

Use of Urea to Correct Immature Urban Composts for Agricultural Purposes

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ABSTRACT

Municipal solid waste composts are often inadequately stabilized for agricultural purposes. In addition, compost quality may be even more reduced by loss of nitrogen (N) during the composting process. We have utilized a compost with a high content of soluble sugars (11 mg g⁻¹, DM, indicating immaturity) and a low N concentration (0.95%, DM). The compost had a low level of heavy metals. Results obtained in a germination bioassay conducted with cress, ryegrass and sunflower in a compost-sand mixture reflected the *immaturity* of the compost. Such composts should be fortified with N (in a complete fertilizer, when possible), at the same time avoiding an intimate contact with the soil (e.g., plowing down). When the compost (and raw wastes and wastes at the 4th week of composting) was mixed with a soil at a heavy rate (2.5 % w:w), ryegrass seedling emergence in pots was not affected, but the plantlets' fresh weight in the compost treatment was significantly lower than that in the control (soil) and lower than that in the raw wastes, probably due to the lower N concentration. As expected, plantlet fresh weight was notably increased by the combination of compost and wastes with a complete fertilizer. The application of compost in combination with a complete fertilizer or urea did not affect either dry matter production or nutrient uptake of ryegrass,

despite the combination's being applied just at sowing (in pots). Results obtained in these experiments indicate that combining immature composts with urea [supplemented with phosphorus (P) and potassium (K), when possible] at a ratio of about 50:1 (about 200 kg urea per 10 t compost) could be sufficient to prevent negative results in crop establishment. Such practices could contribute to overcoming the limited fertilizing capacity of the composts.

INTRODUCTION

The benefits of adding manure and composts to soils are extensively reported (De Bertoldi et al., 1987; Barker, 1997). However, negative effects of using urban composts in agriculture, derived from their pollution capacity and/or immaturity, have also been detected (De Haan, 1981; Mathur et al., 1993; Déportes et al., 1995). Thus, the regulation of compost quality, besides inerts and pollutants (metals, PCB), frequently calls for the control of additional requirements for compost stability. For this purpose, indicators such as volatile solids destruction, spontaneous heating, O₂ uptake rates, toxin production, C/N ratio, seed germination and growth tests, and redox potential have been cited by Gies (1992). If no determination of stability is made, the compost *product* (inerts and pollutants within permissible levels) must be cured for a six month period (Gies, 1992).

However, as pointed out by Zucconi et al. (1985), many of the urban composts used in agriculture are not completely stabilized, because the time required for complete stabilization is often incompatible with the space, handling, equipment and other requirements of industry. This may lead to agricultural problems, being the most serious 'N rob' in plants planted immediately after treating the soil with immature composts (Smith, 1996). Moreover, problems derived from a restriction of O₂ supply for plants may arise when composts are plowing down into compacted soils (Avnimelech et al., 1993). Nitrogen application would be desirable in such cases, especially when an adequate separation between planting and compost loading were troublesome for farmers.

Even when mature, most composts contain relatively low levels of nutrients, with a low rate of mineralization, so that the compost dose needed to meet crop requirements may be too large. This could be a problem for most farmers, because even if such a large rate of compost application were economically viable, it would be outside the limits set by current legislation in many countries (Murillo et al., 1995a; Sikora, 1996). Normal doses of composts must therefore be fortified with complete fertilizers. As pointed out by Sikora (1996), blending residues or composts with fertilizers is an appealing alternative which (a) uses residues at lower rates, (b) reduces the amount of inorganic fertilizer applied to soils, and (c) reduces the accumulation of non-nutrient ingredients in soils. If, moreover, this practice avoids the risks of 'N rob' and of a restricted O₂ supply for plants, it is obvious that fortified compost product shows promise as a fertilizer (Hileman, 1982).

The present paper deals with the effect of N application on ryegrass growth, when a heavy rate of an immature compost is mixed with the soil. The effect (on the dry matter production and composition of a ryegrass) of applying increasing doses of urea together with the compost is also studied to obtain information about reasonable compost:urea ratios for the best use of these immature composts.

MATERIALS AND METHODS

Analyses of the Wastes, Compost, and Plant Material

The urban compost from the 'Villarrasa' facility (Huelva, Spain: fraction <10 mm, available for sale) was utilized as organic fertilizer. A complete 15N-15P₂O₅-15K₂O fertilizer and urea (46% N, supplemented with P and K) were used to fortify the compost. The composts were obtained by a monitored bio-oxidative process taking about 10 to 12 weeks. Ten subsamples were taken at random from several piles of the final compost, in order to obtain a composite sample from each pile. Samples from the initial, raw wastes (fraction <60 mm) and those from wastes at the 4th week of composting (fraction <60 mm) were also sampled. Raw wastes, wastes (4th week), and the final compost will be referred to as W₀, W₄, and C, respectively, hereafter.

Samples were dried at 70°C for 72 h, and visible inerts (>2 mm), such as glass and plastics (fraction that usually ranges from 1-2% in the final compost, but up to 50% in the raw wastes) were removed. Dried samples were then ground to a particle size <2 mm to avoid heterogeneity for analysis. The pH and electrical conductivity were measured by suspending the sample in water in the proportion 1:5 (w:v). Nitrogen of wastes and compost (and plant material) was determined by Kjeldahl digestion, and total organic matter by mineralization at 500°C. The cation exchange capacity was determined in the compost according to Harada and Inoko (1980). Soluble sugars in wastes and compost were extracted with water and determined by the phenol-sulphuric acid method (Ashwell, 1966), and total lipids were determined by extraction with a dichloromethane:methanol solution.

Nutrients (compost and plant material) and heavy metals (compost) were extracted by treating the ashes obtained by mineralization with concentrated HCl on a hot plate. Potassium and sodium (Na) were determined by flame emission, and calcium (Ca), magnesium (Mg), iron (Fe), manganese (Mn), zinc (Zn), copper (Cu), chromium (Cr), nickel (Ni), lead (Pb), and cadmium (Cd) by atomic absorption spectrometry. Phosphorus was determined by colorimetric determination using the phosphovanadomolybdic complex.

Bioassays of Germination and Seedling Emergence

The test was carried out for cress (*Lepidium sativum* L.), ryegrass (*Lolium multiflorum* Lam. cv. 'Tewera') and sunflower (*Helianthus annuus* L. cv. 'Hysum 33') at room temperature, using a sand-compost (or wastes, W₀, W₄) mixture (1:1).

Compost and wastes were homogenized by grinding to a particle size <2 mm. Twenty Petri dishes (diameter: 5.5 cm), containing either the mixture or only sand (controls), were prepared for each species (5 seeds for cress and ryegrass, and 4 seeds for sunflower). The dishes were moistened and lined with a filter paper. Germinated seeds and root length were recorded after 72 h and expressed as a percentage of the control. A germination index was obtained as a product of the two percentages dividing by 100.

Seedling emergence and performance of ryegrass in a representative soil of SW Spain (a light yellowish-brown sandy clay loam soil, pH 7.7, CaCO_3 30%) was also tested in the greenhouse using wastes (W_0 , W_4), and the compost (C), applied alone or fortified with the inorganic 15-15-15 fertilizer (IF). Compost (or wastes)-soil mixtures at 2.5% [w:w, 6.25 g C (or wastes) per pot] were used in pots of 300 g capacity (treatments C, W_0 , W_4 , respectively). Treatment IF (0.42 g IF per pot) was prepared so as to apply roughly the same amount of N per pot as that added in treatments C, W_0 and W_4 (about 0.25 g N kg^{-1} soil).

Treatments W_0 + IF, W_4 + IF, and C + IF were also prepared by adding N at a rate of 0.25 g kg^{-1} [0.21 g IF pot^{-1} and 3.125 g compost (waste) pot^{-1}]. A control was prepared without any fertilizer or waste. Before preparing the test, soil, wastes, and compost were ground to pass a 2-mm sieve for homogenization. A randomized, complete block design with six replicates per treatment and 10 seeds of ryegrass per pot was prepared. Seedling emergence was monitored for 20 days. The five most-developed plantlets of each pot were cut 15 days after sowing and the mean fresh above-ground plantlet weight was immediately obtained using an analytical balance. The other 5 plants were discarded. Three months after the first sowing (bare pots were periodically moistened by subirrigation), each pot was sown again with 10 seeds of ryegrass to obtain the mean fresh plantlet weight 30 days after this second sowing.

Compost Correction with Urea and Inorganic Fertilizer (IF)

Pots of 2 kg capacity, and the same soil used for the mixtures of the seedling emergence assay (1.5 kg per each pot), were used in this experiment, in which N was also applied at a rate of 0.25 g kg^{-1} . Compost was applied at a rate 37.5 g pot^{-1} (2.5% w:w), IF at a rate of 2.52 g pot^{-1} , and urea (treatment U) at a rate of 0.81 g pot^{-1} . Rating of compost, IF and urea was reduced to one half in treatments C + IF (C: 18.75 g pot^{-1} , IF: 1.26 g pot^{-1}) and C + U (C: 18.75 g pot^{-1} , U: 0.40 g pot^{-1}). Soil without any fertilizer was used as a control. Soil and compost were ground to particle size <2 mm to avoid heterogeneity. A randomized, complete block design with four replicates per treatment and one gram of seeds of ryegrass per pot was prepared. Fresh and dry matter (drying at 70°C for 48 h) production per pot was obtained 40 days after sowing and 60 days after the first cut. Plant moisture was obtained by drying a representative sample at 105°C for two hours. After the drying at 70°C, plant material was ground for analysis.

Compost Correction with Increasing Doses of Urea

This assay was conducted to study the effect of three compost:urea ratios on the ryegrass growth. Urea was selected as a cheap source of N. Pots of 2 kg capacity were again used. For avoiding compaction and ensure an adequate oxygen supply, a soil-perlite mixture (5:1, w:w) was used as substrate (control) (750 g of substrate per each pot). Soil was ground to particle size <2 mm, but the compost (fraction <10 mm) was applied without grinding, at a rate of 1.5% (11.25 g pot⁻¹). Compost (1.5%) was combined with increasing doses of urea (46% N) in treatments: C + U₁ (0.245 g urea pot⁻¹, about 0.30 g N kg⁻¹ substrate, in total); C + U₂ (0.49 g urea pot⁻¹, 0.45 g N kg⁻¹ substrate); and C + U₃ (0.735 g urea pot⁻¹, 0.60 g N kg⁻¹ substrate). Compost:urea ratios were, respectively, about 50:1, 25:1, and 12.5:1 in these treatments. Treatments applying only urea (U₁, U₂, and U₃) were also established for comparison. Urea was supplemented with P (0.247 g pot⁻¹ of superphosphate, 35% P₂O₅) and K (0.225 g pot⁻¹ of potassium sulphate, 54% K₂O) in all the treatments.

A randomized, complete block design with four replicates per treatment and one gram of seeds of ryegrass per pot was prepared. Fresh and dry matter production per pot was obtained 40 days after sowing and in two further cuts carried out at monthly intervals.

Statistical Analyses

Data analysis was conducted performing analysis of variance. Mean separations were determined by the Tukey test ($P < 0.05$).

RESULTS AND DISCUSSION

Compost Characterization

Biogenic waste composts have clear advantages for agricultural use, as has been shown in the literature (Fricke et al., 1989; Vogtmann et al., 1993). However, many countries do not have adequate facilities for producing biogenic composts at the rate and price required by local agriculture. A scrupulous separation process at the composting facilities could largely overcome the risks presented by heavy metals and other constraints. Composts produced in facilities receiving wastes from villages, such as the compost used here, may also contain reasonably low amounts of heavy metals, as shown in Table 1. Heavy metal concentrations are lower, except Cr, than the permissible levels indicated by Zucconi and De Bertoldi (1987). Copper and Cr contents are higher than the more restrictive levels of Gies (1992) and Vogtmann et al. (1993) with Pb in the upper limit of these ranges. In fact, concentrations of most heavy metals in the compost from the Villarrasa facility correspond to composts of high quality, according to the data of Genevini et al. (1997).

TABLE 1. Analysis of the compost utilized (particle size <10 mm, mean values of 8 determinations), and normal ranges for the urban composts from the 'Villarrasa' facility (15 batches). Values specified in the literature are also shown (samples were ground in all cases to a particle size <2 mm for homogenization).

Property	Compost	Range	A	B	C
Total organic matter %	38	23 - 42	30 - 35		27.2 - 39.6
N %	0.95	0.55 - 1.35	>0.6		0.88 - 1.47
pH - H ₂ O (compost/water 1/5)	6.9	6.3 - 7.8	5.5 - 8.0		7.2 - 7.8
C/N	22	13 - 25	<22		12.3 - 20.0
Electrical conductivity dS m ⁻¹ (compost/water 1/5)	7.6	4.7 - 9.0			
Cation exchange capacity cmol _c kg ⁻¹ (on organic matter basis)	72.9	57.6 - 82.7			
P ₂ O ₅ %	0.89	0.58 - 1.04	0.5 - 0.9		0.38 - 0.80
K ₂ O %	0.90	0.40 - 1.28	0.2 - 0.8		0.63 - 1.37
CaO %	5.32	3.30 - 9.09	>2.0		2.14 - 5.50
MgO %	0.45	0.28 - 0.85	>0.3		0.38 - 1.09
Na %	0.63	0.28 - 0.88			
Fe %	1.05	0.50 - 1.40			
Cu mg kg ⁻¹	180	93 - 204	300	60	100
Zn mg kg ⁻¹	320	185 - 374	1000	500	400
Mn mg kg ⁻¹	155	119 - 366			
Cr mg kg ⁻¹	190	52 - 267	150	50	100
Ni mg kg ⁻¹	32	23 - 41	50	60	50
Pb mg kg ⁻¹	152	58 - 187	750	150	150
Cd mg kg ⁻¹	<1	<1	5	3	1.5

¹A: Specifications for solid waste composts: recommended limits for heavy metals and desirable levels of nutrients, inerts and organic matter for marketable products (Zucconi and De Bertoldi, 1987).

²B: Specifications for regulating compost quality in Ontario (Gies, 1992).

³C: Levels of organic matter, pH, and nutrients for biogenic waste compost and limits for heavy metals established by BGGK (Vogtmann et al., 1993).

Heavy metals apart, low compost quality can also result from an incomplete stabilization, which can make the compost inappropriate for agricultural use without adequate evolution in the soil prior to planting. Data on soluble sugars content (Table 2), seem to indicate that the compost contained non-stabilized substances, and a high content of total lipids. Table 2 also shows a loss of N during the composting process [also shown by Soliva et al. (1993) in other Spanish facilities],

TABLE 2. Evolution of selected properties during composting in the 'Villarrasa' facility (W_0 , wastes; W_4 , fourth week of composting; C, final compost (particle size <10 mm). Samples were ground in all cases to a particle size <2 mm for homogeneization (mean values of 8 determinations).

Sample	Organic matter (%)	C/N	N (%)	Soluble sugars (mg g^{-1})	Total lipids (%)
W_0	46	17	1.38	15	5.9
W_4	45	22	1.06	11	7.9
C	38	22	0.95	11	8.1

which could enhance a possible 'N rob' in plants (besides possible problems of O_2 restriction, derived from its immaturity) if the compost were applied close to planting and/or plowing down into the soil.

Bioassays of Germination and Seedling Emergence

Chemical analyses apart, bioassays of instant toxicity should always be tested with these kinds of compost to avoid unexpected negative effects on plant development, especially at the plantlet stage. A simple germination bioassay in a sand-compost mixture may be sufficiently informative (Murillo et al., 1995b). This bioassay was carried out, and showed the immaturity of the urban compost used (Table 3). Wastes before composting completely inhibited cress, ryegrass and sunflower germination. After four weeks of composting (treatment W_4), only the sunflower seeds showed some germination capacity, but with an extremely low root elongation, which would corroborate the assertion of Zucconi et al. (1981) that toxicity during composting was strictly associated with the initial 3-4 weeks, and subsequently decreased rapidly, although it had not completely disappeared after two months (residual toxicity).

Finally, the compost obtained (treatment C) completely inhibited the germination of cress and ryegrass. Only sunflower germinated to a certain extent, but with very low root elongation. This showed immaturity of the compost, as symptoms of toxicity are more pronounced at an early stage of root growth (ISTA, 1985), and can certainly cause root shortening, even at low concentrations (Zucconi et al., 1985). The GI for sunflower was thus extremely low, despite the germination percentage's reaching a reasonably high value in relation to the control (Table 3).

When this low-quality compost was incorporated (mixed) into a calcareous soil at a high rate (2.5%, w:w), ryegrass seedling emergence was not affected, despite the germination sensitivity of *Lolium* genus (Murillo et al., 1993). Seedling emergence in soil was not only unaffected by C, but also by W_0 and W_4 treatments (Table 4). The negative effect of wastes and compost (treatments W_0 , W_4 , and C)

TABLE 3. Germination bioassay for cress, ryegrass, and sunflower in a compost/sand mixture (1/1, mean values at 72 hours).

Species	Treatment	Germination (% of control)	Root length (mm)	Germination index (%)
Cress	Control	-	19.9	-
	W ₀	0	0	0
	W ₄	0	0	0
	C	0	0	0
Ryegrass	Control	-	10.8	-
	W ₀	0	0	0
	W ₄	0	0	0
	C	0	0	0
Sunflower	Control	-	7.5	-
	W ₀	0	0	0
	W ₄	18.2	0.2	0.5
	C	63.6	0.5	4.7

TABLE 4. Mean values of seedling emergence of ryegrass in soil (1st sowing, day 21) and fresh weights of the ryegrass plantlets (above-ground part) of the first sowing (day 15) and of the second (day 30), carried out in the same pots three months after the first sowing.

Treatments	Seedling emergence	Plantlets weight (mg)	
	(% of the control)	first sowing	second sowing
Control	100 a*	49.4 c	49.4 bc
IF	104 a	44.1 cb	83.0 d
W ₀	100 a	25.6 a	58.8 c
W ₀ + IF	100 a	45.4 c	121.6 ef
W ₄	112 a	22.8 a	38.3 ab
W ₄ + IF	112 a	42.7 cb	111.3 e
C	100 a	21.0 a	25.5 a
C + IF	108 a	37.3 b	132.4 f

*Values followed by the same letter in the same column do not differ significantly ($P < 0.05$).

was shown by the plantlets' fresh weight, which was, in general, significantly lower than that of the control and, of course, that obtained in treatments IF, $W_0 + IF$, $W_4 + IF$, and C + IF for the two sets of plantlets generated (Table 4).

In the first sowing, however, the increases in plantlet weight caused by treatments $W_0 + IF$, $W_4 + IF$, and C + IF could only be due to the reduction in the concentration of the wastes and compost (from 2.5% to 1.25%), because at this time (15 days after establishing treatments) the inorganic fertilizer did not seem to have had any marked effect on plantlet growth, as inferred from results obtained in treatments including IF. The positive effect of these treatments, especially those including wastes and the compost, was evident in the one-month plantlets of the second sowing (Table 4). The heavy application of compost (2.5%, w:w, treatment C) was still restricting plantlet growth, despite its having been four months in the soil.

Treatment C tended to produce plantlets with even lower weight than those of W_0 and W_4 treatments. The highest weights in these three treatments corresponded to W_0 , with a significant difference in relation to C in the second sowing (Table 4). This could be due, at least in part, to the comparatively high N concentration of the raw wastes (Table 2), which could make their evolution in the soil favorable for plantlets, which in the second sowing achieved a higher weight (mean value of 58.8 mg plantlet⁻¹) than the control (mean value of 49.4 mg plantlet⁻¹).

Compost Correction: Urea 'Versus' a Complete Fertilizer in Ryegrass Growth and Nutrition

An adequate N supply may overcome, at least partially, both the risk of a restriction in O_2 supply, being noticeable when using well particulated composts with low content of visible inerts (as is that studied here), and, at the same time, problems derived from a possible 'N rob'. As expected, the positive effect of a complete fertilizer on plant growth and nutrition was higher than with the sole application of urea (Tables 5 and 6). Treatment IF yielded a total dry matter production twice that obtained in treatment U, the same amount of N per pot being applied in both.

The high rate of compost application (2.5% w:w, treatment C) caused the lowest dry matter production (especially in the first cut, in which it was around 50% lower than that of the control) (Table 5). In the second cut, the two treatments yielded similar dry matter productions, but it is necessary to bear in mind that the ryegrass (sown at a rate of 1 g seed pot⁻¹) accomplished a greater initial consumption of nutrients in the control than in treatment C.

The application of compost together with IF or urea (treatments C + IF and C + U) did not cause any negative effect on plant growth, despite the compost's (and fertilizers') being applied at sowing. It can be seen (Table 5) that these treatments yielded similar (treatment C + IF) or higher (treatment C + U) dry matter production in the first cut than did treatments IF and U. Nevertheless, the consumption of fertilizers by plants of the first cut could reduce the dry matter production of the second cut by around 65% (treatment C + IF) and 42% (treatment C + U). The

TABLE 5. Mean values of dry matter production (above-ground part) and nutrient accumulation (on a dry matter basis) in the above-ground part of the ryegrass.

Treatment	Dry matter production (g pot ⁻¹)			Nutrient accumulation (both cuts, mg pot ⁻¹)			
	1st cut	2nd cut	Total	N	P	K	Mn
Control	0.72 b*	0.24 a	0.96 a	18.9 a	1.8 a	42.1 a	0.18 b
IF	2.13 e	2.88 d	5.01 c	226.0 d	16.0 e	293.6 e	0.81 e
U	1.22 c	1.33 c	2.55 b	111.0 c	4.1 b	131.5 b	0.45 d
C	0.35 a	0.23 a	0.57 a	10.2 a	1.3 a	21.3 a	0.07 a
C + IF	2.11 e	0.73 b	2.83 b	96.2 cb	10.1 d	167.5 d	0.27 c
C + U	1.58 d	0.92 bc	2.51 b	89.1 b	5.5 c	137.6 bc	0.29 c

*Values followed by the same letter in the same column do not differ significantly (P<0.05).

reduction was not found in treatments IF and C (Table 5), both using a twofold amount of fertilizer.

In general, treatments including IF yielded the highest total dry matter productions recorded (two cuts). This was especially so in the case of treatment IF, which led to the highest production (5.01 g pot⁻¹, Table 5). As a consequence, this treatment yielded the highest amounts of accumulated nutrients in the above-ground part of the ryegrass. Nevertheless, urea seemed to be an adequate source of N for fortifying composts, since application (without P and K) yielded a similar dry matter production to that obtained in treatment C + IF.

TABLE 6. Mean values of nutrient accumulation (on a dry matter basis) in the above-ground part of the ryegrass (total of the three cuts).

Treatment	N	P	K	Ca	Mg	Mn	Zn
	----- (mg pot ⁻¹) -----						
Control	12.4 a*	1.8 a	39.1 a	18.3 a	2.6 a	0.12 a	0.05 a
C	22.6 b	3.4 b	75.8 b	15.5 a	3.4 a	0.22 b	0.09 a
U ₁	89.2 c	13.9 c	159.6 c	42.0 b	9.4 b	0.38 c	0.16 ab
U ₂	137.4 d	17.0 ef	211.4 d	55.9 d	12.6 c	0.53 d	0.27 bc
U ₃	159.9 f	15.9 de	197.9 d	71.4 e	15.9 d	0.57 d	0.27 bc
C + U ₁	96.1 c	13.9 c	206.3 d	40.3 b	9.4 b	0.47 cd	0.30 cd
C + U ₂	149.8 e	18.6 f	225.8 e	50.3 c	13.0 c	0.69 e	0.39 de
C + U ₃	175.1 g	15.3 cd	235.7 e	57.3 d	15.1 d	0.71 e	0.44 e

*Values followed by the same letter in the same column do not differ significantly (P<0.05).

The compost tended to enhance (though only slightly) P and K uptake by the ryegrass, since, for a similar total dry matter production, the treatment C + U caused a significantly greater accumulation of P, and also a (non-significantly) greater accumulation of K than did treatment U (Table 5). This effect was especially evident in the second cut (data not shown). In contrast, and also for similar dry matter productions, treatments C + IF and C + U reduced the Mn accumulation around 40% and 35%, respectively, compared with that recorded in treatment U.

This could imply the existence of an antagonism (Jarrell and Beverly, 1981) in the Mn uptake, derived from compost presence. Depressive effects of organic matter on Mn uptake by plants have been reported in the literature (Wallace and Wallace, 1983).

Compost Correction with Increasing Doses of Urea (Supplemented with Phosphorus and Potassium)

In order to minimize possible problems derived from a restricted O₂ supply, the fraction <10 mm was used when correcting the immature, but well particulated composts studied here. Moreover, in this experiment the compost was applied at a rate of 1.5% (not 2.5%) and perlite was included in the substrate, as described above, to prevent soil compaction as much as possible.

In this experiment, ryegrass was cut three times in a period of 100 days (in the above experiment it could be cut only twice, in the same period, due to its poor growth in the control and treatment C). The maximum dry matter production recorded in this experiment was around 5 g pot⁻¹ (treatments U₂, U₃, C + U₂, and C + U₃; Figure 1), similar to that obtained in the treatment IF of the previous assay (Table 5). This corroborates the suitability of using urea as a source of N when correcting composts.

Under the conditions of this experiment, compost (treatment C) always led to higher dry matter production than did the control, despite its application at sowing. On the other hand, the highest doses of urea (treatments U₃ and C + U₃) did not have any positive effect on the dry matter production, when compared with treatments U₂ and C + U₂ (Figure 1).

For similar dry matter productions, treatments C + U₂ and C + U₃ tended to cause higher above-ground accumulations not only of N, K, and Zn than did treatments U₂ and U₃, but also of Mn (Table 6). This could indicate that the above-mentioned depressive effect of compost on Mn uptake could be influenced by, apart from the higher dose of compost applied, the compaction of the substrate. On the other hand, urea (fortified with P and K) seemed to enhance Mn accumulation to some extent, while compost presence tended to depress Ca accumulation (Table 6).

Except for Ca, and bearing in mind that the Ca levels were always satisfactory for covering plant requirements, the application of compost (under the conditions of this experiment) did not interfere in plant nutrition, despite its immaturity and despite its being applied at sowing. Therefore, it seems particularly important not

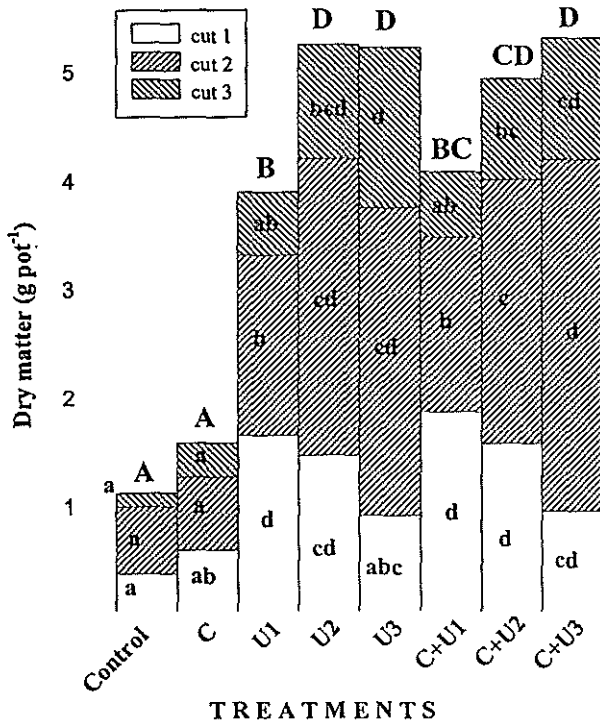


FIGURE 1. Dry matter production of the ryegrass in the different treatments. Bars with the same letter, for each cut (small letters) or in total (capital letters), do not differ significantly.

to restrict aeration of the substrates when using this kind of well particulated compost. Avnimelech et al. (1993) have pointed out that a surface application, or a very shallow mixing with the soil, may ensure a better plant response to compost application.

To avoid an occasional 'N rob', adequate combinations with urea (fortified with P and K when possible) may be an advisable practice when using these composts. This practice would also avoid using large doses of composts for covering crop requirements. In this experiment, in which the compost did not seem to cause any problem of 'N rob', treatment C + U₁ (compost:urea ratio of about 50:1) led to a reasonable dry matter production of ryegrass (a species very demanding for N), not significantly different to that of treatment C + U₂ (compost:urea ratio of 25:1, which yielded one of the maximum productions recorded).

This seems to indicate that for a normal agricultural dose of composts (i.e., 10 t ha⁻¹), recommended for soils in the South of Spain, generally with low organic matter contents, but usually founded incompletely stabilized, the addition of around 200 kg urea ha⁻¹ may be sufficient not only to avoid problems, but also to obtain good results for crop establishment and the first steps of growth. Additional fertilization will depend on each particular situation. The results we are obtaining in the field (in preparation) corroborate this assertion.

CONCLUSIONS

Municipal solid waste compost from 'Villarrasa' facility showed non-stabilized substances which could indicate immaturity, and bioassays of germination proved its instant toxicity for plants. When compost was incorporated into soil, the emergence of plants was not affected, but negative effects were shown by the plantlets' fresh weight. A dose of 200 kg urea ha⁻¹ was sufficient to avoid problems at the first stages of the crop.

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