

13 **Local indicators for global species: pelagic sharks in the tropical northeast Atlantic, Cabo**
14 **Verde islands region.**

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31

32 **Abstract**

33 Pelagic sharks are an important bycatch in pelagic fisheries, especially for drifting longlines
34 targeting swordfish. In the Cabo Verde Archipelago (tropical NE Atlantic), pelagic shark catches
35 can reach a significant proportion of the total catches. Due to the increased concern on the
36 status of pelagic shark species, this study was developed to enhance the current knowledge of
37 those sharks in the Cabo Verde region in comparison to the adjacent areas, especially
38 associated with European Union (EU) pelagic longline fishing activity. Stock status indicators for
39 the two main species, blue shark (*Prionace glauca*) and shortfin mako (*Isurus oxyrinchus*), were
40 developed, based on fisheries data from logbooks and onboard scientific observers, including
41 analysis of size frequency distributions and standardized catch-per-unit-of-effort (CPUE)
42 indexes over time. The standardized CPUEs have been stable or increasing for both species in
43 the past 10 years, indicating no signs of local depletion. In terms of sizes, the blue shark catch is
44 composed mainly of adults, which can be a sign of a stable population. On the contrary, the
45 catch of shortfin mako is composed mainly of juveniles, which in conjunction of a decrease of
46 mean size might be a cause of concern, highlighting possible overfishing on the species in the
47 region. Thirty satellite tags, 25 archival miniPATs and 5 SPOT GPS, were deployed in the Cabo
48 Verde Exclusive Economic Zone (EEZ), showing that those species are highly mobile. The
49 biomass and size distributions were modeled with spatial and seasonal models (GAMs)
50 identifying locations where juveniles are predominantly concentrated and that should be
51 prioritized for conservation. This work presents new information on the status of pelagic sharks
52 in the Cabo Verde region in the context of those highly migratory species, and can now be used
53 to promote more sustainable fisheries in the region.

54

55 **Key-words:**

56 Indices of abundance; longline fisheries; pelagic sharks; population trends; satellite tagging;
57 spatial models.

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59

60 **1. Introduction**

61 Pelagic sharks are an important component in pelagic fisheries catches, especially for drifting
62 longliners targeting swordfish and tunas (Mejuto et al., 2009; Coelho et al., 2018). Depending
63 on the fisheries, areas and seasons, pelagic sharks can be significant in the overall catch. Blue
64 shark (*Prionace glauca*) and shortfin mako (*Isurus oxyrinchus*) are the two main shark species in
65 those fisheries, and can in some areas and season represent more than 50% of the total
66 longline catch and 90% of the total elasmobranch catch (Coelho et al., 2012). In the Cabo Verde
67 Archipelago (tropical NE Atlantic), pelagic shark catches can also be important for pelagic
68 fisheries (Fernandez-Carvalho et al., 2015a); however, the local status of those populations is
69 not currently assessed. Cabo Verde has a large Exclusive Economic Zone (EEZ) of 734,000 km²
70 and, thus, its future sustainability will be largely based on activities related to the use and
71 exploitation of the sea and coastal resources (de Carvalho, 2013). Overall for the Atlantic
72 Ocean, total pelagic fish catches reported to the International Commission for the
73 Conservation of the Atlantic Tunas (ICCAT) in the last few years (2014-2017) were
74 approximately 745,000 t per year. Of those, approximately 76,000 t per year (around 10% of
75 the total catch) represented pelagic sharks, mostly blue shark (approximately 65,000 t per year)
76 followed by shortfin mako (around 6,000 t per year) (ICCAT, 2018).

77 The blue shark is one of the widest ranging of all sharks, found throughout tropical and
78 temperate seas from latitudes of about 60°N to 50°S. It is a pelagic species mainly distributed
79 from the sea surface to depths of about 350 m, even though deeper dives of up to 1000 m
80 have been recorded (Campana et al., 2011). Blue shark is a highly migratory oceanic species,
81 with complex movement patterns and spatial structure probably related to the reproduction
82 cycles and prey distribution (Montealegre-Quijano and Vooren, 2010; Tavares et al., 2012;
83 Coelho et al., 2018). Tagging studies have shown extensive movements with numerous trans-
84 Atlantic migrations probably accomplished by using the major oceanic current systems
85 (Stevens, 1976; Stevens 1990; Queiroz et al., 2005; Silva et al., 2010; Campana et al., 2011). For
86 the north Atlantic, data on the distribution, movements and reproductive behavior suggests a
87 complex reproductive cycle, involving major oceanic migrations associated with mating areas in
88 the north-western Atlantic and pupping areas in the north-eastern Atlantic (Pratt, 1979;
89 Stevens, 1990).

90 The shortfin mako is also a widespread pelagic shark species, occurring in temperate and
91 tropical waters of all oceans from about 60°N to 50°S. It occurs from the surface to at least 500
92 m depth, and is occasionally found close to inshore waters where the continental shelf is

93 narrow (Compagno, 2001). Tagging studies in the northwest Atlantic have shown that shortfin
94 makos can make extensive migrations of more than 3,000 km (Casey and Kohler, 1992), even
95 though it has been suggested that trans-Atlantic migrations are not as common as in the blue
96 shark. In the Atlantic, Casey and Kohler (1992) suggests that shortfin mako core distribution in
97 the northwest Atlantic is between 20° to 40°N, bordered by the Gulf Stream in the west and
98 the mid-Atlantic ridge in the east. In the northeast Atlantic it is presumed that the Strait of
99 Gibraltar might be a nursery ground (Buencuerpo et al., 1998 and Tudela et al., 2005). The area
100 between 17° to 35°S off the coast of Brazil seems to be an area of birth, growth and mating in
101 the southwest Atlantic (Amorim et al., 1998).

102 The blue shark is currently listed as Near Threatened by IUCN, the International Union for the
103 Conservation of Nature (Stevens, 2009), while the shortfin mako is currently listed as
104 Endangered (Rigby et al., 2019). In the Ecological Risk Assessments (ERAs) carried out for pelagic
105 sharks in the Atlantic in 2010 and 2012 (Cortés et al., 2010; Cortés et al., 2015), blue shark was
106 shown to have an overall intermediate vulnerability, because it is the most productive of all
107 pelagic shark species. On the contrary the shortfin mako was one of the most vulnerable of all
108 species analyzed, due to its relatively low productivity and high susceptibility.

109 The latest blue shark stock assessments in the Atlantic were carried out by ICCAT in 2015
110 (Anon., 2015). For the North Atlantic the stock was unlikely to be overfished and subject to
111 overfishing, even though there were very high levels of model uncertainty reported (Anon.,
112 2015). For the shortfin mako the latest stock assessments were carried out by ICCAT in 2017
113 (Anon., 2017). The results for the North Atlantic indicated that stock abundance was either
114 below or very close to B_{MSY} , but that fishing mortality was overwhelmingly above F_{MSY} , with a
115 combined 90% probability of the stock being in an overfished state and experiencing
116 overfishing (Anon, 2017). Although the current biomass of the stock was still not very strongly
117 depleted, current fishing mortality levels are unsustainable and can lead to strong population
118 declines in the near future.

119 Due to the increased concern on the status of pelagic shark species and lack of specific
120 knowledge for the region around the Cabo Verde islands in the tropical NE Atlantic and how
121 local shark components are related to the managed population, this study was developed to
122 enhance the current knowledge of the two main pelagic sharks captured in longline fisheries in
123 the Cabo Verde region, especially associated with European Union (EU) pelagic longline fishing
124 activity. Specifically, the main objectives of this study were to: 1) analyze potential local
125 depletion of sharks in the region, specifically by analyzing trends in the catch composition,

126 catch rates (CPUEs: catch per unit effort) and size distribution for the main species and; 2)
127 identify possible biological and ecological sensitive areas in the region by modeling the spatial
128 distribution of catches of the main species in the region and using the analysis of satellite
129 telemetry tagging data.

130

131 **2. Materials and Methods**

132 ***2.1. Study area and fisheries data collection***

133 The study focused on the Cabo Verde region in the tropical NE Atlantic. Two areas were
134 defined, specifically: i) the Cabo Verde EEZ, and ii) a buffer of 300 nm adjacent to the EEZ
135 (**Figure S1** - Supplemental electronic material).

136 The data collected and analyzed included EU (Portugal and Spain) pelagic longline fleets fishery
137 logbook and scientific fishery observer data. Those data were compiled and used to provide
138 analysis on the sharks catch composition, catch rate trends and size distributions in the region.
139 Data were available and analyzed between 2006 and 2015, with the exception of 2008 that was
140 not included due to issues related with a switch in database format, which did not allow linking
141 the catch, effort and location (Vessel Monitoring System, VMS) data for that year. All fisheries
142 parameters and indicators were compared between the Cabo Verde EEZ and the neighboring
143 300 nm buffer area, according to the study areas previously defined.

144

145 ***2.2. Satellite tagging***

146 A total of 30 satellite tags were deployed within this study, specifically 25 miniPATs and 5
147 Fastloc GPS SPOT tags, both models from Wildlife Computers Inc. One of the GPS SPOT tagged
148 specimens was recaptured after 77 days, and that tag was redeployed on another specimens.
149 As such, of the available tags, 20 miniPATs and 6 GPS SPOTs (5 GPS SPOTs with 1 deployed
150 twice) were deployed in blue sharks and 5 miniPATs were deployed in shortfin makos, all inside
151 the Cabo Verde EEZ during 2016 (**Table S1** - Supplemental electronic material).

152 For the tagging operations, the sharks were restrained alongside the vessel and handled
153 carefully, with those in the best condition selected for tagging in order to maximize post-
154 release survivorship. Each tagged shark was measured, and the sex and maturity stage
155 determined (juvenile vs. adult, see section 2.3.2 with the used sex specific maturity ogives

156 available in the literature, Anon., 2014). Additional data recorded for each tagged specimen
157 included tagging location (latitude and longitude), date and time.

158 The miniPATs were rigged with monofilament leaders secured with stainless steel crimps and
159 encased in plastic tubing. Umbrella-type nylon darts (Domeier et al., 2005) were used to attach
160 the tags to the shark dorsal musculature below the first dorsal fin, using the methodology
161 described by Howey-Jordan et al. (2013). For the SPOT tags, the tags were attached with a
162 plastic fin mount system placed in the first dorsal fin provided by the tag manufacturer.

163 The miniPAT tags archive detailed depth and temperature time-series data and use the light-
164 based information for geo-locations. On the contrary, the Fastloc GPS SPOT tags use GPS based
165 geo-locations that are much faster and provide more precise estimates, but do not record
166 depth or temperature profiles. As such, both tags were used to provide complementary
167 information on the sharks' habitat use and movements.

168 For estimating geographical daily positions, the Fastloc GPS signals are processed by the tags,
169 compressed and then transmitted over the ARGOS satellite system. For the miniPATs, the daily
170 locations were calculated based on the light levels recorded and using state-space statistical
171 models (GPE3 software, processed through the tag manufacturer web portal). The miniPATs
172 provide observations on twilight, sea surface temperature and dive depth, and the state-space
173 modeling approach uses those observations and the corresponding reference data, along with
174 a simple diffusion based movement model, to generate time-discrete gridded probability
175 surfaces throughout the deployment. The corresponding oceanographic reference data used
176 was that from NOAA Optimum Interpolation SST V2 High Resolution for the sea surface
177 temperature, and from NOAA ETOP01 global relief model, Bedrock version, for bathymetry,
178 respectively. The grids used were 0.25*0.25 degrees of latitude*longitude. From those
179 probability surfaces, the most likely animal locations for a given day/time were derived.

180

181 **2.3. Analysis of local shark indicators**

182 *2.3.1. Catch composition*

183 The relative catch composition of sharks, defined as the species specific shark species in
184 relation to the total shark catches, was calculated and analyzed for the general Cabo Verde EEZ,
185 as well as for the 300 nm neighboring area.

186

187 *2.3.2. Size distribution*

188 The annual trends in the size frequency distributions and mean sizes for the main shark species
189 were analyzed and plotted by area, sex and quarter of the year. Size data were tested for
190 normality with Kolmogorov-Smirnov normality tests with Lilliefors correction (Lilliefors, 1967),
191 and for homogeneity of variances with Levene tests (Levene, 1960). Specimen sizes were
192 compared between regions (Cabo Verde EEZ and 300nm adjacent area), sexes and quarters of
193 the year using non-parametric k-sample permutation tests (Manly, 2007).

194 The mean size at first maturity (L_{50}) used to define immature and mature specimens were
195 based on the ICCAT Sharks Working Group report (Anon., 2014) for the North Atlantic shark
196 stocks, as follows:

- 197 • Blue shark (males): mean = 200.1 cm Fork Length (FL);
- 198 • Blue shark (females): mean = 185.1 cm FL;
- 199 • Shortfin mako (males): mean = 182.5 cm FL; range = 180 - 185 cm FL;
- 200 • Shortfin mako (females): mean = 286.5 cm FL; range = 275 - 298 cm FL.

201

202 *2.3.3. CPUE trends and standardization*

203 The time series of catch per unit effort (CPUE) were plotted for blue and shortfin mako sharks,
204 which allowed following the trends over time and assessing seasonality effects in the catch
205 rates. The CPUE time series were standardized in order to remove the fishery-dependent
206 effects (i.e., spatial, seasonal and targeting effects) and estimate relative indexes of abundance
207 that can be used as population status indicators. For the standardization process, the response
208 variable considered was CPUE measured in biomass of live fish (kg) per 1000 hooks deployed.
209 The standardized CPUEs were estimated with statistical models using Generalized Linear
210 Models (GLMs) and Generalized Linear Mixed Models (GLMMs).

211 Blue shark and shortfin mako data had different characteristics, especially with regards to the
212 percentage of zeros in the datasets. Specifically, the blue shark is relatively common in the
213 catches and has a low percentage of fishing sets with zero catches, while the shortfin mako is
214 rarer in the catches and had a much higher percentage of fishing sets with zero catches. The
215 presence of fishing sets with zero catches results in a response variable of $CPUE=0$, that can
216 cause mathematical problems for fitting the models, and as such different approaches were
217 tested and applied in each case.

218 Four different modeling methodologies were initially tested and compared, specifically
219 tweedie, gamma, lognormal and delta lognormal models. For the tweedie models the nominal
220 CPUE was used directly, as the response variable, given that this distribution can handle a
221 certain proportion of zeros (mass) and a continuous distribution for the non-zeros. For the
222 gamma and lognormal models the response variable was defined as the nominal CPUE +
223 constant (c), with c set to 10% of the overall mean catch rate. The value of c=10% of the mean
224 has been recommended by Campbell (2004), as it seems to minimize the bias for this type of
225 adjustments. Further, and in a comparative study, Shono (2008) showed that when the
226 percentage of zeros in the dataset is low (<10%), the method of adding a constant to the
227 response variable performs relatively well. The final tested approach was a delta-lognormal
228 model that uses and combines two different models, specifically a binomial model for the
229 proportion of positive catches and a lognormal model for the expected CPUEs in the positive
230 sets.

231 The covariates considered and tested in the models were:

- 232 • Year: analyzed between 2006 and 2015;
- 233 • Seasonal effects (quarters of the year, 4 categories): 1 = January to March, 2 = April
234 to June, 3 = July to September, 4 = October to December;
- 235 • Spatial/area effects: tested as 5*5 or 10*10 degree grids;
- 236 • Targeting effects: based on the SWO/SWO+BSH ratio of captures.

237 Interactions between pairs of variables were considered and tested in the analysis and used in
238 the final models, if significant. Specifically, interactions not involving the year factor were
239 considered as fixed factors in GLM type models, while interactions involving the year factor
240 were considered as random variables within GLMM models.

241 The significance of the explanatory variables, as well as the interactions, were assessed with
242 likelihood ratio tests (LRT) comparing each univariate model to the null model (considering a
243 significance level of 5%), and by analyzing the deviance explained by each covariate. Goodness-
244 of-fit and model comparison was carried out with the Akaike Information Criteria (AIC) and the
245 pseudo coefficient of determination (R^2). Model validation was carried out with a residual
246 analysis. The final estimated indexes of abundance were calculated by least square means
247 (LSMeans or Marginal Means), that for comparison purposes were scaled by the mean
248 standardized CPUE in the time series.

249

250 **2.4. Spatial models for prediction of catch rates and sizes**

251 Generalized Additive Models (GAM) were used to predict the expected blue shark and shortfin
252 mako shark catch rates (CPUEs) and size distribution as a function of location (latitude and
253 longitude) and quarter of the year. The models used were lognormal GAMs for modeling the
254 CPUEs and Gaussian with identity link for modeling the sizes.

255 The predictors in the models were given by the smooth functions of latitude and longitude plus
256 a parametric component for the quarters of the year. The smooth terms for the location
257 covariates were estimated by maximum likelihood with thin plate regression splines (Wood,
258 2003). The significance of the model parameters was tested with LRT comparing nested
259 models, including the significance of the interactions between latitude, longitude and quarter
260 of the year. Goodness-of-fit was assessed with the Akaike Information Criterion (AIC; Akaike,
261 1973) and with the final deviance explained. A residual analysis was carried out for model
262 validation. The expected mean catch rates and sizes were mapped along the study area and for
263 each quarter of the year.

264 All analysis in this study was carried out using the R language for statistical computing v3.4.0 (R
265 Core Team, 2017), with the following additional libraries: "car" (Fox and Weisberg, 2011),
266 "ggplot2" (Wickham, 2009), "gmodels" (Warnes et al., 2013), "KernSmooth" (Wand, 2015),
267 "lme4" (Bates et al., 2013), "lsmmeans" (Lenth, 2014), "maps" (Becker et al., 2013), "mgcv"
268 (Wood, 2006, 2011), "perm" (Fay and Shaw, 2010), "plyr" (Wickham, 2011), "raster" (Hijmans,
269 2016) and "tweedie" (Dunn, 2013).

270

271 **3. Results**

272 **3.1. Local shark indicators**

273 *3.1.1. Catch composition*

274 The catch composition of elasmobranchs in the study area is largely dominated by blue shark
275 (BSH) followed by shortfin mako (SMA) (**Table 1**). Other less frequent elasmobranch species
276 occasionally captured in the region include the bigeye thresher (BTH) (*Alopias superciliosus*),
277 silky shark (FAL) (*Carcharhinus falciformis*), longfin mako (LMA) (*Isurus paucus*), oceanic
278 whitetip (OCS) (*Carcharhinus longimanus*), crocodile shark (PSK) (*Pseudocarcharias kamoharai*)
279 and smooth hammerhead (SPZ) (*Sphyrna zygaena*) (**Table 1**). Most of those other species are

280 either no-retention species in ICCAT and/or listed in CITES, and therefore are mostly released or
 281 discarded.

282

283 **Table 1.** Catch composition (percentage, in weight) of major shark species captured in the Cabo
 284 Verde EEZ and adjacent waters of the tropical NE Atlantic.

FAO code	Species	Species composition (%)		
		EEZ	300 nm	Combined
BSH	Blue shark	93.4	94.5	94.1
SMA	Shortfin mako	4.7	3.3	3.8
SPZ	Smooth hammerhead	0.2	0.3	0.3
FAL	Silky shark	0.3	0.3	0.3
OCS	Oceanic whitetip	0.1	0.1	0.1
BTH	Bigeye thresher	<0.1	<0.1	<0.1
SKH	Other elasmobranchs	1.2	1.4	1.3

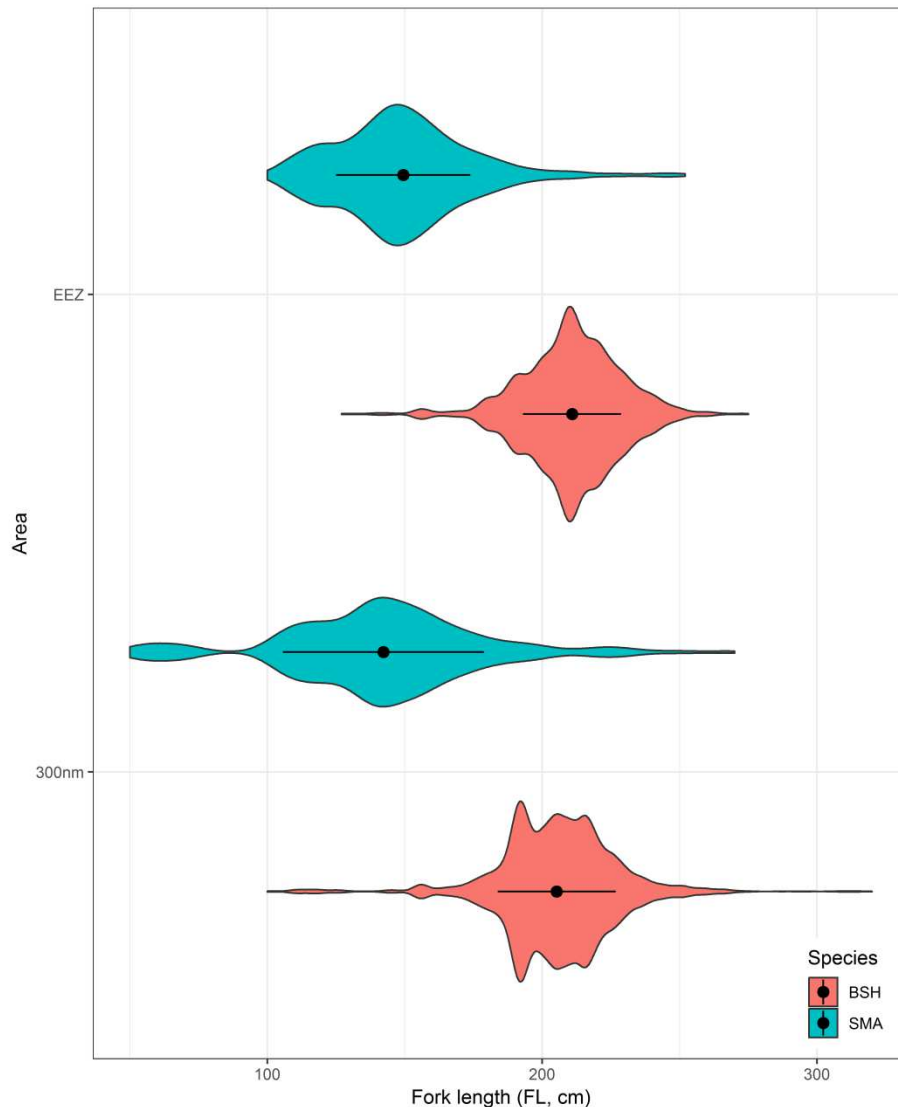
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286 *3.1.2. Size distribution of the major shark species*

287 In terms of size distribution, blue sharks caught in the Cabo Verde region are relatively large
 288 specimens with mean sizes of 210.9 cm FL (SD=17.9) inside the Cabo Verde EEZ and 205.3 cm
 289 FL (SD=21.5) in the 300 nm adjacent waters (**Figure 1**). Those differences observed in the mean
 290 sizes in the two areas were significant (K-Sample Asymptotic Permutation Test: Chi2 = 88.9, df =
 291 1, p-value < 0.001), meaning that in the Cabo Verde EEZ the blue sharks are significantly larger
 292 than in the adjacent waters. Considering that the estimated blue shark mean size at first
 293 maturity for the North Atlantic is 185 cm FL for females and 200.1 cm FL for males (Anon.,
 294 2014), the catch of blue sharks in the Cabo Verde EEZ and neighboring waters is likely
 295 composed mainly by adults. The size distribution is also narrower in the EEZ.

296 For the shortfin mako the mean sizes were 149.4 cm FL (SD=24.5) inside the Cabo Verde EEZ
 297 and 142.2 cm FL (SD=36.6) in the adjacent waters. In this species the observed differences
 298 between areas were also statistically significant (K-Sample Asymptotic Permutation Test: Chi2 =
 299 13.4, df = 1, p-value < 0.001), meaning that in the Cabo Verde EEZ shortfin makos are also
 300 significantly larger than in the adjacent waters (**Figure 1**). Considering that the estimated
 301 shortfin mako mean size at first maturity for the North Atlantic is 275-298 cm FL for females

302 and 180-185 cm FL for males (Anon., 2014), the catch of shortfin makos in both regions is likely
303 composed mainly by juveniles. Again for shortfin mako the size distribution of the catches
304 showed a wider distribution outside the EEZ of Cabo Verde.



305

306 **Figure 1.** Size frequency distributions of the main shark species (BSH - blue shark and SMA -
307 shortfin mako) in the Cabo Verde EEZ and 300 nm adjacent waters. The point and lines inside
308 the plots represent the mean \pm standard deviations. (*Note do editor: Color figure provided for*
309 *the online version of the paper and a grayscale version is provided for the print version- this*
310 *applies to all figures in the manuscript*).

311

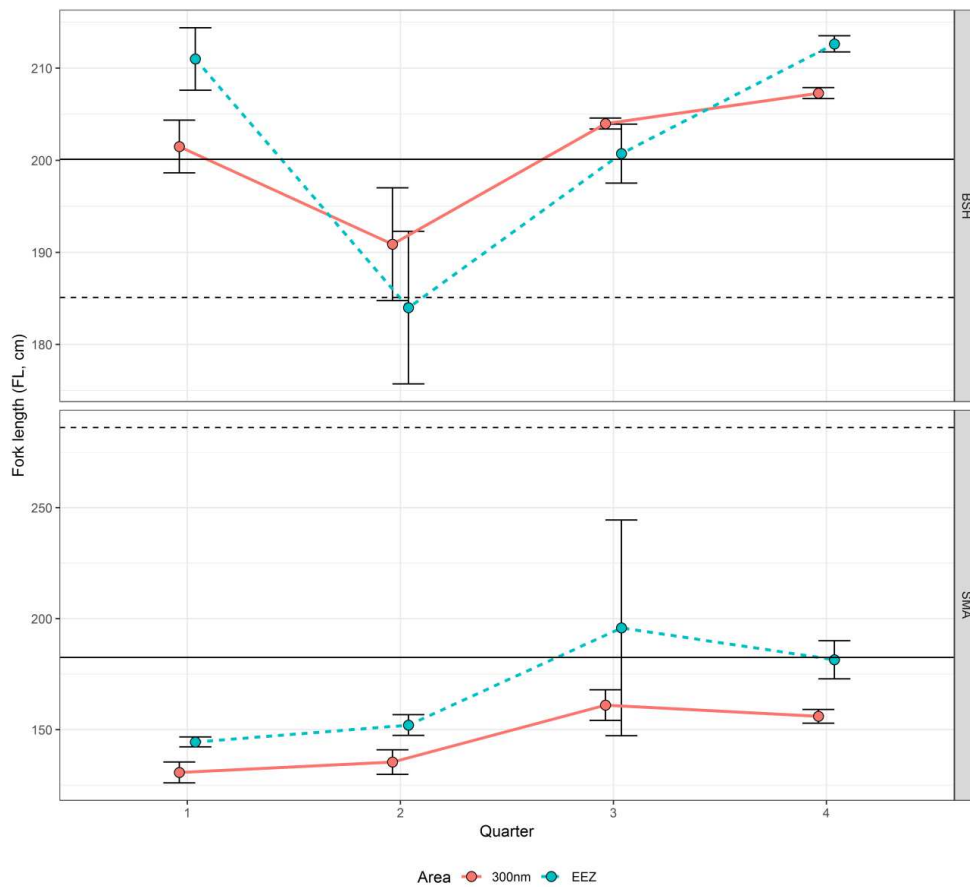
312 In terms of seasonality, larger blue sharks were captured in the 1st, 3rd and 4th quarters of the
313 year, while smaller specimens were caught mainly in the 2nd quarter (**Figure 2**), with those

314 seasonal differences being statistically significant (K-Sample Asymptotic Permutation Test: Chi2
315 = 180.1, df = 3, p-value < 0.001).

316 For the shortfin mako the smaller size specimens were captured in the 1st and 2nd quarters),
317 while larger specimens are captured later in the year, in the 3rd and 4th quarters (**Figure 2**).

318 Those differences were also statistically significant (K-Sample Asymptotic Permutation Test: Chi2
319 = 113.9, df = 3, p-value < 0.001).

320



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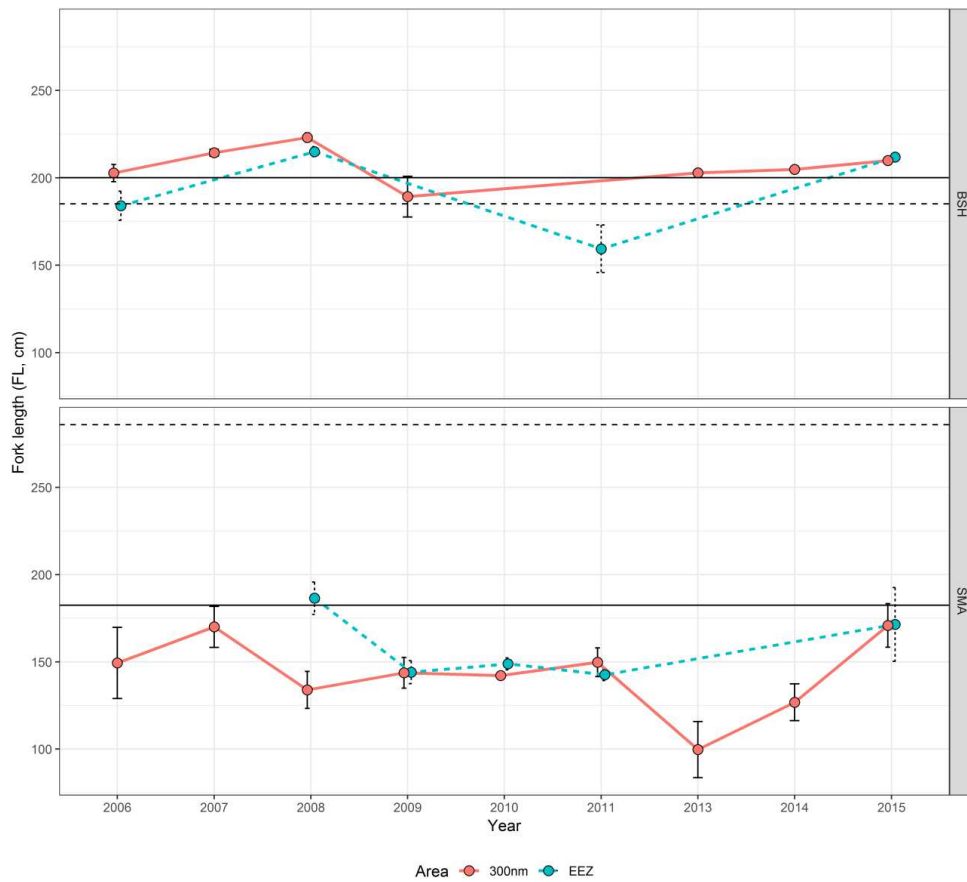
322 **Figure 2.** Seasonal mean sizes of the main shark species (BSH - blue shark and SMA - shortfin
323 mako) in the Cabo Verde EEZ and 300 nm adjacent waters. The error bars refer to the 95%
324 confidence intervals (CI). Mean sizes at first maturity (L_{50}) of each species are indicated for
325 males (horizontal solid lines) and females (horizontal dashed lines) (Anon., 2014).

326

327 In terms of trends in the size distribution, blue shark mean size was relatively stable along the
328 time series (**Figure 3**) in both areas, even though those relatively small differences were

329 significant (K-Sample Asymptotic Permutation Test: $\chi^2 = 573.3$, $df = 7$, $p\text{-value} < 0.001$). By
 330 contrast, there was a general decreasing trend in the mean size of shortfin mako during the
 331 study period, especially in the 300 nm adjacent waters, except in the most recent years (2014-
 332 2015) when mean sizes increased to similar levels of the initial years (**Figure 3**). The mean
 333 annual differences in the shortfin mako sizes were also statistically significant (K-Sample
 334 Asymptotic Permutation Test: $\chi^2 = 136.4$, $df = 8$, $p\text{-value} < 0.001$).

335



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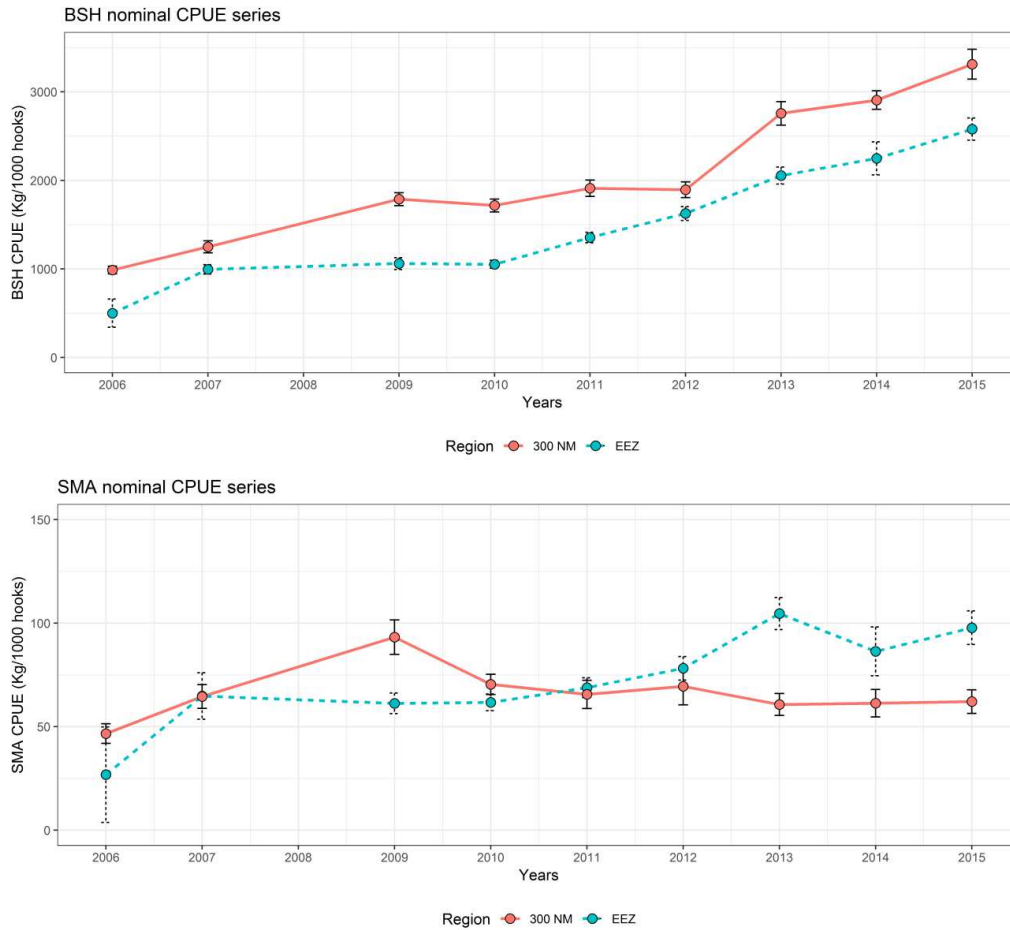
337 **Figure 3.** Time series trends of the mean sizes of the main shark species (BSH - blue shark and
 338 SMA - shortfin mako) in the Cabo Verde EEZ and 300 nm adjacent waters. The error bars refer
 339 to the 95% confidence intervals (CI). Mean sizes at first maturity (L_{50}) values are indicated for
 340 males (solid line) and females (dashed line) of each species (Anon., 2014).

341

342 3.1.3. Nominal catch per unit of effort (CPUE) distribution and trends

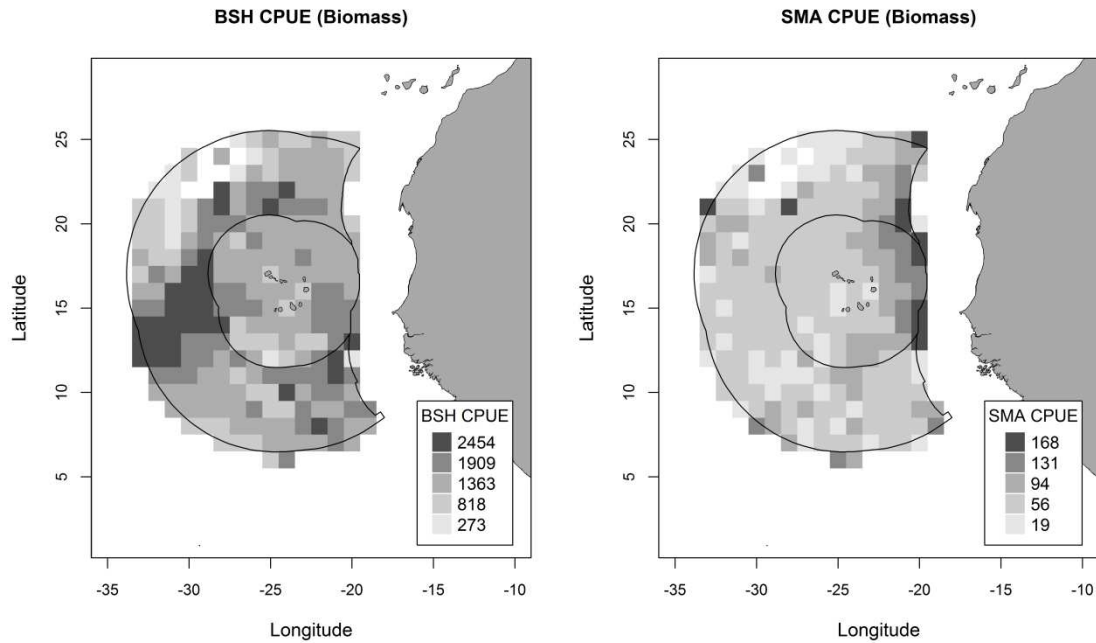
343 General increasing CPUE trends for both blue shark and shortfin mako were observed along the
 344 study period in both areas. More specifically, for blue shark there was a progressive increase

345 between 2006 and 2015, while for shortfin mako there was an increase mainly in the earlier
 346 years, between 2006 and 2009, and then a more stable period between 2009 and 2015 (**Figure**
 347 **4**).



348
 349 **Figure 4.** Time series of nominal catch per unit of effort (CPUE, biomass in Kg/1000 hooks) for
 350 blue shark (above) and shortfin mako (below) in each of the study areas, Cabo Verde EEZ and
 351 the 300 nm adjacent waters. The error bars refer to the 95% confidence intervals (CI).

352
 353 In terms of the spatial CPUE distribution, higher blue shark catch rates occurred mainly outside
 354 the Cabo Verde EEZ, in the 300 nm adjacent waters, especially in the western area (**Figure 5**).
 355 For shortfin mako, higher CPUEs were also observed in the limits and outside the Cabo Verde
 356 EEZ, but mainly in the eastern areas towards the African western coast (**Figure 5**).



357

358 **Figure 5.** Spatial distribution (by 1°*1° squares) of catch per unit of effort (CPUE, biomass in
 359 kg/1000 hooks) in the Cabo Verde EEZ and 300 nm adjacent waters, for blue shark (left plot)
 360 and shortfin mako (right plot). Data was combined for the period 2006-2015.

361

362 3.1.4. CPUE standardization of blue shark

363 The percentage of fishing sets with zero catches of blue shark was low (2.9%). There was a
 364 slight decrease in the sets with zero catches in the earlier period until 2011, followed by a slight
 365 increase in the more recent years (**Figure S2** - Supplemental electronic material). In terms of
 366 data distribution, the nominal blue shark CPUEs were highly skewed to the right and became
 367 more normal distributed in the log-transformed scale (**Figure S3** - Supplemental electronic
 368 material).

369 Of the various models tested, the best fit was obtained with a lognormal GLMM model. All the
 370 explanatory variables tested for the CPUE standardization were significant and contributed
 371 significantly for explaining part of the model deviance, including the effects for year, quarter,
 372 area and targeting (**Table S2** - Supplemental electronic material). The interactions of quarter
 373 with targeting were also significant and included in the model as a fixed variable, as well as the
 374 interaction between year and quarter included as a random effect. On the final fitted model,
 375 the factors that contributed most for the deviance explanation were targeting, followed by
 376 year, quarter and area (**Table S2** - Supplemental electronic material). In terms of model

377 validation, the residual analysis, including the residuals distribution along the fitted values, the
378 QQ plots and the residuals histograms, showed a good model fit without major outliers or
379 trends in the residuals (**Figure S4** - Supplemental electronic material).

380 The final standardized index of abundance for the blue shark in the Cabo Verde EEZ between
381 2006 and 2015 shows an overall increase along the entire time series period, similar to what is
382 observed in the nominal CPUE series (**Figure 6**).

383

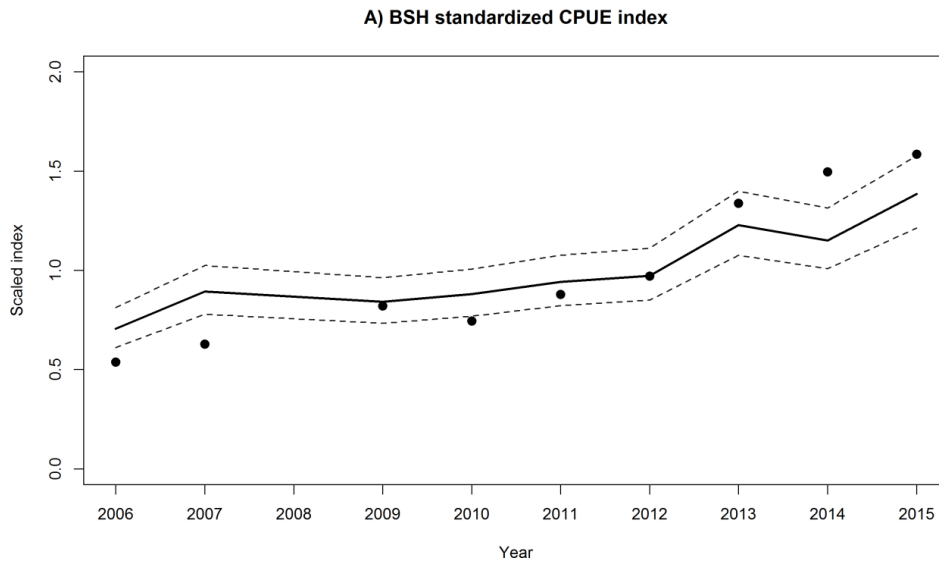
384 *3.1.5. CPUE standardization of shortfin mako*

385 The overall percentage of fishing sets with zero catches of shortfin mako was 37.7%. Higher
386 proportion of sets with zero shortfin mako catches were observed in the earlier years, and a
387 progressive decrease for the most recent years (**Figure S5** - Supplemental electronic material).

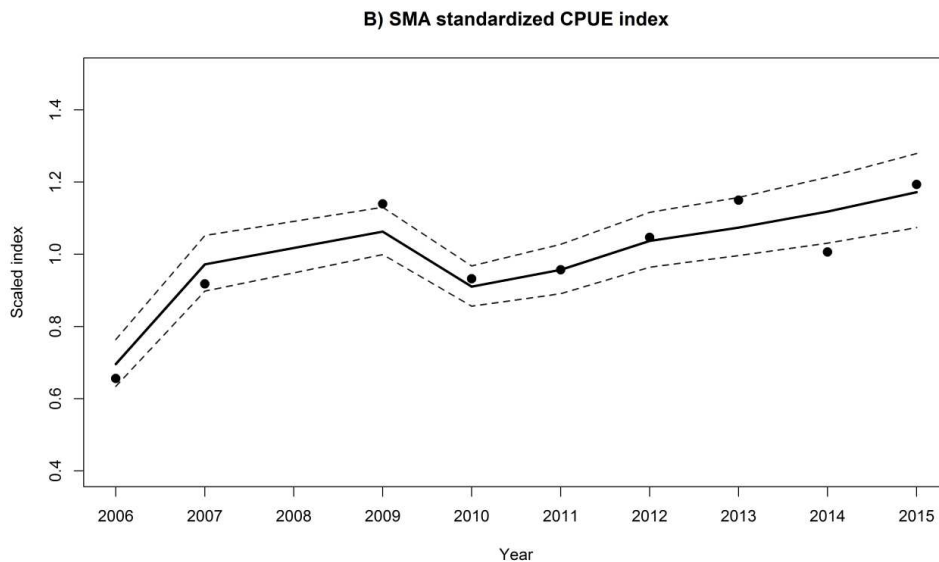
388 The CPUE distribution was also highly skewed to the right and became more normal distributed
389 in the log-transformed scale (**Figure S6** - Supplemental electronic material).

390 Given the high percentage of zeros in the data and the shape of the distribution, the best fitted
391 model was a tweedie GLM. All the explanatory variables tested were significant and
392 contributed significantly for explaining part of the model deviance, including the effects for
393 year, quarter, area and target (**Table S3** - Supplemental electronic material). The interactions of
394 quarter with targeting were also significant and included in the model as a fixed effect. On the
395 final fitted model, the factors that contributed most for the deviance explanation were the
396 area, followed by quarter, year and targeting (**Table S3** - Supplemental electronic material). In
397 terms of model validation, the residual analysis, including the residuals distribution along the
398 fitted values, the QQ plots and the residuals histograms, showed a good model fit without
399 major outliers or trends in the residuals (**Figure S7** - Supplemental electronic material).

400 The final standardized index of abundance for the shortfin mako shark in the Cabo Verde EEZ
401 between 2006 and 2015 shows an increase in the earlier years until 2009, followed by a
402 decrease in 2010, and then a slight increased again in the most recent years until 2015 (**Figure**
403 **6**).



404



405

406 **Figure 6.** Standardized catch per unit effort (CPUE) series for blue shark (A - top) and shortfin
 407 mako (B - below) in the in the Cabo Verde region. The solid line represents the standardized
 408 CPUE, the dashed line represents the 95% confidence intervals of the standardized CPUE, and
 409 the black dots represent the nominal CPUE. Each series is scaled by the mean standardized
 410 CPUE

411

412 **3.2. Satellite tagging**

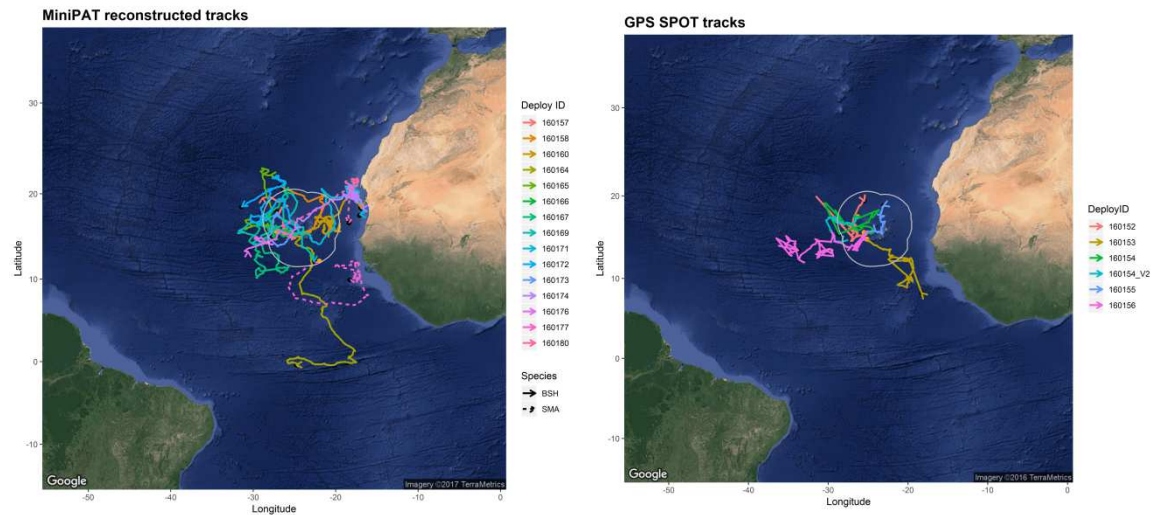
413 The SPOT tags (6 deployments) in blue shark had duration periods between 22 and 88 days,
 414 with 2 of the tagged specimens fished (recaptured) after 50 and 77 days at liberty. Overall, the

415 deployed SPOT tags recorded data on 328 tracking days for the 6 blue shark specimens (**Table**
416 **S4** - Supplemental electronic material).

417 Of the 25 deployed miniPAT tags (20 on blue shark and 5 on shortfin mako), 10 tags reached
418 the full deployment period and popped-up on the expected date after 120 days (8 blue sharks
419 and 2 shortfin makos), 10 specimens suffered post-release mortality (7 blue sharks and 3
420 shortfin makos) between 1 and 26 days at liberty, 1 blue shark was recaptured by a fishing
421 vessel after 71 days at liberty, 2 blue sharks dived to the maximum tag depth (~1850m) after 56
422 and 76 days at liberty and the tags released automatically to avoid damage due to excessive
423 pressure, 1 tag had premature release (shedding) after 12 days, and 1 tag failed to transmit.
424 Overall, a total of 1,296 tracking days were recorded for blue sharks and 258 days for shortfin
425 mako (**Table S4** - Supplemental electronic material).

426 From the miniPAT most likely estimated tracks, it was possible to see that most blue sharks
427 moved substantial distances, on most cases to areas outside the EEZ (**Figure 7**). There was not
428 a defined pattern in the movements, as there were cases of sharks moving towards the east,
429 west, north and south. Particularly noteworthy was a blue shark that was tagged inside the
430 Cabo Verde EEZ close to the Islands and that moved a significant distance towards the
431 equatorial waters. Similar patterns were obtained with the GPS SPOT tags, showing blue sharks
432 also moving outside the Cabo Verde EEZ, in this case mainly towards western and southeastern
433 areas (**Figure 7**).

434 For the shortfin mako, most specimens also tended to move outside the Cabo Verde EEZ, but in
435 this case mostly towards areas closer to the western African shelf, east of the Cabo Verde
436 Islands. One particular specimen that was tagged inside the EEZ southeast of the Islands
437 moved southeast, towards the region closer to the continental shelf at the latitude of the
438 Bijagós Islands in Guiné Bissau (**Figure 7**).



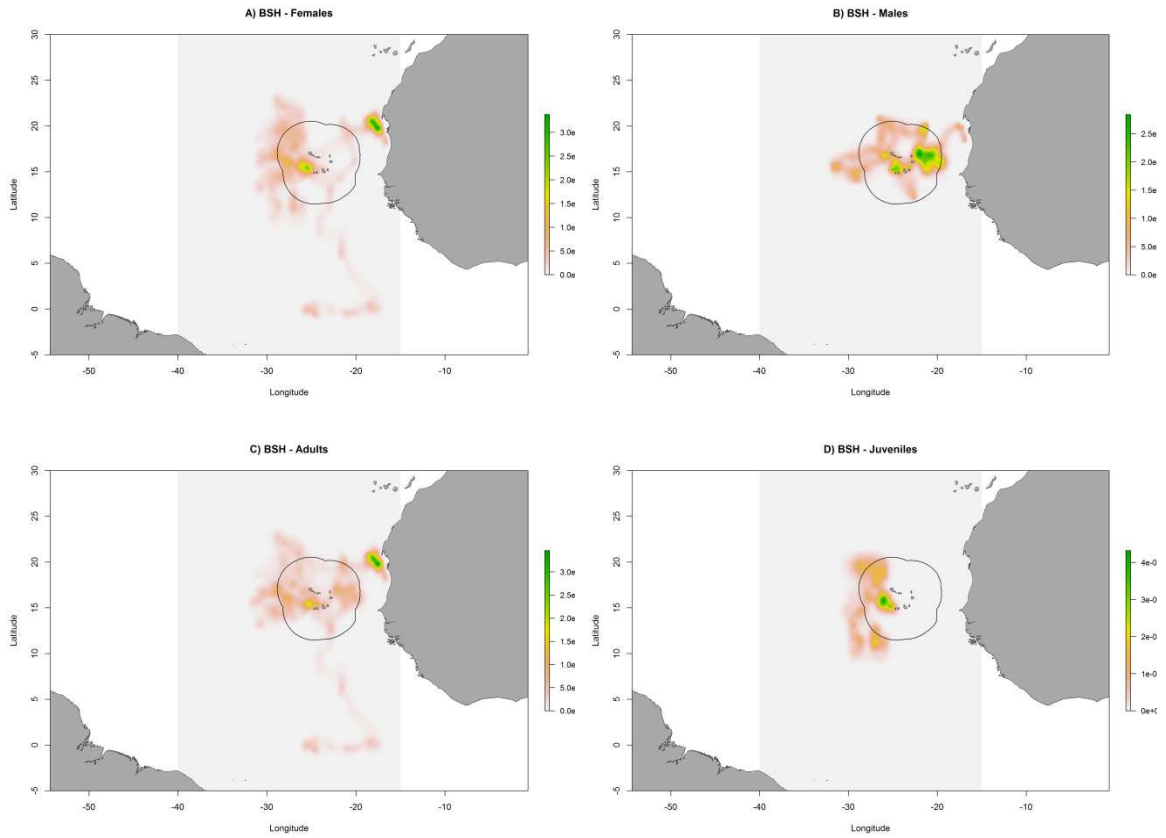
439 **Figure 7.** Reconstructed tracks for miniPAT (left panel) and GPS SPOT tags (right panel) for blue
 440 shark and shortfin mako. Only tags with tracking days > 26 days are shown, in order to exclude
 441 specimens that suffered post-release mortality after tagging and/or were fished (recaptured)
 442 very close to tagging location.

443

444 For blue shark there were differences in the spatial distribution of the satellite tagged sharks,
 445 when comparing between males and females. In general, males moved less and stayed closer
 446 to the islands, while females showed wider movement and traveled greater distances to other
 447 areas (**Figure 8**). When comparing maturity stages, there were also differences, with adults
 448 travelling greater distances than juveniles (**Figure 8**).

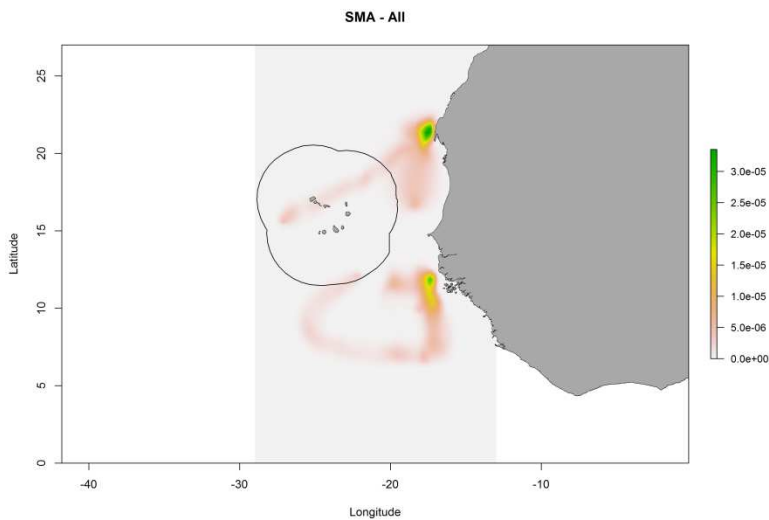
449 For shortfin mako the probability of distributions was closer to the African continental shelf,
 450 mainly outside the Cabo Verde EEZ (**Figure 9**). For this species, as less specimens were tagged,
 451 the analysis was made jointly and not separated by sex or maturity stage.

452



453 **Figure 8.** Probability surfaces of the spatial distribution of satellite tagged blue sharks (BSH) in
 454 the Cabo Verde region, tropical NE Atlantic. The plots represent females (A - top left), males (B
 455 - top right), adults (C - bottom left) and juveniles (D - bottom right). The colors in the legend
 456 refer to the distribution of the density of the specimens, ranging from red (lower density) to
 457 green (higher density).

458



459

460 **Figure 9.** Probability surfaces of the spatial distribution of satellite tagged shortfin mako shark
461 (SMA) in the Cabo Verde region, tropical NE Atlantic. The colors in the legend refer to the
462 distribution of the density of the specimens, ranging from red (lower density) to green (higher
463 density).

464

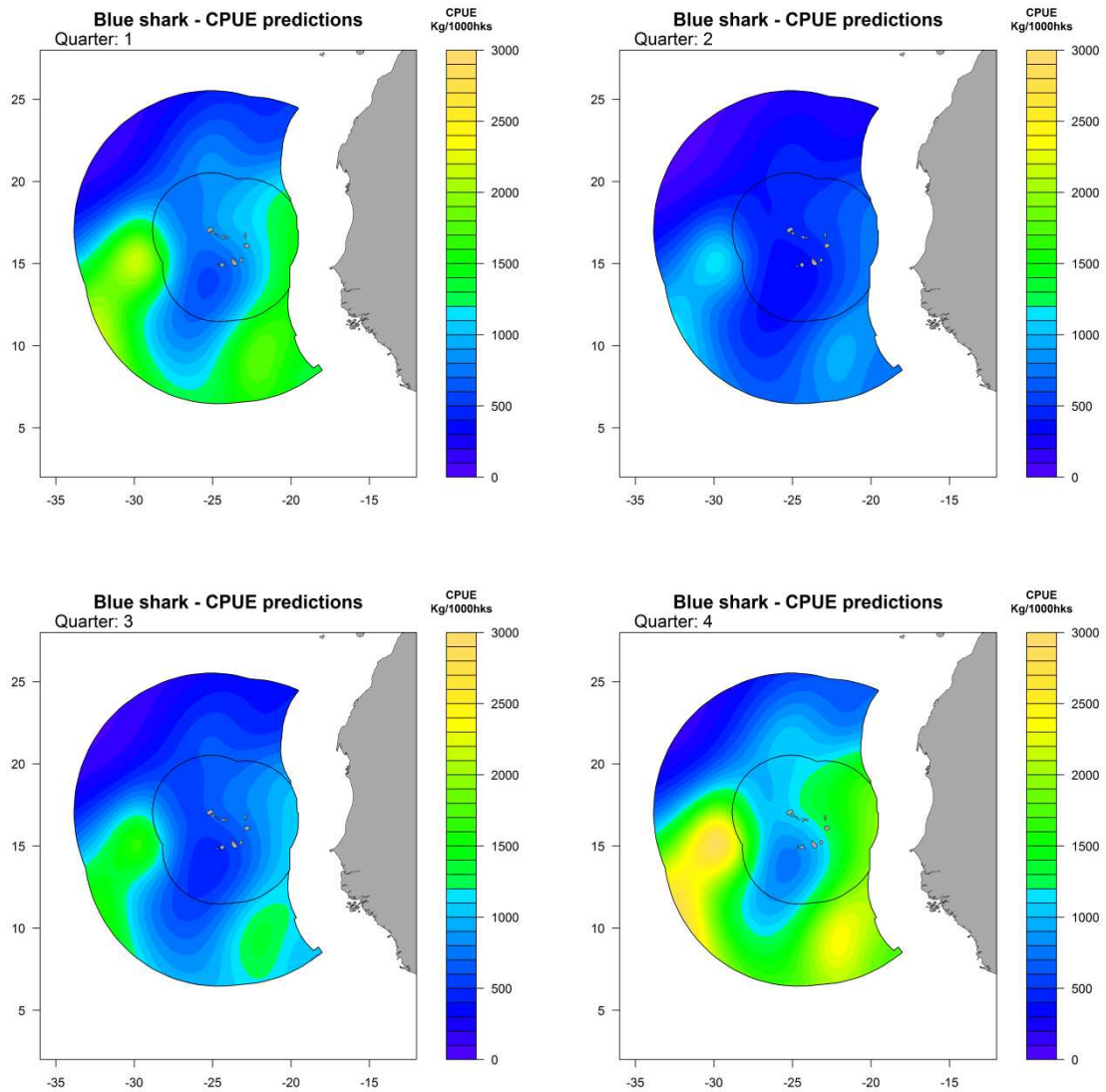
465 **3.3. Spatial models**

466 *3.3.1. Modeling and predicting catch rates*

467 There was considerable variability in the expected catch rates (CPUEs) of both blue shark and
468 shortfin mako in the study area when taking into consideration the location (spatial effects)
469 and quarter of the year (seasonal effects).

470 For blue shark, overall higher CPUEs were predicted outside the Cabo Verde EEZ, especially in
471 the south and southwest regions, while lower CPUEs were expected both in the EEZ and also in
472 the northern areas outside the EEZ (**Figure 10**). Higher CPUEs were predicted during the winter
473 and autumn (quarters 1 and 4), while much lower overall CPUEs were predicted in late spring
474 and summer, especially during quarter 2 (**Figure 10**).

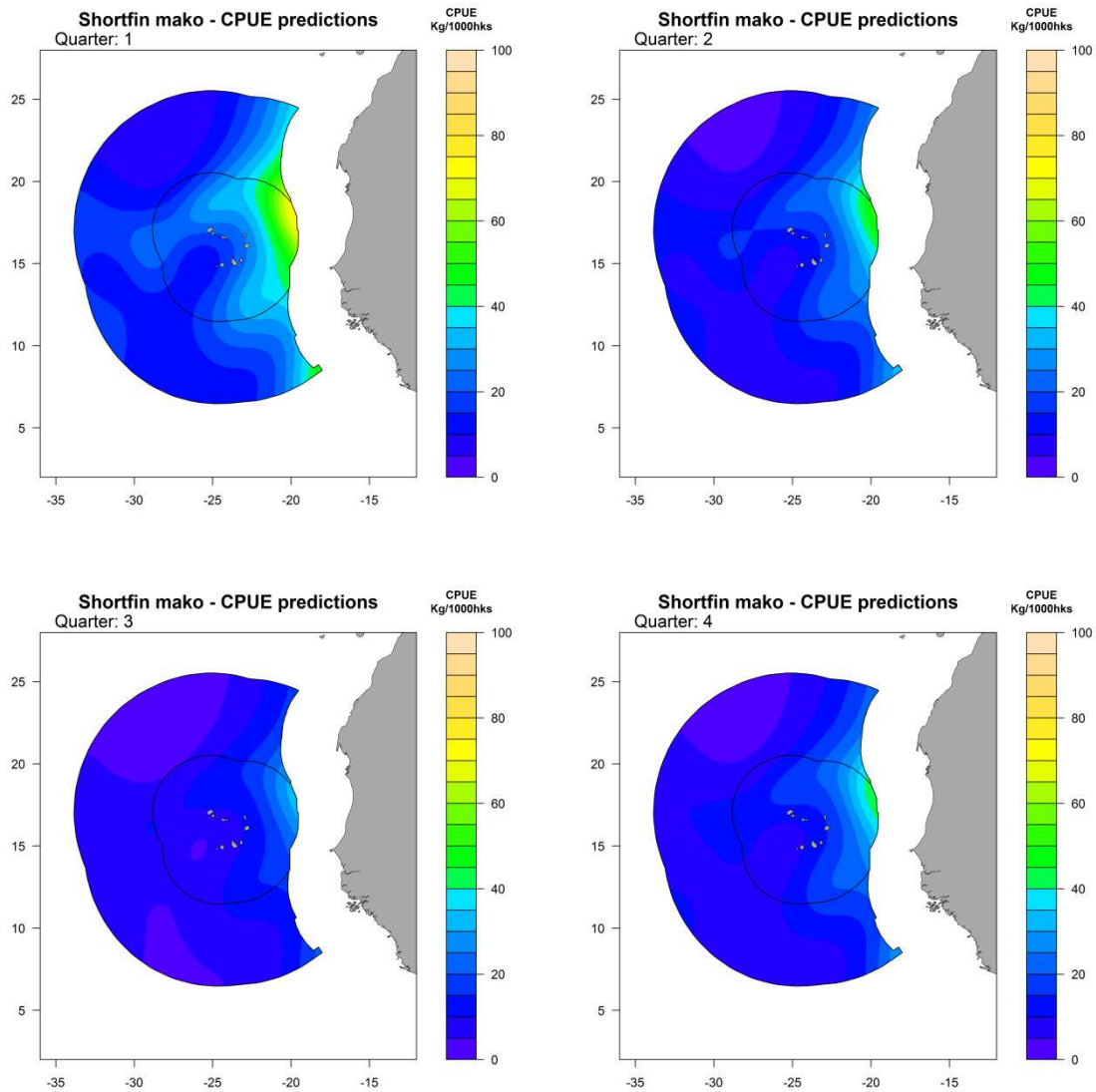
475



476 **Figure 10.** Seasonal prediction of the catch rates (CPUEs) of blue shark in the Cape Verde EEZ
 477 and 300nm adjacent waters. The predicted values are the result of a Generalized Additive
 478 Model (GAM) with lognormal distribution, taking into consideration the smooth terms of catch
 479 location estimated with thin plate regression splines and the quarter of the year used as a
 480 parametric term.

481

482 For shortfin mako very low CPUE along most of the study area was predicted, including both
 483 the Cabo Verde EEZ and most of the adjacent waters. The higher CPUEs for this species were
 484 predicted in the eastern areas, closer to the African continental shelf waters (**Figure 11**). The
 485 seasonal effects were not as noticeable as for the blue shark, with the overall trends mostly
 486 constant and low along all quarters of the year (**Figure 11**).



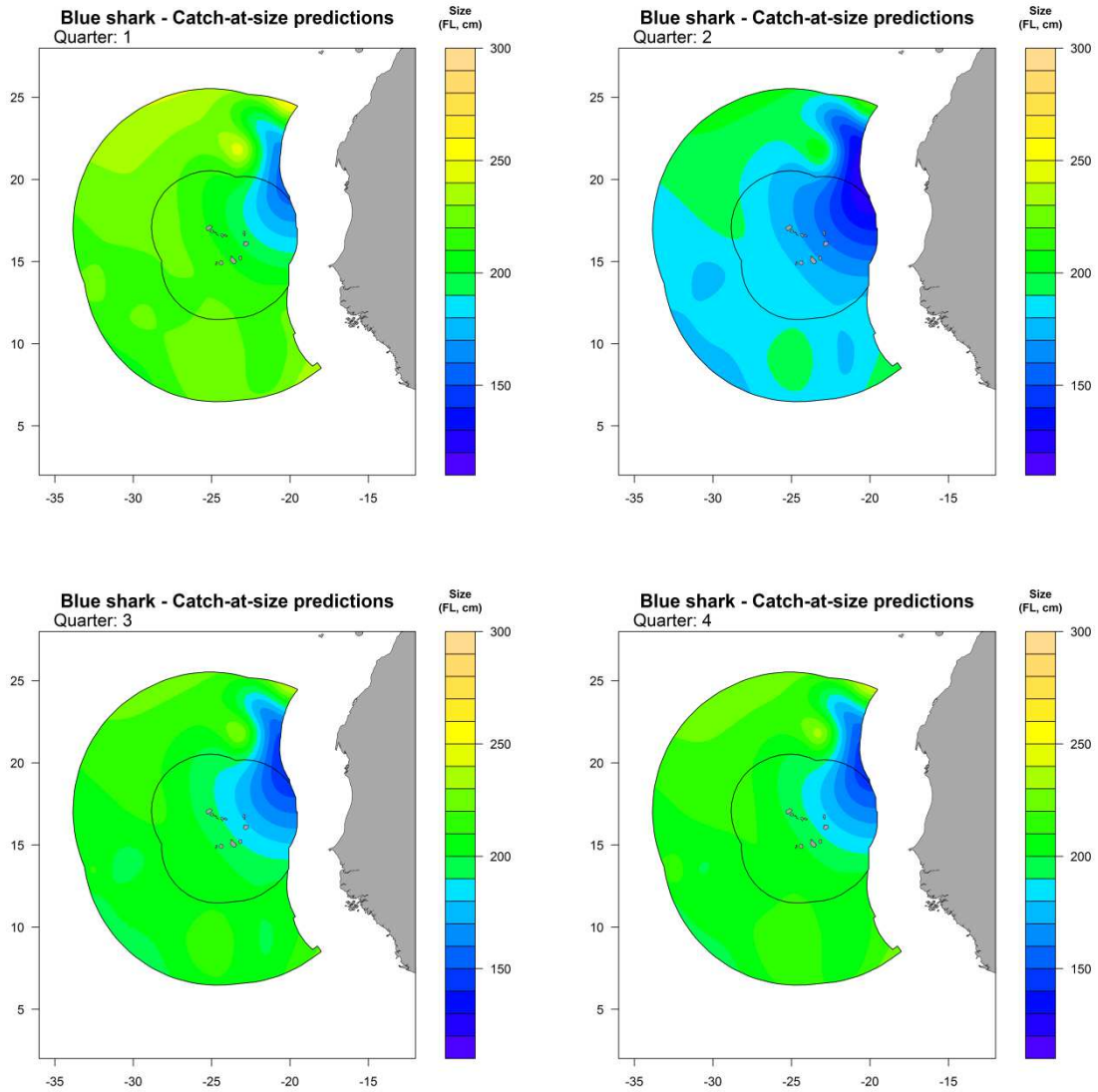
487 **Figure 11.** Seasonal prediction of the catch rates (CPUEs) of shortfin mako in the Cabo Verde
 488 EEZ and 300nm adjacent waters. The predicted values are the result of a Generalized Additive
 489 Model (GAM) with lognormal distribution, taking into consideration the smooth terms of catch
 490 location estimated with thin plate regression splines and the quarter of the year used as a
 491 parametric term.

492

493 3.3.2. Modeling and predicting catch sizes

494 For blue shark, smaller specimens were predicted both inside the Cabo Verde EEZ and the 300
 495 nm adjacent waters, especially in the northeastern areas, as well as outside the study area
 496 towards the southwest. Seasonality was important in the blue shark predicted sizes, with
 497 overall smaller specimens expected during the spring months, in quarter 2 (**Figure 12**).

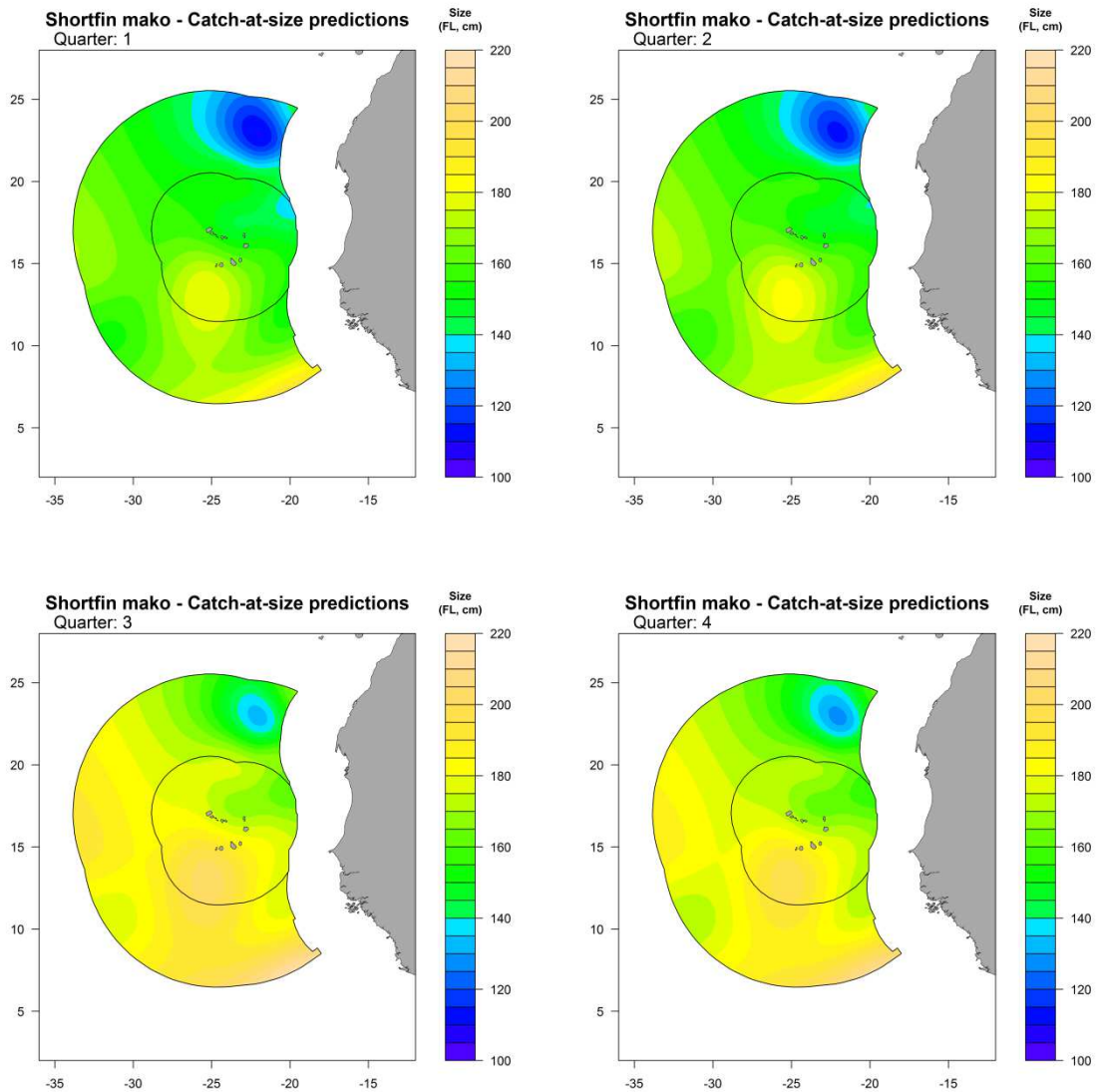
498 Nonetheless, it is important to note that the overall predicted blue shark sizes are relatively
499 large for the species, given that blue sharks mature at 185.1 cm FL (females) and 200.1 cm FL
500 (males). As such, most of the blue shark sizes predicted to occur in the study area along the
501 entire year corresponds to large juveniles or sub-adults, and adults.



502 **Figure 12.** Seasonal prediction of the size distribution of blue shark in the Cabo Verde EEZ and
503 adjacent waters (300 nm). The predicted values are the result of a Generalized Additive Model
504 (GAM) with Gaussian distribution and identity link function, taking into consideration the
505 smooth terms of catch location estimated with thin plate regression splines and the quarter of
506 the year as a parametric term.

507

508 For shortfin mako, there were also marked spatial effects in the predicted size of the
 509 specimens, in this case with smaller specimens expected inside the study area, especially in the
 510 northeastern waters. Smaller specimens were expected to occur in quarters 1 and 2 and larger
 511 specimens were expected to occur mainly in the 2nd semester (**Figure 13**). Contrary to blue
 512 shark, for shortfin mako the overall expected specimen sizes corresponded mostly to small size
 513 specimens. As such, most of the shortfin mako specimens expected to occur in the study area
 514 along the entire year, particularly the females, correspond to juveniles.



515 **Figure 13.** Seasonal prediction of the size distribution of shortfin mako shark in the Cabo Verde
 516 EEZ and 300 nm adjacent waters. The predicted values are the result of a Generalized Additive
 517 Model (GAM) with Gaussian distribution and identity link function, taking into consideration
 518 the smooth terms of catch location estimated with thin plate regression splines and the
 519 quarter of the year used as a parametric term.

520

521 **4. Discussion**

522 **4.1. Shark indicators in the regional (NE Atlantic) context**

523 In the region of the Cabo Verde EEZ and adjacent waters, the elasmobranch catch composition
524 from pelagic fisheries is largely composed by blue shark, followed by shortfin mako. This is
525 common in pelagic longline fisheries operating in other regions of the Atlantic Ocean (e.g.,
526 Mejuto et al., 2009; Coelho et al., 2012; Frédou et al., 2015). However, other less frequently
527 captured elasmobranch species, which include threshers, silky shark, longfin mako, oceanic
528 whitetip, hammerheads and the crocodile shark, were also recorded in similar proportions in
529 pelagic longlines operating in other Atlantic regions (e.g., Coelho et al., 2012).

530 A CPUE standardization procedure was carried out for the two main shark species using
531 statistical models, specifically Generalized Linear Models (GLMs) and Generalized Linear Mixed
532 Models (GLMMs). Such standardization procedure was carried out to remove the fishery-
533 dependent effects of the nominal CPUE data (i.e., spatial, seasonal and targeting effects), which
534 allows the estimation of relative indexes of abundance that can then be used as population
535 status indicators (Hilborn and Walters, 1992). The value of such standardization lies in the
536 improvement in the proportionality between the derived index and true abundance (Ye and
537 Dennis, 2009). Standardized CPUE series are also often used in stock assessment models by
538 most RFMOs (Regional Fisheries Management Organizations). In this case the results from the
539 CPUE standardization process showed an increase for the blue shark index of abundance over
540 the entire time series, between 2006 and 2015. For the shortfin mako the abundance index
541 was more variable, showing an increase in the earlier years of 2006-2009, followed by a
542 decrease in 2010, and then an overall slight increase in the more recent period until 2015.

543 In terms of targeting effects, the differences in fishing strategy used in the models reflect the
544 increased economic importance of sharks among the EU pelagic longline fleets, which
545 traditionally targeted swordfish almost exclusively. These changes in target species were
546 incorporated into the model by a proxy based on the ratio of the swordfish catch and the
547 combined swordfish and blue shark catches by set. This ratio is in general considered a good
548 proxy indicator of target criteria more clearly directed at swordfish *versus* a more diffuse
549 fishing strategy aimed at the two main species (i.e., swordfish and sharks). Moreover, this
550 methodology has been consistently applied to both EU fleets (Portuguese and Spanish) that
551 have a similar method of operation, including applications to the Atlantic and Indian Oceans

552 (e.g., Mejuto et al., 2013; Coelho et al., 2014). Other approaches for including targeting effects
553 into the CPUE standardization process have been tested in the past. Specifically, for the
554 Portuguese pelagic longline fishery, Coelho et al. (2015a) tested a cluster analysis based on the
555 catch composition of the 10 major species or species-groups, in an analysis, as suggested by He
556 et al. (1997), that has been successfully applied for CPUE standardization of other fleets (e.g.
557 Wang and Nishida, 2014; Hoyle et al., 2018). However, Coelho et al. (2015a) demonstrated that
558 for the Portuguese pelagic longline fleet (and EU fleets in general), given that the catches are
559 largely dominated by the two major species (i.e., swordfish and blue shark) the use of ratios or
560 clusters resulted in very similar results.

561 Size distribution trends can also be used as stock status indicators (Tu et al., 2018) and it was
562 observed that the catch of blue shark in the Cabo Verde region is mainly composed of relatively
563 large adult specimens. There were no major variations in the time series trends, with the mean
564 sizes relatively stable along the time period, both in the Cabo Verde EEZ and adjacent waters.
565 This further suggests that there are no signs of population declines. Both the CPUE and size
566 indicators for blue shark seem to indicate an apparently stable population in the region.

567 The catches of shortfin mako were mainly composed of small juvenile specimens, and there
568 was a general decreasing trend in the mean catch size during the time period. This catch
569 composed mainly of juvenile specimens and the general decreasing trend in mean sizes might
570 be an indicator of overfishing for this species in the region. The relatively large catch rates of
571 juvenile shortfin mako may also indicate that the Cabo Verde region and West African
572 continental shelf is functioning as an aggregation area for juvenile specimens, that become
573 vulnerable to the fisheries taking place in the region. Thus, fisheries indicators for shortfin
574 mako should be closely monitored, preferably based on fishery observer programs.

575 One important aspect of this study is that it used detailed data exclusively from the EU fleets
576 that operate in the Cabo Verde region (Portugal and Spain) but it should be noted that other
577 fleets from other countries also operate in the region (e.g. Asian fleets). As such, the results
578 presented here should be interpreted as representative only of the EU fleet component, while
579 the effects of other fleets operating in the region were not considered. Although it should be
580 noted that the Asian fleets traditionally target albacore and tropical tunas, setting their gear in
581 deeper water and thus having lower catch rates of pelagic sharks.

582 It is important to put these results and conclusions within a wider Atlantic perspective. For the
583 blue shark, the Ecological Risk Assessments carried out both in the Atlantic (Cortés et al., 2010;
584 Cortés et al., 2015) and Indian Oceans (Murua et al., 2013, 2018) showed that this species is

585 one of the most productive of all pelagic shark species and therefore capable of sustaining
586 relatively high levels of fishing mortality. Still, the overall vulnerability status was determined to
587 be intermediate, mainly due to the also relatively large susceptibility of blue shark to pelagic
588 fisheries, predominately pelagic longlines. The latest Atlantic blue shark stock assessments
589 carried out by ICCAT in 2015 showed that for the north Atlantic the stock was unlikely to be
590 overfished or subject to overfishing, even though there were high levels of uncertainty (Anon.,
591 2015). This contrasts with the South Atlantic stock where it was not possible to discount that in
592 recent years the stock may have been at levels near B_{MSY} and that fishing mortality has been
593 approaching F_{MSY} , implying that future increases in fishing mortality in the southern stock could
594 push the stock to an overfished state (Anon., 2015). The standardized CPUE increasing trends
595 observed in this study for the blue shark in the Cabo Verde region are in line with the trends
596 from the other fleets used in the last stock assessment by ICCAT. Specifically, for a number of
597 fleets that operate in the North Atlantic, including both eastern and western regions (Portugal,
598 Spain, Japan, US, Chinese-Taipei, Venezuela and Ireland). This general increasing trend has also
599 been registered since the mid-2000s.

600 For shortfin mako the Ecological Risk Assessments carried out in the Atlantic (Cortés et al.,
601 2010; Cortés et al., 2015) and Indian Ocean (Murua et al., 2013, 2018) ranked the species as
602 one of the most vulnerable of all pelagic sharks, mainly due to its very low productivity and
603 high susceptibility to fisheries, especially pelagic longlines. In the latest shortfin mako stock
604 assessment carried out by ICCAT in 2017 (Anon., 2017), the results indicated that there were
605 high probabilities that the North Atlantic stock was overfished and experiencing overfishing
606 (Anon, 2017). In terms of the CPUE indexes used on that assessment, most fleets showed
607 increases in stock abundance between 2000 and 2009, followed by reductions since then.
608 These results have been recently updated by ICCAT (Anon., 2019), which again highlighted the
609 poor stock condition of the North Atlantic stock. This is contrary to the results obtained in our
610 study, where the series between 2006 and 2015 was mostly stable or showing an increasing
611 trend. The reasons for the differences obtained might be related with the location of the
612 fisheries, as the series used for the stock assessment were coming from other fleets operating
613 mostly in different regions of the North Atlantic. As such, it is possible to hypothesize that even
614 though the shortfin mako biomass in the North Atlantic has experienced overall reductions due
615 to overfishing, the fraction of the population in the tropical NE Atlantic still seems to be
616 relatively stable in terms of biomass. This could be either because it is still in better condition
617 and/or because the region is a core area for the species in the Atlantic, where specimens tend

618 to aggregate and therefore signals in population declines might take longer time to be
619 detected.

620

621 **4.2. Satellite tagging**

622 During this study, tagged blue shark and shortfin mako showed considerable movement in the
623 region. Specifically the tagged blue sharks (tagged inside the Cabo Verde EEZ) showed very
624 variable movements in all directions, with sharks moving both inside and outside the EEZ. It
625 was noteworthy that for blue sharks, the females and adults tended to move further away from
626 the islands, while on the contrary the males and juvenile blue sharks tended to aggregate more
627 around the Cabo Verde islands.

628 As regards shortfin mako, a clearer pattern of movements was observed, with the sharks
629 tagged in the Cabo Verde EEZ tending to move mostly towards the West African continental
630 shelf. This corroborated the observations from the catches and the prediction models using
631 data from the commercial fisheries, where higher catch rates were also predicted for the
632 eastern parts of the study area. For shortfin mako, therefore, it seems that areas closer to the
633 African continental shelf, outside the Cabo Verde EEZ but in the EEZ of other African
634 continental countries, are of particular importance.

635 Other pelagic shark species are also present in the region. While those other species, such as
636 oceanic whitetip, silky shark, bigeye thresher, hammerheads and crocodile shark are not as
637 common in the region, they are also accidentally by-caught in pelagic longline fisheries, though
638 most of these species are now discarded due to ICCAT prohibition of retention and CITES
639 regulations. Some previous studies have focused on satellite tagging and habitat use for the
640 less common species, such as Coelho et al. (2015b) for the bigeye thresher and Santos and
641 Coelho (2018) for the smooth hammerhead. Still, the knowledge on those more rare species is
642 substantially lower than for the main shark species and therefore more effort should be put
643 into continued tagging for those species in the future.

644

645 **4.3. Spatial models and predictions**

646 Higher catch rates (in weight) were predicted outside the Cabo Verde EEZ for both blue and
647 shortfin mako. For shortfin mako, in particular, considerably higher CPUEs are predicted along
648 the African continental shelf, in areas outside the Cabo Verde EEZ but inside the EEZ of other

649 West African countries. As noted in section 4.2 above, this was corroborated with the satellite
650 tagging data that also showed that those areas along the West African shelf are of particular
651 importance for this species. These results show that even though the shortfin mako is an
652 oceanic and pelagic species, it seems to have a strong relation with continental shelf areas,
653 especially the juveniles. A recent study using satellite telemetry to map the movements of
654 shortfin mako shark in the West Atlantic (US and Mexico) concluded that shortfin mako
655 displayed very region-specific movements, with little distributional overlap between the Gulf of
656 Mexico/Caribbean Sea and the western North Atlantic (Vaudo et al., 2017). In the eastern
657 Atlantic, our study now seems to have reached similar conclusions and a comparable situation
658 might be occurring off West Africa, with shortfin makos showing the same type of region-
659 specific movements mainly along the West African continental shelf area.

660 There was considerable variability in the expected mean size for both species taking into
661 account spatial and seasonal effects. One important note, however, is that even though those
662 spatial and seasonal effects are important, in general the overall size of blue sharks was
663 expected to be composed mainly of relatively large adult individuals; whereas the overall size
664 of shortfin makos was expected to be mainly composed of relatively small juveniles. This was
665 consistent over the entire region and throughout the year. Both the spatial and seasonal effects
666 were influential in the expected mean size, in the case of the blue shark with the smaller
667 specimens occurring in the area mainly during spring months (quarter 2), while) and in the
668 case of shortfin makos the smaller specimens are expected to occur mainly in the 1st semester
669 during the winter and spring months.

670 When comparing those results within an Atlantic wide perspective, it is important to note that,
671 although there is some information available for blue shark, there is little information available
672 on the shortfin mako and for other pelagic shark species. Blue shark shows a strong size
673 latitudinal stratification pattern in all oceans, with a tendency for the larger adult specimens to
674 occur in warmer equatorial and tropical regions and the smaller juveniles occurring in colder
675 temperate waters (Coelho et al., 2018). However, for some other species the opposite pattern
676 has been found, as for example for the bigeye thresher in the Atlantic, where smaller and
677 younger sharks tend to concentrate predominantly in the tropical regions, while the larger
678 specimens seem to prefer temperate areas of the northern and southern Atlantic (Fernandez-
679 Carvalho et al., 2015b).

680

681 **5. Conclusions and recommendations**

682 As a final conclusion, the Cabo Verde region appears to be part of the Atlantic wide
683 distributional cycle where those shark species move through their life cycles. The blue shark
684 shows widespread and large scale movements in and out of the Cabo Verde EEZ as well as into
685 wider regions. The presence of the large adults in the Cabo Verde region corroborates the
686 previously hypothesized distributional patterns in the North Atlantic, with the large adult
687 specimens occurring mainly in warmer tropical waters (Coelho et al., 2018). In the case of the
688 shortfin mako, although the entire region appears to be an aggregation area for juveniles, the
689 region closer to the African continental shelf seems to be of particular importance to this
690 species, with large aggregations of small juvenile specimens.

691 The following are the main conclusions and recommendations from this study:

- 692 • Blue shark and shortfin mako are the main shark species captured in the pelagic
693 longlines, both in Cabo Verde archipelago EEZ as well as in other regions; this is the
694 same case of most oceanic-wide waters fished by pelagic longline gears;
- 695 • For both species the estimated indices of abundance for the Cabo Verde region showed
696 overall increases over the time series period (2006-2015);
- 697 • Blue sharks captured in the region are mainly large adults and there were no major
698 trends in mean sizes over time. By contrast, the shortfin makos caught in the region are
699 relatively small juveniles, and there were some indications of possible declines in the
700 mean sizes over time;
- 701 • Considering the abundance indexes, local depletion effects do not seem to be
702 occurring for those two shark species in the region as there are no signs of decreasing
703 local abundance (biomass). However, for the shortfin mako there are signs of a
704 decreasing trend in the mean sizes that can indicate overfishing on this species;
- 705 • The satellite tagged sharks showed high mobility of the specimens with movements
706 both inside and outside the Cabo Verde EEZ. In some cases, the sharks moved
707 considerable distances over the tagged periods, especially in the case of blue shark;
- 708 • The shortfin mako sharks seem to have marked region-specific movements and habitat
709 use, mainly along the West African continental shelf. This type of region-specific
710 movements has also been recently hypothesized for this species in the West Atlantic;
- 711 • The presence of the large adult blue shark in the Cabo Verde region corroborates the
712 hypothesis of the distributional patterns of this species in the North Atlantic, with large
713 adult specimens occurring mainly in warmer tropical waters and juveniles in colder
714 temperate and more coastal waters;

715 • For the shortfin mako the areas closer to the African continental shelf seem to be of
716 particularly importance, with large aggregations of small juvenile specimens. Such
717 areas should be of priority focus for the species conservation.

718

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736

737 **7. References**

738 Akaike, H. 1973. Information theory and an extension of the maximum likelihood principle. In:
739 2nd International Symposium on Information Theory (eds N.B. Petrov & F. Csáki). Akadémia
740 Kiadó, Budapest, pp. 267-281.

741 Amorim, A.F., Arfelli, C.A., Fagundes, L. 1998. Pelagic elasmobranchs caught by longliners off
742 southern Brazil during 1974-97: an overview. *Marine and Freshwater Research*, 49: 621-632.

743 Anon. 2014. Shark Species Group Inter-Sessional Report. ICCAT - International Commission for
744 the Conservation of Atlantic Tunas. March 10-14, 2014, Piriapolis, Uruguay. 11 pp + 21
745 appendices.

746 Anon. 2015. Report of the 2015 ICCAT blue shark stock assessment session. ICCAT -
747 International Commission for the Conservation of Atlantic Tunas. July 27-31, 2015. Lisboa,
748 Portugal. 115pp.

749 Anon. 2017. Report of the 2017 ICCAT shortfin mako assessment meeting. ICCAT - International
750 Commission for the Conservation of Atlantic Tunas. June 12-17, 2017. Madrid, Spain. 64pp.

751 Bates, D., Maechler, M., Bolker, B., Walker, S. 2013. lme4: linear mixed-effects models using
752 eigen and S4. R package version 1.0-5. <http://CRAN.R-project.org/package=lme4>.

753 Becker, R.A., Wilks, A.R., Brownrigg, R., Minka, T.P. 2013. maps: draw geographical maps, R
754 package version 2.3-6. <http://CRAN.R-project.org/package=maps>.

755 Buencuerpo, V., Rios, S., Moron, J. 1998. Pelagic sharks associated with the swordfish, *Xiphias*
756 *gladius*, fishery in the eastern North Atlantic Ocean and the Strait of Gibraltar. *Fishery Bulletin*,
757 96: 667-685.

758 Campana, S.E., Dorey, A., Fowler, M., Joyce, W., Wang, Z., Wright, D., Yashayaev, I. 2011
759 Migration pathways, behavioural thermoregulation and overwintering grounds of blue sharks
760 in the northwest Atlantic. *PLoS ONE*, 6: e16854.

761 Campbell, R.A. 2004. CPUE standardisation and the construction of indices of stock abundance
762 in a spatially varying fishery using general linear models. *Fisheries Research*, 70: 209–227

763 Casey, J.G., Kohler, N.E. 1992. Tagging studies on the shortfin mako shark (*Isurus oxyrinchus*) in
764 the western North Atlantic. *Australian Journal of Marine and Freshwater Research*, 43: 45-60.

765 Coelho, R., Fernandez-Carvalho, J., Lino, P.G., Santos, M.N. 2012. An overview of the hooking
766 mortality of elasmobranchs caught in a swordfish pelagic longline fishery in the Atlantic Ocean.
767 *Aquatic Living Resources*, 25: 311–319.

768 Coelho, R., Santos, M.N., Lino, P.G. 2014. Blue shark catches by the Portuguese pelagic longline
769 fleet between 1998-2013 in the Indian Ocean: Catch, effort and standardized CPUE. Working
770 Party on Ecosystems and Bycatch, IOTC Technical Paper. IOTC-2014-WPEB10-24. 32pp.

771 Coelho, R., Santos, M.N., Lino, P.G. 2015a. Standardized CPUE of blue shark in the Portuguese
772 pelagic longline fleet operating in the North Atlantic. ICCAT Sharks Working Group, SCRS
773 Document, ICCAT SCRS/2015/037.

774 Coelho, R., Fernandez-Carvalho, J., Santos, M.N. 2015b. Habitat use and diel vertical migration
775 of bigeye thresher shark: overlap with pelagic longline fishing gear. *Marine Environmental*
776 *Research*, 112: 91-99.

777 Coelho, R., Mejuto, J., Domingo, A., Yokawa, K., Liu, K-M., Cortés, E., Romanov, E., da Silva, C.,
778 Hazin, F., Arocha, F., Mwilima, A.M., Bach, P., Ortiz de Zarate, V., Roche, W., Lino, P.G., García-
779 Cortés, B., Ramos-Cartelle, A.M., Forselledo, R., Mas, F., Ohshimo, S., Courtney, D., Sabarros,
780 P.S., Perez, B., Wogerbauer, C., Tsai, W-P., Carvalho, F., Santos, M.N. 2018. Distribution patterns
781 and population structure of the blue shark (*Prionace glauca*) in the Atlantic and Indian Oceans.
782 *Fish and Fisheries*, 19: 90–106.

783 Compagno, L.J.V. 2001. Sharks of the world. An annotated and illustrated catalogue of shark
784 species known to date. Vol. 2. Bullhead, Mackerel and Carpet Sharks (Heterodontiformes,
785 Lamniformes and Orectolobiformes). FAO, Rome. 269pp.

786 Cortés, E., Arocha, F., Beerkircher, L., Carvalho, F., Domingo, A., Heupel, M., Holtzhausen, H.,
787 Santos, M.N., Ribera, M., Simpfendorfer, C. 2010. Ecological risk assessment of pelagic sharks
788 caught in Atlantic pelagic longline fisheries. *Aquatic Living Resources*, 23: 25-34.

789 Cortés, E., Domingo, A., Miller, P., Forselledo, R., Mas, F., Arocha, F., Campana, S., Coelho, R.,
790 Silva, C.D., Holtzhausen, H., Keene, K., Lucena, F., Ramirez, K., Santos, M.N., Semba-Murakami,
791 Y., Yokawa, K. 2012. Expanded ecological risk assessment of pelagic sharks caught in Atlantic
792 pelagic longline fisheries. *Collective Volume Scientific Papers ICCAT*, 71(6): 2637-2688.

793 de Carvalho, J.M.C. 2013. Elaboration of the Third International Conference on Sustainable
794 Development in Small Island States in Development. UNDP, Praia, Cabo Verde. Available at:
795 [https://sustainabledevelopment.un.org/content/documents/1064300CaboVerde_Report.ENGv](https://sustainabledevelopment.un.org/content/documents/1064300CaboVerde_Report.ENGversion.pdf)
796 [ersion.pdf](https://sustainabledevelopment.un.org/content/documents/1064300CaboVerde_Report.ENGversion.pdf).

797 Domeier, M.L., Kiefer, D., Nasby-Lucas, N., Wagschal, A., O'Brien, F. 2005. Tracking Pacific
798 bluefin tuna (*Thunnus thynnus orientalis*) in the northeastern Pacific with an automated
799 algorithm that estimates latitude by matching sea-surface temperature data from satellites
800 with temperature data from tags on fish. *Fishery Bulletin*, 103: 292–306.

801 Dunn, P.K. 2013. tweedie: Tweedie exponential family models. R package version 2.1.7.

802 Fay, M.P., & Shaw, P.A. 2010. Exact and asymptotic weighted logrank tests for interval censored
803 data: the interval R package. *Journal of Statistical Software*: 36, 1-34.

804 Fernandez-Carvalho, J., Coelho, R., Santos, M.N., Amorim, S. 2015a. Effects of hook and bait in
805 a tropical northeast Atlantic pelagic longline fishery: Part II—Target, bycatch and discard fishes.
806 *Fisheries Research*, 164: 312–321.

807 Fernandez-Carvalho, J., Coelho, R., Mejuto, J., Cortés, E., Domingo, A., Yokawa, K., Liu, K.M.,
808 García-Cortés, B., Forselledo, R., Ohshimo, S., Ramos-Cartelle, A.M., Tsai, W.P., & Santos, M.N.
809 2015b. Pan-Atlantic distribution patterns and reproductive biology of the bigeye thresher,
810 *Alopias superciliosus*. *Reviews in Fish Biology and Fisheries*, 25: 551–568.

811 Frédou, F.L., Tolotti M., Frédou, T., Carvalho, F., Hazin, H., Burgess, G., Coelho, R., Waters, J.D.,
812 Travassos P., Hazin F. 2015. Sharks caught by the Brazilian tuna longline fleet: an overview.
813 *Reviews in Fish Biology and Fisheries*, 25: 365–377.

814 Fox, J., Weisberg, S. 2011. An R Companion to Applied Regression (2nd ed). Thousand Oaks, CA,
815 Sage.

816 He, X., Bigelow, K.A., Boggs, C.H. 1997. Cluster analysis of longline sets and fishing strategies
817 within the Hawaii-based fishery. *Fisheries Research*, 31: 147-158.

818 Hijmans, R.J., 2016. raster: Geographic Data Analysis and Modeling. R package version 2.5-8.
819 <http://CRAN.R-project.org/package=raster>.

820 Hilborn, R., Walters, C.J. 1992. Quantitative Fisheries Stock Assessment. Choice, Dynamics and
821 Uncertainty. New York: Chapman and Hall.

822 Howey-Jordan, L.A., Brooks, E.J., Abercrombie, D.L., Jordan, L.K.B., Brooks, A., Williams, S.,
823 Gospodarczyk, E., Chapman, D.D. 2013. Complex movements, philopatry and expanded depth
824 range of a severely threatened pelagic shark, the oceanic whitetip (*Carcharhinus longimanus*)
825 in the western north Atlantic. *PLoS ONE*, 8: e56588.

826 Hoyle, S.D., Chassot, E., Fu, D., Kim, D.N., Lee, S.I., Matsumoto, T., Satoh, K., Wang, S-P., Yeh, Y-
827 M., Kitakado, T. 2018. Collaborative study of yellowfin tuna CPUE from multiple Indian Ocean
828 longline fleets in 2018. IOTC Document IOTC–2018–WPM09–12.

829 ICCAT, 2018. ICCAT Task I - Nominal catch information. Available online at:
830 <https://www.iccat.int/en/accesingdb.html>.

- 831 Lenth, R.V. 2014. lsmeans: least-squares means. R package version 2.11. [http://CRAN.R-](http://CRAN.R-project.org/package=lsmeans)
832 [project.org/package=lsmeans](http://CRAN.R-project.org/package=lsmeans).
- 833 Levene, H. 1960. Robust tests for equality of variances. In: Contributions to Probability and
834 Statistics: Essays in Honor of Harold Hotelling (eds I. Olkin, S.G. Ghurye, W. Hoeffding, W.G.
835 Madow & H.B. Mann). Stanford University Press, pp 278-292.
- 836 Lilliefors, H.W. 1967. On the Kolmogorov-Smirnov test for normality with mean and variance
837 unknown. *Journal of the American Statistical Association*, 62: 399-402.
- 838 Manly, B. 2007. Randomization Bootstrap and Monte Carlo Methods in Biology (3rd ed). New
839 York: Chapman & Hall/CRC.
- 840 Mejuto, J., García-Cortés, B., Ramos-Cardelle, A., de la Serna, J.M., 2009. Scientific estimations
841 of by-catch landed by the Spanish surface longline fleet targeting swordfish (*Xiphias gladius*) in
842 the Atlantic Oceans, with special reference to the years 2005 and 2006. *Collective Volume of*
843 *Scientific Papers ICCAT*, 64: 2455-2468.
- 844 Mejuto, J., García-Cortés, B., Ramos-Cardelle, A., De la Serna, J.M., González-González, I. 2013.
845 Standardized catch rates of shortfin mako (*Isurus oxyrinchus*) caught by the Spanish surface
846 longline fishery targeting swordfish in the Atlantic Ocean during the period 1990-2010.
847 *Collective Volume of Scientific Papers ICCAT*, 69(4): 1657-1669.
- 848 Montealegre-Quijano, S., Vooren, C.M. 2010. Distribution and abundance of the life stages of
849 the blue shark *Prionace glauca* in the Southwest Atlantic. *Fisheries Research*, 101: 168-179.
- 850 Murua, H., Coelho, R., Santos, M.S., Arrizabalaga, H., Yokawa, K., Romanov, E., Zhu, J.F., Kim,
851 Z.G., Bach, P., Chavance, P., Delgado de Molina, A., Ruiz, J. 2012. Preliminary Ecological Risk
852 Assessment (ERA) for shark species caught in fisheries managed by the Indian Ocean Tuna
853 Commission (IOTC). Working Party on Ecosystems and Bycatch. IOTC Document IOTC-2012-
854 WPEB08-31: 16pp.
- 855 Murua, H., Santiago, J., Coelho, R., Zudaire, I., Neves, C., Rosa, C., Zudaire, I., Semba, Y., Geng.,
856 Z., Bach, P., Arrizabalaga, H., Bach, P., Baez, J.C., Ramos, M. L., Zhu, J.F, Ruiz, J. 2018. Updated
857 Ecological Risk Assessment (ERA) for shark species caught in fisheries managed by the Indian
858 Ocean Tuna Commission (IOTC). IOTC Scientific Committee. IOTC Document IOTC-2018-SC21-
859 14: 28 pp.
- 860 Pratt, H.W. 1979. Reproduction in the blue shark, *Prionace glauca*. *Fishery Bulletin*, 77: 445-470.

861 Queiroz, N., Lima, F.P., Maia, A., Ribeiro, P.A., Correia, J.P., Santos, A.A. 2005. Movement of blue
862 shark, *Prionace glauca*, in the north-east Atlantic based on mark - recapture data. *Journal of*
863 *the Marine Biological Association of the United Kingdom*, 85: 1107-1112.

864 R Core Team. 2017. R: A language and environment for statistical computing. Version 3.4.0. R
865 Foundation for Statistical Computing, Vienna, Austria. <http://www.R-project.org>.

866 Rigby, C.L., Barreto, R., Carlson, J., Fernando, D., Fordham, S., Francis, M.P., Jabado, R.W., Liu,
867 K.M., Marshall, A., Pacoureaux, N., Romanov, E., Sherley, R.B., Winker, H. 2019. *Isurus*
868 *oxyrinchus*. The IUCN Red List of Threatened Species 2019: e.T39341A2903170.
869 <http://dx.doi.org/10.2305/IUCN.UK.2019-1.RLTS.T39341A2903170.en>.

870 Santos, C., Coelho, R. 2018. Migrations and habitat use of the smooth hammerhead shark
871 (*Sphyrna zygaena*) in the Atlantic Ocean. *Plos One*, 13(6): e0198664.

872 Shono, H. 2008. Application of the Tweedie distribution to zero-catch data in CPUE analysis.
873 *Fisheries Research*, 93: 154–162.

874 Silva, C.D., Kerwath, S.E., Wilke, C.G., Meyer, M., Lamberth, S.J. 2010. First documented
875 southern transatlantic migration of a blue shark *Prionace glauca* tagged off South Africa.
876 *African Journal of Marine Science*, 32: 639–642.

877 Stevens, J.D. 1976. Preliminary results of shark tagging in the north-east Atlantic, 1972-1975.
878 *Journal of the Marine Biological Association of the United Kingdom*, 56: 929-937.

879 Stevens, J.D. 1990. Further results from a tagging study of pelagic sharks in the north-east
880 Atlantic. *Journal of the Marine Biological Association of the United Kingdom*, 70: 707-720.

881 Stevens, J. 2009. *Prionace glauca*. In: IUCN 2012. IUCN Red List of Threatened Species. Version
882 2012.1. Available from: www.iucnredlist.org.

883 Tavares, R., Ortiz, M., Arocha, F. 2012. Population structure, distribution and relative
884 abundance of the blue shark (*Prionace glauca*) in the Caribbean Sea and adjacent waters of the
885 North Atlantic. *Fisheries Research*, 129-130: 137-152.

886 Tu, C.Y., Chen, K.T., Hsieh, C.H. 2018. Fishing and temperature effects on the size structure of
887 exploited fish stocks. *Scientific Reports*, 8: 7132.

888 Tudela, S., Kai, A., Maynou, F., El Andalossi, M., Guglielmi, P. 2005. Driftnet fishing and
889 biodiversity conservation: the case study of the large-scale Moroccan driftnet fleet operating in
890 the Alboran Sea (SW Mediterranean). *Biological Conservation*, 121: 65-78.

891 Vaudo, J.J., Byrne, M.E., Wetherbee, B.M., Harvey, G.M., Shivji, M.S. 2017. Long-term satellite
892 tracking reveals region-specific movements of a large pelagic predator, the shortfin mako shark,
893 in the western North Atlantic Ocean. *Journal of Applied Ecology*, 54: 1765–1775.

894 Wand, M. 2015. KernSmooth: Functions for Kernel Smoothing Supporting Wand & Jones
895 (1995). R package version 2.23-15. <http://CRAN.R-project.org/package=KernSmooth>

896 Wang, S-P., Nishida, T. 2014. CPUE standardization with targeting analysis for swordfish (*Xiphias*
897 *gladius*) caught by Taiwanese longline fishery in the Indian Ocean. IOTC Working Party on
898 Billfishes, Technical Paper. IOTC–2014–WPB12–22. 25pp.

899 Warnes, G.R., Bolker, B., Lumley, T., Johnson, R.C. 2013. gmodels: various R programming tools
900 for model fitting. R package version 2.15.4.1. <http://CRAN.R-project.org/package=gmodels>.

901 Wickham, H. 2009. ggplot2: Elegant Graphics for Data Analysis. New York: Springer.

902 Wickham, H. 2011. The split-apply-combine strategy for data analysis. *Journal of Statistical*
903 *Software*, 40, 1-29.

904 Wood, S.N. 2003. Thin plate regression splines. *Journal of the Royal Statistical Society: Series B*,
905 65, 95-114.

906 Wood, S.N. 2006. Generalized Additive Models: An Introduction with R. Chapman and
907 Hall/CRC.

908 Wood, S.N. 2011. Fast stable restricted maximum likelihood and marginal likelihood estimation
909 of semiparametric generalized linear models. *Journal of the Royal Statistical Society: Series B*,
910 73: 3-36.

911 Ye, Y., Dennis, D. 2009. How reliable are the abundance indices derived from commercial catch-
912 effort standardization? *Canadian Journal of Fisheries and Aquatic Sciences*, 66: 1169-1178.