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## Effects of wind farms on Montagu's harrier (*Circus pygargus*) in southern Spain

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24 **Abstract**

25

26 To study the potential impact of wind turbines and associated structures on Montagu's harriers  
27 *Circus pygargus*, we located 111 nests over five years (18 – 28 nests per year) and compared their  
28 distances to several features (natural and anthropogenic) between wind farm pre- and post-  
29 construction periods using a before-after (BA) study design. In addition, we deployed an impact  
30 gradient (IG) study design to examine the potential impact of turbines on Montagu's harrier nest  
31 locations. We analysed abundance and density of nests and colonies through the study period. We  
32 also fitted a predictive model of nest occurrence using distance-to-feature variables and habitat as  
33 predictors. Lastly, Montagu's harrier fatalities from collision with wind turbines were estimated. No  
34 differences were detected between pre- and post-construction periods in nest or colony abundances.  
35 We found that harriers nested closer to the locations of wind turbines and power lines after wind  
36 farm construction, although distance to closest track did not change. We detected a higher spatial  
37 aggregation of Montagu's harrier nests after wind farms were installed, when the distribution of  
38 nests was mostly explained by habitat and distance to the closest track. Distance to wind turbine was  
39 apparently not influential. Fatality through collision was relatively uncommon during the study  
40 period. Our findings demonstrate that the construction, operation and maintenance of wind farms  
41 did not seem to adversely affect Montagu's harrier nesting decisions in our study area. However, we  
42 encourage further studies including reproductive parameters and foraging strategies of Montagu's  
43 harrier to provide a complete investigation of potential impacts of wind farms on this species.

44

45 **Key-words:** BA design, IG design, bird collision, avoidance distance, disturbance, Tarifa.

46

47 **Introduction**

48

49 Numerous studies have focused on wind farm impacts on wildlife due to the spectacular growth in  
50 capturing wind energy as means to obtain 'clean' energy around the world in last decades  
51 (Smallwood and Thelander, 2005; Everaert and Stienen, 2007). One of the most studied impact of  
52 wind farms is bird mortality caused by collision with turbine blades (De Lucas et al., 2008; Drewitt  
53 and Langston, 2008; Marques et al., 2014). These deaths represent 0.01 – 0.02 % of annual avian  
54 mortality rates, and so lower than due to other human activities, such as electrocution at power lines  
55 (Calvert et al., 2013; Sovacool, 2013; Erickson et al., 2014). However, some discrepancy has arisen  
56 because of the high variation in indices of mortality caused by collisions with blades regarding to  
57 species or locations (Stewart et al., 2007), being raptors among the most affected species (Barrios  
58 and Rodríguez, 2004; Hoover and Morrison, 2005; Kuvlesky et al., 2007; Loss et al., 2013),  
59 independent of their abundance in the potentially affected area (Orloff and Flannery, 1992;  
60 Thelander et al., 2003; Drewitt and Langston, 2006; Hötter et al., 2006; De Lucas et al., 2008;  
61 Smallwood et al., 2009; Ferrer et al., 2012; Zimmerling et al., 2013; Erickson et al., 2014; Everaert,  
62 2014).

63

64 Wind farms may also act as sources of disturbance for birds, such that the vicinity of wind farm  
65 infrastructures may be avoided (Langston and Pullan, 2003), causing the displacement of local fauna  
66 from the surroundings of wind farms, and consequently may reduce their reproduction and/or  
67 survival rates (De Lucas et al., 2004; Kuvlesky et al., 2007; Larsen and Guillemette, 2007). Despite  
68 awareness of such indirect effects, they have been rarely evaluated because of, on the one hand, the  
69 difficulty in their detection and, on the other hand, the shortage of standardized protocols that  
70 complicates comparison between studies (Garvin et al., 2011; Hale et al., 2014). Anderson et al.  
71 (1999) outlined potential study designs to encourage common standards across studies investigating  
72 wind farm impacts on birds. The Before-After design (BA) evaluates an impact of a target event by

73 comparing conditions before and after the target event appeared using the same protocols in the  
74 same study area (Anderson et al., 1999). This approach usually entails difficulties for research on  
75 wind farm impacts because it requires repeated measures across several years before and after the  
76 installation of a wind farm, and these requirements are unusually met (Gove et al., 2013). The  
77 Impact-Gradient (IG) design quantifies the impact of a target event by measuring specific biological  
78 traits across a distance gradient from the target event location (Anderson et al., 1999).

79

80 The Montagu's harrier is a migratory medium-sized raptor that breeds in small colonies widespread  
81 in the Palearctic (Cramp and Simmons, 1980). Montagu's harrier has been declining globally in recent  
82 decades primarily due to agricultural intensification, especially in Western Europe (BirdLife  
83 International, 2015). In Spain, this species broadly shows a high spatial coincidence between its  
84 breeding distribution and wind farms (Tellería, 2009) but, as far as we know, there is an absence of  
85 studies examining wind farm impacts on Montagu's harrier. Evaluations of wind farm impacts on a  
86 closely related species, the hen harrier *Circus cyaneus*, have reported effects of small-scale  
87 displacement of foraging birds from around wind turbines (Johnson et al., 2000; Madders and  
88 Whitfield, 2006; Whitfield and Madders, 2006a). Hen harriers typically forage flying close to the  
89 ground and below the sweep of turbine blades (Whitfield and Madders, 2006b), and so collision  
90 victims are rarely found, even when foraging activity may be high (Smallwood and Thelander, 2005;  
91 Whitfield and Madders, 2006a; Smallwood and Karas, 2008; Garvin et al., 2011). However, most  
92 studies have involved foraging hen harriers, and turbines near breeding locations may incur a greater  
93 risk of collision, through acrobatic display flights that occur higher above the ground early in the  
94 breeding season around prospective breeding sites. Montagu's harrier has a pre-breeding display  
95 similar to the hen harrier.

96

97 We used a long term monitoring data set of a Montagu's harrier breeding population to evaluate its  
98 persistence and distribution during a time when several wind farms were installed in the same area

99 of southern Spain. We conducted a BA study to analyse: (A) abundance and density of nests and  
100 colonies; and (B) nesting habitat selection through the study period. We also carried out an IG study  
101 to quantify any wind turbine impact on Montagu's harrier nest location. In addition, we developed a  
102 multivariate model to examine the spatial variables that best explained nest locations after wind  
103 farm construction. Lastly, we also estimated the number and timing of Montagu's harriers killed by  
104 collision with turbine blades.

105

## 106 **Methodology**

107

### 108 Study area and wind farms

109

110 The study was conducted in Tarifa (Cadiz, Andalusia, southern Spain), close to the Strait of Gibraltar.  
111 This area is considered the main migratory passage for western Palearctic bird species between  
112 European breeding areas and African wintering quarters (Bernis, 1980). The hinterland of the Strait  
113 of Gibraltar is also considered an area with major potential for wind power development in Spain due  
114 to the strong winds that constantly lash the region (IDAE). Since 1998, 22 wind farms have been  
115 installed in the study area with their associated structures of tracks and power lines (see Table 1). A  
116 total of 342 wind turbines are arranged in North to South oriented rows within the study area (Figure  
117 1). The area is dominated by extensive crops alternated with some grazing patches; scrubland and  
118 woodland are mostly restricted to mountain ridges. The area is predominantly flat without steep  
119 slopes, situated 10 – 150 m above sea level (a.s.l.), with some hills in the near vicinity that do not  
120 exceed 430 m a.s.l.

121

### 122 Study species

123

124 The Montagu's harrier is a ground-nesting species that traditionally breeds in lowland heaths, dunes,  
125 hay-meadows, and steppes. However, habitat loss and land-use changes derived from agricultural  
126 intensification in recent decades has caused an increase in Montagu's harriers breeding in arable  
127 crops (Arroyo et al., 2004). It has been estimated that 70-90% of breeding pairs in western Europe  
128 breed in agricultural habitats since the 1990s, with cereal crops being the most used nesting habitat  
129 (Cramp and Simmons, 1980). Several conservation problems of the Montagu's harrier have been  
130 described, with the overlap between breeding season and cereal harvesting one of the major causes  
131 of mortality due to the high rate of nest destruction (Arroyo et al., 2002). The breeding season of  
132 Montagu's harrier in southern Spain predominantly extends from late March to the beginning of  
133 June.

134

#### 135 Study design

136

137 The study was carried out during Montagu's harrier breeding seasons in two periods regarding wind  
138 farm construction in the study area: pre-construction period (years 1995 and 2002) and post-  
139 construction period (years 2008, 2009, and 2010). The first wind farm of the study area was already  
140 installed by 2002 (see Table 1), although this year was considered belonging to pre-construction  
141 period in order to increase the sample size. We did this only after confirming that there were not any  
142 statistical differences (Mann-Whitney tests,  $p > 0.07$ ) between years 1995 and 2002 in distances from  
143 Montagu's harrier nest to closest elements of that wind farm (turbines, power lines, and tracks), or  
144 future elements in the case of year 1995 because it was not installed yet.

145

146 The study area was visited every study year during the pre-laying period to detect potential  
147 Montagu's harrier breeders. If no birds were seen, those areas were considered not occupied. All  
148 areas were visited again at least once during the nestling and post-fledging periods, to ensure that no  
149 nests had gone undetected. Observations were made with 10 x 42 binoculars and 20-60 x 80

150 telescopes during twenty minutes repeated every two kilometres along the abundant roads and  
151 tracks that crossed the study area. A nest location of a breeding pair was determined when a female  
152 brought nest material and/or when a male provisioned with food his mate or chicks, and was  
153 registered using hand-held GPS. We distinguished three categories of habitat in which nests were  
154 found: cereal crops (mostly wheat), hay crops (mixture of herbaceous plants cropped for livestock  
155 feeding) and others (non-cereal, fallow or pastures).

156

### 157 *Nest and colony abundances and densities*

158

159 Nests were included in a single colony when the distance to the closest conspecific nest was lower  
160 than 250 m, which is the most conservative distance which adults were registered hunting according  
161 to García & Arroyo (2005). Abundances were calculated as the number of nests and colonies  
162 detected each year within the study area. Nest density was calculated by dividing the number of  
163 nests detected by the area suitable for nesting each year in the study area. We considered areas  
164 suitable for nesting to be areas of non-irrigated crops and pastures, which Montagu's harrier pairs  
165 use to nest, and we excluded areas occupied by irrigated crops, water streams, woodland or  
166 scrubland, and human-made structures within the study area. We estimated the reduction suffered  
167 by the area suitable for nesting due to wind farm installation each year by subtracting those areas  
168 occupied by wind turbines, power lines and tracks from the total area suitable for nesting within the  
169 study area using Geographic Information Systems (GIS). We compared nest abundance, density, and  
170 distance to the closest conspecific nest between pre- and post-construction period using the BA  
171 analyses.

172

### 173 *Nesting habitat selection*

174

175 We entered the location of each nest (and its associated habitat) into the GIS, then we calculated the  
176 distance from nest to the closest different elements we anticipated may affect their location: water  
177 streams, woodland or scrubland, buildings (isolated constructions), villages, roads (paved ways),  
178 tracks (unpaved ways), wind turbines and power lines (or future tracks, wind turbines and power  
179 lines in year 1995). We obtained two topographical characteristics (altitude above sea level and  
180 slope) of nest locations from a 10 m-resolution DEM of Andalusia (Agriculture Department, Junta de  
181 Andalucía, 2010-2011).

182

183 We conducted the BA analysis to compare distance from nests to the closest elements of wind farms  
184 and associated structures (wind turbines, power lines, and tracks) between pre- and post-  
185 construction periods. We performed the IG analysis to calculate the avoidance distance from nest to  
186 the closest wind turbine, power line, and track in the post-construction period. The avoidance  
187 distance was considered as the median distance from nests to each of these structures, that is the  
188 distance at which 50 % of nests were excluded from wind farms and associated structures in the  
189 post-construction period (adapted from Larsen and Madsen, 2000).

190

191 We compared locations of nest and random points in order to reveal the spatial variables explaining  
192 nest location after wind farm construction. Using GIS, we generated the same number of random  
193 points within the study area as for the number of nests detected in each year of the post-  
194 construction period. Because of the clumped spatial distribution of nests, first we generated random  
195 points in the same number of colonies detected each year and used them as a centre of simulated  
196 colonies whose circular area was the mean area occupied by real colonies (49 ha, n = 14). Then, the  
197 “true” random points used in analysis were generated inside such simulated colonies. All distance  
198 variables measured for nest locations were also obtained for these random points (see the beginning  
199 of this section). Habitat of the random points was recorded by visiting their spatial coordinates  
200 obtained from the GIS and asking farmers what they cultivated there in the corresponding year. We



201 obtained such habitat information for 48 of the total of 74 random points, discarding from the  
202 analysis those random points for which we could not obtain habitat information.

203

204 All spatial analysis and distance measurements were performed using ArcGIS 9.3 (ESRI, Redlands,  
205 USA) and aerial photographs of the study area, which are available from the regional government  
206 (<http://www.juntadeandalucia.es/medioambiente>).

207

### 208 *Fatality collisions with turbines*

209

210 Mortality data was recorded at every wind turbine in a daily basis during the breeding, dispersal and  
211 migratory periods of the Montagu's harrier (March–September) after wind farms were installed in  
212 the study area. The trained observers were evenly distributed throughout the area covered by the  
213 wind farms every day from dawn to dusk, thoroughly searching an area of 50 m on both sides along  
214 wind turbine strings in fixed routes on foot (Morrison and Sinclair, 1998). No carcass disappearance  
215 experiments were conducted because observers were looking for mid- and large-sized bird carcasses  
216 that have a long persistence in the region (De Lucas et al., 2012). Data gathered consisted of age, sex,  
217 distance to the closest wind turbine, and date (for more details about carcass searches see Ferrer et  
218 al., 2012).

219

### 220 Statistical methods

221

222 We used one-way ANOVA and Mann-Whitney tests to compare variables normally and non-normally  
223 distributed under BA design. Some variables were transformed (logarithmically or square-rooted) to  
224 accord to a normal distribution for parametric statistical assumptions. A generalized linear model  
225 (GLM) was fitted to a binary response variable (0 = "random point", 1 = "nest") to model the  
226 probability of Montagu's harrier nest occurrence regarding the distances to several elements and

227 habitat. We used a binomial distribution of errors and a logit link function. We included the distances  
228 to several closest elements (see *nesting habitat selection* section) as continuous predictors and  
229 habitat as a categorical predictor with three levels (see above). The model was fitted following a  
230 backward-stepwise procedure, removing one-by-one non-significant variables until all remaining  
231 variables were significant. To assess the predictive success of the final model we used receiver  
232 operating characteristics (ROC) plots, which depict on the y-axis sensitivity (true positive rate of the  
233 model confusion matrix) against 1 – specificity (true negative rate) on the x-axis for all possible  
234 thresholds (Fielding and Bell, 1997). The threshold acts as a limit above and below which the  
235 responses are considered as one or the other alternative of the binomial model. Then, the area  
236 under the curve (AUC) was obtained as a measure of the model accuracy. AUC ranges from 0.5,  
237 corresponding with the ROC plot diagonal of a random performance, and 1.0, which represents the  
238 perfect fitted model. So, the greater AUC value, the higher explicative ability of the model.

239

240 Statistical analyses were conducted with STATISTICA version 8 and a level of 0.05 was used to  
241 evaluate significance of results. All tests were two-tailed.

242

## 243 **Results**

244

### 245 *Nest and colony abundances and densities*

246

247 A total of 111 nests of Montagu's harriers were located within the study area (Figure 1). We did not  
248 find any statistical differences in nest abundance (one-way ANOVA:  $F_{1,3} = 3.89$ ,  $p = 0.14$ ), colony  
249 abundance ( $F_{1,3} = 2.14$ ,  $p = 0.24$ ) or the number of nests per colony ( $F_{1,3} = 1.81$ ,  $p = 0.19$ ) between  
250 pre- and post-construction periods (Table 2).

251

252 The area suitable for nesting did not decrease significantly from pre to post-construction period ( $F_{1,3}$   
253 = -9.03,  $p = 0.07$ ; mean  $\pm$  standard deviation:  $\text{area}_{\text{pre}} = 59.41 \pm 0.14 \text{ km}^2$ ,  $\text{area}_{\text{post}} = 58.51 \pm 0.01 \text{ km}^2$ ).  
254 Nest density did not show statistical differences between the two study periods ( $F_{1,3} = 2.63$ ,  $p = 0.11$ ;  
255  $\text{density}_{\text{pre}} = 0.31 \pm 0.01 \text{ nests/km}^2$ ,  $\text{density}_{\text{post}} = 0.42 \pm 0.07 \text{ nests/km}^2$ ). Distance to the closest  
256 conspecific nest decreased significantly from pre to post-construction periods ( $F_{1,109} = 13.54$ ,  $p <$   
257  $0.001$ ;  $\text{distance}_{\text{pre}} = 253 \pm 315 \text{ m}$ ,  $\text{distance}_{\text{post}} = 181 \pm 560 \text{ m}$ ) (Figure 2).

258

### 259 *Nesting habitat selection*

260

261 We found statistical differences between the two periods in the distances from nests to the closest  
262 power line ( $Z = 5.04$ ,  $p < 0.001$ ; mean  $\pm$  standard deviation:  $\text{distance}_{\text{pre}} = 1471 \pm 1053 \text{ m}$ ,  $\text{distance}_{\text{post}}$   
263  $= 580 \pm 337 \text{ m}$ ) and to the closest wind turbine ( $F_{1,109} = 23.69$ ,  $p < 0.001$ ;  $\text{distance}_{\text{pre}} = 1621 \pm 869 \text{ m}$ ,  
264  $\text{distance}_{\text{post}} = 893 \pm 546 \text{ m}$ ). However, we did not find any statistical differences between pre and  
265 post-construction periods in the distance from nests to the closest track ( $F_{1,109} = 0.10$ ,  $p = 0.75$ ;  
266  $\text{distance}_{\text{pre}} = 404 \pm 265 \text{ m}$ ,  $\text{distance}_{\text{post}} = 417 \pm 196 \text{ m}$ ). The IG model estimated the avoidance  
267 distance for wind turbines at 655 m (first and third quartiles: 385 – 1414 m, range 122 – 2000 m), for  
268 power lines at 657 m (235 – 860 m, range 4 – 1156 m), and for tracks at 426 m (260 – 547 m, range  
269 73 – 936 m) in the post-construction period ( $n = 74$ ).

270

271 We built the GLM with data from 74 nests detected in the post-construction period, and 48 random  
272 points ( $N = 122$ ). The backward-stepwise fitting procedure resulted in discarding every distance and  
273 topographic variable included in the original set apart from habitat and the distance to the closest  
274 track. There was no significant relationship with distance to wind turbine. The model showed that  
275 the probability of nest occurrence increased significantly according to hay presence and significantly  
276 decreased with the presence of cereals and other less abundant habitats (mainly pastures).  
277 Additionally, the probability of nest occurrence increased significantly with distance to the closest

278 track (Table 3). The model explained 42.94% of the deviance and classified correctly 79.03% of the  
279 points (72.92% of random points and 85.13% of nest). ROC plots showed high accuracy of  
280 nest/random point classification by the proposed model, with an AUC value of 0.91 (Figure 3).

281

#### 282 *Fatality collisions with wind turbines*

283

284 A total of seven Montagu's harrier carcasses were found during the post-construction period: four  
285 birds in 2008, and three birds in 2010 (Figure 1). Thus, the mean mortality rate was  $0.007 \pm 0.006$   
286 birds/turbine/year. Carcass decomposition limited data collection of sex and age. Only four carcasses  
287 could be sexed, two were males and two were females. Three carcasses were aged as adult and two  
288 as fledged young. Regarding date, three carcasses were detected in August and the other four  
289 carcasses were detected in April, June, July and September, respectively. Carcasses were found at  
290  $45.8 \pm 59.6$  m as the mean distance ( $\pm$  standard deviation) from the closest wind turbine. The low  
291 number of carcasses prevented any statistical analysis.

292

#### 293 **Discussion**

294

295 Our results suggested that nesting decisions of Montagu's harriers was not adversely affected by the  
296 construction, operation and maintenance of wind farms, although these structures have been  
297 repeatedly considered as a disturbance source for birds (Kuvlesky et al., 2007; Smallwood and  
298 Thelander, 2004). Hötter et al. (2006) pointed out that any negative effect on bird breeding  
299 populations has yet to be verified, in agreement with our results.

300

#### 301 *Nest and colony abundances and densities*

302

303 Nest and colony abundances and nest density remained stable throughout the study period. Our  
304 results are concordant with several studies that reported no negative effects of wind farms on bird  
305 populations (De Lucas et al., 2005; Douglas et al., 2011; Hale et al., 2014; Hatchett et al., 2013),  
306 although other studies did show important negative impacts (Larsen and Guillemette, 2007; Leddy et  
307 al., 1999; Pearce-Higgins et al., 2009). Raptor abundances normally decreased during and after the  
308 construction of wind farms compared to the initial situation (Garvin et al., 2011). Stewart et al.  
309 (2007) proposed that the longer the period of wind turbine operation, the greater the decline in bird  
310 abundance due to lack of habituation to wind farms (although see Madsen and Boertmann, 2008).  
311 On the contrary, nest abundance increased, although not significantly, since wind turbines were built  
312 in our study area. Drewitt & Langston (2006) opined that the apparent absence of wind farm  
313 displacement impacts on birds concluded by several studies may be due to the high philopatry and  
314 longevity of study individuals returning to former breeding sites. Montagu's harrier is considered as  
315 having relatively low breeding philopatry (Arroyo et al., 2004), so this hypothesis can be discarded as  
316 an explanation of our findings. Thus, the relatively stable density of pairs over time in our study  
317 would in large part be due to a high turnover with new recruits establishing as breeders in the study  
318 area, and this would not support the existence of disturbances caused by wind farms that were  
319 overridden by strong breeding site fidelity.

320

321 Nevertheless, while wind farm installation did not seem to affect nest abundance and density, a  
322 change in spatial distribution of nests took place. Breeding pairs tended to nest closer together after  
323 wind farms were installed (see Figure 1). The desertion of the most distant colonies in the west and  
324 southwest of the study area before wind farm installation and the establishment and aggregation of  
325 colonies in the north of the study area after wind farm installation resulted in the observed reduction  
326 of distances between conspecific nests. The south-western colony disappeared during the pre-  
327 construction period, so at least for this colony we do know that the installation of a wind farm in its  
328 surroundings was not the cause of its desertion by breeding pairs.

329

330 *Nesting habitat selection*

331

332 BA analysis of distances from nests to closest wind turbine and power line differed between pre- and  
333 post-construction periods, both showing higher values in the pre-construction period. Breeding pairs  
334 located their nests closer to wind turbines and power lines when all wind farms were installed in the  
335 study area than when only one wind farm was present. On the one hand, this result may simply be  
336 due to chance, or that more powerful influences (distribution of preferred nesting habitat) were at  
337 play. On the other hand, does support the idea that Montagu's harriers were not avoiding the  
338 surroundings of wind turbines to nest, contrary to results of previous studies (e.g. Pearce-Higgins et  
339 al., 2009). However, the distance from nest to closest track did not vary through the study period  
340 despite the greater length of tracks installed after wind farm construction. This suggest that tracks  
341 could have some effect on Montagu's harrier decisions on nest location, and this may explain why  
342 nest sites were more concentrated at the north part of the study area after wind farms installation,  
343 where wind farm (and associated track) density was lowest.

344

345 The IG model showed higher avoidance distances for wind turbines in comparison to other studies  
346 on the closely related hen harrier (Madders and Whitfield, 2006; Whitfield and Madders, 2006a;  
347 Pearce-Higgins et al., 2009), although those studies estimated avoidance distances from bird flight  
348 activity, mainly foraging, rather than nest location. Therefore, we probably overestimated the real  
349 avoidance distance for wind turbines of Montagu's harriers as they might forage closer to these  
350 structures than where they locate nests. Anyhow, observed avoidance distances for wind farms,  
351 especially for wind turbines and power lines (on average up to 650 m), were high and could suggest  
352 some effect on harrier nesting decisions. However, reduced nesting densities of Montagu's harriers  
353 in our study ( $< 0.5$  nest/km<sup>2</sup>), and extensively of raptors as top predator of terrestrial ecosystems  
354 (Newton, 1979), seem more plausible to explain the results than an avoidance behaviour for these

355 structures of nesting harriers. Notably since our BA analysis found that harriers tended to nest  
356 slightly closer to turbine locations after their installation (BA being usually more powerful as an  
357 investigative tool than IG: Anderson et al., 1999) and the distance to closest turbine was not an  
358 important predictor defining nest locations. Our IG analysis, moreover, could not account for the  
359 availability of suitable habitat with varying distance to turbines and other wind farm features: the  
360 availability of one habitat (hay fields) appeared to be a major factor in nest site selection according to  
361 additional analyses. It is also possible that fundamental preferences for suitable locations of turbines  
362 and nests differed between wind farm developers and Montagu's harriers, respectively, at a scale  
363 that we could not attribute. For example, developers likely prefer turbine locations where wind  
364 strengths are higher but nesting harriers likely prefer more sheltered locations where vegetation  
365 growth is lush, and so taller. Overall, our study does show that developers' selection of turbine  
366 locations did not apparently interfere with nest site choices of Montagu's harriers, even though some  
367 suitable harrier nesting areas were within a few hundred metres of turbines. The IG results, thanks to  
368 the wider temporal context provided by our additional BA studies do not necessarily indicate, as  
369 could be interpreted on face value, that nesting harriers 'avoid' turbines within the derived range of  
370 values. Our study thereby provides an instructive example of the potential danger in placing too  
371 much emphasis on simplistic and singular post-construction analyses. Nevertheless, our study was  
372 restricted to nesting decisions and potential changes in breeding numbers: further studies on  
373 foraging behaviours and home ranges of breeding Montagu's harriers would, therefore, be useful in  
374 the environs of operating wind farms.

375

376 The GLM fitted to the binary response variable (random points/nests) was relatively robust. Habitat  
377 and distance to closest track were the only variables with significant effects on nest occurrence after  
378 wind farm constructions. The high predictive ability of habitat included in the model suggests that  
379 this species primarily considers a preferred habitat to locate nests rather than the distances to other  
380 features which may act as sources of disturbance. The model highlighted a preference for nesting in

381 hay crops and avoidance of cereal crops, in disagreement with previously stated preference of  
382 cereal crops for nesting (Cramp and Simmons, 1980). However, it has been reported elsewhere that  
383 in areas where hay fields are available they are used in greater proportion than expected (Barros and  
384 Benítez, 1995). Thus, the negative effect of cereal crops on the probability of nest occurrence may be  
385 explained by this preference in Montagu's harrier nesting habitat selection in our study area. The  
386 third type of habitat (mainly pastures) also had a negative effect on nest location. Montagu's harrier  
387 breeding pairs tend to nest in areas dominated by plants characterized by fast growth in height that  
388 act as a defence against predators by hiding nests (Limiñana et al., 2006). However, pastures were  
389 often heavily grazed by livestock in our study area, restricting growth of graminoid and herbaceous  
390 plants. Consequently, pastures would not reach sufficient height to protect Montagu's harrier nests  
391 and so breeding pairs should avoid this habitat, as suggested by our model. The distance from nest to  
392 closest track arose as an important predictor with a negative effect on the probability of nest  
393 occurrence. It has been demonstrated that noise and vibrations from tracks and roads are important  
394 negative effects on raptors and other birds (Benítez-López et al., 2010; Martínez-Abraín et al., 2010).  
395 As a ground-nesting species, the Montagu's harrier can be specially affected by vibrations caused by  
396 vehicle users of tracks. This may explain why the distance from nest to closest tracks remained  
397 constant over the study period, contrary to what we found with wind turbines and power lines, and  
398 the probability of nest occurrence increase with that distance (see BA results). Although agricultural  
399 habitat changes were not detected at greater spatial scale (province of Cadiz,  
400 [www.juntadeandalucia.es/agriculturaypesca/portal/servicios/estadisticas/agrarias/superficies-y-  
402 producciones.html](http://www.juntadeandalucia.es/agriculturaypesca/portal/servicios/estadisticas/agrarias/superficies-y-<br/>401 producciones.html)), we cannot reject that some habitat changes would have happen in our study  
403 area that, in conjunction with the proliferation of tracks associated to wind farms, would have  
404 caused the observed displacement and aggregation of nesting Montagu's harrier throughout the  
405 study period.

405

406 *Fatality collisions with wind turbines*



407

408 The mortality rate reported per wind turbine and year in our study agreed with avian collision rates  
409 recorded in the literature, that show a wide variation from 0.01 to 23 birds/turbine/year (Drewitt  
410 and Langston, 2006). Fatality rate we obtained was relatively low (0.007 bird/turbine/year), like most  
411 assessment of wind farm impacts on birds (De Lucas et al., 2008). A weakness of our results is that  
412 they lacked any measure of searching efficiency due to the vegetation height during some periods of  
413 the year (Smallwood, 2013). Nonetheless, we were confident that the trained observers had  
414 experience and time to account for variation in vegetation height; and the time interval between  
415 searching events was only one day, increasing detection probability.

416

417 The Montagu's harrier is characterized by high aerial agility and manoeuvrability, conferring lower  
418 risk to collision with wind turbines than larger raptors (Brown et al., 1992; Lucas et al., 2012). Hen  
419 harrier, a closely related species to Montagu's harrier, tends to fly at low altitudes, with most aerial  
420 activity occurring less than 20 m above the ground (Whitfield and Madders, 2006b). Montagu's  
421 harrier has very similar flight behaviour. The heights of the turbine blades in our study ranged  
422 between 27 and 65 m from the ground to the highest point, except one wind farm whose turbines  
423 had blades with a lowest sweep only 15 m high. Thus, flight behaviour of harriers typically avoids  
424 collisions with blades and this could be the reason for the low mortality rate we found. We did not  
425 find any concentration of collision victims at the beginning of the breeding season when pre-  
426 breeding birds will have been conducting elaborate aerial displays higher above the ground (rather  
427 casualties tended to occur towards the end of the breeding season). This suggests that such displays  
428 may not carry a risk of collision greater than other forms of flight behaviour.

429

430 **Conclusion**

431

432 Our results suggest that wind farms did not cause adverse negative effects on nesting decisions of  
433 Montagu's harriers in our study area, contrary to what has been described for other bird species. The  
434 abundance and densities of Montagu's harrier nests remained constant through the study period  
435 although a change in nest spatial distribution occurred, from being more widespread before wind  
436 farm construction to being more aggregated after construction. We could detect no population-wide  
437 consequence of this apparent increased aggregation, however. Breeding pairs seemed to decide  
438 where to build their nest mostly on suitable vegetation availability instead of distances from  
439 potential disturbance sources, notably wind turbines. Only distance to tracks seemed to be  
440 influential in nest location decisions of breeding pairs, presumably because of vibrations caused by  
441 vehicular use of tracks. Habitat changes through the study period within the study area might be the  
442 main driver of nest aggregation, although birds tend to choose nest sites at a particular distance from  
443 tracks. Nevertheless, we encourage further studies of reproductive parameters and foraging  
444 behaviour of Montagu's harriers with the aim to provide supplementary research that we could not  
445 address and examine further our findings of no negative effect of wind farms and associated  
446 structures on this species.

447

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455 harrier population.

456

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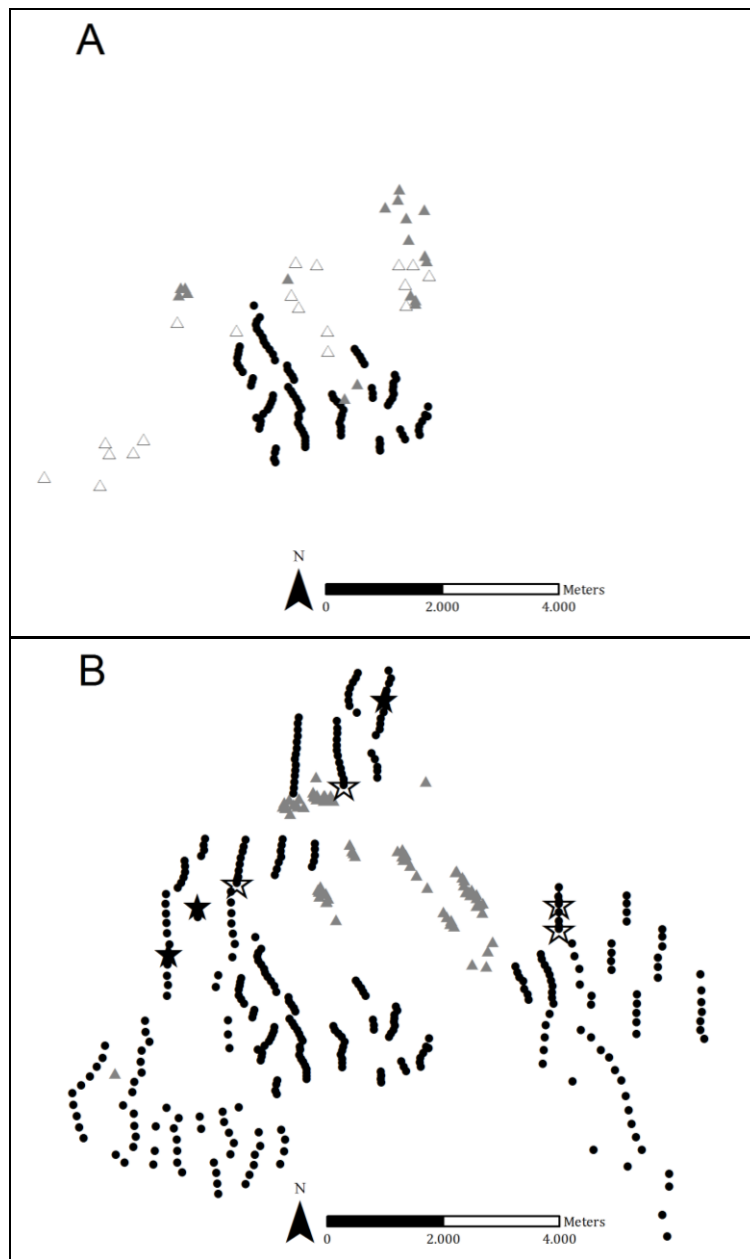
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656 **Figure 1:** Spatial distribution of wind farms (circles), Montagu's harrier nests (triangles), and bird

657 carcasses (stars) in the study area during the pre-construction period (panel A) and during the post-

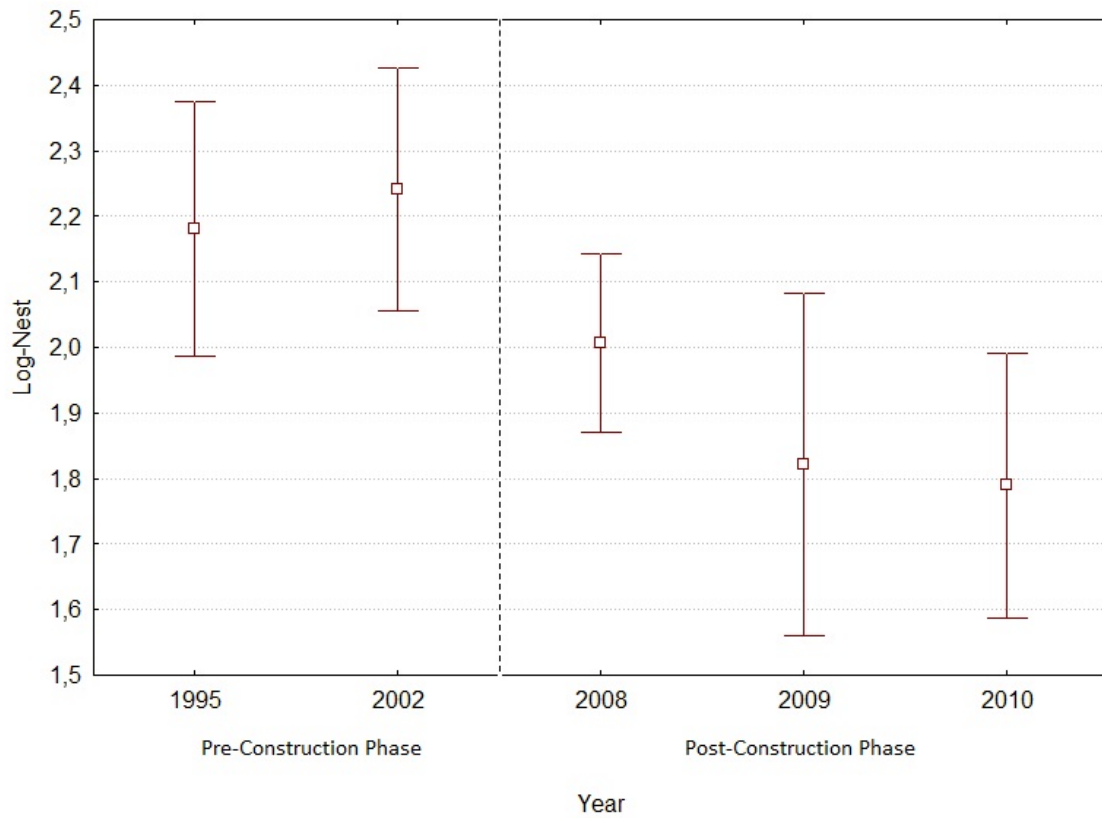
658 construction period (panel B). In panel A, open and grey triangles represent Montagu's harrier nest

659 locations detected in years 1995 and 2002, respectively. In panel B, white stars represent the closest

660 wind turbine from Montagu's harrier carcass locations. Because we did not obtain the spatial

661 coordinates of three of the seven carcasses, only the wind farm where they were detected, we

662 represent the wind turbine situated in a central point within those wind farms with a black star.

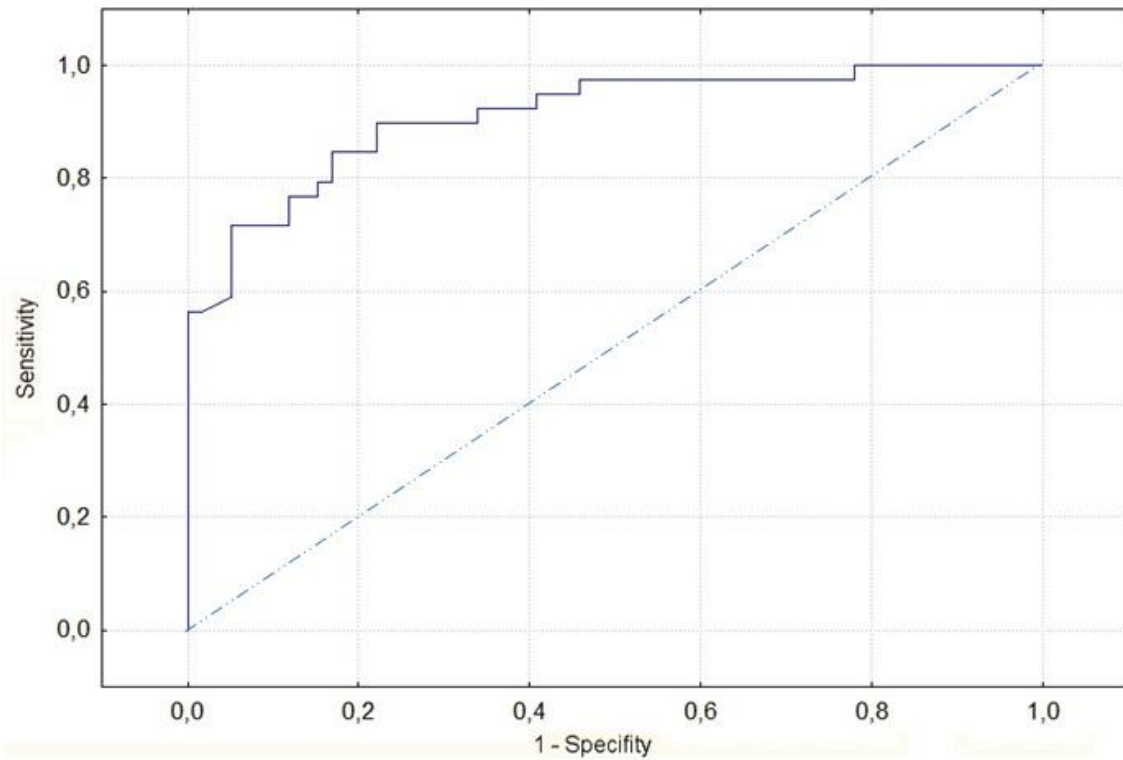


663

664 **Figure 2.** Distance from nest to closest conspecific nest throughout the study period. Mean distances  
 665 and standard deviation per year were plotted because of the greater clarity provided. The dashed  
 666 line separates the two defined periods of the study.

667





668

669 **Figure 3.** ROC plot derived from the fitted binomial GLM. The solid line represents the values of  
670 sensitivity and 1-specificity for all possible threshold of the proposed model. The dotted line  
671 represents the values of a random performance model, which are not able to discriminate between  
672 the two alternatives of the logistic model.

673

674

675 **Table 1.** Wind farms installed in the study area along the study period. We specified information  
 676 about year of operating starting, number of wind turbines, power per wind turbine, hub height, and  
 677 the diameter of the rotor of each wind farm.

<b>Wind farm</b>	<b>Year</b>	<b>Nº turbines</b>	<b>Power turbine (kW)</b>	<b>Hub height (m)</b>	<b>Rotor diameter (m)</b>
WF 1	1998	100	300	30	30
WF 2	2005	11	2000	67-100	87
WF 3	2005	17	800	50-60	59
WF 4	2005	30	800	50-60	59
WF 5	2005	20	800	50-60	56
WF 6	2005	28	1670	70-80	80
WF 7	2005	15	800	50-60	59
WF 8	2005	6	1670	60-80	74
WF 9	2005	16	800	50-60	56
WF 10	2006	4	1500	80-95	72
WF 11	2007	10	2000	95-125	90
WF 12	2007	10	2000	67-100	87
WF 13	2007	7	1800/2000	80-105/95-125	90
WF 14	2007	4	1800/2000	80-105/95-125	90
WF 15	2007	6	2000	60-100	80
WF 16	2007	8	2000	67-100	87
WF 17	2007	6	2000	67-100	87
WF 18	2007	10	2000	95-125	90
WF 19	2007	10	2000	65	70.6
WF 20	2007	8	1800/2000	80-105/95-125	90
WF 21	2007	10	2000	65	70.6
WF 22	2008	6	2000	60-100	80

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686 **Table 2.** Nest and colony abundances and average number of nests per colony during the study  
 687 period.

688

689 **Pre-Construction period**

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Number of Nests		Number of Colonies		Average Number of Nests/Colony	
1995	2002	1995	2002	1995	2002
19	18	6	5	3.1	3.6

694

695 **Post-Construction period**

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Number of Nests			Number of Colonies			Average Number of Nests/Colony		
2008	2009	2010	2008	2009	2010	2008	2009	2010
28	20	26	5	4	5	5.6	5.0	5.2

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701

702

703 **Table 3.** Parameters of the fitted binomial GLM. The intercept included the effect of cereal plots  
704 within habitat variable. Sample size = 122 points.

705

706

707		$\beta$	Standard Error	Wald Statistic	p-value
708	Intercept	-1.38	0.58	5.68	0.02
709	Distance-Track	0.003	0.001	6.52	0.01
710	Non-cereal plots	-1.82	0.47	15.16	<0.001
711	Hay plots	2.21	0.38	34.69	<0.001

712

713 Null Deviance: 163.54 with 121 degrees of freedom.

714 Residual Deviance: 93.31 with 118 degrees of freedom.