



i-sp pollen-transport network for the species Xylocopa violacea $143 x 186 mm \ (300 \ x \ 300 \ DPI)$

1	
2	
3	
4	Downscaling pollen-transport networks
5	to the level of individuals
6	
7	Cristina Tur ^{1*} , Beatriz Vigalondo ¹ , Kristian Trøjelsgaard ² , Jens M. Olesen ²
8	and Anna Traveset ¹
9	
10	¹ IMEDEA- Institut Mediterrani d'Estudis Avançats (CSIC-UIB), Miquel Marqués 21,
11	07190 Esporles, Illes Balears (Spain)
12	² Department of Bioscience, Aarhus University, Ny Munkegade 114, DK-8000 Aarhus
13	C (Denmark)
14	
15	*Corresponding author: Cristina Tur. IMEDEA- Institut Mediterrani d'Estudis
16	Avançats (CSIC-UIB), Miquel Marqués 21, 07190 Esporles, Illes Balears (Spain). E-
17	mail: cris.tur.espinosa@gmail.com; Tel: +34 971 611719; Fax: +34 971 611761.
18	
19	Running headline: Downscaling from species to individuals.
20	

21 Summary

22

1. Most plant-pollinator network studies are conducted at species level whereas little is known about network patterns at the individual level. In fact, nodes in traditional species-based interaction networks are aggregates of individuals establishing the actual links observed in nature. Thus, emergent properties of interaction networks might be the result of mechanisms acting at the individual level.

28

29 2. Pollen loads carried by insect flower-visitors from two mountain communities were 30 studied to construct pollen-transport networks. For the first time, these community-wide 31 pollen-transport networks were downscaled from species-species (*sp-sp*) to individuals-32 species (*i-sp*) in order to explore specialization, network patterns and niche variation at 33 both interacting levels. We used a null model approach to account for network size 34 differences inherent to the downscaling process. Specifically, our objectives were: (i) to 35 investigate whether network structure changes with downscaling, (ii) to evaluate the 36 incidence and magnitude of individual specialization in pollen use, and (iii) to identify 37 potential ecological factors influencing the observed degree of individual specialization. 38

39 3. Network downscaling revealed a high specialization of pollinator individuals, which 40 was masked and unexplored in *sp-sp* networks. The average number of interactions per 41 node, connectance, interaction diversity and degree of nestedness decreased in *i-sp* 42 networks, because generalized pollinator species were composed of specialized and 43 idiosyncratic conspecific individuals. An analysis with 21 pollinator species 44 representative of two communities showed that mean individual pollen resource niche 45 was only c. 46% of the total species niche.

47 4. The degree of individual specialization was associated to inter- and intraspecific
48 overlap in pollen use and it was higher for abundant than for rare species. Such niche
49 heterogeneity depends on individual differences in foraging behaviour and likely has
50 implications for community dynamics and species stability.

51

52 **5.** Our findings highlight the importance of taking inter-individual variation into 53 account when studying higher-order structures such as interaction networks. We argue 54 that exploring individual-based networks will improve our understanding of species-55 based networks and will enhance the link between network analysis, foraging theory 56 and evolutionary biology.

57

58 Key-words: species-based networks, individual-based networks, individual
59 specialization, ecology of individuals, linkage level, niche overlap, pollen-load analysis,
60 foraging behaviour, resource partition, generalization

62 Introduction

63

64 In the last couple of decades, community-studies of plant-animal interactions have been 65 explored with the aid of network theory (e.g. Memmott 1999; Jordano et al. 2003). In 66 such interaction-networks, animal and plant species are nodes, and links represent the 67 interactions between them. As each node is a different species, these networks are 68 species-based. However, each node in a species-based network consists of a population 69 of conspecific individuals, which are the true interacting actors in nature. Operating 70 exclusively at species level may obscure individual behaviour, resulting in loss of 71 valuable information (Ings et al. 2009). The relevance of scaling down species-based 72 pollination networks to *individual-based* networks has previously been stressed (e.g. 73 Olesen et al. 2010; Dupont et al. 2011), as patterns and forces acting at the individual 74 level may be important drivers of structure and dynamics at species level. Moreover, 75 relative invariance of network patterns and lack of differences found among different species-based networks (e.g. Jordano et al. 2006; Petanidou et al. 2008) could be a 76 77 consequence of not resolving networks at the proper scale.

78 Despite the recognized importance of individual variation within natural 79 populations for many ecological processes (Bolnick et al. 2011; Dall et al. 2012; Sih et 80 al. 2012; Wolf & Weissing 2012), only a few empirical studies to date have applied 81 network theory as a tool for exploring ecological interactions at the individual level. For 82 example, individual-based networks have been used to study intrapopulation patterns of 83 resource partition (Araújo et al. 2010; Pires et al. 2011; Tinker et al. 2012), changes in 84 foraging preferences at increasing levels of intraspecific competition (Araújo et al. 85 2008), body size effects in prey-predator interactions (Woodward & Warren 2007; 86 Yvon-Durocher et al. 2010), patterns in roosting sites of bats (Fortuna et al. 2009) and disease dynamics (Perkins *et al.* 2009). In the case of pollination networks, the number of individual-based network studies is very scarce, all focusing on one or two species (Fortuna *et al.* 2008; Dupont *et al.* 2011; Gómez *et al.* 2011, 2012). A likely explanation for the scarcity of such studies is the labour-intensive sampling required to resolve community-wide species networks into individual-based networks covering all species present.

93 Network data at individual level may be used to test important niche breadth 94 questions, and this is a natural progressional step and promising avenue for future 95 network research (Ings et al. 2009). It is well known that conspecific individuals vary in 96 their resource use (Van Valen 1965; Roughgarden 1972). Individual specialization 97 occurs when individuals exploit only a small subset of the total resources used by the 98 entire population, and it is a ubiquitous phenomenon in both vertebrate and invertebrate 99 populations (Bolnick et al. 2003). As individuals within a population vary genetically 100 and phenotypically, their resource choice may differ as well. For example, physiology, 101 criteria of optimal diet shifting, behavioural skills or social status (see Araújo, Bolnick 102 & Layman 2011 for a review on the ecological causes of individual specialization), all 103 influence individual resource use and preferences. Several indices have been developed 104 to quantify the degree of individual specialization (Bolnick et al. 2002), allowing 105 researchers to compare the magnitude of individual specialization across a variety of 106 ecological situations in nature. Indeed, intrapopulation variation in resource use has 107 been proposed to affect population dynamics and ecological interactions (Bolnick et al. 108 2011), whereas these in turn affect the magnitude and incidence of intrapopulation niche 109 variation in a feedback loop way (Araújo, Bolnick & Layman 2011).

Pollination networks are usually built based on field observations of plant andflower-visitor interactions. However, some studies have used data from pollen loads

112 carried by flower visitors (e.g. Bosch et al. 2009; Alarcón 2010) finding that they give 113 complementary information. The analysis of pollen loads, which provides a record of 114 individual foraging patterns across time, addresses one component of pollinator 115 effectiveness and actually gives a good indication of probable pollinators of a given 116 plant species (Popic, Wardle & Davila 2013). Construction of pollen-transport networks 117 has also been useful in applied studies (e.g. Forup & Memmott 2005; Gibson et al. 118 2006; Lopezaraiza-Mikel et al. 2007). Here we used pollen-loads from insect flower-119 visitors (pollinators hereafter) in two mountain scrublands to construct pollen-transport 120 networks at both the species level (species-species network; hereafter sp-sp), and the 121 individual level of the pollinators (*individuals-species* network; hereafter *i-sp*). Plants 122 were only analysed at species level. To our knowledge, no previous study has 123 downscaled a whole pollination network using this approach.

124 Our first goal was to investigate whether network structure changes when 125 downscaling from *sp-sp* to *i-sp* and if so, which network topological parameters change 126 in particular. The network-downscaling process inherently implies an increase in the 127 number of network nodes, because species are composed of groups of conspecific 128 individuals. Given that most network metrics are influenced by network size (Dormann 129 et al. 2009), we built null models that account for size-related differences. Changes in 130 network structure when downscaling might be expected due to differences in individual 131 pollen use resulting from different foraging patterns, individual behaviours and trait 132 variability among conspecifics. We defined niche of a pollinator species both 133 qualitatively (number of interactions to plant pollen types, i.e. linkage level) and 134 quantitatively (diversity of pollen types carried by insects). Given that linkage level of a species (L_{sp}) expresses the sum of all links established by its individuals (L_i) , we 135 136 hypothesize that pollinator species niche may be determined by means of two possible

mechanisms: (i) individuals are as generalized as their species, i.e. all individuals have similar feeding niche (null hypothesis: $L_{sp} \approx L_i$), or (ii) individuals are more specialized than the species (alternative hypothesis: $L_i << L_{sp}$). If the alternative hypothesis is true, we would predict changes in topological parameters beyond those related to size differences when downscaling from *sp-sp* to *i-sp* networks.

142 Our second goal was to evaluate the incidence and magnitude of individual 143 specialization in pollen use. First, we compared species and individual generalization 144 levels for all pollinators (i.e. population vs. individual niche). Moreover, for a subset of 145 abundant and representative pollinator species, we compared the mean empirical 146 linkage levels with those obtained from the null models and tested whether the 147 distribution of their species linkage level among conspecific individuals was nested, as 148 found previously in other studies (Araújo et al. 2010; Pires et al. 2011). In addition, we 149 measured quantitatively the relative degree of individual pollen specialization and 150 determined its significance.

151 Finally, our third goal was to identify potential ecological factors influencing 152 individual specialization in pollen use. Specifically, using structural equation modeling 153 we tested the effect of factors known to influence individual specialization from other 154 studies; (i) inter- and intraspecific overlap (competition, i.e. amount of resources shared 155 by individuals of different or same species), (ii) insect species abundance and (iii) insect 156 species phenophase, i.e. temporal extent of network membership. In the context of 157 Optimal Foraging Theory (Stephens & Krebs 1986), we expected a high degree of 158 individual specialization to be associated to: (i) low amounts of interspecific overlap, 159 because release from resource competition favours species niche expansion through 160 increased variation in resource use between individuals (Costa et al. 2008; Bolnick et al. 161 2010); (ii) high amounts of intraspecific overlap, because strong intraspecific

Journal of Animal Ecology

162 competition promotes resource use diversity among conspecifics (e.g. Svanbäck & Bolnick 2005; Svanbäck & Bolnick 2007; Araújo et al. 2008); (iii) high abundance of 163 164 species, because at high densities of foragers the availability of preferred resources 165 decreases causing individuals to add different subsets of resources (e.g. Fontaine, Collin 166 & Dajoz 2008; Tinker, Bentall & Estes 2008; Svanbäck et al. 2011; Tinker et al. 2012); 167 and (iv) long species phenophases, because species with short phenophases might be .divid. 168 composed by short-lived individuals using similar and narrower subsets of resources 169 due to temporal restriction in resource pool compared to species with long phenophases.

171 Materials and methods

172

173 STUDY SITES AND DATA COLLECTION

174 Plant-pollinator interactions were studied at two sites on the highest mountain in 175 Mallorca (Puig Major, 1445 m): (i) Sa Coma de n'Arbona (CN) at 1100 m a.s.l. 176 (39°48'05" N 2°47'9" E) and (ii) Passadis de Ses Clotades (PC) at 1400 m a.s.l. (39°48'34" N 2°47'50" E). Plant-pollinator interactions were surveyed during the entire 177 178 flowering season, from May to August 2010, on clear and calm days. Pollinator 179 censuses of 5 min, focusing upon randomly selected plant individuals of every species 180 in bloom, were carried out between 10 am and 5 pm three times a week. During each 181 census we recorded: (i) taxonomic identity of plant species under observation; (ii) 182 taxonomic identity of insect pollinators (species name if possible or morphospecies otherwise; hereafter species), and (iii) number of individuals of each insect species 183 184 observed. After each census, flower-visiting insects were captured, placed separately in 185 clean vials and stored in a freezer for later pollen analysis and expert identification. We carried out a total of 458 censuses at CN (38 h 10 min) and 377 at PC (31 h 25 min), 186 capturing 73 and 61 insect species, respectively. At each site, abundance and 187 188 phenophase of each insect species were estimated as total number of individuals 189 observed in all censuses and total number of days between first and last observation 190 date of flower-visiting individuals, respectively.

In the laboratory, each captured insect individual was examined for pollen loads by washing, identifying and counting all pollen grains from its body surface. Pollen was identified at species, species cluster or morphospecies level (pollen types hereafter). See Appendix S1 in Supporting Information for a detailed description of the pollen load analysis procedure. At *CN*, a total of 190 individuals (71 Diptera, 83 Hymenoptera, 33 196 Coleoptera, 3 Hemiptera) were examined for pollen, and 55 pollen types were 197 identified. At *PC*, a total of 137 individuals (43 Diptera, 64 Hymenoptera, 26 198 Coleoptera, 4 Hemiptera) carried 49 pollen types. Lepidopterans were excluded because 199 their pollen load could not be analysed according to our quantitative methodology 200 protocol (they could not be washed).

201

202 CONSTRUCTION AND ANALYSIS OF SP-SP AND I-SP POLLEN-TRANSPORT

203 NETWORKS

Data from pollen load analyses were used to construct plant-pollinator interaction matrices at two levels of resolution: (i) species-species (*sp-sp* network), representing interactions between insect species and pollen types and (ii) individuals-species (*i-sp* network), representing interactions between insect individuals and pollen types. Interaction weight was the number of pollen grains of a given type carried by either the insect species or the individual.

210 We selected the following parameters to describe *sp-sp* and *i-sp* network 211 structure: number of pollinator nodes (A), number of pollen type nodes (P), total 212 number of nodes (T), total number of interactions (I), linkage level (L), network size 213 (N), linkage density (LD), connectance (C), nestedness (NODF), interaction diversity 214 (H_2) and interaction evenness (E_2) (see Table S1 in Supporting Information for 215 definitions). The significance of the NODF metric was assessed against 1,000 216 randomizations using the fixed row and column totals constrained model, i.e. node 217 linkage level was fixed. We calculated the probability of randomly obtaining higher 218 NODF values than that of the empirical matrix with a one-tailed Z-test. All network 219 metrics were obtained with the *bipartite* (version 1.17, Dormann, Gruber & Fründ 2008) and *vegan* (version 2.0-6, Oksanen *et al.* 2012) packages implemented in the *R*statistical software version 2.15.0 (*R* Development Core Team 2012).

222 To test whether network structure changes when downscaling from *sp-sp* to *i-sp*, 223 accounting for differences in network size, we built 1,000 null *i-sp* pollen-transport 224 networks of the same size and species composition as the empirical *i-sp* networks. 225 These null networks were build combining simulated *i-sp* submatrices for each species 226 generated with the Patefield algorithm (i.e. observed marginal totals maintained for 227 matrix rows and columns). Each null *i-sp* subnetwork simulated that conspecific 228 individuals act as generalists as their species, sampling each pollen type at a rate 229 proportional to the corresponding species pollen use distribution. Thus, in null *i-sp* 230 submatrices with a rows and p columns (a was number of individuals of species S and p 231 number of plant pollen types carried by species S), each individual was reassigned the 232 same pollen load as observed but pollen grains were randomly distributed among pollen 233 types with a probability equal to the observed pollen type proportion used by the 234 species. We calculated the abovementioned parameters also for the 1,000 null *i-sp* 235 networks. When parameter values of empirical *i-sp* networks did not fall into 95% 236 confidence intervals of values for the null *i-sp* networks, differences were thus 237 attributed to individual specialization and not to a network size artifact.

238

239 EVALUATION OF INCIDENCE AND MAGNITUDE OF INDIVIDUAL240 SPECIALIZATION

We compared linkage level of each insect species (L_{sp}) with those of their individuals (L_i) to explore specialization at both levels. For species with ≥ 5 sampled individuals (14 spp. at *CN*, 7 spp. at *PC*), we calculated average $\overline{L_i}$ of a given species, and compared it to the $\overline{L_i}$ obtained in null networks. We concluded that a species was

composed of individuals being significantly more specialized than the species when empirical $\overline{L_i} < 95\%$ of 1,000 null $\overline{L_i}$. Within a species, both generalist and specialist individuals were frequently found, so we evaluated the presence of a nested pattern in species linkage level partition among conspecifics. We followed the same procedure explained above for *NODF* calculation with a set of 21 *i-sp* submatrices (matrices for species with \geq 5 sampled individuals) (see Fig. 3 for an example).

251 Relative degree of individual specialization in pollen use was estimated for all those 21 species with ≥ 5 sampled individuals in each network. We followed the 252 253 intrapopulation niche width variation model suggested by Roughgarden (1972) and 254 indices developed by Bolnick et al. (2002, 2007). Total niche width (TNW) can be 255 partitioned into two components: a within-individual component (WIC) and a between*individual component (BIC)*, thus TNW = WIC + BIC. WIC is the average variance in 256 257 the range of pollen types each conspecific individual is using, i.e. the average of 258 individuals' niche breadth. BIC is the variance in mean pollen use among individuals of 259 the particular species, i.e. it represents the niche variation among individuals. Relative 260 degree of individual specialization is measured as the proportion of total niche width 261 (TNW) explained by the within-individual component, WIC/TNW, and is thus a species-262 level metric. This metric approaches 1 when individuals' niches include the full range of 263 pollen types used by their species, and tends to 0 when individuals use smaller, non-264 overlapping subsets of their species' resources. We applied Monte Carlo re-sampling 265 procedures (Bolnick et al. 2002; Araújo et al. 2008, 2010) to test whether the observed 266 individual specialization was significant, i.e. observed WIC/TNW values were less than 95% confidence interval of WIC/TNW values obtained for the 1,000 null i-sp 267 268 submatrices of each species (null hypothesis that all individuals act as generalized as the 269 species).

At the individual level, we estimated specialization by calculating a proportional similarity index (PS_i), which measures similarity in the use of pollen (diet overlap) between an individual *i* and its corresponding species *S*. All indices were obtained using *R* (version 2.15.0, *R* Development Core Team 2012), following formulae described in detail in Appendix S2.

275

276 ANALYSIS OF ECOLOGICAL FACTORS AFFECTING THE DEGREE OF277 INDIVIDUAL SPECIALIZATION

278 Interspecific overlap

279 To measure interspecific overlap, we transformed our two-mode pollen-transport 280 networks into one-mode networks depicting the pattern of shared pollen resources. By 281 definition, two-mode networks represent the pattern of interactions among two different 282 subsets of nodes (e.g. pollinators and pollen types), whereas one-mode networks 283 represent interactions among nodes of the same set (e.g. pollinators). We constructed 284 one-mode projections of the *sp-sp* two-mode network matrices at each study site 285 following the co-occurrences projection method (Opsahl 2009a; Padrón, Nogales & 286 Traveset 2011), which entails counting the number of pollen types shared among each 287 pair of different insect species. We thus obtained a square symmetric matrix with a zero 288 diagonal and with s rows and s columns, where s is the number of insect species and the 289 value in each cell w_{ii} is the number of pollen types shared between them. Thus, total 290 interspecific overlap for a species s_i can be defined as the sum of all its weighted links $\sum_{i} w_{ij}$ (degree for weighted networks in Barrat *et al.* 2004). For example, a species will 291 292 get a total interspecific overlap of six by sharing one pollen type with six different 293 insect species but also by sharing six different plant pollen types with one insect 294 species. In order to compare species among sites, we calculated a standardized measure

of interspecific overlap (species-species overlap, *SPO*), by dividing total interspecific overlap by the maximum overlap that a species can achieve in its network (i.e. when an insect species shares all pollen resources with all other insect species). Thus,

298
$$SPO = \frac{\sum_{j} w_{ij}}{p \cdot (s-1)}$$
 eqn 1

where p is total number of pollen types in the community and s is total number of insect species. *SPO* ranges from 0 (no interspecific overlap) to 1 (maximum overlap). For simplicity, we ignored phenological and size constraints and assumed all species were equally likely to share resources. *SPO* was calculated for each of the 21 selected insect species (Appendix S3 gives an example of the procedure). All one-mode network analyses were done with the *tnet R* package (version 3.0.5, Opsahl 2009b).

305 Intraspecific overlap

306 To estimate the degree of intraspecific overlap for each species, i.e. the amount of 307 overlap in pollen use among its individuals, the 21 two-mode submatrices (consisting of 308 conspecific individuals and their pollen types) were transformed into one-mode 309 weighted networks following the same approach as above. Each one-mode matrix 310 consisted of a rows and a columns, with a being number of conspecific individuals of 311 species S and the cell value (w_{ii}) was number of pollen types shared by the individuals i 312 and *j*. The standardized measure of intraspecific overlap (individual-individual overlap, 313 *IO*) was calculated as

314
$$IO = \frac{\sum_{i} \sum_{j} w_{ij}}{a \cdot n \cdot (a - 1)}$$
 eqn 2

where $\sum_{i} \sum_{j} w_{ij}$ is the total overlap among all conspecific individuals in the subnetwork of *S* (sum of the link weights for all individuals of *S*) and the denominator is the maximum overlap for the subnetwork of *S*, i.e. when all the conspecific individuals 318 share all pollen types used by the species (n = total number of pollen types used by S319 and a = total number of conspecific individuals of species S). See Appendix S4 for an 320 illustrative example of the detailed calculation method of intraspecific overlap. It is 321 important to note that our estimates of inter- and intraspecific overlap in pollen use are 322 only a 'proxy' of competition, as either individuals or species, respectively, might share 323 the same flower species but use a different resource (e.g. pollen or nectar). We thus 324 prefer to use the neutral term 'overlap' instead of competition because the real sign of 325 the interaction is unknown.

326 Statistical analysis

327 Structural equation modeling (SEM) was used to determine the effects of inter- and 328 intraspecific overlap, insect species abundance and species phenophase on the degree of 329 individual specialization (WIC/TNW). This technique explores causal relationships among ecological variables and it allows decomposing total effects into direct and 330 331 indirect effects (Grace 2006). We proposed a model with intra- (10) and interspecific 332 overlap (SPO) directly influencing WIC/TNW and also abundance and phenophase 333 connected directly to WIC/TNW as well as indirectly through effects on IO and SPO 334 (see full path diagram in Fig. 5). Standardized path coefficients were estimated with 335 maximum likelihood method, significance of each one determined with a Wald test and error terms expressed as $\sqrt{1-R^2}$. Indirect effects were calculated by the product of the 336 path coefficients connecting two variables of interest, and total effects were defined as 337 338 the sum of direct and indirect effects. To assess whether the model fits the observed data we performed a Chi-square test of goodnes-of-fit (χ^2) and calculated the standardized 339 root mean residuals (SRMR). A non-significant χ^2 indicates that predicted covariance 340 341 among variables in the model is not distinguishable from the observed covariances, while SRMR calculates deviations between observed and predicted covariances. 342

343 Variables were *log* transformed when necessary to meet the normality assumption. All
344 analyses were done in *R* version 2.15.0 (R Development Core Team 2012) with package
345 *lavaan* (Rosseel 2012).

346

347 SAMPLING LIMITATIONS AND STUDY CAVEATS

348 Downscaling a community-wide pollination network from sp-sp to i-sp is a 349 methodologically complex task, and consequently several sampling limitations must be 350 noted. First, owing to the difficulty in identifying pollen grains to species level, we 351 made 'pollen type clusters', referring to pollen from closely related species (see 352 Appendix S1 for details). This means specialization and overlap levels might be 353 overestimated in some cases. Downscaling may improve our understanding of 354 networks, but achieving sufficient sampling to resolve them is hard, even more than in 355 sp-sp networks (Chacoff et al. 2012). For that reason, the estimation of individual 356 specialization is restricted to a small proportion of the total number of species in the 357 networks (19% for CN and 11% for PC). The number of replicates per species is 358 relatively low (5-10 sampled individuals), mainly because quantifying pollen loads is a 359 highly time-consuming task. However, we evaluated completeness of interaction 360 detection for each species (see Appendix S5) and 69% of the expected interactions were 361 detected on average. Studies on the degree of individual specialization have not yet 362 determined the minimum number of individuals of a population needed to be sampled to 363 get reliable estimations (Araújo, Bolnick & Layman 2011). By applying Monte Carlo 364 resampling procedures, we partly overcame the problem of overestimating the measures 365 of individual specialization due to low sample sizes (Bolnick et al. 2002, Araújo et al. 366 2008).

367 Temporal consistency of the observed individual specialization is important. Here, each individual pollen load sample, rather than a snapshot of the individual 368 369 foraging behaviour, is a picture of the extended flower visitation history of the 370 individual (Bosch et al. 2009). Thus, pollen loads can be reasonable estimators of 371 individual's long-term foraging patterns because pollen grains can remain attached to 372 insect bodies for long periods (Courtney et al. 1981). Obviously, the attachment time 373 depends upon species-specific grooming behaviour and body surface hairiness, as well 374 as pollen grain surface structure. Although we lack information about specific pollen 375 attachment times, we identified pollen grains on insects even one month after the last ρει flowering plant of a given species was blooming in the area (e.g. Rosmarinus 376 377 officinalis).

379 **Results**

380

381 STRUCTURAL PARAMETERS OF SP-SP AND I-SP NETWORKS

382 Downscaling from *sp-sp* to *i-sp* modified most network topological parameters studied 383 (Table 1). The direction and magnitude of the observed changes depended on the 384 foraging behaviour of insect individuals, as shown by the contrasting results obtained 385 between null models and empirical networks. Empirical *i-sp* networks were obviously 386 larger in size than *sp-sp*, as most species were resolved into several individuals except 387 singleton observations (i.e. insect species observed only once and therefore consisting 388 of one individual). Consequently, at both study sites, downscaling increased total 389 number of interactions (1) 1.5-fold (Fig. 1), although significantly less than the expected 390 with null *i-sp* networks. Linkage density (*LD*) and network connectance (*C*) in empirical 391 *i-sp* networks was nearly half the expected compared to the null hypothesis. Therefore, such differences between *sp-sp* and *i-sp* can be attributed to a significant decrease in the 392 393 number of links per pollinator node in empirical *i-sp* networks (Table 1), rather than to 394 an effect of increasing network size. Single individuals had a narrower foraging niche 395 than their corresponding species. Mean pollinator linkage level (L_A) in *i-sp* networks 396 was c. 50% lower than that predicted by the null model. Mean interaction diversity for 397 pollinators (H_A) was also significantly smaller when downscaling, because individuals 398 transported fewer and a more variable load of pollen types than their corresponding 399 species. Because of this idiosyncratic and heterogeneous individual behaviour, changes 400 in network nestedness were also observed. Both sp-sp and i-sp networks were 401 significantly nested; however, NODF values were consistently lower in empirical *i-sp* 402 than in null *i-sp* networks. Furthermore, minor but statistically significant decreases of 403 interaction diversity (H_2) were observed in empirical *i-sp* networks compared to null 404 models, due to differences in the number of interactions, whereas interaction evenness 405 (E_2) increased showing a reduction in the skewness in the distribution of link 406 frequencies.

407

408 PARTITION OF SPECIES LINKAGE LEVEL AMONG CONSPECIFIC409 INDIVIDUALS AND RELATIVE DEGREE OF INDIVIDUAL SPECIALIZATION

410 In general, species' linkage level - or species' niche width - was partitioned among specialized conspecific individuals. Most individuals had a much lower L_i than their 411 412 species (L_{sp}) (Fig. 2), i.e. individuals were always more specialized than their 413 corresponding species (average ratio $L_i/L_{sp} = 0.55$). This was also observed when insect 414 orders were treated separately (Figure S1). For the 21 species with \geq 5 individuals, L_i 415 was significantly lower than expected under the null hypothesis (Table S2). Therefore, a 416 generalist species was composed of specialist individuals using different resources, and 417 only in a few cases did individuals have a similar feeding niche as their corresponding 418 species (Fig. 3). When examining how resources are partitioned among individuals 419 within a species, we found a nested pattern only in five out of 14 species at CN [NODF: 420 Halictus spp. (H. scabiosae and H. fulvipes) = 60.39, Halictus vestitus = 52.71, Oedemera flavipes = 50.29, Paragus tibialis = 48.77, Stomorhina lunata = 49.46; $P < 10^{-10}$ 421 0.001] and in one out of seven species at PC (*NODF*: Halictus spp. = 50.90, P < 0.001). 422 423 Thus, for most species, individuals were highly heterogeneous in their use of pollen, 424 and interactions of the most specialized individuals were rarely proper subsets of those 425 of more generalized individuals.

426 On average, an individual niche represented c. 46% of total species niche (mean 427 $WIC/TNW \pm$ sd; $CN = 0.45 \pm 0.25$, $PC = 0.48 \pm 0.27$; empirical WIC/TNW < null 428 WIC/TNW for all species) (Fig. 4, Table S2). At both sites, a large fraction of

individuals (63.5% in *CN* and 54% in PC) had a niche which differed strongly from their species' niche, i.e. $PS_i < 0.5$ (Fig. S2). However, the intraspecific variability of PS_i was high (average coefficient of variance in mean PS_i of species was 59% at *CN* and 86% at *PC*). Thus, most species consisted of both generalist and specialist individuals (examples in Fig. 3).

434

435 FACTORS AFFECTING DEGREE OF INDIVIDUAL SPECIALIZATION

436 Indices of inter- and intraspecific overlap in pollen use are summarized in Table S2. 437 Results for the proposed SEM model are reported in Fig. 5. Observed data fitted reasonably well the proposed model ($\chi^2 = 0.198$, d.f. = 1, P = 0.66). High levels of inter-438 439 and intraspecific overlap reduced individual specialization (increased WIC/TNW 440 values). From all ecological factors included in the model, insect species abundance showed the strongest total effect on individual specialization (-0.561, P = 0.002), partly 441 442 mediated through its significant negative association with intraspecific overlap 443 (association with interspecific overlap was non-significant). This suggests that as 444 species abundance increases, individuals use a smaller subset of the whole species 445 niche, thereby reducing overlap between conspecifics. Likewise, species phenophase 446 significantly affected intraspecific overlap, but not interspecific overlap, i.e. species 447 with short phenophases had individuals with greater overlap among conspecifics than 448 species with long phenophases. This relationship caused a negative indirect effect on 449 WIC/TNW (-0.358, P = 0.02), although the total effect was non-significant (-0.168, P =450 0.35).

452 **Discussion**

453

454 Our analyses showed that when downscaling from sp-sp to *i*-sp pollen-transport 455 networks different structural parameters changed significantly, specifically linkage 456 density, connectance, nestedness and interaction diversity. The rationale for such 457 changes appears to be the high degree of individual specialization for most pollinator 458 species. This heterogeneity in pollen use and foraging behaviour among conspecific 459 individuals has been overlooked in network studies, despite the potential 460 misinterpretation of ecological dynamics and intra- and interspecific interactions 461 occurring in the community. We discuss these results suggesting possible causes and 462 implications of the main findings.

463

464 NETWORK DOWNSCALING AND INDIVIDUAL FORAGING BEHAVIOR

465 Downscaling to the individual level revealed a high degree of specialization and 466 heterogeneity hidden within *sp-sp* networks. Generalist pollinator species were actually found to be composed of specialist individuals. Results showed empirical *i-sp* networks 467 468 had lower linkage density, connectance, nestedness and interaction diversity than 469 predicted by null models, because conspecific individuals were idiosyncratic in their 470 food plant choice and foraging behaviour. Species linkage level was usually partitioned 471 among specialist individuals (Fig. 2), and this was true for all pollinators combined and 472 also when separating species into orders (Hymenoptera, Diptera and Coleoptera, Fig. 473 S1). Therefore, the most common mechanism for pollinator species to achieve a broad 474 niche (here high L_{sp}) was to have individuals with different and narrow niches ($L_i <<$ 475 L_{sp} (Fig. 3a). Quantitative measures of individual specialization (*WIC/TNW*, *PS_i*) also 476 confirmed this pattern and let us to reject the null hypothesis of individual generalist

477 sampling from species pollen use distribution. Our average values of *WIC/TNW* were 478 within the range reported by Araújo *et al.* (2011) for a broad array of taxa. Nevertheless, 479 both generalist and specialist individuals were commonly found within a species. The 480 frequency distribution of individual niche width was highly skewed, i.e. common 481 species had only a few individuals with a wide niche. However, intraspecific partition of 482 resources was quite overdispersed (i.e. non-nested) and the specialized individuals were 483 not carrying a subset of the pollen types carried by more generalized conspecifics.

484 Network downscaling from species to individuals seems a promising way to 485 connect pollination networks to pollinator foraging behaviour by further exploring 486 mechanisms underlying the observed patterns. The foraging behaviour of insect 487 pollinators is very flexible and a complex array of strategies for efficient collection of 488 pollen and nectar have been described (Goulson 1999). For instance, generalist Apis 489 *mellifera* individuals ($L_i/L_{sp} = 6/12$) may be scouts searching for new food resources 490 independently of each other, whereas foraging workers, which are guided to food 491 resources by specific waggle dances, only carry pollen from one or two species (Seeley 492 1983; Dupont et al. 2011). Furthermore, individual specialization in pollinator species 493 may reflect the individual foraging behaviour described as flower constancy (Waser 494 1986). Flower constancy occurs in many pollinators when individuals restrict their visits 495 to certain flowers, even ignoring more rewarding alternatives, although explanation of 496 this behaviour is still in debate (Chittka, Thomson & Waser 1999). This individual 497 specialization is likely to be beneficial to plants, since it might decrease heterospecific 498 pollen deposition on conspecific stigmas thus preventing stigma clogging, as well as 499 conspecific pollen loss on heterospecific flowers (Morales & Traveset 2008).

500

501 FACTORS INFLUENCING INDIVIDUAL SPECIALIZATION

Variation in individual specialization depends both on intrinsic (e.g. sex, age, morphology, behavior and physiology) and extrinsic factors (e.g. ecological interactions, population density and diversity of resources) (Bolnick *et al.* 2003; Araújo, Bolnick & Layman 2011). Among the extrinsic factors explored, we found evidence of a significant relationship between ecological interactions and inter-individual variation. Such relationship is likely to have ecological consequences for population and community dynamics (Bolnick *et al.* 2011; Wolf & Weissing 2012).

509 Low levels of interspecific overlap increased the degree of individual 510 specialization, thus supporting the notion that in the absence of competing species 511 individuals switch to different resources depending on their phenotypes (Costa et al. 512 2008; Bolnick et al. 2010). Populations can expand their diet breadth when individuals 513 expand their niche and/or specialize on different niches (Bolnick et al. 2007; Tinker, 514 Bentall & Estes 2008). Intraspecific competition has been documented to increase 515 individual specialization (Svanbäck & Bolnick 2007; Araújo et al. 2008). However, this 516 outcome depends on the type of rank preference variation among individuals, i.e. 517 whether individuals have the same or different primary and secondary preferred 518 resources (Svanbäck & Bolnick 2005; Araújo et al. 2011). Our results suggest a 519 scenario where at low intraspecific overlap levels individuals are using different 520 preferred resources, whereas at high levels they expand their niches adding the same 521 resources, thereby reducing individual specialization.

In addition, population densities affect individual foraging decisions as each individual's choice depends on those made by other individuals depleting the floral resources. Our SEM model suggested that as species abundance increases, conspecific individuals become more specialized and heterogeneous in pollen choice, which is concordant with other studies (e.g. Svanbäck & Bolnick 2007; Tinker *et al.* 2012).

527 Unexpectedly, species abundance was negatively associated to intraspecific overlap 528 level. However, this might be explained when considering insect phenologies. Two 529 species might be equally abundant by producing either a cohort with a high number of 530 individuals during a short period or several cohorts with low numbers of individuals 531 during a long period, but intraspecific competition would be stronger in the former case. 532 Finally, degree of individual specialization was not influenced by species 533 phenophase. Species with intermediate-long phenophases (range 40-100 days) showed 534 both high and low individual specialization, probably depending on whether they 535 consisted of short-lived individuals, long-lived or both. Given that species phenophase 536 was associated to intraspecific overlap (Fig. 5), we might expect species with very short 537 phenophases to consist of short-lived individuals with broad niches (relative to the 538 corresponding species). On the other hand, species with long phenophases might consist 539 of either short-lived individuals with narrow and non-overlapping niches or a 540 combination of individuals with different phenophases and degree of specialization. 541 Individual-based networks would certainly be a more informative tool to examine 542 seasonal dynamics if, for instance, individuals present at the beginning of the season 543 interact with early flowering plants whereas those present towards the end of the season 544 do it with another set of plants.

545

546 RELEVANCE OF NETWORK DOWNSCALING

547 Downscaling networks from species to individuals is important in our efforts to explore 548 mechanisms acting at the individual level, which further may upscale and shape species 549 network structure (Olesen *et al.* 2010). Indeed, the individual-based networks reported 550 here provide useful information to improve the understanding of species-based 551 networks because most *sp-sp* networks contain a substantial proportion of singleton

552 observations, which means they are based on observations of only one individual (e.g. 553 23% of an arctic network in Olesen et al. 2008 was based on singletons). Commonly, 554 rare species in networks appear to be more specialized than they really are due to 555 insufficient sampling of the rare interactions (Vázquez & Aizen 2003; Dorado et al. 556 2011). In concordance with this, our results suggest that a specialized behaviour of 557 individuals compared to the species might be a possible explanation for the sampling 558 bias in the estimation of linkage level in rare species. Similarly, most abundant species 559 tend to be also the most generalized in pollination networks (e.g. Elberling & Olesen 560 1999; Olesen et al. 2008), but as we have shown here, this might actually cover a 561 scenario where the conspecific individuals are specialized on different resources. 562 Everything else being equal, individuals of abundant insect species are observed more 563 frequently than those of rare species, and as new individuals are collected proportionally 564 more new links are added to the species due to the specialized behaviour of the 565 individuals. Thus, some broadly described specialization patterns in *sp-sp* pollination 566 networks might have their origin in *i-sp* networks. Because ecological specialization is 567 not a fixed species attribute and much variation exists within species, more studies are 568 needed to explore ecological specialization across scales (Devictor *et al.* 2010).

569 Our findings highlight the importance of also taking inter-individual variation 570 into account when studying higher-order structures such as networks, as part of our 571 understanding of network structure and dynamics hidden on adjacent scale levels. For 572 example, the high heterogeneity in pollen use among conspecifics enforces a high 573 heterogeneity in interaction strength in species-based networks as well, which, so far, 574 has been completely neglected. The strength of a sp-sp interaction depends on the 575 number of individuals taking part in the interaction and the degree of their involvement. 576 This has obviously important implications in the interpretation of community structure

577 and dynamics (Bolnick et al. 2011; Sih et al. 2012). For instance, a pollen type is more 578 likely to be an important resource to a species than other pollen types if a larger 579 proportion of the population is using it. Consequently, having generalized species 580 decomposed into dissimilar and specialized individuals might increase stability of 581 species to the loss of a resource, because only a small proportion of the population 582 would be affected (Wolf & Weissing 2012). By contrast, high individual specialization 583 increases species vulnerability to the loss of individuals. A species composed of 584 generalized individuals may lose a substantial proportion of its population before any 585 effects are seen in the species-based network, whereas a species composed of idiosyncratic individuals loses 586 specialized individuals and network links 587 simultaneously. Therefore, our cross-scale level study suggests that individual foraging mode affects species persistence and, further, network stability. The demonstrated 588 589 variation in individual resource use will affect the network outcome of disturbances, and 590 information about the specific kind of disturbance will be important in our efforts to predict how network stability in detail is affected. 591

592 The intraspecific heterogeneity in pollen use might be related to trait variability 593 among individuals, and most of these traits are subject to natural selection. Because 594 there are trait-matching constraints in how links are distributed in networks, 595 incorporating traits into models which predict species interactions have already helped 596 to gain more insight in network structure and properties (e.g. Petchey et al. 2008; Stang 597 et al. 2009; Ibanez 2012). Therefore, further research on individual-based networks 598 would enable us to link network theory to evolutionary biology by working at the 599 proper scale where natural selection takes place. Exploring all the potential bottom-up 600 processes determining the emergent properties of interaction networks seems a 601 promising avenue for future studies.

603 Acknowledgements

604 We are very grateful to David Baldock, Xavier Canvelles, Leopoldo Castro, Andreas Werner Ebmer, Xavier Espadaler, David Gibbs, Gerard Le Goff and Jordi Ribes for 605 606 taxonomical insect identifications and to Sandra Garcés for help with pollen 607 identifications. Servei de Protecció d'Espècies, Espais de Natura Balear (Conselleria 608 de Agricultura, Medi Ambient i Territori) and the military from Acar Puig Major/EVA 609 n°7 (Ministry of Defence) provided permission to work at the study sites. We thank 610 Sébastien Ibanez, Jimena Dorado and two anonymous reviewers for providing valuable 611 comments on the manuscript. This work was financially supported by the Spanish 612 Ministry of Science and Innovation (projects CGL2007-61165/BOS and CGL2010-613 18759/BOS to AT), CT received an FPU grant from the Spanish Ministry of Education, 614 JMO was funded by the Danish Science Research Council, and KT by Aarhus University Research Foundation. 615

616

617 Data Accessibility

618 R scripts for analysis and data can be found in: DRYAD entry doi:xx.xxxx/dryad.xxxx

622	Alarcón,	R.	(2010)) Congruence	between	visitation	and	pollen-transpor	t networks	in a

- 623 California plant-pollinator community. *Oikos*, **119**, 35-44.
- 624 Araújo, M.S., Guimarães Jr., P.R., Svanbäck, R., Pinheiro, A., Guimarães, P., Dos Reis,
- 625 S.F. *et al.* (2008) Network analysis reveals contrasting effects of intraspecific
- 626 competition on individual vs. population diets. *Ecology*, **89**, 1981-1993.
- 627 Araújo, M.S., Martins, E.G., Cruz, L.D., Fernandes, F.R., Linhares, A.X., Dos Reis,
- 628 S.F. *et al.* (2010) Nested diets: a novel pattern of individual-level resource use.
 629 *Oikos*, **119**, 81-88.
- Araújo, M.S., Bolnick, D.I. & Layman, C.A. (2011) The ecological causes of individual
 specialisation. *Ecology Letters*, 14, 948-958.
- 632 Barrat, A., Barthélemy, M., Pastor-Satorras, R. & Vespignani, A. (2004) The
- architecture of complex weighted networks. *Proceedings of the National Academy*of Science of the United States, 101, 3747-3752.
- Bastian M., Heymann S. & Jacomy, M. (2009) Gephi: an open source software for
- 636 exploring and manipulating networks. *International AAAI Conference on Weblogs*
- 637 *and Social Media*. Available at: <u>https://gephi.org/</u>
- Bolnick, D.I., Yang, L.H., Fordyce, J.A., Davis, J.M. & Svanbäck, R. (2002) Measuring
 individual-level resource specialization. *Ecology*, 83, 2936-2941.
- 640 Bolnick, D. I., Svanbäck, R., Fordyce, J.A., Yang, L.H., Davis, J.M., Hulsey C.D., et al.
- 641 (2003) The ecology of individuals: Incidence and implications of individual
 642 specialization. *American Naturalist*, 161, 1-28.
- 643 Bolnick, D. I., Svanbäck, R., Araújo M.S. & Persson, L. (2007) Comparative support
- 644 for the niche variation hypothesis that more generalized populations also are more

heterogeneous. Proceedings of the National Academy of Science of the United
States, 104, 10075-10079.

- 647 Bolnick, D.I., Ingram, T., Stutz, W.E., Snowberg, L.K., Lee Lau, O. & Paull, J.S.
- 648 (2010) Ecological release from interspecific competition leads to decoupled changes
- 649 in population and individual niche width. Proceedings of the Royal Society B-
- 650 *Biological Sciences*, **277**, 1789-1797.
- 651 Bolnick, D.I., Amarasekare, P., Araújo, M.S., Burger, R., Levine, J.M., Novak, M., et
- 652 *al.* (2011) Why intraspecific trait variation matters in community ecology. *Trends in*
- 653 *Ecology and Evolution*, **26**, 183-192.
- Bosch, J., Martín González, A.M., Rodrigo, A. & Navarro, D. (2009) Plant-pollinator
- 655 networks: adding the pollinator's perspective. *Ecology Letters*, **12**, 409-419.
- 656 Chacoff, N.P., Vázquez, D.P., Lomáscolo, S.B., Stevani, E.L., Dorado, J. & Padrón, B.
- 657 (2012) Evaluating sampling completeness in a desert plant-pollinator network.
 658 *Journal of Animal Ecology*, 81, 190-200.
- 659 Chittka, L., Thomson, J.A. & Waser, N.M. (1999) Flower constancy, insect psychology,
- and plant evolution. *Naturwissenschaften*, **86**, 361-377.
- 661 Costa, G.C., Mesquita, D.O., Colli, G.R. & Vitt, L.J. (2008) Niche expansion and the
- Niche Variation Hypothesis: does the degree of individual variation increase in
 depauperate assemblages? *The American Naturalist*, **172**, 868-877.
- 664 Courtney, S.P., Hill, C.J. & Westerman, A. (1981) Pollen carried for long periods by
- 665 butterflies. *Oikos*, **38**, 260-263.
- 666 Dall, S.R.X., Bell, A.M., Bolnick, D.I., Ratnieks, F.L.W. & Sih, A. (2012) An
- 667 evolutionary ecology of individual differences. *Ecology Letters*, **15**, 1189-1198.

- Devictor, V., Clavel, J., Julliard, R., Lavergne, S., Mouillot, D., Thuiller, W. et al. 669 (2010) Defining and measuring ecological specialization. Journal of Applied 670 Ecology, 47, 15-25.
- Dorado, J., Vázquez, D., Stevani, E.L. & Chacoff, N.P. (2011) Rareness and 671 672 specialization in plant-pollinator networks. Ecology, 92, 19-25.
- 673 Dormann, C.F., Gruber, B. & Fründ, J. (2008) Introducing the bipartite Package:
- 674 Analysing Ecological Networks. R news, 8/2, 8-11.
- 675 Dormann, C.F., Fründ, J., Blüthgen, N. & Gruber, B. (2009) Indices, Graphs and Null
- 676 Models: Analyzing Bipartite Ecological Networks. The Open Ecology Journal, 2, 7-24. 677
- Dupont, Y. L., Trøjelsgaard, K. & Olesen, J.M. (2011) Scaling down from species to 678
- 679 individuals: a flower-visitation network between individual honeybees and thistle 680 plants. Oikos, 120, 170-177.
- 681 Elberling, H. & Olesen, J.M. (1999) The structure of a high latitude plant-flower visitor 682 system: the dominance of flies. *Ecography*, **22**, 314-323.
- 683 Fontaine, C., Collin, C.L. & Dajoz, I. (2008) Generalist foraging of pollinators: diet 684 expansion at high density. Journal of Ecology, 96, 1002-1010.
- 685 Fortuna, M.A., García, C., Guimarães, P.R. & Bascompte, J. (2008) Spatial mating 686 networks in insect-pollinated plants. *Ecology Letters*, **11**, 490-498.
- 687 Fortuna, M.A., Popa-Lisseanu, G., Ibáñez, C. & Bascompte, J. (2009) The roosting
- 688 spatial network of a bird-predator bat. *Ecology*, **90**, 934-944.
- 689 Forup, m.L. & Memmott, J. (2005) The restoration of plant-pollinator interactions in
- 690 hay meadows. Restoration Ecology, 13, 265-274.

- Gibson, R.H., Nelson, I.L., Hopkins, G.W., Hamlett, B.J. & Memmott, J. (2006)
- 692 Pollinator webs, plant communities and the conservation of rare plants: arable
 693 weeds as a case study. *Journal of Applied Ecology*, 43, 246-257.
- 694Gómez, J.M., Perfectti, F. & Jordano, P. (2011) The Functional Consequences of695Mutualistic Network Architecture. PLoS ONE, 6(1), e16143.
- 696 doi:10.1371/journal.pone.0016143
- 697 Gómez, J.M. & Perfectti, F. (2012) Fitness consequences of centrality in mutualistic
- 698 individual-based networks. *Proceedings of the Royal Society B-Biological Sciences*,
- **6**99 **279**, 1754-1760.
- 700 Goulson, D. (1999) Foraging strategies of insects for gathering nectar and pollen, and
- implications for plant ecology and evolution. *Perspectives in Plant Ecology*, 2, 185209.
- Grace, J.B. (2006) *Structural Equation Modeling and Natural Systems*. Cambridge
 University Press, Cambridge, United Kingdom.
- 705 Ibanez, S. (2012) Optimizing size thresholds in a plant-pollinator interaction web.
- Towards a mechanistic understanding of ecological networks. *Oecologia*, **170**, 233242.
- 708 Ings, T.C., Montoya, J.M., Bascompte, J., Blüthgen, N., Brown, L., Dormann, C.F. et
- 709 *al.* (2009) Ecological networks beyond food webs. *Journal of Animal Ecology*, 78,
 710 253-269.
- Jordano, P., Bascompte, J. & Olesen J.M. (2003) Invariant properties in coevolutionary
 networks of plant-animal interactions. *Ecology Letters*, 6, 69-81.
- Jordano, P., Bascompte, J. & Olesen, J.M. (2006) The ecological consequences of
 complex topology and nested structure in pollination webs. *Plant-pollinator*

- 715 *interactions: from specialization to generalization* (eds N.M. Waser & J. Ollerton),
- pp. 173-199. University of Chicago Press, Chicago, USA.
- 717 Lopezaraiza-Mikel, M.E., Hayes, R.B., Whalley, M.R. & Memmott, J. (2007) The
- 718 impact of an alien plant on a native plant-pollinator network: an experimental
 719 approach. *Ecology Letters*, **10**, 539-550.
- 720 Memmott, J. (1999). The structure of a plant-pollinator food web. *Ecology Letters*, 2,
- 721 276-280.
- 722 Morales, C.L. & Traveset, A. (2008) Interspecific pollen transfer: magnitude,
- prevalence and consequences for plant fitness. *Critical Reviews in Plant Sciences*,
 27, 221-238.
- 725 Oksanen, J., Blanchet, F.G., Kindt, R., Legendre, P., Minchin, P.R., O'Hara, R.B. et al.
- (2012) *vegan*: Community Ecology Package. R package version 2.0-3. Available at:
 http://CRAN.R-project.org/package=vegan
- Olesen, J. M., Bascompte, J., Elberling, H. & Jordano, P. (2008) Temporal dynamics in
 a pollination network. *Ecology*, **89**, 1573-1582.
- 730 Olesen, J.M., Dupont. Y.L., O'Gorman, E.J., Ings, T.C., Layer, K., Melián, C.J. et al.
- 731 (2010) From Broadstone to Zackenberg: Space, Time and Hierarchies in Ecological
- Networks. *Advances in Ecological Research: Ecological Networks* (ed. G.
 Woodward), pp. 1-69. Elsevier Academic Press Inc, San Diego, USA.
- 734 Opsahl, T. (2009a). Projecting two-mode networks onto weighted one-mode networks.
- 735 Available at: http://toreopsahl.com/tnet/two-mode-networks/projection/ Last
- 736 accessed 17 May 2013.
- 737 Opsahl, T. (2009b) tnet: Software for Analysis of weighted and longitudinal networks.
- 738 *Structure and Evolution of Weighted Networks*. PhD thesis, pp. 104-122. University

- of London (Queen Mary College), London, UK. Available at:
 http://toreopsahl.com/publications/thesis/
- Padrón, B., Nogales, M. & Traveset, A. (2011) Alternative approaches of transforming
 bimodal into unimodal mutualistic networks. The usefulness of preserving weighted
 information. *Basic and Applied Ecology*, 12, 713-721.
- 744 Perkins, S.E., Cagnacci, F., Stradiotto, A., Arnoldi, D. & Hudson, P.J. (2009)
- Comparison of social networks derived from ecological data: implications for
 inferring infectious disease dynamics. *Journal of Animal Ecology*, 78, 1015-1022.
- 747 Petanidou, T., Kallimanis, A.S., Tzanopoulos, J., Sgardelis, S.P. & Pantis, J.D. (2008)
- Long-term observation of a pollination network: fluctuation in species and
 interactions, relative invariance of network structure and implications for estimates
 of specialization. *Ecology Letters*, 11, 564-575.
- 751 Petchey, O.L., Beckerman, A.P., Riede, J.O. & Warren, P.H. (2008) Size, foraging, and
- food web structure. *Proceedings of the National Academy of Science of the United States*, **105**, 4191-4196.
- 754 Pires, M.M., Guimarães Jr., P.R., Araújo, M.S., Giaretta, A.A., Costa, J.C.L., & Dos
- Reis, S.F. (2011) The nested assembly of individual-resource networks. *Journal of Animal Ecology*, **80**, 896-903.
- Popic, T.J., Wardle, G.M. & Davila, Y.C. (2013) Flower-visitor networks only partially
 predict the function of pollen transport by bees. *Austral Ecology*, **38**, 76-86.
- 759 R Development Core Team (2012) R: A language and environment for statistical
- 760 computing. R Foundation for Statistical Computing, Vienna, Austria. Available at:
- 761 <u>http://www.R-project.org/</u>
- 762 Rosseel, Y. (2012) *lavaan*: An R Package for Structural Equation Modeling. *Journal of*
- 763 Statistical Software, **48**, 1-36.

- Roughgarden. J. (1972) Evolution of niche width. American Naturalist, 106, 683-718.
- 765 Seeley, T.D. (1983) Division of labor between scouts and recruits in honeybee foraging.
- 766 *Behavavioral and Ecological Sociobiology*, **12**, 253-259.
- Sih, A., Cote, J., Evans, M., Fogarty, S. & Pruitt, J. (2012) Ecological implications of
 behavioural syndromes. *Ecology Letters*, 15, 278-289.
- 769 Stang, M., Klinkhamer, P.G.L., Waser, N., Stang, I. & van der Meijden, E. (2009) Size-
- specific interaction patterns and size matching in a plant-pollinator interaction web.

771 *Annals of Botany*, **103**, 1459-1469.

- 772 Stephens, D.W. & Krebs, J.R. (1986) Foraging Theory. Princeton University Press,
- 773 Princeton, New Jersey, USA.
- 774 Svanbäck R. & Bolnick, D.I. (2005) Intraspecific competition affects the strength of
- individual specialization: an optimal diet theory model. *Evolutionary Ecological Research*, 7, 993-1012.
- 777 Svanbäck, R. & Bolnick, D.I. (2007) Intraspecific competition drives increased resource
- visual results of the Royal Society B-
- 779 Biological Sciences, **274**, 839-844.
- 780 Svanbäck, R., Rydberg, C., Leonardsson, K. & Englund, G. (2011) Diet specialization
- in a fluctuating population of *Saduria entomon*: a consequence of resource or
 forager densities? *Oikos*, **120**, 848-854.
- 783 Tinker, T., Bentall, G. & Estes, J.A. (2008) Food limitation leads to behavioral
- diversification and dietary specialization in sea otters. *Proceedings of the National Academy of Science of the United States*, **105**, 560-565.
- 786 Tinker, T., Guimarães Jr., P.R, Novak, M., Darcie Marquitti, F.M., Bodkin, J.L.,
- 787 Staedler, M. et al. (2012) Structure and mechanism of diet specialisation: testing

- models of individual variation in resource use with sea otters. *Ecology Letters*, **15**,
- 789 475-483.
- 790 Van Valen, L. (1965) Morphological variation and width of ecological niche. American
- 791 Naturalist, **99**, 377-390.
- Vázquez, D. & Aizen, M. (2003) Null model anlyses of specialization in plantpollinator interactions. *Ecology*, 84, 2493-2501.
- Waser, N. (1986) Flower constancy: Definition, cause and measurement. *The American Naturalist*, **127**, 593-603.
- Wolf, M. & Weissing, F.J. (2012) Animal personalities: consequences for ecology and
 evolution. *Trends in Ecology and Evolution*, 8, 452-461.
- 798 Woodward, G. & Warren, P.H. (2007) Body size and predatory interactions in
- freshwaters: scaling from individuals to communities. *Body size: The Structure and*
- 800 Function of Aquatic Ecosystems (eds A.G. Hildrew, D. Raffaelli & R. Edmonds-
- 801 Brown), pp. 98-117. Cambridge University Press, Cambridge.
- 802 Yvon-Durocher, G., Reiss, J., Blanchard, J., Ebenman, B., Perkins, D.M., Reuman, D.C.
- 803 *et al.* (2011) Across ecosystem comparisons of size structure: Methods, approaches,
- and prospects. *Oikos*, **120**, 550-563.
- 805

806 Supporting Information

- 807 The following supporting information is available for this article online:
- 808 **Table S1.** Definition of network parameters.
- 809 Table S2. List of individual specialization and overlap indices calculated for the 21
- 810 insect species of flower-visitors selected from our networks.
- Fig S1. Relation between linkage level of species (L_{sp}) and individuals (L_i) for the main
- 812 insect orders.

- 813 **Fig. S2.** Frequency histogram of the proportional similarity indices of individuals (PS_i)
- 814 from the 21 selected species in our networks.
- 815 Appendix S1. Insect pollen load analysis.
- 816 Appendix S2. Indices for quantifying individual specialization and niche components.
- 817 Appendix S3. Measuring interspecific overlap in one-mode weighted networks.
- 818 **Appendix S4.** Measuring intraspecific overlap with one-mode weighted networks.
- Appendix S5. Evaluation of species sampling completeness. 819
- 820

821 **Table 1.** Structural properties of the empirical *sp-sp* networks, empirical *i-sp* networks

and null *i-sp* networks.

823

	sp-sp n	etworks	i-sp networks					
	Emp	irical	Emp	irical	Null model			
Qualitative network parameters	CN	PC	CN	PC	CN	PC		
Number of pollinator nodes (A)	73	61	190	137	190	137		
Number of pollen type nodes (P)	55	49	55	49	55	49		
Total number of nodes $(T=A+P)$	128	110	245	186	245	186		
Network size $(N=AxP)$	4,015	2,989	10,450	6,713	10,450	6,713		
Total number of interactions (I)	434	360	681^{\dagger}	506^{\dagger}	1,342.48	881.78		
Linkage density $([LD=I/(A+P)])$	3.39	3.27	2.78^{\dagger}	2.72^{\dagger}	5.48	4.74		
Connectance $(C=I/AxP)$	0.108	0.120	0.065^{\dagger}	0.075^{\dagger}	0.12	0.13		
Nestedness (NODF)	34.45*	38.65^{*}	26.99 ^{*,†}	29.67 ^{*,†}	44.29	43.77		
Quantitative network parameters								
Interaction diversity (H_2)	3.38	2.89	3.63†	3.18^{\dagger}	3.95	3.36		
Interaction evenness $(E_2 = H_2/H_{max})$	0.56	0.49	0.56^{\dagger}	0.51^{\dagger}	0.55	0.49		
Node parameters								
Mean pollinator linkage level (L_A)	5.9	5.9 [†]	3.5 [†]	3.6†	7.07	6.44		
Mean pollinator interaction diversity (H_A)	0.79	0.70	0.54^{\dagger}	0.54^{\dagger}	0.99	0.78		
Mean pollen type linkage level (L_P)	7.89	7.35	12.38 [†]	10.33 [†]	24.41	18		
Mean pollen type interaction diversity (H_P)	0.62	0.57	0.76^{\dagger}	0.69^{\dagger}	1.46	1.21		

⁸²⁴

825 * Significance p-value < 0.001. That is the probability of getting by random a higher value of

826 nestedness than the empirical one. See text for more details on the calculation.

* Observed values were outside of 95% confidence intervals of values obtained for 1,000 null *i*-

sp networks.





Plant pollen type





856 Figure legends

857

858 Fig. 1. Pollen-transport networks at two scales of resolution: (A) species-species (sp-sp) 859 and (B) *individuals-species (i-sp)*. Pale green nodes are plant pollen types whereas the 860 rest of nodes are pollinators. Coloured and labelled nodes are pollinators with ≥ 5 861 individuals captured (see Table S2 for full species names). In network (A) each 862 pollinator node represents an insect species with a size proportional to the number of 863 individuals captured in the field. The result of decomposing each insect species node 864 into its conspecific individuals is seen in network (B), where each pollinator node represents an insect individual and individuals of the same species are clumped together 865 866 in the graph. Networks were drawn with Gephi 0.8beta (Bastian et al. 2009).

867

Fig. 2. Comparison of species linkage level (L_{sp}) and individual linkage level (L_i) . Cells 868 869 represent combinations of L_i and L_{sp} (data from both sites, n = 327) with colour intensity 870 correlating with number of individuals found at each position. Matrix diagonal is the 871 species-individual isocline (i.e. perfect matching of species and individuals niche width, 872 $L_i = L_{sp}$), while deviations to the left indicate individuals, being more specialized than 873 their species $(L_i < L_{sp})$. The figure shows a high density of individuals with a linkage 874 level lower than their species, although some individuals are positioned on the species-875 *individual isocline* (mainly species captured one or a few times). Notice that the upper 876 right region representing highly generalized species with generalized individuals is 877 completely empty.

Fig. 3. *i-sp* binary pollen-transport matrices of two dipteran species representing linkage
level partitioning among conspecifics. There are two alternative mechanisms to evolve a

881 wide species niche: (a) generalist species composed of relatively specialized individuals 882 using different pollen types and (b) generalist species composed of relatively 883 generalized individuals using broad and similar subsets of resources. Species with high 884 heterogeneity in pollen use among conspecifics (example a) were more common. Bar 885 plots on top of the matrices show percentage of conspecific individuals carrying each 886 pollen type, so generalized species with specialized individuals have a long tail. (L_{sp} : species linkage level; L_i : individual linkage level; PS_i : proportional similarity index; 887 WIC/TNW: degree of individual specialization). 888

889

Fig. 4. Total niche width (*TNW*) partition into within-individual (*WIC*) and betweenindividual (*BIC*) components for species with ≥ 5 individuals sampled (n = 21, 10Hymenoptera spp., 7 Diptera spp. and 4 Coleoptera spp.) at both study sites (CN = *Coma de n'Arbona*, PC = *Passadís de Ses Clotades*).

894

895 Fig. 5. Path diagram showing the relative effect of several ecological factors (insect 896 phenophase, insect abundance, inter- and intraspecific overlap (SPO and IO 897 respectively) on the degree of individual specialization (*WIC/TNW*). Positive effects are 898 indicated by solid lines and negative effects by dashed lines. Thickness of arrows is 899 proportional to the standardized path coefficients indicated with numbers next to each 900 path. Significant paths are coloured in black, whereas non-significant ones are in grey. R^2 and error terms are shown for each endogenous variable. Statistics of goodness of fit 901 for this model are: $\chi^2 = 0.198$, df = 1, P = 0.656; SRMR = 0.02. 902

Appendix S1 Insect pollen load analysis

2

3 In the laboratory, we studied the pollen loads of each insect individual captured. For 4 frequently captured species, a maximum of 10 individuals per species was included in 5 the analysis. An isotonic water solution (1.5 ml) was added to each vial containing a 6 specimen and shaken for 5 sec in a vortex and then washed by agitation for 15 min in an 7 ultrasonic bath to remove pollen grains from their body surface. For honeybees, pollen 8 clumps in the corbicula of the hind legs were removed in advance. Afterwards, insects 9 were pinned for later identification by taxonomist experts. Vials containing the pollen 10 load solution were centrifuged at 13.000 rpm for 15 min and inspected for any presence 11 of a pollen pellet at the bottom. In vials with a large pollen pellet, to facilitate pollen 12 counting, we added a *Lycopodium* spore tablet containing 18,584 spores (batch no. 13 177745, Lund Univ., Sweden), whereas in the other vials the supernatant was carefully 14 removed, and the droplet with the pollen was suspended on a microscopic slide. Pollen 15 concentrated samples were homogenized with vortex agitation in order to dissolve the 16 spore tablets and then three drops (replicates) were mounted on slides. We added a 17 small pink fuchsine-stained jelly cube to the pollen smear, melted it and covered the 18 final smear with a cover slip (Kearns & Inouye 1993). To avoid any pollen 19 contamination, laboratory instruments were cleaned with ethanol between manipulation 20 of samples. Slides were analyzed under microscope (100-400x) to estimate total number 21 of pollen grains from each insect individual. All pollen grains were identified to the 22 lowest possible taxonomic level and the number of pollen grains of each species was 23 counted (on average 60% of total slide area was inspected for pollen). In slides from 24 pollen-dense samples, we counted the number of pollen grains and also the number of 25 spores. In these samples, total pollen number carried by the insect was estimated as the

26 mean of the three replicates, calculated as "number of pollen grains counted in a droplet
27 x (18,584/number of spores counted in droplet)" (Jakobsson, Padrón & Traveset 2008).

28 A pollen reference collection was made during the field season in order to facilitate 29 pollen identification. Pollen grains of each species were measured and photographed 30 under the microscope at 400x. Pollen grains of some species were indistinguishable 31 from those of closely related species, and therefore classified into 'pollen type clusters' 32 including more than one species: 'Galium' (including G. cinereum, G. crespianum and 33 G. balearicum), 'Teucrium' (including T. marum and T. asiaticum), 'Geraniaceae' 34 (including Geranium colombinum, G. lucidum, G. molle, G. purpureum and Erodium 35 cicutarium), and 'Asteraceae' (including all Asteraceae at study sites except Bellium 36 bellidioides, Carlina corymbosa and Santolina chamaecyparissus, which were 37 identified to species). Unknown pollen grains were assigned to morphotype species.

38

39 References in Appendix S1:

40 Jakobsson, A., Padrón, B. & Traveset, A. (2008) Pollen transfer from invasive

41 *Carpobrotus* spp. to natives - A study of pollinator behaviour and reproduction success.

42 Biological Conservation, 141, 136-145.

43 Kearns, C.A. & Inouye, D.W. (1993) Techniques for pollination biologists. University

44 Press of Colorado, Colorado, USA.

1 **Appendix S2** Indices for quantifying individual specialization and niche components

2

3 We used formulae from Bolnick et al. (2002) to calculate each niche component (TNW, 4 WIC and BIC; Roughgarden 1972, 1974) using categorical data on pollen types carried 5 by insect specimens. Shannon diversity index is used as a proxy for variance in niche 6 width, so the niche widens with addition of new resources and with increasing evenness 7 in resource use. For each species with ≥ 5 individuals sampled (14 spp. at CN, 7 spp. at 8 PC), we constructed a matrix with a rows and p columns, where a is the number of 9 individuals of species S, p the plant pollen types found on the body of individuals of S, 10 and the value in each cell (n_{ij}) is the number of j's pollen grains carried by the *i*th 11 individual. Then, using those matrices for each species S, we calculated niche 12 components as

13
$$WIC = \sum_{i} p_{i} \cdot \left(-\sum_{j} p_{ij} \cdot \ln(p_{ij})\right)$$

14

$$BIC = -\left(\sum_{i} p_{i} \cdot \ln(p_{ij})\right) - \left[\sum_{j} q_{j} \cdot \left(-\sum_{i} \gamma_{ij} \cdot \ln(\gamma_{ij})\right)\right]$$
15

$$TNW = -\sum_{i} q_{j} \cdot \ln(q_{j})$$

15
$$TNW = -\sum_{i} q_{i} \cdot \ln(q_{i})$$

16
$$p_{ij} = \frac{n_{ij}}{\sum_{j} n_{ij}}$$

$$p_i = \frac{\sum_{j} n_{ij}}{\sum_{i} \sum_{j} n_{ij}}$$

$$q_j = \frac{\sum_{i} n_{ij}}{\sum_{i} \sum_{j} n_{ij}}$$

$$\gamma_{ij} = \frac{n_{ij}}{\sum_{i} n_{ij}}$$

17

where, p_{ij} is the proportion of pollen type j in the total pollen load of individual i of 18 species S, p_i is the proportion of all pollen types used by S that are used by individual i, 19 20 q_i is the proportion of pollen type j in the total pollen load of S, and γ_{ij} is the proportion 21 of total pollen type *j* used by S that was carried by individual *i*. Relative degree of

individual specialization was obtained by dividing WIC by TNW, i.e. proportion of total 22 23 niche width explained by the within-individual component. 24 Following Bolnick et al. (2007) we also calculated overlap in pollen resource 25 use among each individual and the whole population as $PS_i = 1 - 0.5 \cdot \sum_j |p_{ij} - q_j|$ 26 27 where PS_i is the proportional similarity index, measuring the similarity in pollen 28 resource use distribution of individual *i* and the corresponding species S (Feisinger *et al.* 29 1981). Individuals using pollen resources in the same proportion as their species have a 30 $PS_i = 1$, whereas $PS_i = q_i$ if they are specialized in just one pollen resource *i*.

31

32 References in Appendix S2:

Bolnick, D.I., Svanbäck, R., Araújo, M.S. & Persson, L. (2007) Comparative support
for the niche variation hypothesis that more generalized populations also are more
heterogeneous. *Proceedings of the National Academy of Sciencie of the United States of America*, 104, 10075-10079.

- 37 Bolnick, D.I., Yang, L.H., Fordyce, J.A., Davis, J.M. & Svanbäck, R. (2002) Measuring
- individual-level resource specialization. *Ecology*, **83**, 2936-2941.
- 39 Feinsinger, P., Spears, E.E. & Poole, R.W. (1981) A simple measure of niche breadth.
- 40 *Ecology*, **62**, 27-32.
- 41 Roughgarden. J. (1972) Evolution of niche width. American Naturalist, 106, 683-718.
- 42 Roughgarden J. (1974) Niche width: Biogeographic patterns among lizard populations.
- 43 *American Naturalist*, **108**, 429-442.
- 44

1 Appendix S3 Measuring interspecific overlap in one-mode weighted networks.

2



3

Here we present a simple hypothetical example to illustrate the procedure applied when estimating interspecific overlap (*SPO*) for insect species in our networks. *Network 1* is a two-mode binary pollen-transport network depicting interactions between five insect species (red nodes) and four plant pollen types (green nodes). Insect species are linked to plant pollen types if they carried pollen grains. *Network 1* can also be represented as an interaction matrix (*Matrix 1*) with *s* rows and *p* columns, where *s* is the number of insect species (s = 5), *p* is the number of plant pollen types (p = 4) and the value in each

11 cell entry is 0 or 1 (i.e. absence or presence of pollen grains in insect's body, resp.). 12 Linkage level of species (L_{sp}) is the total number of plant pollen types carried (i.e. 13 matrix 1 row sum). This two-mode binary network was transformed into a one-mode 14 weighted network by counting the total number of plant pollen types shared among 15 species (co-occurrences projection method as in Opsahl 2009 and Padrón et al. 2011). 16 Network 2 is the one-mode projection of network 1, where insect species are linked if 17 pollen grains of the same plant pollen type were found on the body of both insects, and 18 the weight of each link is the number of plant pollen types in common. For instance, *sp1* 19 shares three pollen types with sp2, two pollen types with sp3 and only one with sp4 and 20 *sp5*. Therefore, the maximum link weight possible in this kind of network projection is 21 always the total number of plant pollen types present in the community (here p = 4). 22 This one-mode weighted network depicting the pattern of shared pollen types among 23 insect species can be also represented by an interaction matrix with s rows and s 24 columns (*Matrix 2*) where the value in each cell entry (w_{ii}) is the number of plant pollen 25 types shared. Therefore, sum of link weights ($\sum w_{ij}$, i.e. matrix 2 row sum) is the overlap 26 in plant pollen types among a certain species and all other species in the network. To get 27 a standardized measure of interspecific overlap ranging from 0 to 1 (SPO), we divided 28 $\sum w_{ii}$ by the maximum interspecific overlap possible for a species in the network, which 29 was calculated as p(s - 1), i.e. sum of node link weights $\sum w_{ii}$ in an hypothetical case 30 where all insect species carried all plant pollen types from the community and therefore 31 shared all (Matrix 3). We illustrate the complete calculation of species-species overlap 32 (SPO) for two species in our example.

33

34 **References in Appendix S3:**

- 35 Opsahl, T. (2009) *Projecting two-mode networks onto weighted one-mode networks*.
- 36 Available at: <u>http://toreopsahl.com/tnet/two-mode-networks/projection/</u>
- 37 Padrón, B., Nogales, M. & Traveset, A. (2011) Alternative approaches of
- 38 transforming bimodal into unimodal mutualistic networks. The usefulness of
- 39 preserving weighted information. *Basic and Applied Ecology*, **12**, 713-721.
- 40

1 **Appendix S4** *Measuring intraspecific overlap with one-mode weighted networks.*



2

3 Here we present an example illustrating the methodology used to estimate intaspecific 4 overlap (IO) for insect species in our networks. Network 1 is the two-mode binary pollen-transport network for sp1 (see previous example in Appendix S3) depicting 5 6 interactions between five insect individuals (red nodes) of sp1 and three plant pollen 7 types (green nodes). Therefore, network 1 represents within-species partition of pollen 8 resources among insect individuals. Insect individuals are linked to plant pollen types if 9 they carried pollen grains. Network 1 can also be represented as an interaction matrix 10 (Matrix 1) with a rows and n columns, where a is the number of sp1's insect individuals

(a = 5), n is the number of plant pollen types visited by sp1 $(n = 3, i.e. L_{sp1})$ and the 11 value in each cell entry is 0 or 1 (i.e. absence or presence of pollen grains on insect's 12 13 body, respectively). Linkage level of the corresponding species (L_{sp}) is the maximum linkage level possible for conspecific insect individuals (L_i , i.e. matrix 1 row sum). This 14 15 two-mode binary network (*Network 1*) was transformed into a one-mode weighted 16 network (*Network 2*) by counting the total number of plant pollen types shared among 17 individuals and using it as link weight w_{ii} (co-occurrences projection method as in 18 Opsahl 2009 and Padrón et al. 2011). In the example, *ind1* and *ind2* have two pollen 19 types in common and all other individuals share a pollen type among them. This one-20 mode weighted network can also be represented by an interaction matrix with a rows 21 and a columns (*Matrix 2*) where the value in each cell entry (w_{ii}) is the number of plant 22 pollen types shared among *ind_i* and *ind_j*, so $\sum w_{ij}$ (i.e. *matrix 2* row sum) represents the 23 total number of pollen types shared among ind_i and all other conspecific individuals. Therefore, sum of all matrix link weights ($\sum_{i} \sum_{j} w_{ij}$, i.e. *matrix 2* row and column sum) 24

is the overlap in plant pollen types among all conspecific individuals (i.e. intraspecific overlap). To get a measure of intraspecific overlap ranging from 0 to 1 (*IO*), we divided by the maximum intraspecific overlap possible for the corresponding species (see *Matrix 3*). As maximum number of pollen types in common (w_{ij}) among two individuals of *sp1* is *n*, maximum intraspecific overlap will occur when all conspecific individuals have the same linkage level as the species and therefore share all *n* pollen resources among them as represented in *Matrix 3* (i.e. $\sum w_{ij} = n x (a - 1) = 12$ and $\sum_{i} \sum_{j} w_{ij} = a x$

32 $\sum w_{ij} = 12 \times 5 = 60$). Complete formulae for calculation of intraspecific overlap (i.e. 33 individual-individual overlap, *IO*) is shown for *sp1* as example.

35 **References in Appendix S4:**

- 36 Opsahl, T. (2009). Projecting two-mode networks onto weighted one-mode networks.
- 37 Available at: http://toreopsahl.com/tnet/two-mode-networks/projection/
- 38 Padrón, B., Nogales, M. & Traveset, A. (2011) Alternative approaches of transforming
- 39 bimodal into unimodal mutualistic networks. The usefulness of preserving weighted
- 40 information. Basic and Applied Ecology, 12, 713-721.
- 41

ppliea

1 **Appendix S5.** Evaluation of species sampling completeness

2

3 We aimed to determine the extent to which the number of individuals sampled per species allowed for a sufficient description of species interactions (L_{sp}) . To evaluate 4 5 completeness of our sample sizes we first computed sampled-based rarefaction curves 6 for each species (Gotelli & Colwell 2001) using the package vegan (version 2.0-6, 7 Oksanen et al. 2012) in the R program (version 2.15.0, R Development Core Team 8 2012). Following Chacoff et al. (2012), for each of the 21 species studied, we calculated 9 the percentage of estimated asymptotic richness detected as,

10

11 %
$$S_{OBS} = 100 * S_{OBS} / S_E$$

12

13 where S_{OBS} is the observed pollen type richness in the samples (i.e. species linkage level, L_{sp}) and S_E is the asymptotic estimated pollen type richness (i.e. estimated species 14 15 linkage level). To compute S_E we used the Chao 2 non-parametric estimator in its bias-16 corrected form,

eqn 1

17

18

 $S_E = S_{OBS} + [(a-1)/a] * [u(u-1)/2(d+1)]$ egn 2

19

20 where *a* is the sample size (i.e. number of individuals of each species sampled for pollen 21 load analysis), u is the number of uniques (i.e. plant pollen types that occur only in one 22 sample) and d is the number of duplicates (i.e. plant pollen types that occur in two 23 samples) (Chao 2005). Chao 2 index relies on the principle that rare species in the 24 samples carry most information on the number of un-observed species (Chao 1984) and

1	is one of the least biased estimates for small sample sizes (Colwell & Coddington
2	1994).
3	
4	References in Appendix S5:
5	Chacoff, N.P., Vázquez, D.P., Lomáscolo, S.B., Stevani, E.L., Dorado, J. & Padrón,
6	B. (2012) Evaluating sampling completeness in a desert plant-pollinator network.
7	Journal of Animal Ecology, 81 , 190-200.
8	Chao, A. (1984) Nonparametric estimation of the number of classes in a population.
9	Scandinavian Journal of Statistics, 11, 265-270.
10	Chao, A. (2005) Species richness estimation. Encyclopedia of Statistical Sciences
11	(eds. N. Balakrishnan, C. Read & B. Vidakovic) Vol. 12, pp. 7907–7916. Wiley,
12	New York.
13	Colwell, R.K. & Coddington, J.A. (1994) Estimating terrestrial biodiversity through
14	extrapolation. Philosophical Transactions of the Royal Society (Series B), 345, 101-
15	118.
16	Gotelli, N.J. & Colwell, R.K. (2001) Quantifying biodiversity: procedures and
17	pitfalls in the measurement and comparison of species richness. Ecology Letters, 4,
18	379-391.
19	Oksanen, J., Blanchet, F.G., Kindt, R., Legendre, P., Minchin, P.R., O'Hara, R.B. et
20	al. (2012) vegan: Community Ecology Package. R package version 2.0-3. Available
21	at: <u>http://CRAN.R-project.org/package=vegan</u>
22	R Development Core Team (2012) R: A language and environment for statistical
23	computing. R Foundation for Statistical Computing, Vienna, Austria. Available at:
24	http://www.R-project.org/
25	



3 Fig. S1. Relation between linkage level of species (L_{sp}) and individuals (L_i) for the main 4 insect orders. Each matrix corresponds to a different insect pollinator order: 5 Hymenoptera (N_{HYM} =147 individuals), Diptera (N_{DIP} =114) and Coleoptera (N_{COL} =59). 6 Colours represent number of individuals with a given L_i and L_{sp} configuration, so figures 7 show where the highest density of individuals is in each matrix. Matrix diagonal 8 (species-individual isocline) represents perfect matching of individual niche width and 9 species niche width ($L_i = L_{sp}$), while deviations to the left indicate individuals being 10 more specialized than their species ($L_i < L_{sp}$). Filled cells are located in the upper region 11 of the diagonal because of the constraint $L_i \leq L_{sp}$. For all species in the different orders, specialist individuals predominate $(L_i/L_{sp} < 1)$. The trend is more marked for beetles, 12 13 probably because of their lower mobility.



2

Fig. S2. Frequency histogram (N = 122) of the proportional similarity indices of individuals (*PS_i*) from the 21 selected species in our networks. Individuals with narrower niches than their corresponding species are more frequent (60.7% individuals had a *PS_i* < 0.5), although some highly generalized individuals can be found as well.

Table S1. Definitions of parameters used to describe *sp-sp* and *i-sp* networks. The first five metrics are qualitative, whereas the last two are

quantitative.

Parameter	Definition
Linkage level (L)	Number of interactions of each network node, i.e. number of interactions per species (L_{sp}) in <i>sp-sp</i> networks and number of interactions per individual (L_i) in <i>i-sp</i> networks.
Network size (N)	Total number of possible interactions in the network, i.e. the number of cell entries in the interaction matrix (number of i rows multiplied by number of j columns).
Linkage density (LD)	Mean number of links per network node.
Connectance (C)	Realized proportion of all possible links (Dunne et al. 2002).
Nestedness (NODF)	Nestedness metric based on overlap and decreasing fills (Almeida-Neto <i>et al.</i> 2008). It measures to what extent the interaction pattern resembles a perfectly nested pattern where specialist species interact with a proper subsets of the species with which more generalized species interact. It ranges from 0 (non-nestedness) to 100 (perfect nestedness).
Interaction diversity (H)	Shannon diversity of links for a network node $i (H_i = -\sum_j p_{ij} \cdot \ln p_{ij})$, where $p_{ij} = \frac{n_{ij}}{\sum_j n_{ij}}$ and n_{ij} is the interaction
	frequency between node <i>i</i> and <i>j</i> ; here number of <i>j</i> 's pollen grains carried by insect node <i>i</i>) or Shannon diversity of
	links for the whole network $(H_2 = -\sum_i \sum_j q_{ij} \cdot \ln q_{ij})$, where $q_{ij} = \frac{n_{ij}}{\sum_i \sum_j n_{ij}}$ (Bersier <i>et al.</i> 2002).
Interaction evenness (E_2)	Shannon's evenness of link frequency distribution in the whole network $(E_2 = \frac{H_2}{\ln(I)})$, where <i>I</i> is the total number of
	links in the network and $\ln(I)$ the maximum diversity possible in the network i.e. H_{max}). It measures the heterogeneity of interaction frequencies, ranging from 0 (uneven network) to 1 (uniform network). An uneven network is one with high skewness in the distribution of link frequencies (Tylianakis <i>et al.</i> 2007).

References in Table S1:

- Almeida-Neto, M., Guimarães, P., Guimarães Jr., P.R., Loyola, R.D. & Ulrich, W. (2008). A consistent metric for nestedness analysis in ecological systems: reconciling concept and measurement. *Oikos*, **117**, 1227-1239.
- Bersier, L.F., Banašek-Richter, C. & Cattin, M.F. (2002) Quantitative descriptors of food-web matrices. *Ecology*, 83, 2394-2407.
- Dunne, J.A., Williams, R.J. & Martínez, N.D. (2002) Food-web structure and network theory: The role of connectance and size. *Proceedings of the National Academy of Science of the United States*, **99**, 12917-12922.
- Tylianakis, J.M., Tscharntke, T. & Lewis, O.T. (2007) Habitat modification alters the structure of tropical host-parasitoid food webs. *Nature*, **445**, 202-205.



Insect species	Label	Order	Site	a	\mathbf{L}_{sp}	$\overline{L_i}$	null $\overline{L_i}$	TNW	WIC	BIC	WIC/TNW*	SPO	ю
Apis mellifera	api.mel	HYM	CN	8	12	3.25	9.84	0.76	0.12	0.63	0.16	0.05	0.08
<i>Attalus</i> sp.	att.sp	COL	CN	6	4	1.00	3.59	1.13	0.00	1.13	0.00	0.02	0.0
Eristalis tenax	eri.ten	DIP	CN	5	15	4.20	13.59	1.84	1.11	0.73	0.60	0.06	0.0
Eupeodes corollae	eup.cor	DIP	CN	5	10	3.40	7.08	0.71	0.40	0.31	0.56	0.06	0.1
Exoprosopa bowdeni	exo.bow	DIP	CN	5	9	4.80	5.16	0.83	0.79	0.04	0.95	0.06	0.3
Halictus spp.	hal.sp	HYM	CN	5	10	4.40	9.80	1.19	0.37	0.82	0.31	0.06	0.2
Halictus vestitus	hal.ves	HYM	CN	5	11	5.40	11.00	1.76	0.89	0.87	0.51	0.06	0.2
Lasioglossum nitidulum hammi	las.nit	HYM	CN	6	12	4.00	9.09	1.41	0.89	0.52	0.63	0.06	0.1
Mordellistena sp.	mor.sp	COL	CN	9	10	2.44	9.10	1.35	0.60	0.75	0.44	0.06	0.0
Oedemera flavipes	oed.fla	COL	CN	10	13	4.20	12.95	1.97	0.81	1.16	0.41	0.06	0.1
Paragus tibialis	par.tib	DIP	CN	5	12	5.40	11.61	1.69	1.00	0.68	0.59	0.07	0.2
Plagyolepis pygmaea	pla.pyg	HYM	CN	5	5	1.80	3.72	1.00	0.09	0.91	0.09	0.04	0.1
<i>Sphaerophoria</i> sp.	sph.sp	DIP	CN	6	15	4.17	12.96	1.32	0.54	0.78	0.41	0.05	0.0
Stomorhina lunata	sto.lun	DIP	CN	5	7	3.00	6.90	0.92	0.52	0.40	0.56	0.05	0.2
Anthidium cingulatum	ant.cin	HYM	РС	5	13	5.00	9.84	1.42	1.04	0.37	0.74	0.07	0.1
Apis mellifera	api.mel	HYM	PC	5	14	3.80	11.47	0.48	0.13	0.35	0.27	0.07	0.0
Halictus spp.	hal.sp	HYM	PC	5	10	4.80	9.22	0.91	0.52	0.39	0.57	0.06	0.2
Megachile pilidens	meg.pil	HYM	PC	6	13	5.00	8.78	0.93	0.80	0.13	0.86	0.07	0.1
Oedemera flavipes	oed.fla	COL	PC	6	14	3.00	12.99	1.81	0.71	1.11	0.39	0.05	0.0
Sphaerophoria sp.	sph.sp	DIP	PC	5	14	5.00	9.84	0.40	0.17	0.23	0.42	0.08	0.1
Tapinoma nigerrimum	tap.nig	HYM	РС	6	7	2.00	2.80	0.14	0.01	0.14	0.05	0.05	0.0

1 Table S2 List of individual specialization and overlap indices calculated for the 21 insect species of flower-visitors selected from our networks.

2

3	HYM: Hymenoptera; COL: Coleoptera; DIP: Diptera; CN: Coma de n'Arbona; PC: Passadís de Ses Clotades; a: number of individuals captured for pollen
4	load analysis; L_{sp} : linkage level of species; $\overline{L_i}$: mean linkage level of conspecific individuals; null $\overline{L_i}$: mean linkage level of conspecific individuals after

5 1,000 randomizations under the null hypothesis that individuals act as generalists sampling from species pollen resource distribution proportions; TNW: total

- niche width; WIC: within-individual niche component; BIC: between-individual niche component; WIC/TNW: degree of individual specialization; SPO: 1
- 2 interspecific overlap; IO: intraspecific overlap.
- * All WIC/TNW empirical values reported were below null WIC/TNW values obtained from 1,000 randomizations. 3
- 4