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Hypersaline lakes harbor more active bacterial communities

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Article type: Original research

14 <u>Section</u>: Extreme Microbiology

Running head: Hypersaline lakes exhibit reduced dormancy

16 Keywords: extremophiles, Great Salt Lake, phosphorus, salinity, seed banks

ABSTRACT

- 18 Extremophiles employ a diverse array of resistance strategies to thrive under harsh environmental conditions but maintaining these adaptations comes at an energetic cost. If
- 20 energy reserves to drop too low, extremophiles may enter a dormant state of reduced metabolic activity to survive. Dormancy is frequently offered as a plausible explanation
- 22 for the persistence of bacteria under suboptimal environmental conditions with the prevalence of this mechanism only expected to rise as stressful conditions intensify. We
- estimated dormancy in ten hypersaline and freshwater lakes across the Western UnitedStates. To our surprise, we found that extreme environmental conditions did not induce
- 26 higher levels of bacterial dormancy. Based on our approach using rRNA:rDNA gene ratios to estimate activity, halophilic and halotolerant bacteria were classified as inactive
- 28 at a similar percentage as freshwater bacteria, and the proportion of the community exhibiting dormancy was considerably lower (16%) in hypersaline than freshwater lakes
- 30 across a range of cutoffs estimating activity. Of the multiple chemical characteristics we evaluated, salinity and, to a lesser extent, total phosphorus concentrations influenced
- 32 activity. But instead of dormancy being more common as stressful conditions intensified, the percentage of the community residing in an inactive state decreased with increasing
- 34 salinity in freshwater and hypersaline lakes, suggesting that salinity acts as a strong environmental filter selecting for bacteria that persist and thrive under saltier conditions.
- 36 Within the compositionally distinct and less diverse hypersaline communities, abundant taxa were disproportionately active and localized in families Microbacteriaceae
- 38 (Actinobacteria), Nitriliruptoraceae (Actinobacteria), and Rhodobacteraceae(Alphaproteobacteria). Our results demonstrate that extreme environments may not

40 necessarily be stressful or suboptimal for highly adapted extremophiles causing them to need dormancy less often to survive.

42

INTRODUCTION

- 44 Bacteria in extreme environments survive and often thrive in environmental conditions that are outside the range experienced by the majority of life (Wardle et al., 2004).
- 46 Extremotolerant and extremophilic bacteria, which are found in virtually all harsh environments, have motivated a wide range of research including the metabolic functions
- that have contributed to the evolution of Earth's biosphere (Javaux 2006; Pikuta et al.,2007); novel enzymes for biotechnological applications in chemical, food,
- 50 pharmaceutical industries (van den Burg 2003; Ferrer et al., 2007); and astrobiological clues for discovering life elsewhere in the universe (Rothschild and Mancinelli, 2001). In
- 52 addition, extremophiles provide insight into the physiological adaptations and functional traits that affect microbial performance along environmental gradients (Feller and
- 54 Gerday, 1997; Nealson and Conrad, 1999; Pakchung et al., 2006). For example, extremotolerant and extremophilic bacteria have evolved a diverse array of resistance
- 56 mechanisms, such as the upregulation of organic osmolytes to deal with hypersalinity (Detkova and Boltyanskaya, 2007), heat-shock proteins to combat high temperatures
- 58 (Solow and Somkuti, 2000; Pakchung et al., 2006), and antifreezes to survive in subzero conditions (D'Amico et al., 2006; Struvay and Feller, 2012). However, all of these
- 60 adaptations come at an energetic cost, and if environmental conditions cause energy reserves to drop too low, extremophiles may need to rely on other strategies to ensure
- 62 survival.

One mechanism, dormancy, is frequently offered as a plausible explanation for the

64 persistence of bacterial populations under suboptimal or harsh conditions (Stevenson 1978; Nicholson et al., 2000; Dworkin and Shah, 2010; Lennon and Jones, 2011). As a

- 66 bet-hedging strategy, dormancy builds "seed banks" or reservoirs of inactive individuals that may resuscitated in the future under a different set of conditions (Lennon and Jones,
- 2011). This mechanism not only protects taxa from extinction (Kalisz and McPeek, 1992;
 Honnay et al., 2008), alters species interactions (Chesson and Warner, 1981), and
- 70 influences ecosystem processes (Aanderud et al., 2015), but is prolific, with >90% of biomass and >50% of all bacterial taxa residing in a state of inactivity at any time
- 72 (Alvarez et al., 1998; Lennon and Jones, 2011; Wang et al., 2014). However, the empirical evidence for this mechanism is lacking under the harshest and most adverse
- 74 conditions—extreme environments. In general, the prevalence of dormancy is expected to rise as stressful conditions intensify (Lennon and Jones, 2011). Therefore, bacteria in
- 76 environments at the margins of life should be overly dormant. Even though extremophiles may not just *tolerate* their extreme condition but actually *require* it for
- 78 optimal growth and metabolism (Madigan and Marrs, 1997; Harrison et al., 2013), the activity of extremophiles and extremotolerant bacteria is sensitive to abiotic factors with
- 80 many of these taxa only become metabolically active when a specific set of environmental conditions are met (Pikuta et al., 2007; Zeldovich et al., 2007; Canganella
- 82 and Wiegel, 2011). Thus, if conditions stray outside these boundaries, extremophiles may employ dormancy to survive.
- 84 Hypersaline lakes and their more benign analogs, freshwater lakes, not only offer an ideal setting to identify the extent extremophiles employ dormancy but also the abiotic
- 86 cues structuring bacterial activity. In extreme environments, the primary adverse conditions defining the environment (e.g, salinity, acidity, temperature) vary both
- seasonal or episodically in intensity (Ferris et al., 2003; Detkova and Boltyanskaya, 2007;

Yucel et al., 2013). There is evidence that halophilic organisms are capable of using

- 90 dormancy as a way of contending with hypersalinity and the osmotic stress that it induces. For example, an experimental reduction of hypersaline conditions in lagoon
- 92 water allowed previously undetected protozoa species to emerge from seed banks(Esteban and Finlay, 2003). Also, as ancient hypersaline lakes disappear, haloarchaea
- 94 may survive rising hypersalinity levels in subterranean salt-remains in a dormant state lasting thousands of years (Grant et al., 1998). Further, with the high energetic costs of
- 96 maintaining resistance strategies to combat osmotic stress (Oren 1999), halophiles and halotolerant bacteria may be poorly suited to weather seasonal or episodic changes in
- 98 other abiotic conditions, thus increasing the need for dormancy. In freshwater lakes, the activity of bacteria may track nutreint cues. Dormancy in freshwater systems is common
- 100 with >50% of freshwater biomass and >40% of bacterial taxa potentially residing in a state of reduced metabolic activity (Lennon and Jones, 2011). P availability, in particular,
- influences both bacterial activity (Schindler, 1978; Cole et al., 1993; Jones et al., 1998)and dormancy (Jones and Lennon, 2010). Therefore, under periods of nutrient limitation
- 104 bacteria may decrease their metabolic activity to avoid competition, starvation, and potentially death (Jones and Lennon, 2010).

106 In this study, we tested whether dormancy was more prevalent in extreme hypersaline than freshwater environments and identified the differences in lake chemistry that

- influenced activity. We estimated the dormancy of individual taxa from the recovery of
 16S rRNA transcripts of metabolically active bacteria and 16S rRNA genes of all
- 110 potentially active bacteria (Jones and Lennon, 2010) in five freshwater and five hypersaline lakes across the Western United States. We employed two different

- 112 measurements of dormancy: the percentage of bacterial taxa exhibiting dormancy in each lake and the total relative recovery represented by these dormant taxa within the
- 114 community. We related our dormancy metrics to a suite of chemical characteristics including dissolved O₂, pH, salinity, total nitrogen (TN), and total phosphorus (TP), and

116 temperature.

118 MATERIALS AND METHODS

Lakes and water chemistry

- 120 We sampled water from five hypersaline and five freshwater lakes located in seven states (i.e., AZ, CA, CO, ID, OR, UT, WA) across the Western United States in the early
- summer (17 May 23 June 2012). We selected hypersaline lakes based on salinity (\geq 3.0%) and freshwater lakes that were comparable to at least one of the hypersaline lakes
- in terms of mean depth. The hypersaline lakes included: Great Salt Lake, North Arm(UT); Great Salt Lake, South Arm (UT); Salton Sea (CA); Abert Lake (OR); Mono Lake
- 126 (CA); and the freshwater lakes included: Mormon Lake (ID); Riffe Lake (WA); Arivaca Lake (AZ); Lily Lake (CO); and Silverwood Lake (CA). Supplemental Table 1 provides
- 128 additional information on the elevation, surface area, mean depth, and location of the lakes. Water samples were removed 1.0 m below the lake surface approximately 200 m
- 130 from the shoreline. We measured electrical conductivity, dissolved oxygen (O₂), and temperature *in situ* with an OAKTON EcoTestr EC Low Meter (Oakton Instruments Inc.,
- Vernon Hills, Illinois, USA) and YSI EcoSense DO 200 meter (YSI Inc., Yellow Springs, OH, USA). All other data were collected on lake water samples after being transported on
- 134 ice back to the laboratory. We measured salinity with a conductivity bridge (Beckman,

Brea, CA, USA) and pH with a Thermo Orion Model 410 pH meter (Thermo Scientific,

- 136 Beverly, MA, USA). Total nitrogen (TN) was measured by oxidation and subsequent chemiluminescence using a Shimadzu TOC-V equipped with a TNM-1 unit (Shimadzu,
- 138 Kyoto, Japan). We measured total phosphorus (TP) by persulfate oxidation of organicphosphorus to phosphate followed by colorimetric analysis (Koroleff 1983). We tested
- 140 for differences between hypersaline and freshwater lake chemistry using multiple t-tests and a Benjamini-Hochberg correction to control for the false discovery rate associated
- with multiple comparisons (Benjamini and Hochberg, 1995). Last, in the field, wecollected bacterial biomass for molecular analyses from 2.0 L of water on 142 mm 0.2
- μm filters (Supor[®] PES membrane, Pall Life Sciences, Port Washington, New York,
 USA) using a pressure filtration system (Advantec MFS Inc., Tokyo, Japan). Filters were

146 immediately flash frozen with liquid nitrogen and stored at -80 °C.

148 rDNA and rRNA bacterial communities

We characterized lake bacterial communities using RNA- and DNA-based approaches to

- 150 make inferences about the activity of bacterial taxa. Because ribosomal RNA has a relatively short half-life and is required for protein synthesis (Flardh et al., 1992;
- Bernstein et al., 2002; Steglich et al., 2010), we assumed that bacteria identified fromRNA transcripts were metabolically active, while the bacteria recovered from 16S rDNA
- 154 genes reflect the taxa with varying levels of activity, including organisms that are slow growing and/or dormant (Hugoni et al., 2013, Campbell and Kirchman, 2013). For the
- 156 remainder of the paper, we refer to communities based on the 16S rRNA gene as "rDNA" and 16S rRNA transcripts as "rRNA." Nucleic acids were extracted from filters using a

- 158 PowerSoil DNA Isolation Kit and a RNA PowerSoil Total RNA Isolation Kit (MoBio Corporation, Carlsbad, CA, USA). We reverse transcribed RNA transcripts to cDNA
- using a SuperScript III, one-step RT-PCR kit (Invitrogen Corporation, Carlsbad, CA, USA). We PCR amplified the V3-V4 region of the16S rRNA gene and cDNA using
- bacterial specific primer set 515F and 806R with unique 12-nt error correcting Golaybarcodes (Aanderud and Lennon, 2011). The thermal cycle conditions were: an initial
- 164 denaturation step at 94 °C for 3 min followed by 35 cycles of denaturation at 94 °C for 45 sec, annealing at 50 °C for 30 sec, and an extension at 72 °C for 90 sec. After purifying
- 166 (Agencourt AMPure XP PCR Purification Beckman Coulter Inc., Brea, CA, USA) and pooling PCR amplicons at approximately equimolar concentrations, samples were
- 168 sequenced at the Brigham Young University DNA Sequencing Center (http://dnac.byu.edu/) using a 454 Life Sciences genome sequence FLX (Roche,
- 170 Branford, CT, USA). We analyzed all sequences using mothur (v. 1.29.2), an opensource, expandable software pipeline for microbial community analysis (Schloss et al.,
- 172 2009). After removing barcodes and primers, we screened sequences to remove short reads, chimeras, and non-bacterial sequences. First, we eliminated sequences < 250 bp in
- 174 length and sequences with homopolymers longer than 8 bp. Second, we removed chimeras using UCHIME (Edgar et al., 2011) and eliminated chloroplast, mitochondria,
- archaeal, and eukaryotic 16S rRNA gene sequences based on reference sequences fromthe Ribosomal Database Project (Cole et al., 2009). We then aligned sequences against
- 178 the SILVA database (Pruesse et al., 2007) with the SEED aligner, created operational taxonomic units (OTUs) based on uncorrected pairwise distances at 97% sequence
- 180 similarity, and determined the phylogenetic identity of OTUs with the SILVA database.

To characterize variability in bacterial community composition among lakes, first, we

- used Principal Coordinates Analysis (PCoA) and permutational multivariate analyses of variance (PERMANOVA, Anderson 2001). The PCoA was based on a Bray-Curtis
- 184 distance matrix using the 'vegan' package in R (R Development Core Team 2013). While the PCoA aided in the visualization of communities, we tested for the main effects and
- 186 interactions between lake type (hypersaline versus freshwater) and nucleotide type (rDNA and rRNA) with PERMANOVA using the *adonis* function also in the 'vegan'
- 188 package of R. Second, we quantified the alpha diversity of communities as the inverse Simpson index (Haegeman et al., 2013) after rarefaction by a common sequence number
- (5,846) to remove any bias due to differences in sequencing depth among samples(Nipperess and Matsen 2013). We examined differences in alpha diversity between lake
- (hypersaline vs. freshwater) and nucleotide (rDNA vs. rRNA) type using two-wayANOVA with a Tukey's HSD test. Third, we calculated the relative recovery of eleven
- 194 phyla and four subclasses in rDNA communities to identify differences in the distribution of major taxonomical groups (recovery $\geq 1.0\%$) between hypersaline and freshwater
- 196 lakes. Taxonomic trends were shown with a heat map with hierarchal clustering using the *heatmap* function in the 'gplot' package in R (Oksanen et al., 2013). Last, to evaluate if
- 198 hypersaline and freshwater environments supported similar numbers of bacteria, we estimated abundance as the number of 16S rRNA gene copies in lakes using quantitative
- 200 PCR and the bacterial specific primer set 515F and 806R (Aanderud and Lennon, 2011).We tested for differences between hypersaline and freshwater lakes using a t-test.
- 202

Bacterial dormancy estimates

- We used rRNA: rDNA ratios as a proxy to estimate if a given taxa was dormant or active (Franklin et al., 2013, Jones and Lennon, 2010). Specifically, in each lake, we estimated
- 206 the dormancy of individual OTUs as 1 (rRNA recovery / rDNA recovery). From each of the resulting values, we classified OTUs as either dormant or active based on a cutoff.
- 208 The classification of dormant versus active OTUs is sensitive to the specific cutoff selected (Franklin et al., 2013). Therfore, we estimated dormancy across a range of
- 210 cutoffs from 0.1 0.9. From each of these cutoffs, we estimated bacterial dormancy as the percentage of dormant OTUs occurring in each lake and as the total relative recovery
- 212 represented by these dormant taxa within the community. Specifically, the percentage of dormant OTUs was calculated as the number of dormant OTUs divided by the total
- 214 OTUs present in a given lake × 100, while the relative recovery of dormant OTUs was calculated as the sum of all dormant OTUs in each lake. To determine if dormancy was
- 216 more prevalent in hypersaline environments, we used an indicator variable in multiple regression where lake type (hypersaline vs. freshwater) was treated as a categorical
- 218 predictor variable. Differences in the slopes or intercepts between lake type suggest that hypersalinity differentially affected dormancy responses across the cutoffs (Lennon and

220 Pfaff, 2005).

222 Environmental drivers of bacterial dormancy

We identified the lake chemical characteristics that influenced bacterial dormancy in

- 224 hypersaline and freshwater lakes using multiple regression with lake as a categorical predictor variable (Neter et al., 1996, Lennon and Pfaff, 2005; Lennon et al., 2013). We
- tested whether or not a variable (i.e., dissolved oxygen, pH, salinity, TN and TP) related

to the percentage of dormant bacteria occurring in the five hypersaline and five

- 228 freshwater lakes and the recovery of dormant taxa using forward selection procedure and Akaike's information criterion (AIC; Akaike 1998). For these analyses, we used the
- 230 median cutoff value to classify OTUs as either dormant (≥ 0.5) or active (< 0.5). At this cutoff ratio, the total recovery of an OTU (active and inactive cells) was at least double
- 232 the recovery of RNA transcripts being produced. Therefore, we assumed that in dormant OTUs no more than half of the bacteria were metabolically active and producing RNA
- 234 transcripts. In indicator multiple regression, lake chemistry variables were treated as continuous predictor variable and lake type (hypersaline vs. freshwater) was treated as a
- 236 categorical predictor variable. Differences in the slopes or intercepts between lake type suggest that hypersalinity differentially affected dormancy responses to the chemistry
- 238 variables. The chemical characteristics were checked for collinearity using the *vif* function in the 'car' package in R.
- 240

Rare and abundance bacteria and dormancy

- 242 We classified dormant and active OTUs into abundance categories to gain insight into the contribution of rare and abundant taxa to bacterial dormancy. In our study, rare OTUs
- 244 were defined as OTUs with a relative recovery $\leq 0.1\%$ and all other OTUs were considered abundant with a relative recovery > 0.1% in rDNA communities. Justification
- 246 for this is based on rank abundance curves of bacterial communities from sequencing efforts. In these curves, the bacterial recovery of 0.1% often represents a visible
- 248 demarcation between the few abundant OTUs with relatively high recoveries and the thousands of rare OTUs with relatively low recoveries (Pedros-Alio 2012). Similar to

- 250 indicator variables multiple regression analyses, OTUs were classified as either dormant (≥ 0.5) or active (< 0.5). We tested for the effects of lake type (hypersaline versus
- 252 freshwater) and activity (dormant versus active) on the percentage and recovery of rare and abundant OTUs in communities using two-way ANOVA and Tukey's HSD tests.
- Further, to evaluate whether dormancy was restricted to specific OTUs, we estimated the number and recovery of dormant rare and abundant OTUs in forty-five bacterial families.
- 256 Similar to dormancy in lakes, we estimated dormancy in families as the percentage of dormant OTUs occurring in a given family in each lake and summed the relative recovery
- 258 of dormant OTUs for these taxonomical groups. Differences in dormancy among taxonomical groups and lakes were shown in heat maps with hierarchal clustering using
- 260 the *heatmap* function in the 'gplot' package in R (Oksanen et al., 2013).

262 **RESULTS**

Water chemistry

- 264 Salinity clearly distinguished the extreme conditions in hypersaline lakes from the more benign environmental conditions in freshwater lakes, but other chemical variables also
- 266 differed between lake types (Table 1). On average, salinity was twenty-four-times higher in hypersaline than in freshwater lakes. In addition, electrical conductivity was nineteen-
- times higher in hypersaline lakes, and pH was 8.7 ± 0.47 in hypersaline and 7.0 ± 0.18 in freshwater lakes (mean ± SEM). Conversely, O₂ levels were 23% lower in hypersaline
- 270 than freshwater lakes. Based on concentrations of TN and TP, the trophic status of freshwater and hypersaline lakes varied widely from oligotrophic to hypereutrophic,

- 272 resulting in no differences in total resources between lake types (Vollenweider and Kerekes, 1980; Bachmann et al., 2013).
- 274

Bacterial communities in hypersaline and freshwater lakes

- 276 Hypersaline environments had strong effects on the composition of active and total bacterial communities. This inference was based on the recovery of 570,013 quality
- 278 sequences and 4,212 unique OTUs with samples possessing an average sequencing coverage of $97\% \pm 0.01$. The PCoA results distinctly separated hypersaline from
- 280 freshwater bacterial communities in ordination space along PCoA axis 1, which explained 32.4% of the variation (Figure 1). Hypersaline communities were further
- 282 separated along PCoA axis 2, which explained 18% of the variation. The PERMANOVA results supported the ordination demonstrating a compositional difference between
- hypersaline and freshwater communities (PERMANOVA, lake type, F = 5.33, P = 0.005, df = 1), and also revealed a significant difference between active and total bacterial
- communities (PERMANOVA, nucleotide type, F = 1.9, P = 0.03, df = 1).

Despite having similar bacterial densities as freshwater communities, hypersaline

- 288 communities were less diverse and compositionally similar. Specifically, bacterial diversity was 58% lower in hypersaline than freshwater rDNA communities (two-way
- ANOVA, lake × nucleotide type, F = 15.1, P = 0.001, df = 1, **Supplemental Figure 1**). The distribution of six phyla and three Proteobacteria subclasses distinguished
- 292 hypersaline from freshwater communities; while rDNA and rRNA communities closely grouped together only within hypersaline lake (**Figure 2**). The recovery of
- Alphaproteobacteria and Cyanobacteria was at least 2.5- and 1.7-times higher in

hypersaline than freshwater rDNA and rRNA communities, respectively. Alternatively,

- 296 the recovery of Bacteriodetes was 7.1-times lower in hypersaline rDNA and rRNA communities. Based on qPCR of rDNA, hypersaline $(5.8 \times 10^6 \pm 4.41 \times 10^6 \text{ copies } 16\text{ S})$
- 298 rDNA L⁻¹ water) and freshwater lakes $(1.1 \times 10^7 \pm 1.00 \times 10^7 \text{ copies } 16\text{S rDNA L}^{-1}$ water) bacterial densities were comparable (t-test, t = 0.24, P = 0.64, df = 1).

300

Bacterial dormancy estimates in lakes

- 302 Dormant bacteria were detected in hypersaline and freshwater lakes and dormancy was more prevalent in freshwater than extreme hypersaline environments. Based on indicator
- 304 linear regression results describing the relative recovery of dormant OTUs ($R^2 = 0.82$, $F_{86,8} = 133$, P < 0.001, **Figure 3**), the effect of lake type on dormant OTUs was reflected
- 306 in a difference between the y-intercepts in the equations for each lake type (equation 1 and 2; P < 0.001) where dormancy was 16% lower (percent decrease based on y-
- 308 intercepts) in hypersaline than freshwater lakes across a wide range of cutoffs.

310Freshwater: recovery of dormant OTUs (%) =
$$80.4 - 59.2$$
 (cutoff)(1)Hypersaline: recovery of dormant OTUs (%) = $67.6 - 62.4$ (cutoff)(2)

312

In general, the recovery of bacteria exhibiting dormancy decreased linearly as the

- 314 cutoff estimating dormancy increased, and there were no interactions between the slopes and intercepts, suggesting that the dormancy conclusions were robust across the entire
- 316 range of cutoffs. Alternatively, the effect of lake type on the number of dormant OTUs was similar leading to the overall model:

318

% Dormant OTUs =
$$74.0 - 60.8$$
 (cutoff) (3)

320

 $(R^2 = 0.68, F_{86,8} = 186, P < 0.001,$ Supplemental Figure 2). As the cutoff increased or

became more stringent, the number of OTUs exhibiting dormancy decreased with values ranging from 59.3 $\% \pm 6.92$ to 39.6 $\% \pm 1.22$.

324

Environmental drivers of bacterial dormancy

- 326 Salinity influenced dormancy in both hypersaline and freshwater lakes. The multiple regression model that best predicted the relative recovery of dormant OTUs differed by
- 328 lake type for salinity (equation 4 and 5; P < 0.05) but also included TP to a lesser extent (P < 0.09, $R^2 = 0.96$, $F_{8,1} = 50.0$, P < 0.001; Figure 4). Of all possible models, this one
- 330 generated the lowest AIC score (54) with a Δ AIC of 4.4 units, and resulted in the following equations:

332

Hypersaline: recovery of dormant OTUs (%) = 57.6 - 9.10 (salinity %) + 0.04 (TP) (4)

334 Freshwater: recovery of dormant OTUs (%) = 33.0 - 0.58 (salinity %) + 0.04 (TP) (5)

336 Dormancy decreased with increasing salinity across both lake types, but the effect of salinity on dormancy was more pronounced in freshwater systems where the slope

- describing salinity's impact on dormancy was 16-times higher in freshwater thanhypersaline lakes, substantially contributing to a 43% decline in dormancy (percent
- decrease based on y-intercepts, P < 0.001) in freshwater to hypersaline lakes. We found

no significant model describing the impact of environmental drivers on the number of

dormant OTUs.

344 Abundance and dormancy

A greater percentage of abundant bacteria were classified as active rather than dormant in

- 346 hypersaline lakes. Specifically, the recovery of abundant OTUs (> 0.1% relative recovery) comprised the majority of rDNA in extreme communities ($95\% \pm 1.1$) with
- 348 2.4-times more of this recovery being dormant than active (two-way ANOVA, lake × active versus dormant, F = 78.1, P < 0.0001, df = 1, Figure 5). Further, the recovery of
- 350 abundant and active taxa was 61% higher in extreme than freshwater communities. The relative recovery of rare active and dormant OTUs were similar across lakes (two-way
- 352 ANOVA lake × active versus dormant, F = 0.32, P = 0.58, df = 1), with values ranging from $6.9\% \pm 2.5 - 1.7\% \pm 0.40$.
- 354 Between hypersaline and freshwater lakes, there were robust taxonomical differences in abundant active and dormant taxa. Differences in abundant and active bacteria between
- lake types were localized in five families within two phyla (i.e., Actinobacteria andProteobacteria), which accounted for upwards of 46% of the recovery in any lake (Figure
- 6). For example, the percentage of abundant and active OTUs in the Microbacteriaceae (Actinobacteria, hypersaline = $23\% \pm 7.5$, freshwater = $0.86\% \pm 0.84$), Nitriliruptoraceae
- 360 (Actinobacteria, hypersaline = $13\% \pm 7.6$, freshwater = 0%,), and Rhodobacteraceae (Alphaproteobacteria, hypersaline = $2.0\% \pm 0.68$, freshwater = $0.49\% \pm 0.28$) were at
- 362 least 26-times higher in hypersaline than freshwater lakes, while Burkholderiaceae and Comamonadaceae (Betaproteobacteria) were absent in hypersaline lakes but accounted

18

- 364 for $4.6\% \pm 2.1$ and $12\% \pm 3.5$ of the community in freshwater environments, respectively. Alternatively, taxonomical patterns among abundant and dormant taxa were
- 366 apparent in freshwater systems where the recovery of families: Verrucomicrobiaceae (Verrucomicrobia, hypersaline = $0.20\% \pm 0.11$, freshwater = $13\% \pm 6.0$),
- 368 Flavobacteriaceae (Bacteriodetes, hypersaline = $0.41\% \pm 0.33$, freshwater = $6.1\% \pm 4.8$), and an unclassified Frankineae family (Actinobacteria, hypersaline = $0.50\% \pm 0.40$,
- 370 freshwater = $8.5\% \pm 3.9$) were at least 14-times higher in freshwater than hypersaline lakes.
- 372

DISCUSSION

- 374 Bacterial dormancy is often assumed to be a survival mechanism allowing taxa to contend with harsher environmental conditions (Stevenson 1978; Nicholson et al., 2000;
- 376 Dworkin and Shah, 2010; Lennon and Jones, 2011). However, in this study, we found bacterial dormancy to be less common in extreme hypersaline habitats than seemingly
- 378 more benign freshwater habitats. To our surprise, bacteria in hypersaline lakes were classified as dormant at a similar percentage as freshwater bacteria, and the proportion of
- 380 the community exhibiting dormancy was 16% lower in hypersaline than freshwater lakes across a range of cutoffs describing activity. In both lakes, activity was influenced by
- 382 salinity and to a lesser extent P. Instead of dormancy being more common as stressful conditions intensified, the percentage of the community residing in an inactive state
- 384 decreased with increasing salinity. Taken together, dormancy might not be an advantageous mechanism to weather extreme conditions if bacteria are adapted to harsh
- 386 conditions that are consistently stressful.

388 Hypersaline bacterial communities

Hypersaline bacterial communities supported relatively low levels of bacterial diversity

- 390 consisting of taxa from one dominant phylum and two Proteobacteria subclasses.Hypersaline environments are generated as waters containing high concentrations of salts
- 392 flow into an endorheic lake and are concentrated as evaporation outputs exceed precipitation inputs (Ollivier et al., 1994; Boutaiba et al., 2011). Across these lakes, as
- 394 with all extreme environments from acid seeps and deep-sea thermal vents to glacial ice and acid mine drainage, adverse abiotic conditions select for bacterial communities
- 396 composed of extremotolerant bacteria and extremophiles (Baker and Banfield, 2003;Miroshnichenko and Bonch-Osmolovskaya, 2006; Seufferheld et al., 2008; Anesio and
- 398 Laybourn-Parry, 2012). In hypersaline lakes, the primary adverse condition, hypersalinity and the osmotic stress that it induces, selects a subset of halophilic taxa from regional
- 400 species pool (Wu et al., 2006; Wang et al., 2011; Logares et al., 2013; Tazi et al., 2014). We found evidence supporting this as hypersaline conditions selected for unique
- 402 assemblages of bacteria that were compositionally distinct from freshwater communities. Our lakes contained a range of salinity from 3.0 - 15%, and, thus communities contained
- 404 both "salt loving," halophilic and halotolerant bacteria, which can exist in water up to 15% salinity (Pikuta et al., 2007). Specialized halophiles contributed to hypersaline
- 406 communities that were 50% less diverse than their more benign analog with Cyanobacteria, Alphaproteobacteria, and Gammaproteobacteria contributing upwards of
- 408 70% of the sequences. All three of these taxonomical groups are consistently dominant in

other saline environments as well (Tourova et al., 2007; Jiang et al., 2010; Lefort and

410 Gasol, 2013).

412 Hypersaline lakes contain more active microbes

The prevalence of dormancy did not rise as stressful conditions intensified as expected.

- 414 Based on our approach using rRNA: rDNA ratios to classify the activity of bacterial taxa, we found that the proportion of the community exhibiting dormancy was 16% lower in
- 416 hypersaline than freshwater lakes across a range of cutoffs, and species were classified as dormant at a similar percentage in hypersaline as freshwater lakes. Harsher or more
- 418 extreme saline conditions did not select for higher levels of dormancy. The reverse was actually true. As extreme environments became more hypersaline, a greater proportion of
- 420 the community was active. An explanation for this result may stem from halophiles being highly adapted to hypersalinity for optimal metabolism and growth (Madigan and Marrs,
- 422 1997; Harrison et al., 2013). Extreme environments may not necessarily be stressful or suboptimal for highly adapted extremophiles causing them to need dormancy less often to
- 424 survive. For example, hypersaline environments generally select for a wide range of metabolic diversity, such as oxygenic and anoxygenic phototrophs, obligate and
- 426 facultative aerobic heterotrophs, fermenters, denitrifiers, sulfate reducers, and methanogens (Ollivier et al., 1994; Burke and Knott 1997; Ciulla et al., 1997). However,
- 428 as salinity continues to rise, metabolic diversity dramatically decreases (Oren 2002;Pikuta et al., 2007). Thus, as hypersalinity intensifies, the resulting bacterial communities
- 430 may become more specialized, perform fewer functions but remain predominately active.

432 Salinity and P drive bacterial dormancy

Surprisingly, saltier conditions in both freshwater and hypersaline lakes increased

- 434 activity. A steep rise in salinity spanning less than a single percent (0.01 0.87%) across freshwater lakes corresponded to a 17% decrease in the recovery of dormant taxa
- 436 exhibiting dormancy. Additionally, a five-fold increase in salinity (3.0 15%) among hypersaline lakes corresponded to a 30% decrease in the recovery of dormant taxa. Thus,
- 438 salinity seemed to act as a strong environmental filter selecting for not only active extremophiles but also active freshwater bacteria able to contend with and thrive under
- 440 saltier conditions. The immense effects of salinity on activity may be best explained or mirrored by the importance of salt concentrations to community composition. For
- 442 example, across multiple biomes and in freshwater lakes, microbial community structure and diversity are primarily structured by salinity rather than temperature, pH, or other
- 444 physical and chemical factors (Lozupone and Knight, 2007). As for hypersaline lakes, salinity levels exert immense selective pressure on bacterial species (Canganella and
- 446 Wiegel, 2011). To remain active under saltier conditions, we expect most bacteria to cope with increasing or toxic levels of Na⁺ and potential desiccation stress by using "salting
- out" osmoregulation resistance strategies (Oren 2002; Oren 2008; Canganella and
 Wiegel, 2011). This coping mechanism requires energy to actively export Na⁺ ions and
- 450 synthesize and/or accumulate organic compatible solutes such as polyols, sugars, amino acids and amines (Detkova and Bolyanskaya, 2007; Canganella and Wiegel, 2011). In
- 452 light of these costs, it may be more beneficial/efficient for salt tolerant or salt loving taxa to just remain active and cope with osmotic stresses instead of entering a state of
- 454 dormancy.

P availability appeared to regulate bacterial activity even in extreme environments.

- 456 Regardless of lake type, we found a trend (P < 0.09 marginally significant) where higher TP concentrations related to a lower percentage of the community exhibiting dormancy.
- 458 Our results support previous evidence that low concentrations of TP influence dormancy in freshwater systems (Jones and Lennon, 2010). In more benign systems, like freshwater
- 460 lakes, selection pressures may drive bacteria to develop different patterns of resource use or functional traits allowing them to occupy different niches. Under low resource
- 462 availability, dormancy confers a competitive advantage to bacteria as they avoid starvation and escape death. As nutrient levels decline, bacteria enter dormancy by
- 464 forming cysts or endospores (Sussman and Douthit, 1973; Segev et al., 2012), creating persister cells (Rotem et al., 2010; Maisonneuve et al., 2011), or, simply suspending
- 466 normal metabolic activity (Lennon et al., 2011). Although the role competition is controversial in extreme environments and the impact of resources availability on activity
- 468 is remains unclear, there is a tendency for competition to decline as conditions become more stressful (Fiser et al., 2012) due to extreme conditions limiting the number of niches
- 470 for bacteria to occupy (Oren 2002). In addition to stresses, we propose that competition for essential resources in extreme environments exerts some pressure on competition and,
- 472 ultimately, activity.

474 Abundant taxa are disproportionately active

Communities are unevenly distributed with a few dominant species being numerically

476 abundant and contributing overwhelmingly to the overall community, while rare taxa are thousands in number and contribute little in terms of abundance (Pedros-Alio 2012).

- 478 However, overlaid on top of this seemingly universal shape to bacterial rank abundance curves lies the uncertainty of activity (Shade et al., 2014; Aanderud et al., 2015). We
- 480 found the same ubiquitous rank abundance curve in both hypersaline and freshwater lakes, but hypersaline abundant taxa were disproportionately active. These dominant and
- 482 active taxa were localized in families adapted to a range of salinities. For example,Microbacteriaceae are predominantly aerobic, planktonic, and halotolerant bacteria with
- the potential to persist and thrive at multiple salinity levels (Han et al., 2003; Jiang et al., 2010; Jang et al., 2013), and Rhodobacteraceae of the Alhpaproteobacteria are
- 486 halotolerant, moderately thermophilic chemoorganotrophs and photoheterotrophs common in water, biofilms, and microbial mats that withstand fluctuations in salinity
- 488 (Denner et al., 2006; Farias et al., 2013; Lindermann et al., 2013). Abundant and active Nitriliruptoraceae are haloalkaliphic bacteria that require high salt levels to decompose
- 490 organic C containing nitrile groups (Sorokin et al., 2009). The activity of abundant freshwater taxa may relate to differences in C source availability. For example, the
- 492 Comamonadaceae, which were abundant and active in freshwater lakes, are associated with the decomposition of cyanobacteria biomass, particularly *Microcystis* species, and
- 494 wastewater streams (Li et al., 2012; Krustok et al., 2015). Alternatively,Verrucomicrobiaceae and Flavobacteriaceae, which were abundant but dormant in
- 496 freshwater lakes, are associated with high levels of laminarin and xylan from algal sources. Collectively, Verrucomicrobiaceae and Flavobacteriaceae produce six endo-
- 498 acting polysaccharide hydrolases facilitating the decomposition of polysaccharides and cell wall constituents (Cardman et al., 2014). Unfortunately, we did not measure C

- 500 sources in our water samples. Our results suggest that dormancy is unnecessary for taxa to achieve dominance in extreme conditions.
- 502

Reasonable and robust dormancy estimates

- 504 Our estimates of dormancy represent an approximation of bacterial activity with the cutoffs separating active from dormant taxa conserved over a wide range of values. We
- 506 classified taxa as either active or dormant bacteria based on rRNA: rDNA ratios using the recovery of individual taxa in active (rRNA) and total (rDNA) bacterial communities.
- 508 Inferring activity based on rRNA is a commonly applied approach to characterize growing or active bacteria (Campbell and Kirchman, 2013; Hugoni et al., 2013);
- 510 however, rRNA alone may not be a reliable indicator of the metabolic state of a bacterium (Blazewicz et al., 2013). In the case of dormancy, inactive bacteria may
- 512 contain measureable amounts of rRNA, and in specific cases, specialized cells (e.g., akinetes of some Cyanobacteria), may contain more rRNA in an inactive than active state
- 514 (Sukenik et al., 2012). By estimating bacterial dormancy for each taxa independently using rRNA: rDNA ratios, we compensated for potential taxonomic discrepancies
- 516 associated with rRNA and activity (Blazewicz et al., 2013). Our results are based on a rRNA: rDNA ratio cutoff of \geq 0.5. But even if the cutoff was more strict, where the total
- 518 recovery of an OTU (active and inactive cells) was at least 10-times the recovery of RNA transcripts being produced (cutoff = 0.9), our major findings were the same. We do
- 520 concede that our approach does not perfectly discriminate between dormant and extremely slow-growing bacteria populations (Jones and Lennon, 2010). But slow-
- 522 growers and dormant individuals may respond, grow, and resuscitate similarly following

changes in environmental cues (Kjelleberg et al., 1987; Choi et al., 1999). Thus, we feel our estimate is an appropriate metric quantifying the baseline effects of dominant ecosystem characteristics on activity.

526

524

CONCLUSION

- 528 Halophilic and halotolerant bacteria may employ dormancy to facilitate their long-term persistence and maintain bacterial diversity in extreme environments, but a lower
- 530 proportion of extreme communities utilize dormancy. The overarching adverse condition, hypersalinity, not only structured the less diverse and distinct bacterial communities, but
- 532 also activity levels. However, instead of dormancy being more common as extreme conditions intensified, the percentage of the community residing in an inactive state
- 534 deecreased with increasing hypersalinity. Environmentally adapted halophiles seem to be able to capitalize on their "extreme," yet highly selective set of conditions, allowing them
- to thrive and employ dormancy less often to survive.

538 AUTHOR CONTRIBUTIONS

ZTA, JCV, DPB, and ARH designed the study. ZTA, JCV, and TWM conducted the

- 540 experiments. ZTA, JCV, JTL, TWM, DPB, and ARH analyzed and interpreted the data, ZTA, JCV, JTL, TWM, DPB, and ARH helped write and review the manuscript. ZTA
- 542 agrees to be accountable for all aspects of the work in ensuring that questions related to the accuracy or integrity of any part of the work are appropriately investigated and
- 544 resolved.

546 ACKNOWLEDGMENTS

We would like to thank Dr. Bonnie Baxter for her assistance in registering our Great Salt

- 548 Lake samples with the Great Salt Lake Institute.
- 550 **CONFLICT OF INTEREST:** The authors declare no conflict of interest.

552 FIGURE 1



P1 - Percent variation explained 32.4%

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NOT PEER-REVIEWED

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566 **REFERENCES**

Aanderud, Z.T., Lennon, J.T. (2011). Validation of heavy-water stable isotope probing

- for the characterization of rapidly responding soil bacteria. *Appl. Environ. Microbiol.*77, 4589-4596. doi: 10.1128/AEM.02735-10
- 570 Aanderud, Z.T., Jones, S.E., Fierer N., Lennon, J.T. (2015). Resuscitation of the rare biosphere contributes to pulses of ecosystem activity. *Front. Microbiol.* 6:24. doi:
- 572 10.3389/fmicb.2015.00024

Akaike, H. (1998). "Information theory and an extension of the maximum likelihood

- 574 principle," in *Selected Papers of Hirotugu Akaike*, eds. E. Parzen, K. Tanabe, G. Kitagawa (New York: Springer), 199-213.
- 576 Alvarez, C. R., Alvarez, R., Grigera, S., Lavado, R. S. (1998). Associations between organic matter fractions and the active soil microbial biomass. *Soil. Biol. Biochem.*
- 578 30, 767-773. doi: 10.1016/S0038-0717(97)00168-5
 Anderson, M.J. (2001). A new method for non-parametric multivariate analysis of
- 580 variance. *Austral. Ecol.* 26, 32-46.Anesio, A.M., Laybourn-Parry, J. (2012). Glaciers and ice sheets as a biome. *Trends*
- 582 *Ecol. Evol.* 27, 219-225. doi.org/10.1016/j.tree.2011.09.012
 Bachmann, R.W., Hoyer, M.V., Canfield, D.E. (2013). The extent that natural lakes in
- the United States of America have been changed by cultural eutrophication. *Limnol.Oceanogr.* 58, 945-950.
- Baker, B.J., Banfield, J.F. (2003). Microbial communities in acid mine drainage. *FEMS Microbiol. Ecol.* 44, 139-152. doi: 10.1016/S0168-6496(03)00028-X

588	Benjamini, Y., Hochberg, Y. (1995). Controlling the false discovery rate – a practical and		
	powerful approach to multiple testing. J. R. Stat. Soc. Series B Stat. Methodol. 57,		
590	289-300.		

Bernstein, J.A., Khodursky, A.B., Lin, P.H., Lin-Chao, S., Cohen, S.N. (2002). Global

- analysis of mRNA decay and abundance in Escherichia coli at single-gene resolutionusing two-color fluorescent DNA microarrays. *Proc. Natl. Acad. Sci. U.S.A.* 99,
- *9697-9702.*

Blazewicz, S.J., Barnard, R.L., Daly, R.A., Firestone, M.K. (2013). Evaluating rRNA as

- an indicator of microbial activity in environmental communities: limitations and uses.
 ISME J. 7, 2061-2068. doi:10.1038/ismej.2013.102
- 598 Boutaiba, S., Hacene, H., Bidle, K.A., Maupin-Furlow, J.A. (2011). Microbial diversity of the hypersaline Sidi Ameur and Himalatt Salt Lakes of the Algerian Sahara. *J. Arid*
- 600 *Environ.* 75, 909-916.

Burke, C.M., Knott, B. (1997). Homeostatic interactions between the benthic microbial

- 602 communities and the waters of a hypersaline lake, Lake Hayward, Western Australia. *Mar. Freshw. Res.* 48, 623-631.
- 604 Campbell, B.J., Kirchman, D.L. (2013). Bacterial diversity, community structure and potential growth rates along an estuarine salinity gradient. *ISME J.* 7, 210-220. doi:
- 606 10.1038/ismej.2012.93

Canganella, F., Wiegel, J. (2011). Extremophiles: from abyssal to terrestrial ecosystems

608 and possibly beyond. *Naturwissenschaften*. 98, 253-279. doi: 10.1007/s00114-011-0775-2

- 610 Cardman, Z., Arnosti, C., Durbin, A., Ziervogel, K., Cox, C., Steen, A.D., et al. (2014). Verrucomicrobia are candidates for polysaccharide-degrading bacterioplankton in an
- arctic fjord of Svalbard. *Appl. Environ. Microbiol.* 80, 3749-3756. doi:
 10.1128/AEM.00899-14
- 614 Chesson, P. L., Warner, R. R. (1981). Environmental variability promotes coexistence in lottery competitive-systems. *American Naturalist* 117, 923-943.
- 616 Choi, J.W., Sherr, B.F., Sherr, E.B. (1999) Dead or alive? A large fraction of ETSinactive marine bacterioplankton cells, as assessed by reduction of CTC, can become
- 618 ETS-active with incubation and substrate addition. *Aquat. Microb. Ecol.* 18, 105-115. Ciulla, R.A., Diaz, M.R., Taylor, B.F., Roberts, M.F. (1997). Organic osmolytes in
- 620 aerobic bacteria from Mono Lake, an alkaline, moderately hypersaline environment.*J. Appl. Environ. Microbiol.* 63, 220-226.
- 622 Cole, J.J., Pace, M.L., Caraco, N.F., Steinhart, G.S. (1993). Bacterial biomass and cellsize distributions in lakes - more and larger cells in anoxic waters. *Limnol. Oceanogr.*
- 62438, 1627-1632.

Cole, J.R., Wang, Q., Cardenas, E., Fish, J., Chai, B., Farris, R.J., et al. (2009). The

- Ribosomal Database Project: improved alignments and new tools for rRNA analysis.
 Nucleic Acids Res. 37, D141-D145. doi: 10.1093/nar/gkn879
- 628 D'Amico, S., Collins, T., Marx, J.C., Feller, G., Gerday, C. (2006). Psychrophilic microorganisms: challenges for life. *EMBO Rep.* 7, 385-389.
- Denner, E.B.M., Kolari, M., Hoornstra, D., Tsitko, I., Kampfer, P., Busse, H.J.,
 Salkinoja-Salonen, M. (2006) *Rubellimicrobium thermophilum* gen. nov., sp nov., a

632	red-pigmented, moderately thermophilic bacterium isolated from coloured slime		
	deposits in paper machines. Int. J. Syst. Evol. Microbiol. 56, 1355-1362.		
634	Detkova, E.N., Boltyanskaya, Y.V. (2007). Osmoadaptation of haloalkaliphilic bacteria:		
	Role of osmoregulators and their possible practical application. Mikrobiologiia. 76,		
636	511-522.		
	Dworkin, J., Shah, I.M. (2010). Exit from dormancy in microbial organisms. Nat. Rev.		
638	Microbiol. 8, 890-896. doi:10.1038/nrmicro2453		
	Edgar, R.C., Haas, B.J., Clemente, J.C., Quince, C., Knight, R. (2011). UCHIME		
640	improves sensitivity and speed of chimera detection. Bioinform. 27, 2194-2200. doi:		
	10.1093/bioinformatics/btr381		
642	Esteban, G.F., Finlay, B.J. (2003). Cryptic freshwater ciliates in a hypersaline lagoon.		
	<i>Protist.</i> 154, 411-418.		
644	Farias, M.E., Rascovan, N., Toneatti, D.M., Albarracin, V.H., Flores, M.R., Poire, D.G.,		
	et al. (2013). The Discovery of stromatolites developing at 3570 m above sea level in		
646	a high-altitude volcanic Lake Socompa, Argentinean Andes. PLoS One 8, 15. doi:		
	10.1371/journal.pone.0053497		

- Feller, G., Gerday, C. (1997). Psychrophilic enzymes: molecular basis of cold adaptation.*Cell. Mol. Life Sci.* 53, 830-841.
- 650 Ferrer, M., Golyshina, O., Beloqui, A., Golyshin, P.N. (2007). Mining enzymes from extreme environments. *Curr. Opin. Microbiol.* 10, 207-214.
- 652 doi:10.1016/j.mib.2007.05.004

Ferris, M.J., Magnuson, T.S., Fagg, J.A., Thar, R., Kuhl, M., Sheehan, K.B., et al. (2003).

- 654 Microbially mediated sulphide production in a thermal, acidicalgal mat community in Yellowstone National Park. *Environ. Microbiol.* 5, 954-960.
- 656 Finke, D.L., Snyder, W.E. (2008). Niche partitioning increases resource exploitation by diverse communities. *Science* 321, 1488-1490. doi: 10.1126/science.1160854
- Fiser, C., Blejec, A., Trontelj, P. (2012). Niche-based mechanisms operating withinextreme habitats: a case study of subterranean amphipod communities. *Biol. Lett.* 8,
- 660 578-581. doi: 10.1098/rsbl.2012.0125

Flardh, K., Cohen, P.S., Kjelleberg, S. (1992). Ribosomes exist in large excess over the

- apparent demand for protein synthesis during carbon starvation in marine *Vibrio sp.*Strain CCUG-15956. *J. Bacteriol.* 174, 6780-6788.
- 664 Franklin, R.B., Luria, C., Ozaki, L.S., Bukaveckas, P.A. (2013). Community composition and activity state of estuarine bacterioplankton assessed using differential staining
- and metagenomic analysis of 16S rDNA and rRNA. *Aquat. Microb. Ecol.* 69, 247261. doi:10.3354/ame01635
- 668 Grant, W.D., Gemmell, R.T., McGenity, T.J. (1998). Halobacteria: the evidence for longevity. *Extremophiles*. 2, 279-287.
- 670 Haegeman, B., Hamelin, J., Moriarty, J., Neal, P., Dushoff, J., Weitz, J.S. (2013). Robust estimation of microbial diversity in theory and in practice. *ISME J.* 7, 1092-1101.
- 672 Han, S.K., Nedashkovskaya, O.I., Mikhailov, V.V., Kim, S.B., Bae, K.S. (2003) *Salinibacterium amurskyense* gen. nov., sp nov., a novel genus of the family
- 674 Microbacteriaceae from the marine environment. *Int. J. Syst. Evol. Microbiol.* 53, 2061-2066.

- 676 Harrison, J.P., Gheeraert, N., Tsigelnitskiy, D., Cockell, C.S. (2013). The limits for life under multiple extremes. *Trends Microbiol.* 21, 204-212.
- 678 doi:10.1016/j.tim.2013.01.006

Honnay, O., Bossuyt, B., Jacquemyn, H., Shimono, A., Uchiyama, K. (2008). Can a seed

- bank maintain the genetic variation in the above ground plant population? *Oikos* 117, 1-5. doi: 10.1111/j.2007.0030-1299.16188
- Hugoni, M., Taib, N., Debroas, D., Domaizon, I., Dufournel, I.J., Bronner, G., et al.(2013). Structure of the rare archaeal biosphere and seasonal dynamics of active
- 684 ecotypes in surface coastal waters. *Proc. Natl. Acad. Sci. U.S.A.* 110, 6004-6009. doi:
 10.1073/pnas.1216863110
- 686 Jang, G.I., Cho, Y., Cho, B.C. (2013) Pontimonas salivibrio gen. nov., sp nov., a new member of the family Microbacteriaceae isolated from a seawater reservoir of a solar
- 688 saltern. *Int. J. Syst. Evol. Microbiol.* 63, 2124-2131. doi: 10.1099/ijs.0.043661-0 Javaux, E.J. (2006). Extreme life on Earth - past, present and possibly beyond. *Res.*
- 690 *Microbiol.* 157, 37-48.

Jiang, H.C., Huang, Q.Y., Deng, S.C., Dong, H.L., Yu, B.S. (2010) Planktonic

- 692 actinobacterial diversity along a salinity gradient of a river and five lakes on the Tibetan Plateau. *Extremophiles* 14, 367-376. doi: 10.1007/s00792-010-0316-5
- 694 Jones, B.E., Grant, W.D., Duckworth, A.W., and Owenson, G.G. (1998). Microbial diversity of soda lakes. *Extremophiles* 2, 191-200. doi: 10.1007/s007920050060.
- 696 Jones, S.E., Lennon, J.T. (2010). Dormancy contributes to the maintenance of microbial diversity. *Proc. Natl. Acad. Sci. U.S.A.* 107, 5881-5886. doi:
- 698 10.1073/pnas.0912765107

Kalisz, S., McPeek, M. A. (1992). Demography of an age-structured annual-resampled

- projection matrices, elasticity analyses, and seed bank effects. *Ecology* 73, 1082-1093.
- 702 Kjelleberg, S., Hermansson, M., Marden, P., Jones, G.W. (1987). The transient phase between growth and nongrowth of heterotrophic bacteria, with emphasis on the
- 704 marine environment. Annu. Rev. Microbiol. 41, 25-49.Koroleff, F. (1983). "Determination of phosphorus," in *Methods of seawater analysis:*
- *second, revised, and extended edition*, eds. K. Grasshoff, M. Ehrhardt, K. Kremling(Weinheim, Germany: Verlag Chemie), 125-131.
- 708 Krustok, I., Truu, J., Odlare, M., Truu, M., Ligi, T., Tiirik, K. et al., (2015) Effect of lake water on algal biomass and microbial community structure in municipal wastewater-
- based lab-scale photobioreactors. *Appl. Microbiol. Biot.* 99, 6537-6549. doi:
 10.1007/s00253-015-6580-7
- Lefort, T., Gasol, J.M. (2013). Global-scale distributions of marine surfacebacterioplankton groups along gradients of salinity, temperature, and chlorophyll: a
- 714 meta-analysis of fluorescence in situ hybridization studies. *Aquat. Microb. Ecol.* 70, 111-130. doi: 10.3354/ame01643
- 716 Lennon, J.T., Pfaff, L.E. (2005). Source and supply of terrestrial organic matter affects aquatic microbial metabolism. *Aquat. Microb. Ecol.* 39, 107-119.
- 718 doi:10.3354/ame039107

Lennon, J.T., Jones, S.E. (2011). Microbial seed banks: the ecological and evolutionary

implications of dormancy. Nat. Rev. Microbiol. 9, 119-130. doi:10.1038/nrmicro2504

Lennon, J.T., Hamilton, S.K., Muscarella M.E., Grandy, A.S., Wicking, K., Jones, S.E.

- (2013). A source of terrestrial organic carbon to investigate the browining of aquatic ecosystems. *Plos One* 8, e75771, doi: 10.1371/journal.pone.0075771
- 724 Li, H.B., Xing, P., Wu, Q.L.L. (2012) Characterization of the bacterial community composition in a hypoxic zone induced by Microcystis blooms in Lake Taihu, China.
- *FEMS Microbiol. Ecol.* 79, 773-784.

Lindemann, S.R., Moran, J.J., Stegen, J.C., et al. (2013) The epsomitic phototrophic

- 728 microbial mat of Hot Lake, Washington: community structural responses to seasonal cycling. *Front. Microbiol.* 4, 17.
- Zogares, R., Lindstrom, E.S., Langenheder, S., Logue, J.B., Paterson, H., Laybourn-Parry, J., et al. (2013). Biogeography of bacterial communities exposed to progressive
- long-term environmental change. *ISME J.* 7, 937-948. doi: 10.1016/j.cub.2014.02.050
 Lozupone, C.A., Knight, R. (2007). Global patterns in bacterial diversity. *Proc. Natl.*
- Acad. Sci. U.S.A. 104, 11436-11440. doi: 10.1073/pnas.0611525104.
 Madigan, M.T., Marrs, B.L. (1997). Extremophiles. Sci. Am. 276, 82-87.
- 736 Maisonneuve, E., Shakespeare, L.J., Jorgensen, M.G., and Gerdes, K. (2011). Bacterial persistence by RNA endonucleases. *Proc. Natl. Acad. Sci. U.S.A.* 108, 13206-13211.
- 738 doi: 10.1073/pnas.1100186108.

Miroshnichenko, M.L., Bonch-Osmolovskaya, E.A. (2006). Recent developments in the

- thermophilic microbiology of deep-sea hydrothermal vents. *Extremophiles*. 10, 85-96.Nealson, K.H., Conrad, P.G. (1999). Life: past, present and future. *Philos. Trans. R. Soc.*
- 742 Lond., B, Biol. Sci. 354, 1923-1939.

744

Neter, J., Kutner M., Nachtsheim, C.J., Wasserman, W. (1996). *Applied Linear Statistical Models*. Chicago: Irwin.

Nicholson, W.L., Munakata, N., Horneck, G., Melosh, H.J., Setlow, P. (2000). Resistance

- of *Bacillus* endospores to extreme terrestrial and extraterrestrial environments.*Microbiol. Mol. Biol. Rev.* 64, 548-572.
- 748 Nipperess, D.A., Matsen, F.A. (2013). The mean and variance of phylogenetic diversity under rarefaction. *Methods Ecol. Evol.* 4, 566-572.
- Oksanen, J., Guillaume, F., Kindt, B., Kindt, R., Legendre, P., Minchin, P. R., et al.(2013). Vegan: Community Ecology Package. R package version 2.0-7 R Foundation
- for Statistical Computing, Vienna, AustriaOllivier, B., Caumette, P., Garcia, J.L., Mah, R.A. (1994). Anaerobic bacteria from
- hypersaline environments. *Microbiol. Rev.* 58, 27-38.Oren, A. (1999). Bioenergetic aspects of halophilism. *Microbiol. Mol. Biol. Rev.* 63, 334-

756 348.

Oren, A. (2002). Molecular ecology of extremely halophilic Archaea and Bacteria. Fems

758 *Microbiol. Ecol.* 39, 1-7.

Oren, A. (2008). Microbial life at high salt concentrations: phylogenetic and metabolic

- diversity. *Saline Syst.* 4:2. doi: 10.1186/1746-1448-4-2. doi: 10.1186/1746-1448-4-2
 Pakchung, A.A.H., Simpson, P.J..L, Codd, R. (2006). Life on earth. Extremophiles
- 762 continue to move the goal posts. *Environ. Chem.* 3, 77-93.
 Pedros-Alio, C. (2012). The Rare Bacterial Biosphere. *Ann. Rev. Mar. Sci.* 4, 449-466.
- Pikuta, E.V., Hoover, R.B., Tang, J. (2007). Microbial extremophiles at the limits of life.
 Crit. Rev. Microbiol. 33, 183-209. doi: 10.1146/annurev-marine-120710-100948

TT ... 1 TT

(1 (2007)

/00	Pruesse, E., Quast, C., Knittel, K., Fuchs, B.M., Ludwig, W.G., Pepiles, J., et al. (2007		
	SILVA: a comprehensive online resource for quality checked and aligned ribosomal		
768	RNA sequence data compatible with ARB. Nucleic Acids Res. 35, 7188-7196. doi:		
	10.1093/nar/gkm864		
770	R Development Core Team (2013). R: A language and environment for statistical		
	computing. R Foundation for Statistical Computing, Vienna, Austria, URL		
772	http://www.R-project.org.		
	Rotem, E., Loinger, A., Ronin, I., Levin-Reisman, I., Gabay, C., Shoresh, N., et al.		
774	(2010). Regulation of phenotypic variability by a threshold-based mechanism		
	underlies bacterial persistence. Proc. Natl. Acad. Sci. U.S.A. 107, 12541-12546. doi:		

г 1

T 1 · W/O

р 1° т

776 10.1073/pnas.1004333107.

Rothschild, L.J., Mancinelli, R.L. (2001). Life in extreme environments. Nature 409,

778 1092-1101. doi :10.1038/35059215

Schindler, D.W. (1978). Factors regulating phytoplankton production and standing crop

- in worlds freshwaters. *Limnol. Oceanography* 23, 478-486.
 Schloss, P.D., Westcott, S.L., Ryabin, T., Hall, J.R., Hartmann, M., Hollister, E.B., et al.
- 782 (2009). Introducing mothur: Open-source, platform-independent, communitysupported software for describing and comparing microbial communities. *Appl.*
- *Environ. Microbiol.* 75, 7537-7541. doi: 10.1128/AEM.01541-09
 Segev, E., Smith, Y., and Ben-Yehuda, S. (2012). RNA dynamics in aging bacterial
- 786 spores. *Cell* 148, 139-149. doi: 10.1016/j.cell.2011.11.059

Seufferheld, M.J., Alvarez, H.M., Farias, M.E. (2008). Role of polyphosphates in

- microbial adaptation to extreme environments. *Appl. Environ. Microbiol.* 74, 58675874. doi: 10.1128/AEM.00501-08
- Shade, A., Jones, S.E., Caporaso , G.J., Handeslman, J., Knight, R., Fierer, N., et al.(2014) Conditionally rare taxa disproportionately contribute to temporal changes in
- microbial diversity. *mBio* 5:e01371-14. doi: 10.1128/mBio.01371-14
 Solow, B.T., Somkuti, G.A. (2000). Comparison of low-molecular-weight heat stress
- proteins encoded on plasmids in different strains of Streptococcus thermophilus.*Curr. Microbiol.* 41, 177-181.
- 796 Sorokin, D.Y., van Pelt, S., Tourova, T.P., Evtushenko, L.I. (2009). Nitriliruptor alkaliphilus gen. nov., sp nov., a deep-lineage haloalkaliphilic Actinobacterium from
- soda lakes capable of growth on aliphatic nitriles, and proposal of Nitriliruptoraceae
 fam. nov and Nitriliruptorales ord. nov. *Int. J. Syst. Evol. Microbiol.* 59, 248-253. doi:
- 800 10.1099/ijs.0.002204-0

Steglich, C., Lindell, D., Futschik, M., Rector, T., Steen, R., Chisholm, S.W. (2010).

- 802 Short RNA half-lives in the slow-growing marine cyanobacterium Prochlorococcus. *Genome Biol.* **11**:5. doi: 10.1186/gb-2010-11-5-r54.
- Stevenson, L.H. (1978). Case for bacterial dormancy in aquatic systems. *Microb. Ecol.* 4, 127-133.
- 806 Struvay, C., Feller, G. (2012). Optimization to low temperature activity in psychrophilic enzymes. *Int. J. Mol. Sci.* 13, 11643-11665. doi:10.3390/ijms130911643

44

- 808 Sukenik, A., Kaplan-Levy, R.N., Welch, J.M., Post, A.F. (2012). Massive multiplication of genome and ribosomes in dormant cells (akinetes) of *Aphanizomenon ovalisporum*
- 810 (Cyanobacteria). *ISME J.* 6, 670-679. doi: 10.1038/ismej.2011.128
 Sussman, A.S., and Douthit, H.A. (1973). Dormancy in microbial spores. *Annu. Rev.*
- 812 Plant Physiol Plant Mol Biol 24, 311-352. doi:
 10.1146/annurev.pp.24.060173.001523
- 814 Tazi, L., Breakwell, D.P., Harker, A.R., Crandall, K.A. (2014). Life in extreme environments: microbial diversity in Great Salt Lake, Utah. *Extremophiles*. 18, 525-
- 816 535. doi: 10.1007/s00792-014-0637-x

Tourova, T.P., Spiridonova, E.M., Berg, I.A., Slobodova, N.V., Boulygina, E.S., Sorokin,

- 818 D.Y. (2007). Phylogeny and evolution of the family Ectothiorhodospiraceae based on comparison of 16S rRNA, cbbL and nifH gene sequences. *Int. J. Syst. Evol.*
- 820 *Microbiol.* 57, 2387-2398.

van den Burg, B. (2003). Extremophiles as a source for novel enzymes. Curr. Opin.

822 *Microbiol.* 6, 213-218.

Vollenweider, R.A., Kerekes, J.J. (1980). "Background and Summary Results of the

- 824 OECD Cooperative Program on Eutrophication" in *OECD Report*, (Paris: OECD).Wang, J., Yang, D., Zhang, Y., Shen, J., van der Gast, C., Hahn, M.W., et al. (2011). Do
- 826 patterns of bacterial diversity along salinity gradients differ from those observed for macroorganisms? *PLoS One*. 6:11. doi: 10.1371/journal.pone.0027597
- Wardle, D.A., Bardgett, R.D., Klironomos, J.N., Setala, H., van der Putten, W.H., Wall,D.H. (2004). Ecological linkages between aboveground and belowground biota.
- *Science* 304, 1629-1633.

Wu, Q.L., Zwart, G., Schauer, M., Kamst-van Agterveld, M.P., Hahn, M.W. (2006).

- 832 Bacterioplankton community composition along a salinity gradient of sixteen highmountain lakes located on the Tibetan Plateau, China. J. Appl. Environ. Microbiol.
- 834 72, 5478-5485. doi: 10.1128/AEM.00767-06
 Yucel, M., Sievert, S.M., Vetriani, C., Foustoukos, D.I., Giovannell, D., Le Bris, N.
- 836 (2013). Eco-geochemical dynamics of a shallow-water hydrothermal vent system atMilos Island, Aegean Sea (Eastern Mediterranean). *Chem. Geol.* 356, 11-20.
- 838 doi:10.1016/j.chemgeo.2013.07.020

Zeldovich, K.B., Berezovsky, I.N., Shakhnovich, E.I. (2007). Protein and DNA sequence

determinants of thermophilic adaptation. *PLoS Comput. Biol.* 3, 62-72. doi:
 10.1371/journal.pcbi.0030005

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46

844 **TABLE 1** | Chemistry in freshwater and hypersaline lakes. Data are mean \pm SEM (*n*

= 5) between freshwater and hypersaline lakes with significant differences based on t-

tests and a Benjamini-Hochberg adjustment for multiple comparisons (P < 0.05), which resulted in no false discoveries among significant variables.

	Freshwater	Hypersaline	P value
Dissolved O ₂ (µmol L ⁻¹)	233 ± 15.0	174 ± 9.01	0.01
Electrical conductivity (dS m ⁻¹)	4.5 ± 2.5	85 ± 19	0.01
pH	7.0 ± 0.18	8.7 ± 0.47	0.02
Salinity (%)	0.29 ± 0.17	7.2 ± 2.1	0.03
Temperature (°C)	18.0 ± 1.40	20.5 ± 2.12	0.36
Total N (µmol L ⁻¹)	30.6 ± 8.99	125 ± 57.9	0.18
Total P (µmol L ⁻¹)	6.57 ± 5.40	70.4 ± 33.3	0.14

848

FIGURE CAPTIONS

- 850 FIGURE 1 | Extreme hypersaline lakes influenced the composition of active and total bacterial communities. The multivariate ordination was generated using principle
- 852 coordinate analysis (PCoA) on a sample × OTU matrix of rDNA and rRNA (indicated by dashed lines) community libraries (97% similarity cutoff). Lake abbreviations are as
- follows: hypersaline lakes—Great Salt Lake, North Arm (NGSL); Great Salt Lake, South
 Arm (SGSL); Salton Sea (SS); Abert Lake (LA); Mono Lake (ML); and freshwater
- 856 lakes—Mormon Lake (MRL); Riffe Lake (RL); Arivaca Lake (AL); Lily Lake (LL); and Silverwood Lake (SWL).

858

FIGURE 2 | Heat map showing the distribution of six phyla and three

860 Proteobacteria subclasses that contributed $\ge 1\%$ of the relative recovery to rDNA and rRNA lake communities. Values are based on means (n = 5) with hierarchal

864 FIGURE 3 | Bacterial dormancy decreased linearly as the cutoffs estimating dormancy increased or became more stringent and was more prevalent in

- 866 **freshwater lakes.** Indicator linear regression analysis ($R^2 = 0.82$, $F_{86,8} = 133$, P < 0.001, n = 10) was based on the relative recovery of dormant OTUs across a range of cutoffs
- 868 (0.1 0.9) calculated as 1 (rRNA recovery / rDNA recovery) for each OTU from rDNA and rRNA community libraries. Dormancy was 16% lower in hypersaline than
- 870 freshwater lakes measured as the percent decrease between the significantly different yintercepts (P < 0.001) from the equations for each lake.

872

FIGURE 4 | Bacterial dormancy decreases as lake salinity increases. The indicator

- regression analysis ($R^2 = 0.96$, $F_{8,1} = 50.0$, P < 0.001, n = 10) was based on the relative recovery of dormant OTUs at the cutoff of 0.5 from the equation 1 - (rRNA recovery / Provide Normal Arrows (Recovery (Recovery Arrows (Recovery (Recovery Arrows (Recovery Arrows (Recovery Arrows (Recovery Arrows (Recovery Arrows (Recovery (Recovery Arrows (Recovery Arrows (Recovery (Recovery Arrows (Recovery (Recovery Arrows (Recovery (Recovery
- 876 rDNA recovery). Dormancy was calculated for each OTU from rDNA and rRNA community libraries.
- 878

FIGURE 5 | Abundant bacteria were more likely to be dormant than active in

- 880 hypersaline lakes. OTUs with a relative recovery $\leq 0.1\%$ were considered rare, while OTUs with a relative recovery > 0.1 were considered abundant based on rDNA
- community libraries (97% similarity cutoff). Values are means \pm SEM (n = 5) with different letters indicating significant differences (P < 0.05) based on a two-way ANOVA
- and a Tukey's HSD test.

FIGURE 6 | Heat map showing the distribution of abundant active (A) and dormant (B) lake taxa in 16-19 bacterial families. Values are based on means (n = 5) with

hierarchal clustering of lakes (bottom) and families (left) that contributed $\geq 1\%$ of the relative recovery to any rDNA lake community.

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