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Research article

Facilitation stabilizes moisture-controlled alpine juniper shrublines in the central Tibetan Plateau

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23 **ABSTRACT**

24 The Tibetan Plateau hosts one of the world's highest undisturbed alpine juniper
25 shrublines. However, little is known about the dynamics of these shrublines in
26 response to climate warming and shrub-to-shrub interactions. Since growth of
27 shrubline junipers is limited more by moisture availability than by low temperatures,
28 we tested if upslope advancement of alpine juniper shrublines was constrained by
29 warmer temperatures and related recent droughts. We also evaluated whether
30 facilitation among neighboring shrubs, as inferred from spatial analyses, influenced
31 shrubline dynamics. Three rectangular plots crossing the *Juniperus pingii* var. *wilsonii*
32 shrubline were sampled at elevations from 4810 to 4917 m a.s.l. near the Nam Co
33 Lake, central Tibetan Plateau. Location of each stem and its diameter at the root collar
34 and age were measured. We reconstructed the spatial and temporal shrubline
35 dynamics during the past 350 years using standard dendrochronological methods.
36 Independent, long-term summer temperature reconstructions also were associated
37 with shrub recruitment. Point-pattern analyses were used to characterize spatial
38 patterns of different size classes of shrubs. The three shrublines showed little
39 long-term changes despite ongoing warming; no upward shift has occurred in the past
40 100 years. Recruitment was negatively associated with summer temperatures and
41 drought occurrence since the 1920s. Spatial patterns were characterized by clustering
42 at local scales and attraction between the different size classes, suggesting facilitation.
43 We conclude that moisture availability limits the recruitment and elevational advance
44 of junipers in this area of the Tibetan Plateau. Dynamics of alpine shrublines are more

45 contingent on positive interactions and local environmental factors than on regional
46 climatic variability.

47 *Keywords:*

48 Dendroecology; alpine shrubline; climatic warming; conspecific facilitation; drought;
49 Tibetan Plateau.

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53 **1. Introduction**

54 A growing body of evidence shows that global climatic warming has been altering
55 the composition, structure, and distribution of ecosystems worldwide (IPCC, 2014).
56 Among these structural changes, the northern or upward expansion of shrublines—the
57 highest latitude or uppermost altitude at which shrubs occur—into arctic tundra or
58 alpine grasslands is often regarded as fingerprint of global climate warming (Sturm et
59 al., 2001; Post et al., 2009). However, little is known about the range shift of alpine
60 shrublines in mid-latitude mountains subjected to predictable seasonal changes in
61 precipitation and moisture availability, even though alpine shrublands are a major
62 community in these treeless regions (Körner, 2003).

63 Climatic warming is expected to lead to shifts of latitudinal or altitudinal treelines
64 and shrublines (Kullman, 2002; Myers-Smith et al., 2011; Hofgaard et al., 2013)
65 because minimum air and soil temperatures set the frame for the processes at the
66 leading edge of woody plant communities (Fang et al., 2009; Lv and Zhang, 2011; Liu
67 and Yin, 2013). However, in areas subjected to seasonal water deficits, positive effects
68 of warming on local treelines or shrublines could be canceled out without concurrent
69 increases in precipitation that would alleviate heat-induced moisture stress (Daniels
70 and Veblen, 2004; Wahren et al., 2005). For example, where plant growth at alpine
71 treeline or shrubline is limited by moisture availability during the early growing
72 season (Liang et al., 2012, 2014) or the previous winter (Pellizzari et al., 2014), the
73 treelines or shrublines could be at risk of retreat if lack of water leads to growth
74 decline, constrains recruitment, or increases mortality rates.

75 Local site conditions (e.g., disturbance, topography, biotic interactions) also buffer,
76 nullify, or alter climatic impacts on treelines and shrublines (Callaway et al., 2002;
77 Kikvidze et al., 2005; Case and Duncan, 2014), resulting in heterogeneous and locally
78 contingent responses of nearby ecotones to shared climatic forcing (Harsch et al.,
79 2009). Among those local factors, plant-plant interactions increasingly are thought to
80 be major drivers of treeline and shrubline dynamics (Batllori et al., 2009; Grau et al.,
81 2012). However, surprisingly little information is available about whether such
82 interactions play a role in mediating alpine shrubline dynamics as the climate warms
83 (Dullinger et al., 2011).

84 One of the world's highest natural alpine shrublines occurs on the Tibetan Plateau
85 (Wu, 1983; Huang and Zhang, 2011). In these cold, dry, and treeless regions,
86 shrublands are the primary locus of nutrient cycling, water flow regulation, carbon
87 sequestration, and biodiversity (Huang and Zhang, 2011). However, little is known
88 about the impacts of climatic warming on the structure and distribution of alpine
89 shrublines. *Juniperus pingii* var. *wilsonii* (Rehder) Silba is a widespread alpine shrub
90 forming the upper shrubline across the central Tibetan Plateau at elevations over 4900
91 m. The extremely harsh climatic conditions experienced at the upper limit of shrubs
92 lead to a relatively narrow ecotone of less than 100-m between the shrubline and other
93 alpine vegetation devoid of taller shrub species.

94 Here we assess how regional climatic conditions and local-scale biotic interactions
95 (cf. Callaway et al., 2002) affect dynamics at three high-elevation *J. pingii* var.
96 *wilsonii* shrublines. Specifically, we: (1) reconstructed the age structure of individual

97 shrubs; (2) revealed alpine shrubline dynamics on multi-centennial time scales; and (3)
98 inferred underlying processes of shrubline dynamics from spatial point-pattern
99 analyses of different size classes of shrubs. Our earlier research had shown that the
100 radial growth of *J. pingii* var. *wilsonii* is limited by scarce precipitation rather than by
101 low temperature (Liang et al., 2012). Thus, we hypothesized that on the three
102 shrublines we studied that the upslope advancement of the shrubline would be
103 constrained by recent warming-related droughts. Further, because patchy patterns in
104 alpine shrublands across harsh environments can reflect intraspecific facilitation
105 (Choler et al., 2001; Armas and Pugnaire, 2005), we also hypothesized that positive
106 conspecific interactions would be reflected in clustered spatial patterns of individual
107 junipers.

108

109 **2. Materials and methods**

110 *2.1 Study area and climate*

111 The study area is located near the Nam Co Lake (30° 30' - 30° 55' N; 90° 16' - 91°
112 03' E; 4725 m a.s.l.) on the northern flank of the Gangdise-Nyainqintanglha
113 Mountains (Zhu et al., 2010). Climatically, this area is in the transition zone between
114 semi-arid and sub-humid conditions. Between 2006 and 2008, the annual average
115 temperature at 4730 m a.s.l. on the southeastern shore of the lake (AWS in Fig. 1a),
116 was 0.4°C. January (-8.4 °C) and July (9.5 °C) were the coldest and warmest months,
117 respectively (Zhang et al. 2011). The mean annual precipitation was 415 mm, ≈85%
118 of which fell between July and October (Zhang et al., 2011). However, the climate

119 was dry in May and June, when mean monthly precipitation was, respectively, < 15
120 and 35 mm, and mean monthly evaporation was > 130 and 160 mm (Zhang et al.,
121 2011). Average daily wind speed was low (3.6 m.s⁻¹; see also Zhang et al. 2011) and
122 southeasterly dry winds dominated during the growing season. At the meteorological
123 station of Baingoin (4700 m a.s.l.), about 100 km north-west of the Nam Co Lake,
124 monthly mean temperature, monthly total precipitation, and evaporation from 1957 to
125 2013 illustrated the dry and cold climates in central Tibet (Fig. 2). Temperature
126 reconstructions from different regions on the Tibetan Plateau and based on tree-ring or
127 ice-core proxies identified a warming trend in the area over the past 400 years
128 (Bräuning and Mantwill, 2004; Liu et al., 2005; Liu et al., 2009; Yang et al., 2009) but
129 no significant change in reconstructed annual precipitation over the same interval
130 (Yao et al., 2008).

131

132 2.2 Study species

133 *J. pingii* var. *wilsonii* (henceforth “juniper”) is the most widely distributed shrub
134 species on the Tibetan Plateau (Huang and Zhang, 2011). In the central Tibetan
135 Plateau and around the Nam Co Lake, junipers grow as prostrate, multi-stemmed
136 shrubs in cushion-forming patches of multiple individuals on south-facing slopes;
137 isolated individuals are rarely observed. Their maximum crown diameter is ≤ 3 m,
138 they rarely exceed 1.5 m in height (Chen and Yang, 2011), and the stem diameter at
139 the root collar is normally < 20 cm.

140 On the central Tibetan Plateau, junipers grow along an altitudinal gradient from

141 4,740 m to above 4,900 m a.s.l. The most abundant herb species located above the
142 studied juniper shrubline are: *Rhodiola fastigiata* (Hook. f. et Thoms.) S. H. Fu,
143 *Corydalis thyrsoiflora* Prain, *Oxytropis glacialis* Benth. ex Bunge, *Astragalus arnoldii*
144 Hemsl., *Euphorbia stracheyi* Boiss., *Stellera chamaejasme* Linn., *Heracleum*
145 *millefolium* Diels., *Androsace tapete* Maxim., *Phlomis younghusbandii* Mukerjee,
146 *Oreosolen wattii* Hook. f, *Morina kokonorica* Hao and *Carex oxyleuca* V. Krecz.
147 Human disturbance at the upper elevational limits of juniper usually is negligible
148 because they are remote and lack of edible grass for grazing yaks; cover of edible
149 grass species such as *Stipa purpurea* Griseb. accounts for < 5% cover at the shrubline
150 ecotone. We did not observe any sign of grazing by yaks while we were doing
151 fieldwork in 2013 or 2014, nor did we detect other evidence of human disturbances,
152 such as remains of charcoal, fire scars, browsing damage, or stumps in the study sites.

153 Annual growth rate of juniper is very slow; ring widths average 0.29 ± 0.15
154 mm/yr. Series-sectioning, i.e., comparison of ring-growth series from several sections
155 taken along one shoot of the same individual (methodological details in Kolishchuk,
156 1990; Wilmking et al., 2012), confirmed that the cambium along juniper stems
157 remains active. As a result, we were unlikely to miss outer rings in basal stem sections
158 (Liang et al., 2012). Despite juniper's slow growth, annual ring-width series from
159 different individuals can be cross-dated (Liang et al., 2012; see also the robust
160 cross-dating for other similar shrub species reviewed by Myers-Smith et al., 2015).
161 Radial growth of junipers is limited primarily by low moisture availability in
162 May-June during the year of ring formation (Liang et al., 2012).

163

164 *2.3 Field sampling*

165 In July and August, 2013, we sampled three rectangular plots (30 m × 120 m)
166 located near the north-eastern shore of Nam Co Lake (Fig. 1b). Two plots (SE1, SE2)
167 faced southeast (plot SE1: 30.89° N, 90.86° E, 4866 m a.s.l., slope 18°; plot SE2:
168 30.91° N, 90.80° E, 4917 m, slope 13°), whereas the third (SW1) faced southwest
169 (plot SW1: 30.89° N, 90.87° E, 4810 m, slope 20°). Each rectangular plot was located
170 in a topographically uniform part of the shrubline ecotone. The long side of the plot
171 paralleled the maximum slope and extended above the shrubline (see also Camarero
172 and Gutiérrez, 2004). The relative origin for each plot was situated in its lower left
173 corner. Elevations of lower and upper edges of the plots were measured using a GPS
174 calibrated with a barometric altimeter. Locations (as x, y coordinates ± 0.1 m) of the
175 main stem of each individual shrub within each plot were recorded. Two additional
176 variables for each shrub also were recorded: its maximum height and the diameter at
177 the root collar of the thickest stem of each individual. An electronic caliper was used
178 to measure the stem diameter (± 0.01 mm) and tapes were employed to measure shrub
179 height (± 10 cm). We did not find any dead shrubs within any of the plots.

180 The age structure of shrubs within the three study plots was obtained using
181 dendrochronological methods (Camarero and Gutiérrez, 2004), as have been used to
182 document shrubline changes in recent years (Myers-Smith et al., 2015). An increasing
183 number of treeline studies have explored recruitment dynamics using static age
184 structures, which reflect the balance between survival and mortality rate (Camarero

185 and Gutiérrez, 2004; Liang et al., 2011). Due to the dense wood and small size of
186 shrub stems, collecting discs, not increment cores, are used frequently in the Tibetan
187 Plateau and the Arctic to measure their growth rings of shrubs and relate shrub growth
188 to climate (see reviews in Myers-Smith et al., 2015). In fact, we broke several
189 increment borers during field investigation. However, because sampling stem discs
190 destroy the stem, we could not collect discs from every juniper in the plot. Rather, we
191 first randomly collected 22, 36, and 33 discs from individual junipers just outside of
192 plots SE1, SW1 and SE1, respectively. These samples were visually cross-dated using
193 characteristic rings, yielding inter-correlations for samples from sites SE1, SW1, and
194 SW2 of 0.62, 0.58, and 0.61, respectively.

195 Second, we established relationships between age and root-collar of junipers
196 from 80 randomly-sampled individuals in plot SE2. Wood discs were cut from the
197 thickest stem of each individual shrub, as closely as possible to the root collar; disc
198 diameters ranged from 2.0 to 18.3 cm, and ages ranged from 39 to 361 years. Based
199 on linear relationships between stem diameter at the root collar and juniper age (Fig.
200 3a), ages of all junipers measured in all three plots were estimated to the nearest
201 decade.

202

203 *2.4 Characterizing shrubline dynamics*

204 We observed that mature shrubs < 1 m tall may be > 250 years old (Liang et al.,
205 2012). Thus, it made little sense to define the location of the shrubline by shrub height
206 because height growth is very slow and unrelated to age (Fig. 3). Rather, changes in

207 the upper elevation of the shrubline in the three study plots were reconstructed for the
208 past 350 years (i.e., 1664 -2013) in 50-year intervals based on mapping junipers and
209 estimating their establishment dates to the nearest decade (from the relationship
210 shown in Fig. 3a). Establishment dates were then related to reconstructed summer
211 temperatures on the Tibetan Plateau (Thompson et al., 2006).

212

213 *2.5 Point pattern analyses*

214 Spatial statistics are efficient tools to analyze vegetation dynamics using space as
215 a surrogate to test hypotheses and to infer underlying (and usually unmeasured)
216 processes (Fortin and Dale, 2005; McIntire and Fajardo, 2009; Wang et al., 2010).
217 Point-pattern analyses were used to characterize spatial patterns of junipers growing
218 in the three study plots. We used the $O(r)$ statistic (Wiegand and Moloney, 2004;
219 Wang et al., 2010) to compare observed spatial patterns with those expected under a
220 null model of complete spatial randomness (CSR). Aggregated or hyperdispersed
221 (regular) distributions have values of $O(r)$ higher and lower than λ , respectively,
222 where λ is the value obtained for $O(r)$ for a CSR pattern of the same sample size. Note
223 that the $O(r)$ statistic is a scale-dependent probability density function related to a
224 neighbourhood density (Wiegand and Moloney, 2004). Considering that both the
225 mean radius of shrub patches and the distances between the nearest patches were no
226 more than 5 m, respectively, we calculated $O(r)$ at a spatial resolution = 1 m and for
227 spatial scales ranging from 1 to 10 m, which should identify relevant small-scale
228 patterns in the study plots. Last, because the plots are environmentally heterogeneous

229 (in, e.g., soil conditions and microtopography), we used the inhomogeneous version
230 of the $O(r)$ statistic for all point-pattern analyses. The Programita software was used
231 to perform all point pattern analyses (Wiegand and Moloney, 2004, 2014).

232

233 *2.5.1 Univariate point pattern analyses*

234 We used the univariate $O_{11}(r)$ statistic to examine spatial patterns of three
235 different size classes: size 1 (stem diameter at the root collar ≤ 3 cm); size 2 ($3 <$ stem
236 diameter ≤ 8 cm); and size 3 (stem diameter > 8 cm). These three size classes
237 correspond to different age classes, because shrub size was found to be related to
238 shrub age (Fig. 3a). Specifically, size classes 1, 2, and 3 corresponded to the following
239 age classes: age ≤ 80 yrs, $80 \text{ yrs} < \text{age} \leq 150$ years, and age > 150 years, respectively.
240 Such classification agreed well with the distribution of shrub ages at the three study
241 plots (Fig. 3b). We used a heterogeneous Poisson process as the null model to
242 simulate CSR. Edge correction was used to ensure that the number of points in an
243 incomplete circle was divided by the proportion of the area of the circle that lay
244 within the study plot (Wiegand and Moloney, 2004, 2014). Finally, if the calculated
245 $O_{11}(r)$ statistic was above or below the upper or lower 99% simulation envelopes
246 based on 999 Monte Carlo simulations of the original data then the pattern was
247 considered to be significantly aggregated or hyperdispersed (regular), respectively, at
248 the analyzed spatial scale.

249

250 *2.5.2 Bivariate point pattern analyses*

251 We used the $O_{12}(r)$ statistic to study spatial associations between pairs of the three
252 different size classes of shrubs. We assumed that larger shrub individuals (size 3 in
253 this study) could influence smaller individuals (size 1 and 2), but not *vice versa*.
254 Hence, we used an antecedent condition as the null model, which only randomized the
255 locations of small individuals (sizes 1 or 2), while keeping fixed the locations of larger
256 adults (size 3) or mid-size juveniles (size 2), respectively (Wiegand and Moloney,
257 2004, 2014). In this case, the statistic $O_{12}(r) = \lambda_2 g_{12}(r)$, for which $g_{12}(r)$ is the bivariate
258 mark-correlation function, gives the expected number of points of pattern 2 (seedlings
259 or mid-size juveniles) located at distance r from any point of pattern 1 (mid-size
260 juveniles or adults). $O_{12}(r) = \lambda_2$ corresponds to independent patterns, whereas $O_{12}(r) <$
261 λ_2 or $O_{12}(r) > \lambda_2$ corresponds to repulsion (negative association) and attraction
262 (positive association), respectively (Wiegand and Moloney, 2004, 2014). If the $O_{12}(r)$
263 statistic was above or below the 99% upper or lower simulation envelopes, the pattern
264 was considered to show significant positive or negative associations at those spatial
265 scales, respectively.

266

267 **3. Results**

268 *3.1 Size structure of junipers*

269 Stem diameter classes were right-skewed and unimodal in all plots (Fig. 4). The
270 maximum stem diameters at the root collar were 17.2, 20.2, and 25.5 cm in plots SE1,
271 SW1, and SE2, respectively. Most junipers were < 80 cm tall, but the maximum
272 height of individuals within the three plots ranged from 113 to 150 cm (Fig. 3c, d, e).

273

274 *3.2 Climate and juniper establishment*

275 The oldest junipers in plots SE1, SW1, and SE2, respectively, were estimated to
276 be 351, 409, and 512 years in (Fig. 3b), but there were only one or two individuals
277 older than 350 years in any of the plots. Reconstructed recruitment in each plot
278 appeared to be common before 1950, but rare thereafter (Fig. 5). The relationship
279 between reconstructed recruitment and reconstructed climate also differed before the
280 1920s, when recruitment and summer temperature were positively and significantly
281 correlated (SE1, $r = 0.59$; SW1, $r = 0.58$; SE2, $r = 0.65$; in all cases $n = 33$ and $P <$
282 0.001), and after the 1920s, when they were negatively, but not significantly,
283 correlated (SE1, $r = -0.36$; SW1, $r = -0.40$; SE2, $r = -0.39$; in all cases $n = 8$ and,
284 $P > 0.05$).

285

286 *3.3 Shrubline dynamics*

287 Overall, the juniper shrublines in plots SE1, SW1, and SE2 have shifted upwards
288 by 8, 4 and 9 m, respectively, in the past 350 years, but such summary values obscure
289 much of the variability. In plot SE1 (Fig. 6a), the juniper shrubline increased 6 m in
290 elevation during the 1714–1763 period relative to the previous 50 years. It then
291 remained unchanged from 1764 to 1863, but shifted upward again by 1 m between
292 1864 and 1913 and by another 1 m between 1914 and 1963. In the last 50-year period,
293 no change in shrubline elevation was observed in the data. In plot SW1 (Fig. 6b), the
294 juniper shrubline increased in elevation by 2.0 m (1714–1763), was unchanged

295 (1764–1813), advanced again by 2 m (1814–1863), and then remained unchanged
296 thereafter. Finally, in plot SE2 (Fig. 6c), the juniper shrubline increased in elevation
297 by 4.0 m (1714–1763), ascended another 5.0 m (1764–1813), and then stopped
298 advancing.

299

300 *3.4 Spatial patterns of shrub size classes*

301 Considering all size-classes together, junipers in all sampled plots were
302 significantly aggregated at spatial scales of 1–10 m (in plots SE1 and SE2) or 1–7 m
303 (plot SW1). Among the three size classes, aggregated distributions in plot SE1 were
304 detected at spatial scales of 1–10 m (size-classes 1 and 2) or 1–3 m (size-class 3) (Fig.
305 7a). In this same plot, the three analyzed pairings among the size classes (size-class 1
306 vs. size-class 2, size-class 1 vs. size-class 3, and size-class 2 vs. size-class 3) were
307 positively associated at spatial scales from 1 to 6 m. In plot SW1, size-classes 1 and 2
308 were significantly aggregated from 1 to 6 m, but size-class 3 was CSR (Fig. 7b).
309 Bivariate analysis revealed that all size classes were positively associated from 1 to 3
310 m in plot SW1. In plot SE2, all three size classes were significantly aggregated at
311 spatial scales of 1–8 m (size-class 1), 1–3 m (size-class 2) and 1–2 m (size-class 3)
312 (Fig. 7c). In contrast to the other two plots, however, negative spatial associations
313 were found for SE3 between size-classes 1 and 3 at 1–7 m, whereas other pairs
314 (size-class 1 vs. size-class 2, size-class 2 vs. size-class 3) were significantly positively
315 associated at spatial scales from 1 to 10 m. Overall, spatial associations were stronger
316 between the larger size classes (size 2 vs. size 3) than between the smallest and bigger

317 classes (size 1 vs. size 2 or size 1 vs. size 3).

318

319 **4. Discussion and conclusion**

320 Climatic warming is expected to lead to elevational shifts in treelines and
321 shrublines (Myers-Smith et al., 2011; Hofgaard et al., 2013), but our results illustrate
322 that juniper shrublines on the central Tibetan plateau have been unchanged for the last
323 50–200 years. On the central Tibetan Plateau, the climate is characterized by
324 elevated radiation levels and evaporation rates and low precipitation values (You et al.,
325 2010), hence limited moisture rather than low temperatures constrain growth of
326 junipers at shrubline (Liang et al., 2012). Climatic warming without concurrent
327 increases in precipitation will increase the evaporative demand and adversely affect
328 shrub growth and shrubline upward expansion. Thus, upslope advancement of the
329 Tibetan shrubline cannot occur under the warmer and drier conditions that have
330 affected the Tibetan Plateau over the last century (Thompson et al., 2006; Zhu et al.,
331 2011).

332 Shrubline dynamics also depend on the establishment of new individuals
333 (Myers-Smith et al., 2011), and our data illustrate that little or no new recruitment has
334 occurred in our study plots since the 1920s. Future warming (IPCC, 2014) could
335 reduce juniper recruitment even further. Static age structures within treeline ecotones
336 indicate tradeoffs between survival and mortality (Camarero and Gutiérrez, 2004;
337 Wang et al., 2006; Liang et al., 2011), and have also illustrated reduced recruitment at
338 treelines since the 1950s, coincident with warmer and drier climates, in northwest

339 China (Wang et al., 2006).

340 We suggest that age structures of juniper shrub plots also can provide insight into
341 recruitment conditions on the central Tibetan Plateau. We acknowledge that the
342 inferences on shrubline dynamics focused on recruitment and rates of advance would
343 be more robust if complemented with growth and mortality data because mortality of
344 shrubs can bias recruitment estimates through time. However, dead shrubs were not
345 found within our shrub plots, making it impossible to estimate mortality rates. Further
346 studies on shrub death and decomposition of dead woody material would help to
347 provide estimates of juniper mortality. Since published growth data revealed the
348 dominant role of moisture availability in juniper wood formation (Liang et al., 2012),
349 the growth data in this study do not promote a deeper understanding of shrubline
350 dynamics in this region.

351 The slow pace of juniper shrubline change is different from the rapid and
352 extensive shrub expansion observed in Arctic sites where radial growth of shrubs is
353 constrained primarily by low temperatures (Myers-Smith et al., 2011; Naito and
354 Cairns, 2011). However, slow expansion of shrubs also has been reported in the
355 relatively xeric environments of central and eastern Siberia, the interior Alaskan
356 tundra, and continental Mediterranean mountains that receive only seasonal moisture
357 (Wahren et al., 2005; García-Cervigón et al., 2012; Frost and Epstein, 2014). It seems
358 reasonable to suggest, therefore, that warming can exert either negative or positive
359 influences on shrub growth and expansion depending on simultaneous moisture
360 availability (Wahren et al., 2005).

361 In central Tibet, we hypothesize that recent warming has negatively influenced
362 shrubline dynamics by worsening drought stress, because no significant change in
363 precipitation was observed on the central Tibetan Plateau over the past four centuries
364 and severe droughts were frequent from the 1970s through the 1990s (Yao et al.,
365 2012). In fact, if the frequency of drought increases, the juniper shrubline is at
366 increasing risk of down-slope range shifts. Additional climatic changes also could
367 contribute to exacerbate drought stress: warmer winters could lead to a rapid melting
368 of the snowpack and a decrease in the snowpack thickness could induce drought stress
369 in early spring. Finally, other global drivers such as rising atmospheric concentrations
370 of CO₂ could increase drought tolerance or growth by increasing water-use efficiency
371 of plants (Ainsworth and Long, 2005). However, neither the growth trends (Liang et
372 al., 2012) nor the recruitment patterns we report here match with those expectations.

373 Climate plays an important role in determining the uppermost elevational limits
374 of arboreal and woody vegetation distribution at global, continental, and regional
375 scales (Körner, 2003), but specific dynamics and related patterns are contingent on
376 local scale processes (Callaway et al., 2002; Holtmeier, 2009). Mounting evidence has
377 shown that the magnitude of shrub expansion in cold biomes such as Arctic tundra is
378 highly reliant on microsite facilitation (e.g., small-scale geomorphic settings,
379 disturbance regimes, or positive plant-environment feedbacks; Tape et al., 2006;
380 Hallinger et al., 2010; Myers-Smith et al., 2011; Naito and Cairns, 2011; Frost and
381 Epstein, 2014; Hagedorn et al., 2014). The presence of neighboring conspecifics
382 exerts considerable control on further shrub establishment and encroachment

383 irrespective of the existence of additional constraining abiotic drivers (Stueve et al.,
384 2011). For example, local microclimatic warming resulting from shrub encroachment
385 is similar in magnitude to regional warming observed over the past century (He et al.,
386 2014). Spatial analysis of junipers in our plots revealed highly clustered distributions
387 and aggregation among different size classes at relatively small scales (<10 m),
388 consistent with the patchy patterns of shrub distribution observed in the landscape
389 around Nam Co Lake (Liang et al., 2012). We infer that such aggregated spatial
390 patterns are produced by conspecific facilitation as has been found in some other
391 treelines (Camarero et al., 2000). Clumping patterns also may stabilize shrubline
392 dynamics by retarding the advance of the ecotone and leading to shrubline stasis
393 (Harsch et al., 2009).

394 Positive interactions among woody plants in similar harsh and dry environments
395 result from enhanced establishment and growth and reduced evaporation that buffer
396 the extreme temperature ranges and improve soil fertility (Callaway et al., 2002;
397 Kikvidze et al., 2005). Thus, it is not surprising that conspecific facilitation at local
398 scales dominates in juniper shrublines. Positive interactions related either to structural
399 or to growth-form effects may allow shrub populations to persist for hundreds of years
400 by preserving specific regeneration-niche features, even under extremely adverse
401 environmental conditions (Holtmeier, 2009). Enhanced longevity linked to slow
402 growth rates and aggregated shrub patterns could act as stabilizing factors (Crawford,
403 2008). If juniper shrublands mainly respond to changes in water availability, climatic
404 warming could induce drought stress and lead to increased mortality of formerly

405 fast-growing individuals (Nobis and Schweingruber, 2013). In central Tibet, such
406 positive interactions could alleviate drought stress and create suitable regeneration
407 niches, leading to enhanced recruitment and upslope shrubline shifts when climatic
408 conditions are favorable. The relatively large amount of recruitment observed prior to
409 the 1920s matches this scenario and supports our hypothesis that local-scale
410 facilitation drives shrubline dynamics.

411

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579 **Figure legends**

580 **Figure 1.** Locations of sampling plots (SW1, SE1, and SE2 plots) and the automatic
581 weather station (AWS) around the Nam Co Lake in the central Tibetan Plateau (a).
582 The upper left inset shows the spatial location of the study area on the Tibetan Plateau.
583 Landscape view of the alpine shrubline (plot SW1) located at an elevation of 4795 m
584 a.s.l. (photograph taken by Y.F. Wang) (b).

585

586 **Figure 2.** Monthly mean air temperature, monthly total precipitation, and monthly
587 evaporation from 1957 to 2013 in Baingoin (4700 m a.s.l.), about 100 km north-west
588 of the Nam Co Lake.

589

590 **Figure 3.** The linear relationship found between age of *Juniperus pingii* var. *wilsonii*
591 shrubs and stem diameter measured at the root collar (a). These data correspond to
592 individuals sampled in plot SE2 (a). The right part in the first row (b) shows the age
593 distribution of shrubs in 50-year classes for the three study plots. The second and third
594 rows (c, d, e) show scatter plots between height and age of shrub individuals in the
595 three study plots (SE1, SW1, and SE2).

596

597 **Figure 4.** Distributions of stem diameters (2.5 cm size-class) at the root collar of
598 sampled *Juniperus pingii* var. *wilsonii* shrubs in the three study plots (SE1, SW1, and
599 SE2).

600

601 **Figure 5.** Decadal dynamics of *Juniperus pingii* var. *wilsonii* establishment estimated
602 for three shrublines (SE1, white bars; SW1, gray bars; SE2, dark bars) and
603 reconstructed summer temperature variability presented as *Z*-scores (see more details
604 in Thompson *et al.*, 2006). Positive and negative *Z*-scores correspond to warm and
605 cool summer conditions, respectively.

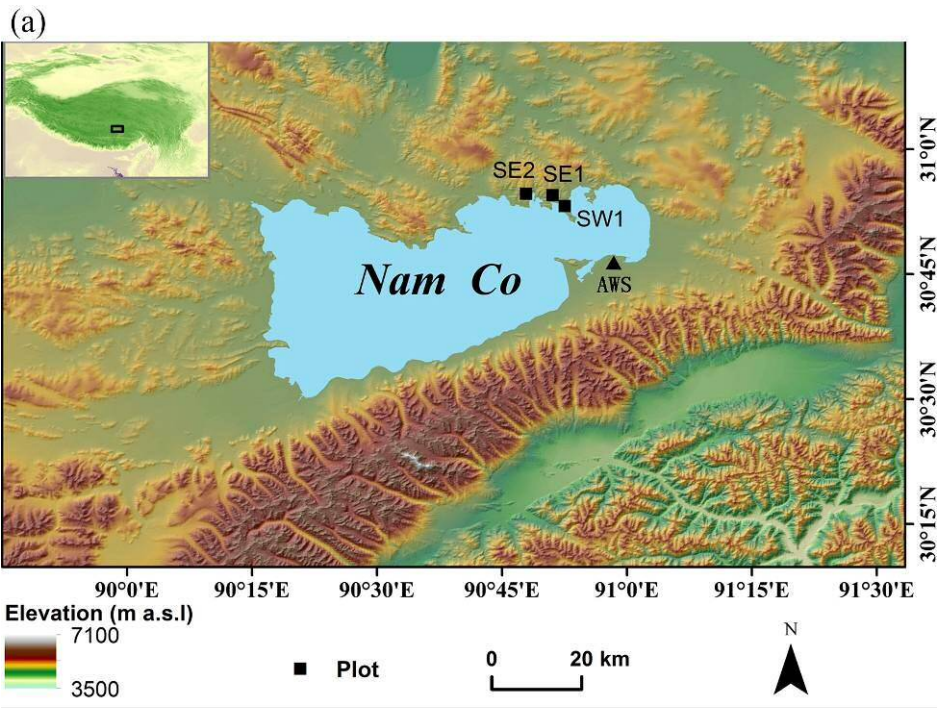
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607 **Figure 6.** Spatio-temporal dynamics of the *Juniperus pingii* var. *wilsonii* shrubline in
608 plots SE1 (a), SW1 (b), and SE2 (c). Each solid symbol corresponds to an individual
609 shrub that germinated during the period shown at the top and open symbols indicate
610 individuals established during the previous period. Different symbols represent
611 different establishment periods. Note that smaller symbols are used in the last two
612 periods, otherwise they are not visible.

613

614 **Figure 7.** Spatial point patterns and related univariate and bivariate point-pattern
615 analyses of shrubs located in the three study plots (SE1, SW1 and SE2). Three size
616 classes of *Juniperus pingii* var. *wilsonii* individuals (size-class 1, stem diameter at the
617 root collar ≤ 3 cm; size-class 2, stem diameter ≤ 8 cm; size-class 3, stem diameter > 8
618 cm) were used in the analysis. Figures in the first row show the spatial positions of the
619 three shrub size classes (note that the plot axes are not drawn at the same scale),
620 whereas figures in the second and third rows show the univariate ($O_{11}(r)$ statistic) and
621 bivariate ($O_{12}(r)$ statistic) point-pattern analyses, respectively. Lines with symbols

622 represent the $O_{11}(r)$ or $O_{12}(r)$ statistics, whereas thin lines correspond to the upper and
623 lower 99% bounds of the simulation envelopes.



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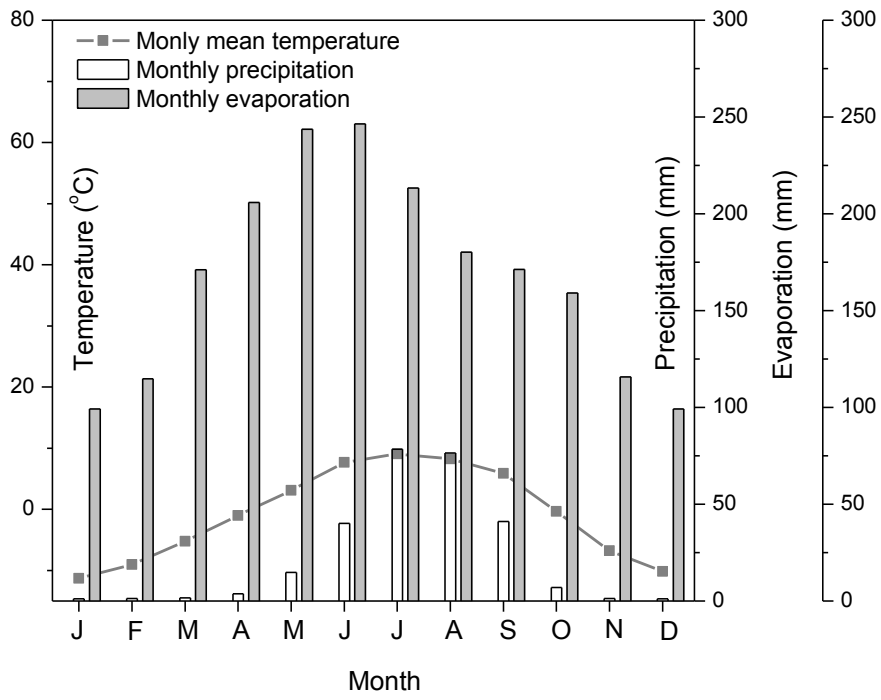
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628 **Figure 1**

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633 **Figure 2**

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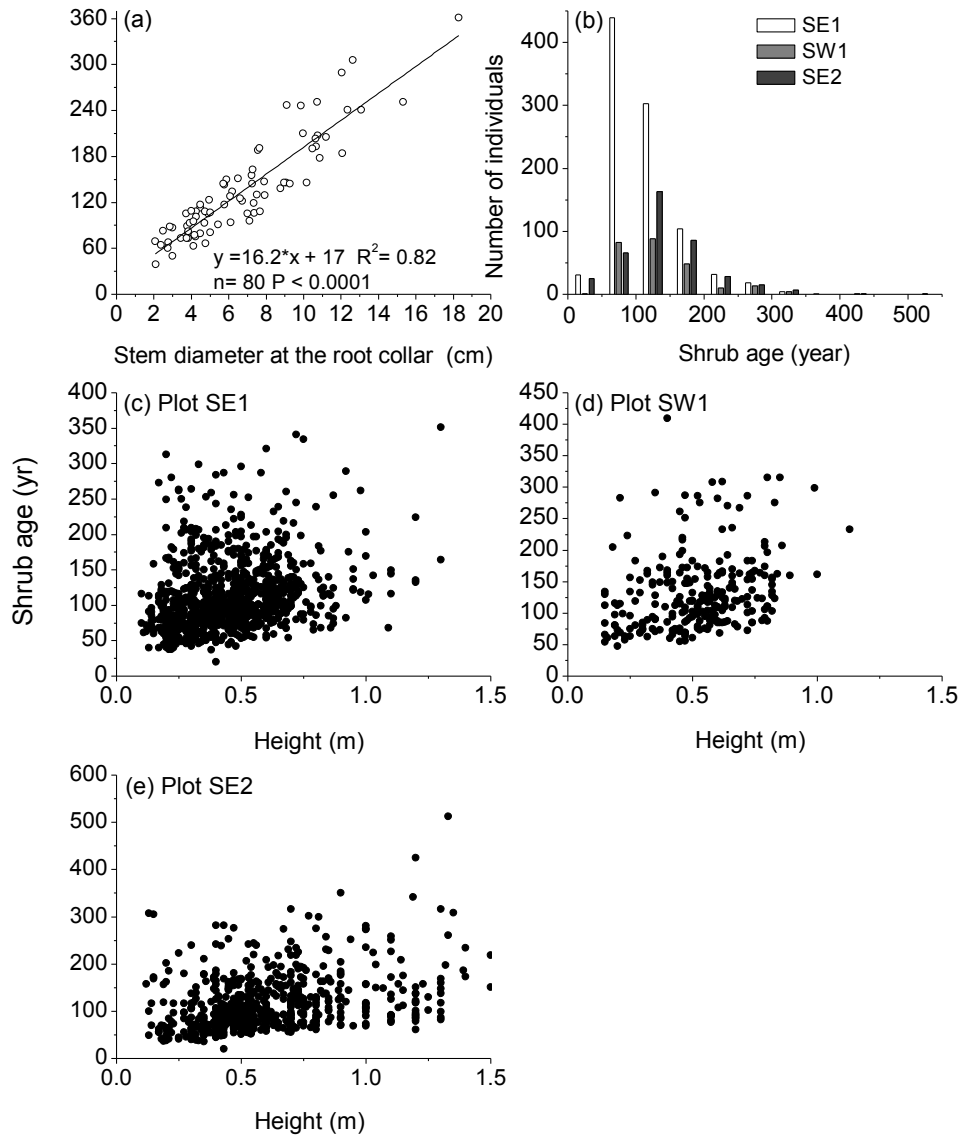
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646 **Figure 3**

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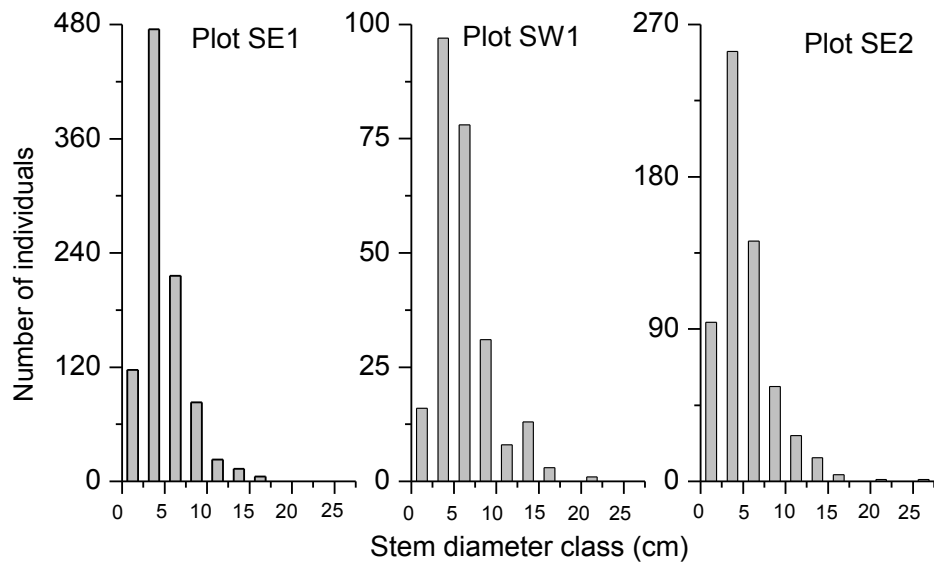
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655 **Figure 4**

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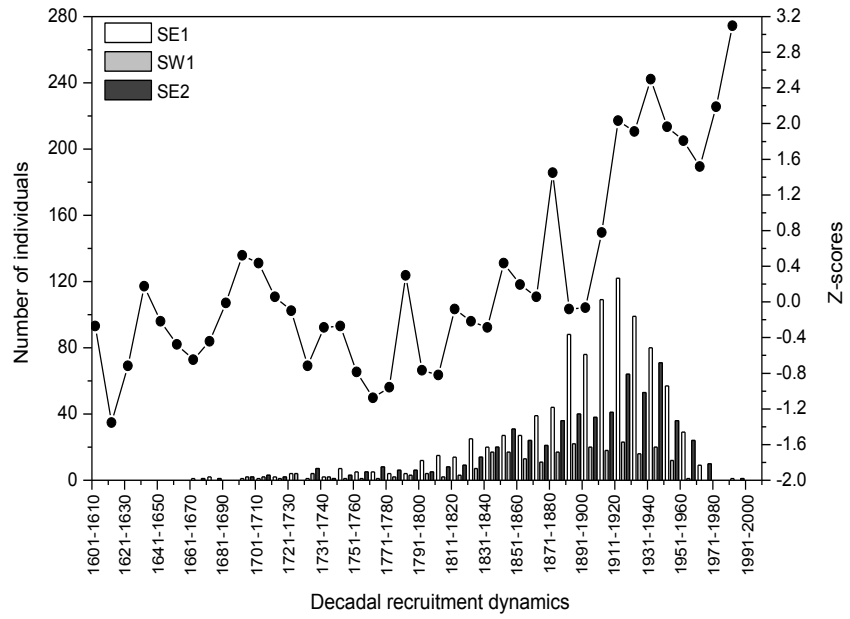
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671 **Figure 5**

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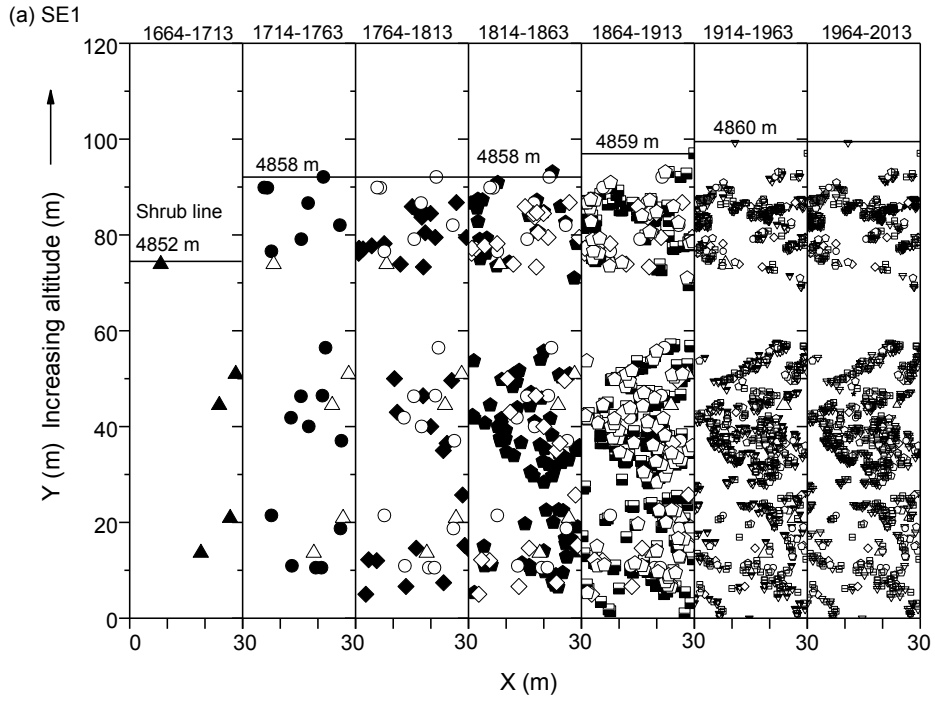
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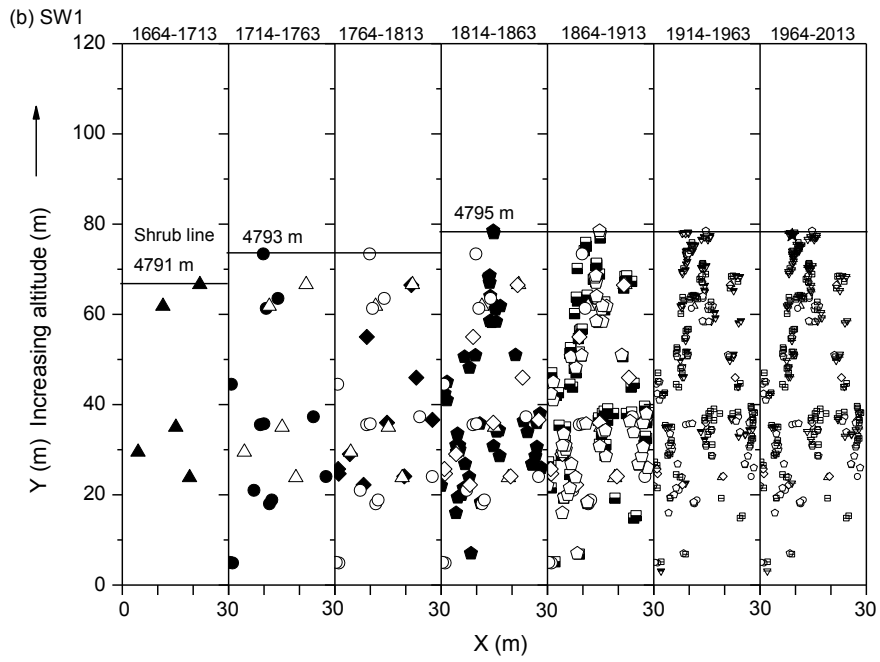
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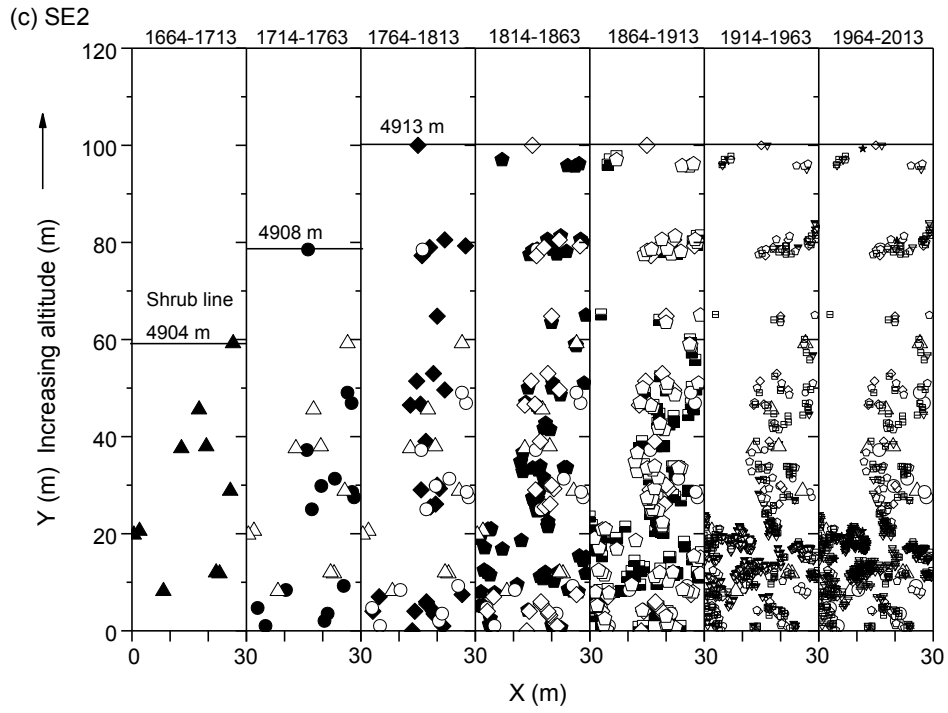


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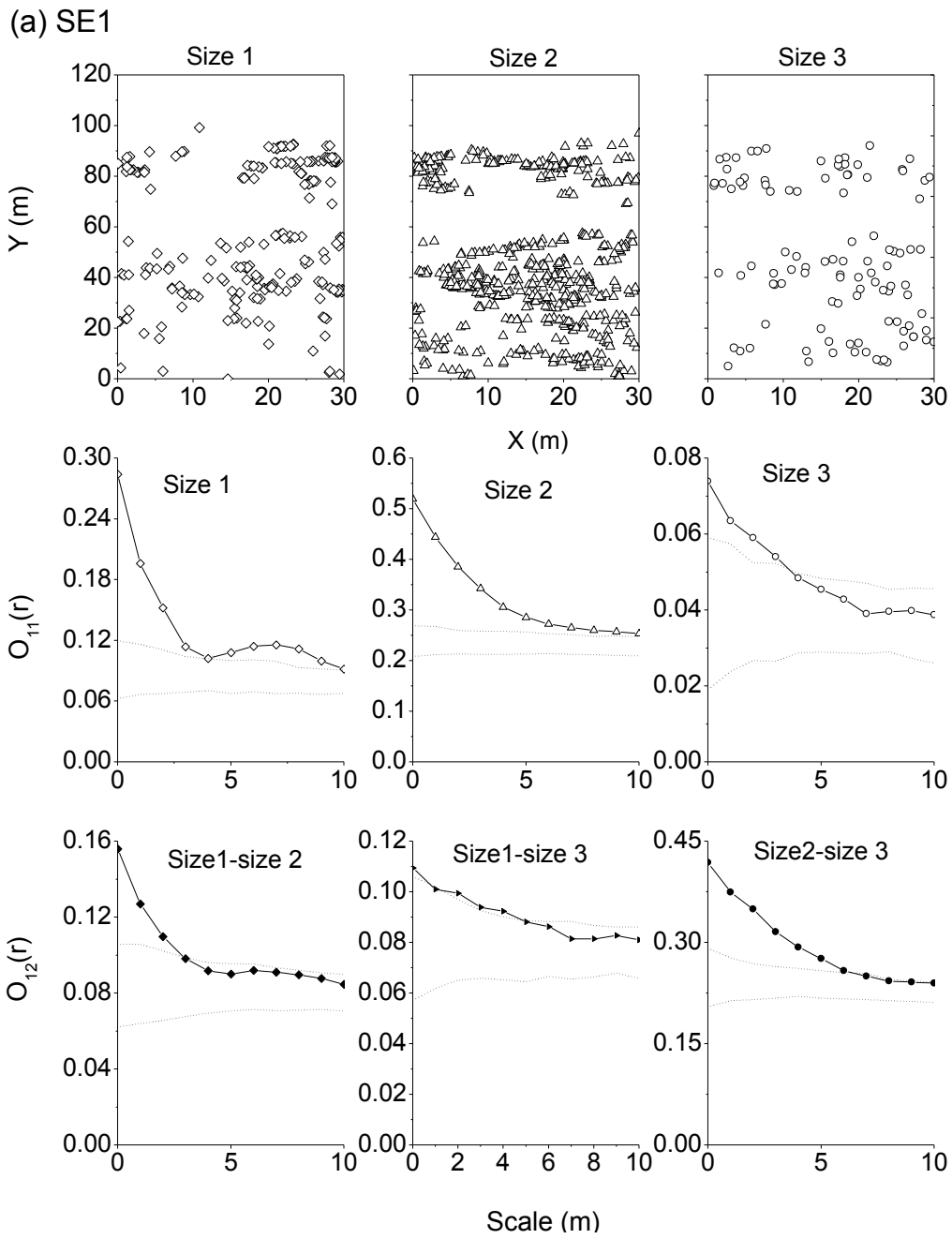


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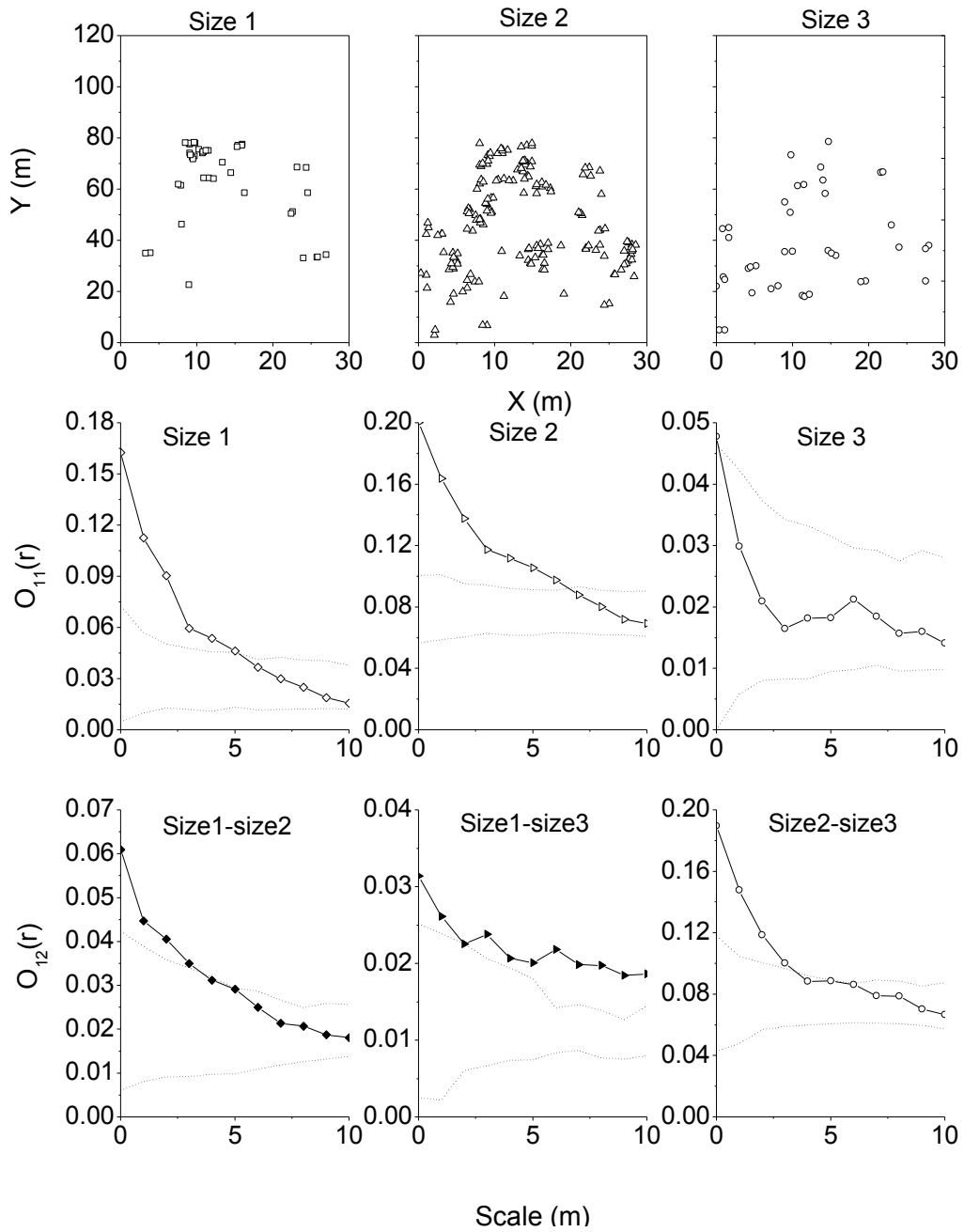
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685 **Figure 6**

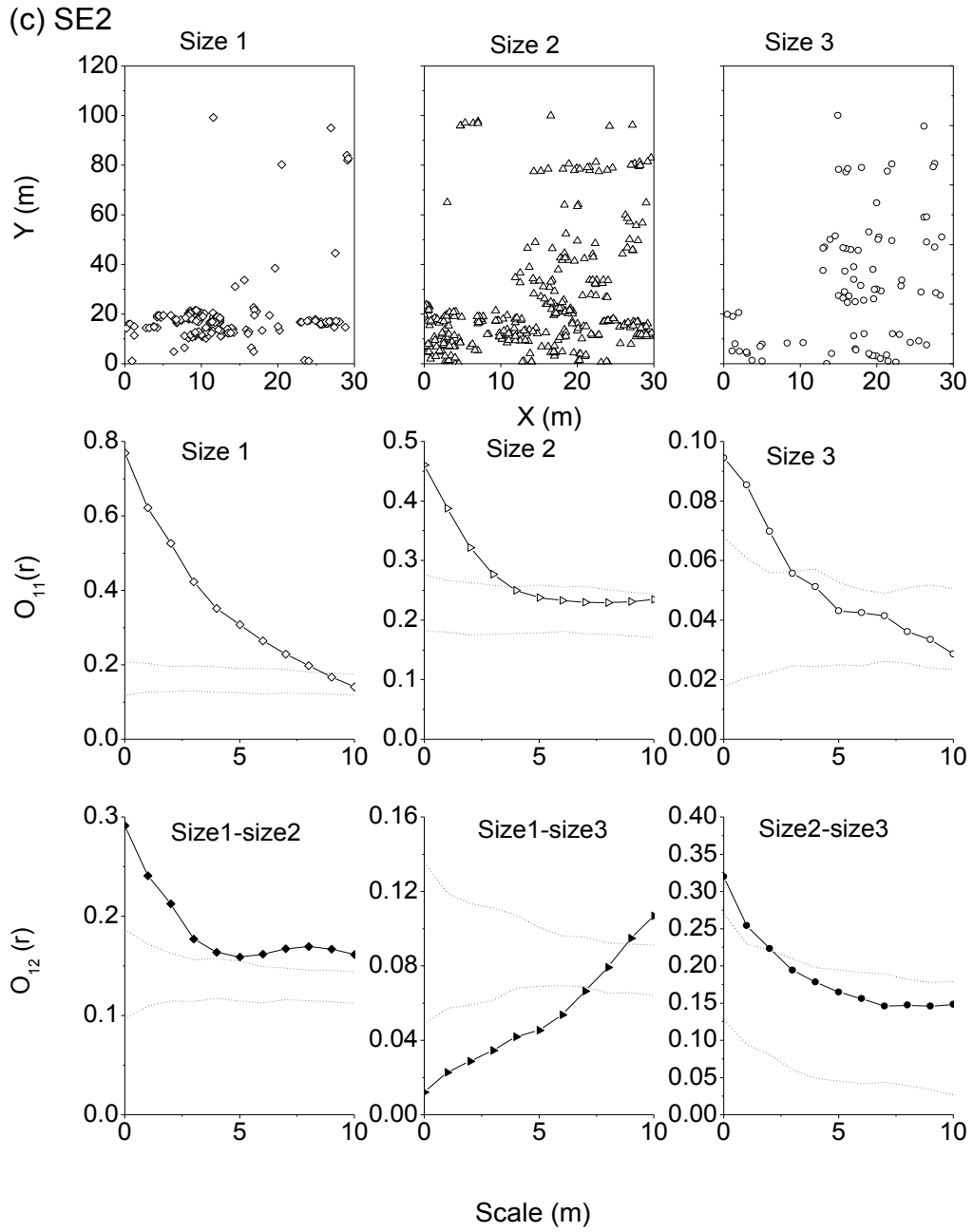


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(b) SW1



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689 **Figure 7**

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