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Kinetics of liquid phase batch adsorption experiments

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Abstract



Batch adsorption experiments are carried out by adding a known amount of adsorbent to a liquid solution at a known initial concentration and following the evolution in time of the concentration of the adsorbate. This is a very common method to obtain equilibrium and kinetic information in liquid systems, but in most cases kinetic results are analysed on the basis of empirical models. Two phenomenological models based on macropore diffusion in beads and shrinking core kinetics are used to generate data that are then interpreted with the widely used *unconstrained* linear regression methods. The results show that for both cases R^2 values close to unity are obtained leading to the incorrect interpretation of the mechanism of mass transport. It is recommended that batch adsorption experiments should be analysed using phenomenological models to obtain physical parameters that are applicable to other systems and to reduce the experiments required to characterise fully the kinetics of adsorption.

Keywords Batch adsorption \cdot Immersion experiment \cdot Pseudo first order kinetics \cdot Pseudo second order kinetics \cdot Elovich kinetics \cdot Langmuir kinetics \cdot Diffusion in particles

List of symbols

- $a_{\rm s}$ Surface to volume ratio of solid (m⁻¹)
- A Constant defined in Eq. 9c
- A_D Slope of short time regression line for the diffusion model (mol m⁻³ s^{-0.5})
- *b* Langmuir affinity $(m^3 mol^{-1})$
- *B* Elovich constant ($mol^{-1} m^3$)
- c_0 Initial concentration in the fluid phase (mol m⁻³)
- c_{∞} Final concentration in the fluid phase (mol m⁻³)
- c Fluid phase concentration (mol m^{-3})
- C Intercept of short time regression line for the diffusion model (mol m^{-3})
- c_P Concentration in the macropores (mol m⁻³)
- \bar{c}_P Average concentration in the macropores (mol m⁻³)
- D_m Molecular diffusivity (m² s⁻¹)
- D_P^{App} Apparent diffusivity from linear and constant concentration model (m² s⁻¹)

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 $D^{E\!f\!f}$ Effective diffusivity defined in Eq. 12 $(m^2 s^{-1})$ D^{P}_{EL} Pore diffusivity linear model defined in Eq. 30b $(m^2 s^{-1})$ Solid diffusivity in shrinking core model $(m^2 s^{-1})$ D_S Pseudo first order kinetic constant (s^{-1}) k_1 Pseudo second order kinetic constant (mol⁻¹ m³ k_2 s^{-1}) K Slope of dimensionless secant of the equilibrium isotherm, $\frac{Q_{\infty}}{c_{\infty}}$ Alternative pseudo second order kinetic constant K_2 $\bar{Q}_{\infty}k_2(\mathrm{s}^{-1})$ Langmuir adsorption rate constant (mol⁻¹ m³ s⁻¹) k_A Langmuir desorption rate constant (s⁻¹) k_D Elovich rate constant (mol $m^{-3} s^{-1}$) k_E Film mass transfer coefficient (m s^{-1}) k_F Linear driving force coefficient (m s^{-1}) k_{LDF} Quadratic driving force coefficient (mol⁻¹ m³ s⁻¹) k_{QDF} Surface resistance mass transfer coefficient (m s^{-1}) k_{S} Vermeulen driving force coefficient ($mol^{-1} m^3 s^{-1}$) k_V M_{S} Mass of solid (kg) Average concentration in the adsorbed phase (mol \bar{q} m⁻³) Defined in Eq. 9d (mol m^{-3}) q_{-} Defined in Eq. 9d (mol m^{-3}) q_+

 \bar{Q} Average concentration in the solid (mol m⁻³)

 \bar{Q}_{eq} Solid concentration at equilibrium with fluid concentration (mol m⁻³)

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- Q_{∞} Final concentration in the solid (mol m⁻³)
- q_S Langmuir saturation capacity (mol m⁻³)
- Q_S Concentration after reaction in shrinking core model (mol⁻³)
- *r* Radial coordinate (m)
- R² Square of the linear correlation coefficient
- R_P Particle radius (m)
- Sh Sherwood number
- t Time (s)
- t_0 Time shift in Elovich linearization defined in Eq. 29 (s)
- V_F Volume of fluid (m³)
- V_{S} Volume of solid (m³)
- z Dimensionless radial coordinate
- z_{∞} Final position of shrinking core interface

Greek symbols

- α Volume ratio, $\frac{V_s}{V}$
- β_n Eigenvalues of the diffusion equation calculated from Eq. 30c
- δ Dimensionless ratio defined in Eq. 14
- ϵ_P Macropore void fraction of the particles
- γ Dimensionless parameter defined in Eq. 9d
- Γ Nonlinearity parameter defined in Eq. 38
- λ Dimensionless parameter defined in Eq. 30c
- Λ Dimensionless ratio defined in Eq. 15
- ρ_S Solid density (kg m⁻³)
- τ Tortuosity

1 Introduction

A batch adsorption experiment from the liquid phase, also known as immersion experiment, is one of the most common tests used to measure adsorption equilibrium and kinetics from solutions. It consists of the addition of a known mass of sample to a fixed volume of liquid at an initial concentration. The liquid is mixed either using a stirrer or the entire cell is agitated to mix the liquid phase. The liquid is sampled or is circulated to a detector and the evolution of the liquid concentration in time is monitored up to equilibrium. This is schematically shown in Fig. 1 for a single cell.

To obtain an adsorption isotherm the experiment is repeated varying either the amount of solid or the initial concentration of the solution or both. Everett (1986) provides an excellent discussion of the different methods to measure adsorption isotherms from the liquid phase, along with recommendations on issues that may be sources of uncertainty.

Here the focus will be on kinetic measurements and consider two general industrially relevant cases:

a Physisorption in porous beads where internal mass transfer is governed by diffusion;



Fig. 1 Schematic representation of a liquid phase batch adsorption experiment

b Chemisorption represented by a shrinking core mechanism.

For chemisorption one can develop more complex kinetic schemes, but these are system dependent and are beyond the scope of this contribution. Furthermore, as diffusion in the stagnant liquid in the macropores is relatively slow, the shrinking core model has been shown to describe well, for example, ion-exchange from concentrated solutions (Phelps and Ruthven 2001) and protein adsorption in porous particles (Zhu and Carta 2016) and is a very relevant limiting case to consider. The uptake curves obtained from the two mechanisms become mathematically very similar when dealing with very sharp isotherms, i.e. the limit of the rectangular isotherm (Teo and Ruthven 1986; Ruthven 2000; Do 1998). Therefore, to confirm which mechanism is relevant imaging of carefully cut beads can be used to determine if the shrinking core model applies (Phelps and Ruthven 2001; Zhu and Carta 2016; Dominguez et al. 2019).

The main scope of this work is to use these phenomenological models to generate liquid batch sorption uptake curves that can then be used to demonstrate the potential pitfalls of correlating these kinetic processes using empirical models that are widely used because they lead to linear plots. The models considered are pseudo first order and pseudo second order kinetic models (Ho and McKay 1998) and the Elovich chemisorption kinetics (Low 1960; Aharoni and Ungarish 1977; Cheung et al. 2000). Although diffusion in a particle would be a phenomenological model, the typical approach is to use the solution valid for short times, crediting either Boyd et al. (1947a) or Weber and Morris (1963). What makes this also an empirical approach is the fact that the analytical solution used applies to the case of constant fluid concentration and linear isotherm and both these assumptions do not apply in the batch adsorption experiment.

These models are widely used in the literature because of the ability to obtain their parameters using unconstrained linear regression, which is easily accessible. The fact that this is a very common methodology in adsorption from solution is evidenced by the fact that (Ho and McKay 1998; Cheung et al. 2000; Boyd et al. 1947a) are cited well over 1000 times. Only considering issues of the journal Adsorption since 2015, 16 contributions reporting parameters based on linearized plots of these empirical models could be found (Câmara et al. 2020; Kong et al. 2016; Lupul et al. 2015; Peralta et al. 2019; Rincon-Silva et al. 2016; Shen et al. 2018; van der Heyden et al. 2018; Bartczak et al. 2016; Bazargan et al. 2017; Ciesielczyk et al. 2016; Marques Fraga et al. 2020; Hubetska et al. 2020; Morales-Perez et al. 2016; Popugaeva et al. 2019; De Smedt et al. 2015; Yuan et al. 2019). This shows that significant effort is being devoted to determining the kinetics of adsorption in liquid systems and it is therefore important to be able to interpret these results correctly and arrive at physical parameters that are directly applicable in the design and simulation of adsorption separation processes.

The paper presents a description of the phenomenological models and a brief outline of the empirical approaches and the corresponding linearised plots. Two well characterised systems that are examples of physisorption and chemisorption will be used to generate batch adsorption curves for dilute and concentrated solutions. Physisorption of xylenes in beads of Y zeolite are well characterised by Santacesaria et al. (1982a), who have also shown the portability of the parameters obtained to the prediction of breakthrough curves (Santacesarla et al. 1982b) and chromatographic pulses (Santacesarla et al. 1982c). For chemisorption Cu++ uptake on a commercial ion-exchange resin was investigated in detail (Phelps and Ruthven 2001) and also in this case the parameters obtained where used to predict breakthrough curves and the performance of an endless belt adsorber (Phelps and Ruthven 2002). Teo and Ruthven (1986) also predict breakthrough curves based on the kinetic measurements for water from an aqueous ethanol solution and use a rectangular isotherm for water on 3A molecular sieves.

1.1 Common basis for batch adsorption from the liquid phase

In a batch adsorption experiment, starting from a fully regenerated adsorbent, the mass balance in the system is given by

$$V_F c_0 = V_F \bar{c}(t) + V_S Q(t) \tag{1}$$

where *c* is the concentration of the fluid and *Q* is the concentration in the solid, while V_F and V_S are the volumes of the

fluid and solid. The overbar indicates the average over the corresponding volume.

$$c(t) = c_0 - \frac{V_S}{V_F} \bar{Q}(t) = c_0 - \alpha \bar{Q}(t) \quad \text{with} \quad \alpha = \frac{V_S}{V_F}$$
(2)

Therefore an uptake curve of $\overline{Q}(t)$ vs. t is easily obtained from c(t) vs. t.

The typical experiment uses a solid that is initially dry, but alternatively the solid could be equilibrated with the pure solvent and then added to the solution. With a dry solid, the initial part of the experiment will have a transient associated with the wetting of the adsorbent and a temperature rise of the solid. This is typically neglected because the thermal mass of the fluid limits the temperature rise and the initial adsorption of the solvent is fast compared to the process being studied. An alternative improved configuration would be to pack a small column with the adsorbent and pump the fluid mixture through the column at high rate (Teo and Ruthven 1986; Babić et al. 2008; Wegmann et al. 2011). In this case the column could be exposed to the pure solvent prior to the start of the experiment. In all the configurations mixing is achieved either with a stirrer or through the rapid circulation. In what follows we make the assumption that the liquid is well mixed and the fluid is considered to be at a uniform concentration, i.e. $c(t) = \bar{c}(t)$.

With a liquid system, the solute molecules will have to reach the solid through a boundary layer. Whatever mass transport mechanism is considered in the solid phase, there will always be a film resistance and

$$\frac{d\bar{q}}{dt} = k_F a_S \left(c - c_s \right) \tag{3}$$

where c_s is the fluid concentration at the external surface of the solid and k_E is the film mass transfer coefficient and a_S is the surface to volume ratio for the solid. Experiments at different stirrer speeds are normally performed, but one should not assume that k_F is negligible once the kinetic response becomes independent of the speed of the stirrer. At high stirrer speeds and corresponding high fluid velocities the particles will be fully fluidized and therefore the *relative* velocity between the fluid and the particles will become effectively independent of the speed of the stirrer. This is one advantage of the use of a fixed bed with a circulation pump, where the relative velocity is always the fluid velocity and results that become independent of the circulation rate could be used to indicate a negligible film resistance (Teo and Ruthven 1986). A further advantage of the fluid circulation system is a reduction in particle attrition (Teo and Ruthven 1986). A simple method to assess the relative importance of the film resistance is to estimate the worst case scenario, where the Sherwood number is 2, which corresponds to the case of a stagnant liquid film.

For the solid phase we consider a biporous adsorbent, with macropores that lead to the adsorbing material. In the macropores a normal liquid solution is present and there is no selectivity to any molecule. The overall concentration in the particles is then

$$\bar{Q} = \varepsilon_P \bar{c}_P + (1 - \varepsilon_P) \bar{q} \tag{4}$$

where ε_P is the macropore void fraction of the particles.

Having defined the fluid phase concentration at the surface of the particle, it is possible to define the adsorbed phase concentration that would be at equilibrium with this fluid concentration, \bar{Q}_{ear} .

$$\bar{Q}_{eq} = \epsilon_P c_s + \left(1 - \epsilon_P\right) \bar{q}_{eq} \tag{5}$$

For the purposes of this contribution the Langmuir isotherm will be used

$$\bar{q}_{eq} = q_S \frac{bc_s}{1 + bc_s} \tag{6}$$

The Langmuir saturation capacity, q_s , and the affinity b are constants and should be determined independently, regressing accurately the final equilibrium data at different fluid concentrations.

 \bar{Q}_{eq} varies during the experiment because the measurement is based on determining the variation of the fluid phase concentration in time, therefore the equilibrium isotherm becomes an integral part of the dynamic model.

To close the problem it is necessary to describe mass transport in the solid using a phenomenological model. Here the emphasis is on models that allow to determine meaningful physical parameters from pure component immersion experiments which can be used to predict reliably multicomponent separations.

1.2 Linear rate model

The linear rate model is the simplest case to consider

$$\frac{d\bar{Q}}{dt} = k_S a_S \left(\bar{Q}_{eq} - \bar{Q} \right) \tag{7}$$

As an example, this model has been used to describe biosorption of dyes and corresponding binary breakthrough curves (Fernandez et al. 2015).

Glueckauf (1955) showed that the surface resistance constant, k_S , can be expressed in terms of the internal diffusion coefficient. For a sphere, $k_S = \frac{5D}{R_p}$. This model can be generalised to multiple resistances in series and becomes a lumped model approach. The relationships needed to include external film resistance, surface resistance, diffusion in macropores, and diffusion in micropores can be obtained from the moments of chromatographic or kinetic experiments and can be found in Ruthven (1984). The film and surface resistances can be combined into a linear driving force constant, k_{LDF} , given by

$$\frac{1}{k_{LDF}} = \frac{1}{k_S} + \frac{\frac{dQ_{eq}}{dc_S}}{k_F} \approx \frac{1}{k_S} + \frac{\frac{Q_{\infty}}{c_{\infty}}}{k_F}$$
(8)

For the case of strong adsorption, where the accumulation in the macropores is negligible compared to the adsorbed amount, i.e. $\bar{Q}_{eq} = \bar{q}_{eq}$, the LDF model can be solved analytically to obtain

$$\ln\left(\frac{q_{+}}{q_{+}-\bar{q}}\right) + A\ln\left[\frac{q_{-}}{q_{+}}\frac{\left(q_{+}-\bar{q}\right)}{\left(q_{-}-\bar{q}\right)}\right] = k_{LDF}a_{S}t \quad \text{if} \quad 1 + bc_{0} > \alpha bq_{s}$$
(9a)

$$\ln\left(\frac{q_{-}}{q_{-}-\bar{q}}\right) - A\ln\left[\frac{q_{-}}{q_{+}}\frac{(q_{+}-\bar{q})}{(q_{-}-\bar{q})}\right] = k_{LDF}a_{S}t \quad \text{otherwise}$$
(9b)

where

$$A = \frac{\sqrt{\phi^2 - 4\gamma}}{2\phi} + \frac{1}{2} \quad \text{with} \quad \phi = \sqrt{(1 + bc_0 + \gamma)^2 - 4\gamma bc_0}$$
(9c)

and

$$q_{\pm} = \frac{1 + bc_0 + \gamma \pm \phi}{2\alpha b} \quad \text{with} \quad \gamma = \alpha b q_s \tag{9d}$$

with the final equilibrium concentration $\bar{q}_{eq}(\infty) = \bar{q}_{\infty} = q_{-}$. When matching this model to data the uptake curve vs. the dimensionless time is generated varying \bar{q} between 0 and a value close to \bar{q}_{∞} . The regression of the data can be carried out by a simple determination of the scaling factor for the time axis.

1.3 Diffusion in a biporous particle

If one considers the fact that normally materials are optimized for a specific separation, the case of macropore diffusion control is the most likely controlling mass transport mechanism in commercial beads. Here we limit the discussion to macropores where the fluid has the same properties as a bulk liquid.

The mass balance in a spherical bead can be written as

$$\varepsilon_P \frac{\partial c_P}{\partial t} + (1 - \varepsilon_P) \frac{\partial q}{\partial t} = \frac{\varepsilon_P}{\tau} \frac{D_m}{r^2} \frac{\partial}{\partial r} \left(r^2 \frac{\partial c_P}{\partial r} \right) \tag{10}$$

where c_P is the concentration in the fluid inside the pores and q is the concentration in the adsorbed phase. The porosity ε_P

and the tortuosity τ are used to relate the diffusion coefficient to the molecular diffusivity D_m in the fluid phase.

If transport in the microporous material is fast compared to the effective diffusional time constant of the macropores, the adsorbed phase can be considered to be at local equilibrium with c_P and

$$\frac{\partial c_P}{\partial r} = \frac{1}{\frac{dQ}{dc_P}} \frac{\partial Q}{\partial r} \tag{11}$$

and Eq. 10 can be rearranged to

$$\frac{\partial Q}{\partial t} = \frac{1}{r^2} \frac{\partial}{\partial r} \left(\frac{\varepsilon_P}{\tau} \frac{r^2 D_m}{\frac{dQ}{dc_P}} \frac{\partial Q}{\partial r} \right) = \frac{1}{r^2} \frac{\partial}{\partial r} \left(D_P^{Eff} r^2 \frac{\partial Q}{\partial r} \right)$$
(12)

This is coupled to the boundary conditions.

$$\left(\frac{\partial Q}{\partial r}\right)_{r=0} = 0 \quad \text{and} \quad (Q)_{r=R_p} = \bar{Q}_{eq}(t)$$
(13)

Equation 12 shows how a concentration dependent effective pore diffusivity D_p^{Eff} is obtained in this case.

What is important to note is that this model, although mathematically more complex than what seen previously, has only the tortuosity that cannot be measured independently. The adsorption isotherm has to be measured in any case and porosity is a known parameter in commercial materials or can be measured using mercury porosimetry (Lowell et al. 2006). Complexity can be added by including surface diffusion, but this will not be considered here. If surface diffusion can be neglected, once kinetic results for

1.4 Shrinking core model

The assumption in this case is that the process is controlled by diffusion in the external layer of the particle with a rapid reaction that effectively takes place at a surface which moves towards the centre of the particle. While more complex mechanisms can be considered, this is in fact a reasonable model for liquid phase systems with chemisorption. It is adopted here given the fact that also in this case particle porosity and tortuosity are the only physical parameters needed, once the equilibrium concentration of the adsorbed phase is known.

As discussed by Ruthven and Phelps (2001) the assumption of finite volume and a shrinking core lead to the following relationship.

$$\frac{D_S}{R_P^2} \frac{c_0}{Q_S} t = \int_z^1 \frac{z(1-z)}{1-\delta+\delta z^3} dz \quad \text{with} \quad \delta = \frac{V_S Q_S}{V_F c_0} \tag{14}$$

The parameter δ represents the ratio of the moles that are chemisorbed by the solid compared to the initial moles in the system when the final fluid concentration is not zero, which is normally the case in immersion experiments where a stoichiometric excess of adsorbate is used, i.e. $\delta < 1$.

Analytical solutions to the shrinking core model are available for both the infinite system (Brauch and Schlünder 1975) and the finite volume/variable concentration case (Teo and Ruthven 1986; Do 1998). To use a consistent notation the integral can either be calculated numerically by a suitable quadrature formula or from.

$$\int_{z}^{1} \frac{z(1-z)}{1-\delta+\delta z^{3}} dz = \frac{\sqrt[3]{\Lambda}}{6\delta} \left\{ \frac{2}{\sqrt[3]{\Lambda}} \ln\left(1-\delta+\delta z^{3}\right) - 2\ln\left(\frac{\sqrt[3]{\Lambda}+1}{z\sqrt[3]{\Lambda}+1}\right) + \ln\left(\frac{\sqrt[3]{\Lambda^{2}}-\sqrt[3]{\Lambda}+1}{z^{2}\sqrt[3]{\Lambda^{2}}-z\sqrt[3]{\Lambda}+1}\right) + \sqrt{12} \left[\operatorname{atan}\left(\frac{2\sqrt[3]{\Lambda}-1}{\sqrt{3}\sqrt[3]{\Lambda}}\right) - \operatorname{atan}\left(\frac{2z\sqrt[3]{\Lambda}-1}{\sqrt{3}\sqrt[3]{\Lambda}}\right) \right] \right\} \text{ with } \Lambda = \frac{\delta}{1-\delta} = \frac{V_{S}Q_{S}}{V_{F}c_{\infty}} \quad (15)$$

one molecule are interpreted correctly and the tortuosity is determined, the kinetic response of other molecules as well as mixtures can be predicted if the adsorption isotherm is known.

For strongly adsorbed components $\frac{dQ}{dc_p}$ decreases rapidly with increasing concentration. Therefore, in an immersion experiment under highly nonlinear conditions the outer layer of the particle will equilibrate rapidly (high diffusivity), while the inner layer will still have a relatively slow transport process. Under these conditions the kinetic response of this model will be qualitatively very similar to a shrinking core model. This can also be seen considering the Langmuir isotherm in the limit of infinite affinity, where the rectangular isotherm (Ruthven 2000) is obtained. At the end of the experiment $z_{\infty} = 0$ for $\delta < 1$, while for $\delta \ge 1$, the shrinking core will stop at $z_{\infty} = \frac{-1}{\sqrt[3]{\Lambda}}$ and $c_{\infty} = 0$, i.e. all the adsorbate will be consumed. This reaction model will have a finite uptake rate at the end of the experiment.

2 Empirical models

2.1 Pseudo first order model

The pseudo first order (PFO) model is derived from the linear rate model assuming that the equilibrium amount is constant throughout the experiment and equal to the final concentration in the solid

$$\frac{dQ}{dt} = k_1 \left(\bar{Q}_{\infty} - \bar{Q} \right) \tag{16}$$

Here the kinetic constant of the PFO model is explicitly defined as a different parameter from k_{LDF} to avoid confusion between the two models. Only for a rectangular isotherm (irreversible adsorption), \bar{Q}_{eq} is a constant if the accumulation in the macropores is negligible.

Equation 13 can be solved analytically using the initial condition ($\bar{Q}_0 = 0; t_0 = 0$) to obtain

$$\ln\left(1 - \frac{\bar{Q}}{\bar{Q}_{\infty}}\right) = -k_1 t \tag{17a}$$

or

$$\ln(\bar{Q}_{\infty} - \bar{Q}) = \ln(\bar{Q}_{\infty}) - k_1 t \tag{17b}$$

While equivalent these two relationships are shown explicitly because Eq. 17a leads to a linear plot that is constrained to start at the origin, while Eq. 17b leads to a linear plot that is *unconstrained*. This distinction is in fact very important, because if $\ln(\bar{Q}_{\infty})$ is not fixed to the actual equilibrium value the PFO model can be used to fit in a semilogarithmic plot any process which results in an exponential decay to equilibrium. Given that in physisorption any mass transport process will eventually conform to an exponential decay to equilibrium, selecting a portion of the uptake for the unconstrained regression of the parameters will lead to the square of the correlation coefficient, R^2 , close to unity if the points selected are close to the final equilibrium point. Without the test of actually plotting the original data on a linear uptake plot (McLintock 1967), R^2 close to unity has very little value in determining which model applies.

We note also that the only apparent advantage of using the approximate Eq. 17 over Eq. 9 is due to the fact that the first allows to convert the data to a form that can be linearized. The kinetic curve from Eq. 9 can be computed easily by setting the value of $\bar{q}(t)$, which varies from 0 to \bar{q}_{∞} , and calculating the corresponding time. Thus a nonlinear regression of the data with Eq. 9 would be straightforward and to be preferred over the PFO model.

2.2 Pseudo second-order model

The pseudo second order (PSO) model is derived from the quadratic driving force (QDF) model

$$\frac{d\bar{Q}}{dt} = k_{QDF} \left(\bar{Q}_{eq} - \bar{Q}\right)^2 = k_{QDF} \left(\bar{Q}_{eq} - \bar{Q}\right) \left(\bar{Q}_{eq} - \bar{Q}\right)$$
(18)

One should note that Vermeulen suggested the use of the following driving force relationship for irreversible equilibrium (Vermeulen 1953)

$$\frac{d\bar{Q}}{dt} = k_V \left(\bar{Q}_{Eq}^2 - \bar{Q}^2\right) = k_V \left(\bar{Q}_{eq} + \bar{Q}\right) \left(\bar{Q}_{eq} - \bar{Q}\right) \tag{19}$$

Glueckauf showed that this model was the best approximation for nearly rectangular isotherms (Glueckauf 1955), when k_V could be related to the internal diffusion coefficient. Eq. 19 is significantly different from the QDF model and is in fact a concentration dependent LDF model, with the mass transfer coefficient increasing with concentration, which is typically true for systems with a Langmuir isotherm. Note that in a batch adsorption experiment $(\bar{Q}_{eq} + \bar{Q}) \approx 2\bar{Q}_{\infty}$ apart from the very initial uptake, therefore Eq. 19 will not be discussed further.

Equation 18 is a kinetic model that has a rate of adsorption that decreases in time monotonically because $(\bar{Q}_{eq} - \bar{Q})$ is constantly decreasing throughout the experiment. The QDF model can be seen as a concentration dependent LDF model where the mass transfer coefficient goes to zero at equilibrium. While the QDF expression could be physically representative of a particular chemisorption process, in the opinion of the author it is not a physically realistic kinetic model of physisorption, where the mass transfer coefficient is not zero at equilibrium.

If the equilibrium concentration is assumed to be constant and equal to the final value, the PSO model is obtained

$$\frac{d\bar{Q}}{dt} = k_2 \left(\bar{Q}_{\infty} - \bar{Q}\right)^2 \tag{20}$$

This can be solved analytically using the initial condition $(\bar{Q}_0 = 0; t_0 = 0)$ to obtain

$$\bar{Q} = \frac{\bar{Q}_{\infty}^2 k_2 t}{1 + \bar{Q}_{\infty} k_2 t} = \frac{\bar{Q}_{eq}^{\infty} K_2 t}{1 + K_2 t} \quad \text{with} \quad K_2 = \bar{Q}_{\infty} k_2 \tag{21}$$

This shows that there are in fact two possible definitions of the pseudo-second order kinetic constant. It is important to note that if k_2 is indeed a constant, then K_2 would depend on the initial concentration in the fluid and the amount of solid added and vice versa.

Equation 21 has a similar structure to the Langmuir isotherm equation and can be linearized as

$$\frac{1}{\bar{Q}} = \frac{1}{\bar{Q}_{\infty}K_2}\frac{1}{t} + \frac{1}{\bar{Q}_{\infty}}$$
(22)

or

$$\frac{t}{\bar{Q}} = \frac{1}{\bar{Q}_{\infty}K_2} + \frac{t}{\bar{Q}_{\infty}}$$
(23)

Note that any kinetic model that is based on a driving force will conform to Eq. 23 as the system approaches equilibrium. For fast systems limited data will be available far from equilibrium and Eq. 23 will give the false impression of accuracy because it will have R^2 close to unity. Similarly to the PFO model, further incorrect interpretations of results arise when \bar{Q}_{∞} is used as an adjustable parameter.

2.3 Langmuir kinetics

Neglecting the accumulation in the macropores, one can consider adsorption and desorption kinetics (Azizian 2004) expressed as

$$\frac{d\bar{q}}{dt} = k_A c \left(q_S - \bar{q} \right) - k_D \bar{q} \tag{24}$$

where k_A and k_D are the adsorption and desorption kinetic constants. While this appears to be a two parameter model, the two constants are related to the affinity in the Langmuir isotherm. At equilibrium one obtains $k_A = bk_D$. It is useful to rearrange Eq. 24 to

$$\frac{d\bar{q}}{dt} = k_D (1+bc) \left(\bar{q}_{eq} - \bar{q}\right) \tag{25}$$

Equation 25 shows that Langmuir kinetics are similar to the LDF model with a linear concentration dependence of the kinetic coefficient. In an immersion experiment the apparent LDF mass transfer coefficient will decrease in time between $k_D(1 + bc_0)$ and $k_D(1 + bc_\infty)$, consistent with two apparent kinetic regimes. When $bc \ll 1$ the model reduces to Eq. 9 with $k_{LDF}a_S = k_D$.

Azizian's analytical solution (Azizian 2004) can be expressed in a more convenient form using the terms defined in Eq. 9b

$$\frac{1}{\phi} \ln\left[\frac{q_-}{q_+}\frac{(q_+ - \bar{q})}{(q_- - \bar{q})}\right] = k_D t \tag{26}$$

As was the case for the LDF model, also here \bar{q} can be varied between 0 and \bar{q}_{∞} to obtain the corresponding time.

Azizian has shown that Eq. 26 can be used to show convergence to either PFO or PSO models depending on the concentrations in the fluid (Azizian 2004). The main result was to show that both models could match the more general model but resulted in concentration dependent k_1 values (linear dependence on c_0) and k_2 (more complex dependence on c_0). Here the approach is quite different, because the aim is to warn against use of these models in systems where the transport mechanisms are known to be <u>not</u> Langmuir kinetics.

2.4 The Elovich model

The Elovich model is discussed in detail by Low (1960) and Aharoni and Ungarish (1977), who also consider the possible derivation from reaction kinetic models and the fact that it does not conserve the number of surface sites, i.e. it is not compatible with the assumptions of Langmuir kinetics. The general expression for the rate is

$$\frac{d\bar{q}}{dt} = k_E e^{-B\bar{q}} \tag{27}$$

This expression is not a "driving force" model since the rate is zero only for $\bar{q} = \infty$. This is not a feasible model for physisorption because the rate will not be zero at equilibrium. In the simulation of an adsorption process one would need to couple Eq. 27 with the equilibrium isotherm and switch to $\frac{d\bar{q}}{dt} = 0$ when the equilibrium concentration is achieved. As such this model should only be used for chemisorption and should give a shape of the response similar to the shrinking core model.

The solution to Eq. 27 is

$$\frac{1}{B}e^{B\bar{q}} = k_E t \tag{28}$$

This can be re-arranged into.

$$\bar{q} = \frac{1}{B} \ln \left(Bk_E t + 1 \right) = \frac{1}{B} \ln \left(Bk_E \right) + \frac{1}{B} \ln \left(t + t_0 \right) \quad t_0 = \frac{1}{Bk_E}$$
(29)

In this case the data are time shifted varying t_0 to obtain a linear plot and an *unconstrained* linear regression is then applied (Cheung et al. 2000).

Qualitatively Eq. 27 corresponds to an initially fast rate, k_E , that progressively decreases. As such one can intuitively suggest that a time-shifted diffusion process should give a linear Elovich plot in a region not close to final equilibrium. The qualitative behaviour away from the final equilibrium is not dissimilar to the PSO model and Langmuir kinetics.

2.5 Linear diffusion model

This model assumes linearity even when the system follows a nonlinear isotherm like the Langmuir isotherm and the immersion experiment is carried out far from the region where Henry's law applies. As such it is included in the list of empirical models, especially when the short time linear plot that results from this model is used to fit batch adsorption experiments.

For the case where the external film resistance is negligible an analytical solution is readily available (Do 1998; Ruthven 1984; Crank 1975) and is given by

$$\frac{\bar{Q}}{Q_{\infty}} = 1 - \sum_{n} \frac{6 \exp\left(-\beta_{n}^{2} \frac{D_{p}^{EL}}{R_{p}^{2}}t\right)}{9\lambda^{2} + 9\lambda + \beta_{n}^{2}}$$
(30a)

where

$$\frac{D_P^{EL}}{R_P^2} = \frac{\varepsilon_P}{\tau} \frac{D_m}{K}$$
(30b)

and

$$\beta_n \cot \beta_n - 1 - \frac{1}{3\lambda} \beta_n^2 = 0 \quad \text{with} \quad \lambda = \alpha K$$
 (30c)

and

$$K = \frac{\Delta Q}{\Delta c} = \frac{Q_{\infty}}{c_{\infty}} = \epsilon_P + \left(1 - \epsilon_P\right) \frac{q_S b}{1 + bc_{\infty}}$$
(30d)

In a truly linear system K becomes the Henry law constant of the particle when $bc_{\infty} \ll 1$.

Equation 30c has a root in each π interval, except the first and requires only a few terms to converge apart from the region close to t = 0. It is therefore useful also to derive the short time expression valid in this region

$$\frac{\bar{Q}}{Kc_{\infty}} = \frac{\bar{Q}}{Kc_0} \frac{c_0}{c_{\infty}} \approx \frac{\bar{Q}_0}{Kc_0} \frac{c_0}{c_{\infty}} \approx (1+\lambda)6\sqrt{\frac{D_P^{EL}}{R_P^2}\frac{t}{\pi}}$$
(34)

The other important difference is the final approach to equilibrium where

$$1 - \frac{\bar{Q}}{Q_{\infty}} \approx \frac{6}{9\lambda^2 + 9\lambda + \beta_1^2} \exp\left(-\beta_1^2 \frac{D_P^{EL}}{R_P^2} t\right)$$
(35)

$$\frac{\bar{Q}}{\bar{Q}_{\infty}} \approx (1+\lambda) \left[6\sqrt{\frac{D_{P}^{EL}}{R_{P}^{2}} \frac{t}{\pi}} - 3(1+3\lambda) \frac{D_{P}^{EL}}{R_{P}^{2}} t + 12\lambda(1+3\lambda) \frac{D_{P}^{EL}}{R_{P}^{2}} t \sqrt{\frac{D_{P}^{EL}}{R_{P}^{2}} \frac{t}{\pi}} + \dots \right]$$
(31)

Equation 31 suggests the plot of the dimensionless uptake response vs. \sqrt{t} , which should be linear for short times.

While the finite volume solution is readily available, as a first approximation one could consider using the solution to the diffusion equation subject to a step change in surface concentration, which corresponds to the limiting case of $\lambda = 0$ and $\beta_n = n\pi$. There are however important differences between Eq. 30 and

$$\frac{\bar{Q}_0}{Kc_0} = 1 - \sum_n \frac{6\exp\left(-n^2 \pi^2 \frac{D_p^{App}}{R_p^2}t\right)}{n^2 \pi^2}$$
(32)

and the corresponding

$$\frac{\bar{Q}_0}{Kc_0} \approx 6\sqrt{\frac{D_P^{App}}{R_P^2}\frac{t}{\pi}} - 3\frac{D_P^{App}}{R_P^2}t + 0\frac{D_P^{App}}{R_P^2}t\sqrt{\frac{D_P^{App}}{R_P^2}\frac{t}{\pi}} + \dots$$
(33)

The short time behaviour differs by a coefficient $1 + \lambda$, therefore the apparent diffusivity, D_P^{App} , will be higher by a factor $(1 + \lambda)^2$ compared to D_P^{EL} and will depend on both the isotherm and the volume of solid added to the solution since $\lambda = \frac{V_S}{V_F}K$. Note also that the third term of this series expansion is zero which indicates that the finite volume solution deviates from the initial linear trend more rapidly than the constant concentration solution.

To see more clearly the origin of this difference consider that for short times the external concentration is close to c_0 and the fact that in the constant concentration case $Q_{\infty} = Kc_0$. The initial trend should be similar, but the normalization to 1 is different. Writing this explicitly

$$1 - \frac{\bar{Q}_0}{Kc_0} \approx \frac{6}{\pi^2} \exp\left(-\pi^2 \frac{D_P^{App}}{R_P^2} t\right)$$
(36)

From Eq. 30c it is possible to see that $1 < \frac{\beta_1}{\pi} < 1.4302$ as λ varies between 0 and ∞ . The long-time asymptote will yield an apparent diffusivity that will be up to double the value of D_p^{EL} and different from the value obtained from the short time uptake, Eq. 33. The incorrect use of the constant concentration equations, could therefore lead to identifying two apparent kinetic regimes for strongly adsorbed species.

What limits further the validity of results found in the literature is the fact that uptake data are regressed using the short time \bar{Q} vs. \sqrt{t} with the empirical expression

$$\bar{Q} \approx A_D \sqrt{t} + C \tag{37}$$

which is often attributed to Weber and Morris (1963) but no distinction is made with respect to the assumption of constant volume or constant concentration. Note also that often Eq. 37 is applied to separate segments of the uptake curve, linking this to the presence of an initial fast process followed by a much slower second kinetic regime. It is not clear how such an interpretation originated, but the \bar{Q} vs. \sqrt{t} plot should only be used in the very first part of the uptake, i.e. for $\bar{Q} < 0.5$.

Boyd et al. (1947a) are also often referenced as the basis for Eq. 37. It is very instructive to read this reference in detail and realise the efforts made to achieve conditions of constant external concentration and linear equilibrium, which are the basis for the model adopted in that study. Boyd's system was not a fixed volume system, but a shallow bed with once-through flow of the solution. This ensured constant external concentration, while the measurement of tracer exchange ensured linearity of the equilibrium isotherm. Using radioactive tracers the adsorbed amounts were measured directly from the particles, not calculated indirectly from the external concentration. Provided that the concentration change across the shallow bed is small, Eq. 32 or the equivalent for diffusion and external film resistance is perfectly valid. The shallow bed with varying concentration in linear conditions would be similar to a liquid phase zero length column system (Brandani and Ruthven 1995). Boyd et al. confirmed that the uptake of different ions on an Amberlite ion-exchange resin could be either controlled by internal diffusion (Eq. 32) for strongly adsorbed species or a film resistance (Eq. 7 or the PFO model). This is consistent with an internal mechanism of macropore diffusion control leading to a strongly concentration dependent internal diffusion coefficient.

The series of papers by Boyd et al. are still today an excellent example of how tocharacterize rigorously the equilibrium properties (1947b) in ion-exchange processes and then carry out careful kinetic experiments (Boyd et al. 1947a) and subsequent breakthrough experiments (Boyd et al. 1947c), checking the independent measurements for consistency. Their aim was the design of ion-exchange separation processes and anything more advanced than what Boyd et al. presented would have required numerical computational power not available between 1943 and 1946 when their work was carried out. Today, one would possibly consider extending the analysis introducing diffusion models that include the interactions between charged species (Wesselingh et al. 1995; Krishna and Wesselingh 1997), but this will not be discussed further as it is beyond the scope of this contribution.

3 Results and discussion

In the description of the empirical models the qualitative discussion of potential pitfalls of the use of linearized plots to extract kinetic information was introduced. To give a quantitative setting to the arguments made, specific examples will be presented based on generating "data" from the full nonlinear model and then applying the empirical approaches to see if the models appear to fit the data well and if any of the models produce parameters that could be considered to be physically meaningful.

3.1 Physical adsorption on zeolites: xylenes on Y zeolite beads

The first in a series of three papers by Santacesaria et al. (1982a) covered immersion experiments to obtain equilibrium properties and formulate a mass transport model based

Table 1	The parameters	used	in	this	study	,
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Parameters	p-xylene	m-xylene	
$\overline{\varepsilon_{P}}$	0.20	0.20	
τ	2.15	2.15	
$R_P (\mathrm{mm})$	0.65	0.65	
$D_m \times 10^9 (\mathrm{m}^2 \mathrm{s}^{-1})$	2.15	2.15	
$k_F \times 10^6 ({\rm m \ s^{-1}})$	1×10^{9a} , 3.31	$1 \times 10^{9a},$ 3.31	
$c_0 ({\rm mol} \;{\rm m}^{-3})$	250, 500	250, 500	
$\rho_S (\mathrm{kg} \mathrm{m}^{-3})$	1400	1400	
$M_S \times 10^3 (\text{kg})$	10, 40	10, 40	
$V_F \times 10^6 (\mathrm{m})$	200	200	
$q_S \pmod{\mathrm{m}^{-3}}$	2450	2450	
$b (\mathrm{m}^3 \mathrm{mol}^{-1})$	0.036	0.006	

^aArbitrary large value to obtain pure diffusion limit. The lower value corresponds to Sh=2



Fig. 2 Batch adsorption uptake curves for p-xylene ($c_0 = 500 \text{ mol m}^{-3}$ and $M_S = 40 \text{ gr}$) and m-xylene ($c_0 = 250 \text{ mol m}^{-3}$ and $M_S = 40 \text{ gr}$). Full lines diffusion model (Sh = ∞); dashed lines combined model and Sh = 2

on combined film resistance and diffusion inside the beads (Table 1).

The supplementary information includes all the results. Here we discuss in greater detail the case of pure diffusion for p-xylene ($c_0 = 500 \text{ mol m}^{-3}$ and $M_s = 40 \text{ gr}$) and m-xylene ($c_0 = 250 \text{ mol m}^{-3}$ and $M_s = 40 \text{ gr}$). p-Xylene and m-xylene were chosen because they are the strongest and weakest adsorbate respectively.

While these are specific examples, the shape of the kinetic response of macropore diffusion control with a Langmuir adsorption isotherm will always conform to being between close to linear (m-xylene example) and strongly nonlinear (p-xylene example). A measure of the nonlinearity is the ratio of the equilibrium concentration of the solid to the value at complete saturation of the adsorbed phase (Brandani 1998), which in this case is given by

$$\Gamma = \frac{Q_{\infty}}{\varepsilon_P c_{\infty} + (1 - \varepsilon_P) q_S}$$
(38)

A value of $\Gamma < 0.5$ is an indication of mild nonlinearity (Brandani 1998). The nonlinearity increases as Γ approaches 1.

Figure 2 shows the uptake curves for the two cases of interest. Qualitatively the shapes of the curves are similar, even when nonlinearity is considered. As expected, the surface resistance modifies the take-off from zero and in general results in a smoother shape. It is interesting to note though that overall, on the uptake plot, the worst case scenario, Sh = 2, is not significantly different from the case of pure diffusion, $Sh = \infty$. Therefore particular care should be used in analysing results with the combined film resistance and diffusion model when experimental scatter will give a large uncertainty on the individual values of the two mass transfer resistances, but the overall effect should still be estimated with sufficient accuracy. To discriminate better between the two mechanisms it would be better to modify the configuration of the experimental system so that the *relative* velocity between the particles and the fluid can be controlled directly. Examples of these systems are discussed in Kärger and Ruthven (1992).

Figure 3 shows the simulated kinetic response for p-xylene, $\Gamma = 0.90$. The data between 3 and 25 min are used to apply the PFO, PSO and Elovich empirical approaches. Figure 3 shows the relevant plots along with the linear regression trend-line. In the Elovich kinetics case one has to plot the data varying t_0 , which introduces a small potential bias from the analyst. The data with $t_0 = 45$ s, conform well to the linear plot, even though there are no chemical reactions involved

The order in which one would rank the models in terms of goodness of fit is PFO > PSO > Elovich. There is very little to discriminate between the three given that the range of R^2 is between 0.9982 and 0.9993. Experimental scatter of the data would reduce R^2 , possibly modifying the order. What is important to emphasise though is that the trend observed is perfectly consistent with the qualitative observations made previously. All models *fit* well the data *in the linearized plots* even though they do not represent the correct transport phenomena.

Figure 3 shows also the test of the fits (McLintock 1967), where the regression parameters are used to generate the actual uptake curves to compare the models to the original data. We now see that the PFO model is in fact a very poor match to the data, whether the equilibrium concentration from the fit is used or the actual equilibrium concentration



Fig. 3 Linearized plot and linear regression of data using the PFO, PSO and Elovich models. Filled points are those used in establishing the trendline. Lower right plot shows the linear plot of the uptake data with the curves calculated from the parameters obtained from

the *unconstrained* linear regression. For the PFO model an additional curve is shown with the correct final concentration. p-Xylene $c_0 = 500 \text{ mol m}^{-3}$ and $M_{\text{S}} = 40 \text{ gr}$



Fig. 4 Regression of uptake data based on linear diffusion model. Filled points are those used in establishing the trendline. Uptake predicted using Eq. 30 passing through the first two points. p-Xylene $c_0 = 500 \text{ mol m}^{-3}$ and $M_S = 40 \text{ gr}$

is used. In the long time region the PSO model appears to be the best match, but in the short time deviations close to 100% can be observed with the actual response being much faster. The Elovich kinetic model matches the initial part well but as expected cannot describe the final approach to equilibrium. This is in fact the *best* model if one decouples the kinetics from the final equilibrium. Plotting the uptake data on linear scales would lead to the incorrect conclusion that one of the two physically unrealistic models describes this system well. It should be clear in this case that the parameters of these models are not physical parameters and therefore of very limited portability to other systems, if any.

Figure 4 shows the \bar{Q} vs. \sqrt{t} plot of the data. Here only the data over the first 10 min are used in the linear regression. Again a high R² is obtained, 0.9896, but this is the lowest for all models. It is worth emphasizing again that the fact that R² is close to unity cannot be used to identify the mechanism of mass transport. If anything, this measure leads to the wrong conclusion, given that in this case it would be better to use the linear diffusion model, even though the match to the data is not as good. This crucial consideration can be understood if the particle size is changed and only the diffusion model would capture the quadratic dependence of the kinetic time constant on the radius of the particles. The empirical models are only *correlative* and not *predictive*, they cannot be used for conditions different from those used in the actual experiment from which the parameters were obtained.

The intercept of the linear regression in Fig. 4 is a significant deviation from 0 and this would be interpreted as evidence of combined surface resistance and internal diffusion (Câmara et al. 2020; Kong et al. 2016; Lupul et al. 2015; Peralta et al. 2019; Rincon-Silva et al. 2016; Shen et al. 2018; Heyden et al. 2018), even though if this was the case the intercept should be negative. The positive intercept



Fig. 5 Internal concentration profiles after 3 and 30 min. Note strong nonlinearity as evidenced by the sharp inflection in the profiles. p-Xylene $c_0 = 500 \text{ mol m}^{-3}$ and $M_S = 40 \text{ gr}$

is actually a result of the fact that the points used are already outside the range where the first term in the series expansion is sufficient to describe the response. Assuming a contribution from film resistance leads again to identifying the incorrect transport mechanism.

Additional information can be inferred when the slope of the short time regression is converted to a diffusional time constant

$$\frac{R_P^2}{D_P^{App}} = \left(\frac{6Q_\infty}{A_D\sqrt{\pi}}\right)^2 \tag{39}$$

to obtain approximately 289 min. The PFO fit of the data can also be interpreted as the long-time asymptote of the linear diffusion equation. From the slope of the PFO regression curve and Eq. 36, it is possible to estimate a value of the apparent time constant of approximately 104 min. This simple check would allow to confirm that the system is strongly non-linear and that the data should be analysed with the full model.

The nonlinearity of this system is also evident from the comparison of the data and the linear model that includes variable fluid concentration, Eq. 30. This comparison, shown in Fig. 4, provides much more information than the empirical linear regression and should always be preferred over the linear short time approach. The practical barrier to generating the curve from Eq. 30 are the roots of Eq. 30c. The simplest and most robust method of generating the β_n values is the bisection method applied in each π interval after the first, given that the sign of the function is always positive to the left of the root and negative to the right. Any solver of a single nonlinear equation will also work, provided that the search for the solution is kept within the correct π interval. Note that λ is the ratio of the moles in the solid to those in the fluid at equilibrium, and is a constant which allows

the calculation of β_n before the attempt to estimate the time constant is carried out. Therefore, rate of convergence to a solution is not an issue and robustness is key.

Figure 5 shows the internal concentration profiles at 3 and 30 min. Note how the profiles are qualitatively similar to a shrinking core model. On close inspection there are however two main differences: the sharp but continuous transition to the initial concentration in the inner part and the fact that the plateau close to the surface shifts down with time because of the decrease in concentration in the fluid phase, i.e. there is desorption taking place in the outer layer of the particle.

Without a detailed analysis of the kinetic response or a comparison with the full solution of the diffusion equation, the strong nonlinearity could be detected from results over a range of adsorbed phase concentrations, always taking into account the finite volume and variable fluid concentration during the experiment.

Given the importance of the nonlinearity of the isotherm on the resulting kinetic responses, an alternative that could be useful to consider would be a desorption experiment. This could be carried out by allowing the solution to equilibrate first and then adding a known quantity of solvent after the adsorption step. Note that this additional experiment is not common, but would allow to distinguish between reversible and irreversible adsorption. It would be recommended when enough sensitivity is available in the concentration detector.



Fig. 7 Internal concentration profiles after 3 and 30 min. m-Xylene $c_0 = 250 \text{ mol m}^{-3}$ and $M_S = 40 \text{ gr}$

Figure 6 shows the results for m-xylene, $\Gamma = 0.44$. Here the nonlinearity is mild as can be seen from the concentration profiles shown in Fig. 7. Also in this case the trends observed match closely the observations made for the p-xylene example. Again R² values close to unity are obtained for all the *unconstrained* linear regressions and the values range from 0.996 to 0.999. When the linear uptake plot is used to check results, the PFO model is seen to match poorly the data. The PSO and Elovich kinetic models would be chosen to represent this system.



Fig. 6 Linearized plot and linear regression of data using the PFO, PSO and Elovich models. Filled points are those used in establishing the trendline. Lower right plot shows the linear plot of the uptake data with the curves calculated from the parameters obtained from

the *unconstrained* linear regression. For the PFO model an additional curve is shown with the correct final concentration. m-Xylene $c_0 = 250 \text{ mol m}^{-3}$ and $M_S = 40 \text{ gr}$

It is possible to see now that the PSO and Elovich models will match the pure diffusion uptake curves with both strong and mild nonlinearity in the isotherm. This simply reflects the nature of the models that start with a fast rate and approach equilibrium with a significantly lower rate. This is a result of the *shape* that these empirical kinetic models can reproduce and the fact that this is consistent with the diffusion model and what happens in the batch adsorption experiment. However, both empirical models are not physically meaningful for physisorption and should not be used to correlate the data.

Figure 8 shows the comparison of the data and the linear diffusion model. Equation 30 is now seen to match closely the data over the entire kinetic response. In this case the linear model would provide an accurate method to determine the tortuosity of the particles.

The addition of a film resistance to mass transport will affect primarily the short time response where the rate is faster. The worst case scenario for external film resistance can be estimated from the Sherwood number set to 2, i.e. $k_F = \frac{D_m}{R_p}$. The uptake curves for the diffusion with and without film resistance cases are qualitatively similar as shown in Fig. 2. Therefore the same trends already observed will apply to the empirical models. The only noticeable difference is the more gradual transition from a fast initial rate to a slower final approach to equilibrium, and this in turn will favour the PSO model for nonlinear systems which becomes the model of choice based on the goodness of fit. Figure 9 shows the uptake response for the p-xylene case with Sh=2along with the match to the Langmuir kinetics and LDF models. Due to the effect of the variation in fluid concentration on the apparent mass transfer coefficient, the Langmuir kinetics model is a very good match to the data and the best



Fig.8 Regression of uptake data based on linear diffusion model. Filled points are those used in establishing the trendline. The model shown is calculated from Eq. 30. m-Xylene $c_0 = 250 \text{ mol m}^{-3}$ and $M_S = 40 \text{ gr}$



Fig. 9 Uptake curves for p-Xylene $c_0 = 500 \text{ mol m}^{-3}$ and $M_S = 40 \text{ gr}$ and Sh=2. Langmuir kinetics and LDF models are calculated adjusting q_S to give the same fluid phase concentration at equilibrium. The PSO model is calculated from the parameters obtained from the *unconstrained* linear regression

model in Fig. 9. Even by eye one can see that the R^2 is very close to unity, but again there are no reactions taking place and Langmuir kinetics is not the physical mechanism of mass transport. The excellent match to the data is only the result of the *shape* of the uptake curves obtained from Langmuir kinetics.

Very similar results are obtained for the m-xylene case with Sh=2. This shows that a diffusion process with or without some surface film resistance can be incorrectly interpreted to conform to the PSO, Elovich and Langmuir kinetic models simply due to the characteristic shape of the uptake curves in a batch adsorption experiment. Figure 10 shows the \bar{Q} vs. \sqrt{t} plot of the data for m-xylene with Sh=2.

Note that in Fig. 10 the intercept is negative, as one would expect in this case. A recent example of Cu^{++} kinetics in a coke derived porous carbon (Yuan et al. 2019) shows R^2



Fig. 10 Regression of uptake data based on linear diffusion model. Filled points are those used in establishing the trendline. m-Xylene $c_0 = 250 \text{ mol m}^{-3}$ and $M_S = 40 \text{ gr and Sh} = 2$

values close to unity for the *unconstrained* linear regressions from all empirical models considered here, with the Elovich being the worst. The data also lead to a negative intercept of the linear diffusion regression. All these results seem to suggest that this particular system was in fact limited primarily by a film resistance, especially because the PFO model shows a calculated equilibrium concentration that is close to the actual measured value. This is very similar to the case p-Xylene $c_0 = 500 \text{ mol m}^{-3}$, $M_s = 10 \text{ gr}$, $\Gamma = 0.94$ and Sh=2 shown in fig. S-2.

We turn now briefly to a case where adsorption is irreversible and generate the data with the shrinking core model. There were no substantial differences observed compared to the p-xylene case, therefore the curves and linear regressions are shown in the Supplementary Information for one of the experiments reported by Phelps and Ruthven (2001). The linear regressions give R^2 values of 0.9978, 0.9989, 0.9959 for the PFO, PSO and Elovich models respectively. Also in this case there is very little to discriminate between the models based on this figure of merit. Figure 11 shows the uptake curve and the calculated responses, including Langmuir kinetics. There is very little difference between the shapes of the PSO and the Langmuir kinetics models. The *unconstrained* regression of the parameters of the PFO model is shown to be the worst model.

Based on Fig. 11 the PSO model would be selected as the kinetic model in this case, but again this would be inconsistent with the actual mechanism, given that Phelps and Ruthven report experiments carried out with different bead sizes that show that for this system uptake is a diffusion controlled process, which at high concentrations is well described by the shrinking core model (Phelps and Ruthven 2001).



Fig. 11 Uptake curves for Cu⁺⁺ in Ionac SR-5 Cation Exchange Resin. $C_0 = 0.0059$ gr Cu/ml; $V_F = 100$ ml; $M_S = 3.6$ gr; $Q_S = 0.09$ gr Cu/gr resin. Langmuir kinetics are calculated setting $b = 10^6$. The other models are calculated from the parameters obtained from the *unconstrained* linear regression

4 Conclusions and recommendations

The analysis carried out has shown that empirical models, that lead to linear regression of uptake data from batch adsorption experiments, can match the data well if the goodness of fit is based on how close to unity R^2 is. This can be very misleading especially if *unconstrained* regression of data is applied, particularly for the pseudo first order model. The models and the data should always be shown on a full uptake plot to avoid the incorrect interpretation of the results.

For two nonlinear models, namely macropore controlled diffusion in beads with a Langmuir isotherm and the shrinking core model, the empirical approaches lead to the incorrect identification of the mechanism if only one uptake experiment is carried out. Experiments in a wide range of conditions will result in coefficients that vary significantly, generating a confusing picture of the kinetic behaviour. This can lead to inefficient experimentation, given that the effect of many different parameters has to be investigated, even when the actual physical mechanism contains only one or two parameters that need to be determined.

When uptake curves conform to the pseudo second order model and the pseudo first order model has an R^2 close to unity, but a low calculated equilibrium amount, it is likely that a diffusion model coupled with an appropriate isotherm will match the data very well. It would be important to carry out experiments with particles of different sizes to confirm that this is indeed the case.

When using the linear plot based on the solution of the diffusion equation for short times, a positive intercept is an indication that the initial rate of uptake is fast and that the data selected are already beyond the range of validity of the approximate model. The resulting apparent diffusivity will be lower than the true value. It is preferable to use the full solution, Eq. 30, in this case. Furthermore, identifying multiple linear regions in the \bar{Q} vs. \sqrt{t} plot is *inconsistent* with the diffusion model.

All the approximate models should be used only to arrive at an initial estimate of the mass transfer coefficients. This may be sufficient if the scope of the investigation is a direct qualitative comparison between materials tested in the same experimental system.

For a more accurate determination of the mass transfer coefficients the full nonlinear models should be used with the measured equilibrium isotherm parametrised independently. Furthermore, given that the experiment relies on measuring the variation of the fluid concentration in time, models based on the assumption of constant fluid concentration should not be used.

Phenomenological models should be preferred over empirical kinetic expressions. This is particularly important when the scope of the investigation is the measurement of kinetic responses to aid the design of separation processes or reactors. In such applications the models used and the corresponding parameters have to be applicable also to multicomponent mixtures and to conditions that can be far from those explored in the batch experiments.

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References

- Aharoni, C., Ungarish, M.: Kinetics of activated chemisorption. Part 2. Theoretical models. J. Chem. Soc. Faraday Trans. I 73, 456–464 (1977)
- Azizian, S.: Kinetic models of sorption: a theoretical analysis. J. Colloid Interface Sci. 276, 47–52 (2004)
- Babić, K., van der Ham, L.G.J., de Haan, A.B.: Sorption kinetics for the removal of aldehydes from aqueous streams with extractant impregnated resins. Adsorption 14, 357–366 (2008)
- Bartczak, P., Zołtowska, S., Norman, M., Klapiszewski, Ł, Zdarta, J., Komosa, A., Kitowski, I., Ciesielczyk, F., Jesionowski, T.: Sawsedge *Cladium mariscus* as a functional low-cost adsorbent for effective removal of 2,4-dichlorophenoxyacetic acid from aqueous systems. Adsorption 22, 517–529 (2016)
- Bazargan, A., Shek, T.-H., Hui, C.-W., McKay, G.: Optimising batch adsorbers for the removal of zinc from effluents using a sodium diimidoacetate ion exchange resin. Adsorption 23, 477–489 (2017)
- Boyd, G.E., Adamson, A.W., Myers, L.S. Jr.: The exchange adsorption of ions from aqueous solutions by organic zeolites. II. Kinetics. J. Am. Chem. Soc. 69, 2836–2848 (1947a)
- Boyd, G.E., Schubert, L., Adamson, A.W.: The exchange adsorption of ions from aqueous solutions by organic zeolites. I. Ion-exchange equilibria. J. Am. Chem. Soc. 69, 2818–2829 (1947b)
- Boyd, G.E., Myers, L.S. Jr., Adamson, A.W.: The exchange adsorption of ions from aqueous solutions by organic zeolites. III. Performance of deep adsorbent beds under non-equilibirum conditions. J. Am. Chem. Soc. 69, 2849–2859 (1947c)
- Brandani, S., Ruthven, D.M.: Analysis of ZLC desorption curves for liquid systems. Chem. Eng. Sci. **50**, 2055–2059 (1995)
- Brandani, S.: Effect of nonlinear equilibrium on ZLC experiments. Chem. Eng. Sci. 53, 2791–2798 (1998)
- Brauch, V., Schlünder, E.U.: The scale-up of activated carbon columns for water purification, based on results from batch tests—II. Theoretical and experimental determination of breakthrough cures in activated carbon columns. Chem. Eng. Sci. 30, 539–548 (1975)
- Câmara, A.B.F., Sales, R.V., Bertolino, L.C., Furlanetto, R.P.P., Rodríguez–Castellón, E., de Carvalho, L.S.: Novel application for palygorskite clay mineral: a kinetic and thermodynamic assessment of diesel fuel desulfurization. Adsorption 26, 267–282 (2020)

- Cheung, C.W., Porter, J.F., McKay, G.: Elovich equation and modified second-order equation for sorption of cadmium ions onto bone char. J. Chem. Technol. Biotechnol. **75**, 963–970 (2000)
- Ciesielczyk, F., Bartczak, P., Jesionowski, T.: Removal of cadmium(II) and lead(II) ions from model aqueous solutions using sol-gelderived inorganic oxide adsorbent. Adsorption 22, 445–458 (2016)
- Crank, J.: The Mathematics of Diffusion, 2nd edn. Oxford University Press, Oxford (1975)
- De Smedt, C., Spanoghe, P., Biswas, S., Leus, K., van der Voort, P.: Comparison of different solid adsorbents for the removal of mobile pesticides from aqueous solutions. Adsorption 21, 243– 254 (2015)
- Do, D.D.: Adsorption Analysis: Equilibria and Kinetics. Imperial College Press, London (1998) Chapter 15.
- Dominguez, M.A., Etcheverry, M., Zanini, G.P.: Evaluation of the adsorption kinetics of brilliant green dye onto a montmorillonite/ alginate composite beads by the shrinking core model. Adsorption 25, 1387–1396 (2019)
- Everett, D.H.: Reporting data on adsorption from solution at the solid/ solution interface (recommendations 1986). Pure Appl. Chem. 58, 967–984 (1986)
- Fernandez, M.E., Bonelli, P.R., Cukierman, A.L., Lemcoff, N.O.: Modeling the biosorption of basic dyes from binary mixtures. Adsorption 21, 177–183 (2015)
- Glueckauf, E.: Theory of chromatography. Part 10. Formulae for diffusion into spheres and their application to chromatography. Trans. Faraday Soc. 51, 1540–1551 (1955)
- Ho, Y.S., McKay, G.: Sorption of dye from aqueous solution by peat. Chem. Eng. J. **70**, 115–124 (1998)
- Hubetska, T., Kobylinska, N., García, J.R.: Efficient adsorption of pharmaceutical drugs from aqueous solution using a mesoporous activated carbon. Adsorption 26, 251–266 (2020)
- Kärger, J., Ruthven, D.M.: Diffusion in Zeolites and Other Microporous Solids, pp. 272–274. Wiley, New York (1992)
- Kong, D., Zheng, X., Tao, Y., Lv, W., Gao, Y., Zhi, L., Yang, Q.H.: Porous graphene oxide-based carbon artefact with high capacity for methylene blue adsorption. Adsorption 22, 1043–1050 (2016)
- Krishna, R., Wesselingh, J.A.: The Maxwell-Stefan approach to mass transfer. Chem. Eng. Sci. 52, 861–911 (1997)
- Low, M.J.D.: Kinetics of chemisorption of gases on solids. Chem. Rev. 60, 267–312 (1960)
- Lowell, S., Shields, J.E., Thomas, M.A., Thommes, M.: Characterization of Porous Solids and Powders: Surface Area, Pore Size and Density. Springer, Dordrecht (2006)
- Lupul, I., Yperman, J., Carleer, R., Gryglewicz, G.: Adsorption of atrazine on hemp stem-based activated carbons with different surface chemistry. Adsorption 21, 489–498 (2015)
- Marques Fraga, T.J.M., Bandeira de Souza, Z.S., dos Santos Marques Fraga, D.M., Nascimento Carvalho, M., de Luna Freire, P., Ghislandi, E.M.M.G., da Motta Sobrinho, M.A.: Comparative approach towards the adsorption of Reactive Black 5 and methylene blue by n-layer graphene oxide and its amino–functionalized derivative. Adsorption 26, 283–301 (2020)
- McLintock, J.S.: The Elovich equation in chemisorption kinetics. Nature **216**, 1204–1205 (1967)
- Morales-Perez, A.A., Arias, C., Ramırez-Zamora, R.-M.: Removal of atrazine from water using an iron photo catalyst supported on activated carbon. Adsorption **22**, 49–58 (2016)
- Peralta, M.E., Nisticò, R., Franzoso, F., Magnacca, G., Fernandez, L., Parolo, M.E., García León, E., Carlos, L.: Highly efficient removal of heavy metals from waters by magnetic chitosan–based composite. Adsorption 25, 1337–1347 (2019)
- Phelps, D.S.C., Ruthven, D.M.: The kinetics of uptake of Cu⁺⁺ ions in Ionac SR-5 cation exchange resin. Adsorption 7, 221–229 (2001)

- Phelps, D.S.C., Ruthven, D.M.: Mass transfer performance of an endless belt counter-current ion exchanger for liquid phase adsorption. Sep. Purif. Technol. 27, 243–256 (2002)
- Popugaeva, D., Manoli, K., Kreyman, K., Ray, A.K.: Removal of aluminum from aqueous solution by adsorption on montmorillonite K10, TiO₂, and SiO₂: kinetics, isotherms, and effect of ions. Adsorption **25**, 1575–1583 (2019)
- Rincon-Silva, N.G., Moreno-Pirajan, J.C., Giraldo, L.: Equilibrium, kinetics and thermodynamics study of phenols adsorption onto activated carbon obtained from lignocellulosic material (*Eucalyptus globulus* labill seed). Adsorption **22**, 33–48 (2016)
- Ruthven, D.M.: Principles of Adsorption and Adsorption Processes. Wiley, New York (1984)
- Ruthven, D.M.: The rectangular isotherm model for adsorption kinetics. Adsorption 6, 287–291 (2000)
- Santacesarla, E., Morbidelli, M., Danise, P., Mercenari, M., Carrá, S.: Separation of xylenes on Y zeolites. 1. Determination of the adsorption equilibrium parameters, selectivities, and mass transfer coefficients through finite bath experiments. Ind. Eng. Chem. Process Des. Dev. 21, 440–445 (1982a)
- Santacesarla, E., Morbidelli, M., Servida, A., Storti, G., Carrá, S.: Separation of xylenes on Y zeolites. 2. Breakthrough curves and their interpretation. Ind. Eng. Chem. Process Des. Dev. 21, 446–451 (1982b)
- Santacesarla, E., Morbidelli, M., Servida, A., Storti, G., Carrá, S.: Separation of xylenes on Y zeolites. 3. Pulse curves and their interpretation. Ind. Eng. Chem. Process Des. Dev. 21, 451–457 (1982c)
- Shen, J., Franchi Evangelista, M., Mkongo, G., Wen, H., Langford, R., Rosair, G., McCoustra, M.R.S., Arrighi, V.: Efficient

defluoridation of water by Monetite nanorods. Adsorption 24, 135–145 (2018)

- Teo, W.K., Ruthven, D.M.: Adsorption of water from aqueous ethanol using 3-Å molecular sieves. Ind. Eng. Chem. Process Des. Dev. 25, 17–21 (1986)
- van der Heyden, S.R.H., van Reppelen, K., Yperman, J., Carleer, R., Schreurs, S.: Chromium(VI) removal using in-situ nitrogenized activated carbon prepared from Brewers' spent grain. Adsorption 24, 147–156 (2018)
- Vermeulen, T.: Theory for irreversible and constant-pattern solid diffusion. Ind. Eng. Chem. 45, 1664–1670 (1953)
- Weber, W.J., Morris, J.C.: Kinetics of adsorption on carbon from solution. J. Sanit. Eng. Divis. 89, 31–60 (1963)
- Wegmann, C., Suárez García, E., Kerkhof, P.J.A.M.: Kinetics of acrylonitrile adsorption from an aqueous solution using Dowex Optipore L-493. Sep. Purif. Technol. 81, 429–434 (2011)
- Wesselingh, J.A., Vonk, P., Kraaijeveld, G.: Exploring the Maxwell-Stefan description of ion exchange. Chem. Eng. J. 57, 75–89 (1995)
- Yuan, X., Im, S.I., Choi, S.W., Lee, K.B.: Removal of Cu(II) ions from aqueous solutions using petroleum coke derived microporous carbon: investigation of adsorption equilibrium and kinetics. Adsorption 25, 1205–1218 (2019)
- Zhu, M., Carta, G.: Protein adsorption equilibrium and kinetics in multimodal cation exchange resins. Adsorption 22, 165–179 (2016)

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