

1 Effects of differently coated silver nanoparticles on
2 the photosynthesis of *Chlamydomonas reinhardtii*
3
4

5 Enrique Navarro^{†*}, Bettina Wagner[‡], Niksa Odzak[‡], Laura Sigg[‡] and Renata Behra[‡]

6 [†]. CSIC, Pyrenean Institute of Ecology, Av. Montañana 1005, Apdo. 13034, 50059 Zaragoza,
7 Spain.

8 [‡].Eawag, Swiss Federal Institute of Aquatic Science and Technology, Überlandstrasse 133, P.O.
9 Box 611, 8600 Dübendorf, Switzerland

10 KEYWORDS: *Chlamydomonas reinhardtii*, silver nanoparticles, dissolved silver, bioavailability,
11 coatings.

12 * Corresponding author

Name:	Enrique Navarro
Address:	CSIC, Pyrenean Institute of Ecology, Av. Montañana 1005, Apdo. 13034, 50059 Zaragoza, Spain.
Phone:	+34 976 369 393
Fax:	+34 976 719 019
E-mail address:	enrique.navarro@ipe.csic.es

13 ABSTRACT

14

15 Various factors have been invoked to explain the toxicity of silver nanoparticles (AgNP) to
16 microorganisms including particle size, the nature of stabilizing coatings as well as the amount of
17 dissolved silver occurring in AgNP suspensions. In this study we have assessed the effects of nine
18 differently coated AgNP (chitosan, lactate, polyvinyl pyrrolidone, polyethelenglycol, gelatin,
19 sodium-dodecyl-benzenesulfonate, citrate, dexpanthenol and carbonate) and AgNO₃ on the
20 photosynthesis of the freshwater algae *Chlamydomonas reinhardtii*. We have thus examined how
21 AgNP effects to algae relate to particle size, measured dissolved silver (Ag_d) and bioavailable
22 silver (Ag_{bioav}). Ag_{bioav} was indirectly estimated in toxicity experiments by cysteine-silver
23 complexation at the EC₅₀. The EC₅₀ calculated as a function of measured Ag_d concentrations
24 showed for some coatings similar values to that of dissolved Ag, while other coated AgNP
25 displayed lower EC₅₀. In all cases, excess cysteine completely prevented effects on photosynthetic
26 yield, confirming the role of Ag_d as a cause of the observed effect on the photosynthesis. Toxicity
27 was neither related to particle size nor to the coatings. For all differently coated AgNP
28 suspensions, the EC₅₀ values calculated as a function of Ag_{bioav} were comparable to the value of
29 AgNO₃. Depending on the coatings Ag_{bioav} was comparable or higher than measured Ag_d.

30

31 INTRODUCTION

32

33 Because of their bactericidal properties ^[1], silver nanoparticles (AgNP) are present in numerous
34 consumer products. During recent years, an increasing number of studies have demonstrated
35 toxicity of AgNP to different microorganisms such as bacteria ^[2-4] or algae ^[5-7]. The toxicity of
36 AgNP to microorganisms has been suggested to result from both the release of silver ions (Ag^+)
37 and interactions between AgNP and cell membrane ^[2, 8-10]. The determining role of Ag^+ in
38 explaining the observed toxicity of AgNP to microorganisms is experimentally supported by the
39 fact that complexation of Ag^+ by thiol ligands, as well as anaerobic conditions prevent toxicity of
40 AgNP ^[5, 7, 11-15]. Thus, risk assessments of AgNP need careful consideration of the contribution of
41 Ag^+ to toxicity. Ionic silver might occur in AgNP suspensions, both as residual from particle
42 synthesis and formed upon particle oxidation. Major factors expected to influence AgNP
43 dissolution in aqueous solutions include pH, ionic strength, and the presence of ligands ^[16, 17].

44 AgNP are typically synthesized with different surface coatings to minimize particle
45 agglomeration in aqueous systems ^[1]. Depending on their chemical nature, coatings might by
46 complexation influence the release of Ag^+ ions from particles ^[18-21], or affect the stability of
47 particles and indirectly, their toxicity to aquatic organisms ^[8, 22-24]. Moreover, intrinsic properties
48 of coatings might contribute to observed effects of AgNP ^[25]. So far, no studies have
49 systematically examined the contribution of size, dissolved silver and of several coatings on AgNP
50 toxicity in a common approach.

51 In this study we have assessed the toxicity of differently coated AgNP and of dissolved Ag
52 (provided as AgNO_3) on the freshwater algae *Chlamydomonas reinhardtii* and examined how
53 effects related to particle size, dissolved silver (Ag_d) and bioavailable silver (Ag_{bioav}) which is the
54 concentration that provokes toxicity. Considered coatings are representative of different chemical
55 families (Table 1) and were selected considering their differences in Ag binding properties and
56 molecular size.

57 We first assessed short-term effects on photosynthesis and measured Ag_d in experimental media
58 used for toxicity testing. The concentration of Ag_{bioav} was assessed indirectly in toxicity
59 experiments by modifying the speciation of Ag_d with cysteine. This use of the cysteine is
60 supported by recent studies showing that the uptake of Ag in presence of cysteine is strongly
61 reduced [15, 26]. Therefore effective concentrations on the base of dissolved and bioavailable silver
62 were compared to infer the influence of coatings on bioavailability and toxicity of AgNP.

63

64 **MATERIALS AND METHODS**

65

66 **AgNP characterization.** AgNP were provided by NanoSys GmbH (Wolfhalden, Switzerland)
67 as aqueous suspensions of 1 g/L total Ag (nominal concentrations). Original AgNP suspensions
68 (see Table 1 for bulk concentrations) were kept in the dark. Experimental concentrations were
69 prepared using 10 mM MOPS (3-morpholine propanesulfonic acid, buffered at pH 7.5). For
70 cysteine experiments, fresh stock solutions of cysteine were prepared in nanopure water, and kept
71 on ice to prevent oxidation. For exposure studies, $AgNO_3$ solutions and AgNP suspensions were
72 freshly prepared before use. Unless otherwise indicated, all chemicals were purchased from
73 Sigma-Aldrich. AgNP were characterized under experimental conditions (10 mM MOPS, pH 7.5)
74 for size and ζ –zeta– potential by Dynamic Light Scattering (DLS) using a Zeta Sizer (Nano ZS,
75 Malvern Instruments).

76 **ICP-MS measurements.** The silver concentration (isotope ^{107}Ag) was measured in acidified
77 solutions (0.1 M HNO_3) by ICP-MS (Element 2 High Resolution Sector Field ICP-MS; Thermo
78 Finnigan). Reliability of the measurements was controlled using a water reference solution
79 (Standard Reference Material 1643e, National Institute of Standards & Technology, USA).

80 **For total Ag measurements** AgNP suspensions were digested with HNO_3 in a microwave oven
81 (MLS 1200 Mega, Microwave Laboratory Systems). Recovery of Ag was 88 – 95 % of the
82 nominal concentrations.

83 **Dissolved silver** present in AgNP suspensions was examined by centrifugal ultrafiltration
84 (Millipore Amicon Ultra-4 3K) through a membrane with a nominal molecular weight limit of 3
85 kDa. Recovery measurements using AgNO₃ and the same experimental media were done, showing
86 that between 90-99% of the silver was recovered after the Ultrafiltration process. Suspensions
87 (5μM AgNP in 0.01 M MOPS at pH 7.5) were centrifuged for 30 min at 3000×g (Megafuge 1.0R,
88 Heraeus Instruments). Concentration of dissolved Ag (Ag_d mostly Ag⁺, see at Supporting
89 Information -SI- Table S1) in the filtrate was expressed as % of the total Ag concentration after 1
90 or 2 h in the experimental media without algae (Table 1). Measurement of Ag_d in the AgNP
91 suspensions were confirmed using diffusive gradients in thin films (DGT) [5, 27], and by
92 ultracentrifugation. Bioavailable Ag (Ag_{bioav}) was on the basis of cysteine-silver complexation
93 experiments (see below).

94 **Speciation calculations.** Calculations of the Ag_d speciation upon cysteine addition to AgNO₃ or
95 AgNP were done using the software ChemEQL 3.0 [28] using the corresponding equilibrium
96 constants, as described in Navarro et al. 2008 [5]. Detailed calculations and the list of stability
97 constants and appropriate references are available in SI Tables S1 and S2 respectively.

98 **Algal culture and photosynthetic yield measurements.** *Chlamydomonas reinhardtii* (wild
99 type, CC-125 137c mt+) were cultured under controlled conditions (25 °C, 90 rpm, 120 μE) in
100 standard growth medium at pH 7.5 [29]. Cell number was measured by an electronic particle
101 counter (aperture 50 μm; Z2 Coulter Counter; Beckman Coulter, Fullerton, CA). While growth is
102 a more sensitive endpoint towards silver with EC₅₀ values of 12 nM [30], photosynthetic yield was
103 chosen as a toxicity endpoint because effects are measurable upon short-term exposure thus
104 avoiding accumulation of secreted biomolecules. The algal photosynthetic yield of the
105 photosystem II in light was measured by fluorometry using a PHYTO-PAM (Heinz Walz GmbH)
106 equipped with an Optical Unit ED-101US/MP. The yield reflects the efficiency of the
107 photochemical energy conversion process [31]. Fluorometry measurements were done at similar
108 light intensities than those used during growth and exposure (140 μmol photons m⁻² s⁻¹ PAR
109 radiation).

110 **Effects of AgNO₃ and AgNP on photosynthetic yield.** Exponentially growing algae in culture
111 media were first centrifuged (2000×g, 10 min) and then resuspended in 10 mM MOPS at pH 7.5
112 to a final volume of 25 mL in glass Erlenmeyers and a density of 1.7×10⁶ cells mL⁻¹.
113 Photosynthetic yield was not affected by the algal translocation into MOPS or by exposure to
114 cysteine ^[5]. Toxicity of AgNP and AgNO₃ to photosynthesis was assessed by concentration-
115 response experiments by exposing algae to increasing concentrations of AgNO₃ and AgNP (see SI
116 Table S3 for a comprehensive list of the experiments and concentrations used). At least 3
117 concentration-response experiments per coating were done. AgNP and AgNO₃ toxicity to the algal
118 photosynthetic yield was measured after short-term exposure (1 and 2 h) in order to minimize
119 accumulation of algal products in the exposure media, and thus changes in the silver speciation.
120 The photosynthetic values of all concentration-response experiments were represented as
121 percentage of the respective controls. Values were plotted as a function of the total Ag or Ag_d
122 concentration measured in AgNP suspensions, in the absence of algae in the suitable concentration
123 range. The Ag_d bound to the algae was estimated to be negligible in comparison with the EC₅₀
124 concentrations, as calculated based on previous studies (ref. 26), see calculations in SI. The Ag_d
125 values were determined by centrifugal ultracentrifugation. Data were fitted to a four parameters
126 logistic curve to determine EC₅₀ values as described below in the Statistics section.

127 **Cysteine-silver complexation experiments.** Complexation experiments were carried out in
128 order to determine the concentration of bioavailable silver (Ag_{bioav}). Based on our previous results
129 ^[5], cysteine concentrations similar to or higher than Ag_d (mostly Ag⁺), were expected to complex
130 Ag⁺ and to decrease the free Ag⁺ concentrations to levels which show no effect to photosynthesis
131 (see detailed speciation calculation at SI Table S1 and S2). Different cysteine concentrations were
132 first added to one concentration of the various AgNP which were in the range of the respective
133 EC₅₀. After ten minutes pre-equilibration (time to reach equilibrium of cysteine-silver complexes)
134 algae were added to the suspensions and their photosynthetic yield measured after 1 h of exposure.
135 Up to 47 complexation experiments were done with various AgNP and cysteine concentrations (SI
136 Table S3). The photosynthetic values were represented as percentage of the respective controls

137 and plotted as a function of cysteine concentration. Points were fitted to the best curve
138 (maximizing R^2 value), using Sigma Plot 12 (© Systat Software Inc.). The models were then used
139 to calculate the lowest amount of cysteine required to completely prevent the toxicity to
140 photosynthesis. For each AgNP, the concentration of bioavailable Ag_{bioav} at EC_{50} (1h) was
141 estimated as the product of the EC_{50} values (based on Ag_d) by the cysteine/ Ag_d ratio ($Ag_{bioav} = Ag_d$
142 $EC_{50} \times Cys/Ag_d$). This approach is detailed in the discussion.

143 The potential effects of cysteine on the dissolution of Ag or the characteristics of AgNP were
144 assessed by measuring dissolved Ag_d , as well as size and ζ - potential of AgNP in the presence of
145 500 nM cysteine which was the highest concentration used in the silver complexation
146 experiments.

147 **Statistics.** Errors of measurements (total number of experiments and replicates are shown in SI
148 Table S3) are expressed as standard deviation (SD). Concentration-response curves were fitted to
149 a four parameters logistic curve using R and drc package ^[32], to obtain the corresponding EC_{50}
150 values and standard error (SE). The same package was also used to perform comparison tests,
151 using R “compPAR” function; the null hypothesis is that the ratio equals 1. The ratio was obtained
152 by dividing EC_{50} value for $AgNO_3$ by EC_{50} for each AgNP. If the ratio significantly differs from 1,
153 the null hypothesis should be rejected, meaning that values are significantly different ($p < 0.05$).
154 The p-values are adjusted using Bonferroni correction for multiple T-tests.

155

156 **RESULTS**

157

158 **AgNP characterization.** In the simple experimental medium used for toxicity experiments (10
159 mM MOPS, pH 7.5), the AgNP displayed average diameters ranging from 17 nm (CIT) to 456 nm
160 (DEX), and ζ - potential from -3.5 mV (LAC) to -36 mV (CO_3) (Table 1). Dissolved silver (Ag_d),
161 measured 1 h after dilution of the original AgNP suspensions in 10 mM MOPS and as percentage
162 of measured total Ag, was between 12 and 19 % in the case of PEG, LAC, DEX and PVP and

163 around 1-3 % in the case of CO₃, CHI, CIT, GEL and NDB (according to both ultrafiltration and
164 DGT). These percentages were similar after 2 h (Table 1).

165 **Effects of dissolved Ag and AgNP on photosynthetic yield.** All differently coated AgNP and
166 dissolved Ag proved to be toxic, decreasing the photosynthetic yield of algae as a function of
167 increasing total Ag concentrations (Figure 1). At the highest applied concentrations (between 20
168 and 50 μM as total silver) most AgNP completely inhibited photosynthesis within 1 h, while in the
169 case of GEL, CIT and CHI coated AgNP, photosynthetic yield remained between 10-30 % of the
170 control. Based on total Ag, the various AgNP strongly differed in their toxicity (Table 2A), with
171 EC₅₀ values ranging from 0.28 (DEX) to 5.25 μM (CIT). Dissolved AgNO₃ showed the highest
172 toxicity to *C. reinhardtii* photosynthetic yield (Figure 1 and Table 2A), with an EC₅₀ of 0.18 μM,
173 up to 30 times lower than that of CIT coated AgNP.

174 Examination of effects of the AgNP on photosynthetic yield as a function of measured Ag_d,
175 showed concentration-response curves and EC₅₀ values to differ less than as a function of total Ag
176 (Figure 2 and Table 2B). Compared to AgNO₃, the EC₅₀ values expressed as function of Ag_d were
177 slightly lower or similar for the CHI, LAC, PEG and PVP coated AgNP, but appeared to be
178 significantly lower for the CO₃, DEX, NDB, CIT and GEL coated AgNP (Table 2B).

179 When expressed as function of measured total Ag, EC₅₀ values were found to decrease with
180 increasing particle size (Figure 3A). No correlations between the EC₅₀ of the various AgNP and
181 the corresponding average particle size have been found, whether expressed as total Ag or as Ag_d
182 (Figure 3).

183 **Silver-cysteine complexation experiments.** After 1 h exposure at the highest applied
184 concentration of cysteine (500 nM), the ζ- potential of most AgNP shifted to less negative values
185 (SI Figure S2), while in the case of polymeric coatings (PEG, PVP and CHI), no effect was
186 observed. The size of most AgNP was not affected by cysteine, and only a few showed a slight
187 increase of the averaged size (SI Figure S2). Cysteine had only small effects on the Ag_d
188 determined in most of the AgNP suspensions (SI Figure S3), with only the GEL AgNP showing a
189 significantly higher Ag_d concentration (4 times).

190 Pre-equilibration of particle suspensions with cysteine resulted for all AgNP in a lowering of
191 toxicity with increasing cysteine concentration (see examples with CO₃, PVP and CHI AgNP in
192 Figure 4). At the highest concentration of 500 nM, cysteine suppressed toxicity of all AgNP
193 completely, indicating absence of toxicity by the coatings.

194 Curves were modeled to calculate the minimal cysteine concentration that prevented toxicity to
195 the photosynthetic yield and which is supposed to correspond to the Ag_{bioav} concentration (Table
196 3, for an example of calculation see SI Figure S1). For most AgNP, the amount of cysteine
197 required to prevent toxicity was between 0.3-0.5 μM and was lower in case of NDB or CIT (0.07
198 or 0.13 μM, respectively). Using the ratio of cysteine concentrations to the measured Ag_d
199 concentrations, toxicity values were corrected to obtain EC₅₀ values as a function of Ag_{bioav} (Table
200 3). Compared to dissolved Ag, most AgNP displayed similar or slightly higher EC₅₀ values, while
201 DEX and NDB coated AgNP appeared to be more toxic.

202

203 DISCUSSION

204

205 **AgNP toxicity as a function of total silver and of dissolved Ag_d.** All AgNP were toxic to the
206 photosynthetic yield of algae, with calculated EC₅₀ values expressed as measured total Ag varying
207 in the range 0.28 – 5.25 μM (Table 2A and Figure 1). These EC₅₀ were all higher than the EC₅₀ of
208 dissolved Ag and thus suggested that various AgNP were less toxic than a similar amount of
209 dissolved Ag. The range of the EC₅₀ expressed as the measured dissolved Ag_d was narrower
210 (0.029 – 0.19 μM, Table 2B and Figure 2), as expected considering the role of dissolved silver in
211 AgNP toxicity to algae^[5]. The size of the nanoparticles did not account for differences in toxicity
212 (see Figure 3B), in contrast to the expected higher toxicity of smaller silver nanoparticles found in
213 other studies^[33, 34].

214 While for some of the AgNP the EC₅₀ based on measured Ag_d became similar to that of AgNO₃,
215 for other the Ag_d based EC₅₀ were lower, indicating for these AgNP an apparently higher toxicity
216 compared to AgNO₃. Factors to be considered for this apparent higher toxicity include intrinsic

217 toxicity of the coatings, direct nanoparticle effects, increased dissolution of NP in contact with the
218 algae and underestimation of Ag_d .

219 Some of the coating compounds, in particular glycols (PEG), CHI, or NDB are toxic for algae
220 [35-37] and bacteria [38] only at high concentrations with EC_{50} ranging from 120 to 1900 mg/L.
221 However, considering the complete suppression of toxicity by cysteine, the various coatings were
222 not affecting photosynthetic yield at the applied experimental concentrations. The suppression of
223 toxicity by cysteine also indicated that direct effects of AgNP were unlikely. Another possibility to
224 explain the apparent higher toxicity of some AgNP is that the concentration of Ag^+ ions did
225 increase during the toxicity experiments upon contact of the particles with algae. As already
226 hypothesized in a previous study that suggested a higher Ag_d based toxicity of carbonate coated
227 AgNP compared to $AgNO_3$, the production and secretion of reactive oxygen species by algae
228 might lead to increased dissolution [6, 39]; as an example, hydrogen peroxide is very efficient in
229 dissolving AgNP [40]. Interestingly, a recent study showed that a close contact of AgNP with
230 bacteria surfaces is required for bactericidal effects, also suggesting the release of Ag^+ from AgNP
231 in close vicinity to the bacteria [13]. Last but not least, underestimation of Ag_d would also lead to
232 the higher apparent toxicity of certain AgNP. The levels of Ag_d in the particle suspensions were
233 measured in the absence of algae to avoid underestimation due to the efficient uptake of Ag^+ ions
234 by algae [15, 26]. In the presence of algae, because of complexation of Ag^+ ions by coating
235 molecules (e.g. in the case of proteins or retention of dissolved Ag^+ ions within a polymeric
236 coating), Ag_{bioav} might be higher than measured Ag_d . However, the Ag^+ bound in polymeric
237 species may be exchangeable with added cysteine, as well as with algal biotic ligands. If the
238 AgNP are in contact with the algae cells, free Ag^+ is rapidly taken up and may then be supplied by
239 dissociation from the complexes with the coating molecules.

240 **Indirect estimation of bioavailable Ag.** The titration experiments with cysteine were used to
241 estimate the concentration of Ag_{bioav} for comparison with measured Ag_d . Cysteine as a very strong
242 ligand for Ag^+ is expected to form complexes with free Ag^+ present at the beginning of the
243 exposure and with Ag^+ adsorbed or complexed to the coatings by competition with ligands such as

244 citrate or functional groups of peptides in gelatin. If Ag^+ is released from AgNP in close vicinity
245 to the algae, these Ag^+ ions are also likely to be rapidly complexed by cysteine (Cys). Formation
246 of AgCys and $\text{Ag}(\text{Cys})_2$ complexes is expected, thus with a ratio of Cys/Ag in the range 1 – 2,
247 depending on the concentrations of Cys and Ag^+ [41]. It was assumed that the concentration of
248 cysteine that suppressed toxicity corresponded to the bioavailable Ag concentration within this
249 ratio. Interactions of cysteine with the nanoparticles surfaces appeared to be of minor importance,
250 as shown by only slight changes of ζ - potential and nanoparticle size in the presence of cysteine
251 (SI Figures S2 and S3).

252 The calculated cysteine concentrations suppressing toxicity, the ratios of cysteine to measured
253 Ag_d and the calculated bioavailable Ag at EC_{50} are presented in Table 3. In the case of AgNO_3 , the
254 Cys/ Ag_d ratio of 1.1 confirmed the speciation calculations for these conditions (SI Table S1 and
255 S2). For the AgNP coated with CHI, LAC, PVP, DEX and NDB, the Cys/ Ag_d ratios were in the
256 range 1.0 to 2.6, which can be considered as similar to the ratio for AgNO_3 . For these coatings,
257 measured Ag_d concentrations under abiotic conditions were in reasonable agreement with the
258 Ag_{bioav} concentrations at which algae were exposed. In the case of other coatings (CO_3 , CIT, GEL
259 and PEG), the higher ratios indicated that Ag_{bioav} was higher than the measured Ag_d . Based on the
260 estimated Ag_{bioav} concentrations (Table 3) the EC_{50} were in a narrow range for most coatings,
261 namely 0.22 - 0.58 μM , and thus close to the EC_{50} for AgNO_3 , but were lower in the case of NDB
262 and DEX coated AgNP.

263 The processes leading to high Cys/ Ag_d ratios need to be considered in detail. The high ratio
264 Cys/ Ag_d (8.2) in the case of GEL indicated that measured Ag_d was likely underestimated, due to
265 complexation of Ag^+ to gelatin proteins, in particular to methionine or cysteine groups [42]. Ag
266 bound to larger proteins would not pass through the ultrafiltration membranes with a cutoff of 3
267 kDa. Ag^+ complexed to the GEL-coating molecules may then be released by cysteine addition and
268 may also be exchanged with the biotic ligands [18] of algae. The excess of cysteine required to
269 prevent toxicity suggests that there was competition between cysteine and proteins as Ag_d ligands.
270 The calculated Ag_{bioav} at EC_{50} was within the range of other coatings. The fact that GEL coated

271 AgNP were the only ones showing a noticeable increase of measured Ag_d upon cysteine
272 interactions (4 fold increase, SI Figure S3), supports all the previous points. GEL coated AgNP
273 did not provoke a complete photosynthetic inhibition, even at higher concentrations, also likely
274 due to the effect of complexation of Ag^+ by proteins (Figure 2).

275 The CO_3 coated AgNP presented the highest Cys/ Ag_d ratio (14.7), and the lowest EC_{50} value as
276 Ag_d . Carbonates are not good silver ligands, so no Ag_d was expected to be complexed to the AgNP
277 coating. These results, together with a strong decrease of EC_{50} from 1 to 2 h (SI Table 5S), suggest
278 some Ag dissolution from AgNP as a result of biological interactions. A tentative explanation, for
279 a better dissolution in this case, is that dissolution of AgNP by ROS produced by the algae is more
280 efficient in the case of CO_3 , which is the coating least adequate to scavenge ROS, in comparison
281 to the organic compounds. Another factor may be the small size of the CO_3 ions, which may
282 facilitate a close contact and subsequent surface interactions between AgNP and algal cells and
283 their by-products ^[13]. Furthermore, specific interactions of the carbonate coating with carbonic
284 anhydrases are also possible. However, the calculated EC_{50} based on Ag_{bioav} (Table 3) was also
285 within the range of the other coatings.

286 In the case of citrate, the Cys/ Ag_d ratio was 3.7 and thus also indicated that complexation of Ag^+
287 by citrate and possibly dissolution occurred. Citrate may complex Ag^+ and thus retains the ions at
288 the AgNP surfaces [18, 19, 23](#), and possibly leads to renewed reduction of dissolved Ag^+ to Ag^0 ^[43, 44].
289 The incomplete inhibition of photosynthesis at high concentrations of CIT coated AgNP may also
290 be due to decreased Ag^+ concentration in the presence of citrate (Figure 2).

291 In the case of PEG, the ratio Cys/ Ag_d (3.1) also indicated some additional interactions, although
292 PEG is not a good ligand for Ag^+ . The Ag^+ can be retained by this polymeric coating close to the
293 nanoparticle surface and be released in the presence of cysteine, as well as by interactions with
294 algae.

295 In the case of NDB and DEX, lower EC_{50} based on Ag_d , as well as based on the Cys/ Ag_d ratio
296 indicate that measured Ag_d was similar to Ag_{bioav} , but that increased toxicity occurred. These two
297 molecules are not likely to strongly complex Ag^+ . However, both NDB as a surfactant and DEX as

298 a solvent may increase membrane permeability, thus facilitating Ag^+ uptake or even leading to the
299 uptake of AgNP [45, 46].

300 **Role of dissolved Ag^+ and of AgNP coatings.** These results thus confirm the predominant role
301 of dissolved Ag for toxicity of AgNP, as also demonstrated in other studies with algae [5, 26], as
302 well as with nematodes [23] and bacteria [12]. Furthermore, Ag uptake in *C. reinhardtii* from AgNP
303 has been shown to be dependent on dissolved Ag, and no evidence of substantial AgNP uptake
304 was found [15, 26]. The role of AgNP coatings in toxicity to algae appears thus to be only minor in
305 acting as an intermediate reservoir of Ag^+ which may subsequently be released, and in modulating
306 interactions of AgNP with algae, which may lead to further release of Ag^+ .

307 **ACKNOWLEDGMENTS**

308 This work was supported by the Swiss Federal Institute of Aquatic Science and Technology
309 (Eawag) and by the Spanish Ministry of Economy and Competitiveness (National Research Plan,
310 reference BFU2010-22053) and by the funding program La Caixa-Government of Aragon
311 (GALC-002/2011). Authors also thank Dr. Andreas Gondikas from the University of Vienna, for
312 his helpful comments on silver-cysteine binding dynamics. The corresponding author also
313 received a research stay grant from Programa Europa XXI from Caja Inmaculada.

314 **SUPPORTING INFORMATION AVAILABLE**

315 This paper includes supporting data as: speciation calculations and the corresponding constants,
316 a complete list of the experiments, tables containing size and ζ - potential measurements for all
317 nanoparticles, the ratio between EC_{50} (calculated as a total Ag) after 1 and 2 hours of exposure. An
318 examples of the calculation of the cysteine required to abolish the toxicity of AgNPs and of the
319 Ag_d adhered by the algae cells, and three figures showing the influence of Cysteine on ζ - potential
320 and size of differently coated AgNPs and in the Ag dissolution from the different AgNPs are also
321 provided. All this information is available free of charge via the Internet at <http://pubs.acs.org>.

322 **FIGURE LEGENDS**

323

324 **Figure 1.** Concentration-response curves of photosynthetic yield after 1 h of exposure to
325 differently coated AgNP and to AgNO₃ as a function of measured total Ag. The dashed line
326 indicates 50% of the control value. Photosynthetic values are expressed as the percentage of the
327 control.

328 **Figure 2.** Concentration-response curves of photosynthetic yield after 1 h of exposure to
329 differently coated AgNP, as a function of measured dissolved Ag_d. The dashed line indicates 50%
330 of the control value. Photosynthetic values are expressed as the percentage of the control.

331 **Figure 3.** EC₅₀ values expressed as measured total Ag (A) and as measured Ag_d (B) as a function
332 of particle size.

333 **Figure 4.** Effect of different concentrations of cysteine on the effect of three differently coated
334 AgNP (CO₃, PVP, CHI; concentrations are indicated in the legend) on photosynthetic yield.
335 Photosynthetic values are expressed as the percentage of the control.

336

337

SUPPORTING INFORMATION

Effects of differently coated silver nanoparticles on the photosynthesis of *Chlamydomonas reinhardtii*

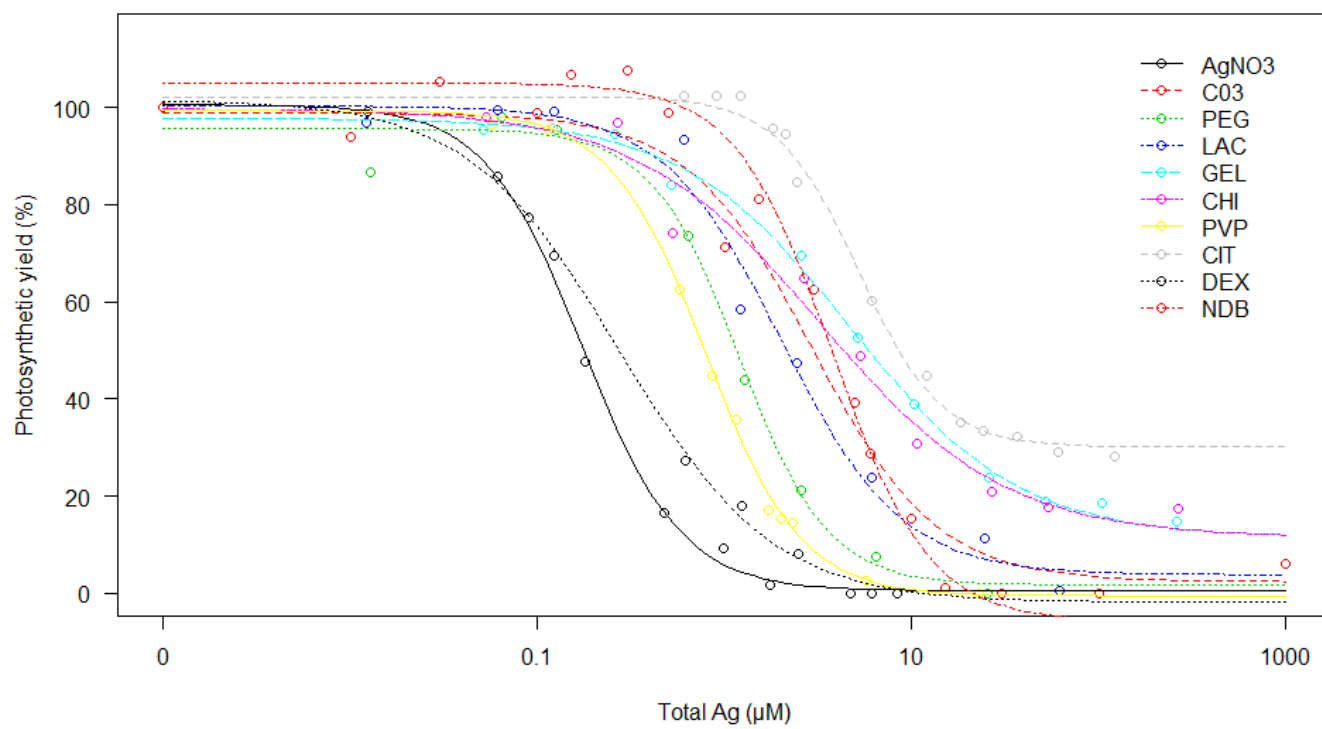
Enrique Navarro[†], Bettina Wagner[‡], Niksa Odzak[‡], Laura Sigg[‡] and Renata Behra[‡]

[†]. CSIC, Pyrenean Institute of Ecology, Av. Montañana 1005, Apdo. 13034, 50059 Zaragoza, Spain.

[‡].Eawag, Swiss Federal Institute of Aquatic Science and Technology, Überlandstrasse 133, P.O. Box 611, 8600 Dübendorf, Switzerland

TABLE OF CONTENTS

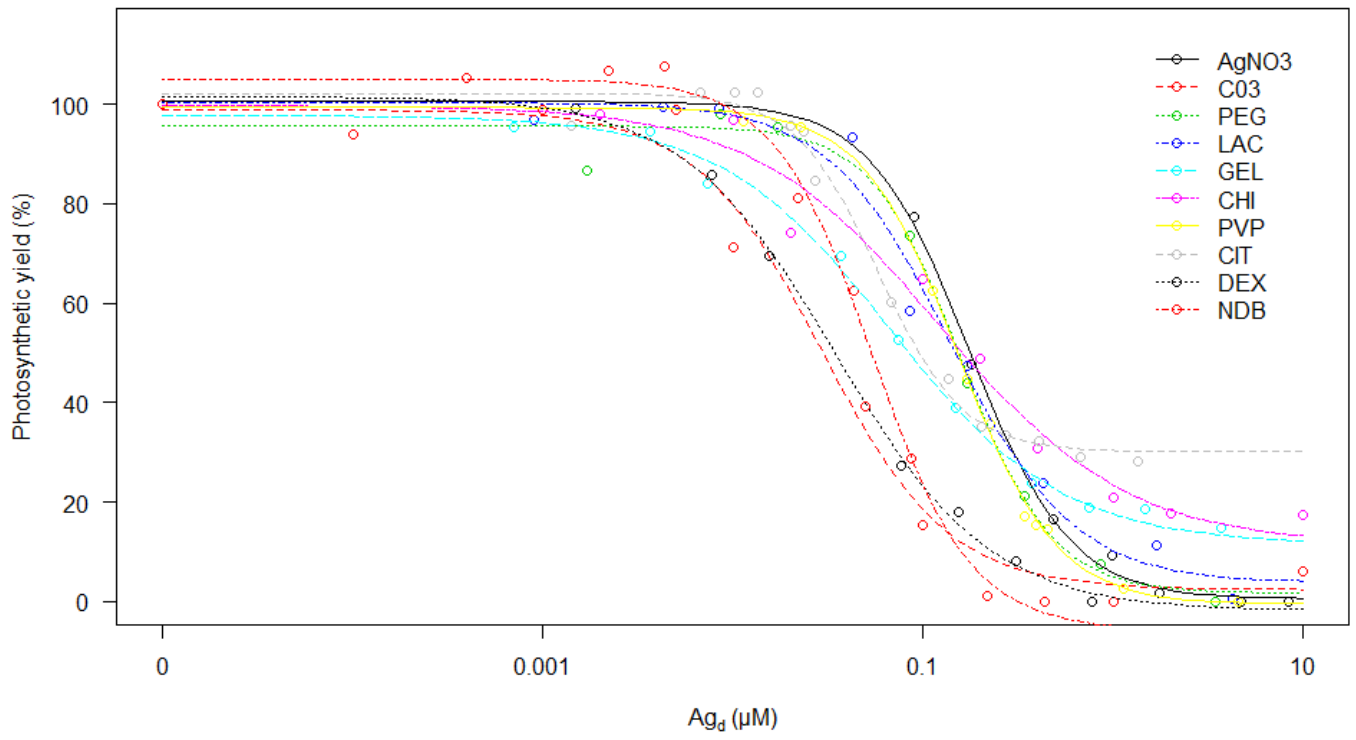
Table S1. Ag⁺ speciation calculations examples using ChemEQL 3.0.
Table S2. Stability constants used to calculate the Ag⁺ concentrations at equilibrium with cysteine.
Table S3. List of experiments, detailing AgNPs (nominal) and cysteine concentrations used.
Table S4. Size (including PDI) and Zeta potentials measurements.
Table S5. Ratio between EC₅₀ (calculated as a total Ag) after 1 and 2 hours of exposure.
Figure S1. Example of the calculation of the cysteine required to abolish the toxicity of AgNPs.
Figure S2. Influence of Cysteine on ζ- potential and size of differently coated AgNPs.
Figure S3. Influence of cysteine on Ag dissolution from the different AgNPs
Calculation S1. Ag_d adhered by the algae



338

Figure 1.

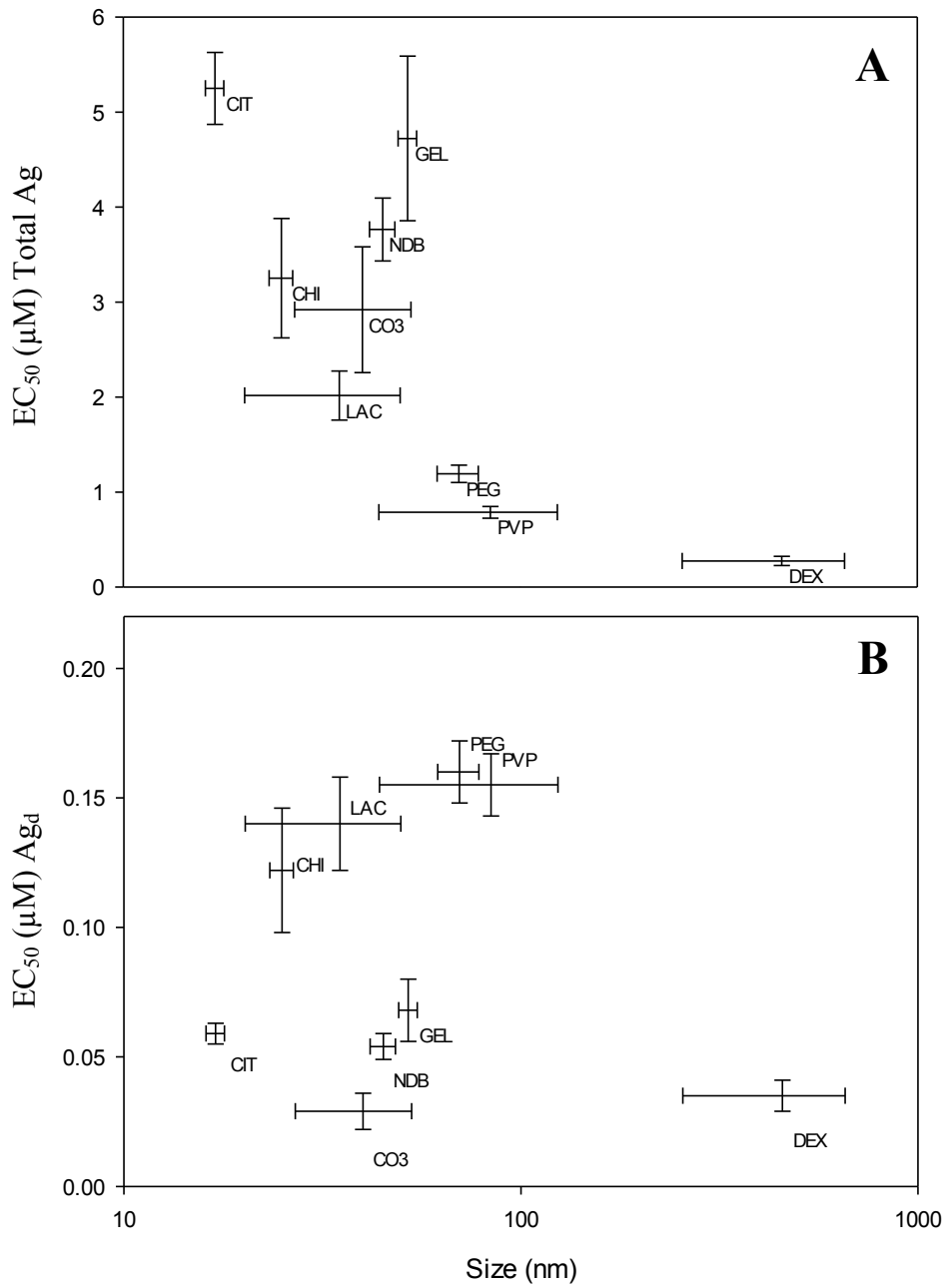
339



340

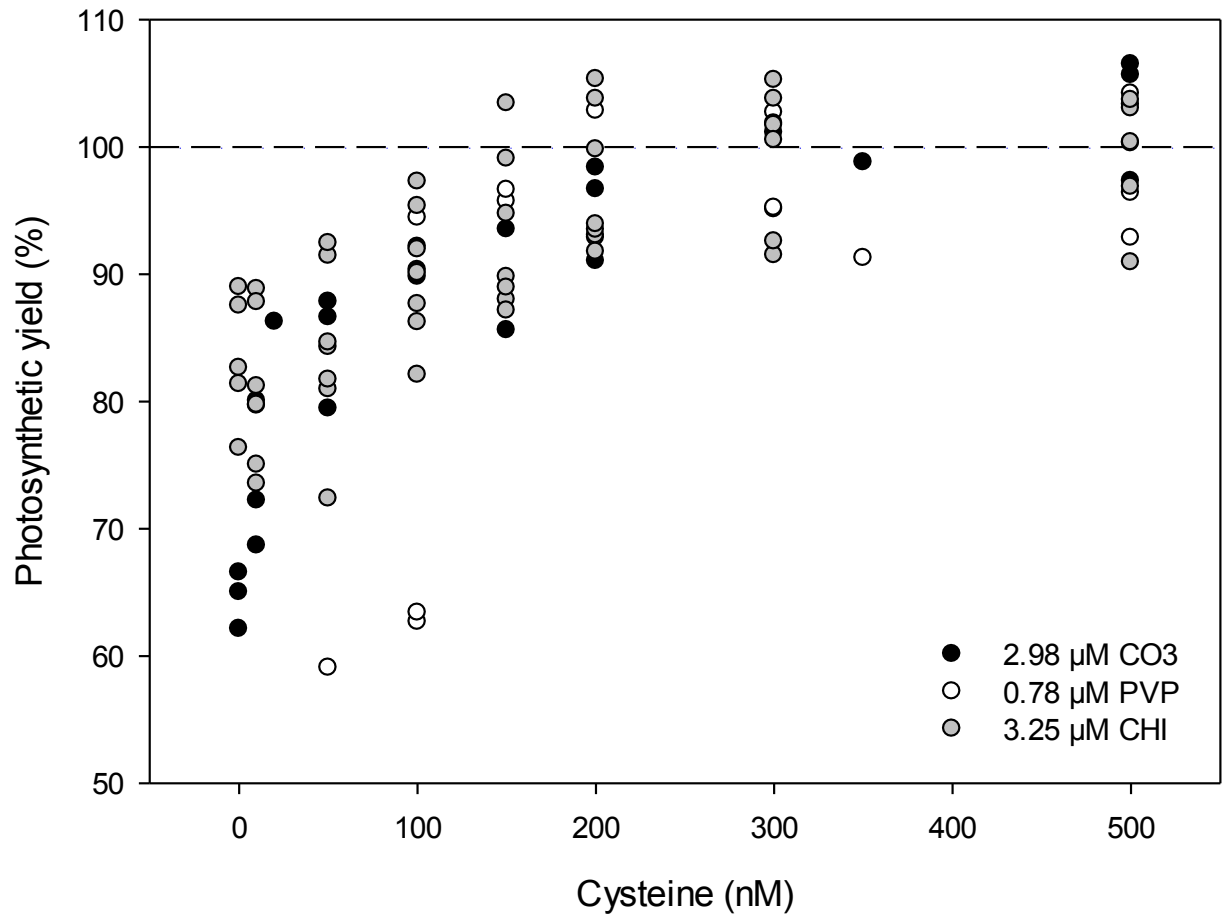
Figure 2.

341



343
344
345

Figure 3.



346
 347
 348
 349

Figure 4.

350 TABLES

351

Coating	Chemical family	Total Ag (mM)	Size (nm)	ζ Pot. (mV)	% of Ag _d 1h	% of Ag _d 2h
Carbonate (CO ₃)	Carbonates	9.27	40 ± 0.5	-36 ± 3.2	0.7 ± 0.1	0.8 ± 0.1
Chitosan (Chi)	Polypeptide+sugar	4.93	25 ± 1.7	-5.1 ± 1.7	3.7 ± 0.2	4.9 ± 0.1
Citrate (Cit)	Carboxylic acid	11.25	17 ± 0.9	-29 ± 3.3	1.1 ± 0.04	1.4 ± 0.03
Dexpanthenol (Dex)	Peptide (imides)	11.42	456 ± 200	-3.8 ± 2.3	12.5 ± 0.69	11.5 ± 0.18
Gelatine (Gel)	Polypeptide	4.79	52 ± 2.8	-6.8 ± 3.0	1.4 ± 0.1	2.4 ± 0.1
Lactate (Lac)	Carboxylic acid	11.35	35 ± 14.8	-3.5 ± 1.5	9.2 ± 2	13.2 ± 3.8
Na-dodecylbenzenesulfonate (NDB)	Detergent	2.80	45 ± 3.3	-29 ± 2.0	1.9 ± 0.1	4.1 ± 2.5
Polyethelenglycol (PEG)	Hydroxi-alcohol polymer	11.88	70 ± 8.3	-5.5 ± 2.4	16.2 ± 2.4	15.5 ± 2.3
Polyvinyl poly pyrrolidone (PVP)	Hydrocarbons (alkenes)	10.67	84 ± 40.0	-6.3 ± 3.1	19.6 ± 0.6	19.8 ± 1.5

352

353 **Table 1.** AgNP characteristics. Total Ag measured in AgNP stock solutions, average AgNP size,354 ζ - potential, and Ag_d (as percentage of total Ag) measured in AgNP suspensions 1 and 2 h after

355 dilution in 10 mM MOPS, pH 7.5.

356

A. Total Ag				B. Dissolved Ag		
AgNP	EC ₅₀ (μM) at 1h	EC ₅₀ Ratio	p-value	EC ₅₀ (μM)	EC ₅₀ Ratio	p-value
AgNO ₃	0.18 ± 0.03	-	-	0.176 ± 0.029	-	-
CO ₃	2.92 ± 0.66	6.01E-02	0.000	0.029 ± 0.007	6.0	0.004
CHI	3.25 ± 0.62	5.40E-02	0.000	0.122 ± 0.024	1.4	0.240
CIT	5.25 ± 0.38	3.34E-02	0.000	0.059 ± 0.004	2.9	0.000
DEX	0.28 ± 0.05	6.35E-01	0.021	0.035 ± 0.006	5.1	0.001
GEL	4.72 ± 0.86	3.72E-02	0.000	0.068 ± 0.012	2.6	0.016
LAC	2.02 ± 0.26	8.71E-02	0.000	0.140 ± 0.018	1.2	0.348
NDB	3.76 ± 0.33	4.66E-02	0.000	0.054 ± 0.005	3.2	0.000
PEG	1.19 ± 0.09	1.47E-01	0.000	0.160 ± 0.012	1.1	0.626
PVP	0.79 ± 0.06	2.23E-01	0.000	0.155 ± 0.012	1.1	0.529

357

358

359 **Table 2:** EC₅₀ values (1 h) and standard error of all AgNP calculated based on measured total (A)360 and measured Ag) concentrations (B), and ratio of EC₅₀ of AgNO₃ to EC₅₀ of each AgNP. The361 result of the statistical comparisons of EC₅₀ values (as a function of Ag_d) of AgNO₃ relative to362 coated AgNP is shown in the last column; if the EC₅₀ ratio significantly differs from 1 (p-value <363 0.05), those EC₅₀ values are significantly different.

364

365

366

367

AgNP	Total Ag (μM)	Ag_d (μM)	Cysteine (μM)	Ratio Cys/ Ag_d	Corrected EC_{50} (Ag_{bioav})
AgNO3	0.50	0.500	0.595	1.19	0.21
CO3	2.98	0.021	0.306	14.67	0.43
CHI	3.25	0.120	0.316	2.63	0.32
CIT	3.20	0.035	0.130	3.69	0.22
DEX	1.52	0.190	0.369	1.94	0.07
GEL	5.17	0.072	0.599	8.28	0.56
LAC	2.02	0.186	0.327	1.76	0.25
NDB	3.77	0.072	0.072	1.01	0.05
PEG	1.19	0.193	0.593	3.08	0.48
PVP	0.78	0.153	0.270	1.74	0.27

369

370 **Table 3.** Total Ag concentrations at which algae have been exposed, measured concentration of
371 Ag_d , estimated cysteine concentration that completely prevents the toxicity of each AgNP, ratio of
372 cysteine to measured Ag_d and EC_{50} values corrected as function of estimated Ag_{bioav} . This has
373 been calculated multiplying the EC_{50} values as a function of Ag_d (from Table 2B) by the
374 cysteine/ Ag_d ratio.

375

- 377 1. Panacek, A.; Kvitek, L.; Prucek, R.; Kolar, M.; Vecerova, R.; Pizurova, N.; Sharma, V. K.;
378 Nevecna, T.; Zboril, R., Silver colloid nanoparticles: Synthesis, characterization, and their
379 antibacterial activity. *Journal of Physical Chemistry B* **2006**, *110*, (33), 16248-16253.
- 380 2. Pal, S.; Tak, Y. K.; Song, J. M., Does the antibacterial activity of silver nanoparticles
381 depend on the shape of the nanoparticle? A study of the gram-negative bacterium *Escherichia coli*.
382 *Appl Environ Microb* **2007**, *73*, (6), 1712-1720.
- 383 3. Lee, W. F.; Huang, Y. C., Swelling and antibacterial properties for the superabsorbent
384 hydrogels containing silver nanoparticles. *Journal of Applied Polymer Science* **2007**, *106*, (3),
385 1992-1999.
- 386 4. Shahverdi, A. R.; Fakhimi, A.; Shahverdi, H. R.; Minaian, S., Synthesis and effect of silver
387 nanoparticles on the antibacterial activity of different antibiotics against *Staphylococcus aureus*
388 and *Escherichia coli*. *Nanomed-Nanotechnol* **2007**, *3*, (2), 168-171.
- 389 5. Navarro, E.; Piccapietra, F.; Wagner, B.; Marconi, F.; Kaegi, R.; Odzak, N.; Sigg, L.;
390 Behra, R., Toxicity of Silver Nanoparticles to *Chlamydomonas reinhardtii*. *Environ. Sci. Technol.*
391 **2008**, *42*, (23), 8959-8964.
- 392 6. He, D.; Dorantes-Aranda, J. J.; Waite, T. D., Silver Nanoparticle-Algae Interactions:
393 Oxidative Dissolution, Reactive Oxygen Species Generation and Synergistic Toxic Effects.
394 *Environ. Sci. Technol.* **2012**, *46*, (16), 8731-8738.
- 395 7. Miao, A. J.; Schwehr, K. A.; Xu, C.; Zhang, S. J.; Luo, Z. P.; Quigg, A.; Santschi, P. H.,
396 The algal toxicity of silver engineered nanoparticles and detoxification by exopolymeric
397 substances. *Environmental Pollution* **2009**, *157*, (11), 3034-3041.
- 398 8. Zhao, C. M.; Wang, W. X., Importance of surface coatings and soluble silver in silver
399 nanoparticles toxicity to *Daphnia magna*. *Nanotoxicology* **2012**, *6*, (4), 361-370.
- 400 9. Morones, J. R.; Elechiguerra, J. L.; Camacho, A.; Holt, K.; Kouri, J. B.; Ramirez, J. T.;
401 Yacaman, M. J., The bactericidal effect of silver nanoparticles. *Nanotechnology* **2005**, *16*, (10),
402 2346-2353.
- 403 10. Sondi, I.; Salopek-Sondi, B., Silver nanoparticles as antimicrobial agent: a case study on *E-*
404 *coli* as a model for Gram-negative bacteria. *Journal of Colloid and Interface Science* **2004**, *275*,
405 (1), 177-182.
- 406 11. Xiu, Z. M.; Ma, J.; Alvarez, P. J. J., Differential Effect of Common Ligands and Molecular
407 Oxygen on Antimicrobial Activity of Silver Nanoparticles versus Silver Ions. *Environ. Sci.*
408 *Technol.* **2011**, *45*, (20), 9003-9008.
- 409 12. Xiu, Z. M.; Zhang, Q. B.; Puppala, H. L.; Colvin, V. L.; Alvarez, P. J. J., Negligible
410 Particle-Specific Antibacterial Activity of Silver Nanoparticles. *Nano Letters* **2012**, *12*, (8), 4271-
411 4275.
- 412 13. Aka Sagliker, H.; Kizildag, N.; Cenkseven, S.; Darici, C.; Kocak, B.; Yarpuz Bozdogan,
413 N.; Daglioglu, N., Effects of Imazamox on Soil Carbon and Nitrogen Mineralization under Two
414 Different Humidity Conditions. *Ekoloji* **2014**, *23*, (91), 22-28.
- 415 14. Patil, S. J.; Adhikari, A.; Baghini, M. S.; Rao, V. R., An ultra-sensitive piezoresistive
416 polymer nano-composite microcantilever platform for humidity and soil moisture detection.
417 *Sensor Actuat B-Chem* **2014**, *203*, 165-173.
- 418 15. Kukumagi, M.; Ostonen, I.; Kupper, P.; Truu, M.; Tulva, I.; Varik, M.; Aosaar, J.; Sober,
419 J.; Lohmus, K., The effects of elevated atmospheric humidity on soil respiration components in a
420 young silver birch forest. *Agr Forest Meteorol* **2014**, *194*, 167-174.
- 421 16. Behra, R.; Sigg, L.; Clift, M. J. D.; Herzog, F.; Minghetti, M.; Johnston, B.; Petri-Fink, A.;
422 Rothen-Rutishauser, B., Bioavailability of silver nanoparticles and ions: from a chemical and
423 biochemical perspective. *Journal of the Royal Society Interface* **2013**, *10*, (87).
- 424 17. Sigg, L.; Behra, R.; Groh, K.; Isaacson, C.; Odzak, N.; Piccapietra, F.; Röhder, L.; Schug,
425 H.; Yue1, Y.; Schirmer, K., Chemical aspects of nanoparticle ecotoxicology. *Chimia* **2014**,
426 *Submitted*.

- 427 18. Liu, J. Y.; Sonshine, D. A.; Shervani, S.; Hurt, R. H., Controlled Release of Biologically
428 Active Silver from Nanosilver Surfaces. *Acs Nano* **2010**, *4*, (11), 6903-6913.
- 429 19. Hurt, R. H.; Liu, J. Y., Ion Release Kinetics and Particle Persistence in Aqueous Nano-
430 Silver Colloids. *Environ. Sci. Technol.* **2010**, *44*, (6), 2169-2175.
- 431 20. Zhang, W.; Yao, Y.; Sullivan, N.; Chen, Y. S., Modeling the Primary Size Effects of
432 Citrate-Coated Silver Nanoparticles on Their Ion Release Kinetics. *Environ. Sci. Technol.* **2011**,
433 *45*, (10), 4422-4428.
- 434 21. Gondikas, A. P.; Morris, A.; Reinsch, B. C.; Marinakos, S. M.; Lowry, G. V.; Hsu-Kim,
435 H., Cysteine-Induced Modifications of Zero-valent Silver Nanomaterials: Implications for Particle
436 Surface Chemistry, Aggregation, Dissolution, and Silver Speciation. *Environ. Sci. Technol.* **2012**,
437 *46*, (13), 7037-7045.
- 438 22. Tejamaya, M.; Romer, I.; Merrifield, R. C.; Lead, J. R., Stability of Citrate, PVP, and PEG
439 Coated Silver Nanoparticles in Ecotoxicology Media. *Environ. Sci. Technol.* **2012**, *46*, (13), 7011-
440 7017.
- 441 23. Yang, X. Y.; Gondikas, A. P.; Marinakos, S. M.; Auffan, M.; Liu, J.; Hsu-Kim, H.; Meyer,
442 J. N., Mechanism of Silver Nanoparticle Toxicity Is Dependent on Dissolved Silver and Surface
443 Coating in *Caenorhabditis elegans*. *Environ. Sci. Technol.* **2012**, *46*, (2), 1119-1127.
- 444 24. Kennedy, A. J.; Hull, M. S.; Bednar, A. J.; Goss, J. D.; Gunter, J. C.; Bouldin, J. L.;
445 Vikesland, P. J.; Steevens, J. A., Fractionating Nanosilver: Importance for Determining Toxicity
446 to Aquatic Test Organisms. *Environ. Sci. Technol.* **2010**, *44*, (24), 9571-9577.
- 447 25. Van Hoecke, K.; De Schampelaere, K. A. C.; Ali, Z.; Zhang, F.; Elsaesser, A.; Rivera-
448 Gil, P.; Parak, W. J.; Smagghe, G.; Howard, C. V.; Janssen, C. R., Ecotoxicity and uptake of
449 polymer coated gold nanoparticles. *Nanotoxicology* **2013**, *7*, (1), 37-47.
- 450 26. Piccapietra, F.; Allue, C. G.; Sigg, L.; Behra, R., Intracellular Silver Accumulation in
451 *Chlamydomonas reinhardtii* upon Exposure to Carbonate Coated Silver Nanoparticles and Silver
452 Nitrate. *Environ. Sci. Technol.* **2012**, *46*, (13), 7390-7397.
- 453 27. Odzak, N.; Kistler, D.; Behra, R.; Sigg, L., Dissolution of metal and metal oxide
454 nanoparticles in aqueous media *Environ. Poll.* **2014**, *191*, 132-138.
- 455 28. Müller, B. *ChemEQL, a program to calculate chemical speciation*; Swiss Federal Institute
456 of Aquatic Science and Technology, Eawag: Kastanienbaum, CH-6047, 1996.
- 457 29. Le Faucheur, S.; Behra, R.; Sigg, L., Phytochelatin induction, cadmium accumulation, and
458 algal sensitivity to free cadmium ion in *Scenedesmus vacuolatus*. *Environ Toxicol Chem* **2005**, *24*,
459 (7), 1731-1737.
- 460 30. Lee, D. Y.; Fortin, C.; Campbell, P. G. C., Contrasting effects of chloride on the toxicity of
461 silver to two green algae, *Pseudokirchneriella subcapitata* and *Chlamydomonas reinhardtii*.
462 *Aquatic Toxicology* **2005**, *75*, (2), 127-135.
- 463 31. Schreiber, U.; Bilger, W.; Neubauer, C., Chlorophyll fluorescence as a nonintrusive
464 indicator for rapid assessment of in vivo photosynthesis. In *Ecophysiology of Photosynthesis*,
465 Schulze, E. D.; Caldwell, M. M., Eds. Springer: Berlin Heidelberg New York, 1994; Vol. 100, pp
466 49-70.
- 467 32. R_Development_Core_Team R: *A language and environment for statistical computing*, R
468 Foundation for Statistical Computing: Vienna, Austria. ISBN 3-900051-07-0, URL [http://www.R-](http://www.R-project.org)
469 [project.org](http://www.R-project.org), 2012.
- 470 33. Bar-Ilan, O.; Albrecht, R. M.; Fako, V. E.; Furgeson, D. Y., Toxicity Assessments of
471 Multisized Gold and Silver Nanoparticles in Zebrafish Embryos. *Small* **2009**, *5*, (16), 1897-1910.
- 472 34. Schneider, M.; Goss, K. U., Volatilization Modeling of Two Herbicides from Soil in a
473 Wind Tunnel Experiment under Varying Humidity Conditions. *Environ. Sci. Technol.* **2012**, *46*,
474 (22), 12527-12533.
- 475 35. Kent, R. A.; Andersen, D.; Caux, P. Y.; Teed, S., Canadian Water Quality Guidelines for
476 glycols - An ecotoxicological review of glycols and associated aircraft anti-icing and deicing
477 fluids. *Environmental Toxicology* **1999**, *14*, (5), 481-522.
- 478 36. Harford, A. J.; Hogan, A. C.; Jones, D. R.; van Dam, R. A., Ecotoxicological assessment of
479 a polyelectrolyte flocculant. *Water Res* **2011**, *45*, (19), 6393-6402.

- 480 37. Konnecker, G.; Regelman, J.; Belanger, S.; Gamon, K.; Sedlak, R., Environmental
481 properties and aquatic hazard assessment of anionic surfactants: Physico-chemical, environmental
482 fate and ecotoxicity properties. *Ecotoxicology and Environmental Safety* **2011**, *74*, (6), 1445-1460.
- 483 38. Potara, M.; Jakab, E.; Damert, A.; Popescu, O.; Canpean, V.; Astilean, S., Synergistic
484 antibacterial activity of chitosan-silver nanocomposites on *Staphylococcus aureus*.
485 *Nanotechnology* **2011**, *22*, (13).
- 486 39. Pospisil, P., Production of reactive oxygen species by photosystem II. *Biochim. Biophys.*
487 *Acta-Bioenerg.* **2009**, *1787*, (10), 1151-1160.
- 488 40. Sigg, L.; Lindauer, U., Silver nanoparticle dissolution in the presence of ligands and of
489 hydrogen peroxide. *Environmental Pollution (submitted)* **2015**.
- 490 41. Adams, N. W. H.; Kramer, J. R., Potentiometric determination of silver thiolate formation
491 constants using a Ag₂S electrode. *Aquat Geochem* **1999**, *5*, (1), 1-11.
- 492 42. Kapoor, S.; Lawless, D.; Kennepohl, P.; Meisel, D.; Serpone, N., Reduction and
493 Aggregation of Silver Ions in Aqueous Gelatin Solutions. *Langmuir* **1994**, *10*, (9), 3018-3022.
- 494 43. Mpourmpakis, G.; Vlachos, D. G., Insights into the early stages of metal nanoparticle
495 formation via first-principle calculations: the roles of citrate and water. *Langmuir* **2008**, *24*, (14),
496 7465-7473.
- 497 44. Rivas, L.; Sanchez-Cortes, S.; Garcia-Ramos, J. V.; Morcillo, G., Growth of silver
498 colloidal particles obtained by citrate reduction to increase the Raman enhancement factor.
499 *Langmuir* **2001**, *17*, (3), 574-577.
- 500 45. Rosen, M. J.; Li, F.; Morrall, S. W.; Versteeg, D. J., The relationship between the
501 interfacial properties of surfactants and their toxicity to aquatic organisms. *Environ. Sci. Technol.*
502 **2001**, *35*, (5), 954-959.
- 503 46. Tulliani, J. M.; Baroni, C.; Zavattaro, L.; Grignani, C., Strontium-Doped Hematite as a
504 Possible Humidity Sensing Material for Soil Water Content Determination. *Sensors-Basel* **2013**,
505 *13*, (9), 12070-12092.

506

507

Species	Stoich. Matrix	Log K	Conc. (mol/l)	Log conc.
Ag+	1 0 0	0.00E+00	4.52E-11	-1.03E+01
AgCysH	1 1 1	2.27E+01	5.00E-07	-6.30E+00
Ag(CysH)2-	1 2 2	3.67E+01	1.10E-11	-1.10E+01
Cys2-	0 1 0	0.00E+00	7.12E-12	-1.12E+01
CysH-	0 1 1	1.08E+01	1.39E-08	-7.86E+00
CysH2	0 1 2	1.91E+01	8.61E-08	-7.07E+00
CysH3+	0 1 3	2.08E+01	1.34E-13	-1.29E+01
H+	0 0 1	0.00E+00	3.10E-08	-7.51E+00
Components	Mode	Initial Conc.	In or out of system	Case of
Ag+	total	5.00E-07	----	AgNO3
Cys2-	total	6.00E-07	----	
H+	free	3.10E-08	7.17E-07	

Species	Stoich. Matrix	Log K	Conc. (mol/l)	Log conc.
Ag+	1 0 0	0.00E+00	1.64E-09	-8.78E+00
AgCysH	1 1 1	2.27E+01	2.98E-07	-6.53E+00
Ag(CysH)2-	1 2 2	3.67E+01	1.08E-13	-1.30E+01
Cys2-	0 1 0	0.00E+00	1.17E-13	-1.29E+01
CysH-	0 1 1	1.08E+01	2.29E-10	-9.64E+00
CysH2	0 1 2	1.91E+01	1.41E-09	-8.85E+00
CysH3+	0 1 3	2.08E+01	2.20E-15	-1.47E+01
H+	0 0 1	0.00E+00	3.10E-08	-7.51E+00
Components	Mode	Initial Conc.	In or out of system	Case of
Ag+	total	3.00E-07	----	AgNO3
Cys2-	total	3.00E-07	----	
H+	free	3.10E-08	3.32E-07	

Species	Stoich. Matrix	Log K	Conc. (mol/l)	Log conc.
Ag+	1 0 0	0.00E+00	7.75E-12	-1.11E+01
AgCysH	1 1 1	2.27E+01	6.00E-08	-7.22E+00
Ag(CysH)2-	1 2 2	3.67E+01	9.26E-13	-1.20E+01
Cys2-	0 1 0	0.00E+00	4.98E-12	-1.13E+01
CysH-	0 1 1	1.08E+01	9.74E-09	-8.01E+00
CysH2	0 1 2	1.91E+01	6.03E-08	-7.22E+00
CysH3+	0 1 3	2.08E+01	9.36E-14	-1.30E+01
H+	0 0 1	0.00E+00	3.10E-08	-7.51E+00
Components	Mode	Initial Conc.	In or out of system	Case of
Ag+	total	6.00E-08	----	CIT-AgNP
Cys2-	total	1.30E-07	----	
H+	free	3.10E-08	2.21E-07	

Species	Stoich. Matrix	Log K	Conc. (mol/l)	Log conc.
Ag ⁺	1 0 0	0.00E+00	2.71E-11	-1.06E+01
AgCysH	1 1 1	2.27E+01	5.40E-08	-7.27E+00
Ag(CysH) ₂ ⁻	1 2 2	3.67E+01	2.15E-13	-1.27E+01
Cys ²⁻	0 1 0	0.00E+00	1.28E-12	-1.19E+01
CysH ⁻	0 1 1	1.08E+01	2.51E-09	-8.60E+00
CysH ₂	0 1 2	1.91E+01	1.55E-08	-7.81E+00
CysH ₃ ⁺	0 1 3	2.08E+01	2.41E-14	-1.36E+01
H ⁺	0 0 1	0.00E+00	3.10E-08	-7.51E+00
Components	Mode	Initial Conc.	In or out of system	Case of
Ag ⁺	total	5.40E-08	----	NDB-AgNP
Cys ²⁻	total	7.20E-08	----	
H ⁺	free	3.10E-08	1.19E-07	

Table S1. Ag⁺ speciation calculations examples using ChemEQL 3.0 (Müller, 1996) First scenario show that a slight excess of cysteine concentrations (0.6 cysteine vs 0.5 AgNO₃), already resulted in extremely low Ag⁺ concentrations (4.5E-11). In the second case, even at an equimolar scenario (0.3 for both compounds), the amount of Ag⁺ is around 1E-9.

REFERENCES

Müller, B. ChemEQL, a program to calculate chemical speciation; Swiss Federal Institute of Aquatic Science and Technology, Eawag: Kastanienbaum, CH-6047, 1996.

Reaction	log K	Reference
$\text{Ag}^+ + \text{Cys}^{2-} + \text{H}^+ \rightleftharpoons \text{AgCysH}$	22.7	1
$\text{Ag}^+ + 2 \text{Cys}^{2-} + 2 \text{H}^+ \rightleftharpoons \text{Ag}(\text{CysH})_2^-$	36.7	
$\text{Ag}^+ + \text{H}_2\text{O} \rightleftharpoons \text{AgOH} + \text{H}^+$	-12	2
$\text{Ag}^+ + 2 \text{H}_2\text{O} \rightleftharpoons \text{Ag}(\text{OH})_2^- + 2 \text{H}^+$	-24	
$\text{Cys}^{2-} + \text{H}^+ \rightleftharpoons \text{CysH}^-$	10.8	
$\text{Cys}^{2-} + 2 \text{H}^+ \rightleftharpoons \text{CysH}_2$	19.1	
$\text{Cys}^{2-} + 3 \text{H}^+ \rightleftharpoons \text{CysH}_3^+$	20.8	

(1) Adams, N. W. H.; Kramer, J. R. Potentiometric determination of silver thiolate formation constants using a Ag_2S electrode. *Aquat. Geochem.* 1999, 5, 1-11.

(2) Martell, A. E.; Smith, R. M.; NIST Standard Reference Database 46, 2001.

Table S2. Stability constants used to calculate the Ag^+ concentrations at equilibrium with cysteine. Constants for AgCysH and $\text{Ag}(\text{CysH})_2$ are calculated from (1) after J. Kramer (personal communication, 2007).

Exp. Number	AgNP coating	(AgNP) μ M	(Cysteine) nM
Exp 1	all AgNP	2	
Exp 2	PEG	0.01, 0.05, 0.1, 0.5, 1, 2, 5, 20	
Exp 3	PEG	0.01, 0.05, 0.1, 0.5, 1, 2, 5, 20	
Exp 4	LAC	0.01, 0.05, 0.1, 0.5, 1, 2, 5, 20	
Exp 5	LAC	0.01, 0.05, 0.1, 0.5, 1, 2, 5, 50	
Exp 6	LAC	0.01, 0.05, 0.1, 0.5, 1, 2, 5, 50	
Exp 7	PEG	0.01, 0.05, 0.1, 0.5, 1, 2, 5, 20	
Exp 8	PEG	0.01, 0.05, 0.1, 0.5, 1, 2, 5, 20	
Exp 9	GEL	0.1, 0.5, 1, 2, 5, 20, 50, 100	
Exp 10	GEL	0.1, 0.5, 1, 2, 5, 20, 50, 100	
Exp 11	GEL	0.1, 0.5, 1, 2, 5, 20, 50, 100, 200, 500	
Exp 12	CHI	0.1, 0.5, 1, 2, 5, 20, 50, 100	
Exp 13	CHI	0.1, 0.5, 1, 2, 5, 20, 50, 100	
Exp 14	CHI	0.1, 0.5, 1, 2, 5, 20, 50, 100	
Exp 15	PVP	0.05, 0.1, 0.5, 1, 2, 5, 20, 50	
Exp 16	PVP	0.1, 0.5, 0.75, 1, 1.5, 1.75, 2, 5	
Exp 17	PVP	0.1, 0.5, 0.75, 1, 1.5, 1.75, 2, 5	
Exp 18	Lac,Gel,Chi, PVP,PEG	2.16, 4.2, 2.84, 0.78, 1.3 respectively	500
Exp 19	CO3	10	0, 10, 20, 50,100,150,200, 350, 500
Exp 20	GEL	10	0, 10, 20, 50,100,150,200, 350, 500
Exp 21	PVP	10	0, 10, 20, 50,100,150,200, 350, 500
Exp 22	CO3	2.98	0, 10, 20, 50,100,150,200, 350, 500
Exp 23	PVP	0.78	0, 10, 20, 50,100,150,200, 350, 500
Exp 24	PEG	1.3	0, 10, 20, 50,100,150,200, 350, 500
Exp 25	PVP	0.78	0, 10, 20, 50,100,150,200,300, 350, 500
Exp 26	PEG	1.3	0, 10, 20, 50,100,150,200,300, 350, 500
Exp 27	CO3	2.98	0, 10, 20, 50,100,150,200,300, 350, 500
Exp 28	PVP	0.78	0, 10, 20, 50,100,150,200,300, 350, 500
Exp 29	CO3	2.98	0, 10,50,100,150,200, 300, 500
Exp 30	CO3	2.98	0, 10,50,100,150,200, 300, 500
Exp 31	PVP	0.78	0, 10,50,100,150,200, 300, 500
Exp 32	PEG	1.3	0, 10,50,100,150,200, 300, 500
Exp 33	LAC	2.16	0, 10,50,100,150,200, 300, 500
Exp 34	CHI	2.84	0, 10,50,100,150,200, 300, 500
Exp 35	CIT	0.1, 0.5, 0.75, 1, 1.5, 1.75, 2, 5	
Exp 36	CIT	0.1, 0.5, 0.75, 1, 1.5, 1.75, 2, 5	
Exp 37	CIT	0.1, 0.5, 0.75, 1, 1.5, 1.75, 2, 5	
Exp 38	DEX	0.01, 0.05, 0.1, 0.5, 1, 2, 5, 20	
Exp 39	DEX	0.01, 0.05, 0.1, 0.5, 1, 2, 5, 20	
Exp 40	DEX	0.01, 0.05, 0.1, 0.5, 1, 2, 5, 20	
Exp 41	CHI	2.84	0, 10,50,100,150,200, 300, 500
Exp 42	CHI	2.84	0, 10,50,100,150,200, 300, 500
Exp 43	LAC	2.16	0, 10,50,100,150,200, 300, 500
Exp 44	CHI	2.16	0, 10,50,100,150,200, 300, 500
Exp 45	CHI	2.84	0, 10,50,100,150,200, 300, 500
Exp 46	CHI	2.84	0, 10,50,100,150,200, 300, 500
Exp 47	CHI	2.84	0, 10,50,100,150,200, 300, 500
Exp 48	LAC	2.16	0, 10,50,100,150,200, 300, 500
Exp 49	DEX	0.29	0, 10,50,100,150,200, 300, 500
Exp 50	DEX	0.29	0, 10,50,100,150,200, 300, 500

Exp 51	DEX	0.29	0, 10,50,100,150,200, 300, 500
Exp 52	CIT	3.53	0, 10,50,100,150,200, 300, 500
Exp 53	CIT	3.53	0, 10,50,100,150,200, 300, 500
Exp 54	CIT	3.53	0, 10,50,100,150,200, 300, 500
Exp 55	CIT	3.53	0, 2, 5, 10, 20, 50,100, 200
Exp 56	CIT	3.53	0, 2, 5, 10, 20, 50,100, 200
Exp 57	CIT	3.53	0, 2, 5, 10, 20, 50,100, 200
Exp 58	DEX	0.29	0, 2, 5, 10, 20, 50,100, 200
Exp 59	DEX	0.29	0, 2, 5, 10, 20, 50,100, 200
Exp 60	DEX	0.29	0, 2, 5, 10, 20, 50,100, 200
Exp 61	NDB	0.1, 0.5, 1, 5, 20, 50, 100, 200	
Exp 62	NDB	0.1, 0.5, 1, 5, 20, 50, 100, 200	
Exp 63	NDB	0.1, 0.5, 1, 5, 20, 50, 100, 200	
Exp 64	NDB	3.77	0, 10,50,100,150,200, 300, 500
Exp 65	NDB	3.77	0, 10,50,100,150,200, 300, 500
Exp 66	NDB	3.77	0, 10,50,100,150,200, 300, 500
Exp 67	LAC	0, 0.02, 2, 40	
Exp 68	DEX	1	0, 10,50,100,150,200, 300, 500
Exp 69	DEX	1	0, 10,50,100,150,200, 300, 500
Exp 70	DEX	1	0, 10,50,100,150,200, 300, 500
Exp 71	GEL	5	0, 20,50,100,150,200, 350, 500
Exp 72	GEL	10	0, 10,50,100,150,200, 300, 500
Exp 73	GEL	10	0, 10,50,100,150,200, 300, 500
Exp 74	GEL	10	0, 10,50,100,150,200, 300, 500
Exp 75	CIT	0, 0.5, 1, 2, 5, 10, 15, 20, 30	
Exp 76	CIT	0, 0.5, 1, 2, 5, 10, 15, 20, 30	
Exp 77	CIT	0, 0.5, 1, 2, 5, 10, 15, 20, 30	
Exp 78	CIT	0, 0.5, 1, 2, 5, 10, 15, 20, 30	
Exp 79	CIT	0, 0.5, 1, 2, 5, 10, 15, 20, 30	
Exp 80	CIT	0, 0.5, 1, 2, 5, 10, 15, 20, 30	
Exp 81	CIT	0, 0.5, 1, 5, 10, 30, 50, 100	
Exp 82	CIT	0, 0.5, 1, 5, 10, 30, 50, 100	
Exp 83	CIT	0, 0.5, 1, 5, 10, 30, 50, 100	

Table S3. List of experiments, detailing AgNP (nominal) and cysteine concentrations used; every experiment was done using 3 replicates. In experiments using cysteine, a control algal treatment without AgNP was used; the rest of algae were exposed to the fix AgNP concentration shown on the third column and to the increasing cysteine concentrations shown in the fourth column. Different AgNP and cysteine concentrations were used in order to obtain the better concentration-response curve possible for modeling purposes.

Coating	Time	(conc) μM	Size ave	St dev	Pdl	St dev	ζ	stdev
CO3	1	920	40.2	0.5	0.4	4.2E-03	-36.0	3.2
CO3	47	920	44.7	1.5	0.3	4.9E-03	-35.3	3.4
CHI	1	2.7	25.7	1.7	0.4	2.3E-03	-5.1	1.7
CHI	2	2.7	23.1	0.6	0.4	1.3E-02	-7.5	1.3
CIT	1	5.2	17.7	0.9	0.6	3.2E-02	-29.4	3.3
CIT	2	5.2	20.5	1.4	0.6	1.8E-01	-26.6	2.4
DEX	1	0.2	456.1	200.1	0.7	7.5E-02	-3.9	2.3
DEX	2	0.2	994.9	999.7	0.8	1.8E-01	-1.0	0.7
GEL	1	10	52.3	2.8	0.3	4.7E-02	-6.8	3.0
GEL	2	10	49.1	1.1	0.3	1.5E-03	-9.6	5.8
GEL	1	50	40.8	0.5	0.2	4.7E-03	-32.1	2.4
LAC	1	10	35.8	14.8	0.2	9.3E-02	-2.6	0.1
LAC	2	10	44.8	4.8	0.4	2.1E-02	-5.9	1.6
LAC	1	1	87.7	55.4	0.6	2.2E-01	-11.5	1.5
LAC	1	5	50.5	1.8	0.6	1.1E-01	-3.9	2.5
NADBS	1	3.7	45.9	3.4	0.4	1.1E-01	-29.5	2.0
NADBS	2	3.7	58.8	3.4	0.4	3.1E-02	-17.6	2.1
PEG	1	10	70.9	8.3	0.3	7.8E-02	-5.5	2.5
PEG	2	10	82.8	15.8	0.4	2.2E-01	-5.6	1.3
PEG	1	1	115.0	21.7	0.7	5.6E-02	-3.1	0.4
PEG	1	5	30.1	1.1	0.5	1.2E-01	-1.3	1.3
PVP	1	2.7	71.7	23.3	0.5	1.7E-01	-6.3	3.2
PVP	2	2.7	97.0	12.7	0.4	1.8E-01	-2.6	0.4

Table S4. Size in nm (including PDI) and ζ - potential measurements. Standard deviations have been calculated using, at least, three replicates. Time units are hours.

Total Ag			
AgNP	EC ₅₀ (μM) at 1h	EC ₅₀ 1h EC ₅₀ 2h	p-value
AgNO ₃	0.176 \pm 0.029	1.1	0.561
CO3	2.920 \pm 0.662	3.3	0.049
CHI	3.252 \pm 0.628	1.2	0.616
CIT	5.250 \pm 0.378	0.7	0.016
DEX	0.276 \pm 0.049	2.0	0.044
GEL	4.723 \pm 0.866	0.9	0.729
LAC	2.017 \pm 0.258	1.2	0.305
NDB	3.765 \pm 0.331	0.7	0.004
PEG	1.193 \pm 0.092	1.6	0.004
PVP	0.788 \pm 0.062	1.3	0.099

Table S5. The second column shows EC₅₀ at 1 hour (error term is standard error) for all AgNP; the third column shows the ratio between EC₅₀ (calculated as a total Ag) after 1 and 2 hours of exposure. The p-values < 0.05 indicate ratios significantly different from 1 and thus time-related differences in toxicity expressed as a function of total Ag. While algae exposed to AgNO₃, and AgNP coated with CHI, GEL, LAC show very similar EC₅₀ after 1 and 2 h of exposure; AgNP coated with CO3, CIT, DEX, NDB, PEG and PVP show a lower EC₅₀ after 2 h.

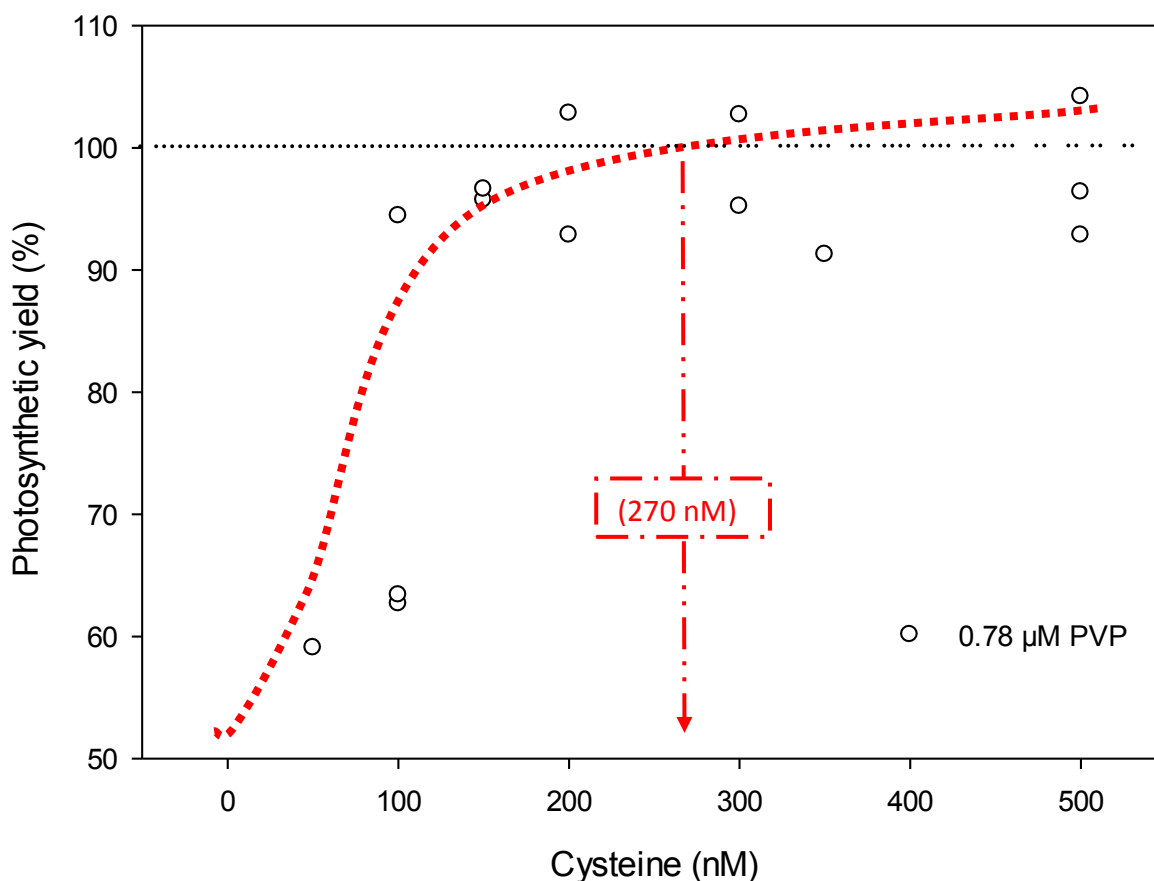


Figure S1. Example of the calculation of the amount of cysteine required to prevent the toxicity of a certain amount of PVP coated AgNP to the photosynthetic yield of *C. reinhardtii*. Graph shows the effect of increasing concentrations of cysteine to the photosynthetic yield. In this example, toxicity data from PVP coated AgNPs were fitted to a four parameters logistic model, that was the one showing the best fitting:

Algae exposed to 0.783 μ M of AgNP coated with PVP				
$y = \min + (\max - \min) / (1 + (x/EC50)^{-Hillslope})$			Using DataFit (Oakley)	
Model fit R sq = 0.91				
(Cys nM)	Yield	Parameter	Value	StdErr
270	100	min	25.25	4.663
		max	110	7.036
		EC50	84.04	14.42
		Hillslope	1.725	0.5055

The PVP toxicity data model is shown using a red dotted line in Figure S1. The calculated parameters are shown in the table. These were later used for estimating the (Cys nM) that resulted in a value of 100% of the photosynthetic yield:

$$\text{Photosynthetic yield } 100 = 25.25 + (110 - 25.5) / (1 + ((\text{cysteine}) / 84.04)^{-1.725})$$

In this case, the amount of cysteine that prevents the toxicity to the photosynthesis (i.e. value of 99.7 % of photosynthesis vs. control) is 270 nM.

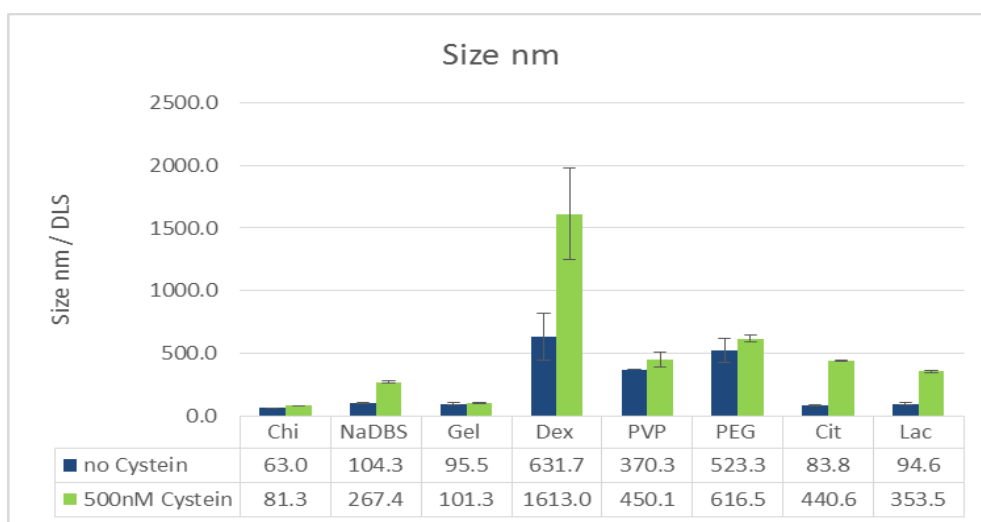
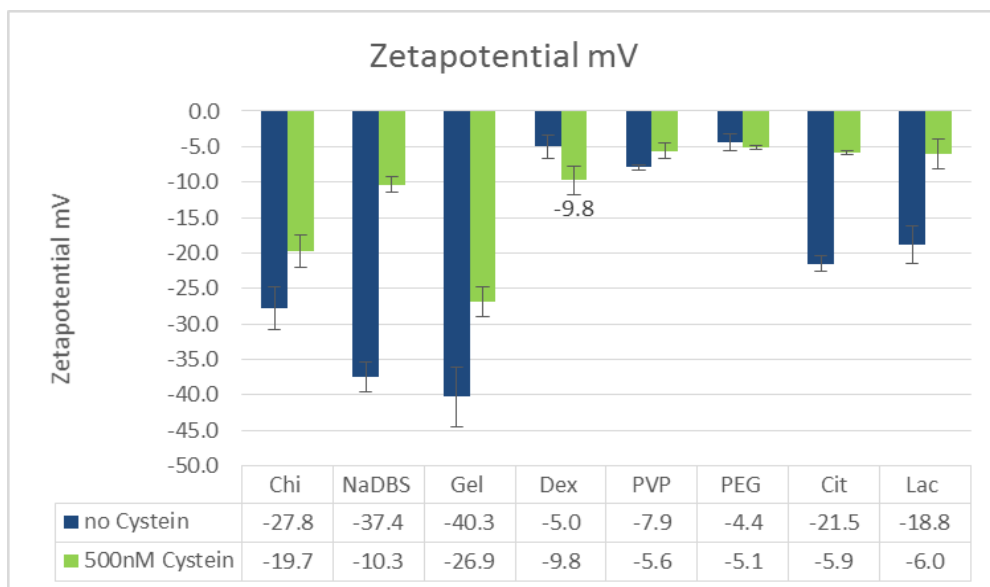


Figure S2. Influence of cysteine on ζ - potential and size of differently coated AgNP assessed at total Ag concentrations corresponding to the EC₅₀ values. These measurements have been carried out 2 years after the original experiments. Thus absolute values were not considered in this new version of the manuscript.

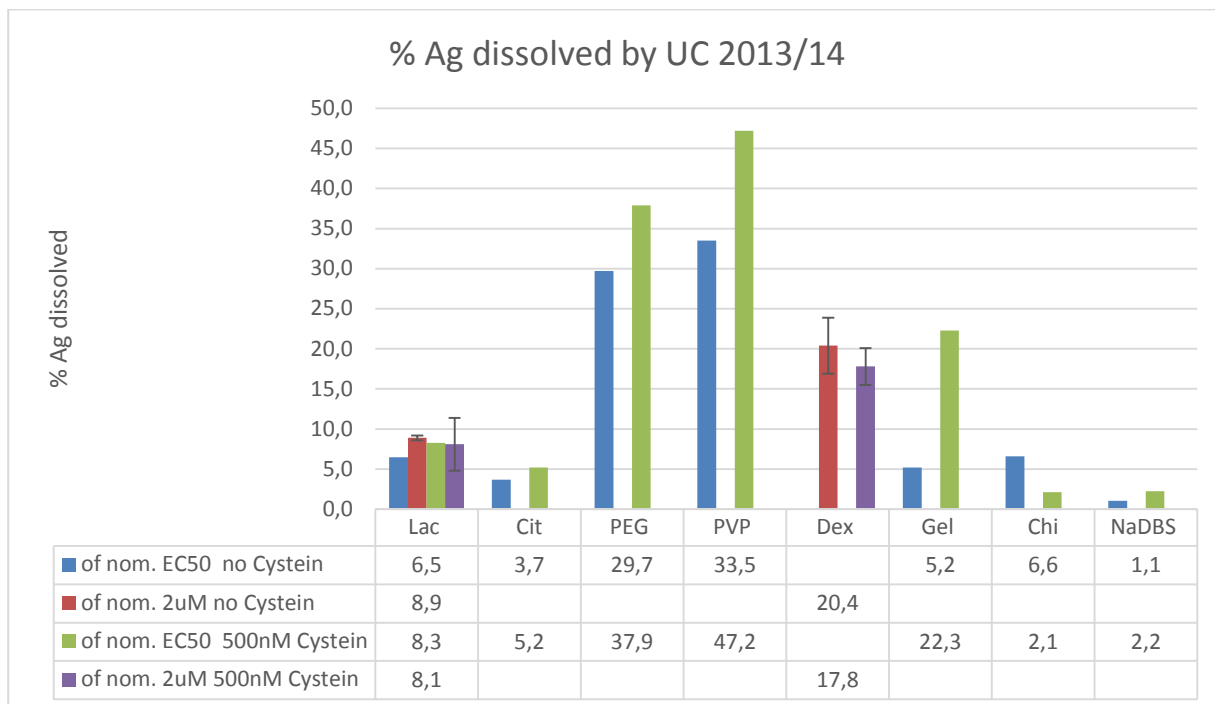


Figure S3. Influence of cysteine on Ag dissolution from the different AgNPs used in this study. Previous data obtained for 5 μM carbonate coated AgNP and different concentrations of cysteine shown that after 1 h dissolution is increased at concentrations equal or higher to 1 μM cysteine. The methods used here are those detailed in the paper; the Ag_d present in AgNPs suspensions (at the corresponding EC_{50} values or at 2 μM as indicated in the legend of the last graph) with and without cysteine (500 nM) was measured by Centrifugal ultrafiltration followed by ICP-MS.

These measurements have been carried out 2 years after the original experiments. Even if the absolute values were not considered in this new version of the manuscript, the influence of cysteine in particle dissolution has been included in the discussion.

Calculation S1. Ag_d ADHERED BY THE ALGAE

Example of calculation with data from Piccapietra's paper (ref. 26):

For 2 μM AgNP intracellular Ag is $5 \times 10^{-4} \text{ mol L}_{\text{cell}}^{-1}$

With the cell volume $V_{\text{cell}} = 100 \text{ fL}$, $\text{Ag}(\text{intracellular}) = 5 \times 10^{-17} \text{ mol/cell}$

With $1.5 \times 10^8 \text{ cells/L}$

$\text{Ag}(\text{cells}) = 7.5 \times 10^{-9} \text{ mol/L}$

The cysteine wash would account for a factor 2-3 for dissolved Ag. Adsorbed AgNP are not relevant in this calculation, as they may contribute to Ag-uptake by dissolution in contact with the cells.