

1 **Bio-silage of mussel work-processing wastes by lactobacilli on**
2 **semi-solid culture.**

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18 **Headline:** Bio-silage of mussel work-processing wastes by lactobacilli.

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20

21 **ABSTRACT**

22 The aim of this work was to evaluate the fermentability of mussel work-
23 processing wastes by lactic acid bacteria in order to remove and to upgrade a
24 material that generate important focuses of pollution in coastal areas. With this
25 perspective, three lactobacilli (*Lactobacillus casei*, *Lactobacillus plantarum* and
26 *Lactobacillus buchneri*) were employed, and the production of metabolites, as
27 well as the nutrient uptake, was evaluated. The effects of inoculum
28 concentration and previous sterilization process were also studied. The kinetic
29 tests were performed in semi-solid cultures and the results indicated the high
30 feasibility of these materials as substrate for bio-silage production. Cultivations
31 of 24 h led to productions of more than 90 g L⁻¹ of lactic acid and 9 g L⁻¹ of final
32 protein. All the fermentation assays were stable for various days without
33 contaminations by other bacteria.

34

35 **Keywords:** mussel wastes; by-products upgrading; lactobacilli; bio-silage; lactic
36 acid production.

37

38 **INTRODUCTION**

39 Mussel culture production is an economical activity of great importance in
40 Galicia (NW, Spain), since in this region there is currently 3360 tray-mussel
41 farms distributed in 50 cultivation areas along the coast. In 2006, 247000 tons
42 of mussels were obtained what it represents 99% of Spanish production and
43 45% of EU. This extensive production generates a large volume of wastes
44 associated with the different steps from mussel work-processing, collection,

45 transformation and canning. Thus, around 80000 tons per year of shells are
46 wasted from these farms and foodstuff companies. This by-product presents a
47 high fraction of calcium carbonate that is management by means of thermal
48 process with the obtaining of a calcareous material of high purity (Barros et al.,
49 2009a). However, another type of residuals appear in the mussel-production
50 chain. These are mainly originated in the primary works of harvest and
51 processing on the tray-farms what from now we will call mussel work-
52 processing (MWP). Due to their high composition in organic material (rests of
53 mussels, epifauna, zooplankton, phytoplankton), heterogeneity of size and high
54 volume (~35000 tons per year) they provoke serious reductions in the efficiency
55 and in the yield of the thermal process for the treatment of the inorganic
56 material of shell wastes (Barros et al., 2009a). Therefore, it is necessary to
57 develop a complementary and environmental friendly process to allow a global
58 use of the different fractions derived from mussel production (Barros et al.,
59 2009a; Barros et al., 2009b).

60

61 In the last years the most relevant alternative for the use of these materials has
62 been associated with their application as amending and nutrient supplement of
63 degraded grounds (mines, forest burnt). Nevertheless, the results have not
64 been positive and satisfactory due to the high cost of transport for these
65 materials. Thus, these wastes are currently being deposited in non-controlled
66 landfill or, more frequently, dumped directly to the sea. Another possible and
67 more realistic alternative could be obtained by means of biological silage using
68 lactic acid bacteria (Pagarkar et al., 2006). This is an easy and low cost process
69 that generates an acid fermentative product with good nutritious qualities,

70 antimicrobial features against pathogen bacteria and high stability for a long
71 time that can be used as protein supplement for animal feeding (Goddard and
72 Perret, 2005). Different lactic acid bacteria (LAB) have been used for obtaining
73 bio-silage of marine by-products: *Lactobacillus plantarum* (Fagbenro and
74 Jauncey, 1995; Lassen, 1995; Pagarkar et al., 2006), *Lactobacillus brevis*
75 (Uchida et al., 2004), yoghurt-bacteria as *Lactobacillus bulgaricus* and
76 *Streptococcus thermophilus* (Yoon et al., 1997), *Lactobacillus delbruecki* spp.
77 *bulgaricus* and *Streptococcus salivarius* spp. *thermophilus* (Martínez-Valdivieso
78 et al., 1996) as well as *Lactobacillus buchneri* and *Lactobacillus casei* (Vázquez
79 et al., 2008a). In all cases, an additional source of carbohydrates was
80 necessarily added in the form of molasses (Fagbenro and Jauncey, 1995;
81 Hammoumi et al., 1998) or dextrose (Vázquez et al., 2008a). Though the
82 fermentative capacity of many organic nitrogen source from marine waste
83 materials has been tested with excellent results (Dufossé et al., 2001; Ellouz et
84 al., 2001; Vázquez et al., 2004a; Vázquez et al., 2004b; Martone et al., 2005;
85 Aspomo et al., 2005a; Aspomo et al., 2005b; Vázquez et al., 2006; Gao et al.,
86 2006; Vázquez et al., 2008b) as well as wastewaters from thermal process of
87 mussel (González et al., 1992; Pastrana et al., 1993; Murado et al., 1993;
88 Guerra and Castro, 2002; Vázquez et al., 2003; Vázquez et al., 2004c; Guerra
89 et al., 2005), the present work is the first one that studies the useful of by-
90 products from MWP as substrate for bio-silage formulations.

91

92 Based on these considerations, in this manuscript a preliminary study of
93 fermentability of MWP wastes by lactic acid bacteria is reported. A quick and

94 easy fermentative process similar to the bio-silage is established in order to
95 propose an environmental friendly solution for this contaminant material.

96

97 **MATERIALS AND METHODS**

98 ***1: Preparation and composition of material from mussel wastes***

99 A representative sample of MWP wastes (10 kg) was collected from mussel-tray
100 farm placed in Boiro (Ría de Arousa, Galicia, Spain). Its composition was
101 principally shells (45%), remnants of meat (20%), epifauna (15 %), algae (15%)
102 and mud (5%). This sample was subsequently homogenized by milling until
103 obtaining a material with particle size of < 3 mm and maintained (15 d at most)
104 at –20 °C until use. Its chemical characterization was: moisture (61.6%), ash
105 (28.8%) and organic matter (9.6%). The composition of this organic matter was
106 45 % of soluble protein-Lowry and 42 % of total sugars (basically glycogen and
107 glucose).

108

109 ***2: Microbiological methods***

110 The micro-organisms used were *Lactobacillus casei* ssp. *casei* CECT 4043
111 (abbreviated key Lb 3.04), *Lactobacillus plantarum* CECT 220 (Lb 8.01) and
112 *Lactobacillus buchneri* CECT 4111 (Lb 10.01). Stock cultures were stored at –
113 75 °C in MRS medium (Hispanlab) with 25% glycerol (Cabo et al., 2001).
114 Inocula (0.1 g of LAB in final medium, that is, 0.6 % w/w of material to silage)
115 consisted of cellular suspensions from 24-h aged cultures on MRS medium,
116 concentrated by centrifugation (4000. g, 10 min) until cell number required
117 ($\sim 10^{10}$ - 10^{11} cfu mL⁻¹).

118

119 Cultures were carried out in duplicate, using 30 mL Pirex tubes with 7.5 g of
120 homogenised waste material, 7.5 mL of a glucose solution of 200 g L⁻¹ and 1
121 mL of the corresponding inoculum. The experimental conditions were
122 temperature at 30 °C and orbital shaking at 200 rpm. In all cases, initial pH was
123 adjusted to 7.0 with 5 N NaOH and media were sterilised at 101 °C for 60 min.

124

125 **3: Analytical methods**

126 At pre-established times, each experimental unit was divided into two aliquots.
127 The first aliquot of 1 g was used for the quantification of viable cells by means of
128 a plate count technique on MRS agar media. Serial tenfold dilutions were
129 prepared in peptone-buffered solutions, and 0.1 mL samples were plated in
130 quadruplicate, incubated at 30 °C for 48-72 h, and manually counted. Results
131 were expressed as colony-forming units per g (cfu g⁻¹). The second aliquot (the
132 rest of the culture) was centrifuged at 6,000 g for 12 min, and the sediment
133 resuspended with 15 mL of distilled water for a second centrifugation at the
134 same previous conditions. Both supernatants obtained from these two
135 centrifugations were mixture for analytical determinations. These determinations
136 were corrected taking into account the dilution generate by the sediment
137 resuspension. The second sediment was also used to determine protein
138 concentration.

139

140 Additional analyses (in duplicate) were: *Proteins*: method of Lowry et al. (1951).
141 For sediments, this method was applied to the samples with a previous alkaline
142 treatment using NaOH (1M) for 24 h at 30°C as well as the standard of bovine
143 serum albumin. *Total sugars*: measured by means of the phenol-sulphuric

144 reaction (Dubois et al., 1956) according to the application of Strickland and
145 Parsons (1968) with glucose as a standard. *Reducing sugars*: 3,5-
146 dinitrosalicylic reaction (Bernfeld, 1951). *Lactic acid*: HPLC, after membrane
147 filtration (0.22 µm Millex-GV, Millipore, USA) of samples, using an ION-300
148 column (Transgenomic, USA) with 6 mM sulphuric acid as a mobile phase (flow
149 = 0.4 mL min⁻¹) at 65 °C and a refractive-index detector.

150

151 **4: Mathematical equations and numerical methods**

152 The profile of *pH* was modelled by means of von Bertalanffy equation (Vázquez
153 et al., 2005):

154

$$155 \quad pH = pH_f + a \cdot \exp(-ct) \quad \text{with} \quad a = pH_0 - pH_f \quad (1)$$

156

157 where, *t* is the time-course (h), *pH_f* is the final value of *pH*, *a* is the *pH* drop, *pH₀*
158 is the initial value of *pH*, *c* is the specific maximum rate of *pH* drop (h⁻¹).

159

160 The mathematical model used to describe kinetically the sigmoid production of
161 lactic acid was as follows (Vázquez et al., 2008c):

162

$$163 \quad L = \frac{L_m}{1 + \exp\left(2 + \frac{4v_m}{L_m}(\lambda - t)\right)} \quad (2)$$

164

165 where, *L* is the lactic acid (g L⁻¹), *L_m* is the maximum production of lactic acid (g
166 L⁻¹), *v_m* is the maximum rate of lactic acid production (g L⁻¹ h⁻¹) and *λ* is the lag-
167 phase of lactic acid production (h).

168

169 An additional calculation of the yields in the formation of lactic acid (L) referred
170 to both the consumption of reducing sugars and of total proteins (in the
171 supernatant + in the sediment) was obtained in the following terms:

172

$$173 \quad Y_{L/RS} = \frac{\Delta L}{\Delta RS} = \frac{L_f - L_i}{RS_i - RS_f} \quad (3)$$

$$174 \quad Y_{L/P} = \frac{\Delta L}{\Delta P} = \frac{L_f - L_i}{P_i - P_f} \quad (4)$$

175

176 where, $Y_{L/RS}$ is the yield of lactic acid production on reducing sugars (g of lactic
177 acid g^{-1} of reducing sugars) and $Y_{L/P}$ is the yield of lactic acid production on total
178 proteins (g of lactic acid g^{-1} of total proteins).

179

180 On the other hand, fitting procedures and parametric estimations calculated
181 from the results were carried out by minimisation of the sum of quadratic
182 differences between observed and model-predicted values, using the non linear
183 least-squares (Levenberg-Marquadt) method provided by Statistica 8.0
184 (StatSoft, Inc. 2007). This software was also used to evaluate the significance
185 of the parameters estimated by the adjustment of the experimental values to
186 the proposed mathematical models and the consistency of these equations.

187

188 **RESULTS AND DISCUSSION**

189 ***1: Fermentability of MWP by lactobacilli***

190 Initially, the selection of LAB used in the present work (Lb 3.04, Lb 8.01 and Lb
191 10.01) was done in relation to the fermentative features that these bacteria
192 showed in the valorisation of fish viscera (Vázquez et al., 2008a). Using wastes

193 from ray, swordfish and shark as protein source for the formulation of complex
194 microbiological media, high lactic acid (more than 80% of yield as function of
195 glucose uptake) and biomass productions were obtained.

196

197 Figure 1 shows experimental result of lactobacilli fermentations on the MPW
198 waste with supplemental glucose. The time course trends were very similar for
199 three bacteria tested here. In all cases, the drop of pH was ~ 2.5 units with pH_f
200 values down to 4.51. The highest specific maximum rate of pH drop was
201 obtained in Lb 8.01 ($c = 0.235 \pm 0.059 \text{ h}^{-1}$) followed by Lb 10.01 ($c = 0.224 \pm 0.044$
202 h^{-1}) and Lb 3.04 ($c = 0.166 \pm 0.037 \text{ h}^{-1}$). Maximum lactic acid productions were up
203 to 80 g L^{-1} and more than 85% of glucose added was consumed. The maximum
204 rate of lactic acid production were also higher with Lb 8.01 ($v_m = 8.64 \pm 3.50 \text{ g L}^{-1}$
205 h^{-1}) than Lb 3.04 ($v_m = 7.52 \pm 5.05 \text{ g L}^{-1} \text{ h}^{-1}$) and Lb 10.01 ($v_m = 6.30 \pm 3.74 \text{ g L}^{-1} \text{ h}^{-1}$).
206 No other metabolites like acetic acid or ethanol were generated. Yields of
207 lactic acid formation on glucose as substrate were up to 0.81 g of lactic acid g^{-1}
208 of reducing sugars and up to 8.70 g of lactic acid g^{-1} of total proteins for the
209 three lactobacilli. As it has been early commented, similar results were obtained
210 when other marine wastes were used as culture media (Vázquez et al., 2008a).
211 Nevertheless, heterofermentative behaviour was observed with these
212 substrates; from 20 h of culture, conversion of lactic acid in acetic acid was
213 developed. Furthermore, Lb 3.04 led to important ethanol concentrations as
214 response to the stress conditions of pH gradient in fed-batch cultures with
215 successive re-alkalisations (Vázquez et al., 2005). In our work, lactic acid was,
216 however, the only metabolite synthesized what could be due to the different

217 formulation of culture broth in relation to the source of proteins and mineral
218 salts.

219

220 On the contrary, the dynamics of LAB biomass did not show the common
221 sigmoid profiles that are habitually generated by the growth of *Lactobacillus* in
222 batch cultivation (Horn et al., 2005; Kedia et al., 2008; Charalampopoulos et al.,
223 2009). The bacteria counts increased at the first 8 hours of culture (1
224 logarithmic unit, log-unit) and subsequently fall until the value of the inocula. For
225 Lb 10.01 this drop was superior to 1.5 log-units from the maximum growth.
226 Regarding the consumption of organic nitrogen source more than 10 g L⁻¹ of
227 final protein (sum of supernatant and sediment concentrations) is not consumed
228 at the end of the kinetic. This fermented material with probiotic properties
229 (Planas et al., 2004; Guerra et al., 2007; Bernárdez et al., 2008a; Bernárdez et
230 al., 2008b) could be used as feed for poultry farming where high concentrations
231 of calcium carbonate in the alimentary substrate are easily assimilable for the
232 animals. Moreover, this formulation was stable to the microbial contamination
233 for 10 days at 20 °C.

234

235 In all cases, the fitting of pH and lactic acid production were satisfactory
236 statistically. The mathematical equations were consistent (Fisher's *F* test) and
237 the parametric estimations were significant (Student's *t*-test). All the values
238 foreseen in the non-linear adjustments produced high coefficients of linear
239 correlation with the experimental values ($r > 0.97$).

240

241 **2: Effect of inoculum in the fermentability of MWP by *Lactobacillus***
242 ***plantarum***

243 In order to reduce the necessity of inoculum, the next step consisted in the
244 study of the effect of initial Lb 8.01 concentration in the MWP fermentation. This
245 strain was selected due to the highest values of kinetic parameters obtained for
246 lactic acid production and *pH* drop. Time course of semi-solid cultures on
247 mussel-based media at different concentrations of Lb 8.01 are depicted in figure
248 2. The results revealed that with the decrease in the initial LAB employed lower
249 ($L_m = 60.92 \pm 14.57$ and 34.71 ± 7.64 g L⁻¹ with 0.075 and 0.01 g L⁻¹ of initial Lb
250 8.01, respectively) and slower production of lactic acid is obtained. The *pH* drop
251 followed the same tendency as well as the maximum specific rate of *pH* drop
252 ($c = 0.306 \pm 0.078$, 0.205 ± 0.032 , 0.055 h⁻¹) for 0.075, 0.05, and 0.01 g L⁻¹ of Lb
253 8.01 inoculum, respectively. Comparing with the results from previous section,
254 the reduction of one order of magnitude in the initial concentration of LAB
255 entailed a fall of 50 g L⁻¹ in the maximum production of lactic acid. In this case,
256 the yields of lactic acid production on glucose were lower with less inoculum:
257 $Y_{L/RS} = 0.70$ and 0.43 g of lactic acid g⁻¹ of reducing sugars for 0.075 and 0.01 g
258 L⁻¹, respectively. Similar differences were obtained for $Y_{L/P}$ with values of 7.93
259 and 5.60 g of lactic acid g⁻¹ of total proteins.

260

261 On the other hand, similar trends of LAB dynamics and protein evolution
262 throughout the time with previous experiment were observed. An initial rise of
263 bacteria until a maximum growth was reached and then decreased gradually
264 until values comparable with the inoculum. Likewise, correlative profiles of
265 biomass production in relation with the initial LAB concentration were obtained.

266

267 **3: Effect of non-sterilisation in the fermentability of MWP by *Lactobacillus***
268 ***plantarum***

269 In the early sections, the tests were carried out using MWP thermally sterilized
270 with the purpose of evaluating their capacity to be fermented by LAB. However
271 this possibility to industrial scale is less realistic than employing raw material
272 with an added LAB-inoculum to lead the lactic acid fermentation since the cost
273 of sterilization process could make the bio-silage too expensive. Therefore,
274 effect of non previous sterilization was assayed.

275

276 In Figure 3 the outcomes obtained for fermentations in non thermal processing
277 substrate are shown. When compared to Figure 2, not very significant
278 differences between pH profiles and numerical parameters in sterilized and non
279 sterilized media ($c= 0.272, 0.210, 0.054 \text{ h}^{-1}$) were noted. Similar lactic acid
280 productions and glucose consumptions were obtained for both initial conditions
281 of wastes and for different LAB initial concentration. These productions and
282 consumptions were significantly dependent with the inoculum, hence, higher
283 productions ($L_m= 67.55\pm 2.85, 61.51\pm 15.72$ and $42.21\pm 6.07 \text{ g L}^{-1}$) and uptakes
284 were obtained at higher Lb 8.01 inoculum. Moreover, the specific maximum rate
285 of growth was much higher with 0.075 g L^{-1} than 0.05 and 0.01 g L^{-1} of initial
286 LAB. However, in these three cases the parabolic profiles showed in Figures 1
287 and 2 were not developed and more common kinetic trends were got.

288

289 As in previous cultures the proposed equations were statistically robust
290 (Fisher's *F*-test and *p*-values < 0.005), and the parametric estimations were

291 significant (Student's t -test $\alpha= 0.05$). The coefficients of linear correlation
292 between predicted and observed values were in all cases > 0.97 .

293

294 Further experiments should be done in order to study the possibility of
295 incorporation of the fermented material in the formulation of fodder for poultry
296 feeding.

297

298 **CONCLUSIONS**

299 The main conclusion of the present study is that MWP wastes can be fermented
300 with LAB, even under non-sterility conditions whenever it is supplemented with
301 an enough amount of fermentable sugar. Thus, with a no longer culture (20-28
302 h) of lactobacilli a material likely suitable for animal feed is obtained. The
303 process developed in this preliminary work could suggest an easy protocol to
304 reduce impact pollution of MWP on marine ecosystem.

305

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474

475 **FIGURE CAPTIONS**

476

477 **Figure 1:** Biological silage of MWP wastes by Lb 3.04 (●), Lb 8.01 (○) and Lb
478 10.01 (□). L: lactic acid, N: colony forming units per gram (cfu g⁻¹), RS:
479 reducing sugars, Pr: protein-Lowry in supernatant, Pr sed: protein-Lowry in
480 sediment. Continuous lines show the fits of the experimental data (points) to the
481 equations (1) and (2); discontinuous lines only represent the experimental
482 profiles. The corresponding confidence intervals of independent experiments
483 are not shown ($\alpha=0.05$, $n=2$), since these were below 10% of the experimental
484 mean value in all cases.

485

486 **Figure 2:** Bio-silage of MWP wastes by Lb 8.01 with different inocula
487 concentrations (●: 0.075 g L⁻¹, ○: 0.05 g L⁻¹, □: 0.01 g L⁻¹) and previous
488 thermal sterilisation of media. Keys as in figure 1.

489

490 **Figure 3:** Bio-silage of MWP wastes by Lb 8.01 with different inocula
491 concentrations (●: 0.075 g L⁻¹, ○: 0.05 g L⁻¹, □: 0.01 g L⁻¹) and without
492 sterilisation of media. Keys as in figure 1.

493

494

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497

498

FIGURE 1

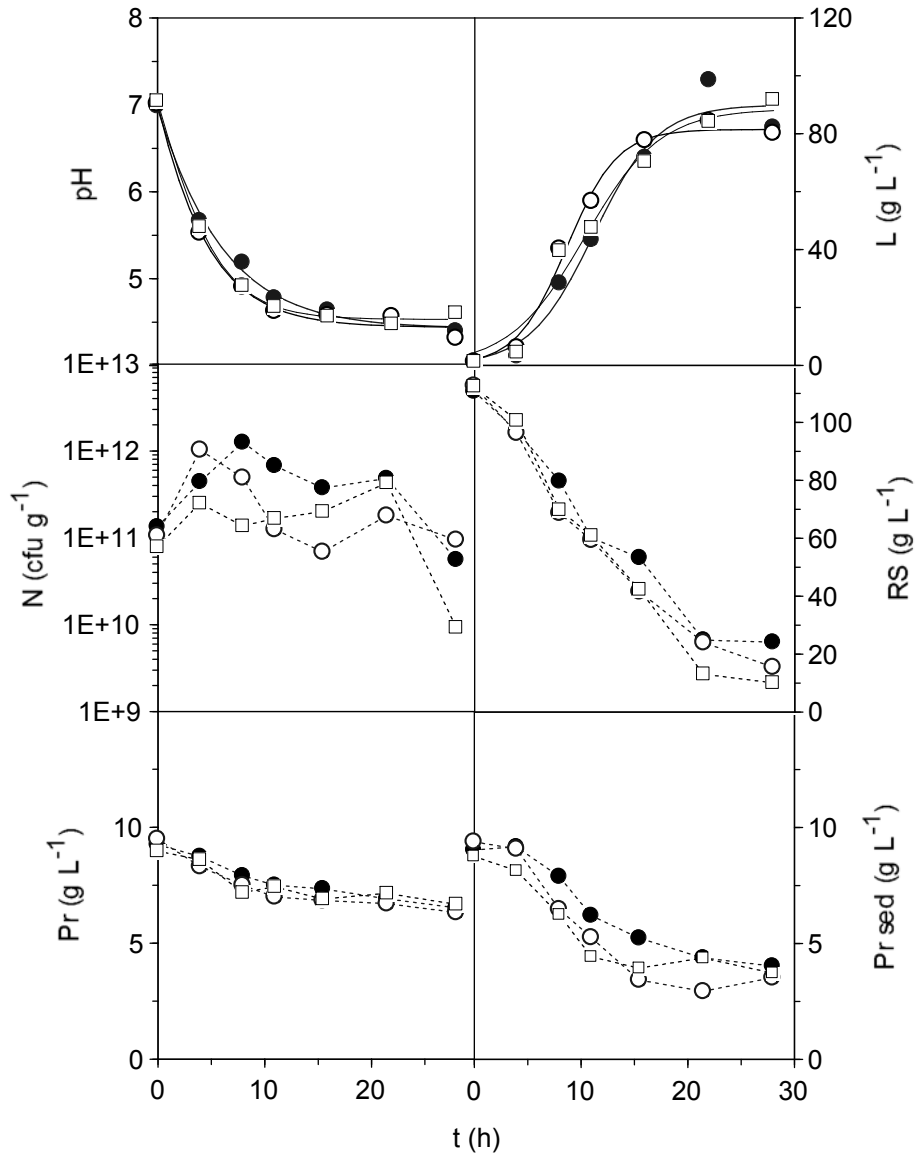


FIGURE 2

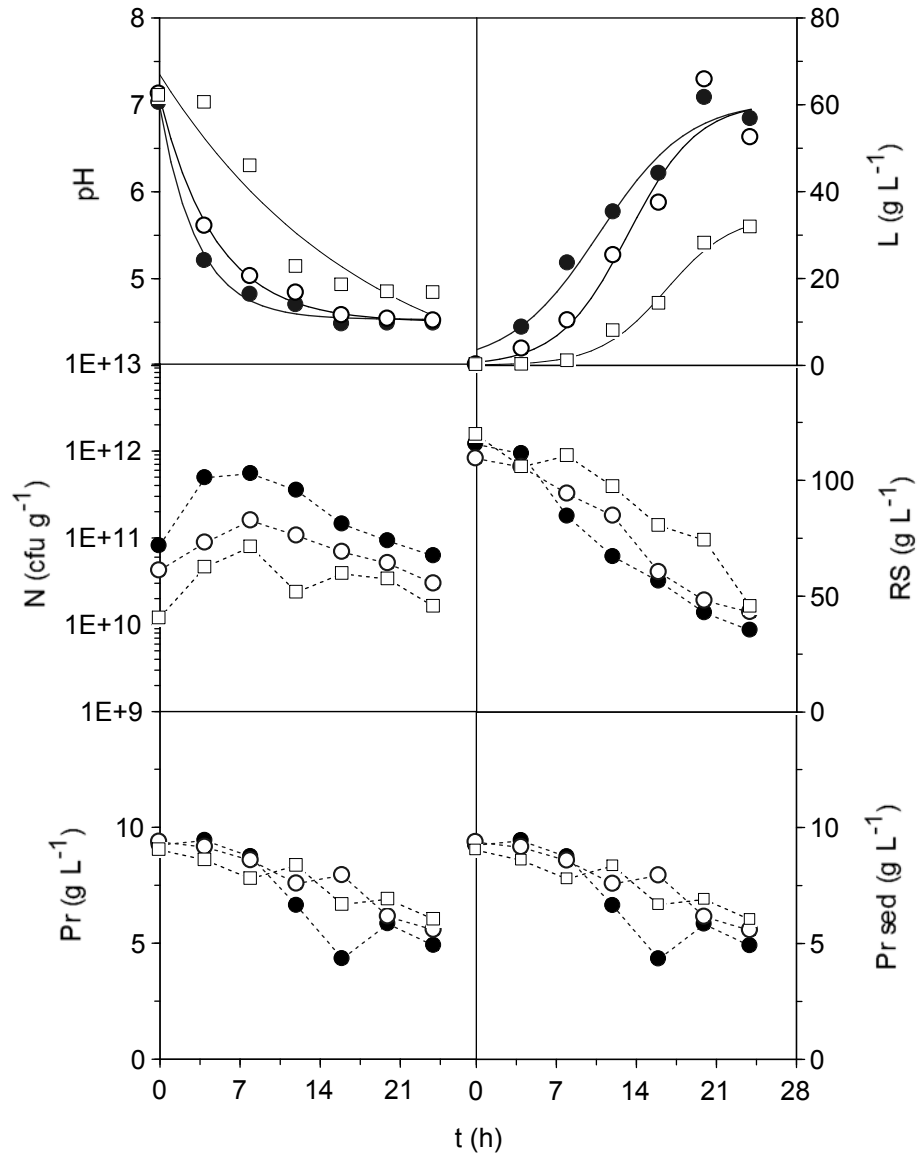


FIGURE 3

