

1 Running title: Compartmentalization in invaded plant–pollinator networks

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3 **Consequences of plant invasions on compartmentalization and species' roles in plant–**
4 **pollinator networks**

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26 Summary

27 Compartmentalization – the organization of ecological interaction networks into subsets of
28 species that do not interact with other subsets (true compartments) or interact more frequently
29 among themselves than with other species (modules) – has been identified as a key property
30 for the functioning, stability and evolution of ecological communities. Invasions by
31 entomophilous invasive plants may profoundly alter the way interaction networks are
32 compartmentalized. We analyzed a comprehensive dataset of 40 paired plant–pollinator
33 networks (invaded vs. uninvaded) to test this hypothesis. We show that invasive plants have
34 higher generalization levels with respect to their pollinators than natives. The consequences
35 for network topology are that rather than displacing native species from the network, plant
36 invaders attracting pollinators into invaded modules tend to play new important topological
37 roles (i.e. network hubs, module hubs and connectors) and cause role shifts in native species,
38 creating larger modules that are more connected among each other. While the number of true
39 compartments was lower in invaded compared to uninvaded networks, the effect of invasion
40 on modularity was contingent on the study system. Interestingly, the generalization level of
41 the invasive plants partially explains this pattern, with more generalized invaders contributing
42 to a lower modularity. Our findings indicate that the altered interaction structure of invaded
43 networks makes them more robust against simulated random secondary species extinctions,
44 but more vulnerable when the typically highly connected invasive plants go extinct first. The
45 consequences and pathways by which biological invasions alter the interaction structure of
46 plant–pollinator communities highlighted in this study may have important dynamical and
47 functional implications, for example, by influencing multi-species reciprocal selection
48 regimes and co-evolutionary processes.

49

50 Key words

51 Biological invasions of mutualistic interaction networks, , exotic species, nestedness,
52 pollination, robustness, specialization

53

54 **Introduction**

55 Despite the crucial role that species interaction networks play for the maintenance of
56 biodiversity [1] and the functioning and stability of ecosystems [2,3] we still know very little
57 about the consequences of different components of global change on its structure and
58 dynamics [4,5].

59 A recurrent property in the organisation of complex biological systems ranging from
60 metabolic [6] to species interaction networks [7-9] is compartmentalization.

61 Compartmentalization of ecological networks refers to the existence of subsets of more
62 closely interacting species with relatively few or no interactions to other subsets [8,10]. In
63 food webs, a long history of research has described such subsets, usually termed
64 compartments [e.g. 2,7] or modules [e.g. 8,9]. Recently, it has been shown that also
65 mutualistic networks, such as pollination [8,11-13] or seed dispersal networks [9], exhibit
66 some level of compartmentalization.

67 Compartmentalization has been predicted to stabilize trophic networks [2] and
68 references therein,14], which has recently been corroborated for the persistence [2] and
69 resilience [15] of antagonistic interaction networks, while the persistence of mutualistic
70 networks may in contrast decrease with increasing compartmentalization [15]. Moreover,
71 modules have been suggested as potentially important units of evolution and co-evolution
72 [8,11], and they may have important functional implications for ecosystems [3]. Finally,
73 modularity is also related to other important network properties such as nestedness and
74 connectance [16], which have been proposed to have strong dynamical implications for the
75 coexistence [16], stability [15] and functioning [3] of ecological communities. Consequently,

76 anthropogenic impacts on these properties of species interaction networks may have profound
77 consequences for ecological and evolutionary dynamics and ecosystem functioning..

78 Biological invasions by alien species may strongly affect species interactions, such as
79 those among plants and their pollinators, both directly and indirectly [17]. Pollination
80 interactions are essential for the reproductive success of many plant species and of crucial
81 importance for the maintenance of the diversity and functioning of most terrestrial ecosystems
82 [18]. Most plant and pollinator species exhibit a higher degree of opportunism and thus
83 generalization in their interactions with mutualistic partner species than previously thought
84 [18], facilitating the integration of alien species into native plant–pollinator interaction
85 networks [19-25]. Pollinators use, but appear not to prefer, alien plant species over natives
86 [26]. However, when these alien plants become invasive (*sensu* 27) and have highly attractive
87 flowers present in high abundances in the community, they may have profound effects on
88 pollinators and their interactions with native plants [28]. Hence, the impact of invasions on
89 network topology may critically depend on the generalization level of the invaders [22].
90 Super-generalist invaders [*sensu* 29] may cause a fusion of previously separate compartments
91 or modules, as predicted by Olesen *et al.* [8], resulting in fewer – but larger – modules (i.e.
92 more species forming a module). However, overall module number may not be altered, but
93 the boundaries among modules may be “blurred”, resulting in more connected modules and
94 thus more cohesive interaction networks. Alternatively, if the invasive species monopolize
95 generalist native partner species from the core of nested networks [22], the invader could lead
96 to a fragmentation of the network into more but smaller modules. Invaders might also form
97 new modules either by usurping species from existing modules, thereby reducing module size,
98 or by attracting pollinators (not previously present) into the invaded communities, thereby
99 increasing network size (i.e. the total number of interacting species). This may not only affect
100 the number of modules and modularity, but also related important structural and dynamical

101 network properties, such as pollinator-plant ratio, nestedness, connectance and network
102 robustness [e.g. 8,30].

103 A further important, yet largely unexplored question is how invasion may impact upon
104 the individual topological roles native species play in the network. Species specific
105 contributions to module formation offer a perfect framework to answer this question. Even if
106 the modular structure of the interaction networks should be robust against the invasion of
107 alien species, invaders are likely to displace native species from topologically important roles.
108 Such role shifts of species may differ between plant and pollinators, possibly resulting in
109 predictable changes in the proportion of plants and pollinators occupying different roles in the
110 network, with potentially profound implications for species persistence, network functioning
111 and reciprocal selection regimes.

112 Here, we investigate consequences and underlying mechanisms of plant invasions on
113 the compartmentalization of plant–pollinator networks and the associated topological roles of
114 the species forming these networks. To this end, we analysed 40 networks including a range
115 of different alien invader species, native communities and geographical regions. We
116 specifically addressed the following questions: (1) how does plant invasion affect network
117 structure (i.e. nestedness, connectance and pollinator-plant ratio), and in particular the level of
118 compartmentalization, and the number and size of compartments and modules and (2) is this
119 modulated by the behaviour (e.g. generalization level) of the invader species in the
120 community? or (3) are these effects driven by increased size of invaded networks, (4) do
121 invasive plants exhibit different topological roles compared to natives and how does plant
122 invasion alter the composition of topological roles played by native plant and pollinator
123 species with respect to network compartmentalization? and, finally, (5) does plant invasion
124 alter the robustness of these networks against secondary species extinctions under different
125 scenarios of species loss?

126

127 Material and Methods128 *Database*

129 We analyzed 40 plant–pollinator networks representing 20 independent pairs of networks
130 from seven study systems, each pair consisting of a network invaded by at least one invasive
131 plant species [*sensu* 27] (hereafter “invaded”) and a network without any alien plants present,
132 or, in four network pairs [see 20], with only alien, non-invasive plants present in the network
133 (“uninvaded”) (Electronic Supplementary Information 1a). The dataset includes our own and
134 published data that fulfilled the following criteria: (1) network data are collected following a
135 paired (invaded vs. uninvaded) approach sampled at different locations within the same region
136 that represent identical habitat types and similar native plant communities; (2) sampling
137 method, period and effort is identical within a network pair. The main characteristics of the
138 analysed plant–pollinator systems are described in Electronic Supplementary Information 1a;
139 for detailed description of field sampling procedures see publications listed there.

140 All plant and pollinator species included in the analysed networks are identified at the
141 species or morpho-species level. Interaction frequency was quantified in all networks as
142 visitation rate, i.e. the total number of visits per sampling time. Flower-visiting animals were
143 regarded as pollinators, if they were observed contacting the reproductive parts of the flower.
144

145 *Network analysis*

146 A traditional measure of compartmentalization is the number of “true” compartments [*sensu*
147 31], defined as the number of subsets in the network with no link to any other subset (i.e.
148 Jordan blocks in the mathematical nomenclature). Although this metric is sensitive to
149 sampling thoroughness, our paired networks have equal sampling effort making relative
150 comparisons meaningful. Furthermore, it has been widely used to analyse
151 compartmentalization in food webs as a “coarse” measure of compartmentalization [e.g. 32].
152 Following [32] and [33] we use the simpler term “compartment” instead of true compartment

153 hereafter. Compartment number was calculated using the *network level* function of the
154 *bipartite* package [34] implemented in R [35].

155 Subsets of highly linked nodes that have much weaker links to other subsets of nodes
156 in the network were defined as “topological modules” (*sensu* [36]; hereafter referred to as
157 “modules”). To determine the level of modularity and the number and composition of
158 modules in the plant–pollinator networks we used the software NETCARTO based on the
159 functional cartography method [6]. Modules are identified by maximising a measure of
160 modularity M using simulated annealing (SA), a stochastic optimization technique based on
161 presence–absence data. M increases with increasing link density within modules and
162 decreasing connectedness between different modules. The results of this algorithm are robust,
163 yielding almost identical partitions in different runs [6,11]. The software identifies modules
164 with an accuracy of 90% [10], and is among the most accurate module-detecting algorithms
165 available to date [10,36]. The algorithm defines modules as subsets of both plants and
166 pollinators that are highly connected to each other, rather than separate subsets of plants and
167 subsets of pollinators as a function of their shared interactions, and thus modules have a clear
168 biological meaning [8,13]. The SA-algorithm also assigns each species to a topological role in
169 a network based on the within-module degree z_i (the standardized number of links a species i
170 is connected with others in its own module) and the among-module connectivity c_i (measuring
171 how connected a species i is to all modules) [6,8]. A network hub is not only highly linked to
172 species of its own but also species of other modules, making it important for the connectivity
173 of among species in both its own module and the entire network [8]. A module hub plays an
174 important role in its own module, increasing its coherence. A connector species is important
175 for among-module connectivity, and consequently network coherence, but plays an inferior
176 role within its own module. Peripheral species have all or most of their relative few
177 interactions within their own module, playing a topologically inferior role in the network [8].

178 For further details of the functional cartography method see Electronic Supplementary
179 Information 1b.

180 In order to examine the role of the generalization level of the invader species driving
181 changes in network compartmentalization we calculated standardized species degree (SD , i.e.
182 the number of interaction partners of a species relative to the maximum possible) as a
183 standardized measure of species generalisation in bipartite ecological networks [37]. For these
184 analyses, five pairs of networks – those studied by [20] and [24] – invaded by several alien
185 plant species, for which an unambiguous assignment to either alien invasive or alien non-
186 invasive was not possible, were excluded .

187 Due to its proposed important dynamical implications and relatedness to
188 compartmentalization [16] we also analysed the degree of nestedness in the compiled plant–
189 pollinator networks. Bipartite ecological networks are nested if little connected species tend to
190 interact with a proper subset of interaction partners of highly connected species (e.g. [16,30]).
191 Nestedness was calculated as *BINMATNEST* temperature using the *R* package *bipartite* [38];
192 results of other nestedness metrics such as *NODF* were qualitatively identical.

193 To assess the robustness of networks to secondary species extinctions we used the
194 robustness index R proposed by [39]. The index is a quantitative modification of the concept
195 of secondary extinction curves (or attack tolerance curve (ATC) [39]) introduced by [40] to
196 assess the tolerance of bipartite mutualistic interaction network to secondary extinctions. The
197 ATC curve is based on the fact that if a given proportion of one mutualistic guild (e.g. plants
198 or pollinators) is eliminated from the network (“attacked”), a certain proportion of species of
199 the partner guild become extinct [39]. The index R calculates the area below the ATC curves.
200 $R \rightarrow 1$ corresponds to a curve that decreases very slowly until the point at which almost all
201 species are eliminated, while $R \rightarrow 0$ corresponds to curve that abruptly declines already after
202 eliminating a single species. In addition to simply cumulatively removing species in a random
203 order from the network [e.g. 41], we tailored the extinction order for the analysis of

204 compartmentalized networks to allow us to account for the topological role of a species with
205 respect to compartmentalization. Thus, either peripheral species were removed first, followed
206 by connectors, module hubs and finally network hubs, or species were removed in the
207 opposite order, starting with network hubs. Secondary extinction was modelled separately for
208 plants and pollinators, using 500 simulations for the calculation of R . It is clear that the
209 modelled species eliminations not necessarily represent real extinctions events in nature,
210 because not all plant species require animal pollination for population persistence, and
211 because plant and pollinator mutualists may “switch” interaction partners to a certain degree
212 following the loss of one or a few species (re-wiring [42,43] and adaptive foraging [44]).
213 Nevertheless, comparisons of such species removal simulations, especially when accounting
214 for the topological role in the removal order of species, can provide relative estimates of the
215 tolerance of mutualistic communities with respect to network functioning [41].

216

217 *Statistical analysis*

218 To address research questions (1) and (5) linear mixed effect models using the *lme*-function of
219 the *nlme* package [45] in *R* were fitted to test the effect of invasion status (invaded vs.
220 uninvaded) on the following response variables: number of compartments (log-transformed),
221 modularity (M , logit-transformed [46]), number of modules, nestedness (log-transformed),
222 connectance (logit-transformed), pollinator-plant ratio (log-transformed) and robustness (R).
223 Invasion status was treated as fixed effect and site nested within study system as random
224 effects. Further, we analysed the effect of invasion status on module size (i.e. the number of
225 species per module) and among-module interactions (i.e. the number of links of all species of
226 a module that are to species of other modules) (both log-transformed) at the module level. For
227 these models, network identity (nested within site and study system) was also included as a
228 random effect in the model. The presence of invasive plant species was determined for each
229 module, and this variable (module with or without invasive plant species) and its interaction

230 with invasion status were included as fixed effects in the model. To examine which factors
231 accounted for the effect of invasion on module size, we further analysed the number of plant
232 species per module and the number of pollinator species per module separately. To test
233 research question (3), i.e. whether the effects of invasion were mediated by changes in
234 network size, we used the same models described above but included network size (before
235 invasion status in the sequentially fitted model) as covariate. Thus, we tested whether the
236 variation explained by invasion status (in the model without the co-variate network size) is
237 actually explained by network size and whether invasion status still explains a significant
238 part of the residual variation (not explained by the co-variate network size). To analyse the
239 effects of invader generalization on the response variables (research question (2)), invaded
240 networks were modelled with *SD* of the invader as fixed effect and the same random effects
241 as described above.

242 To assess how plant invasion promoted shifts in the proportion of species with a
243 particular topological role (network hubs, module hubs, connectors or peripherals) (research
244 question (4)), separate generalized linear mixed-effects models (GLMM) for each role with
245 binomial error distribution and logit link function, were fitted using the *lmer*-function
246 implemented in the *R*-package *lme4* [47] with invasion status as fixed factor and site nested
247 within study system as random effects. To further explore the consequences of invasion on the
248 topological role shifts of individual species, we report species shifts for the subset of all plant
249 and pollinator species occurring in both the uninvaded and the invaded network of each site
250 pair. In order to test whether the probability of role shifts differed between plants and
251 pollinators, a GLMM with a binomial error (change or no change of role) and trophic level
252 (plant or pollinator) as fixed effects, and site nested within study as random effects was fitted.
253 None of the GLMM was overdispersed. Inference was based on likelihood ratio tests [48]. For
254 all analyses, model fit was assessed by plotting the residuals against the predicted values.
255 Means \pm 1 standard error are reported.

256

257 **Results**258 *Impacts of invasion on network compartmentalization*

259 Network size increased in invaded compared to uninvaded networks ($F_{1,19} = 6.22$, $P = 0.022$),
260 and this was caused by an increase in the number of pollinator species from 29.8 ± 5.1 to 39.1
261 ± 5.9 ($F_{1,19} = 7.52$, $P = 0.013$) – but not in the number of plant species (invaded: 12.9 ± 2.5 ,
262 uninvaded: 11.6 ± 3.3 ; $F_{1,19} = 1.58$, $P = 0.224$). The latter result was expected because the
263 sampling was designed to compare sites with similar plant communities. Consequently,
264 pollinator–plant ratio was slightly, but statistically not significantly increased in invaded
265 networks (3.5 ± 0.4 vs. 3.1 ± 0.4 ; $F_{1,19} = 2.58$, $P = 0.125$). Both connectance and nestedness
266 were strongly positively correlated ($r = 0.79$) and negatively related to network size
267 (connectance: $F_{1,18} = 47.00$, $P < 0.001$; nestedness: $F_{1,18} = 87.17$, $P < 0.001$), which accounted
268 for a tendency towards decrease in connectance (0.22 ± 0.01 to 0.19 ± 0.02 ; $F_{1,19} = 3.74$, $P =$
269 0.068) and nestedness (25.7 ± 2.4 to 21.7 ± 1.6 ; $F_{1,19} = 3.57$, $P = 0.074$), which disappeared
270 after accounting for network size (connectance: $F_{1,18} = 0.43$, $P = 0.521$, nestedness: $F_{1,18} =$
271 0.06 , $P = 0.806$). Neither connectance nor nestedness were significantly affected by invader
272 generalization ($P > 0.140$).

273 The number of compartments in invaded networks tended to be lower than in
274 uninvaded ones (invaded: 1.35 ± 0.13 , uninvaded: 1.70 ± 0.18 ; $F_{1,19} = 4.17$, $P = 0.053$, also
275 after accounting for the increased size of the invaded networks ($F_{1,18} = 3.05$, $P = 0.098$).
276 Invader generalization had no detectable effect on compartment number of invaded networks
277 ($F_{1,9} = 0.03$, $P = 0.857$).

278 Neither modularity (M ; $F_{1,18} = 0.12$, $P = 0.738$) nor the number of modules
279 (uninvaded: 5.0 ± 0.3 , invaded: 5.1 ± 0.3 ; $F_{1,18} = 0.15$, $P = 0.703$) were significantly altered
280 following plant invasion. However, the direction and magnitude of invasion effects were
281 contingent on the system studied: M was significantly reduced in three out of seven study

282 systems, whilst it increased in only one system (Fig. 1; invasion \times study system: $F_{6,13} = 3.45$,
283 $P = 0.029$). This variation among study systems was at least partly driven by invader
284 generalization: modularity ($F_{1,9} = 12.96$, $P = 0.006$), but not average number modules ($F_{1,9} =$
285 0.25 , $P = 0.628$), decreased with the level of generalization of the invader species (Fig. 2,3).
286 M was not significantly correlated with nestedness ($r = 0.16$) or connectance ($r = -0.03$) ($P >$
287 0.337).

288 Modules of invaded networks consisted of more species (10.4 ± 0.7) than uninvaded
289 ones (8.4 ± 0.8) ($F_{1,18} = 7.68$, $P = 0.014$; Fig. 3). This increase in module size was mainly due
290 to a higher number of pollinator species within a module (7.8 ± 0.6 vs. 6.0 ± 0.6 ; $F_{1,18} = 7.20$,
291 $P = 0.015$), while the number of plant species per module did not significantly change
292 (invaded networks: 2.6 ± 0.3 ; uninvaded: 2.3 ± 0.3). Modules containing invasive plant
293 species (12.4 ± 1.3) were larger compared to modules without invasive plant species ($8.6 \pm$
294 0.6) (Fig. 3; $F_{1,158} = 7.13$, $P = 0.008$). Indeed, module size of modules without invasive plant
295 species did not significantly differ between invaded (9.0 ± 0.8) and uninvaded (8.4 ± 0.8)
296 networks (Fig. 3). Furthermore, plants and pollinators of invaded networks interacted more
297 with pollinators and plants, respectively, belonging to other – not their own – modules ($F_{1,19} =$
298 9.32 , $P = 0.007$). This connectivity among modules was still higher in invaded compared to
299 uninvaded networks ($F_{1,18} = 4.78$, $P = 0.042$) after accounting for variation explained by
300 network size, but was not significantly influenced by invader generalization ($F_{1,9} = 0.74$, $P =$
301 0.413).

302

303 *Effects of invasion on species roles*

304 Invasive plants were more generalized ($SD = 0.40 \pm 0.03$) with respect to their flower visitors
305 than native plants ($SD = 0.19 \pm 0.01$). In 33% of networks invaded by a single plant species
306 the invader played a new important topological role as a network hub, module hub or
307 connector that was not occupied in the uninvaded network. Indeed, the invader acted as a

308 network hub in 20% of these invaded networks, as a module hub in 33.3%, as a connector in
309 20% and only in 26.7% as a peripheral species. By contrast, 80.0% of native plant species
310 were peripherals. The average proportions of topologically important species (“generalists”
311 *sensu* [11]: network hubs, module hubs and connector species) were slightly, but statistically
312 not significantly (all $P > 0.1$), higher on average in invaded compared to uninvaded networks
313 (Fig. 4). However, invaded networks were more likely to contain a module hub (85%, $\chi^2_1 =$
314 4.24, $P = 0.040$), usually the invader species, or a connector (100%, $\chi^2_1 = 4.07$, $P = 0.044$)
315 than uninvaded networks (65% with module hub, 75% with connector), but not a network hub
316 (25% compared to 15% in uninvaded networks, $\chi^2_1 = 1.87$, $P = 0.175$).

317 Of the 469 native plant and pollinator species present in both the uninvaded and the
318 invaded network within a site pair, 111 species (23.7%) showed a role shift following
319 invasion, with plants showing more shifts (31.9%) than pollinators (19.2%) ($\chi^2_1 = 7.65$, $P =$
320 0.006). Most role shifts of native plant species were from important roles to peripherals
321 (53.1%), while 28.6% were from peripherals to important roles. In contrast, slightly more
322 role shifts of pollinator species were from peripheral to important roles (54.2%), while 45.8%
323 of shifts were from important to peripheral.

324

325 *Impacts of invasion on network robustness*

326 Invaded networks were more robust against the removal of either pollinators ($R = 0.751 \pm$
327 0.025) or plants ($R = 0.629 \pm 0.022$) than uninvaded networks (pollinators removed: $R = 0.678$
328 ± 0.025 , $F_{1,19} = 6.11$, $P = 0.023$; plants removed: $R = 0.562 \pm 0.027$, $F_{1,19} = 5.02$, $P = 0.037$)
329 when peripheral species were removed first and network hubs last. After accounting for
330 variation in network size, there was still a trend for a higher robustness of invaded networks
331 (pollinators removed: $F_{1,18} = 4.14$, $P = 0.057$; plants removed: $F_{1,18} = 3.15$, $P = 0.093$). Similar
332 effects of plant invasion were found when species were removed randomly (plants removed:
333 $F_{1,19} = 8.25$, $P = 0.010$; pollinators removed: $F_{1,19} = 11.94$, $P = 0.003$). In contrast, when

334 network hubs were removed first and peripheral species last, the effect of invasion on
335 robustness against the removal of pollinator (R invaded: 0.666 ± 0.023 , R uninvaded: $0.656 \pm$
336 0.021) or plant species (R invaded: 0.402 ± 0.20 , R uninvaded: 0.388 ± 0.030) was no longer
337 significant (all $P \geq 0.5$).

338

339 **Discussion**

340 Entomophilous invasive plants are usually characterized by showy flowers and high
341 abundances in the communities. We show here that these characteristics tend to confer them
342 higher generalization levels than natives. The consequences for network topology are that
343 invasions by such highly generalized plants decrease compartmentalization and increase
344 connectivity among modules. However, rather than displacing native species from the
345 network, plant invaders tend to play new important topological roles creating larger modules
346 that are more connected among each other. We discuss several structural and dynamical
347 consequences of how plant invasions alter the way these networks are compartmentalized and
348 identify some of the underlying mechanisms.

349

350 *Consequences of invasions on network structure*

351 As predicted, plant invasion resulted in a decline in the number of true compartments,
352 increased connectivity among modules and tended to reduce modularity, but only in the plant-
353 pollination networks invaded by highly generalized alien plants. Indeed, invader
354 generalization showed to be a key predictor of changes in modularity confirms our hypothesis
355 based on existing evidence for a positive relationship of interaction specialization and network
356 compartmentalization mostly from trophic interaction networks [33,49].

357 As a consequence, modules of invaded networks were larger and more connected
358 among each other. The main mechanistic process behind these changes on network structure
359 was that the alien plant invaders attracted new pollinator species into the invaded plant

360 communities resulting in an increase in network size due to a higher number of pollinator –
361 but not plant – species, which in turn was associated with a tendency towards lower
362 connectance and nestedness [21, but see 12,23]. This increase was largely restricted to the
363 modules containing these highly generalized plant invaders, increasing the size of these
364 modules but without significantly altering the size of other modules without alien plant
365 invaders. These findings highlight that alien plant invaders not simply usurp pollinator species
366 from native plant species [17] but also new pollinator species are attracted into the invaded
367 communities. A similar process has been predicted for enhanced pollinator population sizes
368 through increased resource availability offered by abundantly flowering alien plant invaders
369 [26]. With the number of species also the absolute number of interactions between plants and
370 pollinators increased in invaded compared to uninvaded networks, both within and among
371 modules. Consequently, plant invasion did not cause a fusion of modules as hypothesized by
372 Olesen *et al.* [8] but rather resulted in larger modules that were more strongly connected
373 through interactions. At the community level, this increase of realized new links in invaded
374 modules did not compensate the lack of overall realized links associated with the significantly
375 higher number of species of invaded networks, resulting in a tendency towards lower overall
376 network connectance.

377 Thus, high pollinator attraction and level of generalization of the plant invader showed
378 to be a key predictor of changes in modularity. Indeed, most of the principal alien plant
379 invaders acted as super-generalists [*sensu* 29] in the invaded plant–pollinator networks, such
380 as *Carpobrotus affine acinaciformis* or *Opuntia stricta* in Spanish continental mainland [21],
381 *O. maxima* in Balearic island communities [22], and *Impatiens glandulifera* in temperate
382 riparian communities [50]; the only exception seems to be *Opuntia dillenii*, which showed to
383 act as a specialist in the invaded networks in the Canary Islands [22] and tended to increase
384 modularity when compared to uninvaded communities. However, the mechanistic process
385 driving changes in network compartmentalization revealed here for plant invasions may also

386 apply more generally to processes by which mutualistic communities become dominated by a
387 single or a few species showing particularly high abundance and/or attractiveness, e.g. by
388 species with boom and bust cycles, irrespective of whether this dominant species is alien or
389 not. Such positive correlations between species abundance, interaction frequency and
390 generalization is predicted by the theory of interaction neutrality and frequently observed in
391 plant–pollinator networks (see [30] and references therein). Conversely, we show that if alien
392 species do not become dominant (i.e. invasive *sensu* [27]) – as in the studies analysed here –,
393 strong effects on compartmentalization appear unlikely.

394 While compartmentalization in antagonistic interaction networks such as food webs
395 may increase their persistence and resilience [2,15], partly by buffering the propagation of
396 species extinctions throughout the webs [2], recent research suggests that, in contrast, the
397 persistence of mutualistic networks may decline with higher levels of modularity [15]. Here,
398 we found that the modules of invaded plant–pollinator networks were more connected by
399 links with each other, probably making the networks more cohesive and robust against
400 fragmentation into weakly or not at all connected modules. However, the effects of
401 modularity on the functioning and stability of networks is still not fully understood [51]. We
402 tentatively explore this avenue with a simple secondary species extinctions simulation. Our
403 analysis suggests that increased numbers of pollinator species, which tended to act more often
404 as connectors of modules in invaded networks, was a key driver of enhanced robustness
405 against secondary species extinctions of invaded networks. Interestingly, this effect depended
406 on the extinction order and topological role of the removed species: while present when the
407 extinction order was from the least to the most connected species, which may be considered as
408 a realistic extinction scenario in many real-world situations [52], it disappeared if the most
409 connected species, i.e. network hubs, were removed first from the networks. In the latter case,
410 networks collapse faster because the impact of the early loss of key species accounting for
411 most of the network coherence is so strong that subtle differences in the interaction structure

412 between invaded and uninvaded networks is no longer relevant. This corroborates findings of
413 modelling studies identifying connectivity of alien plants as a principal driver of the
414 persistence of species in plant–pollinator networks following simulated alien removal (e.g.
415 [12]). Hence, it is important to note that despite invasive plants appear to enhance some
416 aspects of network robustness, it makes the networks also more dependent on them and hence
417 more vulnerable [53], if the invasive plants have boom and bust cycles [54] or in the face of
418 eradication programs, although flexibility in foraging behaviour of pollinators may mitigate
419 this vulnerability to some degree [42-44].

420

421 *Topological role shifts of plant and pollinator species*

422 The overall proportions of the four different topological roles species were very similar in the
423 uninvaded communities to those reported in other plant–pollinator networks [8,11,12]. Plant
424 species – as a logical consequence of the typically several times higher number of pollinator
425 than plant species in plant–pollinator networks [e.g. 30] – played on average more
426 topologically important roles than pollinator species, which were more often peripherals [8].
427 Indeed, not pollinator species acted as a module hub; this topological role was exclusively
428 occupied by generalist plant species.

429 Most of the principal plant invader species (73%) played topologically important roles,
430 while approximately every fourth played only a peripheral role in the invaded networks,
431 largely confirming previous findings that abundantly flowering invasive plants are generally
432 well integrated in native plant–pollinator networks [12,19-25]. The well-connected principal
433 invaders formed mostly new modules and became module hubs, but also linked existing
434 modules as connectors or did both as super-generalist network hubs. Closer inspection of the
435 network pairs revealed that the invaders either displaced natives from these roles or, in one
436 third of the network pairs, played new important roles as network or module hubs – roles that
437 were not occupied by native species prior to invasion especially in some of the smaller

438 networks. Yet, invasion may also cause individual species to shift their topological role with
439 potentially important dynamical implications for individual species. Although with this
440 dataset we could not directly compare species' roles prior and after invasion, our study of
441 paired networks suggests that a considerable proportion of native pollinator species
442 (approximately 20%), and an even higher proportion of native plant species (roughly 30%)
443 present in both the invaded and the uninvaded network within a geographic pair showed such
444 a role shift. Our analysis also reveals that plant and pollinator species differed in the direction
445 of role shifts. Thus, invasion resulted in a shift in the trophic composition of the connector
446 role that forms the “glue” [8] holding different modules together. While the first finding is in
447 agreement with several studies showing that attractive invasive plant species can usurp some
448 generalized pollinator species from native plants [17], the second result suggests that present
449 pollinators include resources of the invasive plants in their diet and thus become more
450 generalized in their visits of plant species across modules (diet expansion hypothesis).
451 Ecological network data available to date notoriously lack sample completeness, affecting
452 most network descriptors [55-57]. This almost certainly affected the number of unconnected,
453 true compartments detected in the analysed networks. However, sampling effort was identical
454 for the uninvaded and the invaded communities within a site pair. Thus, even though the
455 absolute numbers of true compartments may not precisely represent actual numbers, the
456 significantly lower numbers of such compartments in invaded compared to uninvaded
457 networks sampled with the same effort should provide a robust, “coarse” indication for lower
458 compartmentalization of invaded communities.

459 Super-generalist invaders are predicted to play central roles for the evolution and co-
460 evolution in mutualistic networks by enhancing trait convergence [58]. Indeed, since modules
461 might reflect units of co-evolution in mutualistic assemblages [8,11] and may have functional
462 and stability consequences [3,15], several of our key findings with respect to how plant
463 invasion altered the way plant–pollinator communities are compartmentalized may have

464 important dynamical implications. In particular, the blurred module boundaries and reduced
465 compartmentalization in networks invaded by highly generalized alien invader species and the
466 increase in module size and pollinator-plant ratio of invaded modules may alter pollination
467 functions and community dynamics, and influence multi-species reciprocal selection regimes
468 and co-evolutionary processes in the longer term.

469

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478

479 **Data accessibility**

480 Data sets supporting this article: Dryad doi:10.5061/dryad.m26h2

481

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622

623 **Figure legends**

624 **Fig. 1** Mean (± 1 standard error) modularity (M) of “uninvaded” plant–pollinator networks
625 and networks invaded by one or several alien plant species plotted against the seven study
626 systems. M is a measure of the degree to which a network is organized into clearly delimited
627 modules. “Uninvaded” networks contained no aliens (16 networks) or a significantly lower
628 proportion of alien plant species (study system 3) than invaded networks. Information about
629 study systems is given in Electronic Supplementary Information 1a.

630

631 **Fig. 2** Relationship between species generalization of the principal invader plant, measured as
632 standardized degree (SD), and modularity (M) of invaded plant–pollinator networks.

633

634 **Fig. 3** Example of the modular structure of (a) an uninvaded plant–pollinator network and (b)
635 a network invaded by an alien plant invader (*Carpobrotus affine acinaciformis*; large red
636 square). Interaction networks represent Mediterranean shrubland communities sampled at two
637 locations at Cap de Creus, Spain [for details see 24]. Plants are represented by squares
638 whereas pollinators by circles. Different colours represent different topological species’ roles:
639 peripheral species (yellow), connector (green), module hub (pink), network hub (red).

640

641 **Fig. 4** Mean (± 1 standard error) module size (i.e. the number of species forming a module) of
642 modules of uninvaded plant–pollinator networks ($n = 92$), modules of invaded networks not
643 containing alien plant species ($n = 68$) and modules of invaded networks containing alien
644 plant species ($n = 32$).

645

646 **Fig. 5** Proportion of topological species' roles in uninvaded (open circles) and invaded plant–
647 pollinator networks (closed circles). Roles are defined according to their position in the
648 parameter space of within-module degree z and between-module connectivity c : a network
649 hub ($z > 2.5, c > 0.62$) is highly linked to species within its own module and is well connected
650 to species of other modules, making it important for the coherence of both, its own module
651 and the entire network; a module hub ($z > 2.5, c \leq 0.62$) plays an important role within its own
652 module whilst weakly connected to species of other modules; a connector ($z \leq 2.5, c > 0.62$)
653 species is important for among-module connectivity, and consequently network coherence,
654 but plays an inferior role within its own module; a peripheral species ($z \leq 2.5, c \leq 0.62$) plays a
655 topologically inferior role in the network (Olesen *et al.* 2007).

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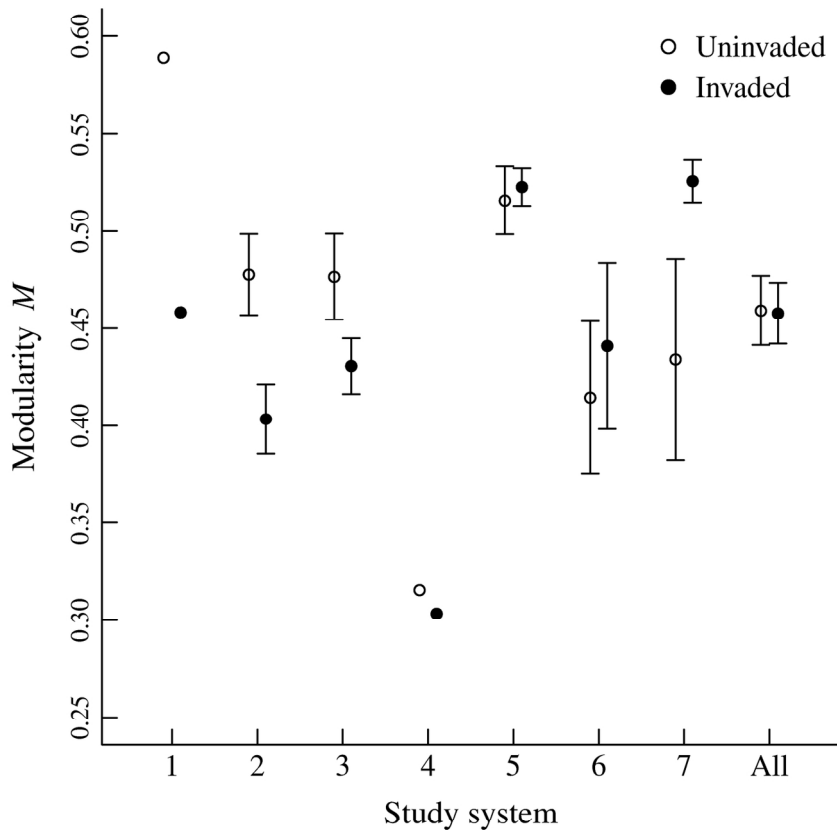
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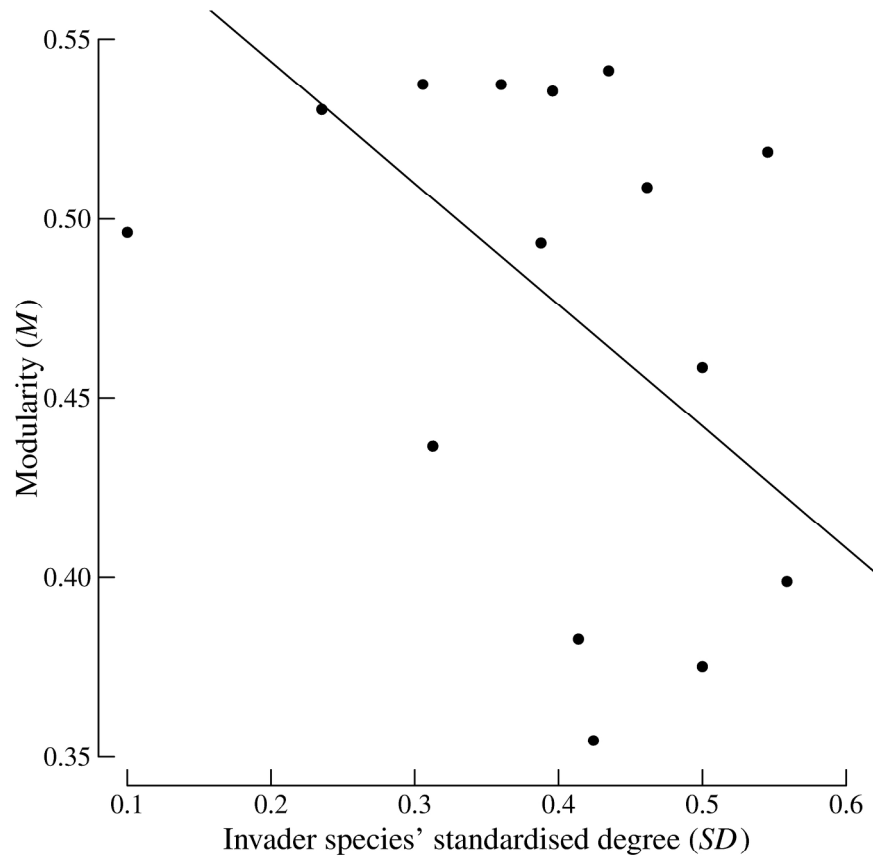
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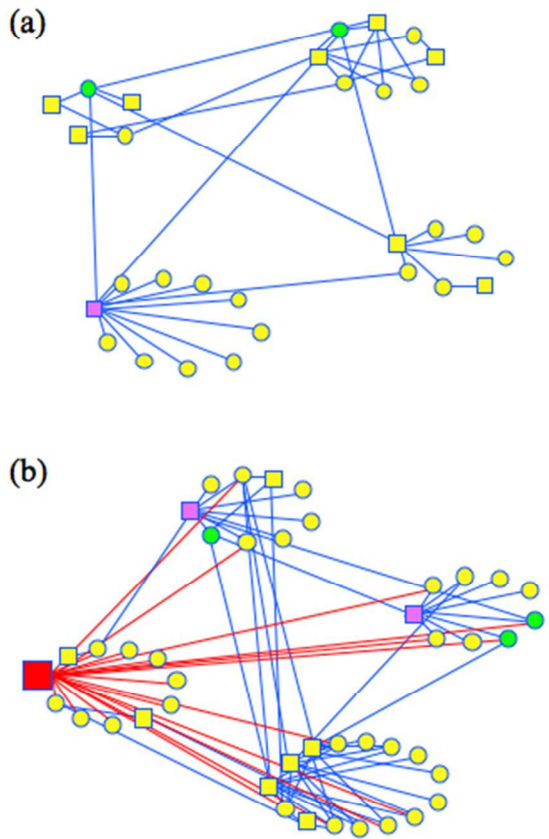
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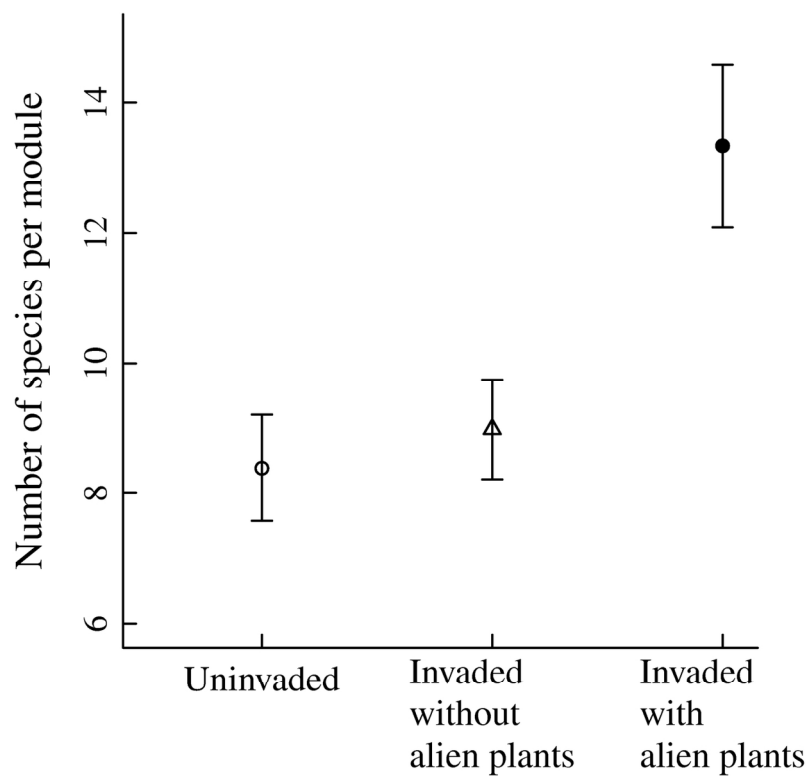


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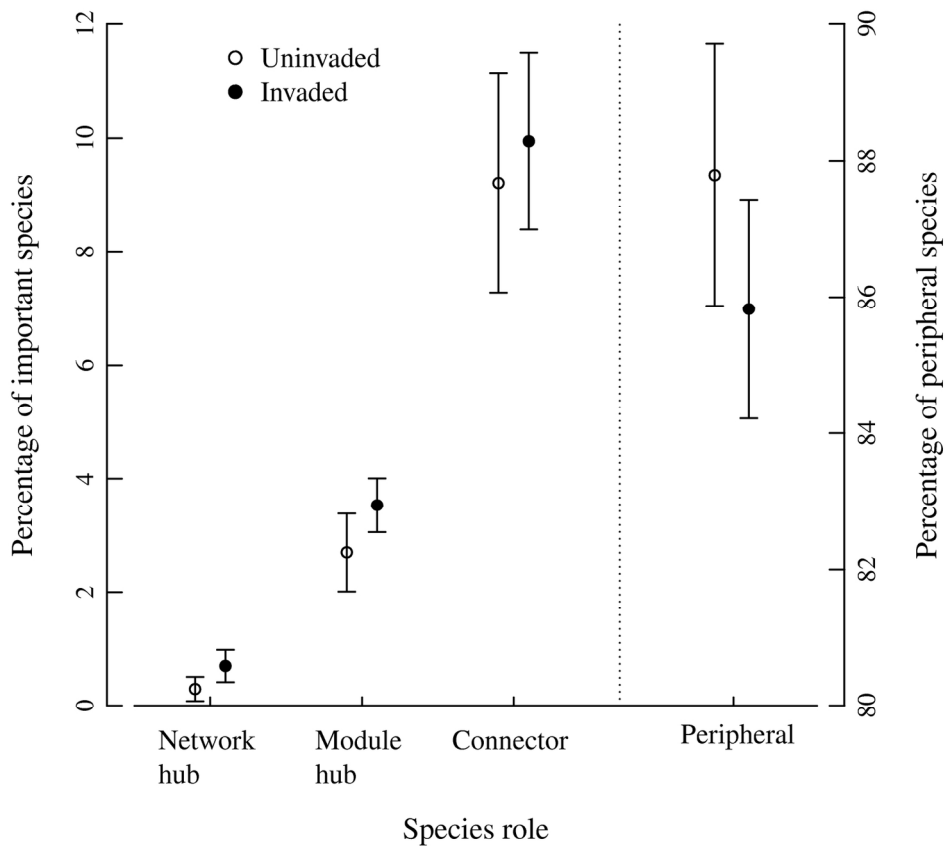


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