

1 **Molecular characteristics of permanganate and dichromate oxidation resistant soil**  
2 **organic matter from a Black C rich colluvial soil**

3

4 Manuel Suárez-Abelenda<sup>a\*</sup>, Joeri Kaal<sup>b</sup>, Marta Camps-Arbestain<sup>c</sup>, Heike Knicker<sup>d</sup>,  
5 Felipe Macías<sup>a</sup>

6

7 <sup>a</sup> Departamento de Edafología e Química Agrícola, Facultade de Bioloxía, Universidade de Santiago de  
8 Compostela, 15782- Santiago de Compostela, Spain

9 <sup>b</sup> Instituto de Ciencias del Patrimonio (Incipit), Consejo Superior de Investigaciones Científicas (CSIC),  
10 Rúa San Roque 2, 15704 Santiago de Compostela, Spain.

11 <sup>c</sup> Institute of Natural Resources, Private Bag 11222, Massey University, Palmerston North 4442, New  
12 Zealand.

13 <sup>d</sup> Instituto de Recursos Naturales y Agrobiología de Sevilla (IRNAS-CSIC), Adva. Reina Mercedes, 10,  
14 41012 Sevilla, Spain.

15

16 \* Corresponding author (M. Suárez Abelenda): Departamento de Edafología e Química Agrícola,  
17 Facultade de Bioloxía, Universidade de Santiago de Compostela, 15782- Santiago de Compostela, Spain.

18 Tel.: +34 881816039. E-mail address: [manuel.suarez@usc.es](mailto:manuel.suarez@usc.es).

19

20 **Abstract**

21 Samples from a colluvial soil rich in pyrogenic material (Black C, BC) in NW Spain  
22 were subjected to  $K_2Cr_2O_7$  and  $KMnO_4$  oxidation and the residual soil organic matter  
23 (SOM) was NaOH-extracted and analyzed using analytical pyrolysis–gas  
24 chromatography–mass spectroscopy (Py-GC/MS) and solid-state  $^{13}C$  cross  
25 polarization–magic angle spinning–nuclear magnetic resonance ( $^{13}C$  CP MAS-NMR) in  
26 order to study the susceptibility of different SOM fractions (fresh, degraded/microbial,  
27 BC and aliphatic) towards these oxidizing agents. NaOH extracts of untreated samples  
28 were also analyzed. Py-GC/MS and  $^{13}C$  NMR indicated that  $KMnO_4$  promotes the  
29 oxidation of carbohydrate products, mostly from degraded/microbial SOM and  
30 lignocellulose, causing a relative enrichment of aliphatic and aromatic structures.  
31 Residual SOM after  $K_2Cr_2O_7$  oxidation contained BC, N-containing BC and aliphatic  
32 structures. This was corroborated by a relatively intense resonance of aromatic C and  
33 some signal of alkyl C in  $^{13}C$  NMR spectra. These results confirm that dichromate  
34 oxidation residues contain a non-pyrogenic fraction mainly consisting of aliphatic  
35 structures.

## 36 1. Introduction

37 Soil organic matter (SOM) is a complex mixture of plant, animal and microbial  
38 tissue, both fresh and at different stages of decomposition (Stevenson 1994; Tabatabai  
39 1996). The molecular structures present in SOM have been a source of debate for many  
40 decades due the analytical difficulties inherent to SOM characterization (Piccolo 1996).  
41 A wide set of methodologies have been used to study SOM composition, including 1)  
42 fractionation methods such as (i) chemolytic techniques (i.e. application of acid-  
43 hydrolysis or extracting agents) that are coupled with colorimetric and/or GC/MS  
44 analyses to identify specific SOM components (polysaccharides, lignin-derived  
45 compounds, amino sugars, extractable- lipids or hydrolysable proteins) (Kögel-Knabner  
46 1995); (ii) physical fractionation into organo- mineral fractions based on particle size  
47 and/or density yields (Christensen 1992; Six et al. 2002); (iii) wet oxidation with  
48 potassium permanganate (KMnO<sub>4</sub>) (Loginow et al. 1987; Tirol-Padre and Ladha 2004),  
49 H<sub>2</sub>O<sub>2</sub> (Eusterhues et al. 2005), Na<sub>2</sub>S<sub>2</sub>O<sub>8</sub> (Eusterhues et al. 2003), NaOCl (Kleber et al.  
50 2005), and K<sub>2</sub>Cr<sub>2</sub>O<sub>7</sub> (Skjemstad and Taylor 1999); and 2) analytical techniques  
51 including (i) spectroscopic techniques, such as infrared (IR) spectroscopy and solid-  
52 state <sup>13</sup>C nuclear magnetic resonance (<sup>13</sup>C NMR) (Wilson et al. 1981; Fründ et al. 1993)  
53 and (ii) pyrolysis-GC/MS (Py-GC/MS) (Sáiz-Jiménez and the Leeuw 1994a). Whereas  
54 IR and <sup>13</sup>C NMR provide information on the environment of carbon atoms (functional  
55 groups) (Baldock and Smernik 2002), more detail on the molecular chemistry is  
56 obtained by pyrolysis-GC/MS.

57 Py-GC/MS is based on thermal degradation in an inert atmosphere (pyrolysis)  
58 and subsequent separation (GC) and identification (MS) of the pyrolyzate, from which  
59 information on macromolecular structures can be extracted (Moldoveanu 1998). For  
60 example, it allows the identification of different sources and degradation states of plant

61 detritus, secondary/microbial material, Black C (BC), and the estimation of their  
62 relative proportions (Nierop et al. 2005; Buurman et al. 2007). Nonetheless, Py-GC/MS  
63 is a semi-quantitative method because of differences in the pyrolyzability of different  
64 organic matter components, not all pyrolysis products are amenable and detectable by  
65 GC, and differences in relative response factors by MS (Sáiz-Jiménez 1994a)

66           The Walkley-Black dichromate oxidation, as modified by Heanes (1984), is a  
67 relatively simple and rapid procedure with minimal equipment needs (Nelson and  
68 Sommers 1996) that has long been used to estimate the organic C (OC) content of soils.  
69 Its major disadvantage is that it incompletely oxidizes soil OC (Gillman et al. 1986),  
70 and has different oxidation efficiencies for different soils (Tabatabai 1996), which  
71 produce considerable and unpredictable deviations from 'true' soil OC content.  
72 Probable causes of deviations in recoveries are (i) spatial inaccessibility of organic  
73 substrates to the oxidation agent (Skjemstad et al. 1996; Six et al. 2002), (ii) binding  
74 with inorganic phases (Eusterhues et al. 2005) and (iii) the presence of chemically  
75 recalcitrant SOM fractions such as BC (Six et al. 2002). In fact, the difference between  
76 total OC and dichromate-oxidizable OC has been used to estimate the BC content of  
77 soils. BC is defined as the product resulting from incomplete thermal combustion of  
78 vegetation and/or fossil fuels, and is relatively resistant to decomposition (Schmidt et al.  
79 2001). However, an unknown portion of BC is actually oxidized while some non-  
80 pyrogenic SOM may survive the oxidation, e.g. non-hydrolyzable aliphatic compounds  
81 that resist aqueous dichromate oxidation (Knicker et al. 2007). This implies that the  
82 properties of the oxidation-resistant residue must be assessed in order to obtain  
83 meaningful estimations of BC content using dichromate oxidation (Knicker et al. 2007).  
84 The stability of BC towards these reagents is of significant interest considering its  
85 upcoming use as a soil amendment (as biochar).

86           The permanganate-oxidizable fraction has been used as a proxy for the labile  
87 fraction of SOM (Loginow et al. 1987, Lefroy et al. 1993), based on the assumption that  
88 the oxidative capacity of  $\text{KMnO}_4$  on SOM is comparable to that of soil microbial  
89 enzymes (Conteh et al. 1997). However, some studies indicated that  $\text{KMnO}_4$ - oxidizable  
90 C may not be a reliable measure of the proportion of labile C because, even though it  
91 efficiently degrades lignin (van Soest and Wine 1986), it has little effect on several  
92 SOM components that are widely recognised as easily degraded by soil  
93 microorganisms, e.g. structural carbohydrates, sugars and amino acids (e.g. Tyrol-Padre  
94 and Ladha 2004). Furthermore, its ability to react with charcoal was also stated  
95 (Skjemstad et al. 2006).

96           SOM is thermodynamically unstable in well-aerated soils (Macías and Camps-  
97 Arbestain 2010). However, SOM stabilized by specific mechanisms can remain as  
98 meta-stable forms in soils for hundreds to thousands of years (Six et al. 2002). These  
99 non-ideal conditions for SOM decay are associated to physical and chemical protection  
100 mechanisms offered by the soil matrix that either impede the access of enzymes to SOM  
101 (e.g., within microaggregates or by creating hydrophobicity) or increase the energy  
102 needed to degrade SOM through interactions with minerals (Eusterhues et al. 2003). In  
103 addition, enhanced SOM preservation may occur when environmental conditions are  
104 not adequate for microbial growth, e.g. the presence of free Al and Fe, low soil pH  
105 and/or due to low nutrient availability (Buurman and Roscoe 2011). Furthermore,  
106 intrinsic recalcitrance of specific SOM components may increase its longevity in soil.  
107 This is most likely the main process responsible for the long turn-over time of highly  
108 condensed aromatic compounds present in BC (Harvey et al. 2012).

109           The nature of permanganate- and dichromate-oxidation resistant SOM is still  
110 poorly understood. Here we study the molecular properties of  $\text{KMnO}_4$ - and  $\text{K}_2\text{Cr}_2\text{O}_7$ -

111 oxidation-resistant SOM from a BC-rich colluvial soil, by Py-GC/MS and solid-state  
112  $^{13}\text{C}$  NMR, which may add to our knowledge on the stability of specific organic  
113 compounds against soil microbial oxidative enzymes.

114

## 115 **2. Material and methods**

### 116 *2.1. Study site and sample descriptions*

117 Soil PRD-4 is a 2.4 m thick Haplic Umbrisol (humic/alumic) according to the  
118 IUSS Working Group (2006) and Humic Pachic Dystrudept according to the SSS  
119 (1998). This soil type is traditionally referred to as Atlantic Ranker (Carballas et al.  
120 1967). Radiocarbon dating showed that the soil gradually accumulated through  
121 colluviation during the last ca. 13,000 yr (Kaal et al. 2011). Soil PRD-4 has a deep  
122 black color owing to a combination of high SOM content and abundance of BC (Table  
123 1). For the present study, three samples were selected from this soil, corresponding to  
124 three periods with radically different ecosystems and hypothetically also different SOM  
125 compositions: S1 (5-10 cm depth) corresponds to recent material (<150 y BP),  
126 evidenced by  $^{14}\text{C}$  dating and the presence of pollen of exotic species *Eucalyptus* sp.  
127 (López-Merino et al. 2012) (Table 1). This soil layer contains considerable amounts of  
128 “fresh” (non- or slightly degraded) SOM and root fragments, as described by Kaal and  
129 van Mourik (2008). The vegetation corresponding to **this sample** is a mosaic of  
130 shrubland (dominated by Ericaceae), pasture and exotic tree species. Sample S2 (95-  
131 100 cm depth) contains large amounts of charcoal from palaeofires (Kaal et al. 2011)  
132 that occurred ca. 5,000 yr ago. Anthracological analysis showed that most charcoal  
133 originates from deciduous *Quercus* sp. This sample is thought to correspond to an oak-  
134 dominated woodland under substantial fire and grazing pressure. Sample S3 (190- 195  
135 cm depth) corresponds to an Early- Holocene phase (ca. 9,700 yr BP) that preceded the

136 colonization of the area by deciduous forest, with “steppe-like” vegetation dominated  
137 by *Betula* sp. (birch), shrubs of the Fabaceae family and herbaceous species.

138

### 139 2.2. Determination of organic C fractions

140 Potassium dichromate-oxidizable organic C ( $OC_{\text{dichro}}$ ) was determined following  
141 the Walkley-Black oxidation method as modified by Wolbach and Anders (1989) and  
142 Knicker et al. (2007). 0.5 g of dry soil (< 2 mm) were oxidized in triplicate with 20 ml  
143 of 0.2 M  $K_2Cr_2O_7$  and 20 ml of concentrated  $H_2SO_4$  at 60 °C in a water bath for 6 h.  
144 **Control samples without soil were also analyzed.** After the reaction, excess dichromate  
145 was determined by titration against 0.033 M  $FeSO_4$ . The amount of dichromate  
146 consumed by the soil was used to calculate the amount of dichromate-oxidizable  
147 organic C ( $OC_{\text{dichro}}$ ) assuming that (i) the oxidation state of soil OC is zero ( $C^0$ ) and (ii)  
148 complete oxidation to  $C^{+4}$  occurs.

149 Potassium permanganate-oxidizable organic C ( $OC_{\text{per}}$ ) was determined, in  
150 triplicate, using 25 ml of 33 mM  $KMnO_4$  solution added in 50 ml centrifuge tubes  
151 containing an amount of dry soil (<2 mm) equivalent to 15 mg organic C (Tirol-Padre  
152 and Ladha 2004). After 24 h shaking, the tubes were centrifuged for 5 min at 2600 g  
153 and the supernatant diluted in distilled water (1:25 v:v). Absorbance was read on a split  
154 beam spectrophotometer at 565 nm. Blanks and a standard soil were analyzed before  
155 each run. For calculation purposes, it was assumed that three moles of C (e.g.  
156 carbohydrates) are oxidized for every four moles of  $Mn^{+7}$  reduced (Tirol-Padre and  
157 Ladha 2004).

158

### 159 2.3. Isolation of SOM fractions

160 The non-oxidized (NO) SOM extraction was considered the control treatment,  
161 consisting of SOM extraction by NaOH as described by Buurman et al. (2007). Briefly,  
162 5 g soil (air- dried fine earth <2 mm) was extracted with 50 ml of 1 M NaOH and  
163 shaken for 24 h under N<sub>2</sub> to prevent oxidation/saponification. The suspension was  
164 centrifuged at 2600 g for 1 h and the extract decanted, after which the extraction was  
165 repeated. The two extracts were combined and the residues discarded. The extracts were  
166 then acidified to pH 1 with concentrated HCl to protonate SOM. One ml of concentrated  
167 HF was added to dissolve silicates and increase C content of the extracted fraction. The  
168 acid mixture was shaken for 48 h, after which it was dialyzed to neutral pH against  
169 distilled water to remove excess salt. Finally, the suspension was freeze- dried.

170 Dichromate oxidation-resistant SOM (CR): 12.5 g soil was oxidized in 500 ml  
171 of 0.2 M K<sub>2</sub>Cr<sub>2</sub>O<sub>7</sub> and 100 ml of concentrated H<sub>2</sub>SO<sub>4</sub> for 6 h at 60 °C using a water  
172 bath. Once cooled, the suspension was centrifuged at 2600 g for 1 h and the supernatant  
173 decanted, after which the sediment was washed with distilled water until the solution  
174 was colorless. The suspension was discarded and the dichromate oxidation-resistant  
175 SOM extracted by 200 ml of 1 M NaOH for 24 h under N<sub>2</sub>. The resultant suspension  
176 was centrifuged at 2600 g for 1 h and the supernatant decanted. This extraction was  
177 repeated twice. Thereafter, the three extracts were combined and acidified to pH 1 with  
178 concentrated HCl. One ml of concentrated HF was added to dissolve silicates and  
179 increase the content of organic C of the extracted fraction. This acid mixture was shaken  
180 for 48 h, dialyzed against H<sub>2</sub>O to neutral pH and finally freeze- dried.

181 Permanganate oxidation-resistant SOM (MN): 1000 ml of 33 mM KMnO<sub>4</sub> were  
182 added to 4.8, 8.2 and 7.0 g dry soil (<2 mm) for samples S1, S2 and S3, respectively,  
183 aiming to add 25 ml of KMnO<sub>4</sub> per 15 mg of organic C (calculated from OC<sub>dichro</sub>  
184 values). After 24 h shaking, the suspension was centrifuged at 2600 g for 1.5 h and the



185 extract decanted, after which the sediment was washed with distilled water until the  
186 supernatant was colorless. The MN in the residue was isolated and purified using 200  
187 ml of 1 M NaOH for 24 h under N<sub>2</sub> atmosphere, analogous to the extraction of CR  
188 described above.

189

#### 190 2.4. Py-GC/MS

191 Platinum filament Py-GC-MS was performed with a Pyroprobe 5000 (CDS  
192 Analytical Inc., Oxford, USA) coupled to a 6890 GC and 5975 MS (Agilent  
193 Technologies, Palo Alto, USA). The non-oxidized (NO) and oxidation-resistant SOM  
194 fractions (MN and CR) were pyrolyzed at 750 °C for 10 s (heating rate 10 °C/ms).  
195 Analyses of sample S2 of the CR series was repeated, first at 400 °C and then at 750 °C,  
196 to distinguish between evaporation and pyrolysis products from volatile and  
197 macromolecular components, respectively. The pyrolysis interface was set at 300 °C  
198 and the GC inlet at 325 °C. The oven of the GC was heated from 50 to 325 °C at 10  
199 °C/min and held isothermal for 5 min. The GC/MS transfer line was held at 325 °C, the  
200 ion source (in electron impact mode, 70 eV) at 230 °C and the quadrupole detector at  
201 150 °C, measuring fragments in the *m/z* 50-500 range. The GC was equipped with a  
202 (non-polar) HP-1 100% dimethylpolysiloxane column. Helium was used as the carrier  
203 gas (constant gas flow, 1 ml/min). The major peaks in the total ion current of all  
204 samples were listed and, if possible, identified using the NIST '05 library and Py-  
205 GC/MS literature (Appendix A). Quantification of these pyrolysis products, 172 in total,  
206 was obtained by using the peak area of the major fragment ions (*m/z*). The sum of these  
207 peaks, i.e. total quantified peak area (TQPA) was set as 100% and the relative  
208 proportions of the pyrolysis products were calculated as the % of TQPA. This is a semi-  
209 quantitative estimate that allows for better comparison among samples than visual

210 inspection of pyrolysis chromatograms (pyrograms) alone, and produces a dataset that  
211 can be treated statistically.

212

### 213 2.5. $^{13}\text{C}$ NMR spectroscopy

214 Solid-state NMR spectroscopy experiments were performed with cross-polarization  
215 magic angle spinning (CPMAS) at 298 K in a 17.6 T Varian Inova-750 spectrometer  
216 (operating at 750 MHz proton frequency) equipped with a T3 Varian solid probe  
217 [Varian, Inc, USA]. Solid NMR samples were prepared in 3.2 mm rotors with an  
218 effective sample capacity of 22  $\mu\text{L}$  which corresponds to approximately 30 mg of the  
219 powdered sample. Spectra were processed and analyzed with MestreC software  
220 (Mestrelab Research Inc.). Carbon chemical shifts were referred to the carbon  
221 methylene signal of solid adamantane at 28.92 ppm. This sample was also used for the  
222 calibration of the 1D CPMAS experiments. 1D CPMAS spectra were acquired for the  
223 samples with the following conditions: the inter-scan delay was set to 0.5 s, the number  
224 of scans was 24000 and the MAS rate was 20 kHz. Heteronuclear decoupling during  
225 acquisition of the FID was performed with Spinal-64 with a proton field strength of 70  
226 kHz. The cross polarization time was set to 1 ms. During cross polarization, the field  
227 strength of the proton pulse was set constant to 75 kHz and that of the  $^{13}\text{C}$  pulse was  
228 linearly ramped during a 10% of the duration with a 20 kHz ramp near the matching  
229 sideband. Spectra were divided into different regions of chemical shift following  
230 Knicker et al (2005). Relative abundances of the various C groups were determined by  
231 integration of the signal intensity in their respective chemical shift regions. The region  
232 between 0– 45 ppm is assigned to alkyl C corresponding to terminal methyl groups and  
233 methylene groups of aliphatic moieties. The *O*-alkyl C region, typically assigned to  
234 carbohydrate-derived structures, between 45– 95 ppm. Here, between 45– 60 ppm, *N*-

235 alkyl C (i.e. in amino sugars and peptide structures) can contribute to the signal.  
236 Between 90 and 160 ppm resonance lines of olefins and aromatic C are detected. The  
237 regions from 160 to 220 ppm and from 220 to 245 are assigned to carbonyl C separated  
238 into carboxyl/amide and aldehyde/ketone groups, respectively. Although often assumed  
239 that solid- state  $^{13}\text{C}$  NMR underestimate BC, recent studies demonstrated that most  
240 charcoals have an atomic H/C ratio  $> 0.5$  and thus provide sufficient protonation for  
241 efficient cross polarization and reliable NMR spectra (Knicker et al. 2005). Because of  
242 the limited sample availability some samples required Al-oxide to fill the rotor, causing  
243 some signal quality deterioration. Samples MN-2, MN-3 and CR-3 were not analyzed  
244 for that reason.

245

#### 246 *2.6. Factor analysis*

247 The relative proportions of pyrolysis products were subjected to factor analysis  
248 using Statistica Version 8 (Statsoft, Tulsa, USA). Factor analysis proved useful in the  
249 interpretation of Py-GC/MS datasets, especially with respect to the sources and  
250 degradation states to which the pyrolysis products correspond.

251

### 252 **3. Results and discussion**

#### 253 *3.1. Py-GC/MS: source allocation*

254 The pyrolysis products were grouped according to their chemical structure into  
255 the following classes: (i) aliphatic compounds (homologous series of *n*-alkanes and *n*-  
256 alkenes, and branched alkenes), (ii) lignin-derived methoxyphenols, (iii) phenols, (iv)  
257 monocyclic aromatic compounds (MAHs), (v) polycyclic aromatic hydrocarbons  
258 (PAHs), (vi) N-containing compounds, (vii) carbohydrate-derived pyrolysis products

259 and (viii) unidentified compounds. Appendix A is a list of the pyrolysis products  
260 identified.

261 *Aliphatic compounds.* *n*-alkane/*n*-alkene pairs, ranging from C<sub>10</sub> to C<sub>28</sub>, originate  
262 largely from aliphatic biopolymers (Eglinton and Hamilton 1967). The other short- and  
263 mid-chain (ca. C<sub>10</sub>-C<sub>20</sub>) alkenes, not from the homologous series and most of which are  
264 probably branched, are considered significant products of charred aliphatic matter  
265 according to recent studies, even though they are not produced exclusively from  
266 pyrolysis of BC (Eckmeier and Wiesenberg 2009; Kaal et al. 2012a). Several *n*-fatty  
267 acids (mainly C<sub>16</sub> and C<sub>18</sub>) seemed to increase disproportionately upon chemical  
268 oxidation, especially in CR. This might be explained by the enrichment of aliphatic  
269 structures upon dichromate oxidation due to their hydrophobicity (Knicker et al. 2007).  
270 However, the use of a HP-1 column, which has a larger internal diameter than  
271 frequently used non-polar columns for Py-GC/MS, may have affected the  
272 'chromatographic mobility' and the relative proportions of these compounds, thus  
273 making its interpretation difficult. Therefore, these compounds were not included in the  
274 statistical analyses neither.

275 *Lignin-derived methoxyphenols.* Methoxyphenols (guaiacyl- and syringyl-based)  
276 are typical products of coniferyl and sinapyl lignin, respectively (Boerjan et al. 2003).  
277 4-vinylphenol was also added to this group because it has frequently been shown to be  
278 marker of coumaryl lignin and the non-lignin coumaric acid in grasses (Sáiz-Jiménez  
279 and de Leeuw 1986). An unknown proportion of 4-vinylphenol and 4-vinylguaiacol  
280 may originate from non-lignin phenolic acids as well (Schellekens et al. 2012), but that  
281 does not influence the interpretation of results here as statistically they behave as the  
282 lignin markers (see below).

283           *Phenols.* The other phenols have multiple origins: phenol and C<sub>1</sub>-C<sub>2</sub>-  
284 alkylphenols may originate from any phenolic precursor including lignin, tannin,  
285 proteinaceous biomass, weakly-charred BC and carbohydrates (Tegelaar et al. 1995;  
286 Stuczynski et al. 1997), while lignin, tannin or thermally demethylated lignin (Kaal et  
287 al. 2012b) are the most likely precursors of 1,2-benzenediol (catechol).

288           *Monocyclic aromatic hydrocarbons.* MAHs include benzene, toluene, styrene,  
289 dimethylbenzenes, linear C<sub>2</sub>-C<sub>4</sub>-alkylbenzenes, a dimethylstyrene and a dimethyl-  
290 methylethylbenzene compound. MAHs are formed from many aromatic and some non-  
291 aromatic precursors (Schulten et al. 1991) but BC is known to produce an exceptionally  
292 high proportion of especially benzene (Kaal et al. 2012a). Indeed, the analysis of the  
293 products of incomplete combustion by Py-GC/MS showed that MAHs and also PAHs  
294 are major pyrolysis products of Black C (Pastorova et al. 1994, Almendros et al. 2003).  
295 Besides a renewed interest in the detection of burning residues in SOM established that  
296 BC is a major source of SOM and that its oxidation products could be a potential source  
297 of highly aromatic humic acids (Hatcher et al. 1989; Skjemstad et al. 1996; Shindo et  
298 al., 2004). On the other hand, the alkylstyrenes most likely originate from the  
299 monoterpenes present in *Eucalyptus globulus* litter (see below).

300           *Polycyclic aromatic hydrocarbon compounds.* The origin of PAHs in SOM  
301 pyrolyzates have been the subject of considerable debate. They were sometimes  
302 considered actual building blocks of humic substances formed upon condensation  
303 reactions during humification (Schulten et al. 1991), or more frequently interpreted as  
304 analytical artefacts because of evidence of their formation during pyrolysis of aliphatic  
305 compounds through cyclization and aromatization (Sáiz-Jiménez 1994b). More  
306 recently, PAHs and particularly the non-alkyl-substituted and >2 ring PAHs (Rumpel et  
307 al. 2007) are considered indicative of (but not markers of) BC in SOM (Kaal and

308 Rumpel 2009; Song and Peng 2010). In the present study, unsubstituted PAHs (indene,  
309 naphthalene, fluorene, biphenyl, phenanthrene and anthracene) and C<sub>1</sub>-C<sub>2</sub> alkyl  
310 analogues of these PAHs were abundant. In addition, a series of C<sub>3</sub>-C<sub>4</sub>  
311 alkylnaphthalenes and C<sub>5:0</sub> and C<sub>5:1</sub> alkylnaphthalenes probably originate from  
312 evaporation and pyrolysis of monoterpenes (e.g. pinene, phellandrene, eucalyptol) and  
313 sesquiterpenoids (aromadendrene, globulol), respectively, present in *Eucalyptus* sp. oil.  
314 Indeed, these compounds were identified in *Eucalyptus globulus* litter (a mixture of  
315 leaves, cortex and branches) pyrolyzates (data not shown). Some of the polysubstituted  
316 PAHs (C<sub>3</sub>-indene, C<sub>2</sub>-C<sub>5:0</sub>- and C<sub>5:1</sub>-naphthalene) provided the largest contributions to  
317 the pyrograms of sample S1 (contrary to many unsubstituted or C<sub>1</sub>- alkylsubstituted  
318 PAHs, largely from BC); this suggests that a significant portion of these compounds  
319 originate from fresh *Eucalyptus* sp. litter. Overall, it should be noted that, in this study,  
320 where distinguishing between BC and non-pyrogenic SOM components will appear to  
321 be important, the interpretation of MAHs and PAHs relies more strongly on the growing  
322 body of knowledge on BC's pyrolysis fingerprints (Kaal et al. 2012a; Fabbri et al. 2012;  
323 Song and Peng 2010) than on previous studies on the structural characteristics of humic  
324 acids (Schulten et al. 1991; Sáiz-Jiménez 1994a).

325 *N*-containing products. Of the 24 N-containing pyrolysis products identified,  
326 benzonitrile and C<sub>1</sub>-benzonitriles were recently proposed as the main products of N-  
327 containing groups in BC (Schnitzer et al. 2007; Song and Peng 2010). In addition,  
328 isoquinoline, phenylpyridine, benzenedicarbonitriles and pyridinecarbonitriles can be  
329 considered as markers of 'Black N' (BN) (Knicker 2007; Kaal et al. 2009). Note that  
330 absence of these products does not imply absence of BN: these compounds can  
331 probably only be detected in high-quality pyrograms of exceptionally BN-rich samples.  
332 Pyrroles, pyridines and indoles are potential products of BN as well but these

333 compounds are common in the pyrolyzates of non-pyrogenic N-moieties. Several  
334 markers of chitin (acetamide and a compound tentatively identified as trianhydro-2-  
335 acetamido-2-deoxyglucose; van der Kaaden et al. 1984; Stankiewicz et al. 1996) and  
336 chitin-entangled protein (diketopiperazine) probably originate from fungal cell walls  
337 and/or arthropod exoskeleta, either way serving as an indication of biologically re-  
338 assimilated (“secondary”) remains in SOM (Gutierrez et al. 1995). Finally, for  
339 picolinamide, cyanobenzoic acid and phthalimide-based compounds no specific origin  
340 has been identified yet.

341 *Carbohydrate compounds.* Of the carbohydrate products identified, levoglucosan,  
342 dianhydro- $\alpha$ -glucopyranose, pyranones and dianhydrorhamnose largely originate from  
343 “fresh” or well-preserved polysaccharides (Stuczynski et al. 1997; Poirier et al. 2005;  
344 Nierop et al. 2005). On the other hand, cyclopentenediones, furans, furfurals,  
345 levoglucosenone and dibenzofuran originate from fresh and/or degraded carbohydrates  
346 (Buurman and Roscoe 2011). This degradation may be either biological or thermal in  
347 nature, the latter especially for the furans, furaldehydes and dibenzofuran (Pastorova et  
348 al. 1994; Boon et al. 1994).

349 *Unidentified compounds.* An unsaturated non-aromatic cyclic compound (U1) was  
350 identified only in the pyrolyzates from S1 (NO-1 and MN-1), which also contained the  
351 pollen of *Eucalyptus* sp. It probably corresponds to  $\alpha$ -phellandrene, which is abundant  
352 in eucalyptus oils (Samat  et al. 1998). Furthermore, several polymethyl-substituted  
353 polycyclic compounds (U3-U6), also detected in the aforementioned fresh eucalyptus-  
354 litter pyrolyzate, probably derived from *Eucalyptus* sp. Finally, a methylated  
355 cyclohexane (U2) of unknown origin was tentatively identified.

356

357 3.2. Py-GC/MS: quantification and interpretation

358 Pyrograms of chemical oxidation resistant SOM fractions and non-oxidized  
359 samples are represented in Fig. 1. The relative contributions to TQPA for identified  
360 groups in the different soil horizons studied are presented in Table 2. In NO-1,  
361 carbohydrate-derived pyrolysis products accounted for 30% of TQPA, with  
362 levoglucosan (Ps13) from intact polysaccharide (Stuczynski et al. 1997; Poirier et al.  
363 2005) being dominant. The presence of 4-hydroxy-5,6-dihydro-(2H)-pyranone (Ps6)  
364 and dianhydrorhamnose (Ps7) confirms the existence of fresh (or well-preserved)  
365 polysaccharides in NO-1 (Nierop et al. 2005). Of the samples studied, these compounds  
366 showed the largest contribution to sample NO-1. The same pattern was observed for  
367 many other indicators of fresh plant material, including the aliphatic compound  
368 producing  $m/z$  83+280, diketodipyrrole, and the lignin-derived products (Suárez-  
369 Abelenda et al. 2011). The large proportion of phenols in the pyrolyzate of this sample  
370 may be explained by the abundance of lignin. Samples NO-1 and MN-1 had the highest  
371 contributions of probably eucalyptus-derived moieties ( $C_{5:0}$ -,  $C_{5:1}$ -alkylnaphthalenes and  
372  $\alpha$ -phellandrene and  $C_3$ - naphthalenes). It is concluded that the SOM of NO-1 is  
373 characterized by a large fraction of well-preserved polysaccharides and lignin, with an  
374 additional contribution of specific eucalyptus-derived substances, and relatively small  
375 proportions of microbial and pyrogenic SOM. The latter is supported by the low  
376 benzene/alkyl-benzenes and PAH/alkyl-PAHs ratios (Table 2), which are indicative of a  
377 low contribution of strongly charred BC to the MAHs and PAHs of these samples (Kaal  
378 and Rumpel 2009; Kaal et al. 2012a).

379 Unsurprisingly, sample NO-2 (ca. 5,000 yr old), produced fewer pyrolysis  
380 products from fresh SOM than NO-1. More specifically, in comparison with NO-1,  
381 among the carbohydrate markers there was a strong increase of furans, furaldehydes,  
382 levoglucosenone and acetic acid, while levoglucosan, pyranones and



383 dianhydrorhamnose diminished, which is a clear indication of a shift of fresh  
384 polysaccharide to degraded/microbial carbohydrates (Sáiz-Jiménez and de Leeuw 1986;  
385 Buurman and Roscoe 2011). Lignin markers were virtually absent. NO-2 sample  
386 produced many N-compounds, including those from chitin (N3 and N22), pyridine (N1,  
387 often associated with microbial SOM; Buurman et al. 2007) and BN (e.g. aromatic  
388 carbonitriles and phenylpyridine). It also gave higher proportions of MAHs and BC-  
389 derived PAHs than the NO-1 sample. It is concluded that the SOM of sample NO-2 was  
390 predominantly composed of degraded/microbial and pyrogenic material.

391 The pyrolyzate of sample NO-3 was dominated by carbohydrate markers, with  
392 acetic acid, 3/2-furaldehyde, 5-methyl-2-furaldehyde, dianhydro- $\alpha$ -glucopyranose, a  
393 furanone and 4-acetylfuran accounting for 62% of TQPA (Table 2). These pyrolysis  
394 products are frequently ascribed to SOM with large proportions of microbial biomass  
395 (Sáiz-Jiménez and de Leeuw 1986; Buurman and Roscoe 2011). The small relative  
396 proportions of MAHs and PAHs suggest that BC accounts for only a minor portion of  
397 the SOM in NO-3. This is supported by the low ratios of benzene/alkyl-benzenes and  
398 PAH/alkyl-PAH (Table 2).

399 In general, the differences in pyrolyzate compositions between NO-samples and  
400 MN-samples were small, yet some are worth mentioning. For sample S1, oxidation with  
401  $\text{KMnO}_4$  (which promoted a decreased of  $2.3 \text{ mg g}^{-1}$  of OC; Table1) caused an increase  
402 in MAHs (from 16.4% in NO-1 to 30.0% in MN-1) and decrease in carbohydrates (from  
403 30.4% to 18.4%) and lignin (from 7.4% to 3.0 %) (Table 2). These results can be  
404 explained by the partial oxidation of fresh SOM (van Soest and Wine 1986; Tirol-Padre  
405 and Ladha 2004) and the relative enrichment of pyrogenic (Almendros et al. 1990) and  
406 aliphatic SOM (González-Vila and Martín 1985; Almendros et al. 1989). Besides,  
407 permanganate oxidation concentrates pyrolysis products from aromatic structures

408 present in humic acids in general (Polvillo et al. 2009), even though cyclization of  
409 aliphatic precursors may play a role as well (González-Vila and Martín 1985). In  
410 sample S2,  $\text{KMnO}_4$  oxidation (MN-2) (with a smaller decrease of OC;  $1.3 \text{ mg g}^{-1}$ )  
411 caused a strong decline in carbohydrate products (from 31.5% to 17.4% of TQPA in  
412 NO-2 and MN-2, respectively) and an increase in aliphatic pyrolysis products (sum of  
413 *n*-alkanes, *n*-alkenes and other aliphatic compounds from 5.3% in NO-2 to 23.1% in  
414 MN-2). These changes are indicative of selective oxidation of (an unknown proportion  
415 of) the degraded/microbial SOM and the relative enrichment of aliphatic precursors  
416 probably from degraded root components (Kaal and van Mourik 2008). In addition,  
417 MN-2 produced higher amounts of N-containing BC markers, as BC and BN are  
418 relatively resistant against oxidation with this reagent. Skjemstad et al. (2006) found  
419 that  $\text{KMnO}_4$  may react significantly with BC, but no evidence of this was found here.  
420 Finally, sample MN-3 (with a decrease of  $1.7 \text{ mg g}^{-1}$  of its OC content) contains a  
421 smaller proportion of carbohydrates than NO-3 (17.7% vs. 62.2% of TQPA),  
422 confirming that the degraded/microbial carbohydrate fraction is relatively susceptible to  
423 this oxidation agent.

424 The pyrolyzates obtained from the residues after dichromate oxidation were very  
425 different from those of NO- and MN-samples. The CR-1 sample ( $27.5 \text{ mg g}^{-1}$  of its OC  
426 was oxidized by dichromate; Table 1) was strongly enriched in aliphatic pyrolysis  
427 products (34% of TQPA), particularly of short-chain ( $<C_{18}$ ) *n*-alkanes/*n*-alkenes  
428 (located on the right side of the broken line in the aliphatic cluster; SE quadrant, Fig. 2)  
429 and branched alkenes, and depleted in lignin-, carbohydrate- and eucalyptus-derived  
430 pyrolysis products in comparison with NO-1 and MN-1. CR-1 also produced the highest  
431 proportions of 3-ring PAHs and higher ratios of benzene/alkyl-benzenes and  
432 PAH/alkyl-PAHs ratios (Table 2), suggesting that a large proportion of the MAHs and

433 PAHs from the CR-1 is pyrogenic. These results are indicative of the enrichment of  
434 pyrogenic SOM in the fraction resistant to  $K_2Cr_2O_7$ - oxidation. Indeed, the partial  
435 resistance of BC to  $K_2Cr_2O_7$  oxidation is well-documented (Knicker et al. 2007, 2008).  
436 The same studies showed the existence of a  $K_2Cr_2O_7$ -resistant alkyl fraction (Knicker et  
437 al. 2007, 2008), which was also supported by the pyrolyzate composition of CR-1. The  
438 increase was also observed for the *n*-fatty acids (data not shown), which are not  
439 considered part of structural aliphatic plant material. These results support the  
440 hypothesis that the enrichment of aliphatic material in the residual fraction of the  
441  $K_2Cr_2O_7$  oxidation residues is produced by the hydrophobic nature of these constituents  
442 (Knicker et al. 2007), possibly in combination with chemical recalcitrance of C-C bonds  
443 in methylene chains. Finally, a decrease was observed for the intact terpene-like plant-  
444 derived PAHs, clearly showing different origin for the un- and methyl-substituted PAHs  
445 (mainly from BC) and the polyalkyl-substituted PAHs from eucalyptus litter. For  
446 sample CR-2, dichromate oxidation oxidized less OC ( $12.3 \text{ mg g}^{-1}$  for CR-2) than for  
447 CR-1. It produced a further decrease in the proportion of lignin markers in comparison  
448 with MN-2, and of microbial products such as acetamide and furans, while the BC and  
449 BN fingerprints were relatively intense (e.g. benzene, unsubstituted PAHs, benzene  
450 carbonitriles, isoquinoline and dibenzofuran). The largest proportion of these BC-  
451 derived pyrolysis products coincides with the highest macroscopic charcoal content of  
452 sample S2 (Table 1). These results confirm the accumulation of BC and BN in the  
453 residues after  $K_2Cr_2O_7$  oxidation. Similar to CR-1, sample CR-2 was enriched in  
454 aliphatic pyrolysis products. Unexpectedly, significant amounts of the markers of well-  
455 preserved polysaccharides such as levoglucosan, were detected in the pyrolyzates of  
456 CR-2 and CR-3. With the information available at this moment, our best explanation to  
457 this observation is the presence of an uncharred cellulose-containing core (Knicker et al.

458 2005) in incompletely charred particles and thereby protected against  $K_2Cr_2O_7$   
459 oxidation. Sample CR-3 (in which dichromate caused a loss of  $14.3 \text{ mg g}^{-1}$  of OC) was  
460 also enriched in pyrogenic SOM with a high contribution of BN markers, and in an  
461 aliphatic component with particularly high contributions of branched alkenes from an  
462 aliphatic SOM fraction, possibly in part pyrogenic.

463

### 464 3.3. Py-GC/MS: factor analysis

465 The first four factors (F1-F4) explained 81% of the variation in the Py-GC/MS  
466 dataset, with F1 and F2 combined accounting for 61%. The loadings of the pyrolysis  
467 products, and the scores of the samples analyzed, are shown in F1-F2 factor space (Fig.  
468 2).

469 *n*-Alkanes/ *n*-alkenes are predominantly represented in the SE quadrant with high  
470 positive loadings on F1. Lower loadings on F1 were observed for the *n*-alkenes  $>C_{20}$   
471 than for  $C_{10}$ - $C_{20}$  *n*-alkenes. Branched alkenes plot between their straight-chain  
472 analogues and the pyrogenic SOM markers, supporting the hypothesis that these  
473 branched alkenes are associated with charred aliphatic precursors (Eckmeier and  
474 Wiesenberg 2009; Kaal and Rumpel 2009).

475 The N-containing compounds are spread throughout the F1-F2 factor space,  
476 which is a result of the diverse origin of the members of this group. One cluster of N-  
477 containing compounds in the NE quadrant is composed of benzonitrile,  $C_1$ -  
478 benzonitriles, benzene dicarbonitriles, cyanobenzoic acid, pyridine,  $C_1$ -pyridine and  
479 pyridinecarbonitrile, clearly reflecting a pyrogenic origin. Indeed, many non-alkyl-  
480 substituted PAHs and dibenzofuran, also associated with BC (Pastorova et al. 1994),  
481 plot in the same region. Chitin-derived N compounds (chitin markers such as  
482 acetamide) and 4-acetylfuran are spread out in the NW quadrant together with microbial

483 polysaccharides whilst diketodipyrrole denotes the presence of fresh SOM in the SW  
484 quadrant (see below). Indoles are spread throughout the SW and SE quadrants, which  
485 may be indicative of a mixed origin.

486 The lignin-derived products (including free phenolic acids), i.e. 4-vinylphenol,  
487 guaiacols and syringols), and catechol occur in the SW quadrant together with an  
488 aliphatic marker of fresh plant material (A113). Most of the remaining pyrolysis  
489 products in this region are associated with fresh or well-preserved SOM components as  
490 well: phenol and alkylphenols (in this case from lignin), C<sub>5:0</sub> and C<sub>5:1</sub> alkyl-naphthalenes  
491 (from eucalyptus litter), dianhydrosucrose (from polysaccharides) and diketodipyrrole  
492 and indoles (typical N-containing pyrolysis products of fresh proteinaceous biomass;  
493 Buurman and Roscoe 2011). Levoglucosan plots between fresh OM and charred  
494 material; this may be attributed to the aforementioned protection of cellulose in interior  
495 parts of charcoal particles in dichromate oxidation residues or an unknown alternative  
496 levoglucosan source. The other carbohydrate products are spread along F2 because they  
497 have multiple sources, most of which corresponding to degraded/microbial SOM. The  
498 carbohydrate products of degraded/microbial SOM are probably those that plot in the  
499 NW quadrant: furans (4-acetylfuran, 3/2-furaldehyde and 5-methyl-2-furaldehyde),  
500 glucopyranose, which is microbial marker (Nierop et al. 2005), and acetic acid. This  
501 interpretation is consistent with the presence of the markers of chitin in this region of  
502 factor space.

503 In summary, F1-F2 separates the pyrolysis products according to their principal  
504 origin. Factor 1 separates the pyrogenic and aliphatic oxidation-resistant SOM fractions  
505 (chemically stable) from the fresh and degraded SOM fractions (chemically labile),  
506 while decomposed and pyrogenic SOM (strongly altered) are separated from fresh and

507 oxidation-resistant aliphatics (resembling plant material) according to their loadings on  
508 F2.

509 The factor scores of the samples can be used to identify the main differences  
510 between the samples analyzed (Fig. 2). As such, the samples with a large fraction of  
511 fresh biomass (NO-1 and MN-1) plot in the SW quadrant. Sample CR-1 plots in the SE  
512 quadrant, as dichromate oxidation eliminated most of the fresh SOM causing the  
513 relative accumulation of aliphatic SOM and weakly charred material. Sample S2 is a  
514 mixture of mainly degraded SOM and BC (with small contributions of fresh and  
515 aliphatic material), which is why NO-2 plots in the NW region, dominated by  
516 degraded/microbial markers, while MN-2 and CR-2 plot in the NE quadrant because of  
517 relative enrichment of BC after chemical treatment. Sample S3 was also rich in  
518 microbial SOM but has a lower content of chemically recalcitrant/hydrophobic aliphatic  
519 SOM and BC, and higher proportion of degraded/microbial SOM; this explains why  
520 NO-3 has a high F2 score while MN-3 and CR-3 plot in the NE quadrant reflecting BC  
521 enrichment after the selective depletion of degraded/microbial SOM. The short distance  
522 in factorial space between MN-3 and CR-3, and the large distance between these  
523 samples and NO-3, suggests a that the abundant degraded/microbial biomass  
524 (carbohydrates, chitin) in this sample is highly susceptible to permanganate and  
525 dichromate treatment.

526 From these results some inferences on the effects of the oxidation agents on SOM  
527 composition can be made. First, the minor differences between NO-1 and MN-1 can be  
528 explained by the relatively small microbial contribution to sample S1. In contrast,  
529  $K_2Cr_2O_7$  thoroughly modified the pyrolysis fingerprint obtained from the residues of S1  
530 by eliminating lignin, polysaccharides and terpenes, and relative enrichment of aliphatic  
531 and pyrogenic structures. In the older samples, where the aliphatic fraction is less

532 dominant while that of degraded/microbial and pyrogenic SOM prevail, chemically  
533 oxidized samples (MN-2, MN-3, CR-2 and CR-3) had positive scores on F1 and F2  
534 mainly because both oxidants concentrate BC, with  $K_2Cr_2O_7$  being the stronger oxidant.

535

### 536 3.4. Solid- state $^{13}C$ NMR spectroscopy: results and comparison with Py-GC/MS

537 Samples are compared by relative intensity of the chemical shift regions. The  
538 spectrum of NO-1 (Fig. 3) was characterized by a dominant signal at 21 ppm and a  
539 shoulder at 29 ppm (combined 29%, Table 3) from alkyl C, which can be ascribed to  
540 aliphatic structures in fatty acids, lipids, waxes, cutan, suberan, cutin and suberin  
541 (Tegelaar et al. 1989) but also peptide structures and short alkyl side-chains. In the O-  
542 alkyl C region (45–110 ppm) a broad peak at 75 ppm (with a contribution of 29%) was  
543 detected, which is generally attributed to cellulose, hemicelluloses and pectins (Gramble  
544 et al. 1994; Kögel-Knabner 1997). The peak at 55 ppm probably corresponds to  
545 methoxyl groups in lignin structures (Kögel-Knabner 1997) but can also have  
546 contributions of N-alkyl from amino sugars and peptides. The O-substituted aromatic C  
547 between 140- 160 ppm may derive from lignin and oxidized BC (Knicker et al. 2005).  
548 Resonance lines of aromatic C-H groups are detected in the chemical shift region  
549 between 110 ppm and 140 ppm (14%) (Knicker and Lüdemann 1995). The chemical  
550 shifts of carbon in carboxylic acids, esters and amides fall within the range between 160  
551 ppm and 220 ppm and represents 15% of the total  $^{13}C$  intensity. There are minor  
552 contributions of carbonyl or aldehydes, giving signals between 220- 245 ppm.

553 The NMR spectra of NO-2 and NO-3 differ considerably from that of NO-1. In  
554 NO-2, the O-alkyl C fraction has the highest values (26%), followed by alkyl C,  
555 carboxyl/amide C and aromatic C. Considering that there is no clear signal in the O-  
556 substituted C region (from 140– 160 ppm), this spectrum can be best explained with a

557 considerable contribution of oxidized charcoal. NO-3 presented the highest contribution  
558 of non-lignin aromatic fraction, probably BC-derived (31%) and a large signal of  
559 carboxyl/amide C (26%), which likely originates from microbial compounds.

560 The chemical oxidants had several effects on the NMR signal obtained from  
561 sample S1. After the  $\text{KMnO}_4$  treatment (MN-1), and comparable to results from Py-GC-  
562 MS, a spectrum similar to that of NO-1 but with slightly higher relative intensities in the  
563 alkyl C (34%) region was acquired, confirming other  $^{13}\text{C}$  NMR studies (Tirol-Padre and  
564 Ladha 2004) in that cellulose is largely resistant to permanganate oxidation. Dichromate  
565 oxidation of sample S1 caused an increase of the relative contribution of aromatic C  
566 (sum of aromatic C-H and aromatic C-O-R) (from 17% in NO-1 to 25% in CR-1) with a  
567 concomitant depletion of methoxyl C/N-alkyl C and O-alkyl C (see Table 3). Note that  
568 differences between NO-1 and CR-1 by NMR are smaller than observed by Py-GC/MS.  
569 This indicates that, as Py-GC/MS data seem to be best supported, NMR results must be  
570 taken carefully.

571 The spectrum of CR-2 showed the highest intensity in the chemical shift region of  
572 aromatic C (45%). Considering the absence of methoxyl C signal, the width of the  
573 signal band (from 90 to 140 ppm) and the composition of the pyrolysis fingerprint, this  
574 aromatic signal originates from BC (see also Skjemstad et al. 1996; Knicker et al.  
575 2005). Moreover, the contribution of alkyl C was strongly reduced upon  $\text{K}_2\text{Cr}_2\text{O}_7$   
576 oxidation (10% in CR-2).

577 In summary, in NO samples,  $^{13}\text{C}$  NMR spectroscopy shows a relative decrease  
578 with depth of aliphatic C, carbohydrates and lignin moieties and a relative increase with  
579 depth of a non-lignin aromatic fraction (probably BC) and carboxyl/amide C (possibly  
580 oxidized BC in combination with N-rich microbial SOM), which is in agreement with  
581 Py-GC/MS data. With regard to the effects of chemical oxidation, both Py-GC/MS and



582  $^{13}\text{C}$  NMR spectroscopy showed a decrease of easily degradable SOM (mainly composed  
583 of fresh lignin and polysaccharides) and an increase of the aromatic fraction. Besides,  
584 lignin was slightly oxidized by  $\text{KMnO}_4$  contrary to that observed by previous studies  
585 (van Soest and Wine 1986; Tyrol-Padre and Ladha 2004; Skjemstad et al. 2006) where  
586 lignin was strongly degraded. The increase of aliphatic moieties upon chemical  
587 oxidation, as suggested by Py-GC/MS, was also observed by NMR spectroscopy in the  
588 superficial sample S1 although it strongly decreased in S2. However here one has to  
589 bear in mind that the alkyl C region does not only contain intensity of lipids but have  
590 considerable contributions of peptide structures or short alkyl side-chains (such as  $\text{C}_3$ -  
591 side chains in lignin). Those moieties may be expected to be relatively susceptible to  
592 oxidation resulting in a relative depletion of the signal intensity in the alkyl C region  
593 even though longer chain aliphatic components may have experienced a relative  
594 enrichment.

595

#### 596 4. Conclusions

597 The molecular study of SOM fractions of three horizons of a colluvial soil  
598 representing ages of 100, 5,000 and 9,700 yr, before and after treatment with  $\text{KMnO}_4$   
599 and  $\text{K}_2\text{Cr}_2\text{O}_7$ , provided detailed information on SOM composition (with regard to  
600 source and degradation/preservation state) and the behaviour of different SOM fractions  
601 towards these oxidation agents. *Eucalyptus*-derived terpenes and sesquiterpenes were  
602 only present in the youngest sample and resisted  $\text{KMnO}_4$  but not  $\text{K}_2\text{Cr}_2\text{O}_7$  oxidation.  
603 Microbial/degraded SOM, mostly composed of carbohydrates and chitin, was especially  
604 abundant in the deeper layers of the soil and appeared highly susceptible to both  
605  $\text{KMnO}_4$  and  $\text{K}_2\text{Cr}_2\text{O}_7$  oxidation. As such,  $\text{KMnO}_4$  could be used as an indication of the  
606 abundance of microbial biomass. Both oxidants,  $\text{K}_2\text{Cr}_2\text{O}_7$  in particular, concentrated

607 two other SOM fractions abundant in this soil: aliphatic and pyrogenic material (BC),  
608 the latter having a significant amount of N-containing functional groups (BN). These  
609 fractions probably survived  $K_2Cr_2O_7$  oxidation because of the chemical stability of  
610 polyaromatic moieties (BC) and resistant C-C bonds in methylene chains and/or  
611 hydrophobicity of the aliphatic fraction (which is probably root-derived). It appeared  
612 that especially  $K_2Cr_2O_7$  oxidation efficiently concentrates BC and oxidation-resistant  
613 aliphatic structures from other SOM sources, and that in combination with Py-GC/MS it  
614 is possible to distinguish between these sources (yet not quantitatively) while  $^{13}C$  NMR  
615 may assist in obtaining estimations of their relative proportions. Finally, BC isolation by  
616 dichromate oxidation and posterior quantification through total digestion (Knicker et al.,  
617 2007) is discouraged as a significant aliphatic fraction resists dichromate producing an  
618 overestimation of its contents.

619

## 620 Acknowledgements

621 We thank Antonio Martínez-Cortizas for the discussion on the statistical analysis.  
622 The contribution of M. Camps-Arbestain to this research was funded by MAF and  
623 NZAGRC. The authors also thank the anonymous reviewers for their constructive  
624 comments on the manuscript.

625

## 626 References

627 Almendros G, González-Vila FJ, Martín F, (1989) Room temperature alkaline  
628 permanganate oxidation of representative humic acids. *Soil Biology and*  
629 *Biochemistry* **21**, 481–486.

- 630 Almendros G, González-Vila FJ, Martín F (1990) Fire-induced transformation of soil  
631 organic matter from an oak forest: an experimental approach to the effects of fire  
632 on humic substances. *Soil Science* **149**, 158–168.
- 633 Almendros G, Knicker H, González-Vila FJ (2003) Rearrangement of carbon and  
634 nitrogen forms in peat after progressive thermal oxidation as determined by solid-  
635 state  $^{13}\text{C}$ - and  $^{15}\text{N}$  NMR spectroscopy. *Organic Geochemistry* **34**, 1559–1568.
- 636 Baldock JA, Smernik RJ (2002) Chemical composition and bioavailability of thermally  
637 altered *Pinus resinosa* (Red pine) Wood. *Organic Geochemistry* **33**, 1093–1109.
- 638 Boerjan W, Ralph J, Baucher M (2003) Lignin biosynthesis. *Annual Review of Plant*  
639 *Physiology and Plant Molecular Biology* **54**, 519–546.
- 640 Boon JJ, Pastorova I, Botto RE, Arisz PW (1994) Structural studies on cellulose  
641 pyrolysis and cellulose chars by PYMS, PYGCMS, FTIR, NMR and by wet  
642 chemical techniques. *Biomass and Bioenergy* **7**, 25–32.
- 643 Buurman P, Petersen F, Almendros G (2007) Soil organic matter chemistry in  
644 allophanic soils: A pyrolysis-GC/MS study of a Costa Rican Andosol Catena.  
645 *European Journal of Soil Science* **58**, 1330–1347.
- 646 Buurman P, Roscoe R (2011) Different chemistry of free light and occluded light and  
647 extractable SOM fractions in soils of Cerrado, tilled and untilled fields, Minas  
648 Gerais, Brazil - a pyrolysis-GC/MS study. *European Journal of Soil Science* **62**,  
649 253–266.
- 650 Carballas T, Duchaufour P, Jacquin F (1967) Évolution de la matière organique des  
651 rankers. *Bulletin de l'École Nationale Supérieure Agronomique de Nancy* **9**, 20–  
652 28.

- 653 Christensen BT (1992) Physical fractionation of soil and organic matter in primary  
654 particle size and density separates. In 'Advances in Soil Science, Volume 20'. (Ed  
655 BA Stewart) pp. 1–90. (Springer: New York).
- 656 Conteh A, Lefroy RDB, Blair GJ (1997) Dynamics of organic matter in soil as  
657 determined by variations in  $^{13}\text{C}/^{12}\text{C}$  isotopic ratios and fractionation by ease of  
658 oxidation. *Australian Journal of Soil Resources* **35**, 881–890.
- 659 Eckmeier E, Wiesenberg GLB (2009) Short-chain *n*-alkanes ( $\text{C}_{16}$ – $\text{C}_{20}$ ) in ancient soil  
660 are useful molecular markers for prehistoric biomass burning. *Journal of*  
661 *Archaeological Science* **36**, 1590–1596.
- 662 Eglinton G, Hamilton RJ (1967) Leaf epicuticular waxes. *Science* **156**, 1322–1335.
- 663 Eusterhues K, Rumpel C, Kleber M, Kögel-Knabner I (2003) Stabilisation of soil  
664 organic matter by interactions with minerals as revealed by mineral dissolution  
665 and oxidative degradation. *Organic Geochemistry* **34**, 1591–1600.
- 666 Eusterhues K, Rumpel C, Kögel-Knabner I (2005) Stabilization of soil organic matter  
667 isolated via oxidative degradation. *Organic Geochemistry* **36**, 1567–1575.
- 668 Fabbri D, Torri C, Spokas KA (2012) Analytical pyrolysis of synthetic chars derived  
669 from biomass with potential agronomic application (biochar). Relationships with  
670 impacts on microbial carbon dioxide production. *Journal of Analytical and*  
671 *Applied Pyrolysis* **93**, 77–84.
- 672 Fründ R, Haider K, Lüdemann HD (1993) Impacts of soil management practices on the  
673 organic matter structure - investigations by CPMAS  $^{13}\text{C}$  NMR-spectroscopy.  
674 *Zeitschrift für Pflanzenernährung und Bodenkunde* **157**, 29–35.

- 675 Gillman GP, Sinclair DF, Beech TA (1986) Recovery of organic carbon by the Walkley  
676 and Black procedure in highly weathered soils. *Communications in Soil Science  
677 and Plant Analysis* **17**, 885–892.
- 678 González-Vila FJ, Martín F (1985) Chemical structural characteristics of humic acids  
679 extracted from composted municipal refuse. *Agriculture, ecosystems &  
680 environment* **14**, 267–278.
- 681 Gramble GR, Sethuraman A, Akin DE, Eriksson KE (1994) Biodegradation of  
682 lignocellulose in Bermuda grass by white rot fungi analysed by solid-state <sup>13</sup>C  
683 Nuclear Magnetic Resonance. *Applied Environmental Microbiology* **60**, 3138–  
684 3144.
- 685 Gutierrez A, Martínez MJ, Almendros G, González-Vila FJ, Martínez AT (1995)  
686 Hyphal-sheath polysaccharides in fungal deterioration. *The Science of the Total  
687 Environment* **167**, 315–328.
- 688 Harvey OR, Kuo LJ, Zimmerman AR, Louchouart P, Amonette JE, Herbert BE (2012)  
689 An Index-Based Approach to Assessing Recalcitrance and Soil Carbon  
690 Sequestration Potential of Engineered Black Carbons (Biochars). *Environmental  
691 Science and Technology* **46**, 1415–1421.
- 692 Hatcher PG, Schnitzer M, Vassallo AM, Wilson MA (1989) The chemical structure of  
693 highly aromatic humic acids in three volcanic ash soils as determined by dipolar  
694 dephasing NMR studies. *Geochimica et Cosmochimica Acta* **53**, 125–130.
- 695 Heanes DL (1984) Determination of total organic-C in soils by an improved chromic  
696 acid digestion and spectrophotometric procedure. *Soil Science and Plant Analysis*  
697 **15**, 1191–1213.

- 698 IUSS Working Group WRB (2006) World Reference Base for Soil Resources. In  
699 'World Soil Resources Reports No. 103, 2nd Edition'. (FAO: Rome).
- 700 Kaal J, van Mourik JM (2008) Micromorphological evidence of black carbon in  
701 colluvial soils from NW Spain. *European Journal of Soil Science* **59**, 1133–1140.
- 702 Kaal J, Rumpel C (2009) Can pyrolysis-GC/MS be used to estimate the degree of  
703 thermal alteration of black carbon? *Organic Geochemistry* **40**, 1179–1187.
- 704 Kaal J, Martínez-Cortizas A, Nierop KGJ (2009) Characterisation of aged charcoal  
705 using a coil probe pyrolysis-GC/MS method optimised for black carbon. *Journal*  
706 *of Analytical and Applied Pyrolysis* **85**, 408–416.
- 707 Kaal J, Carrión Y, Asouti E, Martín Seijo M, Costa Casais M, Martínez Cortizas A,  
708 Criado Boado F (2011) Long-term deforestation in NW Spain: Linking the  
709 Holocene fire history to vegetation change and human activities. *Quaternary*  
710 *Science Reviews* **30**, 161–175.
- 711 Kaal J, Schneider MPW, Schmidt MWI (2012a). Rapid molecular screening of black  
712 carbon (biochar) thermosequences obtained from chestnut wood and rice straw: a  
713 pyrolysis-GC/MS study. *Biomass and Bioenergy* **45**, 115–129.
- 714 Kaal J, Nierop KGJ, Kraal P, Preston CM (2012b) A first step towards identification of  
715 tannin-derived black carbon: conventional pyrolysis (Py-GC-MS) and thermally  
716 assisted hydrolysis and methylation (THM-GC-MS) of charred condensed  
717 tannins. *Organic Geochemistry* **47**, 99–108.
- 718 Kleber M, Mikutta R, Torn MS, Jahn R (2005) Poorly crystalline mineral phases protect  
719 organic matter in acid subsoil horizons. *European Journal of Soil Science* **56**,  
720 717–725.

- 721 Kleber M (2010) What is recalcitrant soil organic matter?. *Environmental Chemistry* **7**,  
722 320–332.
- 723 Knicker H, Lüdemann HD (1995)  $^{15}\text{N}$  and  $^{13}\text{C}$  CPMAS and solution NMR studies of  
724  $^{15}\text{N}$  enriched plant-material during 600 days of microbial degradation. *Organic*  
725 *Geochemistry* **23**, 329–341.
- 726 Knicker H, Totsche KU, Almendros G, González-Vila FJ (2005) Condensation degree  
727 of burnt peat and plant residues and the reliability of solid-state VACP MAS  $^{13}\text{C}$   
728 NMR spectra obtained from pyrogenic humic material. *Organic Geochemistry* **36**,  
729 1359–1377.
- 730 Knicker H (2007) How does fire affect the nature and stability of soil organic nitrogen  
731 and carbon? A review. *Biogeochemistry* **85**, 91–118.
- 732 Knicker H, Müller P, Hilscher A (2007) How useful is chemical oxidation with  
733 dichromate for the determination of "Black Carbon" in fire-affected soils?.  
734 *Geoderma* **142**, 178–196.
- 735 Knicker H, Hilscher A, González-Vila FJ, Almendros G (2008) A new conceptual  
736 model for the structural properties of char produced during vegetation fires.  
737 *Organic Geochemistry* **39**, 935–939.
- 738 Kögel-Knabner I (1995) Composition of soil organic matter. In 'Methods in Applied  
739 Soil Microbiology and Biochemistry'. (Eds P Nannipieri, K Alef) pp. 66–78.  
740 (Academic Press: London)
- 741 Kögel-Knabner I (1997)  $^{13}\text{C}$  and  $^{15}\text{N}$  NMR spectroscopy as a tool in soil organic matter  
742 studies. *Geoderma* **80**, 243–270.

- 743 Lefroy RDB, Blair GJ, Strong WM (1993) Changes in soil organic matter with cropping  
744 as measured by organic carbon fractions and  $^{13}\text{C}$  natural isotope abundance. *Plant*  
745 *and Soil* **155/156**, 399–402.
- 746 Loginow W, Wisniewski W, Gonet SS, Ciescinska B (1987) Fractionation of organic C  
747 based on susceptibility to oxidation. *Polish Journal of Soil Science* **20**, 47–52.
- 748 López-Merino L, Silva-Sánchez N, Kaal J, López-Sáez JA, Martínez-Cortizas A (2012)  
749 Post-disturbance vegetation dynamics during the Late Pleistocene and the  
750 Holocene: an example from NW Iberia. *Global Planetary Change* **92/93**, 58–70.
- 751 Macías F, Camps Arbestain M (2010) Soil carbon sequestration in a changing global  
752 environment. *Mitigation and Adaptation Strategies for Global Change* **15**, 511–  
753 529.
- 754 Moldoveanu SC (1998) An introduction to analytical pyrolysis. In ‘Analytical Pyrolysis  
755 of Organic Polymers. Techniques and instrumentation in analytical chemistry’.  
756 (Ed SC Moldoveanu) pp. 20. (Elsevier: Amsterdam).
- 757 Nelson DW, Sommers LE (1996) Total carbon, organic carbon, and organic matter. In  
758 ‘Methods of Soil Analysis 2, 2nd ed’. (Eds DL Sparks, AL Page, PA Helmke, RH  
759 Loeppert, PN, Soltanpour, MA, Tabatabai, CT, Johnston, ME Sumner) pp. 961–  
760 1010. (American Society of Agronomy, Inc: Madison, WI).
- 761 Nierop KGJ, van Bergen F, Buurman P, van Lagen B (2005) NaOH and Na- $\text{Na}_4\text{P}_2\text{O}_7$ -  
762 extractable organic matter in two allophanic volcanic ash soils of the Azores  
763 Islands-a pyrolysis GC/MS study. *Geoderma* **127**, 36–51.
- 764 Piccolo A (1996) Humus and soil conservation. In ‘Humic Substances in Terrestrial  
765 Ecosystems’. (Ed A Piccolo) pp. 225–264. (Elsevier: Amsterdam)



- 766 Pastorova I, Botto RE, Arisz PW, Boon JJ (1994) Cellulose char structure: a combined  
767 analytical Py-GC-MS, FTIR and NMR study. *Carbohydrate Research* **262**, 27–  
768 47.
- 769 Poirier N, Sohi SP, Gaunt JL, Mahieu N, Randall EW, Powlson DS, Evershed RP  
770 (2005) The chemical composition of measurable soil organic matter pools.  
771 *Organic Geochemistry* **36**, 136–151.
- 772 Polvillo O, González-Pérez JA, Boski T, González-Vila FJ (2009) Structural features of  
773 humic acids from a sedimentary sequence in the Guadiana estuary (Portugal–  
774 Spain border). *Organic Geochemistry* **40**, 20–28.
- 775 Rumpel C, González-Pérez JA, Bardoux G, Largeau C, González-Vila FJ, Valentín C  
776 (2007) Composition and reactivity of morphologically distinct charred materials  
777 left after slash-and-burn practices in agricultural tropical soils. *Organic*  
778 *Geochemistry* **38**, 911–920.
- 779 Sáiz-Jiménez C, de Leeuw JW (1986) Chemical characterization of soil organic matter  
780 fractions by analytical pyrolysis-gas chromatography-mass spectrometry. *Journal*  
781 *of Analytical and Applied Pyrolysis* **9**, 99–119.
- 782 Sáiz-Jiménez C (1994a) Analytical pyrolysis of humic substances; pitfalls, limitations,  
783 and possible solutions. *Environmental Science and Technology* **28**, 1173–1780.
- 784 Sáiz-Jiménez C (1994) Production of alkylbenzenes and alkyl-naphthalenes upon  
785 pyrolysis of unsaturated fatty acids. *Naturwissenschaften* **81**, 451–453.
- 786 Samaté AD, Nacro M, Menut C, Malaty G, Bessiere JM (1998) Aromatic plants of  
787 tropical west Africa. VII. Chemical composition of essential oils of two  
788 Eucalyptus species (Myrtaceae) from Burkina Fasso: Eucalyptus alba Muell. and

- 789 Eucalyptus camaldulensis Dehnhardt. *Journal of Essential Oil Research* **10**, 321–  
790 324.
- 791 Schellekens J, Buurman P, Kuyper TW (2012) Source and transformations of lignin in  
792 Carex-dominated peat. *Soil Biology & Biochemistry* **53**, 32–42.
- 793 Schmidt MWI, Skjemstad JO, Czimczik CI, Glaser B, Prentice KM, Gelinás Y,  
794 Kuhlbusch TAJ (2001) Comparative analysis of black carbon in soils. *Global*  
795 *Biogeochemical Cycles* **15**, 163–167.
- 796 Schnitzer MI, Monreal CM, Jandl G, Leinweber P, Fransham PB (2007) The conversion  
797 of chicken manure to biooil by fast pyrolysis II. Analysis of chicken manure,  
798 biooils, and char by curie-point pyrolysis-gas chromatography/mass spectrometry  
799 (Cp Py-GC/MS). *Journal of Environmental Science and Health* **42**, 79–95.
- 800 Schulten HR, Plage B, Schnitzer M (1991) A chemical structure for humic substances.  
801 *Naturwissenschaften* **78**, 311–312.
- 802 Shindo H, Honna T, Yamamoto S, Honma H (2004). Contribution of charred plant  
803 fragments to soil organic carbon in Japanese volcanic ash soils containing black  
804 humic acids. *Organic Geochemistry* **35**, 235–241.
- 805 Six J, Conant RT, Paul EA, Paustian K (2002) Stabilization mechanisms of soil organic  
806 matter: Implications for C-saturation of soils. *Plant and Soil* **241**, 155–176.
- 807 Skjemstad JO, Clarke P, Taylor JA, Oades JM, McClure SG (1996) The chemistry and  
808 nature of protected carbon in soil. *Australian Journal of Soil Resources* **34**, 251–  
809 271.

- 810 Skjemstad JO, Taylor JA (1999) Does the Walkley-Black method determine soil  
811 charcoal?. *Communications in Soil Science and Plant Analysis* **30**, 2299–2310.
- 812 Skjemstad JO, Swift RS, McGowan JA (2006) Comparison of the particulate organic  
813 carbon and permanganate oxidation methods for estimating labile soil organic  
814 carbon. *Soil Research* **44**, 255–263.
- 815 Soil Survey Staff (1998) ‘Keys to Soil Taxonomy (8th ed.)’. USDA Natural Resources  
816 Conservation Sites (NRCS): Washington.
- 817 Song J, Peng P (2010) Characterisation of black carbon materials by pyrolysis-gas  
818 chromatography-mass spectrometry. *Journal of Analytical Applied Pyrolysis* **87**,  
819 129–137.
- 820 Stankiewicz BA, van Bergen PF, Duncan IJ, Carter JF, Briggs DEG, Evershed RP  
821 (1996) Recognition of chitin and proteins in invertebrate cuticles using analytical  
822 pyrolysis/gas chromatography and pyrolysis/gas chromatography/mass  
823 spectrometry. *Rapid Communications in Mass Spectroscopy* **10**, 1747–1757.
- 824 Stevenson FJ (1994) ‘Humus Chemistry, Genesis, Composition, Reactions 2nd ed’.  
825 (John Wiley and Sons: New York).
- 826 Stuczynski TI, McCarthy GW, Reeves JB, Wright RJ (1997) Use of pyrolysis GC/MS  
827 for assessing changes in soil organic matter quality. *Soil Science* **162**, 97–105.
- 828 Suárez-Abelenda M, Buurman P, Camps Arbestain M, Kaal J, Martínez-Cortizas A,  
829 Gartzia-Bengoetxea N, Macías F (2011) Comparing NaOH-extractable organic  
830 matter of acid forest soils that differ in their pedogenic trends: a pyrolysis-GC/MS  
831 study. *European Journal of Soil Science* **62**, 834–848.

- 832 Tabatabai MA (1996) Soil organic matter testing: an overview. In 'Soil organic matter:  
833 Analysis and interpretation. SSSA Special Publication 46'. (Eds FR Magdoff, MA  
834 Tabatabai, EA Hanlon) pp. 1–9. (SSSA: Madison, WI)
- 835 Tegelaar EW, de Leeuw JW, Sáiz-Jiménez C (1989) Possible origin of aliphatic  
836 moieties in humic substances. *Science of the Total Environments* **81/82**, 1–17.
- 837 Tegelaar EW, Hollman G, Vandervegt P, de Leeuw JW, Holloway PJ (1995) Chemical  
838 characterization of the periderm tissue of some angiosperm species: Recognition  
839 of an insoluble, nonhydrolyzable, aliphatic biomacromolecule (suberan). *Organic  
840 Geochemistry* **23**, 239–251.
- 841 Tirol-Padre A, Ladha JK (2004) Assessing the reliability of permanganate- oxidizable  
842 carbon as an index of soil labile carbon. *Soil Science Society of America Journal*  
843 **68**, 969–978.
- 844 van der Kaaden A, Boon JJ, de Leeuw JW, de Lange F, Wijnand Schuyf PJ, Schulten  
845 HR, Bahr U (1984) Comparison of analytical pyrolysis techniques in the  
846 characterization of chitin. *Analytical Chemistry* **56**, 2160–2165.
- 847 van Soest PJ, Wine RH (1986) Determination of lignin and cellulose in acid-detergent  
848 fibre with permanganate. *Journal of the Association of Official Agricultural  
849 Chemists* **51**, 780–785.
- 850 Wilson MA, Barron PF, Goh KM (1981) Cross polarisation  $^{13}\text{C}$  NMR spectroscopy of  
851 some genetically related New Zealand soils. *Journal of Soil Science* **32**, 419–425.
- 852 Wolbach WS, Anders E (1989) Elemental carbon in sediments: Determination and  
853 isotopic analysis in the presence of kerogen. *Geochimica et Cosmochimica Acta*  
854 **53**, 1637–1647.

855 Figure captions

856 Fig. 1. Example of pyrograms from permanganate- and dichromate-oxidation resistant  
857 soil organic matter, and that from non-oxidized samples.

858 Fig. 2. Projection of the factor loadings of the pyrolysis products and sample scores in  
859 F1-F2 space. The corresponding pyrolysis product codes are given in Appendix A.

860 Fig. 3. Solid- state  $^{13}\text{C}$  NMR spectra of the NaOH-extracts of untreated samples (NO-1,  
861 NO-2 and NO-3) and the NaOH-extracts of potassium permanganate and dichromate  
862 oxidized residues of S1 sample (MN-1 and CR-1) and the NaOH-extract of the  
863 potassium dichromate oxidized residues of S2 sample (CR-2).

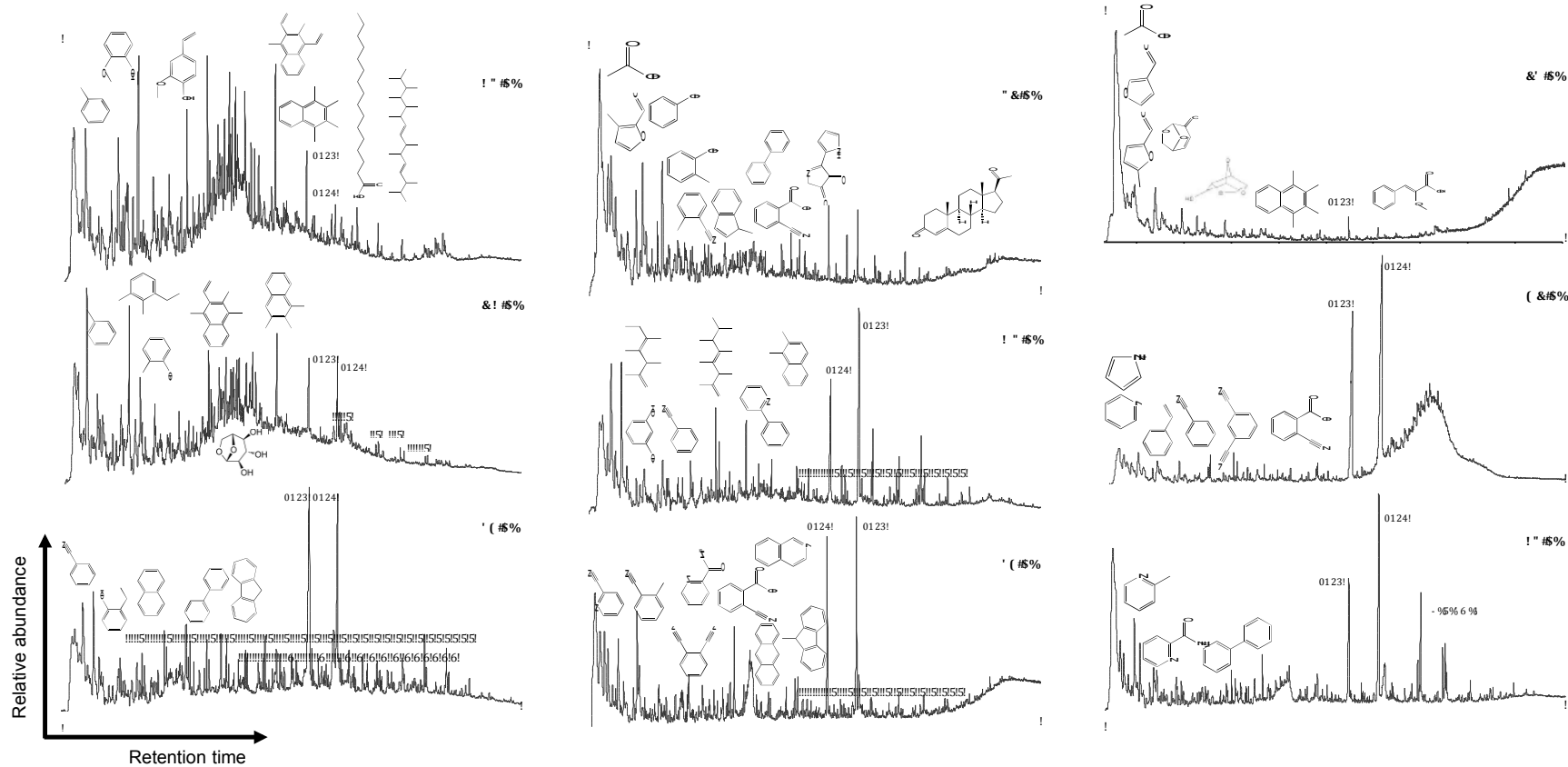


Fig. 1

NO-1: toluene, guaiacol, 4-vinyl guaiacol, C<sub>4:1</sub> naphthalene compound, C<sub>4</sub>-naphthalene, hexadecanoic acid (C<sub>16</sub> *n*-fatty acid) and octadecanoic acid (C<sub>18</sub> *n*-fatty acid), and branched alkene. MN-1: toluene, methyl phenol, 2-ethyl 3,4-dimethyl phenol, C<sub>4:1</sub> naphthalene compound, levoglucosan, C<sub>3</sub>-naphthalene, C<sub>16</sub> *n*-fatty acid and C<sub>18</sub> *n*-fatty acid, and series of pairs of *n*-alkanes/alkenes (\*). CR-1: benzonitrile, ethyl phenol, naphthalene, biphenyl, fluorene and regular series of pairs of *n*-alkanes/alkenes (\*) and *n*-fatty acids (†). NO-2: acetic acid, methyl furaldehyde, phenol, methyl phenol, methyl benzonitrile, methyl indene, biphenyl, cyanobenzoic acid, diketodipyrrole and sterol compound (preg-4-ene 3,20-dione compound). MN-2: branched alkenes, resorcinol, benzonitrile, phenyl pyridine, C<sub>2</sub> naphthalene, C<sub>16</sub> *n*-fatty acid and C<sub>18</sub> *n*-fatty acid and regular series of pairs of *n*-alkanes/alkenes (\*). CR-2: pyridine carbonitrile, methyl benzonitrile, benzenedicarbodinitrile, picolinamide, anthracene, methyl fluorene, cyanobenzoic acid, isoquinoline, C<sub>16</sub> *n*-fatty acid and C<sub>18</sub> *n*-fatty acid and regular series of pairs of *n*-alkanes/alkenes (\*). NO-3: acetic acid, furaldehyde, methyl furaldehyde, levoglucosone, dianhydro- $\alpha$ , D-glucopyranose, C<sub>4</sub>-naphthalene and C<sub>16</sub> *n*-fatty acid. MN-3: pyridine, pyrrole, styrene, benzonitrile, benzenedicarbodinitrile, cyanobenzoic acid and C<sub>16</sub> *n*-fatty acid and C<sub>18</sub> *n*-fatty acid. CR-3: methyl pyridine, picolinamide, biphenyl, C<sub>16</sub> *n*-fatty acid and C<sub>18</sub> *n*-fatty acid and a set of unidentified compounds.

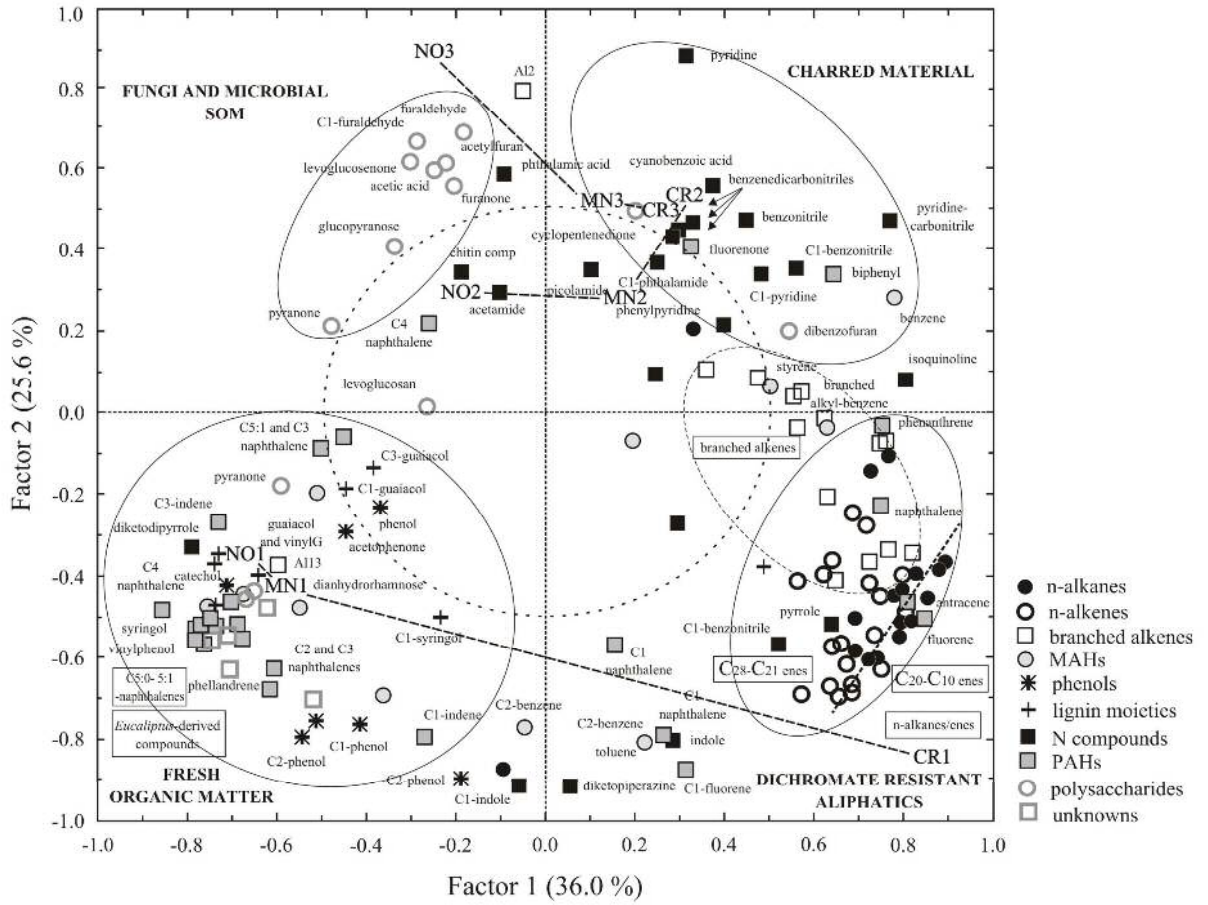


Fig.2



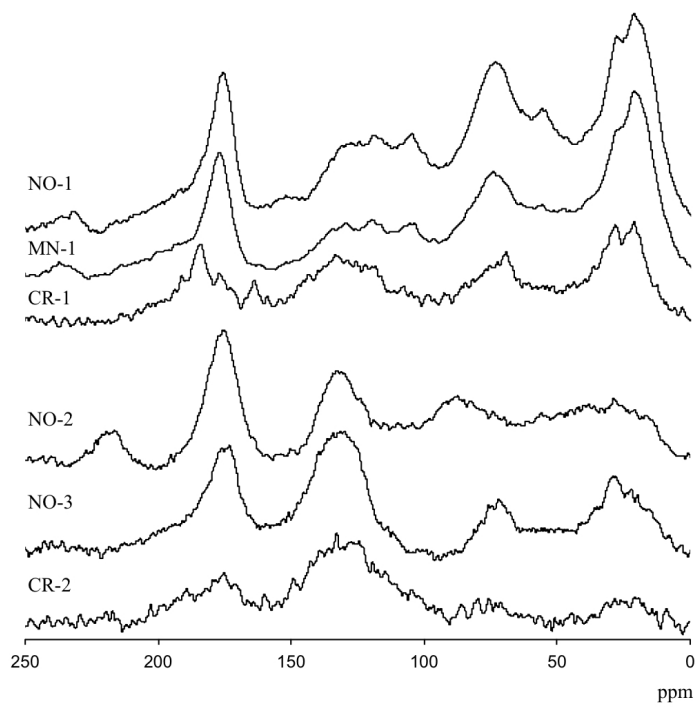


Fig.3

Table 1. General information of samples studied from soil PRD-4.  $OC_{per}$  = permanganate-oxidizable organic C,  $OC_{dichro}$  = dichromate-oxidizable C.

Sample	S1	S2	S3
depth	5-10 cm	95-100 cm	190-195 cm
conventional $^{14}C$ age BP	$104.3 \pm 0.4$ pMC (present)	$4090 \pm 30$	$9760 \pm 50$
radiocarbon sample code	Ua-34719	$\beta$ -299230	$\beta$ -240963
C [ $mg\ g^{-1}$ soil]	$62.3 \pm 0.6$	$36.7 \pm 0.3$	$42.9 \pm 0.4$
$OC_{per}$ [ $mg\ g^{-1}$ soil]	$2.3 \pm 0.1$	$1.3 \pm 0.1$	$1.7 \pm 0.1$
$OC_{dichro}$ [ $mg\ g^{-1}$ soil]	$27.5 \pm 0.6$	$12.3 \pm 0.3$	$14.3 \pm 0.1$
C/N (atomic) [-]	15.9	24.4	23.9
pH- $H_2O$ [-]	4.6	5.0	5.2
charcoal $>2$ mm [ $mg\ g^{-1}$ soil]	0.03	1.97	0.07

For Review Only

Table 2. Relative contributions of pyrolysis product groups and benzene/alkylbenzenes and PAH/alkyl-PAHs ratios of total quantified peak area (% TQPA).

		total n-enes	n- enes >C <sub>18</sub>	n-enes C <sub>18</sub> -C <sub>10</sub>	total n- anes	n-anes >C <sub>18</sub>	n-anes C <sub>18</sub> - C <sub>10</sub>	other aliph <sup>+</sup>	phs	C <sub>2</sub> - ph	total Lg	total MAHs	alkyl- B	total PAHs	ITPB- PAHs	BC- PAHs	total N	BN	total Ps	well pres Ps	degr Ps	U	benz/ alkyl- benz	PAH/ alkyl- PAHs
NO-1	TQPA %	<b>1.7</b>	0.9	0.8	<b>1.9</b>	0.8	1.1	1.4	<b>16.6</b>	1.8	<b>7.4</b>	<b>16.4</b>	6.6	<b>10.6</b>	10.1	0.5	<b>10.7</b>	3.6	<b>30.4</b>	17.8	12.4	2.5	0.2	0.02
	%*		51.4	48.6		40.1	59.9						40.0		95.6	4.4		33.2		58.4	40.7			
MN-1	TQPA %	<b>1.6</b>	0.5	1.1	<b>2.0</b>	0.3	1.7	4.1	<b>14.1</b>	1.9	<b>3.0</b>	<b>30.0</b>	13.4	<b>12.3</b>	11.4	1.0	<b>11.7</b>	4.8	<b>18.4</b>	10.4	7.7	2.7	0.2	0.04
	%*		33.6	66.4		13.4	86.6						44.7		92.2	7.8		40.7		56.6	41.7			
CR-1	TQPA %	<b>9.3</b>	3.0	6.3	<b>10.2</b>	3.3	6.9	14.6	<b>10.9</b>	1.4	<b>1.0</b>	<b>26.3</b>	4.6	<b>4.8</b>	2.2	2.6	<b>15.8</b>	7.4	<b>6.4</b>	0.8	5.4	0.7	1.2	0.60
	%*		32.2	67.8		32.2	67.8						17.5		45.7	54.3		47.0		12.4	84.7			
NO-2	TQPA %	<b>1.8</b>	0.6	1.2	<b>1.8</b>	0.6	1.2	1.7	<b>13.7</b>	0.8	<b>1.8</b>	<b>20.4</b>	4.9	<b>5.7</b>	4.2	1.5	<b>20.9</b>	11.9	<b>31.5</b>	4.0	27.1	0.7	0.9	0.21
	%*		34.2	65.8		34.1	65.9						23.8		73.0	27.0		57.1		12.6	86.3			
MN-2	TQPA %	<b>2.5</b>	0.7	1.9	<b>3.2</b>	0.5	2.7	17.4	<b>11.6</b>	0.9	<b>1.4</b>	<b>20.4</b>	3.6	<b>3.3</b>	2.2	1.2	<b>22.2</b>	13.9	<b>17.4</b>	3.3	13.6	0.5	1.3	0.26
	%*		26.0	74.0		16.7	83.3						17.5		65.0	35.0		62.6		18.7	77.9			
CR-2	TQPA %	<b>4.6</b>	1.6	3.0	<b>2.5</b>	0.9	1.5	7.0	<b>4.8</b>	0.5	<b>0.8</b>	<b>16.9</b>	2.0	<b>3.4</b>	0.9	2.5	<b>32.2</b>	23.8	<b>27.6</b>	15.7	11.5	0.2	3.9	1.31
	%*		35.3	64.7		37.7	62.3						11.6		25.3	74.7		74.1		57.0	41.8			
NO-3	TQPA %	<b>0.9</b>	0.3	0.5	<b>1.0</b>	0.5	0.5	1.5	<b>6.6</b>	0.0	<b>0.7</b>	<b>10.3</b>	2.5	<b>1.4</b>	0.9	0.5	<b>15.0</b>	9.0	<b>62.2</b>	2.7	59.2	0.3	1.3	0.35
	%*		36.6	63.4		47.3	52.7						24.0		65.0	35.0		60.2		4.4	95.2			
MN-3	TQPA %	<b>2.7</b>	0.8	1.9	<b>2.9</b>	1.2	1.7	9.2	<b>7.7</b>	0.0	<b>1.2</b>	<b>17.5</b>	4.1	<b>2.4</b>	1.4	1.0	<b>38.3</b>	28.9	<b>17.7</b>	1.1	16.0	0.4	0.8	0.34
	%*		30.2	69.8		42.0	58.0						23.2		57.9	42.1		75.5		6.5	90.7			
CR-3	TQPA %	<b>3.5</b>	1.2	2.3	<b>2.6</b>	1.1	1.5	10.7	<b>5.0</b>	0.0	<b>2.2</b>	<b>22.3</b>	2.7	<b>2.5</b>	0.7	1.7	<b>22.1</b>	16.5	<b>28.8</b>	16.5	11.4	0.3	2.7	0.89
	%*		35.0	65.0		41.1	58.9						11.9		30.0	70.0		74.8		57.5	39.6			

\*relative proportions within main group (*n*-alkanes/ enes, fatty acids, MAHs, PAHs, nitrogen compounds and polysaccharides). In bold the main groups.

<sup>+</sup>the aliphatic compound with mass 83+280 (likely associated to fresh OM) was not added because is not indicative of the charring effect.

Total *n*-enes: total *n*-alkenes; *n*-enes >C<sub>18</sub>: long-chain *n*-alkenes (>C<sub>18</sub>); *n*-enes C<sub>18</sub>-C<sub>10</sub>: short-chain *n*-alkenes (C<sub>18</sub>-C<sub>10</sub>); total *n*-anes: total *n*-alkanes; *n*-anes >C<sub>18</sub>: long-chain *n*-alkanes (>C<sub>18</sub>); *n*-anes C<sub>18</sub>-C<sub>10</sub>: short-chain *n*-alkanes (C<sub>18</sub>-C<sub>10</sub>); other aliph: other aliphatic compounds (predominantly branched alkenes); phs: phenols; total Lg: total lignin markers; total MAHs: total Monocyclic Aromatic Hydrocarbons; alkyl-B: alkyl-benzenes; total PAHs: total Polycyclic Aromatic Hydrocarbons; ITPB-PAHs: intact terpene-like plant biomass Polycyclic Aromatic Hydrocarbons; BC-PAHs: black carbon derived Polycyclic Aromatic Hydrocarbons; total N: total nitrogen compounds; BN: black carbon derived nitrogen compounds; total Ps: total polysaccharides; well pres Ps: well preserved polysaccharides; degr Ps: degraded polysaccharides; U: unidentified compounds; benz/alkyl-benz: benzene/alkyl-benzenes ratio and PAHs/alkyl-PAHs: total Polycyclic Aromatic Hydrocarbons/ alkylated Polycyclic Aromatic Hydrocarbons ratio.

Table 3. Chemical shift region distribution (relative proportions, %) obtained from solid-state  $^{13}\text{C}$  NMR.

	Alkyl C (0-45 ppm)	N-alkyl C, methoxyl C (45-60 ppm)	O-alkyl C (60-110 ppm)	C-H Aromatic C (110-140 ppm)	COR Aromatic C (140-160 ppm)	Carboxyl C, amide C (160-220 ppm)	Ketone C, aldehyde C (220-245 ppm)
NO-1	29	9	29	14	3	15	1
MN-1	34	8	26	11	2	17	1
CR-1	29	4	22	19	6	20	0
NO-2	23	5	26	19	1	21	4
CR-2	10	2	16	42	3	23	4
NO-3	19	5	11	31	5	26	3

For Review Only

Appendix A. Pyrolysis product list, molecular mass ( $M^+$ ), fragment ions used for quantification and retention times relative to guaiacol (RT).

code	Name	$M^+$	mass	RT/guaiacol
10:1 - 28:1	C10-28 alkene	140 - 392	55+69	0.832 - 3.965
10:0 - 28:0	C10-28 alkane	142 - 394	57+71	0.857 - 3.969
A11	aliphatic compound	n.d.	55+70	0.549
A12	alkane/anal or methylated alkanol	n.d.	57+69+70	1.157
A13	branched alkene	n.d.	55+69	1.160
A14	branched alkene	n.d.	55+69	1.573
A15	branched alkene	n.d.	55+69	1.590
A16	branched alkene	n.d.	55+69	1.610
A17	alkene	n.d.	55+69	1.901
A18	alkene	n.d.	55+69	2.301
A19	branched alkene	n.d.	55+69	2.428
A110	branched alkene	n.d.	55+69	2.941
A111	alkene	n.d.	55+69	3.069
A112	branched alkene	n.d.	55+69	3.272
A113	aliphatic compound	83+280	83+280	3.342
A114	branched alkene	n.d.	55+69	3.565
A115	branched alkene	n.d.	55+69	3.799
Ph1	phenol	94	66+94	0.782
Ph2	acetophenone	120	77+105	0.937
Ph3	C1-phenol	108	107+108	0.941
Ph4	C1-phenol	108	107+108	0.991
Ph5	C2-phenol	122	107+122	1.158
Ph6	C2-phenol	122	107+122	1.202
Ph7	catechol	110	110+64	1.348
Lg1	guaiacol	124	109+124	1.000
Lg2	4-methylguaiacol	138	123+138	1.245
Lg3	4-vinylphenol	120	120+91	1.326
Lg4	4-vinylguaiacol	150	135+150	1.519
Lg5	syringol	154	154+139	1.585
Lg6	4-methylsyringol	168	153+168	1.798
Lg7	C3-guaiacol	196	149+164	1.815
Lg8	Propenoic acid, 3-(4-methoxyphenol)	178	161+178	3.296
Ar1	benzene	78	78	0.347
Ar2	toluene	92	91+92	0.431
Ar3	C2-benzene ethyl benzene	106	91+106	0.559
Ar4	C2-benzene dimethyl benzene	106	91+106	0.574
Ar5	styrene	104	78+104	0.605
Ar6	C2-benzene dimethyl benzene	106	91+106	0.612
Ar7	C3-benzene	120	105+120	0.815
Ar8	C4-benzene	134	91+119	0.887
Ar9	C5-benzene (dimethyl-methylethyl)	148	133	1.160
Ar10	C4:1-benzene ( $\alpha$ -dimethylstyrene)	132	132+117	1.026
Ar11	C7-benzene	176	91+92	1.662
Ar12	branched alkyl-benzene	148	91+119	1.194

Pa1	C1-indene	130	130+115	1.146
Pa2	naphthalene	128	128	1.231
Pa3	C1-naphthalene	142	141+142	1.491
Pa4	C1-naphthalene	142	141+142	1.523
Pa5	biphenyl	154	154	1.668
Pa6	C3-indene	158	143+158	1.723
Pa7	C2-naphthalene	156	141+156	1.772
Pa8	C3-naphthalene	170	155+170	2.052
Pa9	C3-naphthalene	170	155+170	2.056
Pa10	Fluorene	166	165+166	2.090
Pa11	C3-naphthalene	170	155+170	2.103
Pa12	C1-Fluorene	180	165+180	2.314
Pa13	9H-Fluoren-9-one	180	152+180	2.353
Pa14	C4-naphthalene	184	169+184	2.382
Pa15	phenanthrene	178	178	2.442
Pa16	anthracene	178	178	2.461
Pa17	C5-naphthalene	198	183+198	2.470
Pa18	C5-naphthalene (or C2-azulene)	198	183+198	2.643
Pa19	C5:1-naphthalene	202	159+145	1.649
Pa20	C5-naphthalene	204	147+162	1.739
Pa21	C5:1-naphthalene	202	159+202	1.767
Pa22	C5-naphthalene	204	91+105	1.801
Pa23	C5-naphthalene	204	105+133	1.850
Pa24	C5-naphthalene	204	91+105	1.853
Pa25	C5-naphthalene	204	91+105	1.928
Pa26	C5-naphthalene	204	91+105	1.974
Pa27	C5-naphthalene	204	173+189	1.987
Pa28	C4-naphthalene	186	143+171	2.000
Pa29	C5:1-naphthalene	202	159+180	2.037
Pa30	C5:1-naphthalene	202	146+133	2.143
Pa31	C5:1-naphthalene	202	159+202	2.211
N1	pyridine	79	52+79	0.396
N2	pyrrole	67	67	0.402
N3	acetamide	n.d	59	0.421
N4	C1-pyrrole	81	80+81	0.501
N5	C1-pyrrole	81	80+81	0.521
N6	C1-pyridine	93	66+93	0.543
N7	benzonitrile	103	76+103	0.749
N8	pyridinecarbonitrile	104	104+77	0.887
N9	C1-benzonitrile	117	90+117	1.011
N10	C1-benzonitrile	117	90+117	1.067
N11	1,3-benzenedicarbonitrile	128	101+128	1.302
N12	picolinamide	122	79+122	1.320
N13	1,3-benzenedicarbonitrile	128	101+128	1.322
N14	isoquinoline	129	129	1.334
N15	indole	117	90+117	1.457
N16	phthalamic acid	104	104+76	1.457
N17	1,3-benzenedicarbonitrile	128	101+128	1.463
N18	C1-indole	131	130+131	1.664
N19	C1-phthalimide	161	161+76	1.719

N20	cyanobenzoic acid	147	76+147	1.827
N21	phenylpyridine	155	154+155	1.833
N22	chitin-derived compound	167	125+167	1.883
N23	diketodipyrrole	186	93+186	2.280
N24	diketopiperazine compound	194	70+194	2.646
Ps1	acetic acid	60	60	0.203
Ps2	furanone compound	84	54+84	0.434
Ps3	3/2-furaldehyde	96	95+96	0.483
Ps4	acetyl furan	110	95+110	0.613
Ps5	5-methyl-2-furaldehyde	110	109+110	0.686
Ps6	4-hydroxy-5,6-dihydro-(2H)-pyranone	114	58+114	0.755
Ps7	dianhydrorhamnose	128	113+128	0.861
Ps8	cyclopentenedione compound	112	69+112	0.869
Ps9	levoglucosenone	126	68+98	0.989
Ps10	3-hydroxy-2-methyl-4H-pyran-4-one	126	71+126	1.046
Ps11	dianhydro- $\alpha$ ,D-glucopyranose	144	57+69	1.263
Ps12	dibenzofuran	168	168+139	1.953
Ps13	levoglucosan	162	60+73	2.123
U1	alpha phellandrene	136	91+93+136	0.827
U2	U2 (possibly methylcyclohexane)	96	67+96	0.710
U3	U3	157	117+157	1.871
U4	U4	200	185+200	2.186
U5	U5	212	197+202	2.247
U6	U6	203	203	3.152