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9 TITLE

Abandonment of traditional saltworks facilitates degradation of halophytic plant
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12 SHORT TITLE

- 13 Degradation of halophytic communities
- 14

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36 ABSTRACT

Aims: In Mediterranean countries, traditional salt exploitation has been practiced over 37 centuries. However, there is a progressive reduction of active saltworks, causing changes in 38 the adjacent halophytic communities and, ultimately, the invasion by opportunistic plant 39 species. Assessing the impact of land-use change is key to understand and protect these fragile 40 wetland ecosystems. Here, we explore how the abandonment of saltworks is impacting plant 41 communities. We assess if the reduction in saltworks activity alters the composition of 42 protected halophytic communities and favours the invasion by Carpobrotus edulis, an 43 invasive species in many coastal regions throughout the world. 44

45 **Location**: the Natural Park of Ria Formosa (Algarve, Portugal).

Methods: We studied variations in the structure of halophytic communities affected to different degrees by *Carpobrotus edulis* over three saltworks land-use regimes in the Ria Formosa. Plant cover and soil salinity were estimated in a total of 60 transects pertaining to two saltworks complexes harbouring different land-use and hydrologic regimes. We performed a non-metric multidimensional scaling ordination of saltworks based on plant cover and identified the indicator species of each saltworks class.

Results: We found that plant communities significantly varied among types of saltworks according to a pattern of soil salinity and hydrologic regime. We identified *C. edulis* as the main indicator species of the abandoned saltworks' communities, characterized by less saline soils and being desiccated in summer.

Conclusions: Land-use change caused by the abandonment of *salinas* facilitated the transition
of halophytic into psammophytic communities and the invasiveness of *C. edulis*. The

maintenance of traditional saltworks activities is vital for the preservation of this fragilewetland ecosystem.

60 KEYWORDS

Halophytic plant communities, GIS, invasive species, land-use change, saltworks, soilmoisture, soil salinity, wetlands

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66 1. INTRODUCTION

Habitat change is a key driver of biodiversity loss and degradation of coastal wetland 67 ecosystems (Millennium Ecosystem Assessment, 2005). Half of the world's wetlands have 68 69 been lost over human history due to land conversion, infrastructure development, pollution, water withdrawals, overharvesting, and the introduction of invasive alien species; besides, 70 71 their degradation is faster than that for other ecosystems (Millennium Ecosystem Assessment, 2005; Mitsch & Gosselink, 2007). Wetlands provide important ecosystem services, and 72 human-made wetlands such as traditional saltworks (salinas) have also been designated as 73 requiring protection both for ecological and socio-economic reasons (Ramsar Convention, 74 75 www.ramsar.org/).

Saltworks comprise a series of interconnected ponds in which solar energy produce 76 seawater evaporation and precipitation of salts. Traditional salt exploitation has been practiced 77 in the Mediterranean basin over centuries, mainly in its northern coast (Spain, Greece, Italy, 78 France and Portugal). Abandonment of saltworks began in the twentieth century and reached a 79 peak during its second half (Crisman, 1999). Along the Portuguese coast, salinas are recently 80 threatened by destruction, transformation or abandonment, and the number of active saltworks 81 dropped from 1170 in 1960 to 87 in 2000 (Neves 2005, Rodrigues, Bio, Amat, & Vieira, 82 2011). Land-use transformation may result in changes in ecosystem properties and increase 83 opportunities for non-indigenous species (NIS) (Hobbs & Huenneke, 1992; Vitousek, 84 D'Antonio, Loope, & Westbrooks, 1996; Hobbs, 2000; Byers, 2002; Dethier & Hacker, 85 2005). The establishment and early survival of NIS can also be favoured by the stressful 86 nature of saline systems, which provides windows of opportunity for invasion (Dethier & 87

Hacker, 2005). Thus, the change in land-use caused by the abandonment is susceptible to
promote the invasion of *salinas* by non-indigenous plants.

90 This study analyses the expansion of *Carpobrotus edulis* (L.) N.E.Br. during the last decade in the halophytic plant communities on the edges of active and abandoned *salinas* in 91 Southern Portugal. C. edulis is original from South Africa but it is known to compete 92 aggressively with endangered autochthonous species throughout the world (D'Antonio, 1993; 93 Draper, Rosselló-Graell, Garcia, Gomes, & Sérgio, 2003; Stevens & Lanfranco, 2006; Troìa 94 & Pasta, 2006). Although the effects of the invasiveness of *C. edulis* on the coastal dune plant 95 96 communities of the Iberian Peninsula have been explored (e.g. Ley et al., 2007; Maltez-Mouro et al., 2010; Novoa & Gonzalez, 2014), little is known on the mechanisms behind its 97 invasion of coastal saline systems. Our study focuses on saltworks complexes belonging to 98 the Ramsar Site and Natural Park of Ria Formosa (Algarve, Portugal) where the abandonment 99 and conversion of salinas to other uses are considered a threat to waterbird and plant 100 communities (Plano Setorial da Rede Natura 2000; www.icnf.pt). There is evidence of Ria 101 Formosa's saltworks at least since 1812 (Archivo Popular, 1837). Until the end of the 1960s 102 the extraction activity alternated periods of stagnation with periods of increase. Afterwards, 103 the number of active saltworks in the Algarve dropped from 136 in 1960 to 15 in 2000 104 (Neves, 2005). In Ria Formosa, the salt is harvested during the dry season from the 105 crystallizer ponds. In active salinas, water is pumped and circulated through the different 106 basins to regulate its salinity and depth. Thus, natural brackish habitats -adapted to come into 107 contact with saline water that irrigates twice a day Ria Formosa's marshland along the tidal 108 cycle- are transformed into regulated saline habitats. However, once the water regulation 109 ceases to exist in the abandoned salinas, the basins' water level depends on seasonal 110 variations of the groundwater table depth. In spring, the groundwater table begins to descend 111

in the Ria Formosa, reaching its maximum depth in summer (Costa, Lousã, & Espírito-Santo,
113 1996). Therefore, it is frequent to find abandoned *salinas* desiccated during summer.

According to the definitions provided by McDonald, Gann, Jonson, & Dixon 114 (2016), the salinas could be considered as cultural ecosystems because they are composed of 115 local native species but have a human-imposed structure. Human-induced water regulation in 116 salinas has led to an altered composition of the original flora and fauna communities 117 118 (Walmsley, 1999; Bouzillé, Kernéis, Bonis, & Touzard, 2001). The natural halophytic communities of the Iberian Peninsula are dominated by perennial succulent chenopodiaceous 119 120 shrubs both in salt-marshes and saltworks (Rivas-Martínez et al., 2002), so the salinas are suitable habitats for particular halophytic communities constituting high nature value systems 121 (Costa et al., 1996; de Melo Soares, de Assunção, de Oliveira Fernandes, & Marinho-Soriano, 122 2018). The salinas harbour halophytic protected habitats listed in Annex I of the Habitats 123 Directive: "Mediterranean and thermo-Atlantic salt marshes and salt meadows" (habitats 124 1410, 1420, 1430), and "Salt and gypsum inland steppes" (habitat 1510). These halophytic 125 communities are home of several rare, endangered and endemic species (for details see Costa, 126 Monteiro-Henriques, Neto, Arsénio, & Aguiar, 2007; European Commission, 2007). In 127 particular, Ria Formosa's saltworks harbour and contribute to the density of key wetland-128 dependent species of waterbirds such as the black-tailed godwit (Limosa limosa), the Kentish 129 plover (Charadrius alexandrinus) and the pied avocet (Recurvirostra avosetta) (Rufino, 130 Araujo, Pina, & Miranda, 1984; Catry et al., 2011). However, the communities found in 131 halophytic protected habitats are highly threatened, showing unfavourable conservation status 132 in most member states of the European Union, largely due to NIS and human-induced 133 changes in hydrology (ETC/BD, 2014). 134

135 Here, we seek to investigate whether the abandonment of traditional saltworks may have led to an expansion of *Carpobrotus edulis* and to a degradation of protected halophytic 136 communities in the space of a decade. To do that, we combine field data collection with the 137 use of geographical information systems (GIS) and statistical analyses. The specific 138 objectives of this research are: i) to estimate the spatial spread of Carpobrotus edulis from 139 2004 to 2015 throughout the saltworks of Ria Formosa, ii) to analyse the plant community 140 structure of saltworks according to their degree of abandonment, hydrologic regime and soil 141 salinity, and iii) to assess the relationship between the different types of saltworks and the 142 abundance of C. edulis. 143

144 2. METHODS

145 2.1. Study area and survey of Carpobrotus edulis

The Ria Formosa Natural Park (7° 49' W, 37° 1' N) is included in the Natura 2000 Network and in the Ramsar List. It extends along a coastal lagoon system in Algarve (Southern Portugal) (Fig. 1). The park covers an area of 179000000 m² (17900 ha) with high ecological importance due to its variety of habitats and biodiversity.

To estimate the expansion of Carpobrotus edulis throughout the saltworks of Ria 150 Formosa Natural Park, we compared its cover in 2004 with that in 2015 by means of field 151 surveys and GIS. To estimate C. edulis distribution in 2015, we visited all the saltworks of Ria 152 Formosa a total of five times (once per season) from the spring of 2014 to the spring of 2015. 153 We digitized the patches covered by the NIS using ArcGIS Desktop 10 software from ESRI. 154 For the year 2004, digital information on the distribution of *C. edulis* was retrieved from the 155 "Plano de Ordenamento do Parque Natural da Ria Formosa" (www.icnf.pt) in which plant 156 species information is presented as polygon layers. We calculated the total area occupied by 157

C. edulis at each moment (2004 and 2015) using ArcGIS. During the one-year survey we also detected the type of activity of the *salinas* (active *vs.* abandoned), and assessed their hydrologic regimes across the four seasons to identify the abandoned desiccated *salinas* during summer. We used this information to select the most invaded saline complexes to be used as study sites in the subsequent community analyses.

163 2.2. Sampling strategy and soil salinity and moisture measurement

During the initial survey, we found two salt extraction complexes were by far the most 164 affected by the invasion of *C. edulis*: site A ("Faro Airport": 37° 00' N, 7° 58' W) and site B 165 ("Bias do Sul": 37° 02' N, 7° 45' W) (Fig. 1). We carried out the study of plant communities in 166 active and abandoned saltworks located in these two sites. We classified the saltworks into 167 three types according to their observed land-use and hydrologic regime in 2015: i) saltworks 168 where salt extraction activity is carried out ("active saltworks"), ii) "abandoned saltworks", 169 and iii) abandoned saltworks desiccated during summer ("desiccated abandoned saltworks") 170 (Fig. 1 and Fig. 2). At site A, we found no record of change of activity since 2004, with the 171 exception of a small area of ponds which went from being active in 2004 to abandoned in 172 2015 (see Fig. 1). Some of the abandoned ponds at site A have restarted their activity after our 173 survey. At site B, all saltworks were abandoned in 2004, including those that were active in 174 2015. 175

A total of 60 linear transects were systematically distributed over the two sites and saltworks types (10 transects x 3 saltworks types x 2 study sites). Plant sampling was performed during spring 2016 using the point intercept method at each 25-m transect (51 points spaced every 50 cm; see Nunes et al. (2014) for details). Transects were conducted on the dykes of the salt pans (Fig. 2). At each point, a 5 mm diameter rod was stuck in the ground

making a 90° angle. All plant species touching the rod were recorded and cover estimates for
individual species were calculated as the proportion of points intercepted per transect.
Botanical nomenclature follows the "Checklist da Flora de Portugal" (Sequeira et al., 2011),
and species were determined using "Flora Ibérica" (Talavera & Castroviejo, 2000) and "Nova
Flora de Portugal" (Franco, 1984; Franco & Afonso, 2003).

Soil pore water conductivity (ECp) and moisture were simultaneously measured in 186 situ using a WET-2 Sensor and a HH2 Moisture Meter (Delta-T Devices, Cambridge, 187 188 England). Measurements were made at the starting, middle and ending intersect points of each linear transect (points 1, 26 and 51, respectively), next to the roots, at the maximum depth 189 allowed by the moisture sensors. ECp and soil moisture of each transect were estimated as the 190 average of the three recorded conductivity values. Then, ECp was converted to salinity using 191 the Practical Salinity Scale 1978 (Fofonoff & Millard Jr, 1983) and its extension (Hill, 192 Dauphinee, & Woods, 1986) by means of the "ec2pss" function of the "wq" R package. 193

194 2.3. Analysis of the plant communities

We performed a non-metric multidimensional scaling (NMDS) ordination on the matrix of 195 species cover to explore the relationship among the plant communities of each saltworks type, 196 soil salinity and moisture, and the cover of C. edulis. Estimated plant cover values were 197 square-rooted to reduce the influence of large values. In addition, uncommon species 198 occurring on less than 5% of transects were excluded to avoid an excessive influence of rare 199 taxa. In this way, 38 species (out of 67) were retained for analysis (Table S1, Appendix S1). 200 Data were submitted to a Wisconsin double standardization (species were first divided by the 201 maxima, and then locations standardized for total), and the Bray-Curtis dissimilarity index 202 was used to compute the distance matrix. We used permutation tests (n = 999) to determine 203

204 vector fits and assess the correlation coefficient and the significance to the NMDS axes of soil salinity, soil moisture, and species cover. A smooth surface fitting of soil salinity and moisture 205 within the NMDS was estimated by a generalized additive model. Finally, to test if there was 206 207 a significant difference among the communities found in the three types of saltworks, we used an analysis of similarities (ANOSIM; Clarke, 1993). ANOSIM was performed using the Bray-208 Curtis dissimilarity index and 999 permutation tests. ANOSIM's index (R value) ranges from 209 -1 to 1, a positive value indicating higher dissimilarities between groups than within groups. 210 NMDS and ANOSIM analyses were performed using metaMDS, envfit, ordisurf and anosim 211 functions of the R Package "vegan" (Oksanen et al., 2017). 212

To find indicator species for the communities of each type of saltworks, we used the Dufrene-Legendre analysis (Dufrene & Legendre, 1997) computed with the "labdsv" R package (Roberts, 2016). The indicator value (IndVal) quantifies the fidelity and relative abundance of species in each type of saltworks. The index ranges from 0 to 1, and it shows the maximum value when all the individuals of a single species are observed at all sites belonging to a single cluster.

219 3. RESULTS

220 3.1. Spatial spread of Carpobrotus edulis in the salinas of Ria Formosa

We found that *Carpobrotus edulis* had increased 330 per cent its distribution from 2004 to 2015. In 2004, the NIS had occupied 8 patches in the Ria Formosa: one patch inside the saltworks (site A) with a surface of 23400 m², and 7 patches located in sandy soils outside the *salinas* which were not considered in the study (Table S2, Appendix S1). In 2015, *C. edulis* expanded to invade a total of 100467 m² across the *salinas* of Ria Formosa, where we found 136 patches (Fig. 1 and Table S3, Appendix S1). The expansion of *C. edulis* occurred predominantly throughout the abandoned *salinas* (98131 m²), while the invaded area in the active *salinas* was much smaller (2336 m²). Interestingly, we found 78.8 % of the invaded area in the abandoned *salinas* was located in saltworks desiccated during the summer. The totality of the currently invaded areas in the *salinas* corresponded to halophytic communities in 2004. The most invaded *salinas* were those situated near the Airport of Faro (site A) and close to Bias do Sul (site B) (Fig. 1). Thus, we selected these saltworks as study sites for all subsequent analyses.

Soil salinity and moisture were lower in the "desiccated abandoned saltworks" (mean = 3.9 and 17.48, respectively), than in the "abandoned" (mean = 4.27 and 19.19) and the "active" (mean = 4.86 and 23.83) (Table S4 and Fig. S1, Appendix S1). However, there was no significant difference between the means of the "desiccated abandoned" and the other two groups in any case (paired t-tests: p > 0.05).

239 3.2. Plant communities in each type of saltworks

A three-dimensional NMDS provided a representation of plant cover across the three types of 240 saltworks (stress value = 0.16; Fig. 3). The ordination discriminated among the plant 241 communities found in the 60 transects and showed a gradient along the NMDS1 axis from the 242 "active" to the "desiccated abandoned saltworks". The "desiccated abandoned saltworks" 243 244 showed lower soil salinity and moisture, and more Carpobrotus edulis cover than the other two types of saltworks (Fig. 3, Table S4, and Figs. S1 and S2, Appendix S1). C. edulis cover 245 decreased from the "desiccated abandoned saltworks" to the "abandoned saltworks", being 246 practically inexistent in the "active saltworks". 247

ANOSIM found statistically significant compositional dissimilarities among the three types of saltworks (R = 0.42; p < 0.001), the plant communities were significantly more

dissimilar among the types of saltworks than within each type (Fig. 4). Dissimilarities found between sampling sites (A and B) were lower (R = 0.126; p < 0.002).

There was a high coincidence between NMDS and IndVal analyses. Most of the 252 characteristic species identified by IndVal showed also significant correlations with the 253 NMDS ordination axes (Table 1, Fig. 3C, and Table S5, Appendix S1). Considering only 254 those species with highly significant indicator values (IndVal > 0.5; p < 0.001), we found that: 255 i) Mesembryanthemum nodiflorum was the main species representing plant communities of 256 active saltworks, ii) Arthrocnemum macrostachyum was indicator of abandoned saltworks, 257 and iii) Carpobrotus edulis and Vulpia alopecuros were indicator species of desiccated 258 abandoned saltworks (Table 1). 259

260 4. DISCUSSION

The abandonment of saltworks promoted the expansion of *Carpobrotus edulis* and facilitated 261 the transition from halophytic to psammophytic (i.e. growing in sandy soil) communities in 262 the salinas of Southern Portugal. Land-use change caused by the cessation of traditional salt 263 extraction has led to more than a four-fold increase of the cover of C. edulis throughout the 264 abandoned *salinas* over a decade, while the presence of the NIS is practically inexistent in the 265 active saltworks. The three types of saltworks studied showed a significant difference in plant 266 community composition. Plant composition and C. edulis cover varied along a gradient of 267 soil salinity and moisture, and C. edulis cover was inversely related to soil salinity. Thus, 268 well-developed halophytic communities dominated by Mesembryanthemum nodiflorum and 269 with scarce C. edulis cover were found in the active saltworks. Whereas the exotic C. edulis 270 arose as one of the indicator species in the abandoned saltworks desiccated during summer, 271 which are characterized by less soil salinity and moisture. 272

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The Ria Formosa vegetation has essentially been described as halophytic in salt

274 marshes and *salinas*, and psammophytic in the dunes and pine forests (Costa et al., 1996). Therefore, our results could be reflecting a natural turnover between autochthonous 275 communities. The four most significant indicator species (Mesembryanthemum nodiflorum, 276 Arthrocnemum macrostachyum, Carpobrotus edulis and Vulpia alopecuros) are coastal 277 species which share similar thermal conditions in mainland Portugal (SPBotânica, 2014). 278 Despite these bioclimatological similarities, these species differ in their edaphic requirements 279 related to salinity and moisture. *M. nodiflorum*, an indicator of active saltworks, is typical of 280 saltpans dykes and inhabits temporarily inundated soils (SPBotânica, 2014). This succulent 281 plant has been described as the dominant species of the Spergulario bocconei-282 Mesembryanthemetum nodiflori community growing in the dykes of the salinas of Ria 283 284 Formosa (Costa et al., 1996). The main indicator species of the abandoned saltworks (A. 285 *macrostachyum*) is still associated with halophilic environments, occurring in the highest elevations of salt-marshes and saltpans on exceptionally inundated soils (Costa et al., 1996; 286 Rivas Martinez et al. 2001). Whereas C. edulis and V. alopecuros, the indicators of desiccated 287 288 abandoned saltworks, are associated with non-saline sandy soils (Costa et al., 1996; Talavera & Castroviejo, 2000; SPBotânica, 2014). Thus, colonization by non-halophytic species 289 increases with decreasing salinity. In the abandoned desiccated saltworks, there is a lack of 290 halophytes among the indicator species. In addition, the dominance of good colonizers of 291 sandy littoral disturbed lands (i.e. V. alopecuros and C. edulis) together with other 292 psammophilous species (e.g. Briza maxima and Trifolium angustifolium) may evidence the 293 degree of perturbation that affects these saltworks when abandoned and the subsequent 294 transition from halophytic to psammophytic communities. 295

296 Ria Formosa shows a remarkable variability in both temperature and salinity over 297 the year and along the lagoon (Newton & Mudge, 2003). Despite these variations, we found

298 greater dissimilarities across saltworks types than between sampling sites (A and B). The invasion by C. edulis in the salinas of Ria Formosa was mostly associated with abandoned 299 saltworks desiccated during summer, characterized by less soil salinity than the other 300 301 saltworks types. In fact, variations in the depth of groundwater table and in salinity are known factors affecting the distribution of halophytic communities in the Ria Formosa (Costa et al., 302 1996). Moreover, hydrologic alterations can change the distribution of wetland species (Cronk 303 & Fennessy, 2001). Although *C. edulis* also expands through clonal growth, our results are 304 consistent with the fact that its seed germination is inhibited by salt (Weber & D'Antonio, 305 1999). Besides, this NIS has a preference for well-drained soils (DAISIE European Invasive 306 Alien Species Gateway, 2006), circumstance that especially affects "desiccated abandoned" 307 308 saltworks.

309 The abandonment of salinas seems to alter the abiotic conditions of certain saltworks in Ria Formosa. In particular, the long-established human-controlled hydrologic 310 regime disappears, so the saltworks' water level oscillates according to the groundwater table 311 and precipitation. Abandoned solar saltworks are inclined to degrade by desiccation 312 (González-Alcaraz, Aránega, Tercero, Conesa & Álvarez-Rogel, 2014). We found that soil 313 moisture was lower in the "desiccated abandoned saltworks" despite the fact that the measures 314 were taken during spring, when the salt ponds are not yet desiccated. In addition, we found 315 lower soil salinity in saltworks desiccated during summer than in the other types. A possible 316 317 explanation may be that seawater flow is lower in these ponds, especially in summer when the groundwater table is deeper, so salt content in the dykes mainly results from the previous salt 318 extraction. 319

We identified a shift in plant community composition in response to abandonment. Land-use change constitutes a perturbation for the halophytic communities adapted to the 322 salinas, which are shifting toward psammophytic assemblages where ultimately *Carpobrotus* edulis becomes dominant. Disturbance has been previously found to contribute to the 323 invasibility of communities (Hobbs & Huenneke, 1992; Vitousek et al., 1996; Hobbs, 2000; 324 325 Byers, 2002; Dethier & Hacker, 2005). Anthropogenic disturbance is thought to alter habitats and favour invasions by: i) creating new microhabitats, ii) decreasing native populations, iii) 326 introducing non-indigenous species propagules, and iv) placing native species at a 327 competitive disadvantage with non-native species (see Byers, 2002). Non-indigenous species 328 pose a major threat to wetlands biodiversity, especially in communities undergoing habitat 329 modifications (Cronk & Fennessy, 2001). In saline systems, the lack of biotic resistance 330 seems important in the plant invasion process (Dethier & Hacker, 2005). Although the salinas 331 332 are not strictly natural systems, their traditional activity has collaborated to the preservation of 333 relevant halophytic communities. Thus, the cessation of saltworks activity represents a perturbation for these communities maintained by management. Disturbance on the vegetation 334 of coastal communities can facilitate C. edulis propagation by clonal growth (D'Antonio, 335 1993), and once established, *C. edulis* can be competitively superior because it forms thick 336 mats and it can also interfere with the belowground root distribution of resident species 337 (D'Antonio, 1993; D'Antonio & Mahall, 1991). Moreover, C. edulis seedlings have been 338 found to negatively affect the recruitment of native coastal dune species (Novoa & Gonzalez, 339 2014). Consistent with the literature, our results show halophytic species impoverishment in 340 341 saltworks invaded by C. edulis.

The presence of *C. edulis* has been identified as a major threat for endangered coastal species and its eradication has been suggested (e.g. Stevens & Lanfranco, 2006; Troìa & Pasta, 2006). But *C. edulis* removal is time and money-consuming and it requires long-term control of germination and resprout (Chenot et al., 2017) since the species forms a soil seed

346 bank viable for at least two years (DAISIE European Invasive Alien Species Gateway, 2006). Unfortunately, a recent meta-analysis found that recovery of biological structure and 347 biogeochemical functioning of restored wetland ecosystems were lower than in reference sites 348 (Moreno-Mateos, Power, Comín, & Yockteng, 2012). The existing fauna and flora 349 composition adapted to hypersaline habitats could be irremediably damaged if salt extraction 350 activity continues to disappear. The saltworks abandonment and their reconversion to 351 aqualculture ponds, as well as NIS invasions, are among the main concerns for the 352 conservation of the Ria Formosa Natural Park (Plano Setorial da Rede Natura 2000; 353 www.icnf.pt). Despite its artificial origin, saltworks have a great value as conservation areas 354 and their extended abandonment makes necessary a rethink of how to rehabilitate these 355 356 wetland ecosystems (Crisman 1999; Neves, 2005; Crisman, Takavakoglou, Alexandridis, Antonopoulos, & Zalidis, 2009). The case reported by González-Alcaraz et al. (2014) showed 357 that the irrigation with seawater can prevent changes in the vegetation of abandoned 358 saltworks. In Brazil, abandoned saltworks areas were restored to their original mangrove 359 ecosystems (Dos Reis-Neto & Meireles, 2013). Petanidou & Dalaka (2009) proposed a 360 different approach focused on rehabilitation of abandoned saltworks as tourism places, such 361 as information centers and salt museums. During the time-span between our survey and the 362 publication of the present study, some abandoned saltworks restored their activity. This is a 363 good sign, and further investigation would allow us to test if the preexistent halophytic 364 communities could be recovered. Given the high potential for invasiveness of C. edulis in 365 coastal areas, the maintenance of traditional saltworks activities seems vital for the 366 preservation of this fragile wetland ecosystem. 367

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374

375 AUTHOR CONTRIBUTIONS

R.M.C. conceived the idea, R.M.C. and S.C. designed and conducted fieldwork, S.C.
identified the species, R.M.C. and S.C. analysed the data, R.M.C. produced the figures,
R.M.C. led the writing and S.C. made substantial contributions to the writing. All authors
discussed the results and commented on the manuscript.

380

381 DATA ACCESSIBILITY

382 Primary data is presented as Supporting information.

383

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Appendix S1. Tables and additional figures.

TABLES

Table 1 Indicator species of the plant communities found in each type of saltpans analysed.

531 Only significant indicator values (IndVal; p < 0.05) are shown. Correlations with NMDS axes

532 for those species are also shown.

		IndVal	dVal		DS
Type of saltworks / Species	Code	Value	р	r2	р
Active					
Mesembryanthemum nodiflorum L.	Mno	0.6	0	0.34	0
Suaeda vera Forssk. ex J.F.Gmel	Sver	0.25	0.04	0.41	0
Abandoned					
Arthrocnemum macrostachyum (Moric.) Moris	Ama	0.6	0	0.39	0
Sonchus tenerrimus L.	Ste	0.38	0.04	0.22	0
Polypogon maritimus Willd.	Pma	0.28	0.01	0.02	0.65
Desiccated abandoned					
Carpobrotus edulis (L.) N.E.Br.	Ced	0.63	0	0.47	0
Vulpia alopecuros (Schousb.) Dumort.	Val	0.52	0	0.36	0
Salsola vermiculata L.	Sve	0.36	0.03	0.03	0.37
Juncus maritimus Lam.	Jma	0.33	0.01	0.2	0
Trifolium angustifolium L.	Tan	0.26	0.02	0.07	0.11
Briza maxima L.	Bma	0.24	0.03	0.12	0.02

535 FIGURES

536 **Figure 1**

Location of Ria Formosa Natural Park and saltworks complexes where plant composition and soil salinity and moisture were sampled: Site A ("Faro Airport") and site B ("Bias do Sul"). The distribution of *Carpobrotus edulis* in 2015 and the types of saltworks in each site are shown. At Site A, the most recently abandoned saltworks from 2004 to 2015 are bordered in black.



543 **Figure 2**

549

Types of saltworks used in the study: "active saltworks" (upper panel), "abandoned saltworks" (middle panel), and abandoned saltworks desiccated during summer (lower panel). Vegetation was sampled on the dykes of the saltworks. The two types of abandoned saltworks (lower pictures) show the presence of *Carpobrotus edulis*. The inset in the upper right corner shows how soil salinity and moisture were measured.



550 **Figure 3**

Axes 1 and 2 of the three-dimensional non-metric multidimensional scaling (NMDS) 551 ordination of plant community composition across the three types of saltworks sampled: 552 "active saltworks", "abandoned saltworks" and "desiccated abandoned saltworks". (a and b) 553 Dissimilar plant cover was found across the three types of saltworks. The codes for the 554 species are described in Table S1. (c) Vectors represent correlations between species 555 contributing significantly to the NMDS axes, plotted in red (p < 0.001) and black (p < 0.05). 556 Carpobrotus edulis (Ced) showed a relevant contribution to both axes and its cover was 557 higher in desiccated abandoned saltworks. (d) Smooth surface of soil salinity fitted by means 558 of a generalized additive model. Soil salinity is measured in psu (practical salinity unit). An 559 560 inverse relationship between soil salinity and *C. edulis* was found.





Figure 4

Analysis of similarity (ANOSIM) among the three types of saltworks sampled in the study.
Plant compositional dissimilarities between types were significantly higher than within types.
Notches at medians are drawn in each side of the boxes.

