

Demographic compensation among populations: what is it, how does it arise, and what 1 2 are its implications? 3 Jesús Villellas¹, Daniel F. Doak², María Begoña García³, and William F. Morris^{1,4} 4 5 6 1. Department of Ecology and Genetics, Uppsala University, Uppsala 75264, Sweden. E-mail: jesus.villellas@gmail.com. 7 8 2. Environmental Studies Program, University of Colorado Boulder, Boulder, CO 80309, 9 U.S.A. E-mail: daniel.doak@colorado.edu 3. Pyrenean Institute of Ecology (IPE-CSIC), Apdo. 13034, 50080 Zaragoza, Spain. E-mail: 10 11 mariab@ipe.csic.es 4. Dept. of Biology, Duke University, Box 90338, Durham, NC 27708-0338, U.S.A. E-mail: 12 13 wfmorris@duke.edu 14 Authorship: JV, WFM, and DFD originally envisioned the analysis. All authors collected some 15 of the demographic data in the field, and JV extracted additional data from the literature. JV 16 and WFM performed analyses and drafted the manuscript, and all authors contributed to 17 18 revisions. 19 20 **Running title:** Demographic compensation among populations 21 22 Keywords: demographic compensation, environmental gradient, geographical distribution, 23 global climate change, life-history trade-off, negative correlations, population growth rate, 24 range limit, sensitivity, vital rates

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33 Abstract

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Most species are exposed to significant environmental gradients across their ranges, but 35 vital rates (survival, growth, reproduction, and recruitment) need not respond in the same 36 37 direction to those gradients. Opposing vital rate trends across environments, a phenomenon that has been loosely called "demographic compensation", may allow species to occupy 38 39 larger geographical ranges and alter their responses to climate change. Yet the term has 40 never been precisely defined, nor has its existence or strength been assessed for multiple species. Here we provide a rigorous definition, and use it to develop a strong test for 41 42 demographic compensation. By applying the test to data from 26 published, multi-43 population demographic studies of plants, we show that demographic compensation commonly occurs. We also investigate the mechanisms by which this phenomenon arises by 44 45 assessing which demographic processes and life stages are most often involved. 46 Additionally, we quantify the effect of demographic compensation on variation in 47 population growth rates across environmental gradients, a potentially important 48 determinant of the size of a species' geographical range. Finally, we discuss the implications of demographic compensation for the responses of single populations and species' ranges 49 to temporal environmental variation and to ongoing environmental trends, e.g. due to 50 51 climate change. 52

54 Introduction

Environmental conditions typically vary across the geographical ranges of species. In turn, these environmental gradients often cause average vital rates (survival, growth, reproduction, and recruitment) to differ among local populations. The range limit is reached where the combined contributions of the vital rates to the population growth rate at low density are just sufficient to maintain a steady population (Birch 1953). On one side of the limit, populations can increase from low numbers, and on the other side they cannot (for elaborations to this basic scenario, see Pulliam 2000 and Holt 2009).

However, average vital rates may respond in more than one way to environmental 62 63 changes along a geographical gradient. A typical simplifying assumption in analyses of range 64 limits is that all vital rates will decline across the gradient (Fig. 1A) until a point is reached at which the low-density population growth rate equals 1 (i.e., at the range limit L1 in Fig. 1C; 65 66 but see Maurer & Brown 1989). Indeed, many theoretical and empirical studies begin with 67 the expectation that any measure of demographic performance will either decline or remain constant as a range boundary is approached (e.g., Jump & Woodward 2003; Stokes et al. 68 69 2004; Holt et al. 2005). However, vital rates need not respond in lockstep to shifting 70 environmental conditions along a geographical gradient. If they did, we would expect all 71 measures of individual and population performance to decline monotonically from the 72 range center to the periphery, a pattern that is not generally observed (Channell & Lomolino 73 2000; Sagarin & Gaines 2002; Sexton et al. 2009). Instead, vital rates may often change in opposite directions along spatial or environmental gradients (some deteriorating and others 74 improving; Fig. 1B), a phenomenon that has been termed "demographic compensation" 75 76 (Doak & Morris 2010). Vital rates that improve towards range margins will cause the 77 population growth rate to decline less steeply, even more so if those vital rates have a large

78 influence on the population growth rate (i.e., have high sensitivities; Caswell 2001). The 79 result could be a larger geographical range size, or a greater occupancy of sites within the 80 range, than would be the case in the absence of demographic compensation (Fig. 1C,D). 81 Currently, there is a growing interest in large-scale information about species' demographic 82 performance (e.g. Schurr et al. 2012; Merow et al. 2014; Ehrlén & Morris 2015), and understanding how vital rates covary along geographical and environmental gradients 83 84 becomes necessary for correctly predicting shifts in species' ranges in the present situation 85 of global change.

To clarify the phenomenon we are considering here, we note that the term 86 "demographic compensation" has been used in the ecological and evolutionary literature to 87 refer to multiple, disparate processes. For example, in the wildlife management literature, 88 89 focused largely on vertebrates, the term has been used to refer to density-dependent 90 improvements in some vital rates of a population in response to harvesting (e.g., Sterling et 91 al. 1983; Weaver et al. 1996; Boyce et al. 1999; Perón 2013), natural fluctuations in 92 mortality (Gittleman 1993), or disease-driven population declines (Lachish et al. 2009). In 93 this context, the term has sometimes been extended to cover behavioral changes that reduce human impacts on animal populations (McGowan et al. 2011). "Demographic 94 compensation" has also been used to refer to demographic changes within one population 95 96 following extreme climatic events that compensate for the negative impacts of those 97 events, possibly through density-dependent processes (Lloret et al. 2012). Finally, the term has been applied to compensatory changes in parasite density in response to host 98 resistance in host-parasite coevolution models (Nuismer & Kirkpatrick 2003). Given this 99 100 diverse past usage, authors should indicate clearly which definition of "demographic

101 compensation" they are using. Here, we mean changes in opposite directions in mean vital102 rates across populations.

103 Recent studies that have measured all vital rates across continental-scale gradients 104 have found negative correlations between pairs of vital rates among populations for three 105 plant species (Doak & Morris 2010; Villellas et al. 2013b). There are also numerous other 106 examples in the literature of negative vital rate correlations across time (Ferson & Burgman 107 1995; Van Tienderen 1995; Doak et al. 2005; Bell et al. 2013), across space (Elderd & Doak 108 2006; Barbeta et al. 2011; García-Camacho et al. 2012; Martone & Micheli 2012) or both (Jongejans and de Kroon 2005), or between individuals within populations (e.g., Bigler and 109 110 Veblen 2009). Moreover, classical life-history theory is built upon the analysis of negative correlations or tradeoffs between traits or fitness components (Stearns 1989, 1992; Roff 111 112 1992). However, to our knowledge, no previous studies have attempted to find general 113 patterns across multiple species in vital rate correlations among populations. While 114 empirical evidence certainly supports the existence of compensatory changes in vital rates, 115 three key questions about this phenomenon remain unanswered.

First, how should we test for the presence of demographic compensation in a set of demographic data? Precise criteria for its detection were not defined by Doak and Morris (2010) or others using the term in this context. Here, we propose a randomization procedure as a strong test for the presence of demographic compensation, and we apply it

to 26 published, multi-population demographic studies on plants.

121 The second question is: how does demographic compensation arise? Do certain types 122 of demographic processes (survival, growth, reproduction, recruitment and shrinkage rates) 123 or life history stages (e.g., early vs. late in life) tend to be involved in demographic 124 compensation more often than others? Does the importance of different processes or life 125 stages for demographic compensation depend on the longevity of the species? Does it depend on the potential influence of vital rates on population growth rates (sensitivity)? 126 Finally, does compensation tend to involve pairs of vital rates of the same (e.g., early- and 127 128 late-life survival) or different (e.g., survival and growth) types of demographic process, and 129 does it tend to involve pairs of rates affecting the same (e.g., early-life survival and growth) 130 or different (e.g., early-life survival and late-life growth) life stages? Patterns among species 131 in the way demographic compensation arises may indicate that certain kinds of life history 132 adjustments to environmental gradients are easier to attain than others. For example, it may be easier for different types of vital rates than to similar ones to respond in opposite 133 134 ways to the same environmental gradient. Comparisons of the patterns of demographic compensation among species have never been made before, but we begin to do so here. 135 136 The third question is: how effective is demographic compensation at reducing spatial 137 variation in population growth rates? The variance in growth rate among populations is an 138 inverse proxy for the potential range size of species, and thus informs us about the benefits 139 of demographic compensation. Specifically, in the absence of demographic compensation 140 across the sites a species inhabits, the variance in population growth rates would be high because there would be populations with intrinsic growth rate well above one, but also 141 142 locations where the growth rate falls below one (and where the species cannot persist for 143 long on its own). In contrast, in the case of complete compensation, the spatial variance in 144 population growth rate would be zero despite trends in individual vital rates. Low spatial 145 variation in population growth, all else being equal, should translate into a larger 146 geographical range, since more populations would present growth rates above one and thus persist in their locations (Fig. 1C,D). Doak and Morris (2010) have similarly argued that 147 148 compensatory changes among vital rates contributed to the absence of a trend in the

population growth rate across the continental ranges of two tundra plant species, despite
significant trends in the underlying vital rates. However, neither they nor any other authors
of whom we are aware directly quantified the magnitude of the effect of demographic
compensation on population growth across geographical ranges. Here, we propose a
method to gauge the stabilizing effect of demographic compensation on the population
growth rate across space, and we apply it to published multi-population datasets.

After addressing these three questions, we compare our approach to alternative ways of analyzing correlations among traits present in the literature, and we consider broader implications of demographic compensation for responses of populations and geographical distributions to ongoing changes in environmental conditions, for example due to global climate change. Finally, we highlight some future directions for research into demographic compensation and its effects.

161

162 How should we test for demographic compensation?

A good test for demographic compensation should not simply ask if any of the vital 163 164 rates are negatively correlated across populations, but must also account for additional complications. First, even if each vital rate varies randomly among populations, we would 165 166 expect to see some negative correlations arising by chance. Thus a more stringent criterion 167 for demographic compensation would ask: are there *more* negative correlations among the 168 vital rates as a group than we might expect to see if each vital rate varied at random? While demographic compensation involving only a few negative correlations might arise by 169 chance, an excess of negative correlations relative to random expectations would strongly 170 suggest that demographic compensation is intimately linked to life history trade-offs and 171 172 reflects the action of natural selection. Second, we must recognize that spatial variation in

173 population growth rates can be reduced not only by negative correlations in vital rates, but 174 also by an absence of positive correlations. Thus a strong criterion for demographic compensation might be the presence of more negative correlations or fewer positive 175 176 correlations than would be expected if all vital rates varied at random among populations. To handle these multiple needs, we propose using a randomization procedure as a 177 strong test for demographic compensation. Specifically, the estimated values of each vital 178 179 rate from multiple populations of a species are first randomly reassigned among 180 populations, and the correlations between all pairs of vital rates are computed. The numbers of statistically significant (P < 0.05) negative and positive correlations are recorded, 181 182 as judged by one-sided Spearman tests. This procedure is repeated 10,000 times, and the numbers of negative and positive correlations observed in the real demographic data are 183 compared to the percentiles of the randomization distributions. If the observed vital rate 184 correlation matrix includes more negative correlations than the 95th percentile of the 185 distribution of counts of negative correlations, or fewer positive correlations than the 5th 186 percentile of the distribution of counts of positive correlations, we would conclude that the 187 188 observed correlation matrix shows substantial evidence of demographic compensation. To put the basic procedure just described into practice, we modified it in one 189 190 important way, by computing the correlations not between the vital rates directly, but 191 between the "contributions" of each vital rate to variation in population growth rates, as 192 computed from a Life Table Response Experiment (LTRE) analysis (Caswell 2001; García-193 Camacho et al. 2012; see Appendix S1 in Supporting Information). Using LTRE contributions yields the same absolute value of the correlation coefficient between two vital rates, but 194 has two advantages over tests based directly on vital rates. First, contributions account for 195 196 the fact that while most vital rates, e.g. survival, increase the population growth rate

197 (positive sensitivity), some others, e.g. shrinkage to a smaller size, decrease it (negative sensitivity). Thus, our method ensures that compensatory changes between a vital rate with 198 199 positive sensitivity and a rate with negative sensitivity would show a negative correlation 200 between their vital rate contributions, even though the vital rates themselves would be 201 positively correlated. Second, LTRE contributions allow tests for compensation not only 202 between individual vital rates but also between groups of rates classified by demographic 203 process or life stage (e.g., all survival rates vs. all growth rates), since the contributions of 204 individual rates are additive (see Appendix S1, Table S1).

We applied this randomization test to 26 multi-population demographic studies of 205 206 plants (Table 1). Specifically, from the studies reviewed by Crone et al. (2011), we selected 207 all studies that included six or more populations, and we augmented these studies with 208 additional recently published papers that met the same criterion. The studies included plant 209 species of different life forms and habitats, and with populations arrayed at a range of 210 spatial scales, from local to continental. We calculated the vital rates for each population in each study directly from the average projection matrices presented in the original 211 212 publications (which typically did not present the actual vital rates). We focus on the vital rates that are estimated with the largest sample size, and are thus more accurately 213 calculated, by defining the vital rates as in Morris et al. (2008). Details of the estimation and 214 215 selection of vital rates and the computation of their LTRE contributions are given in 216 Appendix S1.

Twenty four out of 26 studies showed more - not fewer - positive correlations than would be expected if vital rates varied independently, suggesting that there is an overall tendency for vital rates to respond similarly to environmental gradients (see Appendix S1). On the other hand, nine studies showed a higher proportion of significant negative 221 correlations among contributions of vital rates than would be expected by chance either 222 when looking at single or grouped vital rates, and two more showed a marginally 223 significantly higher proportion of negative correlations (0.05 ; see Table S2). It is224 thus against a background of frequent positive vital rate correlations that demographic compensation operates through more-than-expected negative correlations. Given these 225 226 patterns, below we only analyze demographic compensation operating through negative 227 correlations. Demographic compensation was found with the same frequency at the level 228 of single vital rates and when vital rates were grouped by demographic process or life stage 229 (Table S2), which shows that similar vital rates can also act together to contribute to this 230 phenomenon.

231 The general prevalence of demographic compensation across taxa is difficult to evaluate, considering the exploratory nature of our study and our limited sample of 232 233 demographic studies. But in a frequentist sense, we would expect to encounter by chance 234 only one or two studies out of 26 (5%) with apparent demographic compensation if vital 235 rate correlations were in fact random. The finding of excess negative correlations in 9 236 studies (or 11 including marginally significant ones) indicates this phenomenon is unlikely to represent fortuitous variation. In addition, our ability to detect demographic compensation 237 238 seems to be impeded by methodological aspects of the studies that affect data quality and 239 statistical power. For example, four of the studies we analyzed were experimental (i.e., the 240 environmental gradient was created at the beginning of the study by the researchers), and 241 none of them showed demographic compensation. The imposed gradient in these studies may not have mimicked the relevant one against which compensatory vital rate patterns 242 have evolved in natural populations, or the experiments may not have allowed enough time 243 244 for demographic compensation to appear. Besides, for the 22 studies utilizing natural rather 245 than experimental gradients, we found that species showing demographic compensation had been monitored for longer than those not showing it (mean: 3.4 vs. 2.2 years; Mann-246 Whitney test, W = 21, P = 0.009; see also Appendix S1), and included on average more 247 populations (mean: 11.4 vs. 7.8 populations; Mann-Whitney test, W = 32, P = 0.074). Longer 248 249 studies should have better estimated the mean vital rates for each population, while imprecision of those estimates for shorter-term studies may have obscured additional 250 demographic compensation that we did not detect. Similarly, a higher number of 251 252 populations would allow smaller but statistically significant negative correlations to be detected. Thus it seems likely that the true frequency of demographic compensation is 253 higher than what we found. 254

There was no relationship between the longevity of the species (short- vs. long-lived) 255 and the occurrence of demographic compensation ($\chi^2 = 0.521$, df = 1, P = 0.471). Indeed, 256 257 compensation was found in annuals, short- and long-lived perennial herbs, and trees, and 258 studies without compensation included a similar variety of life forms (cf. Tables 1 and S2). We also found no difference between studies with vs. without a clearly defined 259 environmental gradient (χ^2 = 1.09, df = 1, P = 0.297), indicating that the underlying gradient 260 may be unapparent to researchers. Interestingly, all four taxa (Plantago coronopus, 261 Polygonum viviparum, Silene acaulis, and Clarkia xantiana ssp. xantiana; Table 1) that were 262 263 monitored across transects that spanned their geographical ranges showed strong evidence 264 of compensation. The ability of compensation to span whole distribution areas implies that it may play a role in increasing range size, even for species such as *Clarkia xantiana* ssp. 265 xantiana that has a relatively small range, but one that spans a steep environmental 266 267 gradient (in this case, in precipitation; Eckhart et al. 2011). Considering the limited number

of studies in our review, the patterns we have pointed out here are only suggestive, but
worthy of future study.

270

271 How does demographic compensation arise?

Having found evidence for demographic compensation in a substantial fraction of
published studies, in this section we analyze the mechanisms whereby demographic

274 compensation most frequently arises, looking for patterns across species. First, we examine

the relative roles of different types of demographic processes (i.e., survival, growth,

shrinkage, fecundity, and recruitment) and life stages in demographic compensation. Then,

277 we explore how the sensitivity, variance among populations, and LTRE contributions of vital

278 rates influence their role in demographic compensation.

279

280 Role of demographic processes and life stages

281 To assess whether some types of demographic processes participated in negative correlations among vital rates more often than others, we first tallied the numbers of vital 282 283 rate correlations including each demographic process that were significantly negative or not. We then used a generalized linear mixed model (GLMM, with binomial errors) to assess 284 whether the frequency of significant negative correlations differed among these processes 285 286 (our model included a random effect of study to account for the fact that the overall 287 fraction of vital rate correlations that were negative - regardless of the vital rates involved -288 differed among studies; see details in Appendix S1). Because the likelihood that a particular 289 process is involved in negative correlations might depend on the longevity of the species, 290 we also included in the GLMM a categorical variable for longevity (short- vs. long-lived; see 291 Table 1) and an interaction between longevity and type of demographic process.

The frequency of negative correlations differed among demographic processes, and 292 293 the pattern of variation differed between species depending on their longevity (see Fig. 2A, 294 Table S3). In short-lived species, survival was more commonly involved in negative correlations than were growth or shrinkage, whereas in long-lived species survival was the 295 296 process least frequently involved in negative correlations. However, fecundity and 297 recruitment were the processes contributing the most to compensation both in short- and 298 long-lived species.

299 Next we asked whether different life stages were more or less likely across species to participate in negative correlations. To make studies more comparable (they differed in the 300 301 number of stages originally reported in the projection matrices), we combined the stages 302 occurring in the first, middle, and last thirds of the life cycle (we refer to the combined stages as "early", "middle", and "late", respectively; see details in Appendix S1). We then 303 304 repeated the preceding analysis, using stage in place of demographic process. As for 305 processes, short- and long-lived species differed regarding the relative importance of 306 different life stages in their contribution to demographic compensation (see Fig. 2B, Table 307 S3). Late stages are most commonly involved in negative correlations in short-lived plants, while both early and late stages are commonly involved in long-lived plants, plants in the 308 309 middle life stage contributing little to demographic compensation in both long- and shortlived species. 310

311 Finally, we evaluated whether demographic compensation tended to involve 312 negative correlations between vital rates that represent the same vs. different types of 313 demographic processes or life stages, again using the GLMM approach. A significantly 314 greater fraction of vital rate pairs were negatively correlated when those rates represented 315 different demographic processes or affected different life stages (see Fig. 2, Table S3).

Although sample size is small, this result suggests that it may be easier to achieve opposing responses to environmental gradients the more different the processes or stages are, or put another way, that it may be difficult for the same type of demographic process or life stage to avoid responding in the same way to an environmental gradient.

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321 Sensitivity, variance and LTRE contributions

322 Fully understanding the roles of different demographic processes or life stages in 323 demographic compensation will require us to grapple with complex relationships between the sensitivity, among-population variance, and LTRE contribution of different vital rates. All 324 325 else being equal, vital rates can play a more effective role in demographic compensation if they make a larger contribution to the differences in population growth rate among 326 327 populations. Otherwise, they would be irrelevant to the among-population variance in 328 population growth, whether or not they are compensated by other rates (although we could 329 also find that multiple rates with low contributions jointly compensate for one or a few 330 rates with high contributions). A higher contribution requires high sensitivity of the 331 population growth rate to that vital rate, high among-population variance in the vital rate, or both (Caswell 2001). However, we might also expect to find that vital rates with high 332 333 sensitivities experience selection to reduce among-population variance in their means, the 334 spatial equivalent of the pattern of reduced temporal variance in high-sensitivity rates 335 pointed out by Pfister (1998). Although consequences of extreme variation in vital rates might be more detrimental at the temporal scale (permanent extinction) than across 336 species' ranges (lack of extinction of a whole species if one population disappears), the 337 338 evolutionary constraints that sensitivity might pose on moderate amounts of both types of 339 variation can be similar. Indeed, we find across all studies that the demographic process

340 with the highest average sensitivity across rates (recruitment) has the lowest average variance among populations, while the demographic process with the second lowest 341 342 average sensitivity (fecundity) has the highest average variance among populations (Fig. 3). 343 In reality, for most of the species we analyzed, the relationship between among-population variance and sensitivity of individual vital rates follows a triangular pattern in which high-344 variance/high-sensitivity rates are uncommon (results not shown). These considerations 345 346 suggest the conceptual model shown in Fig. 4, in which vital rates with high sensitivity could 347 be either more or less likely to play a role in demographic compensation, depending on 348 which path from sensitivity to compensation is stronger.

349 To further explore the role in demographic compensation of vital rates' LTRE contributions, sensitivity and variance, we classified vital rates according to whether their 350 351 LTRE contributions (in absolute value) were above or below the median contribution for 352 each species (for studies that show strong evidence for demographic compensation at the 353 level of single vital rates; see Table S2). We also classified vital rates into four groups based 354 on whether each rate's sensitivity (in absolute value) and variance were above or below the 355 median value of all rates for that species. Then we tested for differences in the percentage of significant negative correlations among these categories, using GLMMs (see Appendix 356 357 S1). Rates that made high LTRE contributions were indeed more often negatively correlated 358 with other rates than were rates with low contributions (see Fig. 5, Table S4), and this 359 difference was consistent across all species (see Table S5). This confirms our prediction that rates with higher LTRE contributions play more of a role in demographic compensation. 360 361 Moreover, rates that had both above-average sensitivity and above-average variance (which 362 would certainly be in the "high contribution" group) showed the highest frequency of 363 negative correlations (see Fig. 5, Table S4). Thus even though the negative correlation

364 between sensitivity and variance (Fig. 3) means these rates are uncommon, they do occur (see also Wisdom et al. 2000), and seem to be disproportionately important in demographic 365 366 compensation. On the contrary, rates with both low sensitivity and low variance were least 367 likely to be negatively correlated with other rates (see Fig. 5, Table S5). The previous finding that fecundity and recruitment rates were most frequently involved in demographic 368 compensation (Fig. 2) was likely related to their high variance and sensitivity, respectively 369 370 (Fig. 3). However, detailed analyses of stage-specific vital rates (rather than rates grouped 371 according to general demographic processes or stages) in short- and long-lived species will 372 be needed to understand the preliminary patterns shown here and their ecological 373 implications. Sensitivity, variance, and LTRE contributions all seem to shape the roles that 374 particular vital rates play in demographic compensation, but in complex ways that warrant 375 further examination using a greater number of studies.

376

377 How effective is demographic compensation at reducing spatial variation in population 378 growth rates?

379 As we noted in the Introduction, we can assess the effectiveness of demographic compensation by quantifying the degree to which it reduces variation in population growth 380 381 rates across space. To do so, we used an approach similar in spirit to the randomization 382 procedure we used to test for strong demographic compensation. Our goal in these 383 randomizations was to eliminate as much as possible the observed negative correlations 384 among vital rates while minimizing changes to the positive vital rate correlations (which 385 showed no role in demographic compensation). First, we determined, for each vital rate 386 involved in significant negative correlations, whether the sum of the LTRE contributions of 387 all of the vital rates with which that rate was negatively correlated was greater than the

388 equivalent sum of contributions for its positively correlated rates. We randomized only those focal vital rates that met this criterion. The correlations between pairs of vital rates 389 that both failed to meet the "sum of contributions" criterion were therefore unchanged. 390 391 This randomization procedure eliminates as much as possible the negative correlations 392 while preserving the important positive correlations, and results in a valid correlation matrix. For each randomized set of vital rates (10,000 replications), we then constructed a 393 394 projection matrix, estimated the population growth rate for each population, and computed 395 the among-population variance in the population growth rate. Finally, we obtained the null 396 distribution for variance in population growth rates. This analysis was applied to each 397 species showing strong demographic compensation at the level of single vital rates. 398 Alternative approaches based on direct manipulations of the vital rate correlation matrices, 399 e.g. converting all negative values to zero, would not be valid due to mathematical 400 constraints on correlation matrices (see Appendix S1).

401 On average, the variance of population growth rates estimated from the observed 402 projection matrices, reflecting actual populations, is 48% of the median variance in the 403 random permutations, indicating that demographic compensation approximately halves the variance in population growth (see Table S2). Such reduction in variance was significant in 404 405 four out of seven studies, and marginally significant in two other studies. However, this may 406 be an underestimate of the importance of demographic compensation: although in the 407 randomizations we deleted virtually all the negative correlations, the constraints of our procedure also removed on average 20% of the positive correlations (the least we could 408 409 achieve), which also influenced variance in population growth rates. We found indeed that 410 the ratio of observed to randomization variance decreased with a higher proportion of the 411 original positive correlations retained in the randomizations (see Fig. S2). For species for

412 which we kept 90% or more of the original positive correlations in the randomizations, the observed variance of the population growth rate was on average about a quarter of the 413 414 median variance in the randomizations. Additionally, we found no differences in the mean 415 population growth rates between observed and randomized data, indicating that our results were due to changes in the correlations per se (results not shown). In summary, the levels of 416 417 demographic compensation we observe in the published studies suggest that it can 418 substantially reduce spatial variation in population growth, and thus may considerably 419 increase the size of a species' geographical range relative to the case in which all vital rates 420 respond similarly to environmental gradients.

421

422 Alternative ways of analyzing trait correlations

423 The search for patterns in the correlations between vital rates to better understand 424 if and how demographic compensation arises is reminiscent of the search for patterns in the 425 correlations between traits of individual organisms, often used to better understand 426 morphological integration (e.g., Cheverud et al. 1989) or to identify suites of traits that have 427 been selected as a whole (e.g., Pérez-Barrales et al. 2014). In this section, we compare the methodological approaches we have used to examine demographic compensation with 428 429 approaches used in trait-based studies to examine trait correlations. Our primary goal is to 430 discuss why several of these trait-based approaches are not appropriate for testing 431 demographic compensation. Readers not interested in this comparison can skip to the 432 following section.

In the trait literature, one approach has been to use Mantel tests to assess the fit of
the observed correlations to a theoretical matrix representing a hypothesis of how the traits
should be related (Cheverud *et al.* 1989). For example, one might hypothesize that traits

436 describing a particular body region or involved in a similar function should be positively correlated while traits of different regions or functions should be uncorrelated, and this 437 438 hypothesis could be represented by a theoretical matrix of 0's and 1's (or -1's if there is 439 reason to expect traits to be negatively correlated). But the approach we take to analyze 440 demographic compensation is different for several reasons. We do not necessarily expect that *all* rates describing a particular demographic process (for example) will be negatively 441 442 correlated with other rates. Instead, we have asked whether negative correlations are more frequent for certain processes or stages than others (or when processes or stages are the 443 444 same or different). Another limitation of the Mantel test approach is that there is a ceiling 445 to the number of negative values that can occur in a correlation matrix, so we could not have a reasonable hypothesis that *all* correlations involving different types of rates would 446 be negative. 447

Another approach that has been applied in the trait literature is to conduct a non-448 449 parametric bootstrap from the original data to place confidence limits on the mean correlation between pairs of traits involved in the same vs. different functions, such as 450 451 insect attraction vs. pollen transfer (Conner & Stirling 1995; Pérez-Barrales et al. 2014). This approach could be applied in studying demographic compensation. For example, we could 452 453 repeatedly sample populations with replacement from the set of populations in a study, 454 compute the vital rate correlations across populations, calculate the mean correlation for all 455 vital rate pairs involving, say, survival, and use the confidence limits for the mean 456 correlation to ask if survival has a lower average correlation with other rates. However, as 457 most vital rate pairs are positively correlated or uncorrelated, calculating the average 458 correlation for a type of vital rate (instead of its frequency of negative correlations, as we

459 compute in the present study) might mask the presence of uncommon negative460 correlations, even though they are important for demographic compensation.

461 An underutilized approach that holds promise for visualizing correlation patterns in both traits and vital rates is to use exploratory methods such as "corrgrams", which are 462 simply colour-coded correlation matrices that reveal groups of variables that show similar 463 correlation patterns with other variables (Friendly 2002). While one can use principal 464 465 components analysis to optimally define groups of variables based on their correlation 466 patterns (for details, see Friendly 2002), perhaps more useful for exploring patterns of 467 demographic compensation are corrgrams in which vital rates are grouped by either life 468 stage or by type of demographic process. Such corrgrams for one of the studied species, Silene acaulis (Fig. 6), reveal that vital rates for a given life stage show a mixture of positive 469 and negative correlations with vital rates for other stages (Fig. 6A), in contrast with the 470 471 general trend across species shown in Fig. 2B. But when vital rates are grouped into 472 different types of demographic process (Fig. 6B), the corrgram reveals frequent negative 473 correlations between vital rates representing different types of process (e.g., survival and 474 fecundity, or survival and growth), in accordance with Fig. 2A. Thus corrgrams are an easy way to see that for Silene acaulis, demographic compensation arises through opposing 475 476 responses of different types of demographic process rather than different life stages.

477

478 Broader implications, caveats, and future directions

In this study, we have searched for patterns in mean vital rates across populations
that indicate the potential for demographic compensation across space. However,
environmental conditions and vital rates vary in both space and time. Opposing patterns of
variation in vital rates across time within a single population have the potential to increase

483 the long-run stochastic population growth rate (Tuljapurkar 1990). These different types of demographic compensation, which might be called spatial vs. temporal compensation, 484 485 respectively, may be produced by the same underlying mechanism. If the effects of 486 particular biotic or abiotic factors (e.g., temperature) on vital rates are consistent in all 487 populations, the same relationship between vital rates and temperature (Fig. 7A) may 488 produce both spatial compensation among populations (Fig. 7B) and temporal 489 compensation within populations (Fig. 7C). For example, Doak and Morris (2010) found that 490 vital rates responded similarly to among-year temperature variation in populations throughout the continent-wide ranges of two tundra plants (although they also observed 491 492 site-specific components of the vital rates). This same vital rate/temperature relationship 493 across a species' range may also produce compensatory responses in the means of different vital rates to trends in average environmental conditions, such as climate warming (Fig. 7D), 494 495 which in turn may buffer local populations against those environmental trends (although as 496 emphasized by Doak and Morris (2010), such buffering may not last indefinitely). 497 However, a given vital rate may not follow the same relationship with an 498 environmental driver in all populations, in particular if there is adaptation to local conditions across a species range. If the responses of one or more vital rates to the same 499 500 environmental driver differ among populations, then we may observe spatial compensation, 501 temporal compensation, both, or neither (Fig. 8). We might also find for some taxa that vital 502 rates covary positively along the most favorable stretches of environmental gradients (perhaps due to fewer resource limitations) and demographic compensation is thus more 503 restricted to environmental – and geographical - extremes. To fully understand which types 504 505 of demographic compensation are operating and in which populations, we must measure 506 the relationships between vital rates and environmental drivers in multiple populations

507 spanning the conditions experienced across the species' range. Although doing so would require longer and larger scale demographic studies, as well as measurements of potential 508 509 drivers, it would yield better insight into the nature and mechanisms of demographic 510 compensation than simply analyzing patterns in the correlations of mean vital rates among populations. Once the relationships between vital rates and environmental drivers are 511 512 established, we could compute mean values of vital rates across species' ranges given local 513 environmental conditions, and test for the presence of demographic compensation with 514 randomization procedures as we did here. However, our ability to do so at present is severely limited by the fact that very few studies have quantified the relationships between 515 516 demographic processes and environmental conditions at large geographical scales or across 517 multiple populations (but see Doak & Morris 2010; Eckhart et al. 2011; Villellas et al. 2013a). More such studies are clearly needed. 518

519 Knowing how vital rates respond to environmental drivers would also be valuable for 520 predicting shifts in species' geographical ranges induced by climate change. The dominant methodology now used to forecast range shifts is species distribution models (SDMs; also 521 522 known as niche or climate envelope models), which correlate known presences (and sometimes absences) of a species with environmental factors (Elith & Leathwick 2009). An 523 524 alternative would be to use multi-site demographic models with vital rates tied to specific 525 environmental drivers to identify locations where populations can increase (Merow et al. 526 2014) or where equilibrium abundance is above zero (Ehrlén & Morris 2015). SDMs and demographic models don't always make the same assessments of the environmental 527 suitability where present-day populations are found (Eckhart et al. 2011, Gerst et al. 2011, 528 Pironon *et al.* 2015), so their predictions about future range shifts may not be consonant. In 529 530 particular, while demographic models can account explicitly for demographic compensation, it would seem a significant challenge for SDMs to correctly predict the consequences of
ongoing compensation in the face of climatic trends of the sort shown in Fig. 7D, as more
extreme climates not experienced by current populations would be inferred by SDMs to be
outside the "climate envelope", even though compensation could allow populations to
persist under those extremes. But it remains to be seen whether more mechanistic,
demographic models will do a better job of predicting future range shifts than purely
correlational, but simpler, SDMs.

538 An issue that has not been addressed in this study is the role played by phenotypic 539 plasticity and local adaptation in the variation in mean vital rates among populations. The 540 relative roles played by phenotypic plasticity vs. local adaptation in shaping the performance of a species across its range are still vigorously debated (Sultan 2004; Sexton et 541 al. 2009; Valladares et al. 2014). Both of these processes, alone or in combination, could be 542 543 responsible for improvements in some vital rates that compensate for deterioration of 544 others near range limits (Fig. 9). Only by performing field reciprocal transplants or 545 laboratory experiments under controlled environmental conditions will we be able to 546 partition the roles played by phenotypic plasticity and local adaptation in demographic compensation. 547

Demographic compensation and life-history tradeoffs are related, but not necessarily identical, phenomena. The strongest indicator of a life-history tradeoff is a negative genetic correlation between two fitness components in a set of genotypes or families, typically from the same population, such as when genotypes with higher early-life reproduction show lower subsequent survival, growth, or reproduction (Reznick 1992; Roff 1992). However, negative correlations between fitness components (also known as vital rates) across populations - or even species - have also been interpreted as the result of natural selection 555 acting on life histories in the face of negative genetic correlations (Stearns 1992). Thus fixed genetic differences between populations could both underlie negative correlations between 556 557 mean vital rates (i.e., demographic compensation) and reflect the consequences of life-558 history tradeoffs. In this context, an interesting question is whether vital rates of different 559 genotypes within populations show negative correlations that are similar to those between mean vital rates across populations. But demographic compensation could also arise if all 560 561 populations contain a single, but plastic, genotype, provided that plasticity in different vital 562 rates leads to opposing changes in mean vital rates across populations. Clearly, the 563 relationship between demographic compensation and life-history tradeoffs deserves more 564 attention in the future.

An important caveat of our analysis is that the demographic studies we analyzed did 565 566 not manipulate intraspecific density, or use statistical methods to distinguish the impact of 567 density from the impact of environmental factors on the patterns of vital rate variation 568 among populations. In fact, given that demographers have a practical tendency to conduct 569 their studies where the focal species is fairly abundant, the populations we analyzed were 570 likely to have been at or even above the local carrying capacity (Buckley et al. 2010), although a few studies claimed that density effects in the monitored populations were 571 572 negligible (Werner & Caswell 1977; Ehrlén 1995; Valverde & Silvertown 1998; Kaye et al. 573 2001). As Birch (1953) and others have argued, the population growth rate at low density 574 (i.e., the finite rate of increase) is perhaps the most cogent indicator of the potential geographical range of a species. Thus to assess the role that demographic compensation 575 576 plays in increasing a species' range, we would want to know how it operates at low density, but what we know best is likely to be how it operates near carrying capacity. However, the 577 578 effect of environmental factors and intraspecific density on the vital rates need not be

additive, and disentangling whether vital rates show similar patterns of demographic
compensation at low and high density remains an important unsolved challenge.

581 Our analysis only involved demographic studies of plants. Fewer multi-population 582 demographic studies have been conducted for animal species than for plants (but see, e.g., Martone & Micheli 2012), making it more difficult to assess the frequency of demographic 583 compensation for animals. Future comparative analysis, as more studies spanning the 584 585 ranges of animals become available, would advance our knowledge of whether 586 demographic compensation is a general feature of living organisms. The growing availability 587 of data sets such as COMPADRE and COMADRE (Salguero-Gómez et al. 2015) may make 588 such analyses increasingly feasible, although these databases would be even more useful if they reported vital rates directly, so that they did not have to be extracted from the matrix 589 elements as we did here. 590

591 The fact that we found strong evidence for demographic compensation when we 592 analyzed demographic studies that measured all vital rates suggests that we may fail to 593 understand the demographic causes of geographical range limits if we consider only one or 594 a few processes in the life cycle. This is especially true in light of evidence presented here that demographic compensation commonly takes place between rates of different 595 596 processes or stages. The demography of multiple populations should be quantified as 597 comprehensively as possible in order to understand the mechanisms responsible for the 598 sizes of species' geographical ranges. On the other hand, if additional studies strengthen 599 generalities about which processes are usually involved in demographic compensation, this 600 information may help us to save time or money in monitoring by targeting efforts on 601 processes that most frequently present opposing patterns of variation across gradients. One well-studied example is the tradeoff between survival and growth rates of conifers that
appears at multiple scales and for multiple species (Bigler & Veblen 2009).

604 Demographic compensation, through negative correlations between vital rates, 605 seems to be a relatively common phenomenon across populations of plant species, 606 although our ability to detect it depends on the spatial and temporal scope of a study. We 607 have started to untangle some mechanisms whereby demographic compensation takes 608 place, such as the involvement of vital rates of differing demographic processes or life 609 stages. In addition, the high sensitivity and variance found in the vital rates involved in 610 negative correlations strongly suggest that demographic compensation has not arisen as a 611 random phenomenon, but as an advantageous process with potential evolutionary implications in terms of species performance and geographical distribution. Of course, the 612 613 detection of demographic compensation in multi-population demographic studies does not 614 discount the possibility that some species may persist in a variable world by means of small 615 ranges with little environmental variation, or by narrowly tracking particular environmental conditions over broader areas (i.e., extreme habitat specialists). Finally, we found in our 616 617 study that, on average, negative correlations between vital rates reduce by half the variance among population growth rates, a reduction likely underestimated due to constraints of the 618 619 analysis. Demographic compensation thus might be crucial for understanding the current 620 distribution of species and accurately predicting range shifts in the face of ongoing 621 environmental changes. At present, a scarcity of studies with large temporal and spatial 622 scales limits our ability to further analyze this phenomenon. However, a growing 623 accumulation of demographic data - particularly for underlying vital rates - for species with a 624 variety of life histories and environments, coupled with information on the biotic and abiotic

625	drivers of those vital rates, can only improve our ability to understand demographic
626	compensation and its consequences in the future.
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639	References
640	1.
641	Barbeta, A., Peñuelas, J., Ogaya, R. & Jump A.S. (2011). Reduced tree health and seedling
642	production in fragmented Fagus sylvatica forest patches in the Montseny Mountains (NE
643	Spain). Forest Ecolog. Manag., 261, 2029-2037.
644	2.
645	Bell, T. J., Powell, K. I. & Bowles, M. L. (2013). Viability model choice affects projection
646	accuracy and reintroduction decisions. J. Wildlife Manag., 77, 1104-1113.
647	3.

- Bigler, C. & Veblen, T. T. (2009). Increased early growth rates decrease longevities of
- conifers in subalpine forests. *Oikos*, 118, 1130-1138.
- 650 4.
- Birch, L.C. (1953). Experimental background to the study of the distribution and abundance
- of insects. 1. The influence of temperature, moisture and food on the innate capacity for
- 653 increase of 3 grain beetles. *Ecology*, 34, 698-711.
- 654 5.
- Boyce, M.S., Sinclair, A.R.E. & White, G.C. (1999). Seasonal compensation of predation and
- 656 harvesting. *Oikos*, 87, 419-426.
- 657 6.
- Bruna, E. M. (2003). Are plant populations in fragmented habitats recruitment limited? Tests
 with an Amazonian herb. *Ecology*, 84, 932-947.
- 660 **7**.
- 661 Buckley, Y.M., Ramula, S., Blomberg, S.P., Burns, J.H., Crone, E.E., Ehrlén, J. et al. (2010).
- 662 Causes and consequences of variation in plant population growth rate: a synthesis of matrix
- 663 population models in a phylogenetic context. *Ecol. Lett.*, 13, 1182–1197.
- 664 **8**.
- 665 Caswell, H. (2001). *Matrix Population Models. Construction, Analysis and Interpretation*. 2nd
- edition. Sinauer, Sunderland, MA, USA.
- 667 9.
- 668 Channell, R., & Lomolino, M. V. (2000). Dynamic biogeography and conservation of
- 669 endangered species. *Nature*, 403, 84-86.
- 670 10.

- 671 Cheverud, J.M., Wagner, G.P. & Dow, M.M. (1989). Methods for the comparative-analysis of
- variation patterns. *Syst. Zool.*, 38, 201-213.
- 673 11.
- 674 Conner, J.K. & Sterling, A. (1995). Testing hypotheses of functional relationships a
- 675 comparative survey of correlation patterns among floral traits in 5 insect-pollinated plants.
- 676 Am. J. Bot., 82, 1399-1406.
- 677 12.
- 678 Crone, E. E., Menges, E.S., Ellis, M.M., Bell, T., Bierzychudek, P., Ehrlén, J. et al. (2011). How
- do plant ecologists use matrix population models? *Ecol. Lett.* 14, 1-8.
- 680 13.
- Davelos, A. L. & Jarosz, A. M. (2004). Demography of American chestnut populations: effects
 of a pathogen and a hyperparasite. *J. Ecol.*, 92, 675-685.
- 683 14.
- Davison, R., Jacquemyn, H., Adriaens, D., Honnay, O., de Kroon, H. & Tuljapurkar, S. (2010).
- 685 Demographic effects of extreme weather events on a short-lived calcareous grassland
- 686 species: stochastic life table response experiments. J. Ecol., 98, 255-267.
- 687 15.
- 688 Doak, D.F. & Morris, W.F. (2010). Demographic compensation and tipping points in climate-
- 689 induced range shifts. *Nature* 467, 959-962.
- 690 16.
- 691 Doak, D.F., Morris, W.F., Pfister, C., Kendall, B.E. & Bruna, E.M. (2005). Correctly estimating
- how environmental stochasticity influences fitness and population growth. Am. Nat., 166,
- 693 E14-E21.
- 694 **17**.

- 695 Eckhart, V.M., Geber, M.A., Morris, W.F., Fabio, E.S., Tiffin, P. & Moeller, D.A. (2011). The
- 696 geography of demography: long-term demographic studies and species distribution models
- reveal a species border limited by adaptation. *Am. Nat.*, 178, S26-S43.
- 698 **18**.
- 699 Ehrlén, J. (1995). Demography of the perennial herb Lathyrus vernus. II. Herbivory and
- 700 population dynamics. *J. Ecol.*, 83, 297-308.
- 701 19.
- 702 Ehrlén, J. & Morris, W.F. (2015). Predicting changes in the distribution and abundance of
- species under environmental change. *Ecol. Lett.*, 18, 303-314.
- 704 20.
- 705 Elderd, B.D. & Doak, D.F. (2006). Comparing the direct and community-mediated effects of
- disturbance on plant population dynamics: flooding, herbivory and Mimulus guttatus. J.
- 707 *Ecol.*, 94, 656-669.
- 708 21.
- 709 Elith, J. & Leathwick, J.R. (2009). Species distribution models: ecological explanation and
- prediction across space and time. *Annu. Rev. Ecol. Evol. Syst.*, 40, 677-697.
- 711 22.
- Endels, P., Jacquemyn, H., Brys, R. & Hermy, M. (2007). Genetic erosion explains deviation
- from demographic response to disturbance and year variation in relic populations of the
- perennial *Primula vulgaris*. J. Ecol., 95, 960-972.
- 715 23.
- 716 Ferson, S. & Burgman, M.A. (1995). Correlations, dependency bounds and extinction risks.
- 717 Biol. Conserv., 73, 101-105.
- 718 24.

Friendly, M. (2002). Corrgrams: exploratory displays for correlation matrices. *Am. Stat.*, 56,

720 316-324.

721 25.

- 722 García-Camacho, R., Albert, M.J. & Escudero, A. (2012). Small-scale demographic
- compensation in a high-mountain endemic: the low edge stands still. Plant Ecol. Divers., 5,
- 724 37-44.
- 725 26.
- 726 Gerst, K.L., Angert, A.L. & Venable, D.L. (2011). The effect of geographic range position on
- demographic variability in annual plants. J. Ecol., 99, 591-599.
- 728 27.
- 729 Ghimire S.K., Gimenez O., Pradel R., McKey D. & Aumeeruddy-Thomas Y. (2008).
- 730 Demographic variation and population viability in a threatened Himalayan medicinal and
- 731 aromatic herb Nardostachys grandiflora: matrix modelling of harvesting effects in two
- contrasting habitats. J. Appl. Ecol., 45, 41-51.
- 733 28.
- 734 Gittleman, J.L. (1993). Carnivore life histories: a re-analysis in the light of new models. In:
- 735 *Mammals as predators* (ed. Dunstone, N. & Gorman, M.L.). Clarendon Press, Oxford, UK, pp.
- 736 65-86
- 737 29.
- Holt, R.D. (2009). Bringing the Hutchinsonian niche into the 21st century: ecological and
- evolutionary perspectives. Proc. Natl. Acad. Sci. U. S. A., 106, 19659-19665.
- 740 30.
- Holt, R.D., Keitt, T.H., Lewis, M.A., Maurer, B.A. & Taper, M.L. (2005). Theoretical models of
- species' borders: single species approaches. *Oikos*, 108, 18-27.

743 31.

- 744 Hunt, L. P. (2001). Heterogeneous grazing causes local extinction of edible perennial shrubs:
- a matrix analysis. J. Appl. Ecol., 38, 238-252.
- 746 32.
- 747 Hyatt, L.A. & Araki, S. (2006). Comparative population dynamics of an invading species in its
- native and novel ranges. *Biol. Invasions*, 8, 261-275.
- 749 33.
- Jacquemyn, H. & Brys, R. (2008). Effects of stand age on the demography of a temperate
- forest herb in post-agricultural forests. *Ecology*, 89, 3480-3489.
- 752 34.
- Jiménez-Sierra, C., Mandujano, M.C. & Eguiarte, L.E. (2007). Are populations of the candy
- 754 barrel cactus (Echinocactus platyacanthus) in the desert of Tehuacán, Mexico at risk?
- Population projection matrix and life table response analysis. *Biol. Conserv.*, 135, 278-292.
- 756 35.
- Jongejans, E., Jorritsma-Wienk, L.D., Becker, U., Dostál, P., Mildén, M. & de Kroon, H. (2010).
- 758 Region versus site variation in the population dynamics of three short-lived perennials. J.
- 759 *Ecol.*, 98, 279-289.
- 760 36.
- Jongejans, E. & de Kroon, H. (2005). Space versus time variation in the population dynamics
 of three co-occurring perennial herbs. *J. Ecol.*, 93, 681-692.
- 763 37.
- Jump, A.S. & Woodward, F.I. (2003). Seed production and population density decline
- approaching the range-edge of *Cirsium* species. *New Phytol.*, 160, 349-358.

766 38.

- 767 Kaye, T.N., Pendergrass, K.L., Finley, K. & Kauffman, J.B. (2001). The effect of fire on the
- population viability of an endangered prairie plant. *Ecol. Appl.*, 11, 1366-1380.
- 769 39.
- 770 Kaye, T.N. & Pyke, D.A. (2003). The effect of stochastic technique on estimates of population
- viability from transition matrix models. *Ecology*, 84, 1464-1476.
- 772 40.
- 773 Knight, T.M., Caswell, H. & Kalisz, S. (2009). Population growth rate of a common understory
- herb decreases non-linearly across a gradient of deer herbivory. For. Ecol. Manage., 257,
- 775 1095-1103.
- 776 41.
- ⁷⁷⁷ Lachish, S., McCallum, H. & Jones, M. (2009). Demography, disease and the devil: life-history
- changes in a disease-affected population of Tasmanian devils (*Sarcophilus harrisii*). J. Anim.
- 779 *Ecol.*, 78, 427-436.
- 780 42.
- 781 Lloret, F., Escudero, A., Iriondo, J.M., Martínez-Vilalta, J. & Valladares, F. (2012). Extreme
- climatic events and vegetation: the role of stabilizing processes. Glob. Change Biol., 18, 797-
- 783 805.
- 784 43.
- 785 Löfgren P., Eriksson O. & Lehtilä K. (2000). Population dynamics and the effect of
- disturbance in the monocarpic herb Carlina vulgaris (Asteraceae). Ann. Bot. Fenn., 37, 183-
- 787 192.
- 788 44.
- Martone, R.G. & Micheli, F. (2012). Geographic variation in demography of a temperate reef
 snail: importance of multiple life-history traits. *Mar. Ecol. Prog. Ser.*, 457, 85-99.

791 45.

- Maurer, B.A. & Brown, J.H. (1989). Distributional consequences of spatial variation in local
 demographic processes. *Ann. Zool. Fenn.*, 26, 121-131.
- 794 46.
- 795 McCauley, R.A. & Ungar, I.A. (2002). Demographic analysis of a disjunct population of
- 796 *Froelichia floridana* in the mid-Ohio River Valley. *Restor. Ecol.*, 10, 348-361.
- 797 47.
- 798 McGowan, C.P., Ryan, M.R., Runge, M.C., Millspaugh, J.J. & Cochrane, J.F. (2011). The role of
- 799 demographic compensation theory in incidental take assessments for endangered species.
- 800 Biol. Conserv., 144, 730-737.
- 801 48.
- Merow, C., Latimer, A.M., Wilson, A.M., McMahon, S.M., Rebelo, A.G. & Silander, J.A.
- 803 (2014). On using integral projection models to generate demographically driven predictions
- of species' distributions: development and validation using sparse data. *Ecography*, 37,
- 805 1167-1183.
- 806 49.
- Morris, W.F. & Doak, D.F. (2002). *Quantitative Conservation Biology: Theory and Practice of Population Viability Analysis*. Sinauer Associates, Sunderland, MA.
- 809 50.
- 810 Morris W.F., Pfister C.A., Tuljapurkar S., Haridas C.V., Boggs C.L., Boyce M.S. et al. (2008).
- 811 Longevity can buffer plant and animal populations against changing climatic variability.
- 812 *Ecology*, 89, 19-25.
- 813 51.

- 814 Münzbergová, Z. (2007). Population dynamics of diploid and hexaploid populations of a
- 815 perennial herb. Ann. Bot., 100, 1259-1270.
- 816 52.
- 817 Nuismer, S.L. & Kirkpatrick, M. (2003). Gene flow and the coevolution of parasite range.
- 818 *Evolution*, 57, 746-754.
- 819 53.
- 820 Pérez-Barrales, R., Simón-Porcar, V.I., Santos-Gally, R. & Arroyo, J. (2014). Phenotypic
- 821 integration in style dimorphic daffodils (Narcissus, Amaryllidaceae) with different
- 822 pollinators. Philos. T. Roy. Soc. B., 369, 20130258.
- 823 54.
- 824 Perón, G. (2013). Compensation and additivity of anthropogenic mortality: life-history
- effects and review of methods. J. Anim. Ecol., 82, 408-417.
- 826 55.
- 827 Pfister, C.A. (1998). Patterns of variance in stage-structured populations: Evolutionary
- predictions and ecological implications. *Proc. Natl. Acad. Sci. U. S. A.*, 95, 213-218.
- 829 56.
- 830 Piñero, D., Martínez-Ramos, M. & Sarukhán, J. (1984). A population model of Astrocaryum
- 831 *Mexicanum* and a sensitivity analysis of its finite rate of increase. J. Ecol., 72, 977-991.
- 832 57.
- Pironon, S., Villellas, J., Morris, W.F., Doak, D.F. & García, M.B. (2015). Do geographic,
- 834 climatic or historical ranges differentiate the performance of central versus peripheral
- 835 populations? *Global Ecol. Biogeogr.*, 24, 611-620.
- 836 58.

- Pulliam, H. R. (2000). On the relationship between niche and distribution. Ecol. Lett., 3, 349-
- 838 361.
- 839 59.
- 840 Reznick, D. (1992) Measuring the costs of reproduction. *Trends Ecol. Evol.*, **7**, 42-45.
- 841 60.
- 842 Roff, D.A. (1992). *The Evolution of Life Histories*. Chapman and Hall, London.
- 843 61.
- 844 Sagarin, R.D. & Gaines, S.D. (2002). The 'abundant centre' distribution: to what extent is it a
- biogeographical rule? *Ecol. Lett.*, 5, 137-147.
- 846 **62**.
- 847 Salguero-Gómez, R., Jones, O.R., Archer, C.R., Buckley, Y.M., Che-Castaldo, J., Caswell, H. et
- al. (2015). The COMPADRE Plant Matrix Database: an open online repository for plant
- 849 demography. J. Ecol., 103, 202-218.
- 850 63.
- Schurr, F.M., Pagel, J., Cabral, J.S., Groeneveld, J., Bykova, O., O'Hara, R.B., et al. (2012). How
- to understand species' niches and range dynamics: a demographic research agenda for
- biogeography. J. Biogeogr., 39, 2146-2162.
- 854 64.
- 855 Sexton, J.P., McIntyre, P.J., Angert, A.L. & Rice, K.J. (2009). Evolution and ecology of species
- range limits. Annu. Rev. Ecol. Evol. Syst., 40, 415-436.
- 857 65.
- Stearns, S.C. (1989). Trade-offs in life-history evolution. *Funct. Ecol.*, 3, 259-268.
- 859 66.
- 860 Stearns, S.C. (1992). The Evolution of Life Histories. Oxford University Press, Oxford, UK.

861 67.

- Sterling, B., Conley, W. & Conley, M.R. (1983). Simulations of demographic compensation in
 coyote populations. *J. Wildlife Manage.*, 47, 1177-1181.
- 864 68.
- Stokes, K.E., Bullock, J.M. & Watkinson, A.R. (2004). Population dynamics across a parapatric
- range boundary: *Ulex gallii* and *Ulex minor*. *J. Ecol.*, 92, 142-155.
- 867 69.
- Sultan, S.E. (2004). Promising directions in plant phenotypic plasticity. *Perspect. Plant Ecol. Evol. Syst.*, 6, 227-233.
- 870 70.
- Tuljapurkar, S. (1990). *Population dynamics in variable environments*. Springer, New York.
 71.
- Valladares F., Matesanz S., Guilhaumon F., Araújo M.B., Balaguer L., Benito-Garzón M. et al.
- 874 (2014). The effects of phenotypic plasticity and local adaptation on forecasts of species
- range shifts under climate change. *Ecol. Lett.*, 17, 1351-1364.
- 876 72.
- 877 Valverde, T. & Silvertown, J. (1998). Variation in the demography of a woodland understorey
- 878 herb (*Primula vulgaris*) along the forest regeneration cycle: projection matrix analysis. J.
- 879 *Ecol.*, 86, 545-562.
- 880 73.
- Van Tienderen, P.H. (1995). Life cycle trade-offs in matrix population models. *Ecology*, 76,
- 882 2482-2489.
- 883 74.

884	Villellas, J., Ehrlen, J., Olesen, J.M., Braza, R. & Garcia, M.B. (2013a). Plant performance in
885	central and northern peripheral populations of the widespread Plantago coronopus.
886	Ecography, 36, 136-145.
887	75.
888	Villellas, J., Morris, W.F. & García, M.B. (2013b). Variation in stochastic demography
889	between and within central and peripheral regions in a widespread short-lived herb.
890	Ecology, 94, 1378-1388.
891	76.
892	Weaver, J.L., Paquet, P.C. & Ruggiero, L.F. (1996). Resilience and conservation of large
893	carnivores in the Rocky Mountains. Conserv. Biol., 10, 964-976.
894	77.
895	Werner, P.A. & Caswell, H. (1977). Population growth rates and age versus stage-
896	distribution models for teasel (Dipsacus Sylvestris Huds.). Ecology, 58, 1103-1111.
897	78.
898	Wisdom, M.J., Mills, L.S. & Doak, D.F. (2000). Life stage simulation analysis: Estimating vital-
899	rate effects on population growth for conservation. <i>Ecology</i> , 81, 628-641.
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908 **<u>Tables</u>**

- 909 **Table 1** Characteristics of the studies used to assess demographic compensation. In Life-form, species were assigned to one of two categories:
- 910 S (short-lived) for annual and short-lived perennial plants, and L (long-lived) for other longer-lived life forms (we used the life-forms provided
- 911 by authors in the studies). We also provide information on habitat and location of the study (US = United States, UK = United Kingdom, CR =
- 912 Czech Republic), on the environmental/geographical factor (unknown or unclear for some studies), and on the number of populations ("N
- 913 pops"), years ("N Yrs"), vital rates ("N rates") and life stages ("N stages"). In number of rates, the parenthesis shows the rates finally used after
- 914 discarding less important/reliable ones (See *Calculation of vital rates* in Appendix S1).
- 915

Study	Species	Life-form	Habitat - Location	Environmental	Ν	Ν	Ν	Ν
				factor	Pops	Yrs	rates	stages
1 - Bruna (2003)	Heliconia	L (long-lived perennial	Amazonian forest; Brazil	Forest	13	2	31 (25)	6
	acuminata	herb)		fragmentation				
2 - Davelos & Jarosz (2004)	Castanea dentata	L (decidious tree)	Eastern deciduous forest;	Pathogen	6	4	30 (26)	8
			Michigan, US					

916

Study	Species	Life-form	Habitat - Location	Environmental	Ν	Ν	Ν	Ν
				factor	Pops	Yrs	rates	stages
3 - Davison et al. (2010)	Anthyllis	S (short-lived	Calcareous grassland;	Unknown	9	3	14 (14)	4
	vulneraria	perennial herb)	Belgium					
4 - Doak & Morris (2010)	Silene acaulis	L (long-lived perennial	Boreoalpine habitats; US,	Range position,	17	5	71 (53)	12†
		cushion)	Canada	climate				
5 - Doak & Morris (2010)	Polygonum	L (long-lived perennial	Boreoalpine habitats; US,	Range position,	16	5	189 (43)	13
	viviparum	herb)	Canada	climate				
6 - Eckhart et al. (2011)	Clarkia xantiana	S (winter annual herb)	Sandy soil slopes;	Range position	20	4	13 (13)	3†
			California, US					
7 - Ehrlén (1995)	Lathyrus vernus	L (long-lived perennial	Mixed forests; SE Sweden	Unclear	6	3	31 (31)	7†
		herb)						
8 - Endels et al. (2007)	Primula vulgaris	L (long-lived perennial	Ditch banks, rivulets; N	Disturbance	10	3	30 (26)	6
		herb)	Belgium					

Study	Species	Life-form	Habitat - Location	Environmental	Ν	Ν	Ν	Ν
				factor	Pops	Yrs	rates	stages
9 - Ghimire et al. (2008)	Nardostachys	L (long-lived	Rocky outcrop,	Harvesting	6	3	27 (24)	6
	grandiflora	perennial	meadows; Himalaya					
		herb)						
10 - Hunt (2001)	Atriplex vesicaria	L (long-lived	Paddocks; S Australia	Distance to water	9	1	16 (15)	4
		perennial shrub)						
11 - Hyatt & Araki (2006)	Polygonum	S (annual vine)	Roadsides, fields;	Unknown	6*	2	11 (11)	7†
	perfoliatum		Pennsylvania, US					
12 - Jacquemyn & Bris (2008)	Primula elatior	L (long-lived	Temperate forests;	Forest age	7	3	15 (15)	5
		perennial herb)	Belgium					
13 - Jiménez-Sierra et al. (2007)	Echinocactus	L (cactus)	Semiarid valley; Mexico	Unknown	6	2	19 (19)	7†
	platyacanthus							
14 - Jongejans et al. (2010)	Carlina vulgaris	S (short-lived	Grasslands; Sweden,	Unknown	8	2	16 (15)	4
		perennial herb)	Germany, CR					

Study	Species	Life-form	Habitat - Location	Environmental	Ν	Ν	Ν	Ν
				factor	Pops	Yrs	rates	stages
15 - Jongejans et al. (2010)	Tragopogon	S (short-lived	Grasslands; Holland, CR	Unknown	6	2	16 (15)	4
	pratensis	perennial herb)						
16 - Jongejans et al. (2010)	Hypochaeris	S (short-lived	Grasslands; Holland,	Unknown	6	2	44 (41)	6
	radicata	perennial herb)	Germany, CR					
17 - Kaye et al. (2001)	Lomatium	L (long-lived	Prairies; Oregon, US	Fire	6	5	32 (25)	6
	bradshawii	perennial herb)						
18 - Kaye & Pyke (2003)	Haplopappus	S (short-lived	Arid rangelands;	Grazing exclosure	10	8	13 (13)	4
	radiatus	perennial herb)	Oregon, US					
19 - Knight et al. (2009)	Trillium	L (long-lived	Decidious forests;	Herbivory	12	3	13 (13)	6
	grandiflorum	perennial herb)	Pennsylvania, US					
20 - Lofgren et al. (2000)	Carlina vulgaris	S (short-lived	Semi-natural pastures;	Unknown	8	2	6 (6)	3
		perennial herb)	Sweden					

Study	Species	Life-form	Habitat - Location	Environmental	Ν	Ν	Ν	Ν
				factor	Pops	Yrs	rates	stages
21 - McCauley & Ungar (2002)	Froelichia	S (annual herb)	Sandy soils; Ohio, US	Unclear	6*	2	11 (11)	5†
	floridana							
22 - Münzbergová (2007)	Aster amellus	L (long-lived	Open xerothermic	Ploidy, habitat	9	3	8 (8)	3
		perennial herb)	habitats; CR.	productivity				
23 - Piñero et al 1984	Astrocaryum	L (palm tree)	Tropical rain forests;	Density	6	1	36 (36)	14†
	mexicanum		Mexico					
24 - Valverde & Silvertown (1998)	Primula	L (long-lived	Decidious temperate	Canopy openness	8	2	20 (19)	5
	vulgaris	perennial herb)	forests; UK					
25 - Villellas et al. (2013)	Plantago	S (short-lived	Coastal habitats; coasts	Range position,	11	3	17 (17)	4
	coronopus	perennial herb)	in Atlantic Europe	climate				
26 - Werner & Caswell (1977)	Dipsacus	S (short-lived	Late-succession fields;	Unknown	8	4	20 (20)	7†
	sylvestris	perennial herb)	Michigan, US					

- 930 Table 1 (cont).
- 931
- 932 * Manipulated (McCauley & Ungar 2002) or invasive (Hyatt & Araki 2006) populations were discarded for this study.
- 933 + Studies including seed bank stages.

934 Figure legends

935

936	Fig. 1. Consequences of demographic compensation for geographical range size
937	(represented by red arrows). (A) In the absence of demographic compensation both vital
938	rates (lines of different colour) decline along the geographical gradient (e.g., latitude or
939	elevation) until the range limit at L_1 is reached (C), beyond which populations cannot
940	increase from low density. (B) With demographic compensation, vital rates change in
941	opposite directions (line thickness is proportional to the influence of each vital rate on the
942	population growth rate, i.e., to that rate's sensitivity). Improvements in one vital rate cause
943	the population growth rate to decline less steeply (D), resulting in range limit L_2 and a larger
944	range size than would be true in the absence of demographic compensation.
945	
946	Fig. 2. Percentage of correlations between vital rate contributions that are significantly
947	negative, and thus involved in demographic compensation. Vital rates are classified by (A)
948	type of demographic process or (B) life stage. Left panels: short-lived species; middle panels:
949	long-lived species; right panels: results for all species comparing grouped pairs of the same
950	vs. different process (A) or stage (B). Groups sharing lower case letters above the bars do
951	not differ significantly within each category, according to post-hoc analyses (see Appendix
952	S1). Abbreviations of 1) Demographic processes: Fec=fecundity; Rec=recruitment;
953	Sur=survival; Gro=growth; Shr=shrinkage; 2) Life stages: Early, Middle and Late; and 3) Vital
954	rate similarity: Diff=different type of process or stage; Same=same type of process or stage.
955	
956	Fig. 3. Sensitivity and variance of vital rates grouped by process. The magnitudes of the

957 sensitivities and variances differed greatly among vital rates and species, so we computed

the median (rather than the mean) across vital rates for each process and then the median
of those across species. Note the logarithmic y axis. Processes are recruitment, survival,
growth, shrinkage, and fecundity.

961

Fig. 4. Proposed relationships between sensitivity, variance, and LTRE contributions of vital
rates, and demographic compensation across multiple populations. Vital rates with high
influence on population growth rates (i.e., high sensitivity) and with high variance among
populations are expected to show a high contribution to the differences in population
growth rates, which in turn is expected to promote the participation of these rates in
demographic compensation. However, sensitivity might also have an opposite indirect effect
through its indirect negative correlation with variance (dashed arrow).

969

970 Fig. 5. Percentage of correlations between vital rate contributions that are significantly 971 negative, and thus involved in demographic compensation, depending on LTRE 972 contributions (C), variance among populations (V) and sensitivity (S). Vital rates were 973 grouped in two ways, depending on whether their C, V and S values were above or below 974 the median across all rates for a species: low vs. high contribution (left panel), and the four 975 possible combinations of high vs. low sensitivity and high vs. low variance (right panel). 976 Groups sharing a letter above the bars do not differ significantly within each category, 977 according to post-hoc analyses (see Appendix S1).

978

Fig. 6. Spatial correlations between mean annual vital rates for populations of the tundra
plant *Silene acaulis* across North America (Doak and Morris 2010). The corrgrams (sensu
Friendly 2002) shown are simply colour-coded correlation matrices; darker red indicates

increasingly negative correlations, and darker blue indicates increasingly positive
correlations. Each row and column represents a different vital rate, and as the correlation
matrix is symmetrical about the diagonal, only the upper right portion of the matrix is
shown. (A) Vital rates grouped by life stage (S = seeds and seedlings; numbers indicate
successively larger size classes for plants older than seedlings). (B) Vital rates grouped by
demographic process (R = rates governing recruitment of seedlings; rates are ordered by
size class within type of process). More pattern is revealed in B than in A.

989

990 Fig. 7. Three types of compensation may all be driven by (A) the same opposing trends in

991 different vital rates (red vs. blue) in response to environmental drivers (such as

temperature) in all populations: (B) spatial compensation across populations; (C) temporal

993 compensation within one population; and (D) compensation against environmental trends,

994 e.g., climate warming, within one population.

995

996 Fig. 8. If the responses of vital rates to environmental drivers (e.g., temperature) are not the same in two populations (in contrast with Fig. 7), we may not observe all types of 997 demographic compensation simultaneously. All combinations of presence/absence of 998 999 spatial and temporal compensation between two vital rates (red vs. blue) for two 1000 populations are shown here (A-D). Spatial compensation involves changes in opposite 1001 directions in the temporal mean vital rates between populations (note that with spatial 1002 compensation, the mean of the red rate is higher in the northern than in the southern 1003 population, but the opposite is true of the blue rate), while temporal compensation involves 1004 changes among years in opposite directions in the two rates within single populations. 1005 Vertical dashed lines show mean temperatures and two-headed arrows show the range of

1006 temperatures experienced by each population. For simplicity, we have assumed that both1007 rates change linearly with temperature.

Fig. 9. Two mechanisms that may produce an improvement in a vital rate (here, individual growth) in peripheral populations. Doak and Morris (2010) found that the growth rates of small individuals of two tundra plant species were higher in populations near the southern range limits, which compensated for deterioration in other vital rates. Higher plant growth could reflect (A) phenotypic plasticity (PP); (B) local adaptation (LA); or (C) a combination of the two processes. Gs, Gn and Ts, Tn correspond to individual growth rates and temperature in southern and northern populations, respectively. Gn* is the growth rate plants from the northern population would show if grown at the mean temperature for the southern population. Populations share genotypes in A), but not in B) and C), where each colour represents a different (locally adapted) set of genotypes. Demographic compensation would require a second vital rate to follow opposite patterns to those shown here.

1030 <u>Figures</u>
1031
1032 Fig. 1
1033



1042 Fig. 2





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1054 Fig. 3
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Fig. 4



1080 Fig. 5















Fig. 9

