wastewater treatment, the resultant sewage sludge is often returned to the land as a
98 fertilizer, hence microfibrils are still released to the environment (Corradini et al.,
99 2019; Gies et al., 2018; Kirchmann et al., 2017). For example, it has been estimated
100 that a secondary WWTP that serves a 650,000 population (Glasgow, UK) with a
101 removal efficiency of 98.41% could release 65 million microplastic particles
102 (including microfibrils) every day (Murphy et al., 2016). A WWTP with a lower

103 retention ability (84%) and a greater population equivalent (1,200,000) could
104 discharge up to 160 million particles per day in its effluent (Magni et al., 2019). It has
105 been reported that the majority of particles detected in WWTPs are microfibres (Gies
106 et al., 2018; Gündoğdu et al., 2018; Leslie et al., 2017).

107

108

109

110 The number of microfibres entering into the marine environment from WWTP is likely
111 to be substantial. Additionally, there are other sources of microfibres into the
112 environment such as tumble drying (Pirc et al., 2016), the wearing of clothes (De
113 Falco et al., 2020) and industrial emissions (Xu et al., 2018). As a consequence,
114 microfibres are now found in aquatic habitats and organisms on a global scale (Avio
115 et al., 2020; Nelms et al., 2019; Obbard et al., 2014; Saturno et al., 2020). Several
116 recent studies revealed the presence of microfibres in various environments,
117 including freshwater and marine surface waters and sediments, as well as terrestrial
118 ecosystems (Ding et al., 2019; González-Pleiter et al., 2020; Liu et al., 2018; Luo et
119 al., 2019; Lusher et al., 2015; Miller et al., 2017; Simon-Sánchez et al., 2019; Taylor
120 et al., 2016; Woodall et al., 2014).

121

122 To mitigate microfibre release in laundry effluent, various devices have been
123 designed to divert and capture released microfibres. These include devices aimed to
124 go in the washing machine drum during a wash cycle and external filters fitted to the
125 washing machine drainpipe to filter microfibres from outgoing effluent. McIlwraith et
126 al., (2019) previously compared the removal efficiency of one in-drum device, the
127 Cora Ball, and one external washing machine filter, the Lint LUV-R. Based on
128 weight, the study reported microfibre reductions into the wastewater by 5% and 80%
129 for the Cora Ball and Lint LUV-R, respectively.

130

131 A range of other products are now available, or are being developed, that have the
132 specific intent to reduce microfibre release. However, there is little data comparing
133 efficacy among such devices. Given the accumulation of plastics in the environment
134 has been associated with a lack of thorough consideration and evaluation of
135 products at the design stage, it is therefore of key importance that any interventions
136 should be appropriately evaluated. Therefore, the overall aim of this study was to

137 examine which devices were the most effective at mitigating the release of
138 microfibers during a typical clothes wash. Efficiency in terms of reducing the release
139 of microfibers to waste water was also compared with control washes that had no
140 device present.

141

142

143

144

145 Our hypothesis assumed that devices would reduce microfibers entering the
146 wastewater from clothes as a consequence of laundering . We chose to quantify the
147 amount of microfibres by analysing the mass collected from the wastewater after
148 washing three jumpers; i.e. microfibres released and that were unsuccessfully
149 captured by the devices.

150

151 2.0 Method

152

153 2.1 Materials

154

155 Three different synthetic fabric types were included in the washing trials to represent
156 a typical mixed load (1.3 ± 0.2 kg). These were medium sized jumpers, sourced from
157 Primark (U.K.), made either of 100% polyester, 100% acrylic or 60% polyester / 40%
158 cotton blend. Each load consisted of a whole garment from each fabric type. In order
159 to identify each fabric type, microfibre samples from five replicates of each jumper
160 type were analysed by FT-IR microscopy in transmission mode with a Hyperion
161 1000 microscope coupled to a Vertex 70 spectrometer (Bruker). Any spectra were
162 recorded with 32 scans in the region of 4000 – 600 cm. The spectra obtained were
163 compared against a spectral database of synthetic polymers (BPAD polymer and
164 synthetic fibres ATR). Napper and Thompson, (2016) had previously shown that
165 garments had an initial peak of microfibre shedding in the first 1-4 washes and then
166 a consistent microfibre shed after the fifth wash. Therefore, prior to data collection,
167 any initial spike in microfibre loss from new clothes was reduced by washing each
168 fabric four times.

169

170 2.2 Devices Tested to Reduce Microfibres Released from Washing

171

172 The devices tested included three in-drum devices: the Guppyfriend washing bag
173 (Langbrett, Germany), a prototype Fourth Element washing bag (Fourth Element,
174 U.K.) and the Cora Ball (Cora Ball, VT, USA). Three external washing machine filters
175 were also tested, including: the Lint LUV-R (Environmental Enhancements, NS,
176 Canada), a prototype XFiltra (Xeros Technology Group, U.K.), and the PlanetCare
177 (PlanetCare Limited, U.K.) (Table 1). All devices were obtained in 2018; however, we
178 understand some manufactures (e.g PlanetCare and Fourth Element Washing Bag)
179 have been working on revised designs. Control washes using the same clothing but
180 without either an in-drum device or external filter were completed following the same
181 methodology. This determined how many fibres were released from the colthign in
182 the absence of any intervention device and allowed is to calculate microfibre capture
183 efficiency.

184

185 There were four replicates of each device and each was used in conjunction with an
186 identical front-loading washing machine of 7 kg capacity (Hotpoint CarePlus
187 WMAOD743P; n = 4).. The mesh used in each device (minus Cora Ball which had
188 no mesh) was visualised by scanning electron microscopy (JEOL, 7001F; Plymouth
189 Electron Microscopy Centre) to assess the pore size.

190

191 Each device and controls were independently tested with four identical replicate
192 mixed clothing loads coupled with four separate washing machines. Each mixed
193 clothing load was washed 10 times, with data recorded after the 1st, 5th and 10th
194 wash (Fig. 1). The washing cycle setting was a 45-minute synthetic wash at 30° C
195 and 1000 R.P.M. This was chosen as a typical automatic programme chosen from
196 the washing machine options (14 programmes available in total). The washing
197 machines did not include weight measurement, so the volume of water used for each
198 wash was consistent throughout (approximately 50 L of water). No detergent or
199 conditioner was used as this would have left deposits affecting any weight change
200 recorded. Additionally, all of the clothing was unwashed and new, so no other foreign
201 contaminants would have affected the weight recorded (i.e. dirt). After washing the
202 mixed loads, each replicate was tumble dried in a condenser dryer using an
203 INDESIT IDC8T3 for 1 hour.

204

205
206
207
208
209
210
211
212
213
214
215
216
217
218
219
220
221
222
223
224
225
226
227
228
229
230
231
232
233
234
235
236
237

2.3 Analysis of Microfibres Captured/Released After Device Testing

For each wash, the mass of microfibres that evaded capture were recorded from each device. After each washing cycle, effluent together with any microfibres which were not caught by the devices were collected in a storage tank and then pumped into a 1 µm filter cartridge (10", Sterner) which was stored in filter housing (AQUAFILTER FHPR1-B1-AQ) (Fig. 2). Aluminium bungs were custom made to block the bottom end of the cartridges; subsequently, the wastewater was pushed through the cartridge leaving any microfibres trapped in its mesh. Cartridges were weighed before and after each wash cycle. The dry weight was recorded for each cartridge after being dried at 30°C to a constant weight and then weighed by a Cubis® precision balance (Sartorius). The cartridges were wrapped in two layers of foil during the drying process to stop microfibre loss or addition of contamination.

The weight of microfibres successfully captured as well as subjective observations on the ease of use of the devices were recorded; this was completed to understand the mechanism of each device, rather than just efficiency testing. For devices where the consumer was expected to visually inspect and then remove the microfibres (Cora Ball, Guppyfriend and Fourth Element washing bag) a timed 5-minute inspection period was used to ensure a sensible and consistent consumer removal effort scenario. This inspection period also provided substantially enough time to remove the majority of collected microfibre mass. This was completed as a consumer would (i.e. without gloves or forceps) and by one person, to reduce variability among individuals. For the PlanetCare filters, the microfibres could not be removed from the device due to being collected into a sealed filter. These filters are

238 intended to be returned to PlanetCare for recycling. Therefore, the dry weight
239 change of the PlanetCare filter itself was recorded.

240

241

242

243

244

245

246

247 2.4. Quality Assurance and Quality Control

248

249 During testing and analysis, all steps were conducted in a regularly cleaned
250 laboratory with controlled access. Care was taken to ensure any potential sources of
251 airborne contamination were minimised (Woodall et al., 2015).. Additionally, all
252 analytical equipment was shielded to mitigate any exposure or contamination
253 throughout the washing and drying process. During analysis (e.g. weighing or
254 sample preparation), procedural blanks were conducted after every 5th sample and
255 confirmed microplastic contamination was minimal with an average of 2 ± 1
256 microfibrils filter⁻¹. This was negligible to the amount of fibres being captured during
257 a wash cycle. After each washing machine cycle which involved mixed clothing
258 loads, cross contamination was minimized between washes, by running the washing-
259 machine at 30 °C, 1000 R.P.M for 45 min with no fabric present.

260

261 2.5 Statistical Analysis

262

263 Normality of the data was confirmed by using QQ plots to examine distribution.
264 Differences between the six devices in terms of the mass of microfibrils captured
265 and released were then analysed using 2-way ANOVAs with device and time point
266 as fixed factors. Examination of residuals of the fitted models indicated the need for
267 transformation (logarithm transformation) of both datasets; residuals were unbiased
268 and homoscedastic after transformation. Post-hoc Tukey tests were used to identify
269 statistically significant differences between devices. Standard error of the mean was
270 used for all analysis.

271

272 3.0 Results

273

274 Washing a mixed load of clothes without any device (control testing), resulted in an
275 average of $0.44 \text{ g} \pm 0.04 \text{ g}$ (mean + S.E) of microfibres being released into the
276 wastewater effluent per wash (Fig. 3A). This estimate (which is assumed to
277 represent 0% success in terms of microfibre shedding mitigation or capture) was
278 then compared against the mass of microfibres collected from wastewater effluent
279 with each device. Higher efficiency (%) equates to a more successful device. When
280 comparing between devices and control, the devices ranged between 21 – 78%
281 efficiency. XFiltra was the most successful device, reducing the number of
282 microfibres being released into the wastewater by $78 \pm 5 \%$. The Guppyfriend
283 washing bag was the second most successful device at $54 \pm 14 \%$. The Cora Ball
284 was the third most successful at $31 \pm 8 \%$. The Lint LUV-R and PlanetCare had
285 similar results at $29 \pm 15 \%$ and $25 \pm 20 \%$, respectively. The Fourth Element
286 washing bag was the least effective at $21 \pm 9 \%$ (Fig. 3A).

287

288 There were significant differences in the mass of microfibres released into
289 wastewater across devices (2-way ANOVA; $p = < 0.008$); these differences were
290 consistent across the three timepoints (Table 2). At the 0.05 level, the Guppyfriend
291 washing bag and XFiltra were the only devices to release significantly less
292 microfibres compared to controls (no device). There were no significant differences
293 between microfibre release by in-drum devices (Fig. 3A). XFiltra also released
294 significantly less microfibres than the Cora Ball, Fourth Element washing bags, Lint
295 LUV-R and Planetcare.

296

297 There was also a significant difference in the mass of microfibres successfully
298 captured by each device type (2-way ANOVA; $p = < 0.000$) (Fig. 3B). There was no
299 significant difference, at the 0.05 level, between Cora Ball, Guppyfriend and Fourth
300 Element washing bags (Fig. 3B). Trying to manually remove the microfibres from
301 devices added to the drum (Cora Ball, Guppyfriend and Fourth Element washing
302 bags) was time consuming as there was a large surface area to analyse and little
303 mass typically collected. With the Guppyfriend washing bag, microfibres typically
304 accumulated in the hem of material. However, for the external filters (XFiltra and Lint
305 LUV-R), microfibres would typically accumulate in a localised area. PlanetCare

306 captured microfibrils were irretrievable due to the devices design; these filters are
307 intended to be returned to PlanetCare for recycling.

308

309 Scanning electron images were obtained to assess the pore size of the mesh used
310 in each device (apart from the Cora Ball, which contained no mesh) (Fig. 4). The
311 largest pore size was the Lint LUV-R, which had 2 pore sizes: 285 μm and 175
312 μm . PlanetCare had the second largest pore size of 200 μm . XFiltra had a pore
313 size of 60 μm . The two bag devices (Guppyfriend and Forth Element washing bag)
314 had the smallest pore size, of 50 μm .

315

316

317 4.0 Discussion

318

319 The XFiltra prototype device was the most successful device, capturing on average
320 78% of the microfibrils per wash. It is possible that this device was more successful
321 firstly because it had the finest mesh pore size (60 μm) compared to the other filters
322 (PlanetCare & Lint LUV-R) which had pore sizes $>175 \mu\text{m}$, and secondarily, because
323 it was the only 'active device', in that it used a motor powered centrifugal separator
324 requiring an external electrical supply to facilitate the flow of the waste water through
325 the filtration mesh. There was also a large variation in efficiency between the
326 Guppyfriend and Fourth Element washing bags; 54% and 21%, respectively. Even
327 though each bag device had similar mesh pore size (50 μm), their shape and design
328 were different which could account for differences in efficiency.

329

330 Additionally, our results found that there was a significant difference in the mass of
331 microfibrils captured by the devices. Devices directly placed into the washing
332 machine drum (Cora Ball, Guppyfriend and Fourth Element washing bags) were all
333 less successful at capturing microfibrils than the filters, but were still found to reduce
334 microfibre emissions into the wastewater by 21 – 54%. This effect seems to have
335 resulted from reduced microfibre shedding by garments during the washing cycle
336 due to the design of these devices.

337

338 Previous research has demonstrated that the Cora Ball and the Lint LUV-R reduced
339 the weight of microfibrils released after a washing cycle by 5% and 80%,

340 respectively (McIlwraith et al., 2019). However, we report that the Lint LUV-R to be
341 less successful at 29%, and the Cora Ball at 31%. One possible explanation for the
342 differences between studies could be because McIlwraith et al., (2019) did not focus
343 on microfibrils smaller than 10 μm , whereas this study had a lower limit of 1 μm .
344 Additionally, there are differences in study design. McIlwraith et al. (2019) used
345 100% polyester fleece blankets, which have been reported to have high shedding
346 rates (Browne et al., 2011; Pirc et al., 2016; Sillanpää and Sainio, 2017). Their
347 research also used a top loading machine which is suspected to shed more
348 microfibrils from clothing/fabric compared to a frontloading machine (Hartline et al.,
349 2016).

350

351 Despite removing 21-78% of outgoing microfibrils, the six devices tested in the
352 present study still released 0.10-0.35 g of microfibrils per wash. As such they do not
353 offer a complete solution and alternative measures will likely still need to be taken to
354 address this issue. A combination of in-drum and external filter technologies used
355 together may cause less shedding and increased microfibre capture, whilst also
356 reducing the need to clean the filter as frequently.

357

358 Additionally, reducing shedding through changes in fabric design could be a more
359 overarching mitigation strategy, as this is likely to help reduce emissions during all
360 use phases: wearing, washing and tumble drying (De Falco et al., 2020; Napper and
361 Thompson, 2016; Pirc et al., 2016). De Falco *et al.*, (2020) estimated the quantity of
362 microfibrils released into the air directly as a consequence of wearing clothes. Their
363 research found that 400 fibres gram^{-1} of fabric could be shed by items of clothing
364 during just 20 minutes of normal activity. Due to this, it is anticipated that
365 atmospheric deposition of microplastics, especially through the wearing of clothes, is
366 a substantial pathway into the environment. Microplastics are potentially transported
367 by wind, because of their small size and low density, from their original source
368 (Bergmann et al., 2019).

369

370

371 Other measures can be put in place to minimise microfibrils shed in the washing
372 cycle. Only washing your clothes when required is a simple way to minimise
373 microfibre shedding. Research has also indicated that delicate wash cycles release

374 more microfibres per wash than a lower water-volume standard wash; showing that
375 simply reducing the water-volume-to-fabric ratio could also have an effect in reducing
376 the amount of microfibres generated (Kelly et al., 2019). Therefore, an effective
377 strategy would be using a combination of modified fabric design together with less
378 aggressive washing cycles and adding washing machine filters/in-drum devices.

379

380 More research is needed to establish how regularly consumers would actually clean
381 the devices (we considered a 5-minute clean to be a reasonably generous amount of
382 time). It is unclear what consumers would do with any microfibres removed; e.g.
383 dispose to landfill or wash them down the sink unintentionally to clean the device.
384 Clear labelling and instructions should be in place to ensure the proper disposal of
385 microfibres. There are further limitations to the widespread implementation of these
386 devices. For the in-drum devices, research should analyse whether garments being
387 laundered receive the same quality of cleaning. Due to the size of the washing bags,
388 the consumer is also limited in the number of clothes able to be laundered, so more
389 washes may be required. Additionally, the external washing machine filters will
390 require potential space for installation in washing machines. All devices vary in price
391 and are currently assumed to be purchased by the consumer, although there is the
392 potential for washing machine manufactures to incorporate filters internally in
393 production.

394

395 Other mitigation strategies that have been promoted include improvements to
396 WWTPs and a switch from synthetic to natural textiles. However, these solutions are
397 more unrealistic. WWTP microplastic removal can already be up to 96% (Carr et al.,
398 2016; Murphy et al., 2016) prior to the water being released to the environment.
399 Upgrading WWTP with more efficient filtering systems could be expensive or
400 potentially not even possible with the system already in place (Conley et al., 2019).
401 Furthermore, replacing synthetic textiles with natural counterparts would typically be
402 more expensive and the impact of non-synthetic microfibres accumulating in the
403 environment is also currently unknown (Dris et al., 2017).

404

405 Many of the issues associated with current levels of plastic pollution have arisen
406 because of inadequate consideration at the industrial design stage of the
407 environmental consequences associated with production, use and disposal. Going

408 forward it is imperative we learn from these mistakes. From the perspective of
409 interventions to tackle current issues with laundering, this needs to be done in terms
410 of their efficacy in addressing the particular issue and potential unintended
411 environmental consequences. From an environmental perspective we can no longer
412 afford to produce devices and products in the hope they will be not be harmful, rather
413 we must rigorously assess performance, prior to release. Industries will continue to
414 develop solutions aimed to stem the flow of or capture plastic getting into the
415 environment. However, it is essential that any proposed solutions are fully tested for
416 their efficiency and evaluated to understand their potential benefit.

417

418 5.0 Conclusion

419

420 There is now considerable agreement and consensus about the issue of plastic
421 waste and pollution. However, some of the key challenges now lie, not just in
422 environmental science to help understand the problem, but robust evidence to inform
423 appropriate solutions. With growing concern about the accumulation of plastic and
424 microplastic (including concern about microfibre pollution) devices are being
425 developed with the intent to reduce the release of microfibres to the environment.
426 These solutions vary in their approach, such as providing consumer ease or being
427 the most effective. They also vary in market readiness. Our study has shown they
428 vary in their ability to address the issue of microfibre contamination. XFiltra and the
429 Guppyfriend washing bag significantly reduced the number of microfibres released
430 into the wastewater compared to no device being present. In order to help minimise
431 some of the avoidable environmental challenges that we currently face, it is essential
432 that technological advance is coupled, at the design stage, to appropriate
433 environmental science, in order to minimise unintended environmental
434 consequences.

435

436

437

438

439

440

References

441

442 Avio, C.G., Pittura, L., d'Errico, G., Abel, S., Amorello, S., Marino, G., Gorbi, S.,
443 Regoli, F., 2020. Distribution and characterization of microplastic particles and
444 textile microfibers in Adriatic food webs: General insights for biomonitoring
445 strategies. *Environ. Pollut.* 258, 113766. doi:10.1016/j.envpol.2019.113766

446 Belzagui, F., Crespi, M., Álvarez, A., Gutiérrez-Bouzán, C., Vilaseca, M., 2019.
447 Microplastics' emissions: Microfibers' detachment from textile garments.
448 *Environ. Pollut.* 248, 1028–1035. doi:10.1016/j.envpol.2019.02.059

449 Bergmann, M., Mützel, S., Primpke, S., Tekman, M.B., Trachsel, J., Gerdts, G.,
450 2019. White and wonderful? Microplastics prevail in snow from the Alps to the
451 Arctic. *Sci. Adv.* 5, eaax1157. doi:10.1126/sciadv.aax1157

452 Boucher, J., Friot, D., 2017. Primary Microplastics in the Oceans: a Global
453 Evaluation of Sources. Gland, Switzerland.

454 Browne, M.A., Crump, P., Niven, S.J., Teuten, E., Tonkin, A., Galloway, T.,
455 Thompson, R., 2011. Accumulation of microplastic on shorelines worldwide:
456 Sources and sinks. *Environ. Sci. Technol.* 45, 9175–9179.
457 doi:10.1021/es201811s

458 Carr, S.A., Liu, J., Tesoro, A.G., 2016. Transport and fate of microplastic particles in
459 wastewater treatment plants. *Water Res.* 91, 174–182.
460 doi:10.1016/j.watres.2016.01.002

461 Cesa, F.S., Turra, A., Checon, H.H., Leonardi, B., Baruque-Ramos, J., 2020.
462 Laundering and textile parameters influence fibers release in household
463 washings. *Environ. Pollut.* 257, 113553. doi:10.1016/j.envpol.2019.113553

464 Conley, K., Clum, A., Deepe, J., Lane, H., Beckingham, B., 2019. Wastewater
465 treatment plants as a source of microplastics to an urban estuary: Removal
466 efficiencies and loading per capita over one year. *Water Res.* X 3, 100030.
467 doi:10.1016/j.wroa.2019.100030

468 Corradini, F., Meza, P., Eguiluz, R., Casado, F., Huerta-Lwanga, E., Geissen, V.,
469 2019. Evidence of microplastic accumulation in agricultural soils from sewage
470 sludge disposal. *Sci. Total Environ.* 671, 411–420.
471 doi:10.1016/J.SCITOTENV.2019.03.368

472 De Falco, F., Cocca, M., Avella, M., Thompson, R.C., 2020. Microfibre release to
473 water, via laundering, and to air, via everyday use: a comparison between
474 polyester clothing with differing textile parameters. *Environ. Sci. Technol.* 54,
475 3288–3296. doi:10.1021/acs.est.9b06892

476 De Falco, F., Gullo, M.P., Gentile, G., Di Pace, E., Cocca, M., Gelabert, L., Brouta-
477 Agnésa, M., Rovira, A., Escudero, R., Villalba, R., Mossotti, R., Montarsolo, A.,
478 Gavignano, S., Tonin, C., Avella, M., 2018. Evaluation of microplastic release
479 caused by textile washing processes of synthetic fabrics. *Environ. Pollut.* 236,
480 916–925. doi:10.1016/J.ENVPOL.2017.10.057

481 Ding, L., Mao, R. fan, Guo, X., Yang, X., Zhang, Q., Yang, C., 2019. Microplastics in
482 surface waters and sediments of the Wei River, in the northwest of China. *Sci.*
483 *Total Environ.* 667, 427–434. doi:10.1016/J.SCITOTENV.2019.02.332

484 Dris, R., Gasperi, J., Mirande, C., Mandin, C., Guerrouache, M., Langlois, V., Tassin,
485 B., 2017. A first overview of textile fibers, including microplastics, in indoor and
486 outdoor environments. *Environ. Pollut.* 221, 453–458.
487 doi:10.1016/j.envpol.2016.12.013

488 Galafassi, S., Nizzetto, L., Volta, P., 2019. Plastic sources: A survey across scientific
489 and grey literature for their inventory and relative contribution to microplastics
490 pollution in natural environments, with an emphasis on surface water. *Sci. Total*
491 *Environ.* 693, 133499. doi:10.1016/j.scitotenv.2019.07.305

492 Gies, E.A., LeNoble, J.L., Noël, M., Etemadifar, A., Bishay, F., Hall, E.R., Ross, P.S.,
493 2018. Retention of microplastics in a major secondary wastewater treatment
494 plant in Vancouver, Canada. *Mar. Pollut. Bull.* 133, 553–561.
495 doi:10.1016/j.marpolbul.2018.06.006

496 González-Pleiter, M., Velázquez, D., Edo, C., Carretero, O., Gago, J., Barón-Sola,
497 Á., Hernández, L.E., Yousef, I., Quesada, A., Leganés, F., Rosal, R.,
498 Fernández-Piñas, F., 2020. Fibers spreading worldwide: Microplastics and other
499 anthropogenic litter in an Arctic freshwater lake. *Sci. Total Environ.* 722, 137904.
500 doi:10.1016/j.scitotenv.2020.137904

501 Gündoğdu, S., Çevik, C., Güzel, E., Kilercioğlu, S., 2018. Microplastics in municipal
502 wastewater treatment plants in Turkey: a comparison of the influent and
503 secondary effluent concentrations. *Environ. Monit. Assess.* 190, 626.
504 doi:10.1007/s10661-018-7010-y

505 Hartline, N.L., Bruce, N.J., Karba, S.N., Ruff, E.O., Sonar, S.U., Holden, P.A., 2016.
506 Microfiber Masses Recovered from Conventional Machine Washing of New or
507 Aged Garments. *Environ. Sci. Technol.* 50, 11532–11538.
508 doi:10.1021/acs.est.6b03045

509 Kelly, M.R., Lant, N.J., Kurr, M., Burgess, J.G., 2019. Importance of Water-Volume

510 on the Release of Microplastic Fibers from Laundry. *Environ. Sci. Technol.* 53,
511 11735–11744. doi:10.1021/acs.est.9b03022

512 Kirchmann, H., Börjesson, G., Kätterer, T., Cohen, Y., 2017. From agricultural use of
513 sewage sludge to nutrient extraction: A soil science outlook. *Ambio* 46, 143–
514 154. doi:10.1007/s13280-016-0816-3

515 Lacasse, K., Baumann, W., 2004. *Textile Chemicals - Environmental Data and*
516 *Facts*. Springer, Berlin, Germany.

517 Leslie, H.A., Brandsma, S.H., van Velzen, M.J.M., Vethaak, A.D., 2017.
518 *Microplastics en route: Field measurements in the Dutch river delta and*
519 *Amsterdam canals, wastewater treatment plants, North Sea sediments and*
520 *biota*. *Environ. Int.* 101, 133–142. doi:10.1016/j.envint.2017.01.018

521 Liu, M., Lu, S., Song, Y., Lei, L., Hu, J., Lv, W., Zhou, W., Cao, C., Shi, H., Yang, X.,
522 He, D., 2018. Microplastic and mesoplastic pollution in farmland soils in suburbs
523 of Shanghai, China. *Environ. Pollut.* 242, 855–862.
524 doi:10.1016/J.ENVPOL.2018.07.051

525 Luo, W., Su, L., Craig, N.J., Du, F., Wu, C., Shi, H., 2019. Comparison of
526 microplastic pollution in different water bodies from urban creeks to coastal
527 waters. *Environ. Pollut.* 246, 174–182. doi:10.1016/J.ENVPOL.2018.11.081

528 Lusher, A.L., Tirelli, V., O'Connor, I., Officer, R., Halsband, C., Galloway, T.S., 2015.
529 *Microplastics in Arctic polar waters: the first reported values of particles in*
530 *surface and sub-surface samples*. *Sci. Rep.* 5, 14947. doi:10.1038/srep14947

531 Magni, S., Binelli, A., Pittura, L., Avio, C.G., Della Torre, C., Parenti, C.C., Gorbi, S.,
532 Regoli, F., 2019. The fate of microplastics in an Italian Wastewater Treatment
533 Plant. *Sci. Total Environ.* 652, 602–610.
534 doi:10.1016/J.SCITOTENV.2018.10.269

535 McIlwraith, H.K., Lin, J., Erdle, L.M., Mallos, N., Diamond, M.L., Rochman, C.M.,
536 2019. Capturing microfibers – marketed technologies reduce microfiber
537 emissions from washing machines. *Mar. Pollut. Bull.* 139, 40–45.
538 doi:10.1016/J.MARPOLBUL.2018.12.012

539 Miller, R.Z., Watts, A.J.R., Winslow, B.O., Galloway, T.S., Barrows, A.P.W., 2017.
540 *Mountains to the sea: River study of plastic and non-plastic microfiber pollution*
541 *in the northeast USA*. *Mar. Pollut. Bull.* 124, 245–251.
542 doi:10.1016/J.MARPOLBUL.2017.07.028

543 Mishra, S., Rath, C. charan, Das, A.P., 2019. Marine microfiber pollution: A review

544 on present status and future challenges. *Mar. Pollut. Bull.*
545 doi:10.1016/j.marpolbul.2019.01.039

546 Murphy, F., Ewins, C., Carbonnier, F., Quinn, B., 2016. Wastewater Treatment
547 Works (WwTW) as a Source of Microplastics in the Aquatic Environment.
548 *Environ. Sci. Technol.* 50, 5800–5808. doi:10.1021/acs.est.5b05416

549 Napper, I.E.I., Thompson, R.C., 2016. Release of synthetic microplastic plastic fibres
550 from domestic washing machines: Effects of fabric type and washing conditions.
551 *Mar. Pollut. Bull.* 112, 39–45. doi:10.1016/j.marpolbul.2016.09.025

552 Nelms, S.E., Barnett, J., Brownlow, A., Davison, N.J., Deaville, R., Galloway, T.S.,
553 Lindeque, P.K., Santillo, D., Godley, B.J., 2019. Microplastics in marine
554 mammals stranded around the British coast: ubiquitous but transitory? *Sci. Rep.*
555 9, 1075. doi:10.1038/s41598-018-37428-3

556 Obbard, R.W., Sadri, S., Wong, Y.Q., Khitun, A.A., Baker, I., Thompson, R.C., 2014.
557 Global warming releases microplastic legacy frozen in Arctic Sea ice. *Earth's*
558 *Futur.* 2, 315–320. doi:10.1002/2014EF000240.Abstract

559 Pirc, U., Vidmar, M., Mozer, A., Kržan, A., 2016. Emissions of microplastic fibers
560 from microfiber fleece during domestic washing. *Environ. Sci. Pollut. Res.* 23,
561 22206–22211. doi:10.1007/s11356-016-7703-0

562 Saturno, J., Liboiron, M., Ammendolia, J., Healey, N., Earles, E., Duman, N., Schoot,
563 I., Morris, T., Favaro, B., 2020. Occurrence of plastics ingested by Atlantic cod
564 (*Gadus morhua*) destined for human consumption (Fogo Island, Newfoundland
565 and Labrador). *Mar. Pollut. Bull.* 153, 110993.
566 doi:10.1016/j.marpolbul.2020.110993

567 Sillanpää, M., Sainio, P., 2017. Release of polyester and cotton fibers from textiles in
568 machine washings. *Environ. Sci. Pollut. Res.* 24, 19313–19321.
569 doi:10.1007/s11356-017-9621-1

570 Simon-Sánchez, L., Grelaud, M., Garcia-Orellana, J., Ziveri, P., 2019. River Deltas
571 as hotspots of microplastic accumulation: The case study of the Ebro River (NW
572 Mediterranean). *Sci. Total Environ.* 687, 1186–1196.
573 doi:10.1016/J.SCITOTENV.2019.06.168

574 Taylor, M.L., Gwinnett, C., Robinson, L.F., Woodall, L.C., 2016. Plastic microfibre
575 ingestion by deep-sea organisms. *Sci. Rep.* 6, 33997. doi:10.1038/srep33997

576 The Fiber Year, 2019. The Fiber Year 2019; World Survey on Textiles & Nonwovens.
577 Frankfurt, Germany.

578 Woodall, L.C., Sanchez-Vidal, A., Canals, M., Paterson, G.L.J.J., Coppock, R.,
579 Sleight, V., Calafat, A., Rogers, A.D., Narayanaswamy, B.E., Thompson, R.C.,
580 2014. The deep sea is a major sink for microplastic debris. *R. Soc. Open Sci.* 1,
581 140317–140317. doi:10.1098/rsos.140317

582 Xu, C.K., Cheng, H., Liao, Z.J., 2018. Towards sustainable growth in the textile
583 industry: A case study of environmental policy in China. *Polish J. Environ. Stud.*
584 27, 2325–2336. doi:10.15244/pjoes/79720

585 Xue, J., Liu, W., Kannan, K., 2017. Bisphenols, Benzophenones, and Bisphenol A
586 Diglycidyl Ethers in Textiles and Infant Clothing. *Environ. Sci. Technol.* 51,
587 5279–5286. doi:10.1021/acs.est.7b00701

588
589
590
591
592