

**THE SURFACE WATERS ACIDIFICATION PROJECT
PALAEO LIMNOLOGY PROGRAMME : MODERN
DIATOM/LAKE-WATER CHEMISTRY DATA-SET**

A.C. Stevenson¹, S. Juggins², H.J.B. Birks², D.S. Anderson³, N.J. Anderson⁴, R.W. Battarbee⁵, F. Berge⁶, R.B. Davis³, R.J. Flower⁵, E.Y. Haworth⁷, V.J. Jones⁵, J.C. Kingston⁸, A.M. Kreiser⁵, J.M. Line⁹, M.A.R. Munro⁵, I. Renberg¹⁰.

¹ Department of Geography, University of Newcastle upon Tyne, Newcastle upon Tyne, NE1 7RU, U.K.

² Botanical Institute, University of Bergen, Allégaten 41, N-5007 Bergen, Norway.

³ Department of Botany, Deering Hall, University of Maine, Orono, Maine 04469, U.S.A.

⁴ Danish Geological Survey, Thoravej 8, DK-2400, Copenhagen NV, Denmark.

⁵ Palaeoecology Research Unit, Department of Geography, University College London, 26 Bedford Way, London, WC1H 0AP, U.K.

⁶ 2920 Leira, Valdres, Norway.

⁷ Institute of Freshwater Ecology, Ferry House, Ambleside, Cumbria, LA22 0LP, U.K.

⁸ Department of Biology, Queen's University, Kingston, Ontario K7L 3N6, Canada.

⁹ University of Cambridge Computer Laboratory, Pembroke Street, Cambridge CB2 3QG, U.K.

¹⁰ Department of Ecological Botany, University of Umeå, S-901 87 Umeå, Sweden.

ISBN. 1 871275 05 9

1991

ENSIS Publishing

London

CONTENTS

Preface	1
Introduction	3
Field and laboratory methods	3
Sample collection and diatom preparation	3
Chemistry and other environmental features of the lakes sampled	4
Diatom taxonomy and nomenclature	11
Introduction	11
Slide exchange and analytical quality control	12
Initial identification of problems	12
a) Nomenclature	12
b) Splitting and amalgamation of taxa	12
c) Identifications	12
Applying and refining the SWAP taxonomic code	14
Protocols for handling unknowns	16
Problems within the genus <i>Aulacoseira</i>	18
Diatom database	18
Numerical analysis of the chemical and other environmental data	19
Basic statistics	19
Principal components analysis	30
Numerical analysis of the diatom data	38
Basic statistics	38
Detrended correspondence analysis	39
Canonical correspondence analysis	42
Reconstruction of lake surface-water chemistry from diatom assemblages	52
Diatom indicator taxa	57
Conclusions	63
Bibliography	64
Appendix A The SWAP taxonomic guide - revised 1989	68
Appendix B List of diatoms in the SWAP training sets and their optima for pH, DOC, and total Al estimated by weighted averaging regression	78

PREFACE

In 1983, when the Surface Waters Acidification Programme (SWAP) was announced, we were asked to design and implement a palaeolimnology sub-project involving scientists from Sweden, Norway, and the UK. Our aim was to reconstruct the acidification history of a range of sites in the three countries and to identify and evaluate the various alternative causes of lake acidification. The results of the project have been published recently (Battarbee *et al.* 1990, Renberg and Battarbee 1990). Although a comprehensive range of palaeolimnological methods and approaches was used in the study we recognised diatom analysis as central to the entire project. We consequently committed considerable effort to improving our diatom methodology and we were especially concerned with the pursuit of a common approach to diatom taxonomy and pH reconstruction. This effort centred on the creation and analysis of a large data-set of surface-sediment diatom assemblages and associated environmental variables from 170 sites representing the full range of lake types in the acid-sensitive and acidified regions of the three countries.

We achieved taxonomic harmonisation by agreeing on taxonomic distinctions in workshop discussions and by testing and reinforcing the decisions we took by "blind" analysis of quality control test slides from the calibration set (Munro *et al.* 1990). For pH reconstruction we achieved a common methodology by using the calibration data-set along with the help and inspiration of Cajo ter Braak to develop a new, statistically robust technique for pH reconstruction (Birks *et al.* 1990a), which is implemented by the program WACALIB (Line and Birks 1990). Most recently this approach has been extended for dissolved organic carbon (DOC) and total aluminium reconstructions (Birks *et al.* 1990b).

This paper describes the calibration data-set, summarises the taxonomic decisions taken, and presents an initial statistical analysis of both the diatom and environmental data. It also includes a summary of the approach taken in the project for water chemistry reconstruction and lists potentially good diatom indicators for pH, DOC, and total aluminium.

The Palaeolimnology Programme within SWAP owes thanks to the many participants in Norway, Sweden, and the UK, to the SWAP Management Committee, and to the sponsors, the Central Electricity Generating Board (UK) and the National Coal Board (UK). At this final point we are especially indebted to Tony Stevenson, Steve Juggins, and John Birks for the time and acumen they have devoted to this ultimate stage of our combined work. We would also like to thank Hazel Juggins for her help with preparing text and camera-ready copy, Sylvia Peglar for her assistance with diagram preparation, Simon Patrick of ENSIS for organising publication, and Cajo ter Braak for providing a

pre-release version of CANOCO 3.10 and for his invaluable statistical advice throughout.

R.W. Battarbee, London

I. Renberg, Umeå

INTRODUCTION

The development of a large calibration, training set of over 170 surface-sediment diatom assemblages and associated environmental variables for the Surface Waters Acidification Programme (SWAP) Palaeolimnology Sub-Project necessitated the amalgamation of several existing regional training sets from Scotland, England, Wales, Norway, and Sweden so that the samples would represent a wide range of water chemistry, catchment sensitivity to acidic deposition, and pollution loadings. This paper documents the SWAP training set, both its diatom assemblages and the associated chemical and physical variables, outlines the methods used in sampling, diatom quality control, and taxonomic harmonisation, and provides a summary of the major patterns of variation in both the environmental and the diatom data using a range of numerical procedures. The joint variation of the diatoms in relation to the environmental variables in the training set is explored using multivariate direct gradient analytical techniques. Finally the ability of the training set to reconstruct lake-water pH, total aluminium, and dissolved organic carbon (DOC) is evaluated, and the possible use of diatoms as "indicator species" is investigated.

FIELD AND LABORATORY METHODS

Sample collection and diatom preparation

Surface-sediment samples were usually taken from the deepest point in each lake using modified Kajak, Hongve, or mini-Mackereth corers operated from boats or from the frozen lake surface. In a few cases the deepest point was not the optimal location and an alternative sampling location was selected nearby. In all cases the top 0.5 cm was used for diatom preparation following the procedures outlined in Stevenson *et al.* (1987). At least 500 valves were counted from each sample or, if planktonic diatoms were common, at least 300 valves of periphyton taxa were counted.

The diatom/water chemistry training set used within the SWAP project is derived from five regional data-sets, some of which were already extant at the commencement of the project (Table 1). A total of 178 samples derived from 170 sites is included in the data-set.

Table 1 Origin of samples by region and analyst

	Number of samples
1. Sweden - provided by I. Renberg & N.J. Anderson	30
2. Norway - provided by F. Berge, D.S. Anderson, & R.B. Davis	51
3. Scotland - provided by R.J. Flower	60
4. Wales - provided by R.J. Flower	32
5. Lake District - provided by E.Y. Haworth	5

Table 2 presents a summary of the site data including the country-specific short code, grid reference, latitude and longitude, lake area, and forest status. The variables forested, unforested, conifer forest, and deciduous forest are nominal (presence/absence) variables. The percentage afforestation is the approximate percentage of the lake's catchment that is afforested, usually by planted conifers. This variable is only relevant for the Welsh, Scottish, and English sites. Sites are individual lakes, and there are some lakes with replicate samples collected at different times. Figure 1 shows the site locations in relation to current sulphur deposition loadings.

Chemistry and other environmental features of the lakes sampled

Water samples were collected from each lake at either the outflow or the lake centre using acid-washed plastic bottles and kept cool and in the dark until analysis.

Most lakes were sampled at three to four monthly intervals over the period 1985 - 1988 and the results presented here are arithmetic means except for pH which is a geometric mean. In some cases, especially for some remote sites and lakes with difficult access, the chemistry is based on one sample.

As far as possible a standardised water-analytical procedure was used between all laboratories involved in the SWAP project. The participating laboratories included:

- i) Scotland Department of Agriculture and Fisheries for Scotland, Pitlochry (R. Harriman)
 Solway River Purification Board, Dumfries (D. Tervet)
- ii) Wales Welsh Water, Swansea (R. West)
- iii) England Institute of Freshwater Ecology, Ambleside (D.W. Sutcliffe)

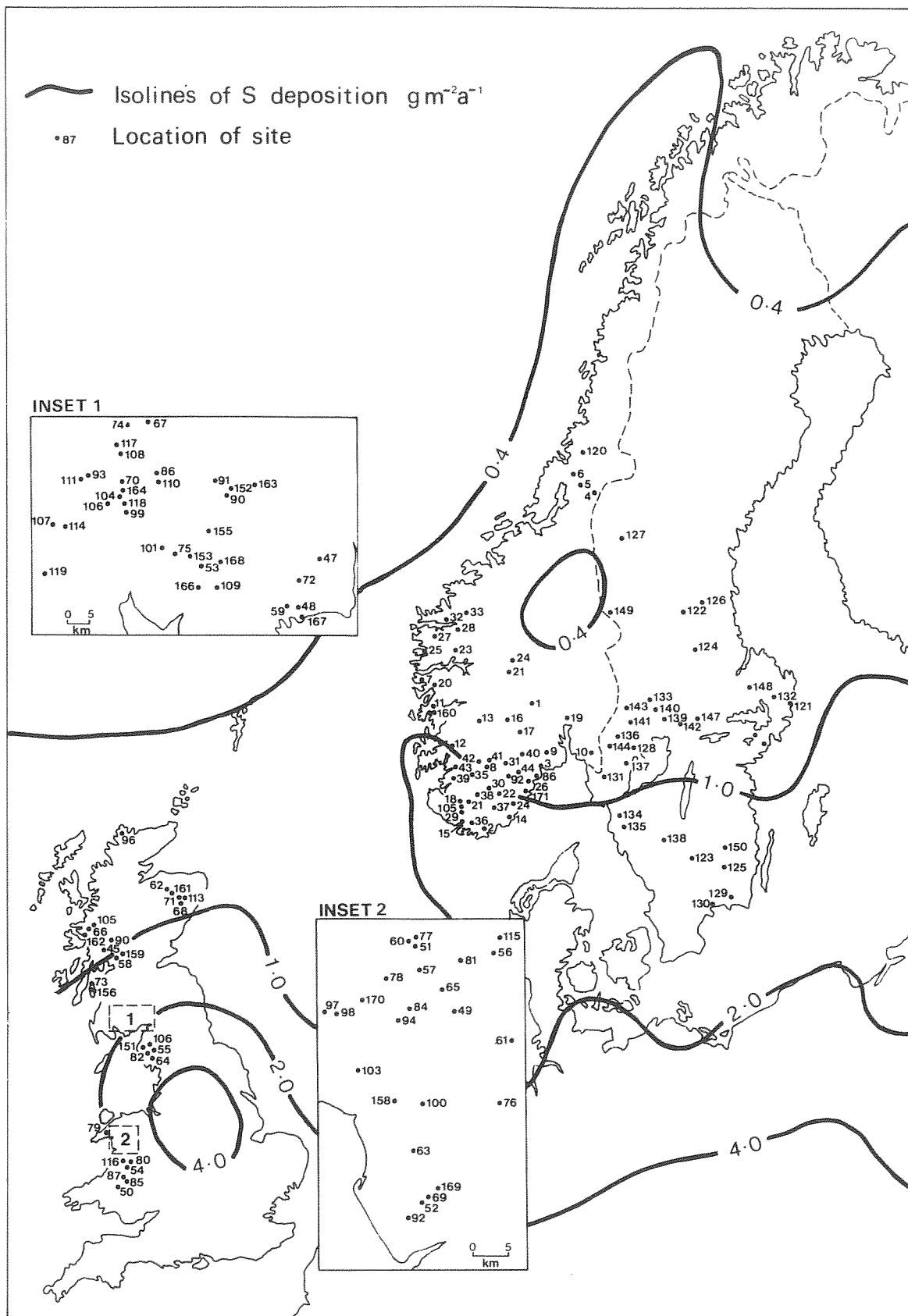


Figure 1 Map of the 170 sites where surface-sediments for diatom analysis and samples for chemistry were obtained. Superimposed are the sulphur deposition isolines derived from Eliassen et al. (1988). Site numbers follow Table 2.

Table 2

Details of the 170 lakes from which the 178 diatom surface assemblages were collected. Site numbers refer to the numbers on Figure 1. Forestry status refers to: P = forestry present, C = conifer forest present, D = deciduous forest present, % = approximate percentage of catchment afforested. Country codes are N = Norway, SCO = Scotland, CYM = Wales, ENG = England, S = Sweden.

Site Number	Site Code	Site Name	Country	Grid	Grid Square	Easting	Northing	Latitude	Longitude	Altitude (m)	Forest			Maximum Depth (m)	Lake Area (ha)
											P	C	D		
1	1	Langtjern	N	32V	NM	540	6692	60.374	9.731	516	1	1	0	12.0	23
2	1.2	Moglandsvatn	N	32V	MK	407	6440	58.090	7.423	176	1	1	0	28.7	77
3	10.2	Langvatn	N	32V	NL	525	6537	58.971	9.435	70	1	1	0	34.0	17
4	111.1	Gronlivatn	N	33W	UM	383	7100	64.006	12.607	514	1	1	0	13.9	215
5	113.2	Store Reinsjoen	N	33W	UM	362	7141	64.365	12.141	454	1	1	0	43.5	170
6	115.1	Grassjoen	N	33W	UM	365	7169	64.617	12.177	153	1	1	0	30.0	60
7	12	Ovre Botnatjonn	N	32V	FH	301	6758	60.907	5.334	180	1	1	0	9.0	5
8	12.1	Brarvatn	N	32V	ML	427	6574	59.297	7.718	902	0	0	0	27.2	125
9	15.1	Myklevann	N	32V	NL	540	6587	59.418	9.705	422	1	1	0	61.0	600
10	17.2	Klosa	N	32V	PL	645	6577	59.306	11.547	172	1	1	0	14.0	100
11	18	Rodlivatn	N	31V	FG	306	6687	60.272	5.492	50	1	1	0	20.0	25
12	18.1	Botnavatn	N	32V	LM	365	6630	59.784	6.595	865	0	0	0	46.0	110
13	19.2	Fjellsjaen	N	32V	MM	425	6663	60.096	7.651	1197	0	0	0	14.2	212
14	2	Hogkleivvatn	N	32V	MK	455	6469	58.361	8.231	160	1	1	0	11.0	8
15	2.1	Handelandsvatn	N	32V	LK	371	6456	58.225	6.804	188	1	1	0	24.7	27
16	20.1	Rosja	N	32V	MM	468	6662	60.092	8.425	1147	0	0	0	8.5	195
17	21.1	Heivann	N	32V	MM	498	6642	59.914	8.964	529	1	1	0	12.3	72
18	3.1	Nedre Malmesvatn	N	32V	LK	373	6489	58.522	6.820	504	1	1	0	25.0	27
19	3.51	Store Gorja	N	32V	NM	596	6661	60.082	10.743	376	1	1	0	25.0	4
20	34.1	Blavatn	N	31V	FH	332	6734	60.705	5.922	837	0	0	0	20.8	50
21	37.1	Reinsennvatnet	N	32V	MN	494	6757	60.947	8.889	923	0	0	0	13.5	100
22	4.1	Hovvatn	N	32V	MK	443	6497	58.608	8.019	493	1	1	0	9.6	120
23	42.1	Gravevatn	N	32V	LP	339	6808	61.390	6.997	600	1	1	0	52.0	130
24	44.2	Fiskeloyse	N	32V	MK	493	6771	61.072	8.870	1037	0	0	0	20.0	110
25	49.1	Langevatn	N	31V	FJ	297	6844	61.674	5.164	70	1	1	0	34.0	70
26	5.1	Sandvatn	N	32V	ML	498	6506	58.693	8.965	150	1	1	0	27.0	30
27	58.1	Movatn	N	32V	LP	352	6876	61.986	6.175	422	1	1	0	27.0	110
28	59.1	Oppljosvatn	N	32V	MP	414	6875	61.996	7.358	1146	0	0	0	60.0	200
29	6	Ovre Malmesvatn	N	32V	LK	373	6490	58.529	6.829	511	1	1	0	15.0	13
30	6.2	Kjosevatn	N	32V	ML	421	6514	58.757	7.635	613	1	1	0	16.0	60
31	60	Oytjern	N	32V	ML	457	6574	59.301	8.245	929	0	0	0	4.8	3
32	65.2	Jutevatn	N	32V	LQ	399	6935	62.531	7.037	525	0	0	0	30.5	70
33	66.1	Ovre Mardalsvatn	N	32V	MQ	452	6933	62.523	8.068	918	0	0	0	54.0	60
34	7	Risvatn	N	32V	MK	457	6479	58.455	8.273	103	1	1	0	26.0	19
35	8.2	Myklevatn	N	32V	ML	407	6548	59.060	7.378	785	1	1	0	33.0	60
36	80.1	Kleivvatn	N	32V	LK	390	6456	58.230	7.127	228	1	1	0	18.5	30
37	81.2	Eilandsvatn	N	32V	MK	441	6478	58.410	7.990	204	1	1	0	9.0	80
38	82.1	Bliksvatn	N	32V	ML	409	6517	58.782	7.426	760	1	1	0	23.5	20
39	83.1	Dorsvatn	N	32V	LL	358	6557	59.127	6.519	850	0	0	0	44.0	50
40	86.1	Tveitvatn	N	32V	NL	510	6591	59.456	9.176	539	1	1	0	17.0	40
41	87.2	Oyusvatn	N	32V	ML	436	6572	59.281	7.877	748	1	1	0	11.2	300
42	88.2	Ovre Brandsvatn	N	32V	ML	419	6596	59.493	7.570	1246	0	0	0	29.2	60
43	89.1	Svinstolvatn	N	32V	LL	366	6591	59.435	6.638	705	0	0	0	47.0	150
44	9.1	Holmvatn	N	32V	ML	480	6555	59.132	8.651	671	1	1	0	27.0	180
45	ACH	Loch na hAchlaise	SCO	NGR	NN	310	480	56.594	-4.753	305	0	0	0	9.2	
46	ARR	Loch Coire nan Arr	SCO	NGR	NG	808	422	57.417	-5.651	130	0	0	0	12.0	
47	ARTH	Loch Arthur	SCO	NGR	NX	904	688	55.002	-3.714	80	1	1	0	13.5	31

Site Number	Site Code	Site Name	Country	Grid	Grid Square	Easting	Northing	Latitude	Longitude	Altitude (m)	Forestry Status				Maximum Depth (m)	Lake Area (ha)
											P	C	D	%		
48	BARE	Loch Barean	SCO	NGR	NX	861	557	54.883	-3.776	40	1	1	0	60	11.0	10
49	BARL	Llyn Barlwyd	CYM	NGR	SH	713	486	53.019	-3.919	558	0	0	0	0	3.0	7
50	BER	Llyn Berwyn	CYM	NGR	SN	743	568	52.195	-3.839	438	1	1	0	60	14.0	13
51	BODG	Llyn Bodgynedd	CYM	NGR	SH	762	593	53.116	-3.850	300	1	1	0	50	17.0	
52	BODL	Llyn Bodlyn	CYM	NGR	SH	648	238	52.794	-4.005	450	0	0	0	0	20.0	16
53	BREC	Lochenbreck	SCO	NGR	NX	643	655	54.966	-4.120	200	1	1	0	70	7.5	16
54	BUGE	Llyn Bugeilyn	CYM	NGR	SN	822	923	52.515	-3.736	457	0	0	0	0	2.3	8
55	BURNMT	Burnmoor Tarn	ENG	NGR	NY	184	44	54.428	-3.258	260	0	0	0	0	13.0	24
56	BYCH	Llyn Bychan	CYM	NGR	SH	753	593	53.116	-3.863	320	1	1	0	50	8.0	2
57	CFYN	Llyn Cwm Ffynnion	CYM	NGR	SH	648	564	53.087	-4.019	380	0	0	0	0	11.0	9
58	CHN	Loch Chon	SCO	NGR	NN	421	51	56.212	-4.546	100	1	1	0	50	25.0	100
59	CLON	Loch Clonyard	SCO	NGR	NX	857	554	54.880	-3.782	34	1	1	0	10	8.5	5
60	CLYD	Llyn Clyd	CYM	NGR	SH	635	597	53.117	-4.040	660	0	0	0	0	6.0	1
61	CON	Llyn Conwy	CYM	NGR	SH	780	463	53.000	-3.818	450	0	0	0	0	22.0	40
62	COR	Loch Coire an Lochan	SCO	NGR	NH	943	4	57.083	-3.744	1000	0	0	0	0	20.0	
63	CWBY	Llyn Cwm Bychan	CYM	NGR	SH	640	313	52.862	-4.020	191	1	0	1	10	14.5	10
64	DEVOKE	Devoke Water	ENG	NGR	SD	163	970	54.361	-3.288	240	0	0	0	0	14.0	
65	DIWA	Llyn Diawaunedd	CYM	NGR	SH	685	536	53.063	-3.963	375	1	1	0	10	19.5	
66	DOI	Loch Doilet	SCO	NGR	NM	808	678	56.750	-5.586	10	1	1	0	41	16.8	53
67	DOON	Loch Doon	SCO	NGR	NX	495	985	55.258	-4.368	215	1	1	0	40	30.5	
68	DUH	Dubh Loch	SCO	NGR	NO	238	828	56.930	-3.252	700	0	0	0	0	21.0	20
69	DUL	Llyn Dulyd	CYM	NGR	SH	662	244	52.800	-3.985	520	0	0	0	0	6.8	2
70	ENO	Loch Enoch	SCO	NGR	NX	445	851	55.136	-4.440	490	0	0	0	0	36.0	50
71	EUN	Loch nan Eun	SCO	NGR	NO	230	854	56.954	-3.266	900	0	0	0	0	23.0	
72	FERN	Loch Fern	SCO	NGR	NX	863	624	54.943	-3.775	85	1	1	0	20	2.5	5
73	FHI	Coire Fhionn Lochan	SCO	NGR	NR	902	459	55.661	-5.336	340	0	0	0	0	9.5	7
74	FINL	Loch Finlas	SCO	NGR	NX	460	983	55.255	-4.423	255	1	1	0	20	13.5	78
75	FLE	Loch Fleet	SCO	NGR	NX	560	697	55.001	-4.252	340	1	1	0	20	16.5	17
76	GARN	Llyn y Garn	CYM	NGR	SH	762	377	52.922	-3.842	510	0	0	0	0	18.5	
77	GEIR	Llyn Geirionydd	CYM	NGR	SH	763	606	53.128	-3.849	228	1	1	0	50	14.5	
78	GLAS	Llyn Glas	CYM	NGR	SH	601	547	53.071	-4.088	520	0	0	0	0	6.5	1
79	GLFR	Llyn Glasfryn	CYM	NGR	SH	402	422	52.953	-4.379	120	0	0	0	0	1.0	
80	GLYN	Llyn Glaslyn	CYM	NGR	SN	826	941	52.532	-3.731	490	0	0	0	0	11.0	23
81	GOD	Llyn Goddionduon	CYM	NGR	SH	754	585	53.109	-3.862	292	1	1	0	100	7.0	6
82	GREENT	Greendale Tarn	ENG	NGR	NY	146	74	54.455	-3.317	400	0	0	0	0	9.0	2
83	GULSPET	Gulspettvatn	N	32V	NL	505	6503	58.668	9.090	56	1	1	0	0	25.0	32
84	GWYN	Llyn Gwynant	CYM	NGR	SH	644	519	53.056	-4.023	70	1	0	1	10	17.0	
85	GYN	Llyn Gynon	CYM	NGR	SN	800	647	52.267	-3.759	420	0	0	0	0	11.0	25
86	HARR	Loch Harrow	SCO	NGR	NX	527	867	55.153	-4.312	250	1	1	0	100	9.0	16
87	HIR	Llyn Hir	CYM	NGR	SN	789	675	52.292	-3.776	435	0	0	0	0	8.8	5
88	HOLET	Holetjorn	N	32V	MK	371	6482	58.479	6.813	485	1	1	0	0	16.0	2
89	HOLMEV	Holmevatn	N	32V	ML	463	6547	59.061	8.355	583	1	1	0	0	28.0	100
90	HOWI	Loch Howie	SCO	NGR	NX	697	834	55.128	-4.044	230	1	1	0	100	13.0	18
91	INVA	Lochinvar	SCO	NGR	NX	659	853	55.144	-4.104	220	1	1	0	5	5.5	33
92	IRD	Llyn Irddyn	CYM	NGR	SH	630	220	52.778	-4.031	300	0	0	0	0	8.8	9
93	KIRR	Loch Kierriereoch	SCO	NGR	NX	363	865	55.146	-4.569	213	1	1	0	75	3.5	8
94	LAG	Llyn Llagi	CYM	NGR	SH	649	483	53.015	-4.014	380	0	0	0	0	16.5	6
95	LAI	Loch Laidon	SCO	NGR	NN	380	542	56.652	-4.643	280	0	0	0	0	39.0	473
96	LAR	Loch na Larach	SCO	NGR	NC	217	583	58.477	-5.058	61	0	0	0	0	8.5	
97	LCSL	Llyn Cwm Silyn Lower	CYM	NGR	SH	512	508	53.033	-4.219	340	0	0	0	0	17.0	4
98	LCSU	Llyn Cwm Silyn Upper	CYM	NGR	SH	515	505	53.031	-4.215	342	0	0	0	0	17.0	5
99	LDE	Loch Dee	SCO	NGR	NX	470	790	55.082	-4.397	230	1	1	0	20	14.5	100
100	LENY	Llyn Llennych	CYM	NGR	SH	655	377	52.920	-4.001	255	0	0	0	0	10.0	3
101	LGR	Loch Grannoch	SCO	NGR	NX	541	691	54.995	-4.281	210	1	1	0	70	20.5	114
102	LJOSV	Ljosvatn	N	32V	MK	366	6476	58.425	6.710	385	1	1	0	0	21.0	11
103	LLDU	Llyn Du	CYM	NGR	SH	564	425	52.960	-4.138	255	0	0	0	0	7.0	2
104	LLGH	Long Loch of Glenhead	SCO	NGR	NX	446	808	55.097	-4.436	300	0	0	0	0	11.5	9

Site Number	Site Code	Site Name	Country	Grid	Grid Square	Easting	Northing	Latitude	Longitude	Altitude (m)	Forestry Status				Maximum Depth (m)	Lake Area (ha)	
											P	C	D	%			
105	LOD	Lochan Dubh	SCO	NGR	NM	895	710	56.783	-5.446	230	0	0	0	0	9.0	9	
106	LOWT	Low Tarn	ENG	NGR	NY	163	91	54.470	-3.292	480	0	0	0	0	3.0	4	
107	MABE	Loch Mayberry	SCO	NGR	NX	286	750	55.040	-4.683	118	1	1	0	40	4.5	70	
108	MACA	Loch Macaterick	SCO	NGR	NX	440	912	55.190	-4.451	290	1	1	0	50	12.5	55	
109	MANN	Loch Mannoeh	SCO	NGR	NX	664	605	54.921	-4.085	128	0	0	0	0	7.0	24	
110	MINN	Loch Minnoch	SCO	NGR	NX	530	857	55.144	-4.307	275	1	1	0	100	7.0	6	
111	MOAN	Loch Moan	SCO	NGR	NX	346	858	55.139	-4.595	200	1	1	0	100	3.0	49	
112	MUCK	Loch Muck	SCO	NGR	NS	513	7	55.278	-4.341	290	1	1	0	5	8.0	11	
113	NAGA	Lochnagar	SCO	NGR	NO	252	859	56.958	-3.230	790	0	0	0	0	24.0	10	
114	OCHI	Loch Ochiltree	SCO	NGR	NX	317	745	55.036	-4.634	107	1	1	0	10	8.0	62	
115	PARC	Llyn y Parc	CYM	NGR	SH	793	587	53.112	-3.803	246	1	1	0	100	11.0		
116	PENR	Llyn Penrhaiadr	CYM	NGR	SN	753	933	52.523	-3.838	410	0	0	0	0	7.0	5	
117	RIEC	Loch Riecawr	SCO	NGR	NX	434	934	55.210	-4.461	280	1	1	0	100	9.0	92	
118	RLGH	Round Loch of Glenhead	SCO	NGR	NX	450	805	55.095	-4.429	300	0	0	0	0	13.5	13	
119	RONA	Loch Ronald	SCO	NGR	NX	265	644	54.944	-4.709	102	1	1	0	40	17.5	37	
120	R\O	Royrtjorna	N		33W	UM	783	646	64.810	12.380	163	1	1	0	14.0	26	
121	S1	Skaren	S					59.370	18.230	49	1	1	0		5.5	29	
122	S10	Hundsjon	S					61.380	15.100	276	1	1	0		15.0	33	
123	S11	Algarydssjon	S					57.080	14.450	201	1	1	0		6.5	33	
124	S12	Ljusacksen	S					60.540	15.250	295	1	1	0		17.0	54	
125	S13	Storasjo	S					56.570	15.170	252	1	1	0		12.0	40	
126	S14	St. Bjorken	S					61.450	15.590	221	1	1	0		11.0	61	
127	S15	V. Helgtjarnen	S					63.100	13.100	648	1	1	0		5.0	29	
128	S16	Botungen	S					59.220	12.430	99	1	1	0		7.5	80	
129	S17	Skaravattnet	S					56.230	15.170	99	1	1	0		10.0	14	
130	S18	Orsjon	S					56.170	14.410	87	1	1	0		10.0	18	
131	S19	Rotehogstjarn	S					58.490	11.370	121	1	1	0		8.5	17	
132	S2	Edasjon	S					59.480	17.540	18	1	1	0		4.5	17	
133	S20	Lill Jangen	S					60.000	13.210	193	1	1	0		9.0	90	
134	S21	Gaffeln	S														
135	S22	Harsvatten	S					58.001	12.030	129					24.0	19	
136	S24	Ulvsjon	S					59.370	12.170	200	1	1	0		23.5	51	
137	S25	Flatsjon	S					59.000	12.240	126	1	1	0		15.0	50	
138	S26	Hagasjon	S					57.320	13.250	177	1	1	0		4.0	80	
139	S27	Vagsjoarna	S					59.480	13.340	187	1	1	0		14.0	59	
140	S28	Stor En	S					59.550	12.250	263	1	1	0		9.0	173	
141	S29	Orvattnet	S					59.440	12.450	276	1	1	0		13.0	72	
142	S3	Limmingsjon	S					59.350	14.300	234	1	1	0		17.0	114	
143	S30	Amten	S					59.570	12.350	273	1	1	0		15.0	47	
144	S31	Skardalsvattnet	S					59.210	11.570	105	1	1	0		11.0	64	
145	S4	Djupa Holmsjon	S					59.110	17.020	60	1	1	0		23.0	20	
146	S5	Hedsjon	S					58.590	17.170	33	1	1	0		6.5	11	
147	S6	Olsjon	S					59.390	15.070	173	1	1	0		7.0	26	
148	S7	Raksjon	S					60.030	17.040	65	1	1	0		9.5	115	
149	S8	Flickersjon	S					61.480	12.250	645	1	1	0		16.5	17	
150	S9	Hagserydssjon	S					57.180	15.260	180	1	1	0		9.5	41	
151	SCOATT	Scoat Tarn	ENG	NGR	NY	159	104	54.482	-3.298	602	0	0	0	0	20.0	5	
152	SKAK	Loch Skae	SCO	NGR	NX	710	837	55.130	-4.020	263	1	1	0	100	13.0		
153	SKE	Loch Skerrow	SCO	NGR	NX	605	682	54.989	-4.181	120	1	1	0	55	10.0	51	
154	SKOMAKV	Skomakarvatn	N		31V	FG	301	6699	60.377	5.399	564	0	0	0	0	3.0	
155	STRO	Loch Stroan	SCO	NGR	NX	644	704	55.010	-4.121	70	1	1	0	70	12.5	23	
156	TANN	Loch Tanna	SCO	NGR	NR	921	428	55.634	-5.303	330	0	0	0	0	3.5	33	
157	TEAN	Loch Teanga	SCO	NGR	NF	818	383	57.324	-7.288	20	0	0	0	0	21.0		
158	TECW	Llyn Tecwyn	CYM	NGR	SH	629	370	52.913	-4.039	100	1	0	1	5	7.0		
159	TINK	Loch Tinker	SCO	NGR	NN	445	68	56.228	-4.509	420	0	0	0	0	9.8	11	
160	TROO	Loch Trool	SCO	NGR	NX	412	798	55.087	-4.488	75	1	1	0	60	17.0	60	
161	UAI	Lochan Uaine	SCO	NGR	NO	1	981	57.063	-3.648	950	0	0	0	0	20.0	4	

Site Number	Site Code	Site Name	Country	Grid	Grid Square	Easting	Northing	Latitude	Longitude	Altitude (m)	Forestry Status				Maximum Depth (m)	Lake Area (ha)
											P	C	D	%		
162	UIS	Loch Uisge	SCO	NGR	NM	808	550	56.636	-5.575	152	0	0	0	0	11.0	18
163	URR	Loch Urr	SCO	NGR	NX	760	845	55.139	-3.946	198	1	1	0	5	13.2	44
164	VAL	Loch Valley	SCO	NGR	NX	445	817	55.105	-4.438	340	0	0	0	0	16.5	35
165	VEREV	Verevatn	N	32V	MK	453	6472	58.383	-8.200	268	1	1	0	0	13.9	9
166	WHIN	Loch Whinyeon	SCO	NGR	NX	625	608	54.923	-4.146	220	1	1	0	30	13.0	41
167	WHIT	White Loch	SCO	NGR	NX	864	547	54.874	-3.771	30	1	1	0	60	13.5	11
168	WOOD	Loch Woodhall	SCO	NGR	NX	673	675	54.984	-4.074	50	1	1	0	50	16.5	67
169	YBI	Llyn y Bi	CYM	NGR	SH	670	265	52.819	-3.974	445	0	0	0	0	3.0	3
170	YGAD	Llyn y Gadair	CYM	NGR	SH	648	564	53.087	-4.019	380	0	0	0	0	3.0	

- iv) Norway NIVA, Oslo
 Botanical Institute, University of Bergen (J.F. Boyle)
- v) Sweden Environmental Protection Agency, Freshwater Section, Uppsala

Up to twenty chemical determinands were measured on each water sample using standard procedures (Golterman 1969, APHA 1980, Ahl 1972). pH and conductivity were measured after the sample had equilibrated to room temperature (20°C) and ambient CO₂ conditions. Ca, Na, Mg, K, Mn, Zn, and Fe were determined, after filtration, by atomic absorption spectrophotometry. SO₄, NO₃, Cl, and total organic nitrogen (TON) were determined by ion chromatography, and SiO₂ was measured colorimetrically. Dissolved organic carbon (DOC) was determined as absorbance at 254nm by spectrophotometry and estimated using the linear calibration of Vik (1982). In most cases only total aluminium was measured. However, for a number of samples labile aluminium concentrations were determined after separation on an ion exchange column packed with DOWEX 50w Na type resin (Driscoll 1984). Total and non-labile monomeric aluminium were measured using the pyrocatechol violet method. However, in this paper the term total aluminium is used to refer to total aluminium *s.s. plus* total monomeric aluminium. All measurements are given as follows: Na, Ca, K, Mg, alkalinity, NO₃, Cl, SO₄ are in µeq l⁻¹; conductivity in µs cm⁻¹, Zn, Fe, total Al, Mn, TON, SiO₂ in µg l⁻¹; DOC in mg l⁻¹.

Before entry to the SWAP diatom-environment database the results of the chemical determinations were screened to identify spurious or discordant values that may have arisen either as a result of analytical error or from unusual meteorological conditions, such as sea-salt events. In addition, because of logistical problems within laboratories not all analyses were performed on all samples (Table 3).

Table 3 Summary of the number of diatom samples with associated environmental determinations for the 178 sample data-set

Determinand	Number of samples	Determinand	Number of samples
pH	178	Conductivity	177
Dissolved Organic Carbon	148	Ca	175
Mg	174	Na	144
K	174	SO ₄	175
Cl	175	NO ₃	126
Total Organic Nitrogen	117	Alkalinity	161
Total Al*	173	Zn	108
Mn	136	Fe	139
SiO ₂	137	Altitude	176
Maximum Water Depth	176	Lake Area	128

* some of these determinations refer to Total Al, others to the monomeric fraction.

DIATOM TAXONOMY AND NOMENCLATURE

Introduction

Despite the general availability of major diatom floras (e.g. Hustedt 1930-1966, Cleve-Euler 1951-1955, Patrick and Reimer 1966, 1975), there can be considerable variation in taxonomic and nomenclatural usage between diatomists in different laboratories. Floras differ in their definitions and use of names, diatomists develop laboratory-based concepts for inter- and intra-specific taxonomic division, and errors can easily be made. Moreover, reference to "type material" is often not practical or, in some cases, not possible. Such differences can hinder the comparison of data between laboratories and, in situations where diatoms are used as environmental indicators, misleading conclusions can arise.

In palaeolimnological projects where diatomists from several laboratories are involved, it is essential to establish agreed protocols for diatom taxonomy and nomenclature. Diatomists involved in the Palaeoecological Investigation of Recent Lake Acidification (PIRLA) project in the USA (Charles and Whitehead 1986) have developed a rigorous system of taxonomic control (Kingston 1986) including a diatom iconograph series (PIRLA 1984-1986), regular workshops, and the circulation of diatom slides between laboratories. In SWAP diatomists from Norway, Sweden, and the UK adopted a similar approach to taxonomic harmonization through regular workshops, diatom slide exchange, and the circulation of agreed taxonomic protocols.

Slide exchange and analytical quality control

Initial identification of problems

The first SWAP diatom taxonomy workshop was held in March 1987. Prior to this, each of the four laboratories involved circulated diatom slides and accompanying count sheets for three sediment samples, to each of the other laboratories. The slides were chosen to represent the range of soft-water floras encountered within SWAP. All 12 slides were then recounted by the other laboratories and a comparison of the results made at the taxonomy workshops. Figure 2a summarizes the results of one of the 12 slides, Lingmoor Tarn. Figure 2b shows the same results after revision of the taxonomy and nomenclature at the 1987 workshop. This example demonstrates the three main problems that arose, namely:

a) *Nomenclature*

In this analysis there was a problem within the genus *Anomoeoneis* with both former and revised names being allocated to the same taxon. *A. vitrea* (Grun.) Ross was used by laboratory I and *A. exilis* Cleve, including var. *lanceolata*, was used by laboratories II, III, and IV for the same diatom. Likewise *Anomoeoneis brachysira* (Bréb.) Grun. and *A. seriens* var. *brachysira* (Bréb) Cleve were both applied to the same taxon. These two taxa have now been transferred to the genus *Brachysira*, *A. vitrea* becoming *B. vitrea* (Grun.) Ross and *A. brachysira* becoming *B. brebissonii* Ross.

b) *Splitting and amalgamation of taxa*

Anomoeoneis exilis var. *lanceolata* was split from the nominate by one laboratory. This has also been transferred to the genus *Brachysira* and after some discussion it was agreed to amalgamate the variety with the nominate taxon within SWAP.

c) *Identification*

The criteria used to separate *Eunotia alpina* (Naeg.) Hust. from *E. lunaris* (Ehr.) Grun. differed between laboratories. A mutually agreed list of criteria was drawn up to differentiate the two taxa. The revised names for *E. alpina* and *E. lunaris* are *E. naegelii* Migula and *E. curvata* (Kütz.) Lagerst., respectively.

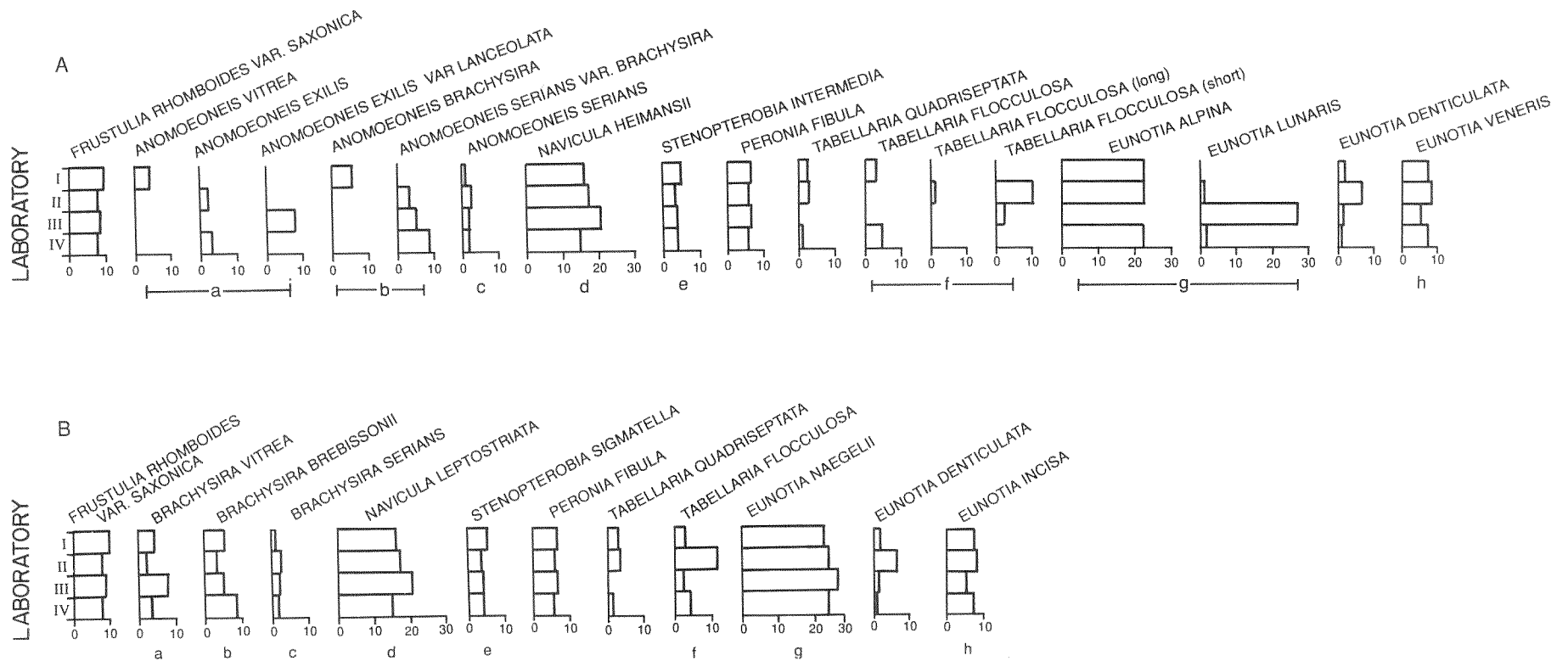


Figure 2 a (upper) Dominant taxa in the Lingmoor Tarn slide illustrating (i) problems of nomenclature in groups a,b,c,d,e,g, and h, (ii) problems of splitting versus amalgamation in groups a and f, and (iii) the use of differing identification criteria in group g. Horizontal scale is percentage occurrence.

b (lower) Dominant taxa in the Lingmoor Tarn slide after taxonomic and nomenclatural revision. Horizontal scale is percentage occurrence.

The agreements resulting from the discussion of these and other taxa were circulated to all the workshop participants in a SWAP taxonomic guide (Appendix A). Decisions on taxonomy and nomenclature were based on the *Checklist of British Diatoms* (Hartley 1986) which formed the framework for *A Coded Checklist of British Diatoms* (Williams *et al.* 1988). The SWAP taxonomic guide also included agreements on definitions for boundaries between certain species and their varieties. If possible, published descriptions were referred to or failing this, the criteria for identification were agreed between the analysts.

Applying and refining the SWAP taxonomic guide

The next stage in the harmonization of diatom taxonomy was to put the protocols agreed at the 1987 workshop to the test. Three slides from lake-sediment samples representing the range of pH values encountered within SWAP (one sample from each of the ranges pH < 5.0, pH 5.0 - 6.0, and pH > 6.0) were circulated to all SWAP diatomists. This time, count sheets were not included so the slides were counted without prior knowledge of the diatom assemblages. The results were discussed at a diatom taxonomy workshop in July 1988.

Many potential problems had been avoided by following the protocols agreed at the previous workshop. However, due to the inclusion of taxa additional to those already encountered, some further problems were raised concerning nomenclature, splitting/amalgamation, and identification. Figure 3 shows the predominant taxa from one slide, Botungen. In this slide the use of both *Cyclotella comta* and *C. radiosa* was identified as a problem in nomenclature and *C. radiosa* Håkansson was adopted then as the valid name for this taxon. Also, the four forms of *Tabellaria flocculosa* had been split by each diatomist using different criteria. In this case, no common set of criteria could be agreed upon for consistently separating the forms so it was agreed to amalgamate the longer forms of this species into a *T. flocculosa* aggregate category within the SWAP data-set. Similarly, it was agreed to combine *Achnanthes minutissima* and *A. microcephala* for the purposes of data analysis within SWAP, although with these and all other combinations of taxa made, it was agreed that diatomists could retain the original distinctions in their own data-sets, if they wished. In addition this workshop produced a guide for the classification of unknown diatoms.

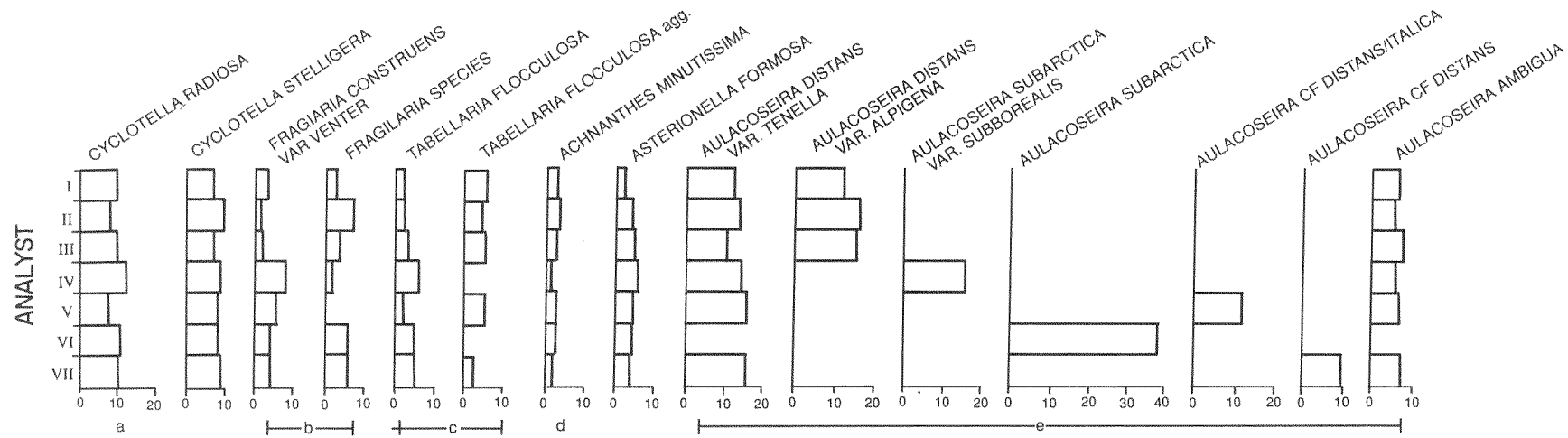
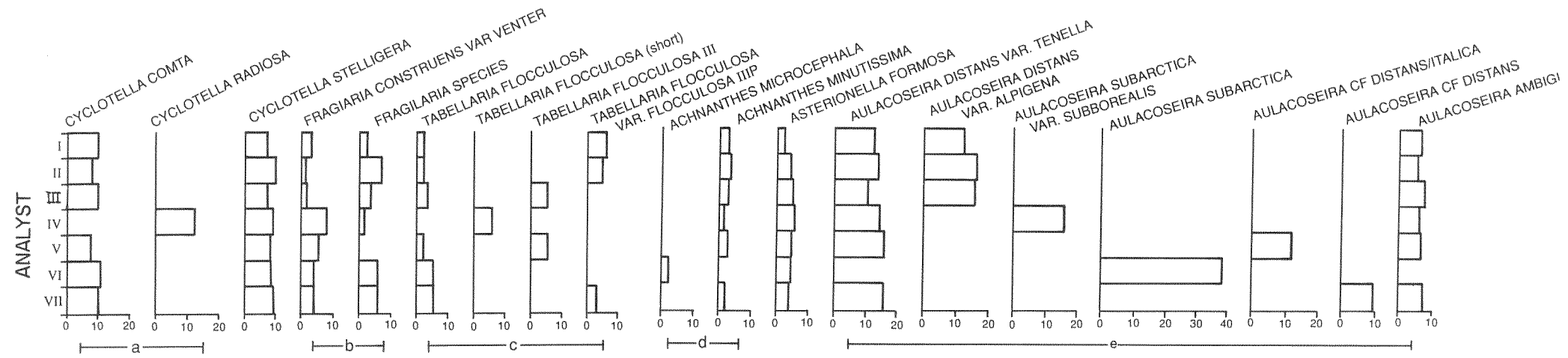


Figure 3 a (upper) Dominant taxa in the Botungen slide illustrating similar problems of nomenclature in groups a,d, and e, problems of splitting versus amalgamation in groups c and e, and the use of differing identification criteria in the same groups. Horizontal scale is percentage occurrence.

b (lower) Dominant taxa in the Botungen slide after taxonomic and nomenclatural revision in 1988. Horizontal scale is percentage occurrence.

Protocols for handling unknowns

There are two main groups of unknown diatoms:

- 1) those that occur infrequently and cannot be allocated a name due to damage or deformation or are too rare to justify detailed taxonomic work, and
- 2) those that occur frequently enough to be identified as a discrete taxon but no published description of the diatom can be found.

Within SWAP it was agreed to name unknown diatoms in group 1 as follows:

i) Diatoms that cannot be identified to genus level due to breakage, being in girdle view or being partly obscured on the slide should be called "unknown pennate", "unknown centric", or simply "unknown".

ii) Diatoms that cannot be identified to species level due to the reasons listed above should be allocated a genus if possible. For example, *Achnanthes* spp., *Pinnularia* spp., "unknown Naviculaceae", or "unknown *Achnanthes/Navicula*".

iii) Unknown diatoms occurring infrequently which do not resemble a particular species but can be allocated to a genus should also be classified as in ii) above.

iv) Diatoms which do not exactly fit a description for a taxon but have some features in common with that taxon and that do not occur often enough to warrant classifying them separately from the taxon they resemble, should be put in a "*cf.*" category. For example, in a count of 500 valves there may be several valves of a diatom which closely resembles *Achnanthes marginulata* but which do not fit the published descriptions of the species. In this case, if the diatoms have not been observed elsewhere it may be convenient to put these valves into *Achnanthes cf. marginulata* rather than list them individually as in group 2.

In group 2 unknown diatoms that occur with sufficient regularity or in sufficient numbers to be identifiable as a discrete taxon should be named using []. The square brackets can contain anything that readily identifies that diatom to the diatomist who finds it. The name of the "type" site for the diatom and the initials of the diatomist making the identification should appear as the "authority" after the name. Some examples are given below:

Unknown [sp.1 *Stauroneis/Navicula*] Loch Doilet AK

In this case the diatom could not be placed in either genus so the two possible genera are included within the brackets. The diatom was first noted at Loch Doilet by Annette Kreiser (AK).

Achnanthes [sp.2 *nodosa*] Loch Chon AK

Here the diatom is clearly an *Achnanthes*. The information in the brackets shows that it is the second unknown *Achnanthes* named at this site and that *A. nodosa* is the closest published description to it.

Aulacoseira [sp.1 PIRLA *distans* var. 5] Llyn Tecwyn RJF

This name indicates that the diatomist (R.J. Flower) has found a similar unknown (*Aulacoseira distans* var. 5) in the PIRLA (1984-1986) diatom iconograph.

Pinnularia [sp.2 small, capitata] Llyn Berwyn AK

In this case the unknown does not resemble any other known or published taxon so other memory-joggers are included in the name.

Information contained within the square brackets only has significance for the diatomist who put it there, unless a full description has been circulated by the diatomist. This is particularly important in the case of a name containing a reference to a published description or illustration. The reference given does not necessarily accurately describe the diatom in question. This allows a diatomist to call an unknown *Eunotia*, *Eunotia* [PIRLA sp.13] because it exactly matches PIRLA *Eunotia* sp.13 (which is not an officially published name so has to remain in square brackets) but also enables another diatomist to use the same name for a diatom from a different site which only resembles PIRLA *Eunotia* sp.13 in certain respects. Clearly in cases like this the two occurrences should be listed separately with the different site names and the diatomist's initials, indicating that they may be different taxa, unless circulated descriptions confirm that they are the same taxon. Descriptions of important unknowns were circulated to other diatomists to help prevent the use of different names for the same taxon. A list of the final taxonomic decisions made for the SWAP training set is given in Appendix A.

As a result of the 1988 workshop, a revised edition of the SWAP taxonomic guide was circulated to all SWAP diatomists. It was also decided to distribute a reference set of the surface-sediment (training data-set) slides to each SWAP laboratory so that particular unknown or problematic taxa could be readily examined by diatomists with reference to their own material.

Problems within the genus Aulacoseira

A major problem encountered at the 1988 workshop was one of identification within the genus *Aulacoseira* (Figure 3b). This genus presents particular taxonomic problems due to the difficulties of matching side (girdle) views of the diatom valves with front (valve) views. Difficulties with identification of the rarer and smaller taxa within the genus, such as those within the *A. subarctica* / *subborealis* group remained unresolved.

As a result of the 1988 workshop it was decided to concentrate on the taxonomy of the *Aulacoseira* genus at the next workshop. Three samples containing *Aulacoseira* were selected from SWAP sites and slides were circulated to all SWAP diatomists. Three hundred valves of *Aulacoseira* were counted from each slide and the results discussed at the diatom taxonomy workshop in February 1989. There were no problems with the most abundant *Aulacoseira* taxa, such as *A. lirata* and *A. distans* var. *nivalis*. For *A. lirata* var. *alpigena* a description was agreed (see Appendix A). Work is continuing on the *A. subarctica* / *A. subborealis* group. For the purposes of SWAP, it was decided to amalgamate these taxa into a *A. [subarctica* agg.] category. Although the analysis of surface and core samples had been completed by this stage, it was agreed to recount samples containing the problematic *Aulacoseira* taxa since considerable revision of their taxonomy had occurred as a result of the 1989 workshop.

This series of workshops and quality-control exercises enabled all the diatom data generated within SWAP to be compatible between laboratories. Nomenclature follows Williams *et al.* (1988) but names for taxa or amalgamated groups of taxa found only within SWAP are also used. This standardization of the diatom data was essential before the data could be stored and manipulated in the computerized data-base.

DIATOM DATABASE

The diatom database (DISCO) at University College London (Munro *et al.* 1990) combines archives of diatom counts, chemical analyses, and catchment descriptions for SWAP sites. It uses the commercial program ORACLE and the standard database language SQL. The database includes a diatom checklist which allocates a code to each taxon on the list using an alphanumeric code to represent the original name of the taxon.

Diatomists submitting counts to the database provided an outline code dictionary for their data. The dictionary was checked to ensure that it contained valid checklist codes, and was compared to previous

lists from the same laboratory. The counts were then converted to percentages of the total number of diatoms counted in each sample; any amalgamations and other re-definitions of taxa were performed. The database also provided a list of the more frequently occurring taxa for each group of samples, defined as those present in at least two of the samples and having a frequency of more than 1% in at least one sample. These lists were compared to help identify unresolved taxonomic problems.

The SWAP training data-set was created by merging full sets of surface-sediment percentages in this way for the regional data-sets from Scotland, Wales, English Lake District, Norway, and Sweden. All the aggregates to genus level and above (e.g. *Navicula* spp.) were deleted and the remainder of the list was used to select the taxa to be exported to the data-analytical and calibration programs.

NUMERICAL ANALYSIS OF THE CHEMICAL AND OTHER ENVIRONMENTAL DATA

The training data-set consists of 170 lakes with diatom, chemical, and other environmental data. Of these, eight have a replicate set of diatom and chemical variables (collected in 1987 from Round Loch of Glenhead, Loch Howie, Loch Grannoch, Loch Skerrow, Loch Dee, Loch Skae, Loch Woodall, Loch Whinyeon; Stevenson *et al.* 1989) to give a basic 178 sample data-set. These replicates differ in collection date by 3 - 4 years. Figure 1 illustrates the geographical coverage of the 170 SWAP training lakes.

Overall, the samples are derived from lakes with varying geologies giving varying sensitivities to acidification and from a range of different pollution settings. Land-use within the modern data-set varies from upland farming with some improvement to unimproved moorland and forest (both natural and planted). In general, the majority of the UK sites are derived from unforested catchments while the Scandinavian sites are predominantly forested.

Basic statistics

The available mean chemical data for all 178 samples with associated diatom counts are tabulated along with the available basic catchment vegetation and site physical variables in Table 4. Frequency histograms of the distribution of selected chemical and site physical variables are presented in Figure 4. Summary statistics for selected variables are given in Table 5 as mean, median, standard deviation, minimum, maximum and lower and upper quartiles. Basic scatter plots for the major chemical determinands are shown in Figures 5 and 6.

Table 5 Summary statistics for selected environmental variables in the SWAP training set (Stan. dev. = standard deviation, min. = minimum, max. = maximum, quar. = quartile, N = number of determinations, afforest = afforestation)

	N	Mean	Median	Stan. dev.	Min.	Max.	Lower Quar.	Upper Quar.
pH	178	5.59	5.40	0.77	4.33	7.25	4.96	6.25
Conductivity	177	44.5	37.2	27.6	7.4	154.0	26.5	57.0
DOC	148	3.9	2.8	3.5	0.1	17.0	1.4	5.5
Ca	175	109.5	66.4	120.2	6.2	938.0	37.5	134.6
Mg	174	67.5	51.1	52.8	7.8	337.3	33.0	85.7
Na	144	171.8	156.7	138.0	16.5	1004.0	88.7	204.4
K	174	9.8	8.0	7.4	2.0	54.1	5.5	11.8
SO ₄	175	106.9	90.0	64.4	18.7	352.4	60.8	145.8
Cl	175	187.4	164.2	161.6	10.4	1160.0	79.0	233.0
NO ₃	126	10.4	6.2	13.3	0.0	91.5	3.1	12.4
TON	117	271.7	245.0	183.4	10.0	982.0	130.0	380.5
Alkalinity	161	57.3	12.8	116.5	0.0	903.0	0.0	58.8
Total Al	173	130.9	112.0	106.2	12.0	802.0	54.2	175.9
Zn	109	26.2	11.0	96.3	2.0	730.0	7.3	17.3
Mn	136	55.0	30.0	78.9	1.5	601.5	12.8	66.8
Fe	139	143.4	70.6	175.7	3.5	837.5	33.0	199.0
SiO ₂	137	1329.9	1128.0	991.7	179.0	7000.0	550.0	1725.0
Altitude (m)	176	349.2	270.5	267.4	10.0	1246.0	165.3	488.7
% afforest	112	22.3	0.0	33.0	0.0	100.0	0.0	50.0
Max Depth (m)	176	15.9	13.5	10.8	1.0	61.0	9.0	20.0

Figure 4 shows that the majority of the sites are acidic, oligotrophic, and clearwater with low modal pH (5.0 - 5.4), conductivity (25 - 35 $\mu\text{S cm}^{-1}$), calcium (< 50 $\mu\text{eq l}^{-1}$), and alkalinity (< 50 $\mu\text{eq l}^{-1}$) classes. As might be expected from this distribution the majority of the lakes are slightly coloured with a modal DOC range of 1 - 3 mg l^{-1} . Strong colour is restricted to a few sites, especially in northern Scotland and a large number of sites from Sweden (e.g. Hagasjön, Vågsjöarna). Throughout the acid end of the data-set, aluminium levels are high (modal class 50 - 150 $\mu\text{g l}^{-1}$ total Al).

Sodium and chloride are in approximate marine proportions (data-set Na/Cl = 0.92 ± 0.21 ; marine ratio = 0.85) throughout the data-set (Figure 6) indicating a contribution predominantly from rainfall. A few lakes from coastal locations such as Loch na'Larach and Loch Teanga in Scotland have very high conductivities, sodium, and chloride levels. A number of lakes subject to past and current mining activities in their catchments have high trace-metal concentrations, e.g. Llyn Geirionydd and Llyn y Parc (Zn >700 $\mu\text{g l}^{-1}$).

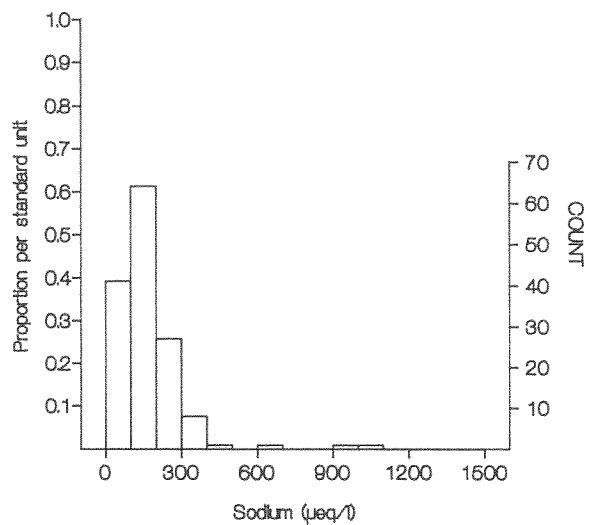
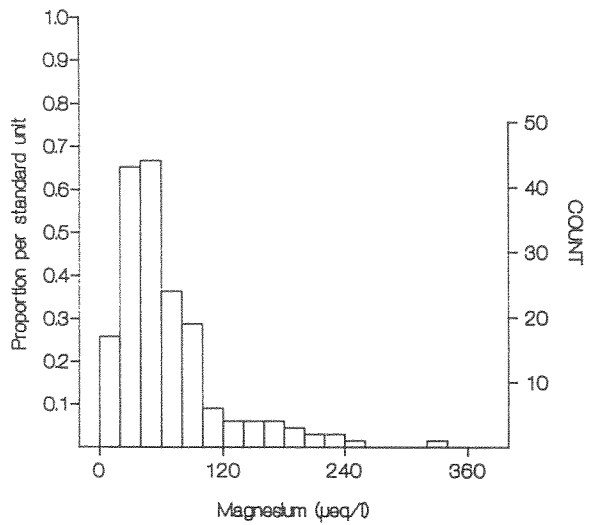
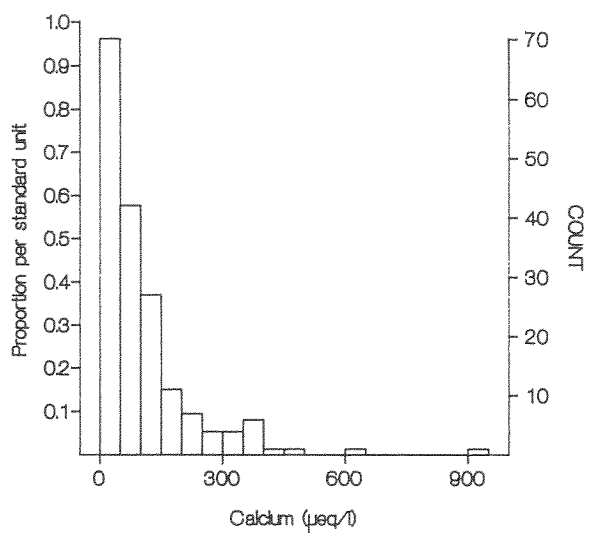
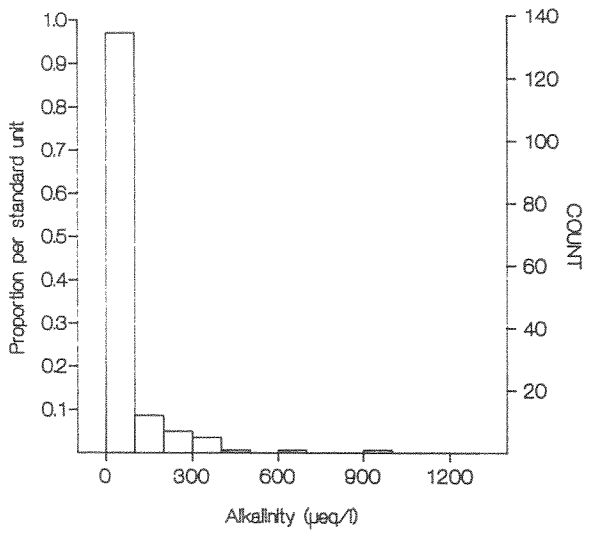
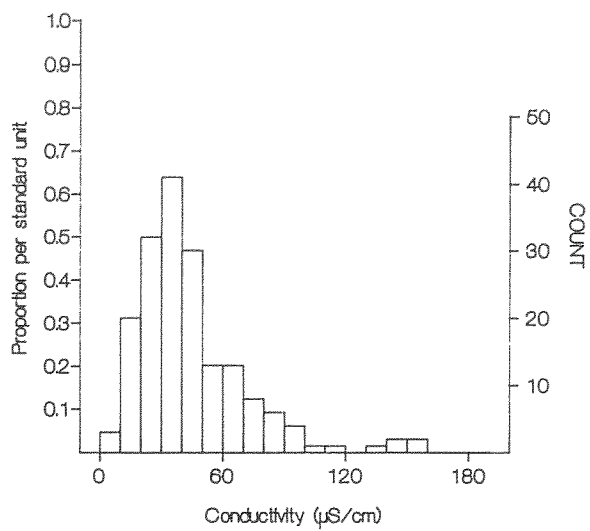
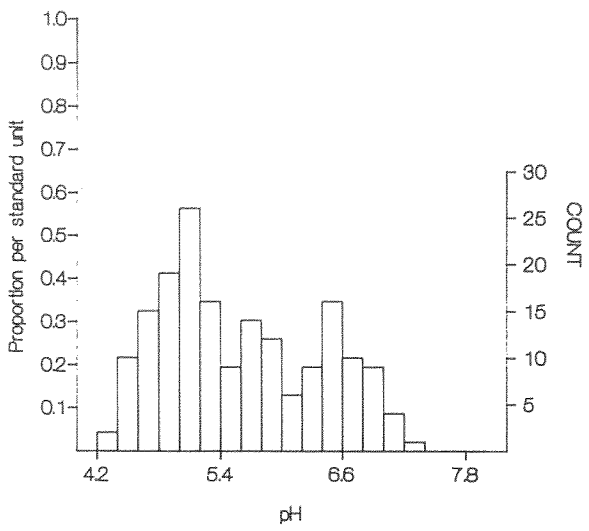


Figure 4 Frequency histograms for major chemical determinands and catchment variables, showing the proportion of different classes for the major variables.

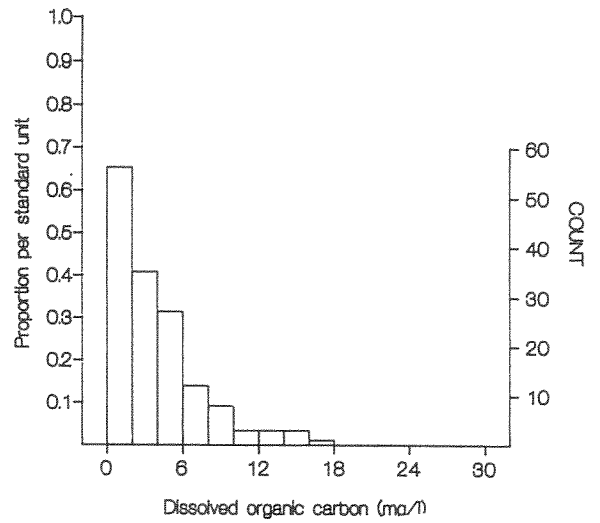
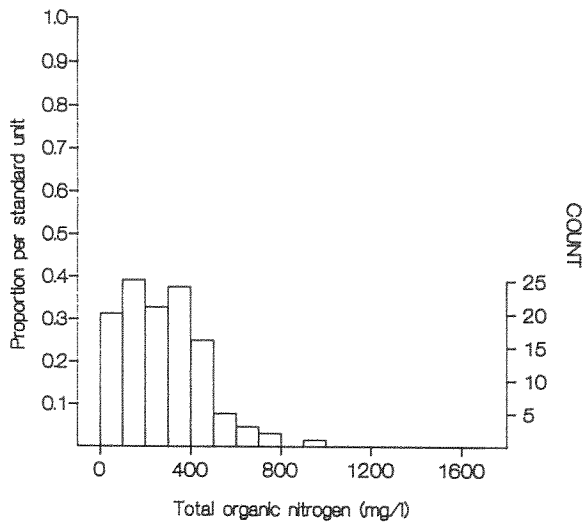
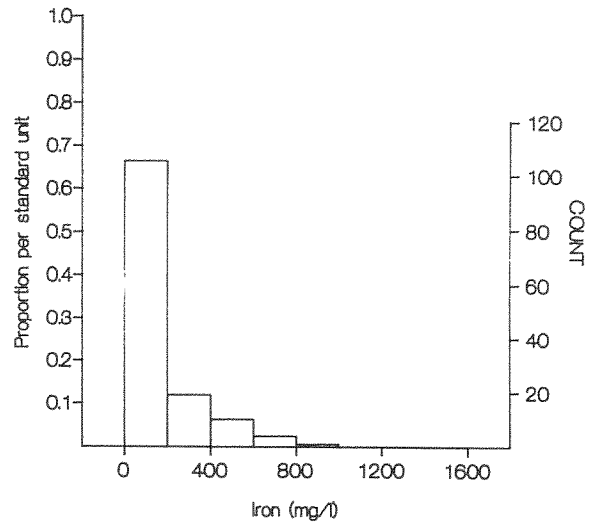
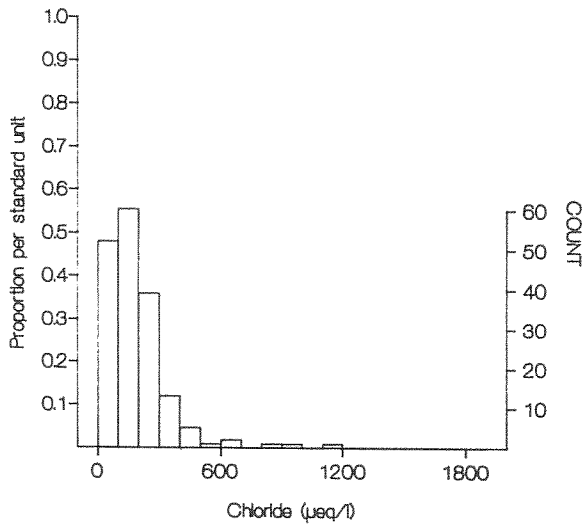
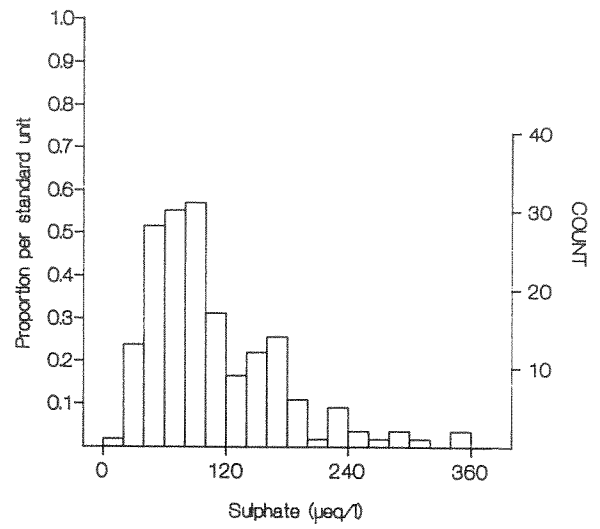
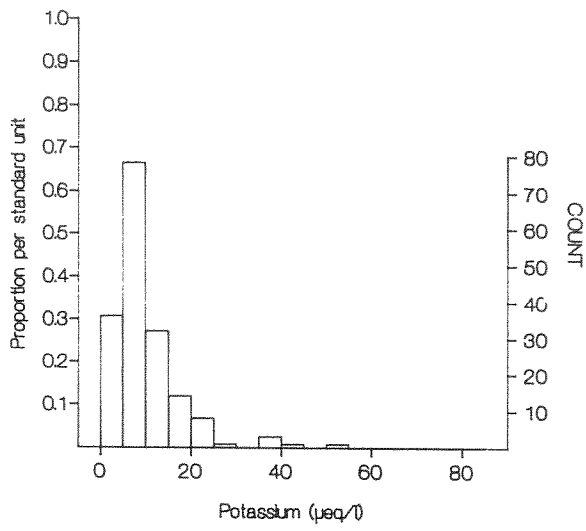


Figure 4 Continued.

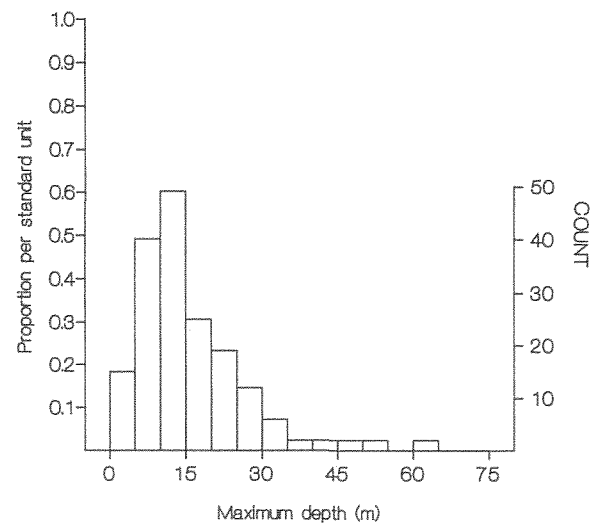
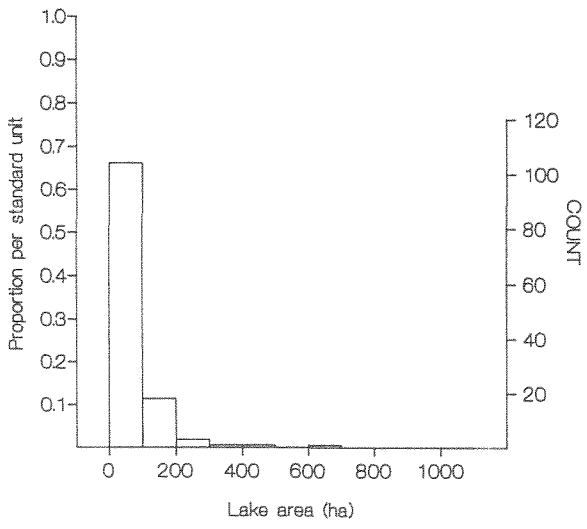
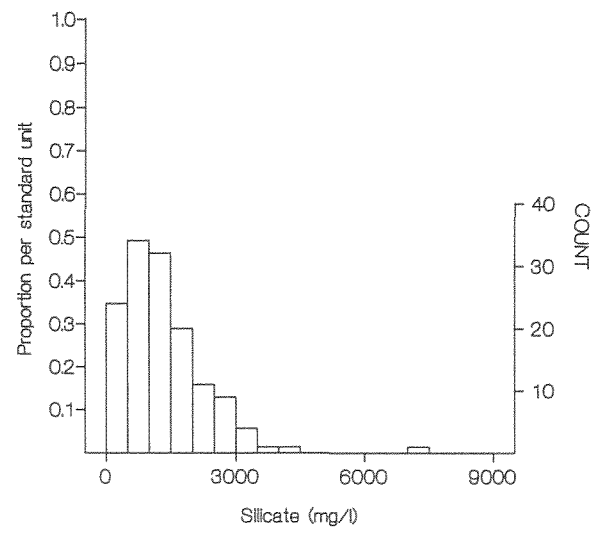
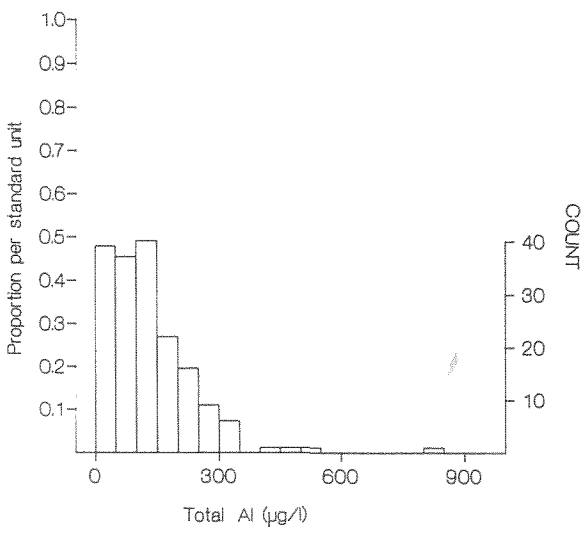


Figure 4 Continued.

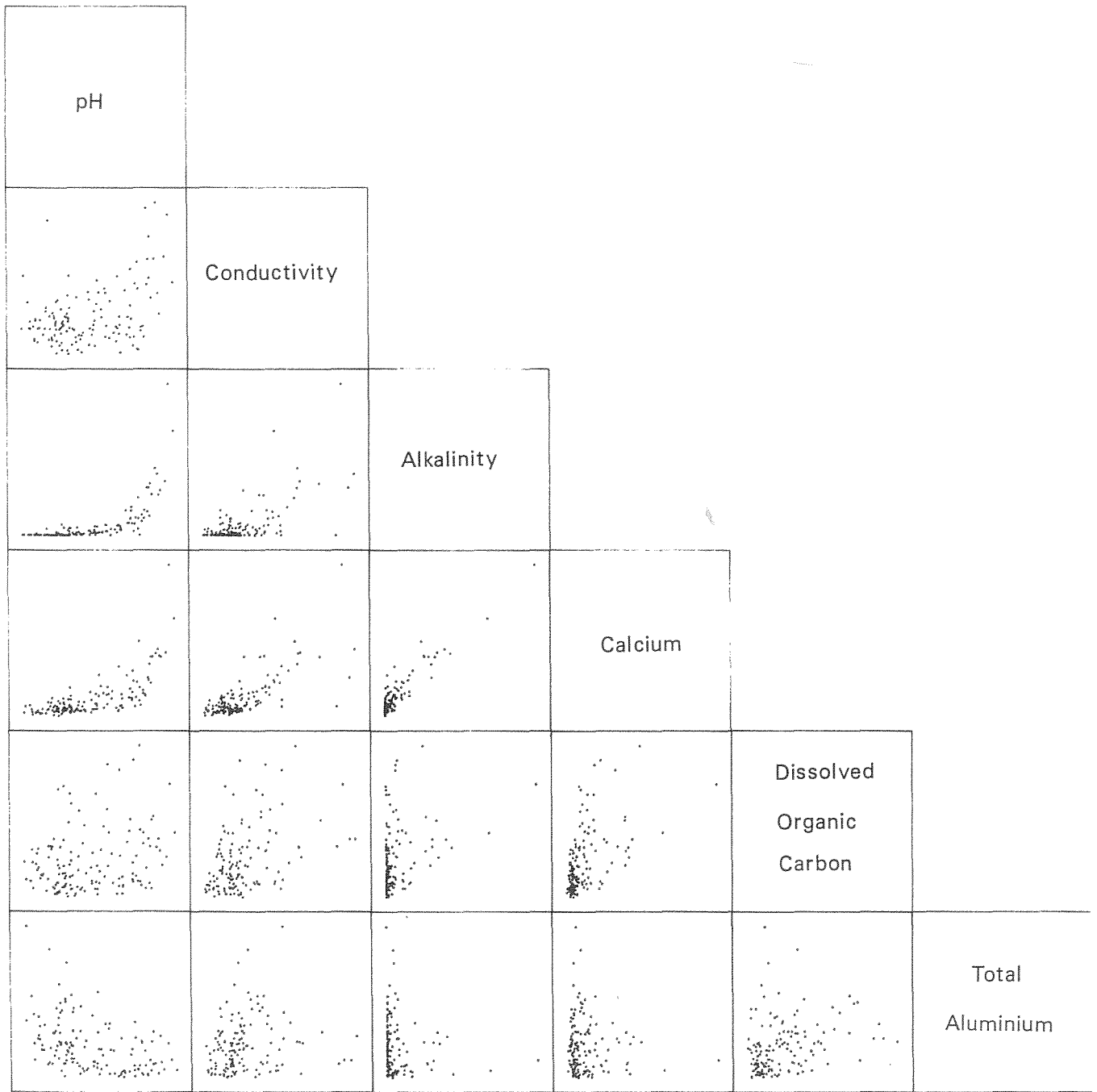


Figure 5 Matrix of uniformly scaled scatter plots for major chemical determinands, showing the patterns between pairs of chemical determinands.

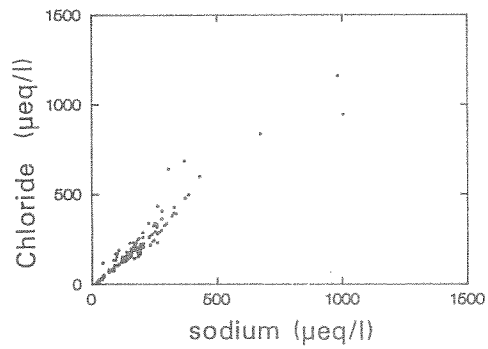
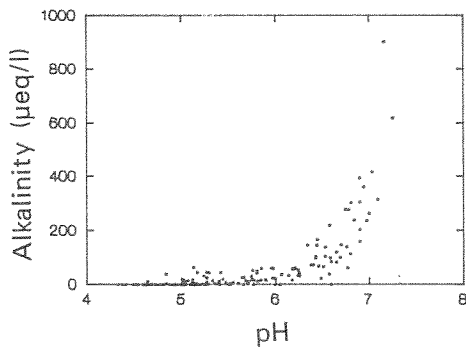
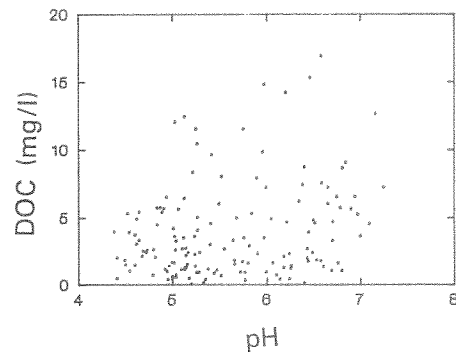
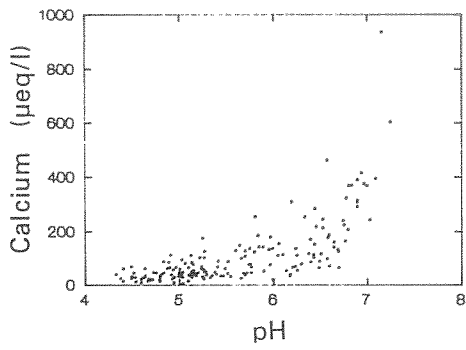
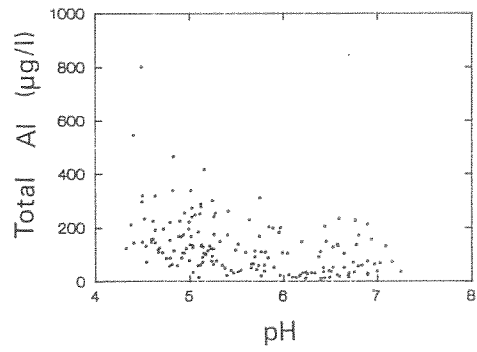


Figure 6 Scatter plots between selected chemical determinands.

The correlation matrix (Table 6) of environmental variables, with the significance levels adjusted for varying sample numbers, shows expected patterns of correlation between the variables. pH is positively correlated with both alkalinity and calcium levels (see also Figures 6 and 7), and negatively correlated with total aluminium. Altitude is negatively correlated with conductivity, DOC, magnesium, sodium, and potassium. The most acid sites are also the highest. Lake depth appears to have little correlation with lake-water chemistry.

Table 6 *Matrix of product-moment correlations between environmental variables*

	pH	Cond	DOC	Ca	Mg	Na	K	SO ₄
Conductivity	0.423*							
DOC	0.260	0.442*						
Ca	0.705**	0.714**	0.531*					
Mg	0.528**	0.856**	0.455*	0.704**				
Na	0.275	0.706**	0.361*	0.511*	0.823**			
K	0.494*	0.712**	0.495*	0.749**	0.755**	0.680**		
SO ₄	0.333*	0.764**	0.509*	0.629**	0.820**	0.672**	0.64**	
Cl	0.211	0.718**	0.138	0.295	0.788**	0.959**	0.517*	0.570*
NO ₃	0.117	0.446 *	-0.039	0.277	0.514*	0.310*	0.474*	0.410*
TON	0.107	0.198	0.542*	0.438*	0.209	0.003	0.423*	0.400*
Alkalinity	0.681**	0.674**	0.354*	0.914**	0.600*	0.424*	0.692**	0.400*
Total Al	-0.437*	0.139	0.279	-0.097	0.079	0.080	0.025	0.271
Zn	0.191	0.146	-0.128	0.159	0.155	0.164	0.019	0.256
Mn	-0.108	0.319*	0.395*	0.161	0.290	0.301*	0.430*	0.394*
Fe	-0.082	0.205	0.708**	0.145	0.234	0.191	0.303*	0.308*
SiO ₂	0.267	0.454*	0.435*	0.376*	0.523*	0.433*	0.496*	0.419*
Altitude	-0.186	-0.576**	-0.470*	-0.374*	-0.560*	-0.559*	-0.449*	-0.571*
Unforested	-0.118	-0.298	-0.503*	-0.308*	-0.249	-0.072	-0.188	-0.407*
Forested	0.118	0.298	0.503*	0.308*	0.249	0.072	0.188	0.407*
Conifer F.	0.099	0.295	0.530*	0.303*	0.249	0.064	0.196	0.400*
Deciduous F.	0.063	-0.001	-0.107	0.007	-0.011	0.025	-0.036	0.012
%afforest.	0.067	0.366*	0.279	0.252	0.261	0.165	-0.012	0.446*
Max Depth	-0.165	-0.394*	-0.312*	-0.280*	-0.337*	-0.299	-0.303*	-0.332*

Table 6 Continued.

	Cl	NO ₃	TON	Alkal	Tot. Al	Zn			
NO ₃	0.375*								
TON	-0.157	0.025							
Alkalinity	0.276	0.277	0.227						
Total Al	0.137	0.191	-0.010	-0.153					
Zn	0.123	-0.027	0.185	0.090	-0.121				
Mn	0.260	0.461*	0.165	0.081	0.252	-0.035			
Fe	0.081	0.102	0.235	0.015	0.443*	-0.090			
SiO ₂	0.348*	0.587*	-0.021	0.343	0.386*	-0.042			
Altitude	-0.498*	-0.171	-0.183	-0.269	-0.195	-0.072			
Unforested	-0.042	-0.058	-0.184	-0.239	-0.323*	-0.103			
Forested	0.042	0.058	0.184	0.239	0.323*	0.103			
Conifer F.	0.038	0.058	0.163	0.240	0.344*	0.105			
Deciduous F.	0.013		0.040	-0.013	-0.090	-0.012			
%afforest.	0.253	0.109	-0.177	0.192	0.428*	0.217			
Max Depth	-0.274	-0.102	-0.174	-0.232	-0.118	-0.055			
	Mn	Fe	SiO ₂	Altit	Unfor	Forest	Conif	Decid	%Aff
Fe	0.517*								
SiO ₂	0.385*	0.466*							
Altitude	-0.221	-0.250	-0.354*						
Unforested	-0.189	-0.272	-0.357*	0.503*					
Forested	0.189	0.272	0.357*	-0.503*	-1.000**				
Conifer F.	0.202	0.300*	0.393*	-0.468*	-0.964**	0.964**			
Deciduous F.	-0.052	-0.107	-0.137	-0.113	-0.098	0.098	-0.170		
%afforest.	0.151	0.149	0.355*	-0.419*	-0.769**	0.769**	0.799**	-0.070	
Max Depth	-0.231	-0.208	-0.159	0.347*	0.042	-0.042	-0.032	-0.037	0.171

(* = significant at $p \leq 0.05$, ** = significant at $p \leq 0.01$; all significance tests have been adjusted for varying sampling size)

(Max = maximum, afforest. = afforestation, F. = forest)

Principal Components Analysis

Principal components analysis (PCA) was used to summarise the major patterns of variation within the environmental data. Since not all determinands were available for all samples or all sites, a reduced data-set of 157 samples with nine chemical variables (pH, Cl, SO₄, conductivity, Mg, K, Ca, alkalinity, total Al) in common was selected for PCA. Figure 7 is a PCA correlation biplot derived from a PCA of the correlation matrix of the untransformed chemical data. The individual sample scores and variable loadings for this analysis are listed in Table 8. The seven catchment and physical variables (maximum depth, altitude, unforested, forested, forested with deciduous trees, forested with

coniferous trees, percentage afforested) were included as passive variables (ter Braak 1987). In a PCA correlation biplot, variables with high positive correlations generally have small angles between their biplot arrows. Variables with long arrows have high variance and are often the most important within the data (Jongman *et al.* 1987).

The PCA and all the other ordinations presented here were implemented by means of the computer program CANOCO 3.10 (ter Braak 1987, 1990).

Table 7 *Eigenvalues and cumulative variance accounted for in a PCA of the 157 sample x 9 chemical variable data-set*

	Eigenvalue	Cumulative variance represented
Axis 1	0.608	60.8%
Axis 2	0.184	78.1%
Axis 3	0.082	86.3%
Axis 4	0.047	91.0%

The variance within the data-set is efficiently captured with over 78% accounted for by the first two components. The first axis (60.8%) is effectively contrasting lakes with high pH/alkalinity and associated variables (Ca, K, Mg, SO₄, Cl) with acid sites, such as Bliksvatn (**82.1** on Figure 7) and the Round Loch of Glenhead (**RLGH**), positioned to the left-hand side of the PCA biplot. Alkaline sites such as Loch Fern (**FERN**), Loch Clonyard (**CLON**), and Edasjön (**S2**) are located on the right-hand side. Lake altitude and maximum depth, included as passive variables, are negatively correlated with this set of variables. Axis two (18.7%) is more complex and may reflect contrasts in aluminium, with high aluminium sites such as Loch Stroan (**STRO**), Loch Minnoch (**MINN**), and Härsvatten (**S22**), located at the positive end of axis 2 and low aluminium sites such as V. Helgtjärnen (**S15**), Flickersjön (**S8**), and Loch Whinyeon (**WHIN**) located at the negative end of axis 2. The contrast in the relative sizes of axes 1, 2, and 3 (eigenvalues 0.608, 0.184, 0.082, respectively) suggests that there is only one overriding gradient of variation in the chemical data, namely axis 1 reflecting the strong gradient from high to low pH, alkalinity, calcium, conductivity, etc. There is very little scatter of sites on axis 2, with the exception of two Swedish sites (**S22**, **S15**), and axis 2 may be primarily influenced by extreme values in these sites.

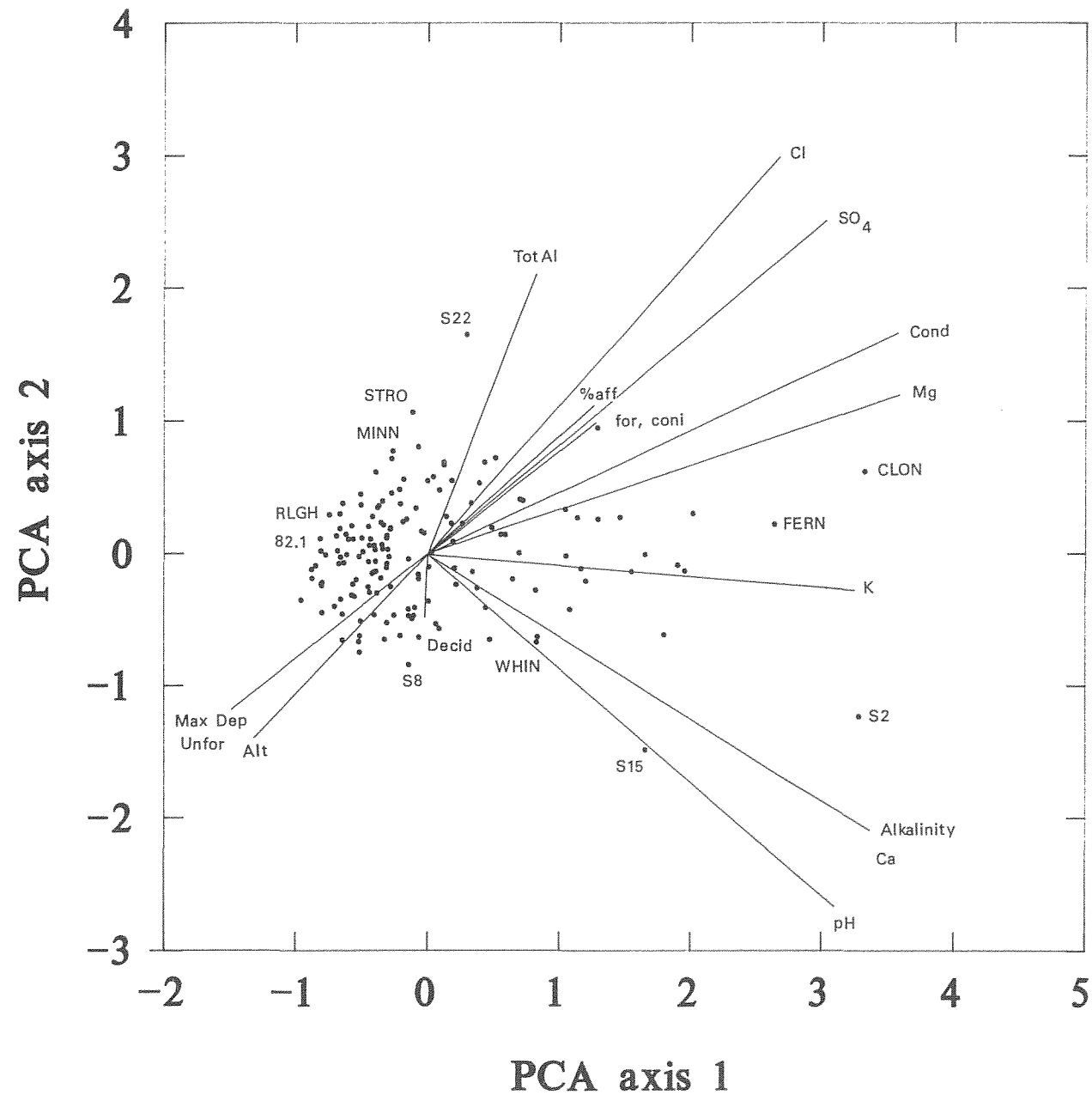


Figure 7 Principal components correlation biplot of the 157 sample x 9 chemical variable data-set. Catchment and other environmental variables are analysed as passive variables. The individual sample scores and variable loadings are listed in Table 8. See Table 2 for site codes. Abbreviations: tot = total, aff = afforestation, for = forested, conifer = conifer, cond = conductivity, alt = altitude, unfor = unforested, decid = deciduous, max dep = maximum depth, PCA = principal components analysis.

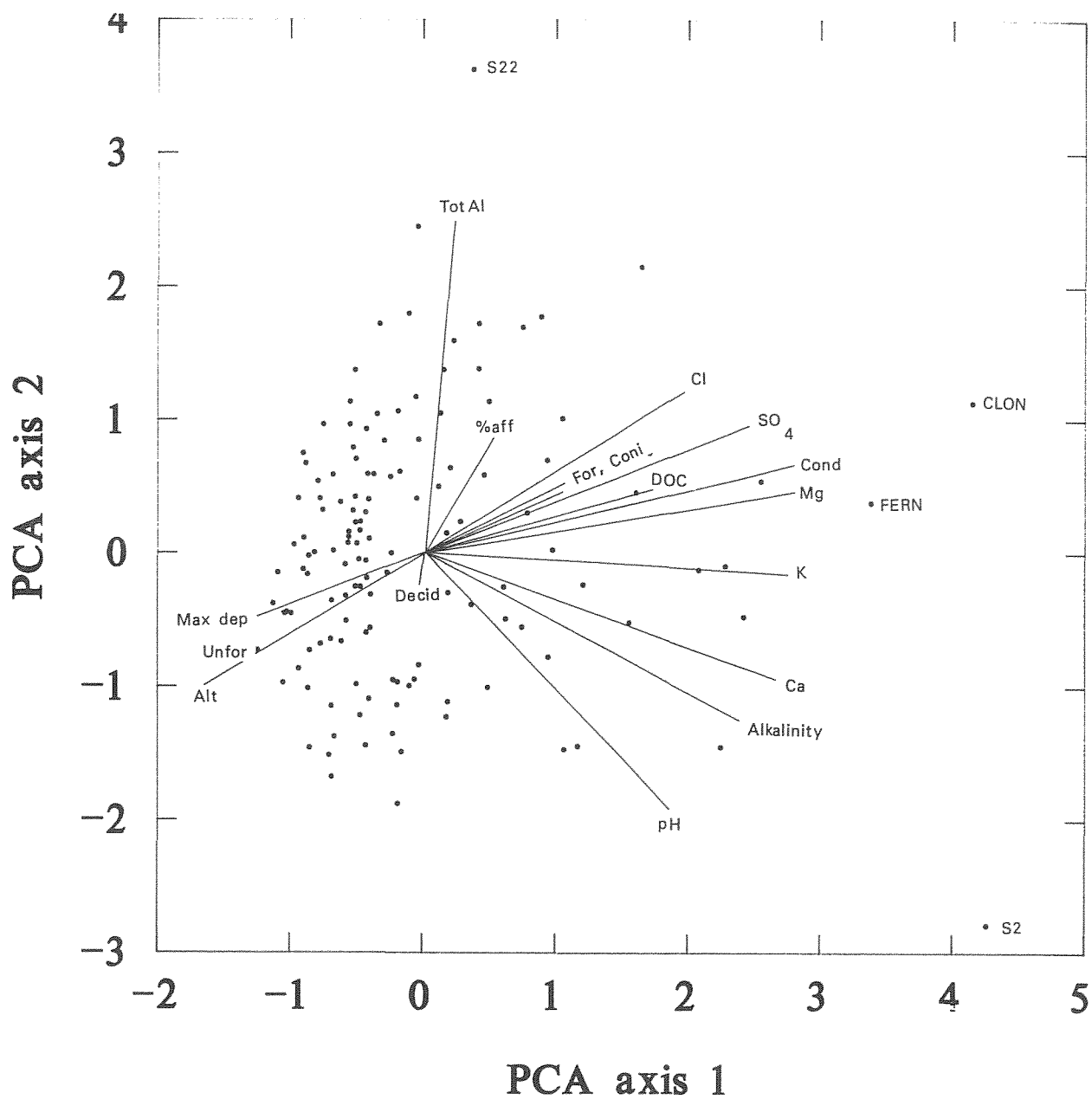


Figure 8 Principal components correlation biplot of the 138 x 10 chemical variable data-set. Catchment and other environmental variables are analysed as passive variables. The individual sample scores and variable loadings are listed in Table 8. See Table 2 for site codes. Abbreviations follow the caption for Figure 7.

Table 8 Sample scores and environmental variable loadings for the PCAs of the 157 and 138 sample data-sets.

	157 sample data-set				138 sample data-set			
	Axis 1	Axis 2	Axis 3	Axis 4	Axis 1	Axis 2	Axis 3	Axis 4
<i>Variable loadings</i>								
pH	0.6219	-0.6453	0.0120	-0.2047	0.8556	-1.3713	0.1815	-1.0496
Conductivity	0.9326	0.2283	0.2078	0.0417	1.1999	0.5589	0.5208	0.4665
DOC					0.5742	0.1637	-0.6619	-0.3308
Ca	0.8858	-0.3118	-0.2258	0.1607	1.1294	-0.6969	-1.0525	-0.0215
Mg	0.9344	0.1599	0.2020	-0.1072	1.2026	0.4038	0.5040	-0.4662
K	0.8770	-0.0480	-0.0444	0.1254	1.0880	-0.0879	-0.5255	0.6240
SO4	0.8187	0.3257	-0.0409	-0.2284	1.0163	0.8415	-0.2766	-2.0332
Cl	0.6516	0.4073	0.5819	-0.0345	0.8965	0.9998	1.7918	0.9214
Alkalinity	0.7929	-0.4151	-0.1132	0.3971	1.0413	-0.8933	-0.8445	1.4136
Total Al	0.0727	0.8320	-0.3688	0.2874	0.0411	1.9092	-1.7487	0.3041
Altitude	-0.5546	-0.3304	0.0320	0.1753	-0.6978	-0.7510	-0.1018	0.4185
Unafforested	-0.3599	-0.2127	0.4031	0.1217	-0.4360	-0.4583	1.0890	1.1294
Forested	0.3532	0.1582	-0.3994	-0.1579	0.4281	0.3382	-0.9932	-1.1553
Conifer F.	0.3535	0.1799	-0.4311	-0.1460	0.4249	0.3797	-1.0689	-1.0496
Deciduous F.	-0.0109	-0.0777	0.1182	-0.0360	-0.0037	-0.1594	0.3041	-0.3359
% afforested	0.1730	0.2895	0.1796	0.0949	0.2839	0.7094	0.4663	-0.1474
Max. depth	-0.4215	-0.1567	0.0221	0.0360	-0.4922	-0.3865	0.0565	0.0582

Sample scores

Sample number	Sample code	Axis 1	Axis 2	Axis 3	Axis 4	Axis 1	Axis 2	Axis 3	Axis 4
1	1.21	-0.0778	0.8051	-0.2540	0.1288	-0.1045	1.8021	0.1107	1.4889
2	10.21	0.1347	0.2803	-0.2154	-0.2352	0.2067	0.6454	-0.3514	-0.1510
4	111.11	-0.5203	-0.6220	0.0846	-0.0426	-0.6633	-1.3766	0.0209	-0.4591
5	113.21	-0.3311	-0.6494	0.0801	0.0843	-0.4264	-1.4435	0.0973	0.0432
6	115.11	-0.3576	-0.1833	0.2784	0.1290	-0.3857	-0.3111	0.2554	-0.6728
7	12.11	-0.8007	-0.2170	-0.0976	-0.0105	-1.0225	-0.4400	-0.2066	0.3170
9	15.11	-0.4983	0.0154	-0.2968	-0.1136	-0.5550	0.1197	-0.9998	-0.0099
10	17.21	-0.0946	0.3418	-0.1329	-0.2082	-0.0278	0.8559	-0.5559	-0.4970
11	18.11	-0.7096	-0.3968	0.0083	0.0355	-0.9317	-0.8625	0.1247	0.2634
13	19.21	-0.6519	-0.6538	0.0102	-0.0858	-0.8519	-1.4595	-0.0519	-0.2365
14	2.11	-0.1635	0.2589	0.1619	0.0323	-0.2397	0.5766	0.8460	0.2405
15	20.11	-0.5231	-0.7463	-0.0291	-0.1375	-0.6855	-1.6774	-0.1813	-0.3511
17	21.11	-0.3888	-0.2969	-0.1631	-0.2522	-0.4202	-0.5953	-0.8573	-0.7865
18	3.11	-0.6712	0.1895	-0.0664	0.1180	-0.7878	0.5414	-0.3380	0.3774
20	34.11	-0.9605	-0.3494	0.0670	0.1053	-1.2460	-0.7249	0.2044	0.3554
22	4.11	-0.6503	0.3787	-0.2868	0.1007	-0.7483	0.9669	-0.7877	0.8015
23	42.11	-0.8082	-0.2254	0.0993	0.1231	-1.0416	-0.4487	0.2927	0.3031
24	44.21	-0.5293	-0.6663	0.0963	-0.3085	-0.7050	-1.5149	0.1319	-0.9994
25	49.11	-0.6803	-0.0738	0.1764	0.1440	-0.8954	-0.1214	0.6529	0.3881
26	5.11	-0.2779	0.7136	-0.4504	-0.0404				
27	58.11	-0.6639	-0.3422	0.1443	0.0336	-0.8493	-0.7257	0.2818	-0.2048
28	59.11	-0.8041	-0.4441	0.0072	-0.0329	-1.0535	-0.9688	0.0682	0.0830

Table 8 Continued.

Sample number	Sample code	157 sample data-set				138 sample data-set			
		Axis 1	Axis 2	Axis 3	Axis 4	Axis 1	Axis 2	Axis 3	Axis 4
36	80.11	-0.3468	0.3951	-0.0976	0.0927	-0.4216	0.9309	0.0348	0.7039
37	81.21	-0.5138	0.4470	-0.2886	0.0708	-0.5439	1.1348	-0.9367	0.5331
38	82.11	-0.8190	0.1137	-0.1834	0.1122	-0.9353	0.4145	-0.8947	0.3161
39	83.11	-0.8496	-0.0893	-0.0089	0.1400	-1.0918	-0.1449	0.1095	0.6483
40	86.11	-0.5708	0.1122	-0.2986	-0.0855	-0.6179	0.3841	-1.1531	0.0054
41	87.21	-0.8060	-0.2399	-0.0522	-0.0250	-0.9913	-0.4492	-0.3648	-0.0590
42	88.21	-0.8767	-0.1849	-0.1530	0.0301	-1.1266	-0.3753	-0.2655	0.6134
43	89.11	-0.6516	-0.4567	0.1478	-0.0561	-0.8631	-1.0111	0.3977	-0.2920
44	9.11	-0.7794	-0.0082	-0.1318	0.0376	-0.9688	0.0636	-0.3294	0.4745
45	ACH1	-0.5458	-0.1944	0.3591	0.1904	-0.6816	-0.3544	0.7492	-0.2603
46	ARR1	-0.1503	-0.4188	0.2994	0.1599	-0.2201	-0.9462	0.8654	-0.1588
47	ARTH1	1.9557	-0.1288	-0.1294	-0.0003	2.4230	-0.4717	0.7026	0.9739
48	BARE1	2.0155	0.3022	0.7497	0.0215	2.5513	0.5461	2.3842	-0.9427
49	BARL1	-0.1217	-0.4914	0.1029	-0.0388	-0.1886	-1.1362	0.3952	-0.0741
50	BER1	-0.2164	0.4815	0.1495	-0.0327				
51	BODG1	0.6487	-0.1912	0.2500	-0.2507	0.7506	-0.5489	1.1684	-0.5334
52	BODL1	-0.4228	-0.1463	0.2338	0.0147	-0.5723	-0.3198	0.7780	-0.1397
53	BREC1	0.7226	0.3983	0.1553	-0.2181				
54	BUGE1	-0.3387	0.2175	0.2122	0.1656	-0.3653	0.5938	0.3290	-0.1236
56	BYCH1	0.4403	-0.4079	0.3146	-0.0403	0.4874	-1.0017	1.2266	-0.2317
57	CFYN1	-0.5799	-0.3119	0.1832	0.0449	-0.7675	-0.6768	0.5518	-0.0501
58	CHN1	-0.3094	0.0335	0.1518	0.0671	-0.4002	0.1070	0.5503	0.0901
59	CLON1	3.3234	0.6187	0.4663	0.0291	4.1512	1.1346	2.6187	0.0634
60	CLYD1	-0.2599	-0.4656	0.2058	-0.0444	-0.4007	-1.0915	0.8243	-0.2172
61	CON1	-0.5274	-0.0207	0.1590	0.1160	-0.6694	0.0172	0.4662	0.1460
62	COR1	-0.6436	-0.0682	-0.1516	0.1124	-0.8630	-0.1582	0.0683	1.0511
63	CWBY1	-0.3145	0.1367	0.0501	0.0182	-0.4271	0.3057	0.5005	0.3498
65	DIWA1	-0.4448	-0.2909	0.1878	-0.0152	-0.6077	-0.6592	0.6540	-0.1577
66	DOI1	-0.0727	-0.1556	0.3546	0.0965	-0.1142	-0.3530	1.0410	-0.3889
67	DOON1	-0.0530	0.1689	0.1099	0.0206	-0.0424	0.4142	0.3693	-0.0980
68	DUH1	-0.6675	-0.0240	-0.1115	0.0593	-0.8581	-0.0204	-0.0715	0.6351
69	DUL1	-0.3529	0.0687	0.1318	0.0564	-0.4702	0.1686	0.6074	0.1876
70	ENO1	-0.6896	0.0231	0.1372	0.2178	-0.8942	0.1143	0.5543	0.6335
71	EUN1	-0.6716	0.2998	-0.3587	0.1176	-0.8791	0.6754	-0.3281	1.5963
72	FERN1	2.6373	0.2218	-0.1908	0.1500	3.3830	0.3834	0.3597	0.7305
73	FHI1	-0.0010	0.5469	-0.0284	0.1450	-0.0468	1.1734	0.7316	1.0814
74	FINL1	0.0874	0.4776	-0.2927	0.1005	0.1368	1.0523	-0.2826	1.0908
75	FLE1	-0.5811	0.2110	0.0069	0.1396	-0.6730	0.5915	-0.1534	0.2702
76	GARN1	-0.1034	-0.4075	0.1835	-0.0995	-0.1871	-0.9628	0.7200	-0.3330
77	GEIR1	0.8166	-0.2764	0.2518	-0.3151	0.9452	-0.7734	1.2701	-0.6088
78	GLAS1	-0.3134	-0.5226	0.2104	-0.0607	-0.4669	-1.2166	0.7768	-0.2761
79	GLFR1	1.7961	-0.6129	0.0994	0.7303	2.2468	-1.4500	0.9890	2.0303
80	GLYN1	-0.4122	0.0630	0.1096	0.0077	-0.5522	0.1587	0.5769	0.1784
81	GOD1	0.1861	0.0916	0.4417	-0.0686	0.1812	0.1527	1.5325	-0.5483
82	GREENT1	-0.3159	0.1186	0.1662	0.0559	-0.4659	0.2356	0.9708	0.4482
84	GWYN1	-0.0713	-0.6325	0.1290	-0.0876	-0.1576	-1.4887	0.5878	-0.1191
85	GYN1	-0.4027	-0.0564	0.1938	0.0370	-0.4767	-0.0479	0.3134	-0.4099
86	HARR1	-0.3980	0.6127	-0.3964	0.1517	-0.5042	1.3708	-0.3754	1.7484
87	HIR1	-0.2860	0.1947	0.1888	0.0218				
90	HOW1	0.3916	0.5300	0.0309	-0.1519	0.4974	1.1404	0.5481	-0.0251

Table 8 Continued.

Sample number	Sample code	157 sample data-set				138 sample data-set			
		Axis 1	Axis 2	Axis 3	Axis 4	Axis 1	Axis 2	Axis 3	Axis 4
95	LAG1	-0.4552	-0.0583	0.2135	0.0721	-0.5784	-0.0845	0.5634	-0.1090
96	LAI1	-0.5617	-0.3206	0.2104	0.0812	-0.6908	-0.6406	0.2866	-0.3727
97	LAR1	1.2840	0.9481	1.4318	0.6788	1.6426	2.1511	3.9893	-1.0271
98	LCSL1	-0.3481	0.0380	0.1240	0.0882	-0.4943	0.0740	0.7797	0.4483
99	LCSU1	-0.3190	-0.1032	0.1651	0.0083	-0.4647	-0.2522	0.8338	0.1122
100	LDE1	-0.3972	-0.1312	0.1009	0.1502	-0.5018	-0.2500	0.3256	0.2910
101	LDE2	-0.5722	-0.2294	0.0309	-0.0474				
102	LENY1	0.0026	-0.3578	0.2090	-0.1499	-0.0252	-0.8327	0.6281	-0.6155
103	LGR1	-0.5116	0.3669	-0.0391	0.2071	-0.5429	0.9657	-0.4091	0.3221
106	LLDU1	0.1968	-0.1088	0.3549	-0.0342	0.1935	-0.2929	1.2900	-0.3542
107	LLGH1	-0.4230	0.2795	0.0525	0.1576	-0.4984	0.7086	0.1727	0.4025
108	LOD1	-0.4564	-0.2491	0.2684	0.1334	-0.5722	-0.5036	0.5688	-0.2793
109	LOWT1	-0.4049	0.0315	0.1817	0.0695	-0.5601	0.0754	0.8338	0.3078
110	MABE1	0.4320	0.6860	0.0860	-0.0204	0.7559	1.6974	-0.5043	-1.1041
111	MACA1	-0.3587	0.2356	0.0031	0.1547	-0.4120	0.5982	0.0350	0.3999
112	MANN1	1.2868	0.2600	-0.0445	-0.0691	1.6005	0.4606	0.7085	0.5176
113	MINN1	-0.2684	0.7714	-0.5659	0.1837	-0.3181	1.7205	-0.7518	2.1448
114	MOAN1	0.1193	0.6873	0.0324	-0.0207	0.2346	1.5969	0.0875	-0.1530
115	MUCK1	0.1793	0.5478	-0.0148	0.1322	0.4209	1.3852	-0.6944	-0.4508
116	NAGA1	-0.6216	-0.0058	-0.0927	0.0543	-0.8179	0.0066	0.0979	0.6962
117	OCHI1	0.7002	0.4071	0.2224	-0.0490	1.0536	1.0125	-0.0266	-1.4011
118	PARC1	1.0438	-0.0143	0.4303	-0.4552	1.2079	-0.2340	1.9178	-1.1270
119	PENR1	-0.4384	0.1118	0.0118	0.0485	-0.5215	0.3201	0.0015	0.1102
120	RIEC1	-0.1921	0.2407	0.0200	0.1028	-0.1701	0.6173	-0.0981	0.0706
121	RLGH1	-0.6077	0.1102	0.1155	0.1826	-0.7534	0.3262	0.3230	0.3775
122	RLGH2	-0.6984	0.1351	-0.0786	0.1213				
123	RONA1	1.4558	0.2734	0.4664	0.0993				
125	S101	0.0600	-0.5312	-0.2806	-0.1015	0.1902	-1.1108	-1.3009	-0.4745
126	S11	0.8223	-0.6709	-0.3789	-0.1850	1.1665	-1.4471	-1.4891	-0.3254
127	S111	0.5131	0.7182	-0.4354	-0.2479	0.8936	1.7773	-1.8247	-0.8881
128	S121	-0.2128	-0.6206	-0.0985	-0.2185	-0.2214	-1.3529	-0.6577	-0.8069
129	S131	-0.0324	0.1558	-0.1132	-0.1633	0.1218	0.5040	-1.0014	-1.0673
130	S141	-0.1486	-0.4699	-0.2083	-0.1224	-0.0580	-0.9410	-1.2831	-0.7979
131	S151	1.6496	-1.4784	-1.0020	0.8583	2.1129	-3.3329	-1.9708	4.4456
132	S161	0.3770	-0.2596	-0.2320	-0.2409	0.6247	-0.4882	-1.2533	-1.1299
133	S171	1.0401	0.3339	-0.0981	-0.7379	1.2769	0.6174	0.4494	-1.2025
134	S181	0.5886	0.1419	0.0646	-0.4657	0.7887	0.3061	0.1521	-1.4535
135	S191	0.1201	0.6652	-0.1762	-0.0156	0.4237	1.7252	-1.5134	-1.0002
136	S201	-0.4162	-0.1397	-0.1799	-0.0488	-0.4169	-0.1857	-0.9944	-0.4566
137	S21	3.2779	-1.2273	-1.2559	0.6987	4.2564	-2.7824	-2.5659	4.7998
139	S221	0.2929	1.6497	-0.5734	0.1067	0.3747	3.6315	-0.1663	2.4726
140	S241	-0.0728	-0.1874	-0.1574	-0.2674	0.0336	-0.3168	-1.0576	-1.1412
141	S251	0.8286	-0.6308	-0.3033	-0.2405	1.0640	-1.4707	-0.6515	0.1554
142	S261	1.5459	-0.1347	-0.3595	-0.1706	2.2845	-0.0933	-2.2978	-1.4738
143	S271	-0.3854	0.3472	-0.5051	-0.0211	-0.1823	1.0669	-2.6607	-0.6701
144	S281	-0.5060	0.1184	-0.3063	0.0091	-0.5049	0.4265	-1.3086	-0.0141
145	S291	-0.4490	0.0631	-0.2079	-0.0868	-0.5038	0.2316	-0.7028	-0.0224
146	S301	-0.3156	-0.0727	-0.2321	-0.1110	-0.2352	0.0038	-1.4078	-0.8056
147	S31	0.0824	-0.5670	-0.1177	-0.2275	0.1797	-1.2249	-0.7761	-0.9447
148	S311	0.2538	0.2265	-0.1919	-0.2759	0.4604	0.5918	-0.9577	-1.0020

Table 8 Continued.

Sample number	Sample code	157 sample data-set				138 sample data-set			
		Axis 1	Axis 2	Axis 3	Axis 4	Axis 1	Axis 2	Axis 3	Axis 4
153	S81	-0.1479	-0.8408	-0.0382	0.0528	-0.1857	-1.8751	-0.2424	0.0810
154	S91	1.1299	0.2711	-0.1630	-0.7587	1.7083	0.7573	-1.6814	-3.1043
155	SCOATT1	-0.3755	0.3596	-0.0865	0.2226	-0.5194	0.7938	0.4549	1.3799
156	SKAK1	0.3259	0.3789	0.0808	-0.1798				
157	SKAK2	0.0050	-0.0997	0.2179	-0.2361				
158	SKE1	0.0390	0.5764	-0.2547	0.0923	0.1611	1.3764	-0.7114	0.6028
159	SKE2	-0.4544	0.2054	-0.0890	0.0599				
160	SKOMAKV1	-0.8156	0.0209	-0.0362	0.1696				
161	STRO1	-0.1199	1.0655	-0.6244	0.0735	-0.0371	2.4521	-1.3273	1.5831
162	TANN1	-0.2140	0.3753	-0.0190	0.1488	-0.2871	0.8465	0.4483	0.8230
164	TECW1	0.3373	-0.1344	0.3597	-0.2683	0.3648	-0.3805	1.2905	-0.9536
165	TINK1	-0.4116	-0.4615	0.1378	-0.0693	-0.4956	-0.9809	0.0326	-0.7185
166	TROO1	-0.2796	0.4546	-0.1120	0.1439	-0.3391	1.0435	0.0758	0.8940
167	UAI1	-0.1524	-0.0393	0.1369	0.0263	-0.2627	-0.1440	0.8966	0.3189
168	UIS1	-0.1121	-0.4680	0.1587	0.1328	-0.0979	-0.9902	0.0980	-0.4431
169	URR1	0.5517	0.1436	-0.0936	-0.2605				
170	VAL1	-0.6271	0.1465	0.0727	0.2162	-0.7729	0.4121	0.2243	0.5361
172	WHIN1	1.1587	-0.1114	-0.2475	0.2437				
173	WHIN2	0.4714	-0.6510	0.0739	-0.0485				
174	WHIT1	1.8959	-0.0859	0.5932	0.5425				
175	WOOD1	1.6483	-0.0054	0.1100	-0.0318	2.0838	-0.1200	0.8519	0.1595
176	WOOD2	1.0748	-0.4202	0.0223	-0.0802				
177	YBI1	-0.2904	0.1841	0.1036	0.0306	-0.4034	0.4051	0.6824	0.3178
178	YGAD1	-0.2852	-0.2484	0.1382	0.0182	-0.3894	-0.5575	0.5169	-0.0319

In order to incorporate DOC into the analysis, the data-set was further restricted to the 138 samples with DOC data. The correlation biplot of an analysis of the correlation matrix is shown in Figure 8. The summary statistics are shown in Table 9 while the variable loadings and sample scores for this analysis are listed in Table 8.

Table 9 *Eigenvalues and cumulative variance accounted for in a PCA of the 138 sample x 10 chemical variable data-set*

	Eigenvalue	Cumulative variance represented
Axis 1	0.574	57.4%
Axis 2	0.176	75.0%
Axis 3	0.106	85.6%
Axis 4	0.050	90.6%

As with the PCA of the 157 sample data-set, the PCA of the 138 sample data-set represents an efficient summary of the data with 75% of the variance of the data-set accounted for by the first two axes (Figure 8). The dominant trends appear to be the same as the 157 lake analysis, namely high to low pH, calcium, etc. Samples with high DOC values lie mid-way along axis 1. The inclusion of a limited range of catchment characteristics albeit run as passive variables, suggests that conifer-forested catchments tend to have high DOC values, and that deep, high-altitude lakes that have unforested catchments have low DOC values (Figure 8). The most acid lakes show no strong correlations with any catchment characteristics.

NUMERICAL ANALYSIS OF THE DIATOM DATA

Basic statistics

A total of 277 diatom taxa was recorded in the 178 samples within the SWAP training set with a relative abundance of greater than 1% in at least one sample. A total of 210 taxa was present in fewer than 25% of all samples. Appendix B lists full names, authorities, codes for all taxa, and environmental optima estimated by weighted averaging regression for selected taxa.

The principal patterns of floristic variation in these data have been explored using detrended correspondence analysis (DCA), whereas the patterns of floristic variation in the data that can be explained statistically by the measured environmental variables are detected by means of canonical correspondence analysis (CCA).

Detrended correspondence analysis

DCA (Hill and Gauch 1980) is one of the most powerful and numerically robust ordination or indirect gradient analytical techniques currently available (Jongman *et al.* 1987). It provides a good approximation to the problem of ordinating samples and species in two or more dimensions when the underlying model of species responses to their environment is unimodal, when the data contain many zero values for samples at which a species is absent, and when the number of species is large (ter Braak and Prentice 1988).

In the DCA presented here (Table 10, Figures 9 and 10), detrending was by segments with nonlinear rescaling of axes. Rare species were downweighted (ter Braak 1987). Not surprisingly with 277 taxa and many zero values in the data, the DCA axes do not capture a large proportion of the cumulative variance of the species data (Table 10). Despite the low (14.6%) of the variance or "inertia" of the species data explained by two dimensions, the ordination plot is still informative. Low percentages explained commonly result from noisy data containing many zero values.

Table 10 *Eigenvalues and cumulative variance accounted for in a DCA of the 178 diatom samples*

	Eigenvalue	Cumulative variance represented
Axis 1	0.598	9.3%
Axis 2	0.338	14.6%
Axis 3	0.256	18.6%
Axis 4	0.187	21.5%

It is not possible statistically to relate these axes to the measured environmental variables by, for example, regression because only one variable (pH) is available for all 178 samples (Table 5). Preliminary interpretation of the DCA axes is therefore attempted solely on the basis of the known ecological preferences of the major diatom taxa. Axis 1 appears to be partly a pH axis with alkaline species and sites positioned on the right-hand side of the plot, e.g. *Aulacoseira ambigua* (Au002a on Figure 10), *Aulacoseira granulata* var. *angustissima* (Au003d), *Cyclostephanus dubius* (Cc001a), Hagsjön (S26 on Figure 9), Vagsjöarna (S25), and Glasfryn (GLAS), and acid species such as *Tabellaria binalis* (Ta003a), *Semiorbis hemicyclus* (Se001a), *Eunotia bactriana* (Eu046c), *Aulacoseira tethera* (Au023a), and *Surirella biseriata* var. *biseriata* (Su004a) and acid lakes, e.g. Llyn y Bi (YBI), Ljosvatn (LJOSV), and Scoat Tarn (SCOAT) positioned to the left. No interpretation is presented for axis 2 or later axes. Interpretation of the diatom data in relation to the available environmental

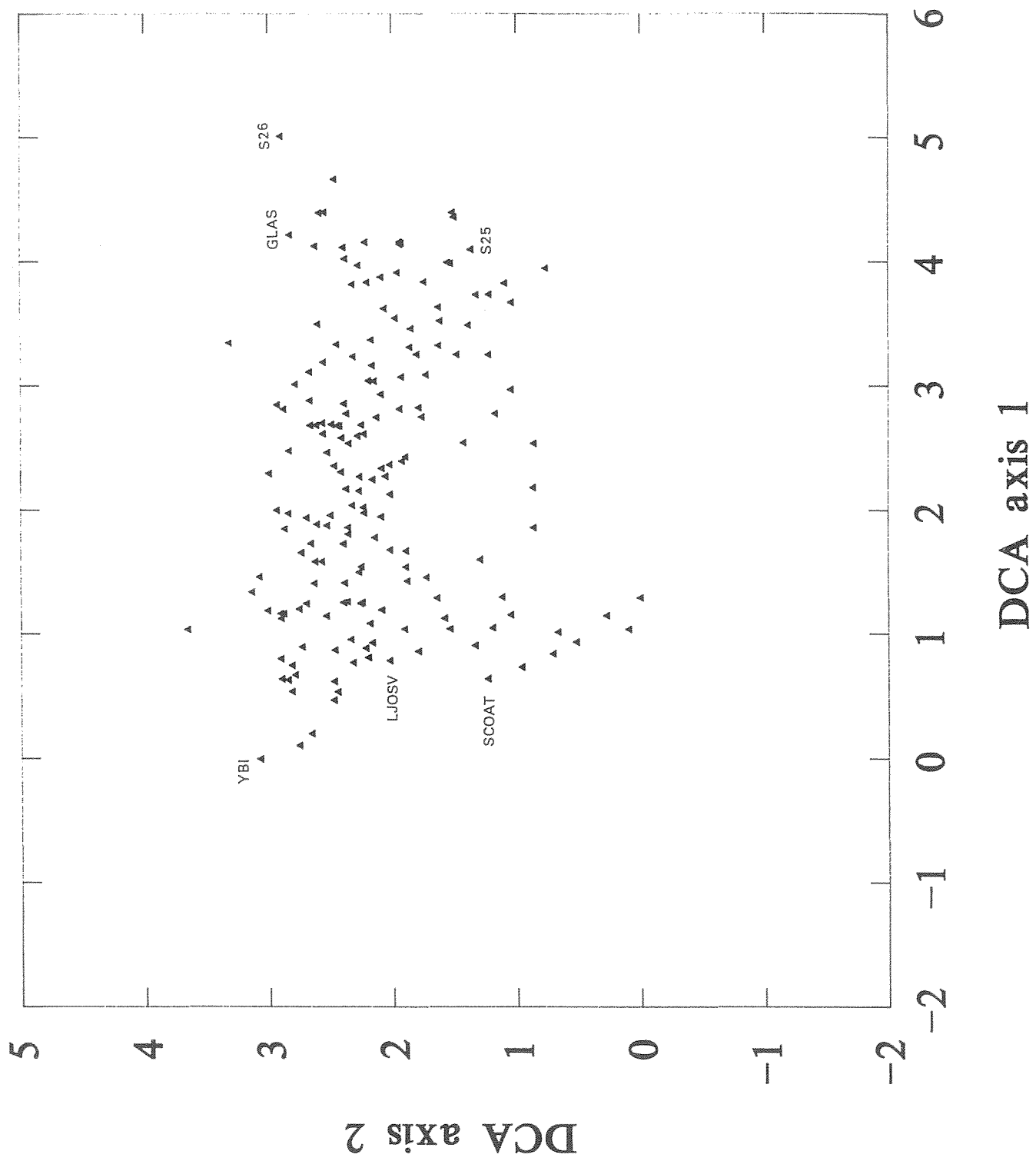


Figure 9 Detrended correspondence analysis (DCA) plot of the 178 diatom samples on axes 1 and 2. See Table 2 for site codes. Only selected samples are labelled.

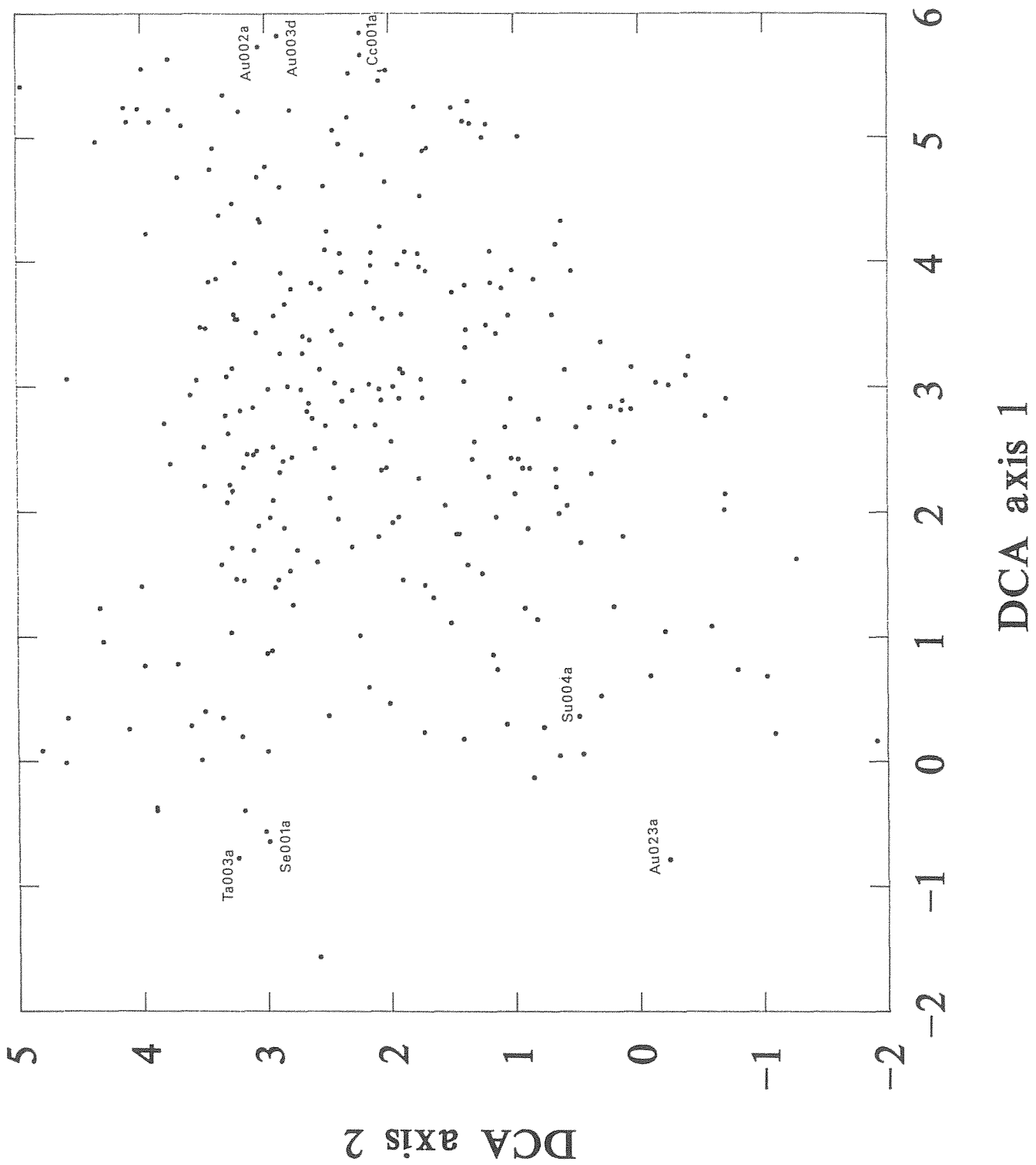


Figure 10 Detrended correspondence analysis (DCA) plot of the diatom taxa in the 178 samples on DCA axes 1 and 2. See Appendix B for taxon codes. Only selected taxa are labelled.

data is more readily attempted by means of the direct gradient technique of canonical correspondence analysis.

Canonical correspondence analysis

Like DCA, canonical correspondence analysis (ter Braak 1986) derives a set of ordination axis scores for species and for samples. For the first axis, species scores and sample scores are chosen to maximise the correlation between them. For subsequent axes, the sample and species scores are also maximally correlated but are chosen to be uncorrelated to species and sample scores on previous axes. In CCA the sample scores are also constrained to be simple linear combinations of the available environmental variables. A CCA ordination diagram can thus include not only the sample and site scores but also vectors representing the different environmental variables. In such a diagram, the vector for an environmental variable points in the direction of maximum variation of that variable across the diagram, and its length is proportional to its importance. Environmental variables with long vectors are more strongly correlated with the ordination axes than those with short vectors, and are more closely related to the patterns of biological variation displayed in the diagram (ter Braak 1986, 1987). A CCA diagram thus simultaneously displays the major patterns of variation in the diatom assemblages, as far as these can be related to the available environmental variables and the major patterns in the weighted averages of the individual species in relation to the environmental variables. CCA assumes a common unimodal response model for all species in relation to the measured environmental gradients.

The same 138 samples that were used for in the PCA of the chemical data were used here. The CCA presented is similar to that shown in Birks *et al.* (1990b) except that here the environmental data were not subjected to logarithmic transformation, all 138 samples were included, and physical variables such as maximum depth and altitude and catchment attributes such as forest status and forest type were included. In both analyses rare species were downweighted. The results using untransformed environmental data are very similar to those obtained with logarithmically transformed data.

The CCA results are shown in Figures 11 and 12. The CCA plot can be interpreted in a broadly similar manner to that of the PCA biplot with small angles between those environmental variables that are highly correlated, and with variables with the highest variance and hence often being the most important having the longest arrows. Tables 11 and 12 present a summary of the CCA statistics while Table 13 presents the full species and site scores and environmental variable loadings. The CCA of the 138 data-set captures the variance in the species abundance-environmental relationship well with

over 46% of the variance of the species weighted averages being accounted for by the first two axes.

Table 11 Summary statistics of the environmental variables included in the CCA of the 138 diatom-environmental data-set

	Weighted mean	Standard deviation	Variance inflation factor			
pH	5.5598	.7050	3.7455			
Conductivity	43.6492	26.2270	46.0034			
DOC	3.8761	3.2846	3.0576			
Ca	101.5286	107.7411	27.4217			
Mg	64.9368	49.5752	15.5114			
K	9.9082	6.9200	3.4507			
SO ₄	105.1257	64.1189	15.6147			
Cl	182.5068	141.3208	22.2756			
Alkalinity	46.2302	98.4989	20.0193			
Total Al	125.3340	94.9714	2.5304			
Altitude	352.1492	262.1623	2.3075			
Unforested	.3984	.4896	46.5526			
Forested	.5938	.4911	55.5266			
Conifer F.	.5709	.4949	12.5739			
Deciduous F.	.0228	.1494	.0000			
Maximum depth	15.7617	11.2323	1.4453			
CCA axes			1	2	3	4
Eigenvalues			.483	.248	.179	.141
Species-environment correlations			.932	.847	.783	.788
Cumulative percentage variance						
of species data			7.8	11.8	14.6	17.0
of species-environment relationship			30.4	46.1	57.4	66.7
(F. = forest)						

The CCA results (Tables 11 and 12, Figures 11 and 12) clearly show that axis 1 is strongly related to pH ($r = 0.84$), along with conductivity (0.58), calcium (0.78), and alkalinity (0.69), and negatively correlated with altitude (-0.45), maximum depth (-0.35), and total aluminium (-0.21). Thus high pH lakes with associated high alkalinity, calcium, potassium, magnesium, and conductivity contrast with low pH lakes of low alkalinity, calcium, potassium, magnesium, and conductivity. The low pH lakes tend to be at high altitudes, to be deep, to have unforested catchments, and to have high total aluminium. Acid lakes such as Loch Enoch (ENO on Figure 11), Hovvatn (4.1), and Övre Brandsvatn

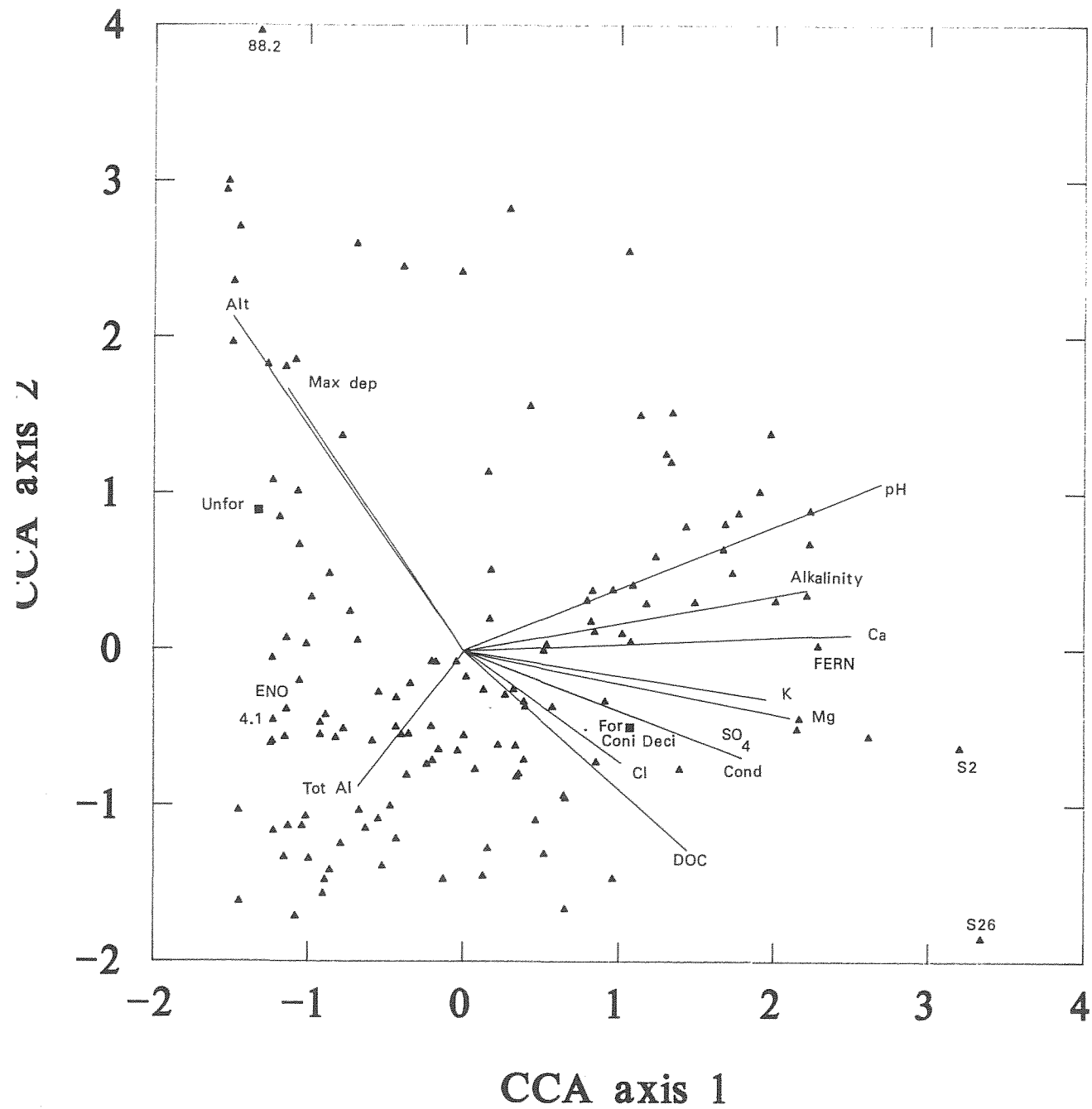


Figure 11 Canonical correspondence analysis (CCA) plot of the 138 samples on CCA axes 1 and 2 in relation to the 16 environmental variables included in the analysis. Continuous environmental variables are shown as vectors from the plot origin, nominal variables are plotted as centroids and shown as solid squares. Only selected samples are labelled. See Table 2 for site codes. Abbreviations follow the caption for Figure 7.

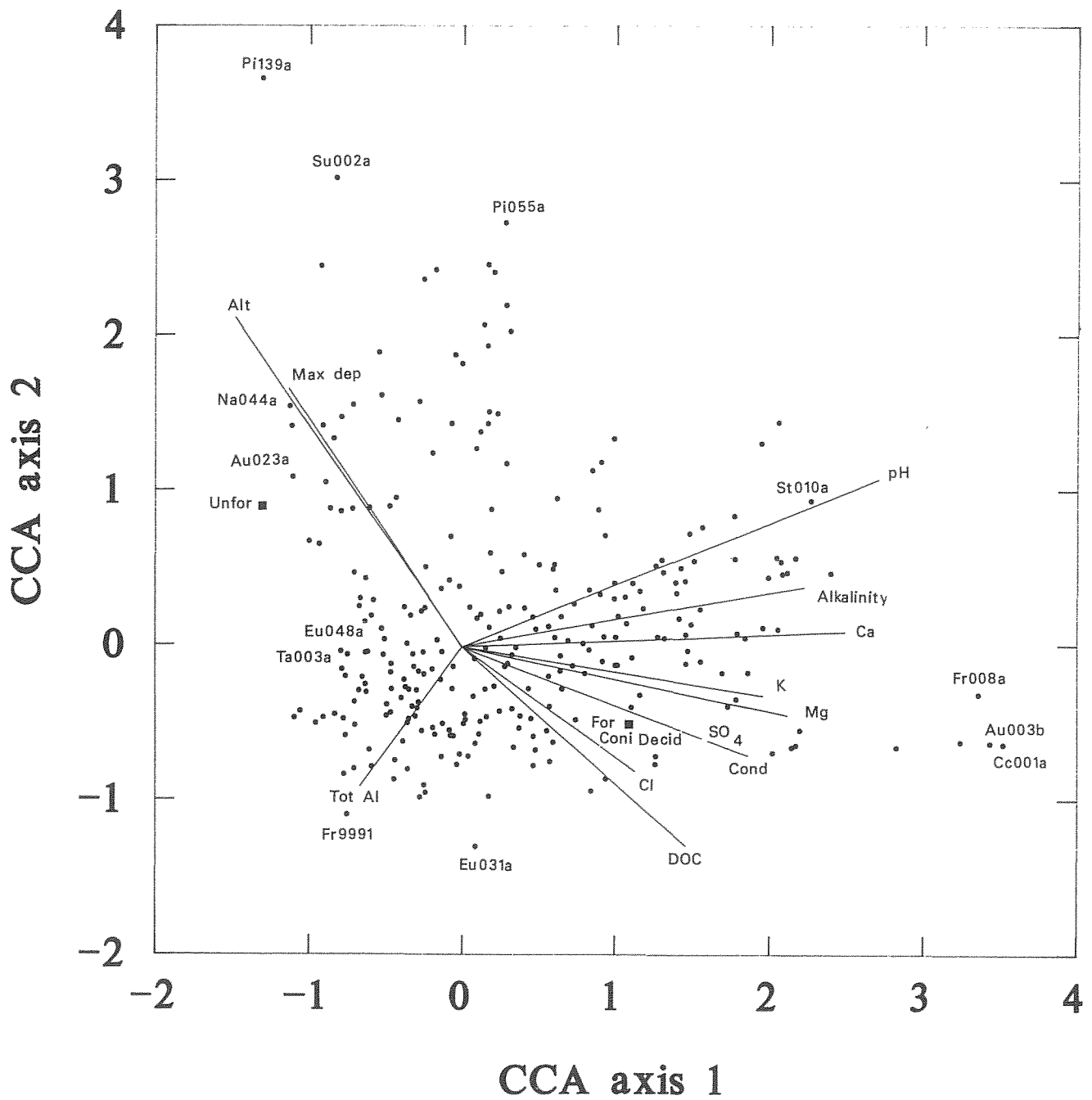


Figure 12 Canonical correspondence analysis (CCA) plot of all diatom taxa included in the CCA of the 138 samples on CCA axes 1 and 2 in relation to the 16 environmental variables included in the analysis. The diagram is constructed in the same way as Figure 11. Only selected taxa are labelled. See Appendix B for taxon codes. Abbreviations follow the caption for Figure 7.

Table 12 Weighted correlations between ordination axes and the environmental variables in the CCA of the 138 sample data-set. The species-environment correlations are 0.932, 0.847, 0.783 and 0.788, respectively.

Species axis 1	1.00									
Species axis 2	0.00	1.00								
Species axis 3	-0.02	0.02	1.00							
Species axis 4	-0.06	-0.03	-0.00	1.00						
Env. axis 1	0.93	0.00	0.00	0.00	1.00					
Env. axis 2	0.00	0.85	0.00	0.00	0.00	1.00				
Env. axis 3	0.00	0.00	0.78	0.00	0.00	0.00	1.00			
Env. axis 4	0.00	0.00	0.00	0.79	0.00	0.00	0.00	1.00		
pH	0.84	0.31	-0.09	-0.02	0.90	0.36	-0.11	-0.02	1.00	
Conductivity	0.58	-0.20	0.08	0.25	0.62	-0.23	0.10	0.32	0.37	1.00
DOC	0.45	-0.36	0.37	-0.30	0.48	-0.43	0.48	-0.38	0.22	0.40
Ca	0.78	0.03	0.29	0.07	0.84	0.03	0.38	0.09	0.70	0.70
Mg	0.66	-0.12	0.11	0.27	0.71	-0.14	0.14	0.34	0.49	0.93
K	0.61	-0.09	0.31	0.25	0.66	-0.10	0.40	0.31	0.50	0.74
SO4	0.56	-0.19	0.14	0.00	0.60	-0.23	0.18	0.00	0.33	0.80
Cl	0.32	-0.20	-0.17	0.35	0.34	-0.24	-0.22	0.45	0.17	0.86
Alkalinity	0.69	0.11	0.28	0.27	0.74	0.13	0.36	0.34	0.67	0.62
Total Al	-0.21	-0.25	0.19	0.09	-0.23	-0.30	0.24	0.12	-0.44	0.18
Altitude	-0.45	0.61	0.11	-0.09	-0.49	0.72	0.14	-0.11	-0.15	-0.58
Unforested	-0.32	0.21	-0.14	0.21	-0.35	0.25	-0.17	0.26	-0.14	-0.28
Forested	0.34	-0.20	0.12	-0.22	0.37	-0.23	0.15	-0.28	0.17	0.25
Conifer F.	0.33	-0.19	0.16	-0.26	0.35	-0.22	0.21	-0.33	0.14	0.25
Deciduous F.	0.05	-0.02	-0.16	0.13	0.05	-0.02	-0.21	0.17	0.09	0.00
Max. depth	-0.35	0.47	0.06	-0.17	-0.37	0.56	0.08	-0.21	-0.17	-0.42
	Spec	Spec	Spec	Spec	Env.	Env.	Env.	Env.	pH	Cond.
	axis 1	axis 2	axis 3	axis 4	axis 1	axis 2	axis 3	axis 4		
DOC	1.00									
Ca	0.49	1.00								
Mg	0.41	0.71	1.00							
K	0.47	0.76	0.76	1.00						
SO4	0.47	0.66	0.83	0.64	1.00					
Cl	0.12	0.32	0.79	0.52	0.56	1.00				
Alkalinity	0.31	0.91	0.61	0.70	0.41	0.31	1.00			
Total Al	0.30	-0.07	0.11	0.06	0.29	0.14	-0.14	1.00		
Altitude	-0.46	-0.36	-0.54	-0.44	-0.57	-0.50	-0.23	-0.25	1.00	
Unforested	-0.50	-0.32	-0.29	-0.23	-0.45	-0.06	-0.21	-0.35	0.50	1.00
Forested	0.51	0.33	0.28	0.23	0.43	0.04	0.22	0.28	-0.47	-0.98
Conifer F.	0.54	0.32	0.28	0.24	0.42	0.04	0.22	0.31	-0.43	-0.94
Deciduous F.	-0.12	0.02	-0.01	-0.05	0.02	0.02	0.00	-0.11	-0.14	-0.12
Max. depth	-0.28	-0.29	-0.38	-0.33	-0.36	-0.35	-0.23	-0.19	0.38	0.04
	DOC	Ca	Mg	K	SO4	Cl	Alk.	Tot. Al	Alt.	Unforest
Conifer F.	0.95	1.00								
Deciduous F.	0.13	-0.18	1.00							
Max. depth	-0.02	-0.01	-0.04							
	Forested	Conif.	Decid.							

(Env. = environmental, max. = maximum, F. = forest)

Table 13 Scores for samples and diatom taxa and environmental variable loadings for the CCA of the 138 sample data-set. Sample codes follow Table 4, species codes follows Appendix B.

	Axis 1	Axis 2	Axis 3	Axis 4		Axis 1	Axis 2	Axis 3	Axis 4
<i>Environmental variable loadings</i>					<i>Sample scores</i>				
pH	0.9034	0.3620	-0.1136	-0.0220	6.21	-1.2255	-0.0560	0.9862	0.0859
Conductivity	0.6216	-0.2325	0.1041	0.3228	65.21	0.4235	1.5593	-1.5094	-2.6249
DOC	0.4845	-0.4289	0.4753	-0.3758	80.11	-1.0089	0.0326	0.2277	0.3841
Ca	0.8375	0.0343	0.3754	0.0941	81.21	-1.1516	-1.3351	1.0141	0.0595
Mg	0.7076	-0.1440	0.1429	0.3446	82.11	-1.1343	-0.3871	1.0484	-0.3473
K	0.6559	-0.1027	0.3995	0.3137	83.11	-1.5118	2.9970	1.2716	-0.0912
SO4	0.6029	-0.2271	0.1779	0.0020	86.11	-0.9211	-0.4700	0.6921	-0.6041
Cl	0.3437	-0.2398	-0.2229	0.4452	87.21	-0.8612	0.4848	0.5990	-1.0980
Alkalinity	0.7441	0.1305	0.3579	0.3401	88.21	-1.5960	4.2539	1.8824	-0.8744
Total Al	-0.2286	-0.2970	0.2373	0.1187	89.11	-0.0140	2.4163	-0.5482	-1.2441
Altitude	-0.4868	0.7165	0.1412	-0.1146	9.11	-1.2212	-0.4524	1.2701	-0.2514
Unforested	-0.3466	0.2499	-0.1735	0.2639	ACH1	-0.4668	-1.0056	-0.3973	-0.2629
Forested	0.3684	-0.2327	0.1492	-0.2762	ARR1	0.5146	-0.0067	-1.2303	1.0377
Conifer F.	0.3498	-0.2239	0.2105	-0.3255	ARTH1	2.0141	0.3096	1.0324	1.6511
Deciduous F.	0.0523	-0.0234	-0.2071	0.1705	BARE1	1.6819	0.8016	-0.9070	0.1328
Max. depth	-0.3713	0.5600	0.0784	-0.2114	BARL1	1.0772	0.0491	-0.9091	0.8466
					BODG1	1.3430	1.5162	-2.1564	-2.5975
<i>Sample scores</i>					BODL1	-0.3949	-0.5468	-0.6654	0.2639
					BUGE1	-0.5194	-1.3916	-0.0129	0.2552
1.21	-1.1471	-0.5617	0.6353	0.3855	BYCH1	1.6687	0.6397	-1.3262	1.7077
10.21	0.1685	0.1936	-0.6498	-0.6172	CFYN1	-0.4280	-1.2173	-0.6664	-0.1417
111.11	0.1561	1.1350	-0.7459	-1.1987	CHN1	-0.2034	-0.4965	-0.8927	0.6296
113.21	0.1728	0.5094	-0.8235	-0.7099	CLON1	2.2385	0.8853	0.5155	2.6972
115.11	-0.4269	-0.3134	0.2246	-1.2361	CLYD1	0.5683	-0.3670	-1.4655	0.1425
12.11	-1.1415	1.8089	1.0289	-0.5172	CON1	-0.1584	-0.6425	-0.8268	-0.6600
15.11	-0.6783	0.0564	-0.0816	-0.2813	COR1	-1.5236	2.9395	1.1582	0.7675
17.21	-0.5435	-0.2769	0.1103	-0.8524	CWBY1	-0.8200	-0.5668	-0.4101	0.8955
18.11	-0.6896	2.5975	0.5552	-0.2183	DIWA1	0.0203	-0.1791	-1.0860	0.3516
19.21	-0.3891	2.4497	-0.1441	-2.2105	DOI1	-0.1735	-0.0830	-0.8156	0.8699
2.11	-0.7273	0.2427	0.1554	-0.2671	DOON1	0.3936	-0.3670	0.2425	0.6098
20.11	0.2888	2.8203	-0.1651	-1.4961	DUH1	-1.2242	1.0827	0.4939	1.0607
21.11	-0.2000	-0.0794	-0.3185	-0.4726	DUL1	-0.5851	-0.5873	-0.2583	0.4362
3.11	-1.1254	-1.1334	0.9413	0.1623	ENO1	-1.2261	-0.5873	0.5281	0.4753
34.11	-1.4422	2.7048	1.5552	-0.4943	EUN1	-1.4776	2.3548	0.9548	2.0384
4.11	-1.2384	-0.6001	1.8519	-0.2491	FERN1	2.5696	0.0915	4.5015	3.8387
42.11	-1.0566	0.6691	0.4842	-0.1637	FHI1	-1.1347	0.0719	-0.1302	2.1341
44.21	1.0556	2.5510	0.1396	-1.2088	FINL1	0.3854	-0.3361	-0.7594	0.4544
49.11	-0.7809	1.3701	0.2307	-0.4350	FLE1	-0.8873	-1.4799	-0.1210	0.2352
58.11	-1.1816	0.8461	0.8887	-0.2904	GARN1	1.1352	1.5006	-2.4793	-3.0029
59.11	-1.3152	3.9592	1.8019	-0.9078	GEIR1	1.2322	0.5953	-2.5188	0.9907

Table 13 Continued.

	Axis 1	Axis 2	Axis 3	Axis 4		Axis 1	Axis 2	Axis 3	Axis 4
<i>Sample scores</i>					<i>Sample scores</i>				
GLAS1					S191	0.1544	-1.2776	0.9721	-2.0814
GLFR1	2.2882	0.0246	1.2022	3.6458	S201	-0.3589	-0.8067	-0.0120	-0.6478
GLYN1	-1.0532	-0.2031	0.0000	0.6694	S21	3.2049	-0.6306	7.2128	1.7697
GOD1	0.4605	-1.0976	-1.2519	-0.8773	S221	-1.4422	-1.0326	1.4979	0.8276
GREENT1	-0.8819	-0.4223	0.0419	0.6327	S241	0.5333	0.0324	-0.0425	-1.3076
GWYN1	1.4296	0.7857	-2.0587	2.2169	S251	1.9078	1.0080	-0.2704	-1.8208
GYN1	-0.1273	-1.4752	-0.7044	-0.7741	S261	3.3389	-1.8518	8.6923	-5.3154
HARR1	-0.3355	-0.2233	-0.7051	0.4806	S271	-0.7847	-1.2494	0.6840	-0.5911
HOWI1	1.3018	1.2509	-2.1252	-1.3549	S281	-0.8558	-1.4173	0.6950	-0.2853
INVA1	1.7270	0.4914	0.3179	2.0218	S291	-0.1962	-0.7128	0.6257	-0.2689
IRD1	0.1282	-0.2612	-0.9386	-0.3910	S301	-0.0349	-0.6498	0.0919	-0.9483
LAG1	-1.0781	-1.7110	-0.2040	0.7220	S31	2.2321	0.6759	1.9231	0.1556
LAI1	-0.3487	-0.5441	-0.5827	-0.2191	S311	0.0043	-0.5514	0.0070	-0.9578
LAR1	-0.5441	-1.0894	-0.5702	0.4182	S41	0.6489	-1.6628	1.4248	-2.8840
LCSL1	-0.9772	0.3340	-0.0239	1.9584	S51	2.6145	-0.5579	4.2589	-0.4855
LCSLU1	-1.0640	1.0108	-0.2573	3.4904	S61	0.6436	-0.9340	-0.3674	-1.2659
LDE1	-0.0451	-0.0807	-1.0411	0.3703	S71	1.4897	0.3028	0.6374	-2.7576
LENY1	0.8174	0.1820	-1.6711	0.2654	S81	1.0228	0.1024	1.4228	-1.0042
LGR1	-1.0373	-1.1337	0.0488	0.7541	S91	1.3904	-0.7665	1.4638	-1.9251
LLDU1	0.9098	-0.3335	-1.2841	0.9025	SCOATT1	-1.4821	1.9646	1.1985	1.1321
LLGH1	-0.6266	-1.1489	-0.3428	0.1284	SKE1	0.3369	-0.8150	-0.8077	-0.6101
LOD1	-0.6671	-1.0323	-0.4304	0.4063	STRO1	-0.4286	-0.4988	-0.4410	0.2891
LOWT1	-0.9183	-0.5482	-0.2469	1.2544	TANN1	-1.2181	-1.1639	0.6951	0.9731
MABE1	0.6508	-0.9521	0.0222	-0.3539	TECW1	0.3855	-0.7076	-1.5762	0.1959
MACA1	-0.2307	-0.7373	-0.7929	-0.0048	TINK1	0.3329	-0.6169	-1.4092	-0.4088
MANN1	1.7710	0.8694	-1.3416	3.4906	TROO1	-0.7679	-0.5100	-0.1374	0.4022
MINN1	0.3192	-0.2579	-1.0182	0.2420	UAI1	-1.0791	1.8543	0.1875	0.9741
MOAN1	0.9592	-1.4696	0.3610	-0.5984	UIS1	0.8414	0.1170	-1.0813	1.2269
MUCK1	0.2229	-0.6128	-0.4098	0.3991	VAL1	-1.0127	-1.0700	0.1326	0.4675
NAGA1	-1.2566	1.8234	0.7883	1.6440	WOOD1	2.2146	0.3460	1.6022	3.2482
OCHI1	0.8494	-0.7185	-0.1555	0.2797	YBI1	-1.4369	-1.6170	0.6667	1.9031
PARC1	0.9579	0.3835	-2.2672	0.5901	YGAD1	0.0760	-0.7692	-1.3832	-0.0943
PENR1	-0.8996	-1.5666	0.0912	0.3248					
RIEC1	0.2671	-0.2951	-0.4266	0.4924	<i>Species scores</i>				
RLGH1	-0.9942	-1.3425	0.0509	0.4663	AC001A	1.6899	-0.1795	1.5185	1.5867
S101	1.1769	0.2943	-0.3208	0.0906	AC002A	1.3914	0.3388	-0.1467	1.1984
S11	2.1663	-0.4443	2.7228	-0.7423	AC004A	0.789	0.0113	-0.9852	-0.2666
S111	0.3536	-0.7962	0.2809	-1.2691	AC013A	0.9944	0.3003	-0.4687	0.5173
S121	1.3345	1.1996	-0.8224	-2.5011	AC014A	-0.7764	-0.8351	-0.147	0.0952
S131	0.1261	-1.4548	0.6285	-1.5302	AC014B	-0.6728	0.2492	-0.0227	0.4389
S141	1.0874	0.4129	0.6541	-1.5422	AC014C	-0.8404	1.3334	0.3423	-0.1842
S151	1.9793	1.3815	0.8897	1.2983	AC017A	-0.4366	0.9487	-0.2655	-0.2601
S161	2.1517	-0.5117	3.2695	-0.8750	AC018A	0.8966	0.3299	0.7267	-0.5418
S171	0.7911	0.3159	-0.2482	-1.9843	AC019A	-0.0461	1.8739	-0.1859	-0.675
S181	0.5154	-1.3095	-0.0136	-2.0656					

Table 13 Continued.

	Axis 1	Axis 2	Axis 3	Axis 4		Axis 1	Axis 2	Axis 3	Axis 4
<i>Species scores</i>					<i>Species scores</i>				
AC022A	-0.8934	1.0497	0.2627	0.1962	AU9983	1.4653	-0.0343	1.0601	0.1300
AC025A	1.0026	0.0521	0.1306	-0.2750	AU9984	0.4651	-0.7769	0.9543	-1.2948
AC028A	-0.0873	0.4214	-0.5534	0.5572	AU9987	0.4662	-0.5886	0.6692	-0.5432
AC029A	-0.1956	1.2400	0.0517	-0.0506	AU9988	-0.6335	-0.2560	0.0643	0.3239
AC030A	0.7973	-0.1837	-0.1701	0.8590	BR001A	0.4828	0.1031	-0.5508	-0.0190
AC034A	0.1173	1.3782	-0.2199	-0.2014	BR003A	-0.7787	-0.4764	0.5013	0.1473
AC035A	1.1584	-0.3236	0.3864	-0.6356	BR004A	-0.1307	-0.5056	0.0891	-0.3544
AC039A	0.8483	1.1260	0.1301	-0.5480	BR005A	1.4196	0.5004	0.9003	0.2126
AC042A	0.4621	0.1834	-0.7294	0.1896	BR006A	-0.2876	-0.1699	-0.0470	0.0521
AC044A	-0.7153	1.5517	0.3013	-0.4542	BR9997	0.0815	-0.6336	0.1409	-0.2974
AC046A	0.2070	-0.2669	-0.2009	-0.1041	CA018A	1.2971	0.5499	0.5456	0.0118
AC048A	-0.2426	0.5060	-0.4113	0.2841	CC001A	3.5205	-0.6385	5.7894	2.5756
AC9964	1.0190	0.1866	-1.6164	0.2052	CM004A	1.0637	0.3120	-0.1595	0.5018
AC9965	-0.5441	1.8883	-0.0912	0.2523	CM010A	-0.3728	-0.2709	-0.2461	0.1346
AC9968	-0.8644	0.8802	0.2628	0.7370	CM013A	1.5485	0.2330	-0.2750	0.3461
AC9969	0.6456	0.1851	-1.2039	0.0431	CM014A	-0.6358	0.1504	-0.2602	0.8818
AC9975	-0.7954	-0.0413	-0.0414	0.4196	CM015A	0.9905	1.3333	0.5498	0.0402
AC9996	0.4068	0.2414	-0.9895	0.2815	CM015B	0.1895	0.8767	-0.5612	0.4271
AM001A	0.2296	1.4951	-0.5444	-0.6865	CM017A	-0.6678	0.3027	0.1392	-0.0302
AM001B	2.0718	0.5413	0.2810	1.9810	CM020A	0.1751	0.1159	-0.3104	0.1169
AM001D	1.2727	0.0521	0.3781	0.3339	CM031A	0.6091	0.3567	0.1409	0.4928
AS001A	1.9559	0.1141	1.2662	0.9593	CM031C	-0.7942	0.8607	0.5024	-0.2224
AS003A	-0.3820	-0.2260	0.1476	0.0699	CM038A	1.7701	0.8351	-0.4588	-0.9251
AT001A	-0.2538	-0.9039	0.6460	-0.6040	CM048A	0.2781	-0.1382	-0.3317	-0.1613
AU001C	0.2428	-0.4280	0.8089	-0.7820	CM050A	1.9456	1.3046	2.5753	0.8923
AU002A	2.0185	-0.6951	2.3578	-1.5464	CM051A	0.7304	0.2687	0.1119	-0.2407
AU003B	3.4368	-0.6324	5.5688	2.3286	CM052A	1.1110	0.4021	0.1356	0.2407
AU003D	2.8253	-0.6595	4.2602	1.4549	CM101B	1.5050	0.5467	1.7817	0.1523
AU004A	-0.1396	0.3659	0.6744	-0.8001	CM9989	-0.2803	-0.9832	0.6341	-0.3240
AU004B	0.6408	-0.1681	0.2916	-0.5637	CM9995	-0.6088	-0.6766	-0.6281	0.6764
AU004C	1.0122	-0.6119	0.0806	-0.7998	CO001A	2.0440	0.5675	1.5978	2.2770
AU004D	-0.3524	-0.4759	-0.4553	0.3050	CO001B	1.8583	-0.1782	0.8582	2.5966
AU005A	-0.9096	1.4167	0.5598	-0.5133	CY001A	1.7893	0.0776	0.8165	-0.0609
AU005B	1.2562	-0.7694	0.9576	-1.4014	CY002A	2.0824	0.4612	2.6205	2.7430
AU005D	0.8397	-0.9422	1.0091	-1.3301	CY003A	2.1145	0.4705	2.9871	2.8811
AU005E	-0.9348	0.6500	0.5300	0.0533	CY004A	1.1053	-0.4011	0.6230	-0.7413
AU005J	0.5691	-0.3965	-1.4348	0.6434	CY007A	1.4767	0.7214	0.9250	-0.4488
AU005L	2.1700	-0.6461	0.5880	-1.4765	CY010A	1.3850	0.4074	-0.4482	-0.8744
AU009A	0.9369	-0.8632	-0.1322	-0.2217	CY9991	0.6017	0.5226	-0.4854	-0.6902
AU009B	0.1715	-0.9776	1.2412	-1.1978	DE001A	1.4473	0.4164	-0.3033	0.5709
AU010A	-0.1366	-0.7201	0.0342	-0.2616	DT002A	-0.0771	0.7017	-0.7876	-0.1078
AU010B	-0.2670	-0.5542	0.3032	-0.0429	DT003A	1.1811	0.2373	-0.3030	1.7729
AU014A	0.4788	-0.6774	-0.0070	-0.8122	DT004B	1.9895	0.4393	2.4925	2.6469
AU022A	3.2431	-0.6231	5.3677	2.1436	EU002A	-0.0600	-0.1388	-0.0377	-0.0259
AU023A	-1.1026	1.0837	0.2047	1.3545	EU002B	-0.1970	-0.1557	-0.0692	0.0781

Table 13 Continued.

	Axis 1	Axis 2	Axis 3	Axis 4		Axis 1	Axis 2	Axis 3	Axis 4
<i>Species scores</i>					<i>Species scores</i>				
EU002D	-0.0867	-0.5482	-0.2298	0.0210	FR006A	1.2592	0.5120	0.7383	-0.2438
EU002E	0.1141	-0.4925	-0.3146	0.2006	FR007A	1.1568	0.3497	0.3446	1.0980
EU002K	-0.2968	-0.4065	-0.0457	0.0636	FR008A	3.3606	-0.3173	4.9776	2.9053
EU003A	-0.3385	0.1925	0.5987	0.2130	FR009F	0.4035	0.5858	-1.0782	0.2434
EU004A	-0.5271	0.1042	0.1646	-0.1855	FR010A	0.1059	-0.5736	0.2689	0.0103
EU009A	-0.5928	0.1885	0.0626	0.0118	FR011A	0.7202	-0.1348	0.2391	-1.0642
EU009C	-0.2451	-0.9517	0.6651	-0.3686	FR015A	0.3354	-0.6579	1.0746	-0.9732
EU011A	-0.3908	-0.6246	-0.2106	0.1964	FR018A	1.7816	-0.3498	0.9768	0.7719
EU013A	0.0956	0.1737	-0.3700	0.3136	FR9991	-0.7581	-1.0974	0.5035	-0.0288
EU014A	-1.0547	-0.4261	0.7125	0.0615	FU002A	-0.3516	-0.2884	0.2166	-0.1877
EU015A	-0.6335	-0.0508	-0.1342	0.5919	FU002B	-0.3009	-0.2927	0.0823	-0.2021
EU016A	-0.3102	-0.4595	0.0187	-0.1364	FU002F	-0.1804	-0.5761	-0.0184	-0.1407
EU017A	0.2507	0.0443	-0.3995	-0.1606	GO003A	0.4344	-0.2829	-0.2046	0.1926
EU019A	-0.1941	-0.5327	0.1597	-0.2360	GO004A	0.3255	-0.0642	-0.4147	0.2290
EU020A	-0.5081	0.0356	0.4325	-0.6128	GO006C	0.9164	-0.1068	0.1295	0.1122
EU021A	-0.4224	1.4526	0.0905	-0.4890	GO013A	0.9262	0.0531	0.1584	1.1142
EU022A	-0.4442	-0.7451	-0.8003	0.1016	GO023A	1.3064	0.4697	0.4130	1.5054
EU025A	-0.7041	-0.5174	0.3829	0.3562	GO025B	1.4884	0.1364	-0.1399	0.3359
EU027A	-1.0008	0.6698	0.4752	-0.2126	GO025F	2.1683	0.5641	1.4358	2.9968
EU028A	-0.7668	-0.2033	0.7263	-0.4168	GY005A	2.1432	-0.6588	2.8632	1.1976
EU028B	-0.9582	-0.5036	0.5717	-0.1383	HN001A	0.6079	0.0490	-0.1574	0.9380
EU031A	0.0822	-1.3041	1.2257	-1.0257	ME019A	0.3717	-0.5344	0.2359	-0.0290
EU034A	-0.3601	-0.8009	0.8198	-0.5964	NA002A	0.3793	-0.4572	0.1487	-0.3683
EU039A	-0.7099	0.4672	0.1177	-0.0696	NA003A	0.9933	0.4016	1.0317	0.3748
EU040A	-0.3633	-0.5010	0.0734	0.0537	NA003B	0.6915	0.0290	0.2571	-0.0399
EU046C	-0.7876	-0.1582	0.2058	0.0038	NA005A	0.5923	0.4918	0.1966	-0.3140
EU047A	-0.4962	-0.3850	-0.0425	0.1131	NA005B	1.4538	-0.1194	0.4481	0.1047
EU048A	-0.6280	-0.3058	0.3789	-0.1044	NA006A	-0.2562	-0.0458	0.0012	-0.1402
EU049A	0.1568	-0.4634	0.0272	0.0802	NA006B	-0.4642	-0.2859	-0.0431	0.7012
EU049B	0.0357	-0.7171	0.4150	-0.3418	NA007A	1.4543	0.0694	0.1552	1.4593
EU051A	-0.3995	-0.3453	-0.0462	-0.0689	NA008A	1.8411	0.0471	1.0247	1.4825
EU051B	-0.0663	-0.2831	0.1815	0.0282	NA013A	0.8272	0.3576	0.0654	0.4566
EU056A	-0.2508	-0.1896	-0.0086	0.3938	NA014A	1.0751	0.1401	0.5760	0.1504
EU057A	-0.3611	0.0101	-0.4935	0.5540	NA015A	-0.2445	0.2417	0.0518	-0.4004
EU058A	-1.0916	-0.4690	0.2542	0.1260	NA016A	0.5924	-0.6264	0.1131	-0.4906
EU9961	-0.4984	-0.4568	-0.4172	0.4457	NA032A	-0.0222	0.3806	-0.3645	-0.3483
EU9962	0.0062	-0.5069	-1.1328	0.0293	NA033A	-0.3186	-0.1398	-0.5263	0.5557
EU9965	-0.7097	-0.7981	-0.1753	0.1021	NA037A	0.0780	-0.0876	-0.6700	0.3761
EU9969	-0.7059	-0.3684	-0.1813	0.6706	NA038A	-0.0392	-0.7703	0.8130	-0.5966
FR001A	1.1103	-0.0821	0.2666	0.6491	NA042A	0.1649	1.4311	0.0367	-0.7125
FR001B	-0.0020	1.8173	-0.2127	-1.0213	NA043A	0.2037	2.4077	-0.0389	-0.9120
FR002A	2.1966	-0.5489	2.6001	-0.4854	NA044A	-1.1233	1.5436	0.5569	-0.1500
FR002C	0.9102	-0.4566	0.3945	-0.0793	NA045A	0.4480	-0.4734	-0.0875	0.2308
FR005A	0.2446	0.2192	0.1928	1.0149	NA046A	0.1526	-0.0190	-0.9131	0.5302
FR005D	0.3249	-0.4105	-0.2088	-0.2903	NA048A	-0.6533	-0.2098	0.1917	0.0226

Table 13 Continued.

	Axis 1	Axis 2	Axis 3	Axis 4		Axis 1	Axis 2	Axis 3	Axis 4
<i>Species scores</i>					<i>Species scores</i>				
NA063A	1.7726	0.5586	0.7394	2.5096	OP001A	1.7273	-0.3946	1.7120	-1.2102
NA068A	0.5721	-0.7508	0.0454	0.0131	PE002A	-0.3250	-0.0590	-0.2476	0.1355
NA084A	0.1647	2.4581	-0.2824	-1.2409	PI005A	-0.0214	-0.7035	0.5575	-0.1703
NA086A	1.3144	0.0432	0.0488	1.5890	PI007A	0.0135	-0.4462	0.0205	-0.0739
NA099A	-0.3307	-0.3958	-0.3238	0.5438	PI011A	-0.2679	0.2199	0.1632	-0.1407
NA101A	1.5591	0.7624	2.7641	0.9421	PI014A	-0.0721	1.4322	-0.2448	-0.6244
NA102A	0.9071	1.1808	0.5036	0.2663	PI015A	0.3488	-0.0142	0.0993	-0.1741
NA112D	0.1379	2.0673	0.0546	-0.9077	PI016A	0.0894	1.2693	-0.3503	-0.5264
NA113A	0.2868	1.1740	0.2575	-1.0593	PI018A	-0.3811	0.2447	0.3234	-0.1105
NA114A	0.1647	2.4581	-0.2824	-1.2409	PI018B	-1.1096	1.4139	0.6403	-0.2014
NA115A	-0.4493	-0.8688	-0.4627	0.4594	PI022B	-0.4720	-0.1788	-0.1012	0.0612
NA129A	-0.1786	2.4249	0.0343	-0.8883	PI023A	-0.0808	-0.5824	-0.1682	0.3830
NA133A	0.1186	0.2004	-0.7875	0.3057	PI055A	0.2775	2.7291	0.0128	-0.5398
NA135A	-0.4686	-0.1237	-0.1459	0.9722	PI056A	-0.0605	-0.5861	0.9291	-1.0021
NA140A	-0.6777	-0.2944	-0.1395	0.6951	PI139A	-1.3046	3.6570	0.9759	-1.2352
NA149A	0.1797	0.5963	-0.7161	-0.0130	PI164A	-0.4804	0.8943	0.0464	-0.5723
NA151A	0.3135	2.0253	0.2249	-0.6206	RH006B	0.1646	1.9319	0.1377	-0.5807
NA156A	-0.4718	-0.4372	-0.1554	0.1721	SA001A	0.0462	0.2461	-0.0839	-0.4066
NA158A	-0.5952	-0.7841	0.2172	0.0977	SA001B	0.3036	0.2504	-0.2850	-0.2457
NA160A	-0.7917	1.4718	0.2943	-0.2962	SA006A	0.4996	0.5214	0.1635	0.3099
NA167A	-0.7538	-0.0627	0.4082	-0.1466	SA042A	-0.6356	0.4317	0.2113	-0.5206
NA170A	-0.2571	2.3630	-0.0496	-0.5952	SE001A	-0.8347	-0.4466	0.5961	0.0681
NA9904	1.2586	-0.7176	0.9303	-1.7391	SP002A	0.0164	-0.4808	0.2002	-0.4488
NA9919	0.6526	-0.2838	0.6052	-0.4190	ST004A	2.3952	0.4667	0.7932	2.4064
NA9955	0.6438	-0.0700	-0.7378	-0.0138	ST010A	2.2691	0.9322	-0.0762	0.8913
NA9963	0.0714	-0.3313	-0.8740	0.2590	SU002A	-0.8256	3.0145	1.1845	-0.9072
NA9964	-0.2876	-0.3691	-0.3745	0.2331	SU004A	-0.7229	0.8787	0.0616	0.6032
NE003A	-0.1637	0.0333	-0.1667	-0.1041	SU005A	-0.1439	-0.2227	0.5034	-0.5068
NE003B	-0.5156	-0.0650	0.3166	-0.3255	SU006A	-0.6193	-0.0464	0.2716	-0.0799
NE003C	0.5523	-0.5526	0.5049	-1.2488	SY002A	0.8466	0.1270	0.3422	-0.2710
NE004A	-0.5806	0.2897	0.2517	-0.1128	SY003A	1.5476	-0.1047	0.7676	-0.1523
NE012A	-0.6081	0.8839	-0.1527	0.2482	SY004A	1.4098	0.1741	1.5038	-0.0385
NE020A	-0.9238	2.4475	0.5267	-0.8178	SY009A	1.0141	-0.1305	0.1470	-1.1588
NE023A	-0.5330	1.6142	0.0895	-0.4764	SY010A	0.8277	-0.0310	0.0631	1.0540
NI002A	0.8847	0.8737	0.1900	0.4981	SY013A	2.0547	0.1031	1.5071	1.6104
NI005A	0.1481	-0.2851	-0.4731	0.0012	SY043A	0.2572	0.4764	0.1217	-1.2400
NI008A	0.6191	0.9456	-0.2453	-0.1595	TA001A	-0.1330	-0.0427	-0.1354	-0.0701
NI009A	0.9314	0.7083	0.4151	-0.0832	TA002A	0.5643	-0.2020	-0.0689	0.8093
NI009B	0.1734	1.5078	-0.2178	-0.8509	TA003A	-0.9081	-0.4661	0.3138	0.3646
NI017A	0.7381	-0.4811	0.4191	-1.2318	TA004A	-0.7676	-0.5835	0.0316	0.1539
NI021A	0.5666	0.1211	0.2759	-0.9816	TA9996	1.0004	-0.1313	0.4291	0.2097
NI026A	0.2981	-0.1201	-0.2399	0.8647					
NI027A	0.2840	2.1940	-0.3789	-1.0883					
NI152A	-0.2825	1.5753	0.0713	-0.4663					
NI9984	2.0570	1.4420	3.9939	1.4340					

(88.2) are located on the left of axis 1 and alkaline sites such as Edasjön (S2), Loch Fern (FERN), and Hagsjön (S26) are located on the right-hand side of axis 1. Acid-tolerant diatoms such as *Navicula krasskei* (Na044a on Figure 12), *Aulacoseira tethera* (Au023a), *Tabellaria binalis* (Ta003a), and *Eunotia naegelii* (Eu048a) occur with negative scores on the left side of CCA axis 1, whereas alkaline-tolerant diatoms such as *Cyclostephanus dubius* (Cc001a), *Aulacoseira granulata* var. *angustissima* (Au003b), *Fragilaria crotonensis* (Fr008a), and *Stephanodiscus parvus* (St010a) are positioned on the right-hand side with positive scores.

Axis 2 is more difficult to interpret. Altitude ($r = 0.61$), maximum depth (0.47), total aluminium (-0.25), DOC (-0.36), and pH (0.31) are the only variables with large correlations with axis 2. The length of the total Al arrow is, however, short. Forward selection and associated Monte Carlo unrestricted permutation tests (99 permutations) of the significance of the environmental variables (ter Braak 1990) suggest that all environmental variables except conductivity, calcium, and alkalinity make significant ($p < 0.05$) contributions to explaining the underlying variation in the diatom assemblages. Axis 3 is correlated with DOC ($r = 0.37$), calcium (0.29), potassium (0.31), and alkalinity (0.28), whereas axis 4 is positively associated with chloride (0.35), conductivity (0.25), magnesium (0.27), potassium (0.25), and alkalinity (0.27) and negatively associated with coniferous forest (-0.26) and DOC (-0.30). No biological interpretations of these axes are presented here.

RECONSTRUCTION OF LAKE SURFACE-WATER CHEMISTRY FROM DIATOM ASSEMBLAGES

The principal reason for developing the SWAP diatom-lake chemistry data-set was to provide a basis for the quantitative reconstruction of lake pH and other chemical variables from fossil diatom assemblages.

Diatoms have long been known to be sensitive ecological indicators of certain lake chemical variables. This feature has been extensively exploited in the last decade to reconstruct pH from fossil assemblages (e.g. Battarbee 1984), and a range of numerical procedures has been developed for quantitative inference of pH. Such reconstructions, in practice, consist of two mathematical stages.

First, the responses of modern diatoms to contemporary lake chemistry are modelled by regression. This involves a modern training set of diatom assemblages ("responses") from surficial sediment samples with associated lake chemical data ("predictors"). Second, these modelled responses are used

to infer past chemistry from the composition of fossil assemblages by means of calibration.

Nearly all the numerical procedures used are variants of the basic multiple linear regression model and usually involve grouping diatoms into ecological or chemical categories. Within SWAP we have used the two procedures developed by ter Braak and van Dam (1989), namely maximum likelihood (ML) and weighted averaging (WA) regression and calibration because they are more sound theoretically and perform better than other, more widely used "ad hoc" reconstruction procedures (e.g. Battarbee 1984). Full details of the ML and WA procedures used in SWAP are given by Birks *et al.* (1990a). In a comparison of the two procedures using the SWAP training set and pH, the computationally simple but heuristic approach of WA gave superior results in terms of a lower root-mean-square error of prediction (RMSE) in cross-validation than the computationally demanding but formal statistical approach of ML Gaussian logit regression and ML calibration (Birks *et al.* 1990a; see also ter Braak and van Dam, 1989). For this reason, we used simple WA regression and calibration to infer separately pH, total Al, and DOC using the SWAP training set (Birks *et al.* 1990a, 1990b) implemented by the computer programs WACALIB 2.1 (Line and Birks 1990) and 3.0 written by J.M. Line in conjunction with C.J.F. ter Braak and H.J.B. Birks.

In any large heterogeneous data-set such as the SWAP set, it is inevitable that some samples are "rogues" or atypical observations, for example with unusual diatom assemblages poorly related to chemistry or with poor or unreliable chemical data. It was therefore necessary to screen the data numerically to detect potential "rogues". The procedures used for data-screening are described by Birks *et al.* (1990a). Using these procedures, the 178-sample pH set was reduced to a training set of 167 samples, the 157-sample total Al set was reduced to a training set of 126 samples, and the 138-sample DOC set was reduced to a training set of 123 samples. Details of the three training sets after data screening are given in Table 14, along with lists of the samples deleted.

Simple WA without tolerance downweighting was used as it consistently gives lower RMSE as estimated by bootstrapping (Birks *et al.* 1990a). An inverse deshrinking regression was used to deshrink the inferred DOC and total Al values because it minimises RMSE in the training set. Classical regression was used to deshrink the inferred pH values because it takes inferred values further away from the mean and in these data the mean lies in the pH interval where lake pH is most variable, and because in acidification studies we require our reconstructions to be most precise at the lower end of the pH range in the training set (Birks *et al.* 1990a).

Root-mean-square of the error ($\hat{x}_i - x_i$) was calculated for each training set, where x_i is the observed

value of chemical variable x in sample i (Table 15). The correlation (r) between \hat{x}_i and x_i was also calculated (Table 15). Plots of x_i and of $(\hat{x}_i - x_i)$ against x_i for pH, DOC, and total Al are shown in Figure 13. As the apparent RMSE is invariably an underestimate when based solely on training sets (ter Braak and van Dam, 1989), bootstrapping was used to estimate a more realistic RMSE of prediction for each training set (Table 15). This is partitioned into two components, s_{i1} , that part of the prediction error due to estimation error in the taxon parameter used in WA, namely its WA optimum, and s_2 , that part of the error due to imperfections in the calibration function, even if the optima are known without error. Diatom assemblages vary even among lakes with the same chemistry or, conversely, because lakes with the same diatom assemblage may differ chemically. Component s_2 also contains any model specification errors (see Birks *et al.* 1990a).

The s_{i1} component is small compared with the s_2 component (Table 15). The ratio of s_{i1}/s_2 is 0.23 (pH), 0.27 (total Al), and 0.26 (DOC). The training sets are thus adequate to yield reliable estimates of taxon optima by WA regression required for WA calibration. As noted by Birks *et al.* (1990b) there is, however, some trend in the differences $(\hat{x}_i - x_i)$ plotted against x_i for DOC and Al (Figure 13). Inferences are too high at low observed total Al or DOC and too low at high total Al or DOC, probably because of functional interactions between pH, Al, and DOC. Such interactions are not accounted for in WA reconstructions of individual variables. Multivariate calibration procedures that permit inference of more than one environmental variable simultaneously are required to allow for confounding and interacting environmental variables.

The pH, total Al, and DOC optima and tolerances of the diatom taxa in the training sets, as estimated by WA regression are tabulated in Appendix B, along with the number of occurrences of each taxon in the training sets. These optima plus the regression equations listed in Table 16 provide the basis for reconstructing pH, total Al, and DOC for any fossil diatom assemblage with identical taxonomy to the SWAP training sets.

Table 14 Details of the three SWAP training sets before and after data screening (sample numbers follow Table 4)

Training set	Number of samples prior to screening	Number of samples after screening	Range	Mean	Median	Standard deviation
pH	178	167	4.23-7.25	5.56	5.27	0.77
Al ($\mu\text{g l}^{-1}$)	157	126	13.0-256.3	103.2	104.3	60.04
DOC(mg l^{-1})	138	123	0.12-11.60	3.17	2.50	2.29

Training set samples before screening

Training set	Samples
pH	All samples
Al	1, 2, 4-7, 9-11, 13-15, 17-18, 20, 22-29, 32, 35-55, 56-63, 65-82, 84-87, 90-103, 106-123, 125-137, 139-162, 164-170, 172-178.
DOC	1, 2, 4-7, 9-11, 13-15, 17, 18, 20, 22-25, 27-29, 32, 36-49, 51, 52, 54, 56-63, 65-82, 84-86, 90, 92, 93, 95-100, 102, 103, 103-121, 125-137, 139-155, 158, 161, 162, 164-168, 170, 175, 177, 178.

Training set samples deleted after data screening

Training set	Samples
pH	4, 27, 72, 79, 90, 91, 114, 124, 150, 157, 177
Al	1, 2, 26, 27, 32, 45, 53, 57, 74, 86, 90, 96, 108, 109, 110, 112, 113, 114, 115, 127, 135, 139, 143, 151, 156, 158, 159, 161, 166, 172, 177.
DOC	72, 110, 115, 117, 127, 135, 137, 142, 143, 146, 148, 150, 151, 152, 154.

Table 15 Apparent root-mean-square error of prediction (RMSE), correlation (r) between observed and inferred chemistry and RMSE estimated by bootstrapping ($\text{RMSE}_{\text{boot}}$) and its error components (s_{11} , s_2) for the three screened SWAP training sets.

Training set	Apparent RMSE	r	RMSE boot	RMSE s_{11}	RMSE s_2
pH	0.297	0.933	0.320	0.072	0.312
Al ($\mu\text{g l}^{-1}$)	37.820	0.777	49.663	12.985	47.935
DOC (mg l^{-1})	1.251	0.837	1.580	0.403	1.527

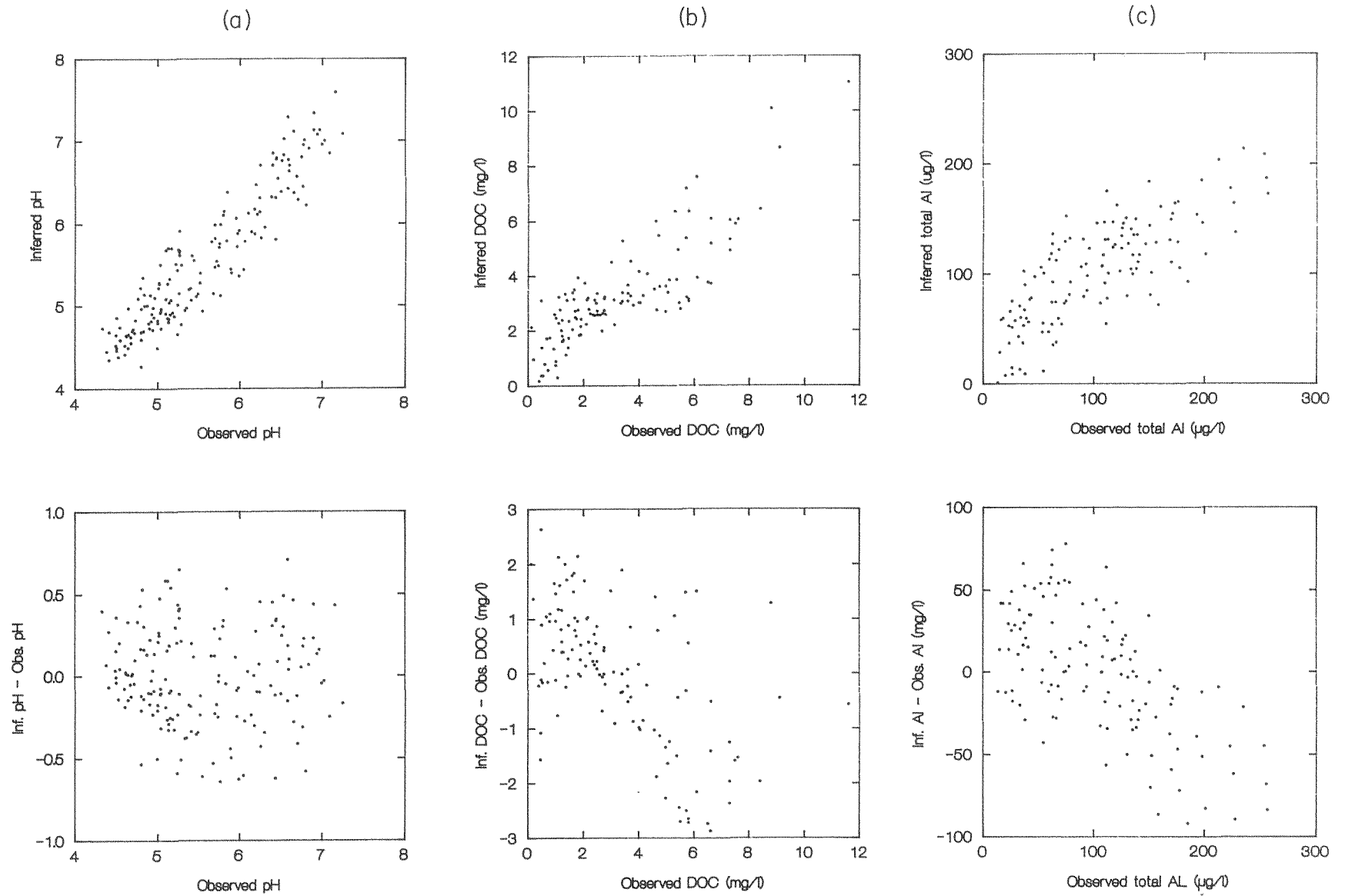


Figure 13 Plots of (a) inferred pH and of the differences between inferred and observed pH against observed pH based on weighted averaging regression and calibration using the screened SWAP pH training set ($n = 167$), (b) inferred DOC and of the differences between inferred and observed DOC against observed DOC based on weighted averaging regression and calibration using the screened SWAP DOC training set ($n = 123$), and (c) inferred total Al and of the differences between inferred and observed total Al against observed total Al based on weighted averaging regression and calibration using the screened SWAP total Al training set ($n = 126$). Abbreviations: *inf* = inferred, *obs* = observed.

Table 16 *Deshrinking regression equations for the three screened SWAP training sets*

	Constant	Coefficient	Regression type
	a	b	
pH	2.630	0.5264	Classical ¹
Al	-151.2	2.461	Inverse ²
DOC	-3.775	2.202	Inverse ²

¹Classical regression $Final\ x_i = (Initial\ x_i - a)/b$

²Inverse regression $Final\ x_i = a + b * Initial\ x_i$

$$Initial\ \hat{x}_i = \frac{\sum_{k=1}^m y_{ik} \hat{u}_k}{\sum_{k=1}^m y_{ik}}$$

where y_{ik} is the abundance of taxon k in fossil sample i ($y_{ik} \geq 0$); u_k is the WA estimate of the optimum of taxon k ; x is the chemical variable to be reconstructed; \hat{x}_i is the inferred value of the chemical variable x for the sample i ($k = 1, \dots, m$ diatom taxa).

DIATOM INDICATOR TAXA

Within a training set such as the SWAP set, some taxa may be sensitive to particular chemical variables, whereas others may be insensitive. Taxa that are sensitive, that have narrow ecological amplitudes ("tolerances") for that variable, and that are frequent may be good "indicator" taxa for particular chemical components. The approach we have adopted in drawing up lists of potential indicator taxa for pH, DOC, and total Al is as follows.

(1) Do separate canonical correspondence analyses of the three screened training sets (Table 14) with pH, Al, and DOC as the sole environmental variables for the three individual sets. This provides a direct gradient analysis of the taxa along the gradient of the single chemical variable of interest.

(2) Examine the fraction of the variance of a taxon fitted by this gradient (ter Braak 1990), namely CCA axis 1. For a taxon to be considered as a potential indicator, we require that 10% or more of a taxon's variance be explained by the chemical variable of interest. Using this criterion (A), there are 31 potential indicators for pH, 8 for Al, and 13 for DOC. The mathematical basis for estimating the fraction of the variance of a taxon fitted or the taxon's relative contribution or the contribution of dimensions to the inertia of the species is given by ter Braak (1990).

(3) Estimate the taxon's tolerance, as estimated by weighted averaging, for the chemical variables of interest, with each tolerance adjusted for the effective number of occurrences of the taxon, N_2 (ter Braak, 1990). The WA tolerances are divided by $(1 - 1/N_2)^{1/2}$ where N_2 is analogous to Hill's (1973) N_2 diversity measure. Select taxa as possible indicators that fulfil criterion A *and* that have tolerances less than the mean tolerance of all taxa in the training set (criterion B). Note that these tolerances are slightly different from the simple WA tolerances listed in Appendix B which are not adjusted for N_2 .

(4) Consider the number of occurrences of each taxon that fulfils criteria A and B, and tabulate the taxa with 10 or more occurrences (criterion C).

These taxa lists (Table 17) are ordered by number of occurrences. For pH 30 taxa fulfil the three criteria but only the 10 most frequent taxa are listed. For Al only 8 taxa fulfil the criteria and for DOC only 7 taxa qualify. Scatter plots of some of these indicator taxa in relation to pH, Al, and DOC for the total available data, rather than the screened data, are shown in Figure 14. Taxa that appear to be "indicator species" have a star in the top of their relevant scatter plots.

Table 17 List of indicator taxa for pH, DOC, and total Al that fulfil the criteria of having 10% or more of their variance explained by the chemical variable concerned and of having a tolerance (t) that is less than the mean tolerance for all taxa in the training sets. \hat{u} = WA optimum, n = number of occurrences. For pH only the 10 most frequent taxa are listed.

pH

	n	t	\hat{u}
<i>Frustulia rhomboides</i> var. <i>saxonica</i>	154	0.61	5.17
<i>Eunotia incisa</i>	145	0.47	5.06
<i>Brachysira brebissonii</i>	143	0.61	5.30
<i>Brachysira vitrea</i>	140	0.68	5.94
<i>Eunotia exigua</i>	121	0.46	5.06
<i>Achnanthes minutissima</i>	118	0.55	6.34
<i>Eunotia naegelii</i>	111	0.41	4.95
<i>Frustulia rhomboides</i>	109	0.55	5.12
<i>Cymbella hebridica</i>	104	0.49	5.09
<i>Cyclotella kuetzingiana</i> agg.	101	0.62	6.29

Dissolved Organic Carbon

	n	t	\hat{u}
<i>Achnanthes marginulata</i>	89	1.22	1.25
<i>Gomphonema acuminatum</i> var. <i>coronatum</i>	31	2.47	4.57
<i>Cyclotella comta</i>	21	2.20	6.12
<i>Aulacoseira distans</i> var. <i>tenella</i>	20	2.91	9.93
<i>Achnanthes pusilla</i>	20	2.23	5.90
<i>Aulacoseira subarctica</i> agg.	18	1.91	6.40
<i>Navicula indifferens</i>	16	2.50	5.44

Total Aluminium

	n	t	\hat{u}
<i>Achnanthes minutissima</i>	94	51.34	70.76
<i>Frustulia rhomboides</i>	77	54.88	133.36
<i>Navicula hoefleri</i>	56	59.46	140.20
<i>Stauroneis anceps</i> f. <i>gracilis</i>	45	52.21	67.87
<i>Brachysira serians</i>	39	57.31	158.51
<i>Stenopterobia sigmatella</i>	38	57.74	156.33
<i>Tabellaria binalis</i>	41	52.03	171.31
<i>Semiorbis hemicyclus</i>	24	45.70	177.20

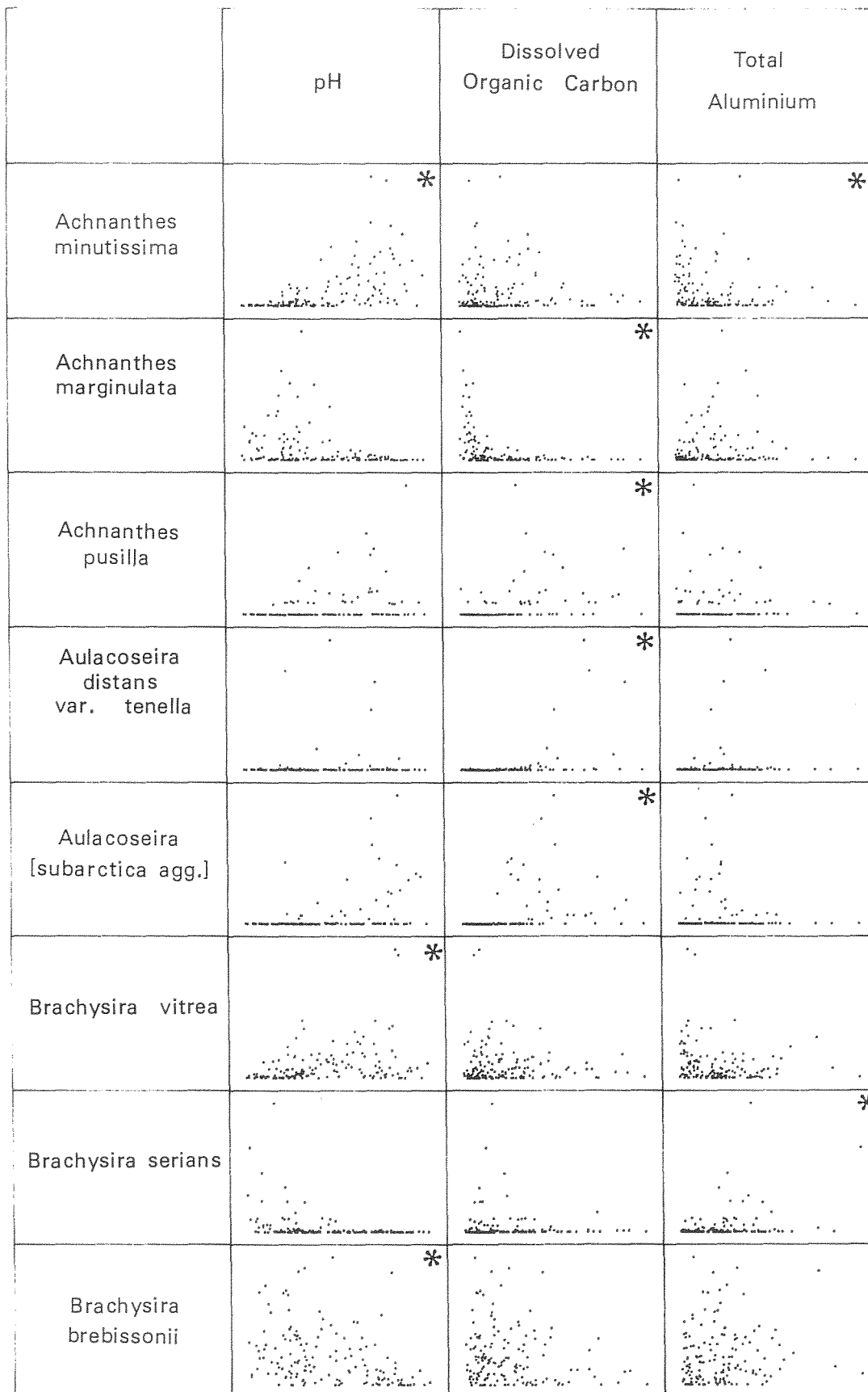


Figure 14 Scatter plots of selected diatom taxa showing their relative abundance (vertical axis scaled from zero to maximum abundance) in relation to pH, dissolved organic carbon, and total Al in the SWAP training sets for pH ($n = 178$), DOC ($n = 138$), and total Al ($n = 157$). * indicate those taxa that satisfy the criteria for potential indicator taxa (see Table 17) when the screened training data-sets are used.

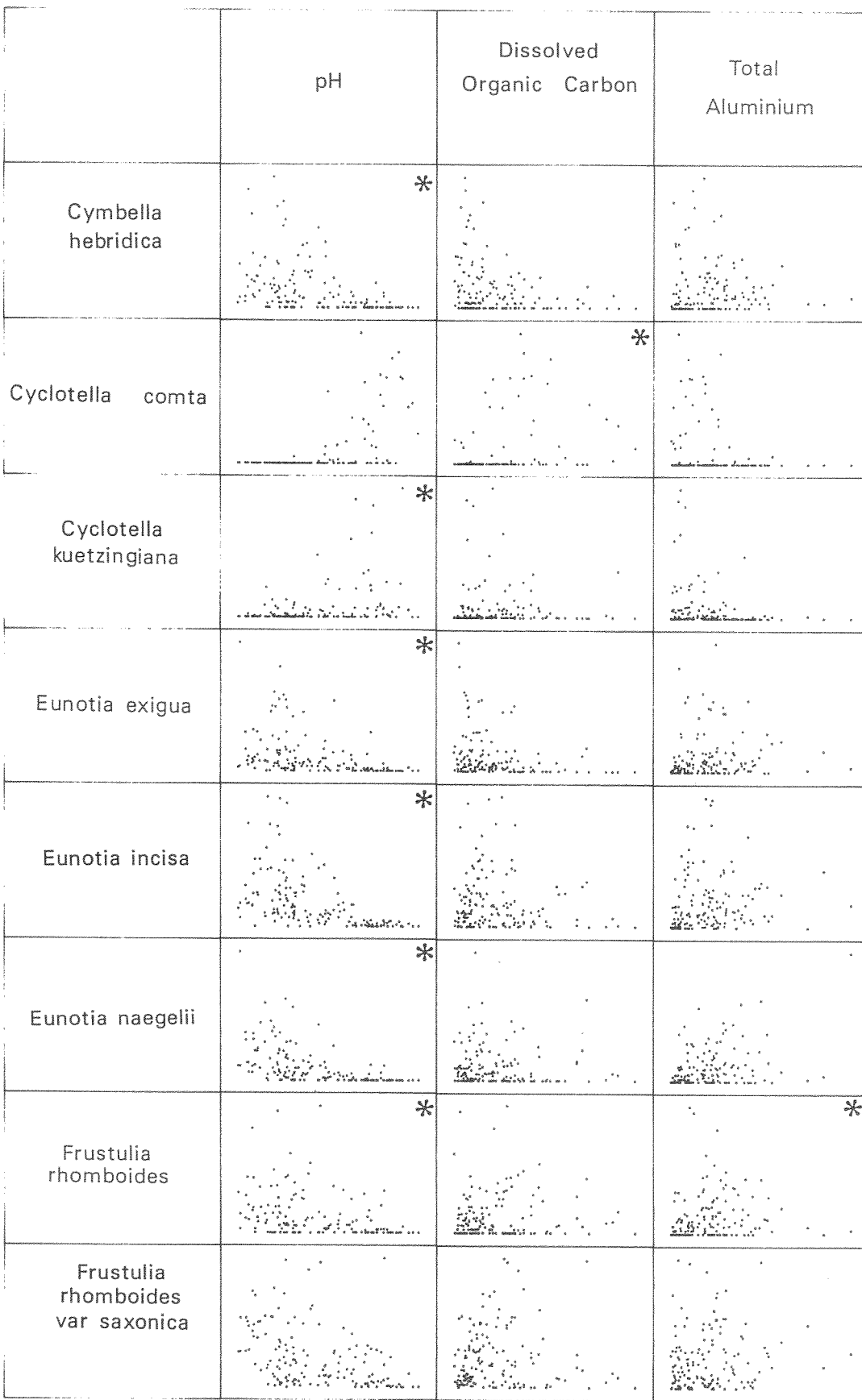


Figure 14 Continued.

	pH	Dissolved Organic Carbon	Total Aluminium	
Navicula hoefleri				*
Stauroneis anceps var. anceps				
Semiorbis hemicyclus				*
Stenopterobia sigmatella				*
Tabellaria binalis				*
Tabellaria quadriseptata				
Tabellaria flocculosa agg.				

Figure 14 Continued.

CONCLUSIONS

The SWAP set of modern diatom assemblages and associated environmental variables has, to date, been used almost exclusively as a lake-water chemical calibration training set. The 178-sample pH set, reduced to a training set of 167 samples after data-screening, provided the pH reconstructions for all the lakes studied in SWAP palaeolimnological sub-project (Birks *et al.* 1990a, Battarbee 1990, Atkinson and Haworth 1990, Renberg 1990, Birks *et al.* 1990c, Renberg *et al.* 1990a, Anderson and Korsman 1990, Kreiser *et al.* 1990, Berge *et al.* 1990, Renberg *et al.* 1990b, Jones *et al.* 1990, Renberg and Battarbee 1990). The 157-sample total Al training set (screened to 126 samples) and the 138-sample DOC set (screened to 123 samples) were used to reconstruct total Al and DOC at Round Loch of Glenhead (Birks *et al.* 1990b).

The SWAP data-set contains an enormous amount of modern but largely unexplored biological information at the individual taxon, samples, and lake scales, in addition to the basic diatom assemblage/lake chemistry information that has been exploited in the SWAP palaeolimnological investigations. Hopefully this wealth of biological information on diatom ecology will begin to be explored by diatomists. Moreover the data-set will continue to expand as new, but taxonomically and methodologically comparable, surface-sediment data are acquired from, for example, Ireland, Norway, and parts of Scotland. A harmonized modern data-set such as this, when explored from ecological and limnological viewpoints may help to strengthen the "bridges between paleolimnology and aquatic ecology" (Smol 1991).

BIBLIOGRAPHY

- Ahl, Th. (1972) Hydrochemical analyses. A catalogue of procedures used within the IHD in the Nordic countries. *Nordic IHD Report 3*, Oslo.
- Anderson, N.J. and Korsman, T. (1990) Land-use change and lake acidification: Iron Age de-settlement in northern Sweden as a pre-industrial analogue. *Philosophical Transactions of the Royal Society of London B* **327**, 373-376.
- APHA (1980) *Standard methods for the examination of water and wastewater*. American Public Health Association; American Water Works Association; and Water Pollution Control Federation (14th Edition), 1193 pp.
- Atkinson, K.M. and Haworth, E.Y. (1990) Devoke Water and Loch Sionascaig: recent environmental changes and the post-glacial overview. *Philosophical Transactions of the Royal Society of London B* **327**, 349-355.
- Battarbee, R.W. (1984) Diatom analysis and the acidification of lakes. *Philosophical Transactions of the Royal Society of London B* **305**, 451-477.
- Battarbee, R.W. (1990) The causes of lake acidification, with special reference to the role of acid deposition. *Philosophical Transactions of the Royal Society of London B* **327**, 339-247.
- Battarbee, R.W., Mason, Sir J., Renberg, I. and Talling, J.F. (1990) *Palaeolimnology and Lake Acidification*. Proceedings of a Royal Society Discussion Meeting held on 25 August 1989. Royal Society, London, 219 pp.
- Berge, F., Brodin, Y-W., Cronberg, G., El-Daoushy, F., Høeg, H.I., Nilssen, J.P., Renberg, I., Rippey, B., Sandøy, S., Timberlid, A. and Wik, M. (1990) Palaeolimnological changes related to acid deposition and land-use in the catchments of two Norwegian soft-water lakes. *Philosophical Transactions of the Royal Society of London B* **327**, 385-389.
- Birks, H.J.B., Line, J.M., Juggins, S., Stevenson, A.C. and ter Braak, C.J.F. (1990a) Diatoms and pH reconstruction. *Philosophical Transactions of the Royal Society of London B* **327**, 37-52.
- Birks, H.J.B., Juggins, S., and Line, J.M. (1990b) Lake surface-water chemistry reconstructions from palaeolimnological data. In Mason, B.J. (Ed.) *The Surface Waters Acidification Programme*. Cambridge University Press, Cambridge, 301-313.
- Birks, H.J.B., Berge, F., Boyle, J.F. and Cumming, B.F. (1990c) A palaeoecological test of the land-use hypothesis for recent lake acidification in south-west Norway using hill-top lakes. *Journal of Paleolimnology* **4**, 69-85.
- Camburn, K.E. and Kingston, J.C. (1986) The genus *Melosira* from soft-water lakes with special reference to North Michigan, Wisconsin and Minnesota. In: Smol, J.P., Battarbee, R.W., Davis, R.B. & Meriläinen, J. (Eds.) *Diatoms and Lake Acidity*. Junk, Dordrecht, 17-34.

- Charles, D.F. and Whitehead, D.R. 1986 The PIRLA project: paleoecological investigations of recent lake acidification. *Hydrobiologia* **143**, 13-20.
- Cleve-Euler, A (1951-1955). *Die Diatomeen von Schweden und Finnland*. K. Svenska Vetenskapsakad. Handl. Fjärde Ser. 2: 1. 3-163, 4: 1. 3-158, 4: 5. 3-255, 5: 4. 3-231, 3: 3. 3-153.
- Driscoll, C.T. (1984) A procedure for the fractionation of aqueous aluminium in dilute acid water. *International Journal of Environmental Analytical Chemistry* **16**, 267-283.
- Eliassen, A., Hor, O., Iversen, T., Saltbones, J., and Simpson, D. (1988) Estimates of airborne transboundary transport of sulphur and nitrogen over Europe. *EMEP/MS-C-W report* no. 1/88.
- Flower, R.J. and Jones, V.J. (1989) Taxonomic descriptions and occurrences of new *Achnanthes* taxa in acid lakes in the UK. *Diatom Research* **4**, 227-239.
- Flower, R.J and Kreiser, A.M. (1988) SWAP Diatom Taxonomy Update Number One - May 1988, unpublished manuscript, University College London.
- Golterman, H.L. (1969) *Methods for chemical analysis of freshwaters*. IBP Handbook No. 8. Blackwell Scientific Publications, Oxford, 166 pp.
- Hartley, B. (1986) A check-list of the freshwater, brackish and marine diatoms of the British Isles and adjoining coastal waters. *Journal of Marine Biological Association U.K.* **66**, 531-610.
- Haworth, E.Y. (1988) Distribution of diatom taxa of the old genus *Melosira* (now mainly *Aulacoseira*) in Cumbrian waters. In Round, F.E. (Ed.) *Algae and the aquatic environment*. Biopress, Bristol, 138-167.
- Haworth, E.Y., Atkinson, K.M. and Newell, P.S. (1988) Distribution of certain diatom taxa in the waterbodies of Cumbria, UK. *Proceedings of the 9th Diatom Symposium, 1986*, 17-28.
- Haworth, E.Y and Hurley, M.A. (1987) Comparison of the Stelligeroid Taxa of the Centric Diatom Genus *Cyclotella*. *Proceedings of the 8th Diatom Symposium, 1984*. 43-58.
- Hill, M.O. (1973) Diversity and evenness: a unifying notation and its consequences. *Ecology* **54**, 427-432.
- Hill, M.O. and Gauch, H.G. (1980) Detrended correspondence analysis: an improved ordination technique. *Vegetatio* **42**, 47-58.
- Hustedt, F. (1930-1966) *Die Kieselalgen Deutschlands, Österreichs und der Schweiz mit Berücksichtigung der übrigen Länder Europas sowie der angrenzenden Meeresgebiete*. *Kryptogamen-Flora* 7 Vol. 1 (1927-1930), 2 (1937-1959), 3 (1961-1966). Geest and Portig, Leipzig.

- Jones, V.J., Kreiser, A.M., Appleby, P.G., Brodin, Y-W., Dayton, J., Natkanski, J.A., Richardson, N., Rippey, B., Sandøy, S. and Battarbee, R.W. (1990) The recent palaeolimnology of two sites with contrasting acid-deposition histories. *Philosophical Transactions of the Royal Society of London B* **327**, 397-402.
- Jongman, R.H.G., ter Braak, C.J.F and van Tongeren, O.F.R. (1987) *Data Analysis in Community and Landscape Ecology*. Pudoc, Wageningen, 299pp.
- Kingston, J.C. (1986) Diatom methods. In Charles, D.F. and Whitehead, D.R. (Eds.) *Paleoecological investigation of recent lake acidification methods and project description*. EPRI EA-4906, Electric Power Research Institute, Palo Alto, California.
- Krammer, K. and Lange-Bertalot H. (1986) Bacillariophyceae. I. Teil: Naviculaceae. In: H. Ettl, J. Gerloff, H. Heynig and D. Mollenhauer (Eds.) *Süsswasserflora von Mitteleuropa*. Fischer, Stuttgart.
- Kreiser, A.M., Appleby, P.G., Natkanski, J., Rippey, B. and Battarbee, R.W. (1990) Afforestation and lake acidification: a comparison of four sites in Scotland. *Philosophical Transactions of the Royal Society of London B* **327**, 377-383.
- Line, J.M. and Birks, H.J.B. (1990) WACALIB version 2.1 - a computer program to reconstruct environmental variables from fossil assemblages by weighted averaging. *Journal of Paleolimnology* **3**, 170-173.
- Meriläinen, J. (1969) The diatoms of the meromictic Lake Valkiajärvi, in the Finnish Lake District. *Annales Botanici Fennici* **6**, 77-104.
- Munro, M.A.R., Kreiser, A.M., Battarbee, R.W., Juggins, S., Stevenson, A.C., Anderson, D.S., Anderson, N.J., Berge, F., Birks, H.J.B., Davis, R.B., Flower, R.J., Fritz, S.C., Haworth, E.Y., Jones, V.J., Kingston, J.C. and Renberg, I. (1990) Diatom quality control and data handling. *Philosophical Transactions of the Royal Society of London B* **327**, 31-36.
- Patrick, R. and Reimer, C. (1966) *The Diatoms of the United States Exclusive of Alaska and Hawaii. Volume I*. Academy Natural Sciences, Philadelphia Monographs 13.
- Patrick, R. and Reimer, C. (1975) *The Diatoms of the United States Exclusive of Alaska and Hawaii Volume II*. Academy Natural Sciences, Philadelphia Monographs 13.
- PIRLA (1984-1986) *PIRLA Diatom Iconograph* (Edited by K.E. Camburn, J.C. Kingston, and D.F. Charles) PIRLA Unpublished Report Series, Report 3, Indiana University, Bloomington.
- Renberg, I. (1976) Palaeolimnological investigations in Lake Prästsjön. *Early Norrland* **9**, 113-159.
- Renberg, I. (1990) A 12600 year perspective of the acidification of Lilla Öresjön, south-west Sweden. *Philosophical Transactions of the Royal Society of London B* **327**, 357-361.
- Renberg, I. and Battarbee, R.W. (1990) The SWAP Palaeolimnology Programme: a synthesis. In Mason, B.J. (Ed.) *The Surface Waters Acidification Programme*. Cambridge University Press, Cambridge, 281-300.

- Renberg, I., Korsman, T. and Anderson, N.J. (1990a) Spruce and surface water acidification: an extended summary. *Philosophical Transactions of the Royal Society of London B* **327**, 371-372.
- Renberg, I., Brodin, Y-W., Cronberg, G., El-Daoushy, F., Oldfield, F., Rippey, B., Sandøy, S., Wallin, J-E. and Wik, M. (1990b) Recent acidification and biological changes in Lilla Öresjön, southwest Sweden and the relation to atmospheric pollution and land-use history. *Philosophical Transactions of the Royal Society of London B* **327**, 391-396.
- Schmidt, A. (1984) *Atlas der Diatomaceen-kunde. Leipzig und Berlin, 1874 - 1959*. Koeltz Scientific Books.
- Simonsen, A. (1987) *Atlas and catalogue of the diatom types of Friedrich Hustedt. Vols I-III*. J. Cramer, Berlin.
- Smol, J.P. (1991) Are we building enough bridges between palaeolimnology and aquatic ecology? *Proceedings of the Vth International Symposium on Palaeolimnology* (in press).
- Stevenson, A.C., Patrick, S.T., Kreiser, A.M. and Battarbee, R.W. (1987) *Palaeoecological evaluation of the recent acidification of susceptible lakes: methods utilised under DoE contract PECD 7/7/139 and the Royal Society SWAP project*. Palaeoecology Research Unit, University College London. Research Paper No 26, 36pp.
- Stevenson, A.C., Birks, H.J.B., Flower, R.J. and Battarbee, R.W. (1989) Diatom-based pH reconstruction of lake acidification using canonical correspondence analysis. *Ambio* **18**, 228-233.
- ter Braak, C.J.F. (1986) Canonical correspondence analysis: a new eigenvector technique for multivariate direct gradient analysis. *Ecology* **67**, 1167-1179.
- ter Braak, C.J.F. (1987) *CANOCO - A FORTRAN program for canonical community ordination by [partial] [detrended] [canonical] correspondence analysis, principal components analysis and redundancy analysis (version 2.1)*. TNO Institute of Applied Computer Science, Statistics Department, Wageningen, Netherlands, 95pp.
- ter Braak, C.J.F. (1990) *Update notes: CANOCO version 3.10*. Agricultural Mathematics Group Wageningen, 35 pp.
- ter Braak, C.J.F. and Prentice, I.C. (1988) A theory of gradient analysis. *Advances in Ecological Research* **18**, 271-317.
- ter Braak, C.J.F. and van Dam, H. (1989) Inferring pH from diatoms: a comparison of old and new calibration methods. *Hydrobiologia* **178**, 209-223.
- Vik, E.A. (1982) *Treatment of potable water containing humus by electroanalytical addition of aluminium*. PhD Thesis, University of Washington, Department of Civil Engineering, Seattle, USA.
- Williams, D.M., Hartley, B., Ross, R., Munro, M.A.R., Juggins, S. and Battarbee, R.W. (1988) *A Coded Checklist of British Diatoms*. ENSIS publishing, London, 74pp.

Appendix A

THE SWAP TAXONOMIC GUIDE - REVISED 1989

The diatom taxa discussed by participants at the 1987, 1988, and 1989 SWAP diatom workshops are listed below in the following format:

- a) Taxon and authority.
 - b) Taxonomic decisions, agreements, or alterations.
 - c) Ecological comments.
-

- a) *Achnanthes altaica* (Poretzky) Cleve-Euler
- b) Supposedly synonymous with *A. recurvata* Hust. although both names appear to be valid on the BM checklist. A small non-curved form of this taxon *A. [altaica var. minor]* has been identified by RJF and is described in the guide circulated at the SWAP workshop, July 1988.

- a) *Achnanthes hyperborea* Grun.
- b) Synonymous with *A. frigida*.

- a) *Achnanthes gibberula* Grun.
- b) Synonymous with *A. biasolettiana*

- a) *Achnanthes pusilla* Grun.
- b) Synonymous with *A. linearis* var. *pusilla* Grun.

- a) *Achnanthes austriaca* var. *helvetica*
- b) Two new forms of this taxon were discussed at the July 1988 workshop; f. *alpina* (formerly *elongata*) and f. *minor*. They are described in Flower and Jones (1989) but will not be used generally until their distribution in the surface sediment data-set has been assessed.

- a) *Achnanthes microcephala* Kütz.
- b) Following the results of the 1988 slide counts it was decided not to differentiate between *A. minutissima* and *A. microcephala* in the central data-set although individual analysts could split the two taxa at the local level if they wished.

- a) *Achnanthes marginulata* Grun.
- b) Counts of this species have possibly included the diatom described as *A. [marginulata f. major]* by VJJ at the SWAP July 1988 workshop.

- a) *Achnanthes [scotica]* Jones & Flower
- b) This small *Achnanthes* has probably been included in counts of *A. marginulata* in the past. Since it has been found by RJF and VJJ to occur independently of *A. marginulata* in cores, it was agreed to re-examine surface sediment slides to determine its distribution in modern material. Taxonomy follows Flower and Jones (1989).

- a) *Achnanthes marginulata* Grun.
- b) Synonymous with *A. [sp. 5]* Loch Uaine (VJJ).

- a) *Amphora ovalis* var. *affinis* Kütz.
 - b) Synonymous with *Amphora ovalis* var. *libyca*.
-
- a) *Asterionella ralfsii* W. Smith
 - b) Problems arise in distinguishing this taxon from *A. formosa*, particularly when only broken valves are present. A category *A. ralfsii/formosa* has been created for use when such uncertainty exists.
 - c) *A. ralfsii* appears to prefer coloured waters and is therefore often associated with basins receiving artificial drainage from peat.
-
- a) *Aulacoseira distans* (Ehr.) Simonsen
 - b) Synonym *Melosira distans*. All transfers from *Melosira* to *Aulacoseira* follow Hartley (1986). Refer to Camburn and Kingston (1986) for a description of the taxon.
-
- a) *Aulacoseira lirata* var. *alpigena* (Grun.) Haworth
 - b) Synonym *Melosira distans* var. *alpigena*, *Aulacoseira distans* var. *alpigena* and *Aulacoseira* [sp. 4].
-
- a) *Aulacoseira* [cf. *distans* var. *distans*] SWAP 1989
 - b) Synonymous with *A. distans* var. *distans* (Ehrenb.) Simonsen 1979.
-
- a) *Aulacoseira distans* var. *nivalis*
 - b) Synonym *Melosira distans* var. *nivalis*, but has not yet been put into *Aulacoseira* in the BM checklist. Taxonomy follows Camburn and Kingston (1986). Bears some similarity to *A. distans* var. *africana* although the two are separate taxa.
-
- a) *Aulacoseira* [*subborealis/subarctica*]
 - b) This taxon has a coarsely punctate valve face and large spines. In girdle view the slightly spiral striae consist of two rows of punctae. The reference site for this taxon is Botungen.
-
- a) *Aulacoseira distans* var. *tenella* (Nygaard) Ross
 - b) Synonym *Melosira distans* var. *tenella*. Taxonomy follows Camburn and Kingston (1986).
-
- a) *Aulacoseira lirata* (Ehr.) Ross
 - b) Synonym *Melosira lirata*. Taxonomy follows Camburn and Kingston (1986). The valve view maybe confused with that of *A. perglabra* var. *floriniae*.
-
- a) *Aulacoseira lirata* var. *lacustris* (Grun.) Ross
 - b) Synonym *Melosira lirata* var. *lacustris*. Taxonomy follows Camburn and Kingston (1986).
-
- a) *Aulacoseira lirata* f. *biseriata*
 - b) Synonym *Melosira lirata* f. *biseriata*. To be transferred to *A. distans* var. *lirata* f. *biseriata*.
-
- a) *Aulacoseira perglabra*
 - b) Synonym *Melosira perglabra*. Taxonomy follows Camburn and Kingston (1986). This taxon is recorded as *A. lirata* var. *perglabra* on the BM checklist but will be transferred to *Aulacoseira perglabra*.

- a) *Aulacoseira perglabra* var. *floriniae*
 - b) Synonym *Melosira perglabra* var. *floriniae*. Taxonomy follows Camburn and Kingston (1986). Valve view may be confused with that of *A. lirata* if one is not familiar with both taxa. This taxon does not appear on the BM checklist but should be added shortly.
- a) *Aulacoseira* [*subarctica* agg.]
 - b) This includes *Aulacoseira subarctica* var. *subborealis*, a description of which is in Haworth (1988). SEM and LM photographs of *M. italica* var. *subborealis* Nygaard can be found in Engelmark *et al.* (1976), p. 151, fig. 4.
- a) *Aulacoseira* [var.5 PIRLA]
 - b) This unknown taxon from the surface sediments of Llyn Tecwyn closely resembles the PIRLA *Aulacoseira distans* var.5 PIRLA (1984-1986) Iconograph plate 51, fig. 5. It has a punctate valve face with a suggestion of curvature and prominent spines. The sulcus is poorly developed and the valve measures 9-12 µm in diameter. In girdle view the parallel striae are 20 in 10 µm. The valves are 5-6 µm deep. This diatom is similar to the *A. distans* var. *laevissima* found by EYH.
- a) *Berkeleya rutilans* (Trentepohl *ex* Roth.) Grun. 1868
 - b) Synonymous with *Amphipleura rutilans* (Trentepohl *ex* Roth.) Cleve 1894.
- a) *Brachysira brebissonii* Ross
 - b) Synonym *Anomoeoneis brachysira* (Bréb.) Grun. It was decided not to use f. *thermalis* since the two taxa cannot be consistently separated.
- a) *Brachysira* [*serians* small]
 - b) This small form has been found by IR and NJA in some Scandinavian lakes. Although diatomists may separate this form in their counts, it was agreed not to include it in the database.
- a) *Brachysira follis* (Ehr.) Ross
 - b) Synonym *Anomoeoneis follis*.
- a) *Brachysira serians* (Bréb. *ex* Kütz.) Round & Mann
 - b) Synonym *Anomoeoneis serians*.
- a) *Brachysira serians* var. *modesta*
 - b) This small variety of *B. serians* follows Cleve-Euler (1951-1955) described in vol. III p. 197, but is not separated from *B. brebissonii* in the data-set.
- a) *Brachysira styriaca* (Grun.) Ross
 - b) Synonym *Anomoeoneis styriaca*.
- a) *Brachysira vitrea* (Grun.) Ross
 - b) Synonym *Anomoeoneis vitrea*.
- a) *Brachysira vitrea* var. *lanceolata*
 - b) Synonym *Anomoeoneis vitrea* var. *lanceolata*. It was agreed to distinguish this taxon from the nominate at the local level but not in the data-set.
- a) *Brachysira* [sp.1 Round Loch]
 - b) It was agreed to include this unpublished *Brachysira* in the data-set.

- a) *Caloneis ventricosa* (Ehrenb.) Meister 1912
- b) Synonymous with *Caloneis silicula* (Ehrenb. (Cleve 1894)).

- a) *Cyclotella comensis* Grun.
- b) This is a separate taxon from *C. kuetzingiana* although it is probable that the two taxa have been confused in the past.

- a) *Cyclotella* [*kuetzingiana* agg.]
- b) There are no agreed criteria within the group for separating *C. kuetzingiana* from its varieties. Therefore, this name should be used for both the nominate and its varieties in the data-set.

- a) *Cyclotella kuetzingiana* Thwaites
- b) Only to be used at the local level (see above).

- a) *Cyclotella kuetzingiana* var. *planetophora* Fricke
- b) Only to be used at the local level (see above).

- a) *Cyclotella kuetzingiana* var. *radiosa* Fricke
- b) Only to be used at the local level (see above).

- a) *Cyclotella comta* (Ehrenb.) Kütz.
- b) Synonymous with *C. radiosa* Håkansson but has not been revised in the BM checklist.

- a) *Cyclotella glomerata* Bachm.
- b) Synonym *C. stelligera* var. *glomerata* Haworth. See Haworth and Hurley (1987). Included in the BM checklist as *C. glomerata*.

- a) *Cymbella aequalis* W. Smith
- b) This taxon may have been confused with *C. angustata* in the past although it is clear only *C. aequalis* is present in the SWAP slides.

- a) *Cymbella gaumannii* Meister
- b) There is some taxonomic confusion due to the similarity with *C. perpusilla*. Taxonomy follows the description in Krammer and Lange-Bertalot (1986). The *C. amphicephala* in Hustedt (1930) is synonymous with *C. gaumannii*.

- a) *Cymbella lunata* W. Smith
- b) Synonym *C. gracilis*. For taxonomy see Patrick and Reimer (1966).

- a) *Cymbella minuta* Hilse ex Rabh.
- b) Synonym *C. ventricosa* Kütz. (the smaller, finer forms). For taxonomy see Krammer and Lange-Bertalot (1986). See also *C. silesiaca*.

- a) *Cymbella perpusilla* A. Cleve
- b) Follows the description in Krammer and Lange-Bertalot (1986). *C. bipartita* Mayer is synonymous with *C. perpusilla*, and was used to describe the oval forms. See discussion on *C. gaumannii* above.

- a) *Cymbella scotica* var. *naviculacea* (Grun. ex Cleve) R. Ross 1947.
- b) Synonymous with *C. naviculacea* Grun. ex Cleve 1881.

- a) *Cymbella silesiaca* Bleisch in Rabenhorst (1864).
 - b) This taxon includes the larger, coarser forms of *C. ventricosa*. See Krammer and Lange-Bertalot (1986).
- a) *Cymbella elginensis* Krammer
 - b) This is synonymous with *C. turgida* Greg. This should not be confused with *C. turgida sensu* Cleve (1894) or *C. turgida sensu* Hustedt (1930) *pro parte*, which are now in *C. mesiana* Cholnoky. See Krammer and Lange-Bertalot (1986) for details.
- a) *Diatoma tenue* var. *elongatum* Lyngb. 1819
 - b) Synonymous with *D. elongatum* (Lyngb.) Ag. 1824
- a) *Eunotia arcus* Ehr.
 - b) No apparent taxonomic problems.
- a) *Eunotia bactriana* Ehr.
 - b) No taxonomic problems.
- a) *Eunotia curvata* (Kütz.) Lagerst.
 - b) Synonym *E. lunaris*. Used for diatoms of the *E. curvata/naegelii* group with width > 3 µm and striae count < 20 in 10µm.
- a) *Eunotia curvata* var. *subarcuata* (Naegeli) Woodhead & Tweed
 - b) This name was probably used for the shorter forms of *E. naegelii* in the past but should now be used with care for the short forms of *E. curvata*, applying the criteria described above. Do not use the Hustedt description of this taxon.
- a) *Eunotia denticulata* (Bréb.) Rabh.
 - b) This name is to be applied to the "classic" *E. denticulata* as described in Hustedt (1930 -1966), vol II, p. 293 figs. a-i.
- a) *Eunotia denticulata/laponica*
 - b) This combination has been formed to accommodate the broader form of *E. denticulata* with the robust apices characteristic of *E. laponica*.
- a) *Eunotia exigua* (Bréb.) Rabh.
 - b) This name has been applied to diatoms showing a number of morphological variations. Refer to Hustedt (1930-1966) vol. II p. 287 fig. 751 a-r as a starting point, being aware that Petersen has removed *E. paludosa* and *E. steineckii* from Hustedt's concept of *E. exigua*. See Flower and Kreiser (1988).
- a) *Eunotia flexuosa* Kütz.
 - b) Two forms of this taxon, fine and coarse have been noted.
- a) *Eunotia incisa* W. Smith *ex* Greg.
 - b) Synonymous with the diatom commonly called *E. veneris*
- a) *Eunotia major* (W. Smith) Rabh.
 - b) Synonym *E. monodon* var. *major*

- a) *E. minutissima* Cleve-Euler
 - b) For description see Cleve-Euler (1955) and SWAP taxonomy update (Flower and Kreiser 1988). It is possible that this taxon has been confused with small, fine forms of *E. naegelii* in the past.
- a) *Eunotia* [sp.13 *minutissima*]
 - b) Used to describe a small *Eunotia* similar to *E. minutissima* but finer (see Flower and Kreiser 1988).
- a) *Eunotia naegelii* Migula
 - b) Synonym *E. alpina sensu* Hustedt in Schmidt Atlas vol II plate 291 figs 7 & 8. This name to be used for diatoms in the *E. curvata/naegelii* group with width <3 µm and striae >20 in 10 µm. It is probable that shorter specimens of this taxon have been assigned to *E. curvata* var. *subarcuata* in the past.
- a) *Eunotia nodosa* Ehr.
 - b) Synonym *E. formica*.
- a) *Eunotia nymanniana* Grun
 - b) Synonym *E. exigua* var. *compacta*. Description in Cleve-Euler (1951-1953), vol. II, fig. 445 a & d.
- a) *Eunotia pectinalis* (Kütz.) Rabh.
 - b) No taxonomic problems.
- a) *Eunotia pectinalis* var. *minor* (Kütz.) Rabh.
 - b) See SWAP taxonomy update (Flower and Kreiser 1988).
- a) *Eunotia pectinalis* var. *ventricosa* Grun.
 - b) Synonym *E. pectinalis* var. *ventralis*.
- a) *Eunotia praerupta-nana* Berg
 - b) It was agreed to keep this taxon.
- a) *Eunotia rhomboidea* Hust.
 - b) It was agreed to keep this taxon although its morphology can come close to *E. incisa*.
- a) *Eunotia serra* Ehr.
 - b) Synonym *E. robusta*.
- a) *Eunotia serra* var. *diadema* (Ehr.) Patr.
 - b) Synonym *E. robusta* var. *diadema*. This taxon also includes *E. robusta* var. *tetraodon*.
- a) *Eunotia tenella* (Grun.) Hust.
 - b) See SWAP taxonomy update (Flower and Kreiser 1988).
- a) *Eunotia tibia* var. *bidens* (W. Smith) Cleve-Euler
 - b) Synonym *E. major* var. *bidens* and *E. monodon* var. *bidens*.
- a) *Eunotia tridentula* Ehr.
 - b) Synonym *E. polydentula* and *E. tridentula* var. *perminuta* (see BM list).

- a) *Eunotia vanheurckii* Patrick
 - b) Synonym *E. faba*.
- a) *Eunotia vanheurckii* var. *intermedia*
 - b) Synonym *E. faba* var. *intermedia*. See SWAP taxonomy update (Flower and Kreiser 1988).
- a) *Eunotia exgracilis* A. Berg ex Cleve-Euler 1953
 - b) Synonymous with *E.* [PIRLA sp.8] PIRLA (1984-1986).
- a) *Fragilaria* [*cf. oldenburgiana*]
 - b) This taxon does not correspond to Hustedt's type material from Bremerhaven, (Simonsen 1987, vol. 3 plate 680 figs. 1-5) although it is agreed the *F. cf. oldenburgiana* in the PIRLA material is the same as that found in Europe. Reference: PIRLA (1984-1986) Iconograph plate 20 figs. 61-62 & plate 38 fig. 107.
- a) *Fragilaria heidenii* Østr. 1910
 - b) Synonymous with *F. inflata* (Heiden) Hust. 1931.
- a) *Fragilaria intermedia* Grun.
 - a) *Fragilaria vaucheriae* (Kütz.) Boye Petersen
 - b) Petersen put these two taxa together but EYH separates them using striae count, TEM structure, and whether or not the cells form colonies. It was agreed to separate them if possible but more information is needed on their differences.
- a) *Fragilaria virescens* var. *exigua* Grun.
 - b) This name is applied within PIRLA to the taxon known as *F. virescens* in Europe (PIRLA (1984-1986) Iconograph plate 25 figs. 105-112) since the taxonomy of *F. virescens* is uncertain. It was decided to adopt the above name for the taxon currently called *F. virescens* until the type material of *F. virescens* has been examined.
- a) *Frustulia* [*rhomboides* agg.]
 - b) It was agreed to use this classification for central areas or ends which can not be confidently assigned to the nominate or one of the varieties.
- a) *Frustulia rhomboides* (Ehr) De Toni
 - b) This is split from the variety *saxonica* by its size (generally >60µm) and striae density (<30 in 10µm) but see unpublished notes by RJF. It differs from the variety *viridula* in that it has a straight raphe, symmetrical central area, and the striae continue around the apices of the valve. A good reference is Krammer and Lange-Bertalot (1986), fig. 95, nos. 2 & 3. (No. 1 is ambiguous). See also Haworth *et al.* (1988).
- a) *Frustulia rhomboides* var. *viridula* (Bréb. ex Kütz.) Cleve
 - b) This taxon can be separated from the nominate by its curved raphe, asymmetrical central area, and gap in the striae at the apices. It is also usually larger than the nominate. See Krammer and Lange-Bertalot (1986), fig. 96 nos. 1-3 and also Haworth *et al.* (1988).
- a) *Frustulia rhomboides* var. *saxonica* (Rab) De Toni
 - b) Can be distinguished from the nominate by its smaller size (<60 µm) and finer striae (>30 in 10 µm).
- a) *Gomphonema olivaceum* (Hornemann) P. Dawson ex R. Ross in Hartley (1986)
 - b) Synonymous with *G. olivaceum* (Hornemann) Bréb. 1838.

- a) *Gomphonema vibrio* var. *intricatum* (Kütz.) R. Ross in Hartley (1986)
- b) Synonymous with *G. intricatum* Kütz. 1844.

- a) *Gomphonema vibrio* var. *pumilum* (Grun.) R. Ross in Hartley (1986)
- b) Synonymous with *G. intricatum* var. *pumilum* Grun. 1880

- a) *Melosira arentii* (Kolbe) Nagumo & Kobayasi
- b) Synonym *Cyclotella arentii*.

- a) *Navicula angusta* Grun.
- b) Synonyms: *N. cari* var. *angusta* and *N. lobeliae*.

- a) *Navicula bremensis* Hust.
- b) The *N. bremensis* seen in the SWAP material does not exactly fit the description in Hustedt (1930-1966), although it corresponds with the *N. bremensis* in the PIRLA (1984-1986) Iconograph, plate 1 figs. 3-4. It was decided to continue using the name for the time being.

- a) *Navicula capitata* var. *hungarica* (Grun.) R. Ross 1947
- b) Synonymous with *N. hungarica* Grun. 1860.

- a) *Navicula cocconeiformis* Gregory
- b) No taxonomic problem.

- a) *Navicula contenta* var. *parallela* Petersen
- b) Synonymous with the *Navicula* sp. at 37 cm in the Tveita core. Reference: Hustedt (1930-1966) vol. III p. 210

- a) *Navicula cumbriensis* Haworth
- b) Synonymous with *N. madumensis* var. 1 (UCL) and *N. subtilissima* var. 2 (I. Renberg). See Haworth (1988). Note that *N. cumbriensis* is strictly an invalid name because it is preempted by Koboyasi and Nagumo's description of this taxon as *N. parasubtilissima* (E.Y. Haworth, personal communication). For the purposes of SWAP, this taxon is still informally termed *N. cumbriensis*.

- a) *Navicula* [*cumbriensis* small]
- b) This is a small variety of the above but is not *N. cumbriensis* var. *minor* as identified by EYH. It was agreed to make the distinction between this and the nominate at the local level but to amalgamate them in the data-set.

- a) *Navicula detenta* Hust. 1943
- b) Synonymous with *N. dicephala* Ehrenb. 1838.

- a) *Navicula cryptocephala* var. *exilis* (Kütz.) Grun.
- b) Synonymous with *N. exilis* Kütz. 1844

- a) *Navicula festiva* Krasske
- b) No taxonomic problem.

- a) *Navicula gysingensis* Foged
- b) Previously called *N. cf. arvensis* in the Öresjön SWAP slides. Reference: PIRLA (1984-1986) iconograph, plate 41 figs. 23-24.

- a) *Navicula hoefleri* Cholnoky
 - b) Not to be confused with *N. [hoefleri sensu Ross & Sims]*. Reference: Hustedt (1930-1966) vol. III, p. 97.
-
- a) *Navicula [hoefleri sensu Ross & Sims]*
 - b) Synonymous with *N. simsii* and *N. subtilissima* var 4, PIRLA (1984-1986) iconograph, plate 42 figs. 75-77.
-
- a) *Navicula leptostriata* Jorgensen
 - b) Synonym *N. heimansii*.
-
- a) *Navicula madumensis* Jorgensen
 - b) Synonymous with *N. subtilissima* var.1 (I. Renberg).
-
- a) *Navicula mediocris* Krasske
 - b) No taxonomic problem.
-
- a) *Navicula minuscula* var. *muralis* (Grun.) Lange-Bertalot.
 - b) Synonymous with *N. muralis* Grun.
-
- a) *Navicula radiosa* var. *tenella* (Bréb.) Grun.
 - b) This includes the concept of *N. cari sensu* Germain, as used at UCL.
-
- a) *Navicula subtilissima* Cleve
 - b) Reference: Hustedt (1930-1966) vol. III, p. 89
-
- a) *Navicula tenuicephala* Hust.
 - b) The SWAP concept of this taxon conforms with that found in Simonsen (1987) vol. 3 plate 457, figs. 11-16. There is some variability in the completeness of the stauros and some specimens have an asymmetrical stauros. There is possibly another taxon with a symmetrical central area.
-
- a) *Neidium affine* (Ehr.) Cleve
 - b) No taxonomic problem.
-
- a) *Neidium glaberrimum* (Østrup) Ross
 - b) Synonym *N. alpinum*
-
- a) *Nitzschia perminuta* (Grun.) M. Perag.
 - b) The *N. perminuta* in the SWAP material is synonymous with *Nitzschia* sp.1 PIRLA (1984-1986) iconograph, (plate 44, fig. 123). However, the taxonomy of this species remains unclear. In the BM checklist *N. perminuta* has been put into *N. frustulum* yet the SWAP *N. perminuta* is clearly not related to *N. frustulum*. The above name will be used in the data-set but further work is required.
-
- a) *Nitzschia pusilla* Grun. 1862
 - b) Synonymous with *N. kuetzingiana* Hilse 1863.
-
- a) *Pinnularia abaujensis* var. *abaujensis* (Pant.) R. Ross
 - b) Synonymous with *P. gibba* (Ehrenb.) Ehrenb. 1843.
-
- a) *Pinnularia biceps* Greg.
 - b) Synonym *P. interrupta*. It was decided not to separate the varieties of *P. biceps*.

- a) *Pinnularia irrorata* (Grun.) Hust.
- b) Decision made not to synonymise *P. irrorata* with *P. appendiculata* although Krammer and Lange-Bertalot (1986) do so. The concept of *P. irrorata* is represented in plate 193 and figs. 21-26. *P. sylvatica* may need to be included in this taxon but more research is required.

- a) *Pinnularia subcapitata* var. *hilseana* (Janisch) O. Mull.
- b) Synonym *P. hilseana*

- a) *Pinnularia rupestris* Hantzsch
- b) Synonym *Pinnularia viridis* var. *rupestris* (Hantzsch) Cleve. Illustrations can be found in Meriläinen (1969), p. 96, fig. 171 and in Renberg (1976), p. 151 plate 5 fig. 12.

- a) *Semiorbis hemicyclus* (Ehr.) Patr.
- b) Synonym *Amphicampa hemicyclus*.

- a) *Stenopteroberia sigmatella* (Greg.) Ross
- b) Synonym *Stenopteroberia intermedia*.

- a) *Surirella delicatissima* Lewis
- b) It was agreed to split fine and coarse forms at the local level, but not in the data-set.

- a) *Tabellaria binalis* (Ehr.) Grun.
- b) For the purposes of the data-set, this name will include both the elliptic and panduric forms although the two varieties will probably be split at the local level.

- a) *Tabellaria fenestrata* (Lyngbye) Kütz.
- b) This may have been confused in the past with the longer forms of *T. flocculosa*. More work is required on morphology and colony structure.

- a) *Tabellaria flocculosa* var. *flocculosa* (Roth.) Kütz.
- b) This name is to be applied to those diatoms conforming to Koppen's strain IV, i.e. length <39 µm and central inflation distinctly wider than those at the apices.

- a) *Tabellaria* [*flocculosa* agg.]
- b) For the purposes of the data-set this category is to be applied to all specimens of *T. flocculosa* which do not fit into the above description. This will include Koppen's strain III and the twisted form IIIp plus var. *linearis*. Analysts may differentiate between these forms at the local level if they wish.

Abbreviations

AK Annette Kreiser
 EYH Elizabeth Haworth
 IR Ingemar Renberg
 NJA John Anderson
 RJF Roger Flower
 UCL University College London
 VJJ Vivienne Jones

Appendix B *List of diatoms in the SWAP training set and their optima for pH, DOC (mg l⁻¹), and Total Al (µg l⁻¹) estimated by weighted averaging (WA) regression. (O = WA optimum, T = WA tolerance, N = number of occurrences in total data-set or the different screened training sets, Max = maximum abundance in the total data-set.)*

Taxon code	Taxon name and authority	Total data-set		pH			DOC			Al		
		N	Max	O	T	N	O	T	N	O	T	N
AC046A	<i>Achnanthes altaica</i> (Poretzky) A. Cleve-Euler 1953	93	5.9	5.7	0.6	88	3.2	2.5	69	98.3	67.1	70
AC014A	<i>Achnanthes austriaca</i> Hust. 1922	7	3.4	4.9	0.4	7	3.5	1.3	7	120.2	36.8	6
AC9965	<i>Achnanthes austriaca</i> var. <i>alpina</i> Uaine (VJJ) 1988	3	6.3	5.8	0.3	3	0.4	0.1	3	103.3	45.3	3
AC014C	<i>Achnanthes austriaca</i> var. <i>helvetica</i> Hust. 1933	94	24.5	5.4	0.4	88	1.5	1.5	73	84.9	47.1	66
AC014B	<i>Achnanthes austriaca</i> var. <i>minor</i> L. Granoch (RJF) 1986	30	5.5	5.1	0.4	29	2.6	2.1	25	124.0	45.6	25
AC9996	<i>Achnanthes</i> cf. <i>levanderi</i>	21	11.2	6.1	0.4	19	1.4	1.3	17	32.7	25.6	15
AC042A	<i>Achnanthes detha</i>	41	21.7	6.1	0.5	38	1.8	2.2	33	41.3	49.5	33
AC039A	<i>Achnanthes didyma</i> Hust. 1933	7	1.0	6.3	0.6	6	3.8	3.3	4	72.9	60.7	6
AC025A	<i>Achnanthes flexella</i> (Kutz.) Brun 1880	23	1.3	6.3	0.6	21	5.1	2.0	13	77.8	35.8	15
AC017A	<i>Achnanthes kryophila</i> J.B. Petersen 1924	11	3.2	6.0	0.7	10	1.5	0.4	7	63.2	40.6	4
AC001A	<i>Achnanthes lanceolata</i> (Breb. ex Kutz.) Grun. 1880	13	0.9	6.4	0.7	11	4.0	1.8	7	87.6	55.9	11
AC018A	<i>Achnanthes laterostrata</i> Hust. 1933	11	1.4	6.2	0.7	9	4.5	3.4	6	43.2	27.9	8
AC044A	<i>Achnanthes levanderi</i> Hust. 1933	36	12.5	5.6	0.6	34	1.5	1.2	27	79.2	57.1	27
AC002A	<i>Achnanthes linearis</i> (W. Sm.) Grun. in Cleve & Grun. 1880	31	4.6	6.4	0.5	26	5.0	2.0	19	102.0	49.8	22
AC022A	<i>Achnanthes marginulata</i> Grun. in Cleve & Grun. 1880	122	46.9	5.2	0.4	118	1.3	1.2	89	124.2	64.9	89
AC9968	<i>Achnanthes marginulata</i> form <i>major</i> Uaine (VJJ) 1988	20	40.5	5.2	0.3	20	0.7	0.5	20	138.4	76.5	16
AC013A	<i>Achnanthes minutissima</i> Kutz. 1833	127	46.0	6.3	0.5	118	3.4	2.2	92	70.8	50.7	94
AC9964	<i>Achnanthes minutissima</i> var. <i>scotica</i> (Carter) RJF 1988	7	1.3	6.2	0.6	6	1.9	0.6	4	49.3	15.8	5
AC019A	<i>Achnanthes nodosa</i> A. Cleve-Euler 1900	17	5.6	6.2	0.4	17	1.9	1.2	14	29.2	15.4	12
AC004A	<i>Achnanthes pseudoswazi</i> J.R. Carter 1963	31	3.4	6.1	0.5	28	2.9	1.3	18	61.4	38.3	22
AC035A	<i>Achnanthes pusilla</i> Grun. in Cleve & Grun. 1880	35	2.0	6.2	0.6	31	5.9	2.1	20	99.7	54.8	24
AC028A	<i>Achnanthes saxonica</i> Krasske in Hust. 1933	11	0.9	5.7	0.5	11	1.8	1.3	10	83.9	68.7	11
AC048A	<i>Achnanthes scotica</i> Jones & Flower	28	5.3	5.6	0.5	27	1.0	0.8	25	70.7	49.6	23
AC029A	<i>Achnanthes sublaevis</i> Hust. 1936	30	8.0	5.8	0.6	26	2.1	1.6	20	57.5	36.2	22
AC034A	<i>Achnanthes suchlandtii</i> Hust. 1933	12	1.7	6.3	0.4	10	1.5	1.9	7	48.7	48.1	10
AC030A	<i>Achnanthes umara</i> J.R. Carter 1970	11	1.2	6.1	0.5	10	3.3	1.3	5	96.6	69.2	8
AC9975	<i>Achnanthes</i> [<i>altaica</i> var. <i>minor</i>] L. Granoch (RJF) 1988	31	20.7	4.9	0.3	31	2.6	2.2	28	160.3	67.7	24
AC9969	<i>Achnanthes</i> [<i>scotica</i> / <i>marginulata</i>] Groningen (RJF) 1988	20	3.0	6.1	0.5	19	1.9	0.9	16	53.4	42.1	18
AT001A	<i>Actinella punctata</i> Lewis	13	1.1	5.1	0.3	13	5.3	1.4	8	133.3	30.4	9
AM001A	<i>Amphora ovalis</i> (Kutz.) Kutz. 1844	4	1.4	6.2	0.5	3	2.8	1.1	4	50.5	38.7	3
AM001D	<i>Amphora ovalis</i> var. <i>affinis</i> (Kutz.) Van Heurck 1885	16	1.8	6.4	0.5	13	4.8	2.4	12	62.5	51.9	16

Taxon code	Taxon name and authority	Total data-set		pH			DOC			Al		
		N	Max	O	T	N	O	T	N	O	T	N
AM001B	<i>Amphora ovalis</i> var. <i>pediculus</i> (Kutz.) Van Heurck 1885	11	1.3	6.8	0.2	11	4.9	1.9	4	108.2	62.1	9
AS001A	<i>Asterionella formosa</i> Hassall 1850	30	18.5	6.7	0.4	27	6.2	1.3	15	107.8	45.9	23
AS003A	<i>Asterionella ralfsii</i> W. Sm. 1856	11	25.5	4.9	0.4	11	3.6	0.7	7	125.0	27.9	8
AU002A	<i>Aulacoseira ambigua</i> (Grun. in Van Heurck) Simonsen 1979	13	67.3	6.5	0.4	12	7.6	2.0	8	114.2	41.9	12
AU005A	<i>Aulacoseira distans</i> (Ehrenb.) Simonsen 1979	31	11.9	5.4	0.6	30	2.0	1.5	20	104.6	57.4	19
AU005L	<i>Aulacoseira distans</i> var. <i>humilis</i> (A. Cleve-Euler) R. Ross in Hartley 1986	2	3.8	6.8	0.1	2	9.1	-	1	69.4	24.7	2
AU005J	<i>Aulacoseira distans</i> var. <i>laevisissima</i> (Grun.) Haworth	2	9.2	6.0	0.1	2	1.7	0.0	2	27.6	11.1	2
AU005E	<i>Aulacoseira distans</i> var. <i>nivalis</i>	50	9.6	5.0	0.5	49	2.0	1.6	38	132.8	64.0	39
AU005B	<i>Aulacoseira distans</i> var. <i>nivaloides</i> Camburn 1987	6	1.1	6.1	0.7	5	9.1	0.9	3	100.1	41.9	4
AU005D	<i>Aulacoseira distans</i> var. <i>tenella</i> (Nygaard) R. Ross in Hartley 1986	30	31.3	5.8	0.6	28	9.9	2.3	20	146.5	30.1	22
AU003D	<i>Aulacoseira granulata</i> (Ehrenb.) Simonsen 1979	3	2.4	7.1	0.2	2				64.6	32.3	3
AU003B	<i>Aulacoseira granulata</i> var. <i>angustissima</i> (O. Mull.) Simonsen 1979	2	2.4	7.0	0.4	2	5.4	-	1	76.7	4.9	2
AU009A	<i>Aulacoseira islandica</i> (O. Mull.) Simonsen 1979	2	2.7	6.5	0.2	2				137.5	18.9	2
AU009B	<i>Aulacoseira islandica</i> subsp. <i>helvetica</i> (O. Mull.) Simonsen 1979	2	1.8	5.4	0.2	2	9.3	1.4	2	144.9	16.1	2
AU001C	<i>Aulacoseira italica</i> var. <i>valida</i> (Grun. in Van Heurck) Simonsen 1979	6	2.4	5.6	0.6	6	6.7	3.1	5	107.7	50.4	5
AU004A	<i>Aulacoseira lirata</i> (Ehrenb.) R. Ross in Hartley 1986	52	14.9	5.2	0.7	51	3.6	2.5	33	101.8	63.6	36
AU004C	<i>Aulacoseira lirata</i> form <i>biseriata</i>	9	2.0	5.7	0.8	9	6.5	2.7	6	85.1	28.7	7
AU004D	<i>Aulacoseira lirata</i> var. <i>alpigena</i> (Grun.) Haworth	10	9.4	5.3	0.4	9	2.5	1.5	10	88.8	28.3	8
AU004B	<i>Aulacoseira lirata</i> var. <i>lacustris</i> (Grun. in Van Heurck) R. Ross 1986	21	3.4	6.0	0.7	21	4.7	3.6	16	78.7	50.9	15
AU014A	<i>Aulacoseira nygaardii</i> Camburn	21	3.7	5.7	0.5	21	5.1	2.6	16	100.1	42.7	19
AU010A	<i>Aulacoseira perglabra</i>	52	5.2	5.2	0.6	49	4.1	2.3	41	113.3	39.8	35
AU010B	<i>Aulacoseira perglabra</i> var. <i>floriniae</i>	45	10.4	5.2	0.4	42	3.6	2.4	33	113.5	36.4	33
AU022A	<i>Aulacoseira subborealis</i> SWAP 1989	4	12.0	6.9	0.6	4	7.2	1.2	3	81.0	15.7	4
AU023A	<i>Aulacoseira tethera</i> Haworth nov. sp. 1989	1	5.3	5.0	-	1	0.5	-	1	223.0		1
AU9988	<i>Aulacoseira</i> [cf. <i>distans distans</i>] SWAP 1989	13	2.8	5.0	0.3	13	3.1	1.9	9	134.4	34.9	10
AU9984	<i>Aulacoseira</i> [lirata <i>alpigena</i> (small)] SWAP Sweden (IR & NJA) 1989	10	6.8	5.6	0.5	10	8.8	1.8	7	140.2	31.3	10
AU9987	<i>Aulacoseira</i> [lirata <i>alpigena/distans</i> girdle view] SWAP 1989	13	1.8	5.6	0.4	12	6.2	4.3	9	121.9	50.3	11
AU9983	<i>Aulacoseira</i> [subarctica <i>agg.</i>] SWAP 1989	33	19.1	6.4	0.6	29	6.4	1.8	18	114.7	48.9	29
BR006A	<i>Brachysira brebissonii</i> R. Ross in Hartley 1986	149	9.2	5.3	0.6	143	3.0	2.1	110	112.1	57.1	103
BR003A	<i>Brachysira serians</i> (Breb. ex Kutz.) Round & Mann 1981	65	14.5	4.8	0.4	64	2.9	1.7	42	158.5	55.2	39
BR004A	<i>Brachysira styriaca</i> (Grun. in Van Heurck) R. Ross 1986	29	1.6	5.2	0.6	29	3.9	1.5	19	115.0	49.1	15
BR001A	<i>Brachysira vitrea</i> (Grun.) R. Ross in Hartley 1986	149	48.3	5.9	0.7	140	3.0	2.1	105	74.8	50.3	102
BR005A	<i>Brachysira zellensis</i> (Grun.) Round & Mann 1981	2	3.1	6.7	0.1	2	7.3	-	1	102.3	15.4	2
BR9997	<i>Brachysira</i> [sp. 1] Round L. Glenhead (VJJ) 1985	18	9.9	5.4	0.5	18	5.0	2.3	14	105.4	29.0	11

Taxon code	Taxon name and authority	Total data-set		pH			DOC			Al		
		N	Max	O	T	N	O	T	N	O	T	N
CA018A	<i>Caloneis tenuis</i> Gregory (Krammer) 1985	3	1.3	6.6	0.0	3	6.6	1.4	3	88.7	32.7	3
CO001A	<i>Cocconeis placentula</i> Ehrenb. 1838	13	3.3	6.7	0.3	12	5.1	1.0	5	137.5	36.5	10
CO001B	<i>Cocconeis placentula</i> var. <i>euglypta</i> (Ehrenb.) Grun. 1884	2	10.1	6.9	-	1	5.7	0.1	2	28.6	18.3	2
CC001A	<i>Cyclostephanus dubius</i> (Fricke in A. Schmidt) Round 1982	2	10.5	7.2	-	1				69.9	13.7	2
CY013A	<i>Cyclotella antiqua</i> W. Sm. 1853	1	1.8	4.5	-	1						
CY010A	<i>Cyclotella comensis</i> Grun. in Van Heurck 1882	32	41.5	6.7	0.3	29	4.9	1.5	21	66.0	43.3	25
CY001A	<i>Cyclotella comta</i> (Ehrenb.) Kutz. 1849	40	15.4	6.7	0.4	39	6.1	2.1	21	85.6	43.8	31
CY007A	<i>Cyclotella glomerata</i> Bachm. 1911	21	19.5	6.7	0.4	20	6.0	1.8	12	88.4	62.5	17
CY9991	<i>Cyclotella kuetzingiana</i> agg.	106	72.3	6.3	0.6	101	3.2	2.1	77	81.1	53.1	82
CY003A	<i>Cyclotella meneghiniana</i> Kutz. 1844	2	5.2	6.9	-	1	6.6	-	1	172.8	2.8	2
CY002A	<i>Cyclotella pseudostelligera</i> Hust. 1939	11	26.1	6.9	0.2	10	5.8	0.7	5	128.3	55.0	9
CY004A	<i>Cyclotella stelligera</i> (Cleve & Grun. in Cleve) Van Heurck 1882	25	7.9	6.2	0.5	24	7.0	2.3	13	91.9	43.4	19
CM014A	<i>Cymbella aequalis</i> W. Sm. ex Grev. 1855	51	20.6	5.1	0.3	48	1.2	1.1	40	166.5	72.7	38
CM015A	<i>Cymbella cesatii</i> (Rabenh.) Grun. in A. Schmidt 1881	24	4.5	6.4	0.5	23	4.1	2.2	16	43.4	23.3	16
CM015B	<i>Cymbella cesatii</i> var. <i>capitata</i> Krieger 1933	2	2.7	6.3	0.1	2	0.5	-	1	16.3	-	1
CM038A	<i>Cymbella delicatula</i> Kutz. 1849	8	1.5	6.3	1.0	8	5.0	1.4	4	58.6	13.9	4
CM052A	<i>Cymbella descripta</i> (Hust.) Krammer & Lange-Bertalot 1985	46	6.0	6.4	0.6	44	4.2	2.3	30	65.9	44.6	31
CM051A	<i>Cymbella elginensis</i> Krammer 1981	11	1.1	6.2	0.6	11	4.6	2.4	8	66.6	46.5	9
CM020A	<i>Cymbella gaeumannii</i> Meister 1934	19	2.6	5.9	0.5	19	2.5	2.5	16	97.0	61.8	13
CM017A	<i>Cymbella hebridica</i> (Grun. ex Cleve) Cleve 1894	109	6.6	5.1	0.5	104	2.3	1.8	78	116.2	54.6	69
CM013A	<i>Cymbella helvetica</i> Kutz. 1844	8	1.4	6.7	0.3	8	3.8	2.4	4	68.7	44.9	6
CM048A	<i>Cymbella lunata</i> W. Sm. in Grev. 1855	124	7.3	5.7	0.6	118	3.2	2.0	92	84.2	56.2	91
CM004A	<i>Cymbella microcephala</i> Grun. in Van Heurck 1880	58	8.8	6.3	0.7	53	3.6	2.4	44	52.0	36.1	45
CM031A	<i>Cymbella minuta</i> Hilse ex Rabenh. 1862	52	4.0	6.1	0.7	46	3.6	2.3	37	89.6	55.5	39
CM031C	<i>Cymbella minuta</i> var. <i>silesiaca</i> (Bleisch ex Rabenh.) Reimer 1975	6	2.5	5.3	0.3	5	1.8	1.4	6	68.2	31.0	4
CM010A	<i>Cymbella perpusilla</i> A. Cleve 1895	127	19.4	5.2	0.5	120	2.7	1.7	91	107.7	63.7	88
CM101B	<i>Cymbella scotica</i> W. Sm. 1853	7	1.4	6.7	0.6	7	6.6	1.3	7	61.6	32.6	7
CM050A	<i>Cymbella subaequalis</i> Grun. in Van Heurck 1880	2	1.2	7.1	0.2	2	5.9	2.5	2	37.5	0.9	2
CM9989	<i>Cymbella</i> [cf. <i>deliculata</i>] SWAP-IR 1988	6	1.3	5.1	0.4	6	4.9	1.4	4	136.2	23.0	4
CM9995	<i>Cymbella</i> [PIRLA sp. 1] PIRLA 1985	4	10.0	5.0	0.3	3	1.6	1.2	4	62.1	1.6	2
DE001A	<i>Denticula tenuis</i> Kutz. 1844	14	4.3	6.8	0.4	13	3.9	2.5	7	66.4	39.2	9
DT002A	<i>Diatoma hyemale</i> (Roth) Heib. 1863	8	1.5	5.8	0.4	8	1.5	0.7	7	43.0	28.1	7
DT004B	<i>Diatoma tenue</i> var. <i>elongatum</i> Lyngb. 1819	15	7.7	6.5	0.5	12	5.5	1.4	6	159.2	30.0	11
DT003A	<i>Diatoma vulgare</i> Bory 1824	5	2.8	6.3	0.3	5	3.6	1.1	4	52.7	50.1	3
EU013A	<i>Eunotia arcus</i> Ehrenb. 1837	47	1.4	5.8	0.7	44	2.4	1.6	26	104.5	61.3	32

Taxon Code	Taxon Name	Total Dataset		pH			DOC			Al		
		N	Max	O	T	N	O	T	N	O	T	N
EU014A	<i>Eunotia bactriana</i> Ehrenb. 1854	42	7.6	4.7	0.3	42	3.0	1.5	26	154.6	57.7	26
EU022A	<i>Eunotia bigibba</i> Kutz. 1849	4	0.4	5.3	0.4	4	2.5	0.2	2	93.0	-	1
EU049A	<i>Eunotia curvata</i> (Kutz.) Lagerst. 1884	60	4.9	5.5	0.6	57	3.9	1.9	40	101.9	49.5	42
EU049B	<i>Eunotia curvata</i> var. <i>subarcuata</i> (Naegeli ex Kutz.) Woodhead & Tweed 1954	34	1.8	5.3	0.5	33	3.6	2.7	21	136.6	39.9	24
EU015A	<i>Eunotia denticulata</i> (Breb. ex Kutz.) Rabenh. 1864	70	9.4	5.1	0.4	69	1.7	1.2	55	143.6	71.6	50
EU9969	<i>Eunotia denticulata</i> /lapponica 1987	5	2.9	5.0	0.1	5	0.7	0.3	5	130.2	56.7	4
EU016A	<i>Eunotia diodon</i> Ehrenb. 1837	20	1.2	5.3	0.5	18	3.9	2.0	14	127.7	43.4	12
EU057A	<i>Eunotia exgracilis</i> A. Berg ex A. Cleve-Euler 1953	38	5.6	5.3	0.5	35	1.3	0.9	27	96.7	47.1	25
EU009A	<i>Eunotia exigua</i> (Breb. ex Kutz.) Rabenh. 1864	126	11.0	5.1	0.5	121	2.4	1.6	93	114.2	54.3	85
EU009C	<i>Eunotia exigua</i> var. <i>tridentula</i>	10	1.3	5.1	0.4	10	5.3	2.2	4	154.5	26.7	6
EU025A	<i>Eunotia fallax</i> A. Cleve 1895	25	1.9	5.1	0.8	24	2.9	2.0	15	129.3	61.4	14
EU017A	<i>Eunotia flexuosa</i> Kutz. 1849	48	1.4	5.7	0.6	48	3.1	2.2	35	97.0	64.0	33
EU019A	<i>Eunotia iatriaensis</i> Foged 1970	35	1.4	5.3	0.6	33	3.6	2.4	22	109.5	41.9	24
EU047A	<i>Eunotia incisa</i> W. Sm. ex Greg. 1854	156	28.0	5.1	0.5	145	3.0	1.7	107	119.9	53.4	109
EU020A	<i>Eunotia meisteri</i> Hust. 1930	54	6.6	5.2	0.5	48	3.7	1.4	34	115.0	49.7	32
EU028A	<i>Eunotia microcephala</i> Krasske ex Hust. 1932	26	3.1	4.9	0.4	25	3.7	2.0	16	138.1	46.7	16
EU028B	<i>Eunotia microcephala</i> var. <i>tridentata</i> (A. Mayer) Hust.	6	1.4	4.7	0.3	6	2.7	2.1	5	117.7	49.7	5
EU056A	<i>Eunotia minutissima</i> A. Cleve-Euler 1934	44	2.6	5.3	0.5	43	2.2	1.7	35	117.1	64.2	34
EU048A	<i>Eunotia naegelii</i> Migula 1907	115	10.9	5.0	0.4	111	2.9	2.0	83	130.2	52.4	78
EU040A	<i>Eunotia paludosa</i> Grun. 1862	38	1.3	5.1	0.6	38	3.4	2.0	29	118.1	52.4	31
EU034A	<i>Eunotia parallela</i> Ehrenb. 1843	9	2.0	5.1	0.4	9	5.6	1.7	6	150.6	45.3	7
EU002A	<i>Eunotia pectinalis</i> (O.F. Mull.) Rabenh. 1864	79	3.2	5.5	0.7	74	3.2	2.2	57	99.4	57.6	57
EU002B	<i>Eunotia pectinalis</i> var. <i>minor</i> (Kutz.) Rabenh. 1864	134	5.2	5.4	0.6	128	3.4	1.8	94	126.7	60.5	95
EU002E	<i>Eunotia pectinalis</i> var. <i>minor</i> form <i>impressa</i> (Ehr.) Hust.	50	1.3	5.5	0.6	48	3.8	1.9	33	115.1	54.5	33
EU002D	<i>Eunotia pectinalis</i> var. <i>undulata</i> (Ralfs) Rabenh. 1864	7	1.1	5.4	0.4	7	2.4	0.5	6	68.9	45.2	6
EU002K	<i>Eunotia pectinalis</i> var. <i>ventricosa</i> Grun. in Van Heurck 1881	58	2.7	5.2	0.6	56	3.3	2.1	45	107.5	51.5	45
EU046C	<i>Eunotia perpusilla</i> Grun. in Van Heurck 1881	4	2.2	4.7	0.2	4	3.0	0.8	2	193.7	8.8	2
EU003A	<i>Eunotia praerupta</i> Ehrenb. 1843	22	3.7	5.4	0.7	20	2.1	2.4	14	64.6	46.5	15
EU011A	<i>Eunotia rhomboidea</i> Hust. 1950	92	7.0	5.1	0.5	86	3.2	1.9	64	112.3	50.2	61
EU058A	<i>Eunotia schwabei</i> Krasske	9	1.7	4.7	0.2	9	2.8	1.9	8	111.2	44.5	7
EU031A	<i>Eunotia septentrionalis</i> Ostr. 1898	7	1.2	5.5	0.6	7	8.8	-	1	113.0	-	1
EU021A	<i>Eunotia sudetica</i> O. Mull. 1898	23	2.5	5.6	0.5	23	2.2	1.9	17	70.0	51.9	14
EU004A	<i>Eunotia tenella</i> (Grun. in Van Heurck) A. Cleve 1895	104	6.9	5.2	0.5	98	3.0	1.6	76	114.4	53.5	69
EU027A	<i>Eunotia trinacria</i> Krasske 1929	24	6.2	4.8	0.4	22	1.9	1.8	16	84.2	52.4	14
EU039A	<i>Eunotia triodon</i> Ehrenb. 1837	25	2.4	5.1	0.6	24	2.2	1.3	18	146.3	61.4	15

Taxon Code	Taxon Name	Total Dataset		pH			DOC			Al		
		N	Max	O	T	N	O	T	N	O	T	N
EU051A	<i>Eunotia vanheurckii</i> Patr. 1958	78	4.0	5.1	0.6	74	3.2	1.5	58	122.5	52.7	54
EU051B	<i>Eunotia vanheurckii</i> var. <i>intermedia</i> (Krasske) Cleve	40	5.1	5.4	0.5	40	3.5	2.8	30	156.8	42.6	25
EU9962	<i>Eunotia</i> [<i>exigua</i> var. 2 (kummer.)] Cranmer Pond (RJF) 1988	5	1.4	5.5	0.5	5	2.3	0.3	4	74.9	30.1	4
EU9965	<i>Eunotia</i> [<i>sp.</i> 10 (<i>minima</i>)] L. Grannoch (RJF) 1988	25	4.7	4.9	0.4	24	3.6	1.6	22	136.6	50.6	15
EU9961	<i>Eunotia</i> [<i>vanheurckii</i> var. 1] Round L. Glenhead (RJF) 1988	39	2.9	5.2	0.5	38	2.4	1.3	31	118.4	58.7	28
FR006A	<i>Fragilaria brevistriata</i> Grun. in Van Heurck 1885	30	9.5	6.5	0.4	28	4.4	3.4	16	58.5	44.0	27
FR009F	<i>Fragilaria capucina</i> var. <i>lanceolata</i> Grun. in Van Heurck 1881	1	2.4	6.2	-	1	0.5	-	1	13.0	-	1
FR010A	<i>Fragilaria constricta</i> Ehrenb. 1843	21	3.3	5.4	0.4	19	5.3	2.8	14	134.4	47.8	14
FR002A	<i>Fragilaria construens</i> (Ehrenb.) Grun. 1862	24	8.0	6.6	0.6	21	5.9	2.9	13	96.8	41.8	20
FR002C	<i>Fragilaria construens</i> var. <i>venter</i> (Ehrenb.) Grun. in Van Heurck 1881	54	27.6	6.2	0.7	49	5.4	2.1	33	104.2	56.1	44
FR008A	<i>Fragilaria crotonensis</i> Kitton 1869	6	1.5	7.0	0.2	6	6.1	0.4	2	89.5	28.7	5
FR018A	<i>Fragilaria elliptica</i> Schum. 1867	35	34.3	6.6	0.5	30	6.2	2.0	21	64.9	51.0	29
FR011A	<i>Fragilaria lapponica</i> Grun. in Van Heurck 1881	4	2.2	6.0	0.4	3	2.9	3.6	2	81.7	65.9	3
FR015A	<i>Fragilaria lata</i> Renberg	23	9.2	5.5	0.5	23	9.1	3.8	11	164.1	23.2	11
FR001A	<i>Fragilaria pinnata</i> Ehrenb. 1843	43	10.4	6.3	0.6	37	4.0	2.3	23	80.8	64.4	32
FR001B	<i>Fragilaria pinnata</i> var. <i>lancettula</i> (Schum.) Hust. in A. Schmidt 1913	2	1.6	6.3	0.7	2	2.3	1.1	2	43.2	35.2	2
FR007A	<i>Fragilaria vaucheriae</i> (Kutz.) J.B. Petersen 1938	65	9.0	6.3	0.6	57	4.4	2.5	40	98.0	54.9	48
FR005A	<i>Fragilaria virescens</i> Ralfs 1843	8	0.5	5.6	0.8	7	2.7	1.8	7	117.8	71.7	6
FR005D	<i>Fragilaria virescens</i> var. <i>exigua</i> Grun. in Van Heurck 1881	122	51.1	5.7	0.5	113	3.4	2.5	87	79.2	48.7	92
FR9991	<i>Fragilaria</i> [cf. <i>oldenburgiana</i> PIRLA pl 20, 61-2] PIRLA 1987	45	27.9	4.7	0.3	45	4.7	2.0	29	187.6	54.0	27
FU002A	<i>Frustulia rhomboides</i> (Ehrenb.) De Toni 1891	114	11.8	5.1	0.5	109	3.4	2.3	82	133.4	54.2	77
FU002B	<i>Frustulia rhomboides</i> var. <i>saxonica</i> (Rabenh.) De Toni 1891	162	32.5	5.2	0.6	154	3.5	2.0	116	116.6	56.2	113
FU002F	<i>Frustulia rhomboides</i> var. <i>viridula</i> (Breb. ex Kutz.) Cleve 1894	99	15.1	5.3	0.5	95	4.0	2.3	71	120.0	66.0	68
GO006C	<i>Gomphonema acuminatum</i> var. <i>coronatum</i> (Ehrenb.) W. Sm. 1853	41	1.0	6.1	0.6	38	4.6	2.4	31	103.3	53.1	31
GO003A	<i>Gomphonema angustatum</i> (Kutz.) Rabenh. 1864	34	3.6	5.8	0.6	31	3.2	1.7	26	82.4	57.1	24
GO004A	<i>Gomphonema gracile</i> Ehrenb. 1838	67	4.2	5.8	0.7	62	3.1	1.9	50	85.1	49.3	51
GO013A	<i>Gomphonema parvulum</i> (Kutz.) Kutz. 1849	23	4.2	6.2	0.6	20	4.3	2.0	15	128.1	61.9	19
GO023A	<i>Gomphonema truncatum</i> Ehrenb. 1832	6	1.3	6.7	0.5	6	4.5	0.8	4	130.9	4.1	4
GO025B	<i>Gomphonema vibrio</i> var. <i>intricatum</i> (Kutz.) R. Ross in Hartley 1986	31	2.7	6.6	0.5	27	5.3	1.9	15	104.1	51.5	20
GO025F	<i>Gomphonema vibrio</i> var. <i>pumilum</i> (Grun. in Van Heurck) R. Ross 1986	2	1.8	6.6	0.5	2	5.8	-	1	138.3	-	1
GY005A	<i>Gyrosigma acuminatum</i> (Kutz.) Rabenh. 1853	2	1.3	7.2	-	1				38.3	13.9	2
HN001A	<i>Hannaea arcus</i> (Ehrenb.) Patr. in Patr. & Reimer 1966	10	1.2	6.0	0.6	9	3.8	1.0	6	108.2	51.8	6
ME019A	<i>Melosira arentii</i> (Kolbe) Nagumo & Kobayasi 1977	15	0.8	5.6	0.6	15	5.4	2.8	10	138.7	38.8	8
NA113A	<i>Navicula acceptata</i> Hust. 1950	10	3.2	6.2	0.7	10	2.7	2.0	7	54.0	51.8	8
NA037A	<i>Navicula angusta</i> Grun. 1860	59	5.6	5.6	0.6	56	2.6	1.6	48	77.0	51.6	42

Taxon Code	Taxon Name	Total Dataset		pH			DOC			Al		
		N	Max	O	T	N	O	T	N	O	T	N
NA038A	<i>Navicula arvensis</i> Hust.	15	1.1	5.2	0.6	14	4.8	2.3	9	127.4	50.2	10
NA084A	<i>Navicula atomus</i> (Kutz.) Grun. 1860	1	1.3	6.6	-	1	1.8	-	1	26.4	-	1
NA099A	<i>Navicula bremensis</i> Hust. 1957	33	2.6	5.0	0.5	31	2.0	1.7	23	119.5	53.9	19
NA045A	<i>Navicula bryophila</i> J.B. Petersen 1928	25	2.4	5.6	0.6	25	4.5	2.1	18	126.1	63.7	16
NA032A	<i>Navicula cocconeiformis</i> Greg. ex Greville 1855	75	3.1	5.7	0.6	72	2.6	1.7	57	81.1	47.1	57
NA046A	<i>Navicula contenta</i> Grun. in Van Heurck 1885	12	0.9	5.8	0.7	12	2.1	2.0	12	83.2	69.5	11
NA007A	<i>Navicula cryptocephala</i> Kutz. 1844	17	1.8	6.6	0.4	15	4.4	2.1	10	108.3	83.2	13
NA158A	<i>Navicula cumbriensis</i> Haworth 1987	42	4.7	4.9	0.3	41	3.3	1.6	32	125.0	44.7	31
NA115A	<i>Navicula difficillima</i> Hust. 1950	11	2.2	5.0	0.3	8	4.3	1.5	8	84.9	32.9	8
NA149A	<i>Navicula digitulus</i> Hust. 1943	17	2.2	6.0	0.6	17	1.6	1.2	15	37.5	27.9	14
NA151A	<i>Navicula gysingensis</i> Foged 1952	3	1.4	6.4	0.5	3	2.0	2.0	2	27.4	25.2	2
NA015A	<i>Navicula hassiaca</i> Krasske 1925	15	1.6	5.4	0.7	15	1.9	1.3	7	98.4	52.6	10
NA167A	<i>Navicula hoefleri</i> Sensu Ross et Sims	88	11.1	4.9	0.4	84	3.1	1.6	65	140.2	58.3	56
NA068A	<i>Navicula impexa</i> Hust. 1961	24	4.7	5.8	0.6	21	4.1	2.5	14	144.9	63.2	18
NA016A	<i>Navicula indifferens</i> Hust. 1942	27	1.8	5.9	0.7	25	5.4	2.3	16	143.0	71.5	19
NA101A	<i>Navicula jaagii</i> Meister 1934	5	1.4	6.9	0.6	4	7.1	0.5	2	57.8	50.2	2
NA002A	<i>Navicula jaernefeltii</i> Hust. 1942	22	1.4	5.5	0.7	22	4.5	1.2	12	120.0	47.3	15
NA044A	<i>Navicula krasskei</i> Hust. 1930	29	13.9	5.3	0.4	28	1.1	0.7	25	94.0	55.5	23
NA102A	<i>Navicula laevissima</i> Kutz. 1844	6	7.3	6.6	0.4	6	3.6	1.2	4	22.9	22.4	4
NA156A	<i>Navicula leptostriata</i> Jorgensen 1948	102	18.7	5.1	0.4	97	2.5	1.7	76	121.7	49.2	68
NA140A	<i>Navicula madumensis</i> E.G. Jorg. 1948	42	6.6	4.9	0.4	40	1.8	1.4	30	148.8	60.9	24
NA006A	<i>Navicula mediocris</i> Krasske 1932	110	2.3	5.4	0.6	105	2.8	2.0	75	104.8	52.8	74
NA006B	<i>Navicula mediocris</i> var. <i>atomus</i> Hust.	14	5.7	5.2	0.5	14	1.5	0.7	13	80.1	24.0	14
NA042A	<i>Navicula minima</i> Grun. in Van Heurck 1880	39	8.7	6.1	0.5	33	2.6	2.0	25	61.0	48.4	25
NA112D	<i>Navicula minuscula</i> var. <i>muralis</i> (Grun. in Van Heurck) Lange-Beralot 1981	7	3.3	6.4	0.3	7	2.0	1.8	5	44.2	44.9	7
NA013A	<i>Navicula pseudoscutiformis</i> Hust. 1930	36	2.5	6.1	0.6	31	3.5	2.2	22	62.4	50.4	26
NA014A	<i>Navicula pupula</i> Kutz. 1844	52	4.0	6.3	0.6	47	4.4	2.3	33	71.9	57.6	39
NA003A	<i>Navicula radiosa</i> Kutz. 1844	62	5.4	6.2	0.8	55	4.9	2.2	36	72.1	57.5	39
NA003B	<i>Navicula radiosa</i> var. <i>tenella</i> (Breb. ex Kutz.) Grun. ex Van Heurck 1885	71	8.4	6.1	0.7	66	3.7	2.6	50	64.7	42.1	55
NA008A	<i>Navicula rhyncocephala</i> Kutz. 1844	21	2.5	6.6	0.4	17	5.9	1.5	12	99.6	62.3	17
NA133A	<i>Navicula schassmannii</i> Hust. 1937	13	4.9	5.9	0.6	13	1.7	2.0	12	40.7	36.3	12
NA129A	<i>Navicula seminuloides</i> Hust. 1937	9	12.3	6.2	0.5	9	1.5	0.3	7	33.3	19.9	9
NA005A	<i>Navicula seminulum</i> Grun. 1860	41	2.4	6.1	0.6	38	3.4	2.7	25	88.1	59.0	34
NA005B	<i>Navicula seminulum</i> var. <i>intermedia</i> Hust.	13	1.2	6.4	0.6	12	7.1	1.6	9	128.5	36.5	10
NA048A	<i>Navicula soehrensensis</i> Krasske 1923	13	24.5	5.1	0.2	13	3.4	2.3	8	230.6	50.3	9

Taxon Code	Taxon Name	Total Dataset		pH			DOC			Al		
		N	Max	O	T	N	O	T	N	O	T	N
NA170A	<i>Navicula strenzkii</i> Hust.	3	4.6	6.1	0.4	3	1.1	0.2	3	22.1	1.6	3
NA043A	<i>Navicula subatomoides</i> Hust. ex Patr. 1945	5	1.4	6.5	0.4	5	2.1	1.2	5	41.9	47.2	5
NA160A	<i>Navicula submolesta</i> Hust. 1949	13	1.4	5.5	0.5	13	1.3	0.8	11	74.1	33.5	12
NA114A	<i>Navicula subrotundata</i> Hust. 1945	2	1.3	6.5	0.2	2	1.8	-	1	26.4	-	1
NA033A	<i>Navicula subtilissima</i> Cleve 1891	30	2.2	5.2	0.6	28	1.5	0.8	19	116.9	66.0	17
NA086A	<i>Navicula tantula</i> Hust. 1943	9	5.5	6.3	0.4	7	4.2	1.8	8	42.8	29.9	7
NA135A	<i>Navicula tenuicephala</i> Hust. 1942	23	32.5	5.3	0.2	23	0.9	1.0	12	87.0	33.3	15
NA063A	<i>Navicula trivialis</i> Lange-Bertalot 1980	5	2.9	6.6	0.1	5	4.8	0.9	2	96.4	30.7	3
NA9964	<i>Navicula</i> [cf. <i>spirata</i>] L. Hir (SF) 1986	16	1.3	5.3	0.6	15	2.4	1.6	13	80.6	36.3	13
NA9955	<i>Navicula</i> [cf. <i>vitiosa</i>] L. Hir (SF) 1986	16	2.3	6.1	0.6	14	3.3	2.2	10	116.2	75.2	11
NA9919	<i>Navicula</i> [cumbriensis (small)] Gulspettvatn (IR) 1988	11	2.7	6.0	0.7	11	6.5	1.1	7	114.7	37.6	9
NA9963	<i>Navicula</i> [sp. 1] L. Hir (SF) 1986	29	7.1	5.6	0.5	27	2.0	1.2	21	62.2	30.0	21
NA9904	<i>Navicula</i> [sp. a Merilainen 1969: fig. 143] SWAP Sweden (IR & NJA) 1989	10	1.6	6.1	0.5	9	7.2	1.6	5	105.7	44.2	8
NE003A	<i>Neidium affine</i> (Ehrenb.) Pfitz. 1871	47	2.2	5.3	0.6	45	3.5	1.8	35	119.4	76.7	34
NE003C	<i>Neidium affine</i> var. <i>amphirhynchus</i> (Ehrenb.) Cleve 1894	17	2.6	5.7	0.5	17	4.9	2.5	11	120.4	33.6	11
NE003B	<i>Neidium affine</i> var. <i>longiceps</i> (Greg.) Cleve 1896	20	2.4	5.0	0.4	18	3.1	2.3	15	152.7	83.0	15
NE004A	<i>Neidium bisulcatum</i> (Lagerst.) Cleve 1894	58	2.6	5.1	0.6	57	2.7	1.7	45	128.1	54.7	44
NE012A	<i>Neidium glaberrimum</i> (Ostr.) R. Ross in Hartley 1986	6	1.9	5.5	0.4	6	0.8	1.3	5	83.9	52.8	5
NE020A	<i>Neidium hercynicum</i> A. Mayer 1917	5	2.1	5.5	0.4	4	1.3	0.8	5	64.4	46.1	5
NE023A	<i>Neidium inconspicuum</i> Hust. 1921	8	1.1	5.7	0.7	7	1.4	0.9	8	58.3	43.9	6
NI021A	<i>Nitzschia acula</i> Hantzsch ex Cleve & Grun. 1880	14	1.7	5.9	0.6	14	5.1	2.3	10	90.0	48.5	10
NI002A	<i>Nitzschia fonticola</i> Grun. in Van Heurck 1881	49	4.9	6.3	0.6	43	3.6	2.3	36	59.4	45.4	39
NI008A	<i>Nitzschia frustulum</i> (Kutz.) Grun. in Cleve & Grun. 1880	23	1.4	6.4	0.4	19	3.1	2.1	18	72.9	54.0	20
NI017A	<i>Nitzschia gracilis</i> Hantzsch 1860	30	2.2	5.9	0.5	28	5.1	2.0	17	107.0	42.4	20
NI027A	<i>Nitzschia microcephala</i> Grun. in Cleve & Grun. 1880	7	8.5	6.5	0.2	7	2.2	1.0	4	35.5	30.2	7
NI009A	<i>Nitzschia palea</i> (Kutz.) W. Sm. 1856	34	2.3	6.2	0.6	30	4.3	2.3	22	79.2	50.4	24
NI009B	<i>Nitzschia palea</i> var. <i>tenuirostris</i> Grun. in Van Heurck 1881	9	2.8	6.2	0.4	9	3.2	1.6	9	64.5	44.7	8
NI005A	<i>Nitzschia perminuta</i> (Grun. in Van Heurck) M. Perag. 1903	73	5.6	5.7	0.5	71	2.8	1.9	55	68.1	37.5	55
NI152A	<i>Nitzschia pusilla</i> Grun. 1862	7	2.7	5.8	0.6	6	2.0	1.3	7	40.9	36.4	6
NI026A	<i>Nitzschia romana</i> Grun. in Van Heurck 1881	12	14.9	6.3	0.5	11	3.8	1.6	7	152.3	21.7	4
NI9984	<i>Nitzschia</i> [cf. <i>gracilis</i>] SWAP Sweden (IR & NJA) 1989	1	3.5	7.3	-	1	7.3	-	1	38.0	-	1
OP001A	<i>Opephora martyi</i> Herib. 1902	6	1.8	6.5	0.5	6	2.2	1.9	3	110.1	40.4	6
PE002A	<i>Peronia fibula</i> (Breb. ex Kutz.) R. Ross 1956	135	9.7	5.3	0.5	128	2.3	1.6	96	112.7	59.6	91
PI015A	<i>Pinnularia abaujensis</i> (Pant.) R. Ross in Hartley 1986	69	3.7	5.6	0.8	67	3.4	1.9	44	91.6	57.7	49
PI014A	<i>Pinnularia appendiculata</i> (Ag.) Cleve 1896	9	2.8	5.8	0.6	8	2.6	1.3	8	76.5	62.5	71

Taxon Code	Taxon Name	Total Dataset		pH			DOC			Al		
		N	Max	O	T	N	O	T	N	O	T	N
PI055A	<i>Pinnularia balfouriana</i> Grun. ex Cleve 1896	1	1.4	6.6	-	1	1.4	-	1	20.0	-	1
PI018A	<i>Pinnularia biceps</i> Greg. 1856	95	4.7	5.2	0.6	93	2.7	2.2	66	122.9	60.0	65
PI018B	<i>Pinnularia biceps</i> form <i>petersenii</i> R. Ross 1947	9	2.0	5.3	0.5	9	1.7	1.5	8	99.5	69.1	8
PI016A	<i>Pinnularia divergentissima</i> (Grun.in Van Heurck) Cleve 1896	9	1.6	5.9	0.5	7	2.6	1.9	7	44.1	31.9	4
PI023A	<i>Pinnularia irrorata</i> (Grun. in Van Heurck) Hust. 1939	47	3.0	5.4	0.7	43	4.4	1.7	28	136.8	69.3	32
PI005A	<i>Pinnularia major</i> (Kutz.) W. Sm. 1853	22	2.2	5.3	0.5	21	4.7	2.2	14	134.3	55.9	16
PI011A	<i>Pinnularia microstauron</i> (Ehrenb.) Cleve 1891	81	3.7	5.4	0.6	80	2.8	2.4	59	100.3	49.5	56
PI139A	<i>Pinnularia obscura</i> Krasske 1932	1	1.2	5.5	-	1	0.7	-	1	37.9	-	1
PI056A	<i>Pinnularia rupestris</i> Hantzsch in Rabenh. 1861	21	3.4	4.9	0.5	21	4.7	2.2	8	120.8	20.7	9
PI022B	<i>Pinnularia subcapitata</i> var. <i>hilsiana</i> (Janisch ex Rabenh.) O. Mull. 1898	73	9.4	5.0	0.5	69	3.1	2.1	54	115.4	53.9	50
PI164A	<i>Pinnularia termitina</i> (Ehrenb.) Patr. in Patr. & Reimer 1966	8	3.4	5.5	0.7	8	2.2	1.1	7	74.0	45.1	6
PI007A	<i>Pinnularia viridis</i> (Nitzsch) Ehrenb. 1843	41	1.4	5.5	0.7	37	4.0	2.5	30	107.7	63.5	30
RH006B	<i>Rhopalodia operculata</i> (Ag.) Hakansson 1979	2	1.4	6.3	0.7	2	2.3	1.7	2	50.1	59.7	2
SE001A	<i>Semiorbis hemicyclus</i> (Ehrenb.) Patr. in Patr. & Reimer 1966	42	24.8	4.8	0.3	42	3.4	1.4	25	177.2	43.1	24
SA001A	<i>Stauroneis anceps</i> Ehrenb. 1843	51	2.1	5.6	0.6	47	3.0	2.0	34	82.5	56.3	36
SA001B	<i>Stauroneis anceps</i> form <i>gracilis</i> Rabenh. 1864	62	1.8	5.9	0.6	58	2.6	1.8	44	67.9	51.3	45
SA042A	<i>Stauroneis gracillima</i> Hust. 1943	16	1.3	5.1	0.5	14	3.1	1.5	16	118.0	63.1	14
SA006A	<i>Stauroneis phoenicenteron</i> (Nitzsch) Ehrenb. 1943	27	1.8	5.9	0.8	24	3.6	2.1	18	100.6	85.0	18
SP002A	<i>Stenopterobia sigmatella</i> (Greg.) R. Ross in Hartley 1986	56	2.8	5.3	0.7	53	3.8	2.3	37	156.3	56.0	38
ST004A	<i>Stephanodiscus minutula</i> (Kutz.) Round	5	2.1	6.9	0.2	4	6.6	1.5	2	108.0	40.2	5
ST010A	<i>Stephanodiscus parvus</i> Stoermer & Hakansson 1984	1	8.6	7.1	-	1	4.6	-	1	133.3	-	1
SU004A	<i>Surirella biseriata</i> Breb. & Godey 1835	21	4.0	5.3	0.4	21	0.8	0.7	20	125.1	48.7	18
SU006A	<i>Surirella delicatissima</i> Lewis 1864	84	8.2	5.0	0.4	81	2.7	2.1	62	129.9	63.2	60
SU005A	<i>Surirella linearis</i> W. Sm. 1853	40	3.6	5.3	0.6	39	4.2	2.2	29	131.9	50.6	28
SU002A	<i>Surirella ovata</i> Kutz. 1844	2	1.0	5.5	-	1	0.7	-	1	37.2	1.7	2
SY003A	<i>Synedra acus</i> Kutz. 1844	27	3.5	6.4	0.5	24	6.0	1.8	16	89.1	52.9	23
SY043A	<i>Synedra famelica</i> Kutz. 1844	2	2.8	5.9	0.3	2	4.8	0.3	2	107.6	-	1
SY010A	<i>Synedra minuscula</i> Grun. in Van Heurck 1881	29	3.6	6.0	0.6	26	3.7	1.5	23	117.3	70.8	23
SY009A	<i>Synedra nana</i> Meister 1912	5	2.7	6.2	0.7	5	5.8	0.5	4	80.8	62.6	3
SY004A	<i>Synedra parasitica</i> (W. Sm.) Hust. 1930	9	1.4	6.7	0.4	8	4.2	1.9	4	51.6	40.0	9
SY002A	<i>Synedra rumpens</i> Kutz. 1844	20	2.6	6.4	0.6	18	4.5	1.8	12	80.6	30.0	14
SY013A	<i>Synedra tenera</i> W. Sm. 1856	26	4.3	6.6	0.7	24	6.0	1.3	12	113.5	43.3	17
TA003A	<i>Tabellaria binialis</i> (Ehrenb.) Grun. in Van Heurck 1881	63	51.5	4.7	0.3	60	2.6	1.7	43	171.3	50.5	41
TA002A	<i>Tabellaria fenestrata</i> (Lyngb.) Kutz. 1844	18	1.0	5.9	0.7	18	4.0	2.1	15	110.6	54.9	13
TA001A	<i>Tabellaria flocculosa</i> (Roth) Kutz. 1844	171	28.3	5.4	0.6	161	3.4	1.9	118	106.0	51.1	122

Taxon Code	Taxon Name	Total Dataset		pH			DOC			Al		
		N	Max	O	T	N	O	T	N	O	T	N
TA9996	Tabellaria flocculosa agg.	44	24.5	6.1	0.7	43	5.6	2.2	28	108.7	41.5	32
TA004A	Tabellaria quadrisepata Knudson 1952	89	47.8	4.9	0.3	87	2.7	1.4	66	119.3	45.0	58