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Tadano, Yara S and Potgieter-Vermaak, Sanja and Kachba, Yslene R and Chiroli, Daiane MG and Casacio, Luciana and Santos-Silva, Jéssica C and Moreira, Camila AB and Machado, Vivian and Alves, Thiago Antonini and Siqueira, Hugo and Godoi, Ricardo HM (2020) Dynamic model to predict the association between air quality, COVID-19 cases, and level of lockdown. *Environmental Pollution*. p. 115920. ISSN 0269-7491

Downloaded from: <https://e-space.mmu.ac.uk/626735/>

Version: Accepted Version

Publisher: Elsevier

DOI: <https://doi.org/10.1016/j.envpol.2020.115920>

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DYNAMIC MODEL TO PREDICT THE ASSOCIATION BETWEEN AIR QUALITY, COVID-19 CASES, AND LEVEL OF LOCKDOWN

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ABSTRACT

Studies have reported significant reductions in air pollutant levels due to the COVID-19 outbreak worldwide due to global lockdowns. Nevertheless, all of the reports are limited compared to data from the same period over the past few years, providing mainly an overview of past events, with no future predictions. Lockdown level can be directly related to the number of new COVID-19 cases, air pollution, and economic restriction. As lockdown status varies considerably across the globe, there is a window for mega-cities to determine the optimum lockdown flexibility. To that end, firstly, we employed four different Artificial Neural Networks (ANN) to examine the compatibility to the original levels of CO, O₃, NO₂, NO, PM_{2.5}, and PM₁₀, for São Paulo City, the current Pandemic epicenter in South America. After checking compatibility, we simulated four hypothetical scenarios: 10%, 30%, 70%, and 90% lockdown to predict air pollution levels. To our knowledge, ANN have not been applied to air pollution prediction by lockdown level. Using a limited database, the Multilayer Perceptron neural network has proven to be robust (with Mean Absolute Percentage Error ~ 30%), with acceptable predictive power to estimate air pollution changes. We illustrate that air pollutant levels can effectively be controlled and predicted when flexible lockdown measures are implemented. The models will be a useful tool for governments to manage the delicate balance among lockdown, number of COVID-19 cases, and air pollution.

Artificial Neural Networks showed to be robust predictive tools to estimate the best equilibrium among COVID-19 cases, lockdown percentage, and air pollutants level.

24 **Keywords:** Artificial Neural Networks; SARS-CoV-2; Air pollution; lockdown flexibility.

25

26 **1. INTRODUCTION**

27 The World Health Organization (WHO) stated that South America is the new
28 epicenter of the coronavirus pandemic (CNBC, 2020), and Brazil, one of the countries with
29 the highest incidence of new cases and the second highest total number of cases in the
30 world. A study done by scientists from Imperial College, London, showed that Brazil had the
31 highest rate of transmission (R_0 of 2.81) among the 48 countries they investigated (The
32 Lancet, 2020). To date (September 3, 2020), 6.6% of Brazil's total cases (3,997,865) were
33 recorded in São Paulo city (262,570). This number constitutes more than 30% of the cases
34 reported in São Paulo state (826,331). On September the 3rd the number of deaths in São
35 Paulo city was 11,554 (4.4% of confirmed cases of COVID-19 led to death), higher than the
36 global (3.3%) (SEADE, 2020).

37 Due to the rapid person-to-person transmission of COVID-19, São Paulo state
38 government ordered lockdown on March 24, 2020, closing all (Secondary schools,
39 Universities, Shopping Malls and, other commercial entities) but essential services (Nakada
40 and Urban, 2020). As expected, beyond the efficiency to suppress the R_0 (Wilder-Smith and
41 Freedman, 2020), these actions led to the scaling down in traffic, industrial and trade
42 activities, and consequent reduction in air pollution levels, therefore improving air quality as
43 a whole (Dutheil et al., 2020).

44 In response to the exponential increase in infection rates of the virus worldwide, local
45 and national governments relaxed environmental legislation. For instance, the US EPA

46 allowed industries and other facilities autonomy to decide and report if they meet the
47 legislated requirements (Wu et al., 2020). Similarly, the Brazilian government has largely
48 negated enforcement of environmental legislation during the coronavirus outbreak (The
49 Guardian, 2020), which resulted in additional industrial air pollution emission, as well as, an
50 increase in deforestation in the Amazon (de Oliveira et al., 2020). The danger is that reduced
51 enforcement will continue past virus's peak to stimulate the economy and therefore put the
52 population at risk.

53 Various scientists reported decreased air pollutant levels, comparing pre- and post
54 COVID-19 air pollution levels using different methods and scales (Chauhan and Singh 2020;
55 Dantas et al., 2020; Le et al., 2020; Li et al., 2020; Muhammad et al., 2020; Nakada and
56 Urban, 2020; Sharma et al., 2020; Shehzad et al., 2020; Tobías et al., 2020). However, the
57 available air pollution studies related to the COVID-19 situation are based on satellite
58 images, air quality modeling and generally comparing lockdown period data with monthly
59 means over the past few years. Worldwide, most studies reported in the literature indicated
60 reductions in NO_x and $\text{PM}_{2.5}$ levels and an increase in O_3 concentration during lockdown
61 (Nakada and Urban, 2020; Sharma et al., 2020; Sicard et al., 2020; Siciliano et al., 2020;
62 Tobías et al., 2020). The following are a few examples of studies using these approaches.

63 Many researchers worldwide reported a reduction in NO_2 concentration levels
64 (Chauhan and Singh, 2020; Muhammad et al., 2020; Zambrano-Monserrate et al., 2020).
65 Zambrano-Monserrate et al. (2020) reported reductions in China, USA, Italy, and Spain,
66 when Copernicus Atmosphere Monitoring Service data for $\text{PM}_{2.5}$ and NO_2 were compared to
67 the previous three years. Rodríguez-Urrego and Rodríguez-Urrego (2020) studied $\text{PM}_{2.5}$
68 profiles of the 50 most polluted countries and reported an average reduction of 12%
69 worldwide. They used the World Air Quality Index platform to obtain data and compared it to
70 the previous 2 years.

71 Closer to home, Dantas et al. (2020) and Nakada and Urban (2020) compared
72 various air pollutants (including CO, O₃, NO₂, NO, PM_{2.5}, PM₁₀, and SO₂) over different time
73 scales (one year to five-year trend) in Rio de Janeiro and São Paulo, respectively. In both
74 cases, local data were used. Both studies indicated a reduction of all pollutants investigated,
75 except for ozone, which increased.

76 These approaches (using satellite images, air quality modeling and generally
77 comparing lockdown period data with monthly means over the past few years) are limited as
78 it provides mainly an overview of past events, with no future predictions.

79 Artificial Neural Networks (ANN), on the other hand, is a nonlinear methodology
80 capable of mapping a set of inputs into an output, which is important to support decisions
81 regarding preventive measures. This approach has been used in air pollution
82 epidemiological studies (Araujo et al., 2020; Kachba et al., 2020; Kassomenos et al., 2011;
83 Polezer et al., 2018). In Araujo et al. (2020) and Kassomenos et al. (2011), the ANN showed
84 a better performance than linear approaches as Generalized Linear Models. Kassomenos et
85 al. (2011) also concluded that ANN is a more flexible and adaptive mathematical approach.

86 In this context, as lockdown status varies considerably across the globe, there is a
87 window of opportunity for mega-cities to determine the optimum level of lockdown to ensure
88 effective management of transmission rates, air quality, and a healthy economy. To our
89 knowledge, ANN have not been applied to air pollution prediction by lockdown level.

90 To that end, we used four Artificial Neural Networks (ANN) (Extreme Learning
91 Machine – ELM; Echo State Network – ESN; Multilayer perceptron – MLP and Radial Basis
92 Function Networks – RBF) to estimate the influence that newly reported COVID-19 cases
93 and lockdown level may have on the local air pollution (CO, O₃, NO₂, NO, PM_{2.5}, and PM₁₀
94 levels) in São Paulo city. After checking compatibility, we simulated four hypothetical partial
95 lockdown scenarios (10, 30, 70, and 90 %) to investigate the relationship between reduced
96 activities and air quality.

97 In the light of evidence that poor air quality may exacerbate COVID-19 symptoms
98 (Wu et al., 2020), and potentially lead to higher mortality rates, the ANN showed to be a
99 useful predictive tool for governments. Using this approach, resumption of industrial and
100 other activities can be managed to ensure a sustainable balance among economic health,
101 air quality, and transmission rate.

102 **2. MATERIALS AND METHODS**

103 The data of São Paulo city was selected to examine the robustness of our approach.
104 São Paulo is the most populous city of Latin America, with around 12.25 million inhabitants
105 (IBGE, 2020), the main hotspot of COVID-19 in Brazil, and one of the most polluted cities in
106 Latin America. The inputs were: daily number of COVID-19 cases, partial lockdown level,
107 and meteorological variables; the outputs were the daily concentration of each air pollutant
108 (CO [ppm], O₃ [µg/m³], NO₂ [µg/m³], NO [µg/m³], PM_{2.5} µg/m³, and PM₁₀ [µg/m³]).

109 Data on the daily number of newly reported COVID-19 cases and lockdown
110 percentages was collected from March 17, 2020 to May 13, 2020 from the Statistical Portal
111 of São Paulo State (SEADE, 2020). The Intelligence Monitoring System of São Paulo has an
112 agreement with mobile phone companies to track people's movement. This georeferenced
113 anonymised information is available on the SEADE website and has been used in this study.

114 Meteorological variables were extracted from the Environmental Company of São
115 Paulo State database (CETESB). These included: relative humidity – RH [%]; maximum
116 temperature – MT [°C]; atmospheric pressure – AP [hPA]; wind speed – WS [m/s] and global
117 solar radiation–GSR [W/m²] (CETESB, 2020).

118 The data on target pollutant levels of CO [ppm], O₃ [µg/m³], NO₂ [µg/m³], NO [µg/m³],
119 PM_{2.5} µg/m³, and PM₁₀ [µg/m³] concentrations were selected from January 01, 2020 to May
120 13, 2020 (134 samples). As a matter of comparison and to improve the ANN performance,

121 we included the data for a period with zero COVID-19 cases and no lockdown (data from
122 January 01, 2020 to March 16, 2020).

123 Daily concentrations were extracted from the CETESB. More than sixty-six percent of
124 the hourly averages were similar to the daily average. The data were ratified by the
125 CETESB, who follows the quality assurance/quality control (QA/QC) procedure approved by
126 the State Council of Environment (CONSEMA) of the State of São Paulo. Beta radiation is
127 used for PM₁₀ and PM_{2.5} measurements, chemiluminescence for NO₂ and NO, non-
128 dispersive infrared for CO, and ultraviolet analysis for O₃ (CETESB, 2020).

129 Data from four CETESB air quality monitoring stations (AQMS) were used due to
130 their locations (Figure 1). The largest data sets could be obtained from D. Pedro II station
131 (blue spot - located in a high demographic density area) and Tietê station (red spot located
132 near a busy ring road). D. Pedro II station is located downtown – high demographic density
133 area; influenced mainly by a light-duty fleet, and Tietê station is near a ring road,
134 characterized mainly by heavy-duty emissions.

135 Table 1 shows that even at these two stations, some data is lacking. PM_{2.5} data from
136 D. Pedro II station had several gaps in the data set for consecutive days, and these were
137 replaced by data from Mooca station (yellow spot) (CETESB, 2020), as the linear correlation
138 of the data with those from D. Pedro II station is 0.95. For missing data from non-
139 consecutive days, the previous day's values were used. Tietê station had no ozone data,
140 and it was supplemented by data from a nearby location USP-Ipen station (green spot).



141
 142 **Figure 1.** Locations of the air quality monitoring stations in São Paulo. The satellite map is from
 143 Google Maps (Map data©2020 Google; <https://www.google.com/maps/place/Brazil/>); the satellite is
 144 from Google Earth Pro (Map data©2020 Google; [https://www.google.com/maps/@-23.6815315,-](https://www.google.com/maps/@-23.6815315,-46.8754814,10z)
 145 [46.8754814,10z](https://www.google.com/maps/@-23.6815315,-46.8754814,10z)). The maps were edited with Microsoft Power Point (version 16.28-19081202).

AQMS	CO	O ₃	NO ₂	NO	PM ₁₀	PM _{2.5}
Tietê*	2	0*	0	0	1	2
D. Pedro II	4	1	0	0	0	10

146 Note: AQMS: Air Quality Monitoring Station; Tietê: ring road; D. Pedro II:
 147 downtown; *Tietê station has no O₃ data and was replaced by data from USP-
 148 Ipen station.

149 **Table 1.** Number of days with no data for each studied AQMS.

150 2.1. Artificial Neural Networks

151 The four ANN used in this study are described below (further details in Araujo et al.
 152 (2020)).

153 2.1.1. Multilayer Perceptron Overview

154 The Multilayer Perceptron (MLP) is a neural model able to map any nonlinear,
 155 continuous, limited, and differentiable function with arbitrary precision, which confers a
 156 characteristic of a universal approximator (Haykin, 2008). The basic structure of an ANN is
 157 the artificial neurons, functional units responsible for processing the information, and
 158 providing the output response (de Castro, 2007).

159 In an MLP, the neurons are distributed in three kinds of layers. The input layer
160 transmits the data to the intermediate (hidden) layers, where the neurons perform a
161 nonlinear transformation, mapping the input signal to another space. Then, the signal is sent
162 to the output layer, in which the output signal is generated based on a linear combination, in
163 most cases. Neurons from the same layer are disconnected, while those from disjoint layers
164 fully exchange information since this is a feed forward model (Siqueira and Luna, 2019).

165 Training a neural model means using an algorithm to determine its free parameters
166 or adjust the neurons' weights. The most known way to solve this task in an MLP is to use
167 the backpropagation algorithm, a general iterative tool based on the steepest descent, a first
168 order unrestricted linear optimization method. In this case, the method reduces the mean
169 square error between the desired response and the output of the network (Haykin, 2008).
170 However, in this work, we address a second-order method that presents computational cost
171 similar to the first: The Modified Scaled Conjugate Gradient (MSCG) (dos Santos and Von
172 Zuben, 1999).

173 We highlight the maximum number of iterations as the stop criterion in training. We
174 also use the hold-out cross-validation method to determine the topology (number of neurons
175 in the hidden layer) and avoid overfitting (Haykin, 2008).

176 2.1.2. Radial Basis Function

177 The Radial Basis Function networks (RBF) are a well-known ANN model. Like the
178 MLP, they are feed forward architectures, and universal approximators, but present only two
179 layers of neurons (Siqueira and Luna, 2019). The first, intermediate, perform a nonlinear
180 input-output mapping using radial basis functions, like the Gaussian function. The second –
181 output layer –performs the model's response, similarly to the MLP (Haykin, 2008).

182 The hidden neurons present two parameters: a centre \mathbf{c}_i (with the same dimension of
183 the number of inputs), and a dispersion σ_i ; Therefore, the output of each neuron is higher to

184 inputs that are spatially closer to the current centre. The dispersion is responsible for
185 modulating the decay of the response concerning the distance between the inputs and the
186 centres. Usually, the Gaussian function is addressed as RBF. A linear combinator is used to
187 perform the output response (Siqueira and Luna, 2019).

188 The training process of an RBF is performed in two steps. The first is the adjustment
189 of the hidden neurons (centers and dispersions), a task performed by the unsupervised
190 clustering method. In this work, we addressed the K-Medoids algorithm. Also, we assumed
191 that all dispersions are the same (Haykin, 2008). The second step is the adjustment of the
192 output neurons. A simple and efficient tool found in the literature is the use of the Moore–
193 Penrose inverse operator (Haykin, 2008).

194 2.1.3. Extreme Learning Machines

195 Extreme Learning Machines (ELM) are feed forward neural models, with a single
196 hidden layer (Huang et al., 2006, 2015). This structure is quite similar to the classic MLP, the
197 training process being the main difference (Siqueira et al., 2018).

198 In an ELM, the intermediate neurons have weights randomly generated, and they are
199 not adjusted during the running time. The insertion of new neurons in the hidden layer leads
200 to a decrease in the output error (Siqueira et al., 2012a).

201 Then, an ELM training is summarized in finding the best set of weights of the output
202 layer. The main manner to overcome this task is to use a minimum square solution,
203 especially the Moore–Penrose generalized inverse operation (Siqueira et al., 2018).

204 2.1.4. Echo State Networks

205 The Echo State Networks (ESN) are architectures of ANN, which present high
206 similarity with the ELM, regarding the structure and training process. However, unlike the
207 previously mentioned networks, this is a recurrent model since it presents feedback loops of

208 information. In this case, the hidden layer, named dynamic reservoir, has such recurrence
209 (Jaeger, 2002, 2001).

210 Jaeger (2002, 2001) demonstrated that the reservoir is a nonlinear transformation,
211 which is influenced by the recent samples of the input signal, so that we can choose the
212 weights in advance if specific conditions are respected. In this work, we used the reservoir
213 design by (Jaeger, 2001).

214 As in the ELM, the training is responsible for determining the weights of the output
215 layer, which may be done using the Moore–Penrose generalized inverse operation, as in the
216 ELM case (Siqueira et al., 2018).

217 **2.2. Computational Details**

218 The computational step involved the seven input variables mentioned above: number
219 of COVID-19 new cases, partial lockdown level, maximum temperature, relative humidity,
220 atmospheric pressure, wind speed, and global solar radiation. The desired signals (target)
221 were each air pollutant's (CO, O₃, NO₂, NO, PM_{2.5}, and PM₁₀) concentration.

222 We evaluated the performance considering all the inputs at the same time; without
223 the number of new COVID-19 cases; and without the number of new COVID-19 cases and
224 partial lockdown, to analyze the robustness of the neural networks on predicting air quality
225 according to COVID-19 variables and using a small database. All cases included the
226 meteorological variables.

227 To perform the computational analysis, we separated the dataset in three subsets:

- 228 ▪ Training: from January 01 to April 23, 2020 (114 samples);
- 229 ▪ Validation: April 24 to May 03, 2020 (10 samples);
- 230 ▪ Test: May 04 to May 13, 2020 (10 samples).

231 The training subset is used to adjust the models, and the validation is applied to
232 verify the overtraining and define the number of neurons in the intermediate layer. Finally,
233 the test subset is used to evaluate the performance of the models. We also verified if the use
234 of the Z-score may bring some performance gain. It is a mathematical treatment that
235 transforms the series of data into approximately stationary. Some studies have presented
236 the importance of using such an approach (Kachba et al., 2020; Siqueira et al., 2018).

237 To apply the Z-score, the value of each sample is subtracted from the mean and
238 divided by the standard deviation. At the end of the ANN execution, the process is reversed
239 to analyze the performances in the original domain.

240 The number of neurons in the hidden layer was defined by empirical tests, varying
241 from 3 to 100 neurons. The best number for each case was chosen based on the lower
242 Mean Square Error (MSE) in the test set. The number of neurons in the hidden layer of each
243 neural model is in Tables A1 and A2 in Appendix A.

244 We followed the premises from the literature of adopting the MSE as the most
245 important error metric because this is reduced during the training (adjustment) of the neural
246 models (Araujo et al., 2020; Kachba et al., 2020; Siqueira et al., 2020, 2018, 2014).

247 The artificial neurons in the intermediate layer of the MLP, ELM, and ESN, use the
248 hyperbolic tangent as an activation function. In the RBF, the Gaussian function is used. The
249 MLP training addressing the Modified Scaled Conjugate Gradient (MSCG) and uses as stop
250 criterion the maximum number of 500 iterations. Also, the K-Medoids in RBF achieved the
251 stop criterion after 10 iterations without modification in the position of the centroids
252 (Figueiredo et al., 2019).

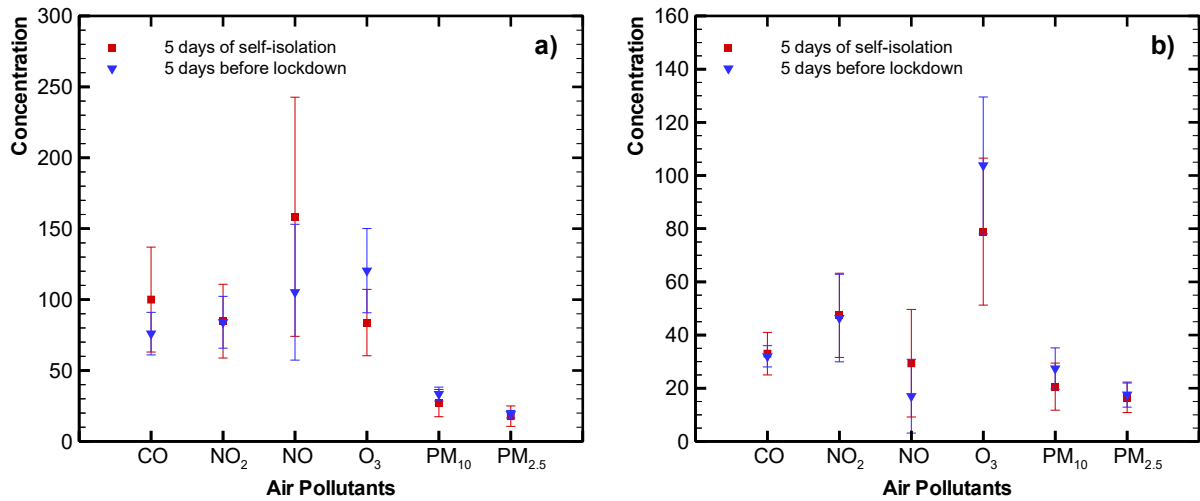
253 **3. RESULTS AND DISCUSSION**

254 For simplicity, we divided this section into three parts. Firstly, the descriptive analysis
255 of the databases, followed by the ANN prediction results, and lastly, the results for the
256 hypothetical scenarios of 10%, 30%, 70%, and 90% of lockdown.

257 **3.1 Descriptive Analysis**

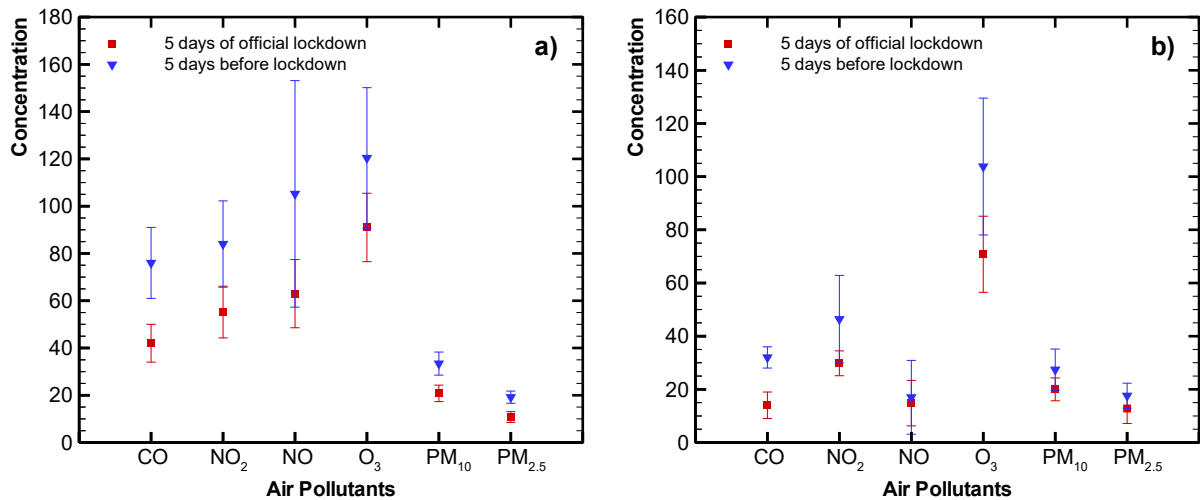
258 The daily concentrations during the studied period, together with the partial lockdown
259 level, are shown in Appendix A - Figure A1. The São Paulo state government officially
260 ordered lockdown on March 24, 2020, however, the population started to self-isolate
261 voluntarily the week before (first available social isolation data – March 17, 2020). From
262 March 17, 2020 to May 13, 2020, the lockdown varied between 38 and 59%, with an
263 average of 51%.

264 To visualize changes in air pollution levels due to voluntary self-isolation and / or
265 lockdown, we compared the five-day average before (12 – 16 March, 2020) voluntary self-
266 isolation with a five-day average during self-isolation (17 – 21 March, 2020) (Figure 2). There
267 is no distinctive change in pollutant levels within experimental error, as may be expected due
268 to a lag in response and a low level of reduced activities. However, comparing a five-day
269 average during the first lockdown period (54 – 56% lockdown from 24 to 28 of March, 2020)
270 with the period before lockdown or self-isolation, we do observe a general decrease in
271 pollutant levels for all pollutants at Tietê and for most at D. Pedro II as is shown in Figure 3.
272 As this period would reflect the changes in the self-isolation period's activities with additional
273 reduction of activities, this finding is not surprising.



274

275 **Figure 2.** Five-day average pollutant levels before and during the voluntary self-isolation period at
 276 Tietê station (a) and D. Pedro II station (b) (CO concentration were multiplied by 100).



277

278 **Figure 3.** Averages comparison between five days of official lockdown with five days before lockdown
 279 for Tietê station (a) and D. Pedro II station (b) (CO concentration were multiplied by 100).

280 From Figure A1 we observe that this trend continues until around the 24th of April,
 281 after which relaxation in lockdown rules corresponds to a steady increase in most of the
 282 pollutant levels. It does seem as though not all the pollutants are similarly influenced by the
 283 lockdown. The particulate matter concentration appears to be influenced by other factors as
 284 well, and reaches much higher values PM_{2.5} towards the end of the lockdown period discussed
 285 here than what it was before. The ozone levels generally increased with a lockdown
 286 percentage increase.

287 Using a neural network to study atmospheric ozone formation in the Metropolitan
 288 Area of São Paulo (MASP), Guardani et al. (1999) found that temperature was the main
 289 factor affecting ozone formation and observed higher ozone levels in regions characterized
 290 by lower emission levels of ozone precursors. Martins and Andrade (2008) evaluated VOC s'
 291 potential for ozone formation using a three-dimensional air quality model and found that
 292 ozone in the MASP is VOC-limited, as commonly observed in urban areas (Li et al., 2019;
 293 Siciliano et al., 2020; Tobías et al., 2020). Under these conditions, a decrease of NO_x can
 294 reduce the removal of O₃ through NO_x titration and/or the effect of radical terminating
 295 reactions, and thereby increasing O₃ formation (Seinfeld and Pandis, 2016; Sillman, 2003,
 296 1999). Furthermore, Andrade et al. (2017), studying the MASP, explain that decreasing NO_x
 297 and CO emissions simultaneously contribute to higher ozone levels. This behavior is also
 298 affirmed in (Gentner et al., 2009; Harley et al., 2005; Marr and Harley, 2002; Stedman,
 299 2004).

300 Table 2 presents the linear correlations between the lockdown level (varying from 38
 301 - 59%) and air pollutant concentrations at Tietê and D. Pedro II stations for March 17, 2020
 302 (first day of available data of social isolation) to May 13, 2020. Bar ozone, all the pollutants
 303 correlated negatively (ranging from -0.14 for CO at D. Pedro II to -0.60 for NO at Tietê) with
 304 the lockdown.

	CO	O ₃ *	NO ₂	NO	PM ₁₀	PM _{2.5}
Tietê	-0.45	0.15	-0.57	-0.60	-0.34	-0.38
D. Pedro II	-0.14	0.11	-0.42	-0.33	-0.23	-0.26

305 Note: *Data from USP-Ipen station

306 **Table 2.** Linear correlations between lockdown and studied air pollutant concentrations for March 17,
 307 2020 to May 13, 2020.

308 Finally, Appendix A - Figure A2 shows the number of daily COVID-19 newly reported
 309 cases. The first day of registered COVID-19 cases was February 25, 2020 and an
 310 exponential increase is observed from the beginning of April onwards.

311 **3.2 ANN Estimation Analysis**

312 Tables 3 and 4 contain the average and standard deviation for each pollutant level
313 obtained from the 3 subsets (training, validation, and test) at the two sites. Although the two
314 monitoring sites are in the same city, the descriptive statistics show significant differences.
315 Tietê station (near highways) has higher average concentrations for all pollutants in
316 comparison to D. Pedro II station (populated city area). The different statistical profiles of the
317 two sites are indicative of robust evaluation of the data, as the model could provide a MAPE
318 of ~ 30%, despite two dissimilar data sets.

319 Tables A1 and A2 (Appendix A) display the ANN computational results for AQMS
320 Tietê (ring road station) and AQMS D. Pedro II (densely populated city area station),
321 respectively. For this purpose, the best (lower Mean Square Error - MSE) of 30 independent
322 executions were considered (de Castro, 2007; Haykin, 2008; Siqueira et al., 2018). The
323 shaded values indicate results with the best performance (lower MSE). The MLP neural
324 model achieved the best results (i.e., lowest MSE) in almost all cases, except for O₃ at D.
325 Pedro II station. The latter was best estimated using the ELM neural model. It is an important
326 observation, as there is no consensus about which ANN is the best. It corroborates with the
327 results achieved by Polezer et al. (2018) and Araujo et al. (2020), both applied to air
328 pollution epidemiological studies.

	Training	Validation	Test
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Pollutant	Average	Standard Deviation	Average	Standard Deviation	Average	Standard Deviation
CO [ppm]	0.69	0.29	0.93	0.46	0.85	0.34
O ₃ [µg/m ³]	70	28	98	21	74	14
NO ₂ [µg/m ³]	68	24	86	31	88	31
NO [µg/m ³]	91	69	124	89	151	84
PM _{2.5} [µg/m ³]	13	5.5	24	12	20	9.7
PM ₁₀ [µg/m ³]	22	8.2	43	19	38	19

329 **Table 3.** Average and standard deviation for each studied pollutant for the 3 subsets (Tietê Station).

Pollutant	Training		Validation		Test	
	Average	Standard Deviation	Average	Standard Deviation	Average	Standard Deviation
CO [ppm]	0.30	0.15	0.62	0.40	0.50	0.36
O ₃ [µg/m ³]	65	24	81	19	59	14
NO ₂ [µg/m ³]	43	17	60	33	64	31
NO [µg/m ³]	21	19	51	63	75	76
PM _{2.5} [µg/m ³]	12	4.7	20	8.8	16	7.8
PM ₁₀ [µg/m ³]	19	7.3	37	14	31	16

330 **Table 4.** Average and standard deviation for each studied pollutant for the 3 subsets (D. Pedro II
331 Station).

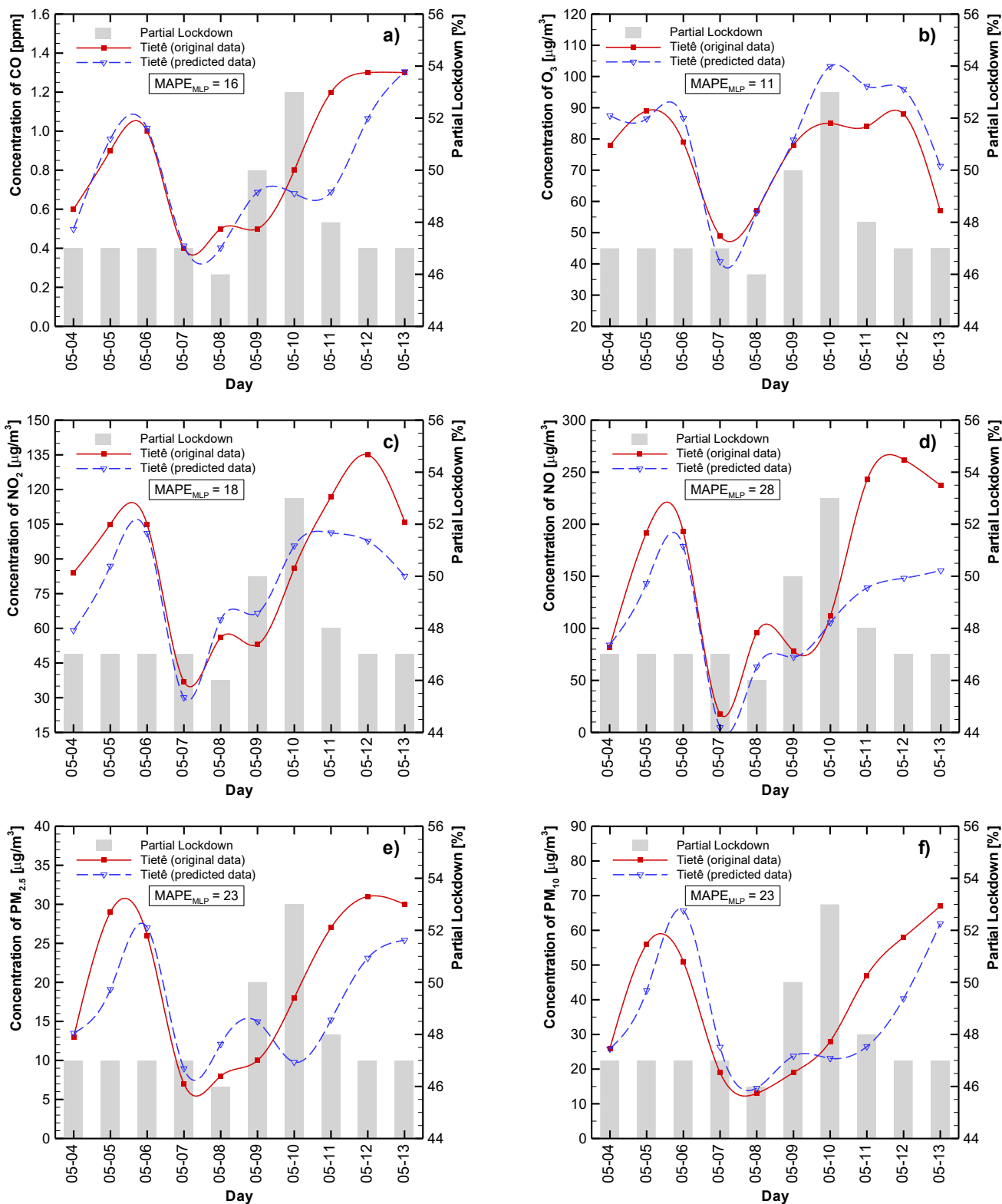
332 It is important to highlight that the best overall ANN results were achieved when the
333 variables “number of new COVID-19 cases” and “partial lockdown” were included (8 out of
334 12 cases). The remaining 4 cases (NO₂ and PM_{2.5} at Tietê, and NO₂ and O₃ at D. Pedro II)
335 showed the best result considering only “partial lockdown”. In both scenarios the
336 meteorological variables were included.

337 To establish if the Z-Score application could result in performance gain, the ANN was
338 also performed with the Z-score (Results shown in Tables A1 and A2). The Z-score's use
339 proved to be beneficial in 2 cases at the Marginal Tietê station, and four cases at the D.
340 Pedro II site. Therefore, it can be considered in addition to increasing the quality of the
341 results of the ANN.

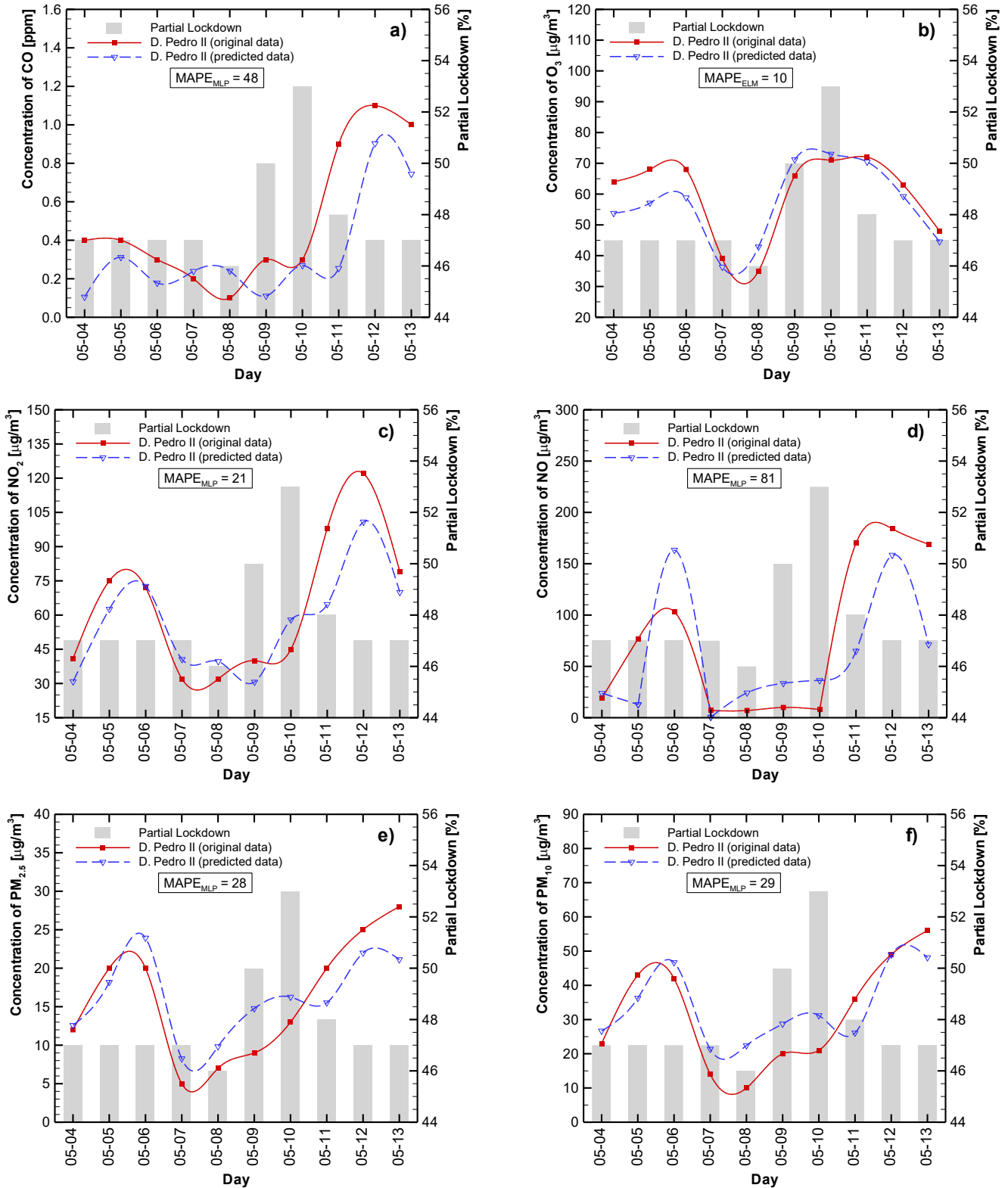
342 Figures 4 and 5 represent the observed (continuous red line) and best estimation
343 (dashed blue line) concentration levels for CO (a), O₃ (b), NO₂ (c), NO (d), PM_{2.5} (e), and
344 PM₁₀ (f) at Tietê and D. Pedro II stations, respectively during the period 4 – 13 May, 2020.
345 The lockdown level is indicated as shaded bars.

346 In general, the predicted results, using this approach, captured the original data
347 tendencies reasonably well, with a mean absolute percentage error (MAPE) of 30% for
348 almost all cases. The exceptions were at D. Pedro II station (CO – 48% and NO - 81%) (see
349 Tables A1 and A2 – Appendix A).

350 It is important to notice two distinct behaviors during the lockdown to the test set
351 period (see Figures 4 and 5). When the lockdown level remains unchanged (first 5 days), the
352 main influence can be ascribed to the meteorological variables (Figure A3 – shows the
353 meteorological raw data for the test period). But after five days in the test set, the percentage
354 lockdown jumps from 46% to 53% in two days. As the temperature and relative humidity
355 were relatively stable in the last five days, one can say that the lockdown is the main
356 contributor to the change in air pollutant level. Observe that ozone concentration has a
357 consistent relation with solar irradiation, with similar profiles. This behavior is in accordance
358 with those observed at the beginning of lockdown (March 17, 2020), as mentioned in section
359 3.1. The importance of maintaining continuous and consistent interventions to curb air
360 pollution is evident from the data displayed here. It is particularly important during extreme
361 air pollution events, and there is enough evidence that lockdown measures will nearly
362 instantly reduce air pollution levels.



363 **Figure 4.** Best estimation to predict CO (a), O₃ (b), NO₂ (c), NO (d), PM_{2.5} (e), and PM₁₀ (f) levels for
 364 Tietê station. Predictions are in dashed lines and observed levels in solid lines. The bars are the
 365 partial lockdown.



366 **Figure 5.** Best estimation to predict CO (a), O₃ (b), NO₂ (c), NO (d), PM_{2.5} (e), and PM₁₀ (f) levels for
 367 D. Pedro II station. Predictions are in dashed lines and observed levels in solid lines. The bars are the
 368 partial lockdown.

369 Each ANN architecture has positive and negative points. As discussed in Section 2.1,
370 the ESN is a recurrent model, presenting feedback loops of information in its hidden layer.
371 This characteristic may be relevant when dealing with data processing since more
372 information is available to form the output response. Additionally, together with the ELM,
373 their training processes require less computational effort than the RBF and MLP, since there
374 are no iterative processes to adjust their weights because the hidden layer is not modified. In
375 addition, other works have presented the capability of such models to overcome traditional,
376 fully trained architectures (Araujo et al., 2020; Siqueira et al., 2018, 2014, 2012a).

377 Despite the advantage and good results found in the literature for ESN, ELM, and
378 RBF (Siqueira et al., 2018, 2012b), the MLP errors were smaller than the others. It seems
379 clear that adjusting the hidden weights is an important step in nonlinear mapping
380 applications, as is presented in this investigation. In this case, there are a set of inputs of
381 variable nature (for example, temperature, humidity, and partial lockdown), and mapping
382 these values to another variable is not a trivial task (Kachba et al., 2020; Polezer et al.,
383 2018).

384 3.3 Hypothetical Scenarios

385 To predict the impact that the partial lockdown has on air quality, four hypothetical
386 scenarios were modeled: a minimum lockdown level (10%); possible vertical isolation (only
387 for COVID-19 high-risk groups – over-60s and people with chronic disease, diabetics,
388 among others) (30%); the considered ideal lockdown percentage (70%); and an extreme
389 isolation action (90%). The results are compared in Figures 6 and 7, with results for AQMS
390 Tietê D. Pedro II, respectively. The red lines correspond to 10% lockdown, the pink lines to
391 30% lockdown, the blue lines to 70% lockdown, and the green lines to 90% lockdown. The
392 pollutant designation (a -f) is the same as for Figures 4 and 5.

393 The data in Figure 6 (Tietê station) indicates that in general higher concentrations are
394 predicted for all pollutants at 10 (red line) and 30 (pink line) % lockdown. A different pattern
395 is observed for May 07 and May 08, whereby the lower lockdown also predicted low

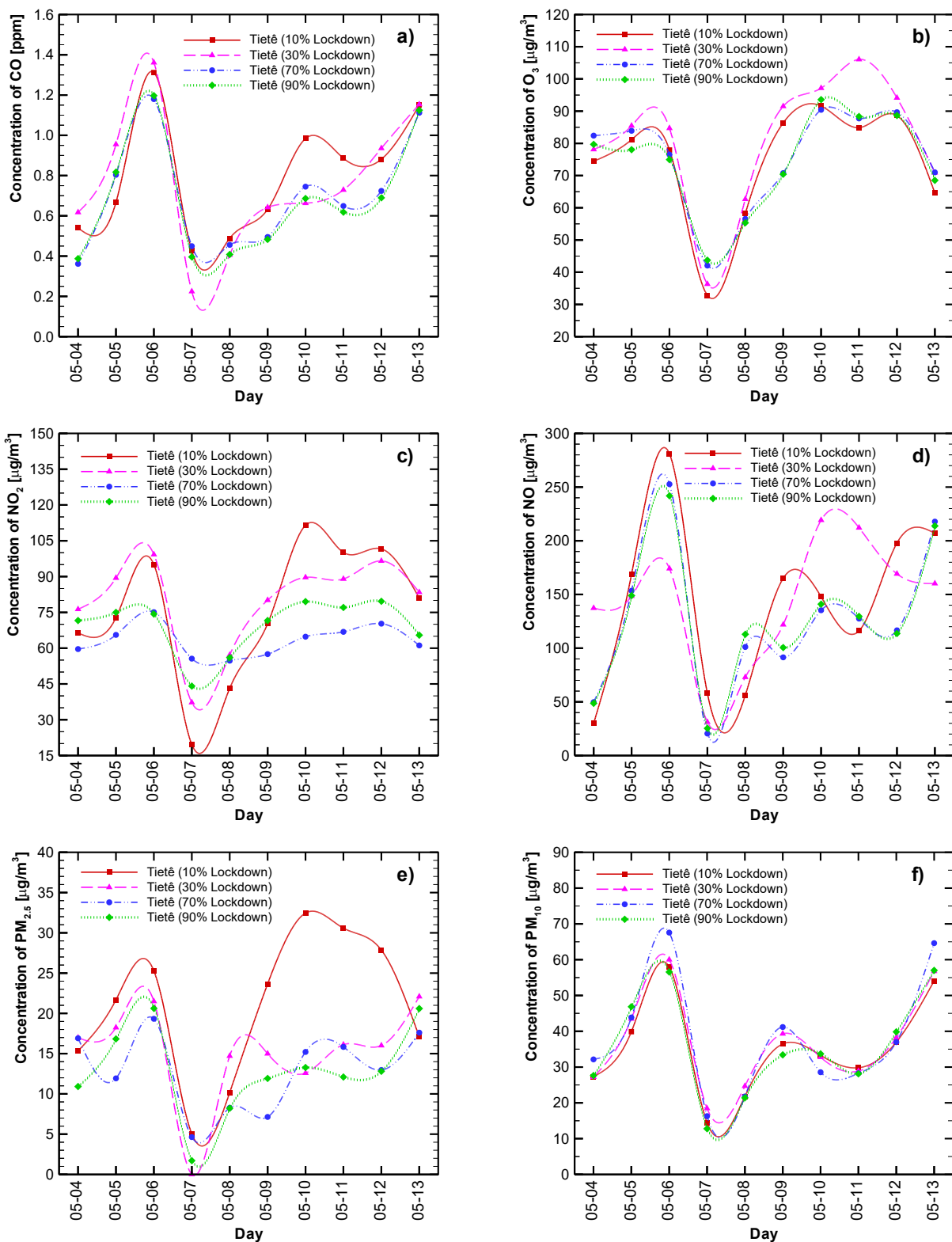
396 pollutant concentrations. During these two days, the meteorological conditions changed
397 abruptly (low temperature and solar irradiation, and high relative humidity - see Figure A3).
398 This scenario exemplifies the complex interdependency of air pollutant levels on several
399 variables. These findings suggest that when abrupt weather conditions are forecasted,
400 lockdown interventions should happen a few days earlier. Our data corroborate with the
401 recent publication of Hong et al. (2019) who reported that extreme weather events might be
402 a crucial mechanism by which air quality is influenced.

403 The predicted ozone concentration at Tietê station (Figure 6b) for the 30% lockdown
404 showed an unexpected behavior, presenting higher concentrations than 70% and 90%
405 lockdown. It may have been a consequence of the complexity of the variables that influence
406 air quality. Although this may be seen as a poor fit for the model, we need to emphasize that
407 this is one case out of twelve.

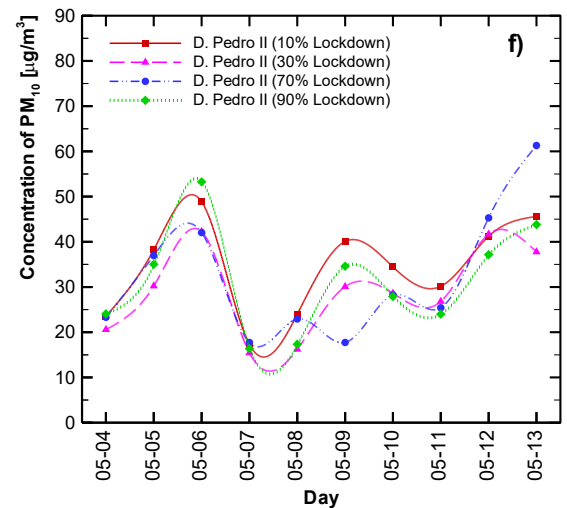
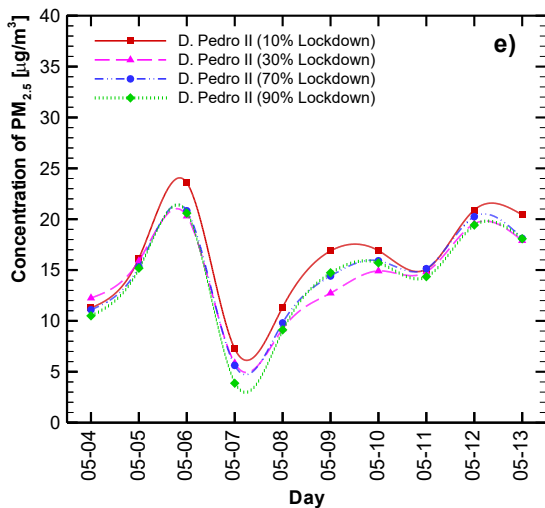
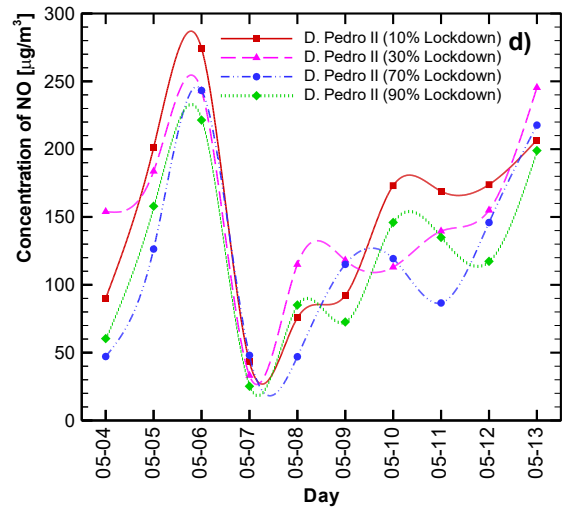
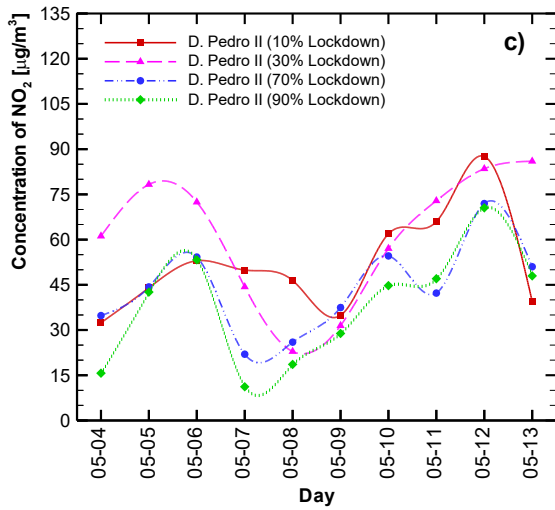
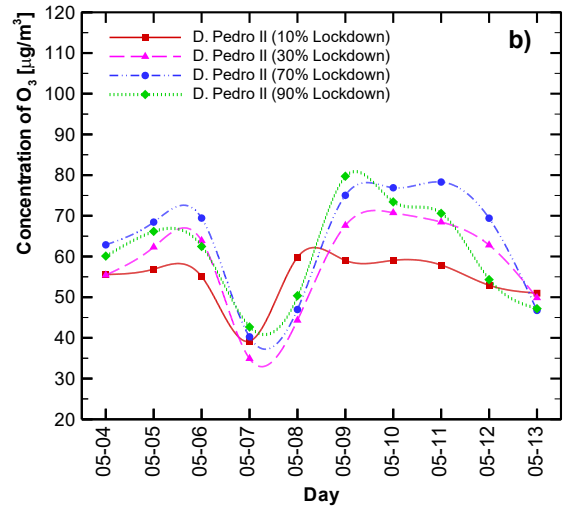
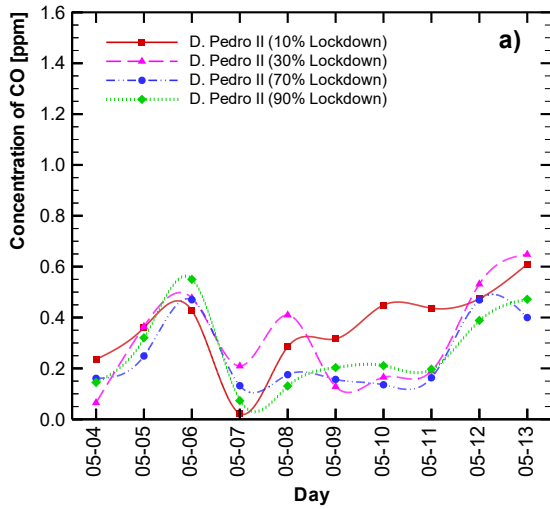
408 Although the same abrupt change in meteorological conditions was observed for 7
409 and 8 May at the D. Pedro II station (Figure 7), the ANN could estimate the response more
410 coherently than for the Tietê station. This may be due to other factors at play, influencing the
411 air pollutant level at this station. Observe that the ozone profiles are as expected, especially
412 for a 10% lockdown. It is important to highlight that the ANN prediction was good as only one
413 of the seven inputs were changed.

414 We also observe that the particulate matter levels are not greatly influenced by
415 lockdown (as reported by Nakada and Urban, 2020), especially the PM₁₀ concentration. At
416 the D. Pedro II station, the PM_{2.5} levels also stay very similar regardless of the lockdown
417 level.

418 We acknowledge that air pollutant levels have a complex set of variables that
419 determine it, and that even a powerful tool such as ANN cannot always accurately predict
420 the level. However, the data presented here provides adequate evidence that ANN can be
421 used successfully to estimate the impact of different levels of lockdown will have on the air
422 quality.



423 **Figure 6.** Hypothetical scenarios considering the impact that 10% (red line), 30% (pink like), 70%
 424 (blue line), and 90% (green line) lockdown would have on CO (a), O₃ (b), NO₂ (c), NO (d), PM_{2.5} (e),
 425 and PM₁₀ (f) levels for AQMS Tieté.



426 **Figure 7.** Hypothetical scenarios considering the impact that 10% (red line), 30% (pink like), 70%
 427 (blue line), and 90% (green line) lockdown would have on CO (a), O₃ (b), NO₂ (c), NO (d), PM_{2.5} (e),
 428 and PM₁₀ (f) levels for AQMS D. Pedro II.

429 **4. CONCLUSION**

430 Artificial Neural Networks were able to predict how changes in the level of lockdown
431 affected air quality in São Paulo City. We have shown that even when using a restricted data
432 set of pollutant levels together with meteorological information, the ANN results showed
433 Mean Absolute Percentage Error (MAPE) around 30%.

434 The result of the ANN approach to four hypothetical scenarios of lockdown (i.e., 10%,
435 30%, 70%, and 90%) showed evidence of the complexity of the calculation problem as a
436 consequence of the abrupt meteorological changes.

437 For the first time, ANN were used as a tool to describe the equilibrium between air
438 pollution, COVID-19 cases, and the partial lockdown, which can be employed in several
439 national contexts. This approach's predictive power allows governmental bodies and policy
440 makers to manage lockdown responsibly ensuring minimal economic impact. This method
441 will lead to improved air pollution control measures (and potentially COVID-19 mortality) by
442 enforcing a lockdown level that will still sustain sufficient economic activities. Furthermore, in
443 the light of the global drive to improve air quality and work towards zero emissions, this
444 approach could also be used in the future to reach emission target levels.

445 **ACKNOWLEDGEMENT**

446 The authors would like to thank the Environmental Company of São Paulo State
447 (CETESB) to the availability of meteorological and air quality data.

448 **REFERENCES**

449 Andrade, M. de F., Kumar, P., de Freitas, E.D., Ynoue, R.Y., Martins, J., Martins, L.D.,
450 Nogueira, T., Perez-Martinez, P., de Miranda, R.M., Albuquerque, T., Gonçalves,
451 F.L.T., Oyama, B., Zhang, Y., 2017. Air quality in the megacity of São Paulo: Evolution
452 over the last 30 years and future perspectives. Atmos. Environ.

453 <https://doi.org/10.1016/j.atmosenv.2017.03.051>

454 Araujo, L.N., Belotti, J.T., Antonini Alves, T., Tadano, Y. de S., Siqueira, H., 2020. Ensemble
455 method based on Artificial Neural Networks to estimate air pollution health risks.
456 Environ. Model. Softw. <https://doi.org/10.1016/j.envsoft.2019.104567>

457 CETESB, 2020. Environmental Company of São Paulo State (in Portuguese: Companhia
458 Ambiental do Estado de São Paulo) <http://cetesb.sp.gov.br/ar/qualar/> (accessed 15
459 May 2020).

460 Chauhan, A., Singh, R.P., 2020. Decline in PM_{2.5} concentrations over major cities around the
461 world associated with COVID-19. Environ. Res.
462 <https://doi.org/10.1016/j.envres.2020.109634>

463 CNBC, 2020. South America is a 'new epicenter' of the coronavirus pandemic, WHO says.
464 [https://www.cnbc.com/2020/05/22/south-america-is-a-new-epicenter-of-the-](https://www.cnbc.com/2020/05/22/south-america-is-a-new-epicenter-of-the-coronavirus-pandemic-who-says.html)
465 [coronavirus-pandemic-who-says.html](https://www.cnbc.com/2020/05/22/south-america-is-a-new-epicenter-of-the-coronavirus-pandemic-who-says.html) (accessed 25 May 2020).

466 Dantas, G., Siciliano, B., França, B.B., da Silva, C.M., Arbilla, G., 2020. The impact of
467 COVID-19 partial lockdown on the air quality of the city of Rio de Janeiro, Brazil. Sci.
468 Total Environ. <https://doi.org/10.1016/j.scitotenv.2020.139085>

469 de Castro, L.N., 2007. Fundamentals of natural computing: an overview. Phys. Life Rev.
470 <https://doi.org/10.1016/j.plrev.2006.10.002>

471 de Oliveira, G., Chen, J.M., Stark, S.C., Berenguer, E., Moutinho, P., Artaxo, P., Anderson,
472 L.O., Aragão, L.E.O.C., 2020. Smoke pollution's impacts in Amazonia. Science.
473 <https://doi.org/10.1126/science.abd5942>

474 dos Santos, E.P., Von Zuben, F.J., 1999. Improved second-order training algorithms for
475 globally and partially recurrent neural networks, in: Proceedings of the International
476 Joint Conference on Neural Networks. <https://doi.org/10.1109/ijcnn.1999.832591>

477 Dutheil, F., Baker, J.S., Navel, V., 2020. COVID-19 as a factor influencing air pollution?
478 Environ. Pollut. <https://doi.org/10.1016/j.envpol.2020.114466>

479 Figueiredo, E., Macedo, M., Siqueira, H.V., Santana, C.J., Gokhale, A., Bastos-Filho, C.J.A.,
480 2019. Swarm intelligence for clustering — A systematic review with new perspectives
481 on data mining. Eng. Appl. Artif. Intell. <https://doi.org/10.1016/j.engappai.2019.04.007>

482 Gentner, D.R., Harley, R.A., Miller, A.M., Goldstein, A.H., 2009. Diurnal and seasonal
483 variability of gasoline-related volatile organic compound emissions in Riverside,

484 California. Environ. Sci. Technol. <https://doi.org/10.1021/es9006228>

485 Guardani, R., Nascimento, C.A.O., Guardani, M.L.G., Martins, M.H.R.B., Romano, J., 1999.
486 Study of atmospheric ozone formation by means of a neural network-based model. J.
487 Air Waste Manag. Assoc. <https://doi.org/10.1080/10473289.1999.10463806>

488 Harley, R.A., Marr, L.C., Lehner, J.K., Giddings, S.N., 2005. Changes in motor vehicle
489 emissions on diurnal to decadal time scales and effects on atmospheric composition.
490 Environ. Sci. Technol. <https://doi.org/10.1021/es048172+>

491 Haykin, S., 2008. Neural Networks and Learning Machines, Pearson Prentice Hall New
492 Jersey USA 936 pLinks. <https://doi.org/978-0131471399>

493 Hong, C., Zhang, Q., Zhang, Y., Davis, S.J., Tong, D., Zheng, Y., Liu, Z., Guan, D., He, K.,
494 Schellnhuber, H.J., 2019. Impacts of climate change on future air quality and human
495 health in China. Proc. Natl. Acad. Sci. U. S. A.
496 <https://doi.org/10.1073/pnas.1812881116>

497 Huang, G. Bin, Zhu, Q.Y., Siew, C.K., 2006. Extreme learning machine: Theory and
498 applications. Neurocomputing. <https://doi.org/10.1016/j.neucom.2005.12.126>

499 Huang, Gao, Huang, Guang Bin, Song, S., You, K., 2015. Trends in extreme learning
500 machines: A review. Neural Networks. <https://doi.org/10.1016/j.neunet.2014.10.001>

501 Jaeger, H., 2002. Short term memory in echo state networks, GMD Report 152.

502 Jaeger, H., 2001. The “echo state” approach to analysing and training recurrent neural
503 networks. GMD Rep.

504 Kachba, Y., de Genaro Chiroli, D.M., Belotti, J.T., Antonini Alves, T., de Souza Tadano, Y.,
505 Siqueira, H., 2020. Artificial neural networks to estimate the influence of vehicular
506 emission variables on morbidity and mortality in the largest metropolis in South
507 America. Sustain. <https://doi.org/10.3390/su12072621>

508 Kassomenos, P., Petrakis, M., Sarigiannis, D., Gotti, A., Karakitsios, S., 2011. Identifying the
509 contribution of physical and chemical stressors to the daily number of hospital
510 admissions implementing an artificial neural network model. Air Qual. Atmos. Heal.
511 <https://doi.org/10.1007/s11869-011-0139-2>

512 Le, T., Wang, Y., Liu, L., Yang, J., Yung, Y. L., Li, G., Seinfeld, J.H., 2020. Unexpected air
513 pollution with marked emission reductions during the COVID-19 outbreak in China.
514 Science. <https://doi.org/10.1126/science.abb7431>

- 515 Li, K., Jacob, D.J., Liao, H., Shen, L., Zhang, Q., Bates, K.H., 2019. Anthropogenic drivers of
516 2013–2017 trends in summer surface ozone in China. *Proc. Natl. Acad. Sci. U. S. A.*
517 <https://doi.org/10.1073/pnas.1812168116>
- 518 Li, L., Li, Q., Huang, L., Wang, Q., Zhu, A., Xu, J., Liu, Ziyi, Li, H., Shi, L., Li, R., Azari, M.,
519 Wang, Y., Zhang, X., Liu, Zhiqiang, Zhu, Y., Zhang, K., Xue, S., Ooi, M.C.G., Zhang, D.,
520 Chan, A., 2020. Air quality changes during the COVID-19 lockdown over the Yangtze
521 River Delta Region: An insight into the impact of human activity pattern changes on air
522 pollution variation. *Sci. Total Environ.* <https://doi.org/10.1016/j.scitotenv.2020.139282>
- 523 Marr, L.C., Harley, R.A., 2002. Spectral analysis of weekday-weekend differences in ambient
524 ozone, nitrogen oxide, and non-methane hydrocarbon time series in California. *Atmos.*
525 *Environ.* [https://doi.org/10.1016/S1352-2310\(02\)00188-7](https://doi.org/10.1016/S1352-2310(02)00188-7)
- 526 Martins, L.D., Andrade, M.D.F., 2008. Ozone formation potentials of volatile organic
527 compounds and ozone sensitivity to their emission in the megacity of São Paulo, Brazil.
528 *Water. Air. Soil Pollut.* <https://doi.org/10.1007/s11270-008-9740-x>
- 529 Muhammad, S., Long, X., Salman, M., 2020. COVID-19 pandemic and environmental
530 pollution: A blessing in disguise? *Sci. Total Environ.*
531 <https://doi.org/10.1016/j.scitotenv.2020.138820>
- 532 Nakada, L.Y.K., Urban, R.C., 2020. COVID-19 pandemic: Impacts on the air quality during
533 the partial lockdown in São Paulo state, Brazil. *Sci. Total Environ.*
534 <https://doi.org/10.1016/j.scitotenv.2020.139087>
- 535 Polezer, G., Tadano, Y.S., Siqueira, H. V., Godoi, A.F.L., Yamamoto, C.I., de André, P.A.,
536 Pauliquevis, T., Andrade, M. de F., Oliveira, A., Saldiva, P.H.N., Taylor, P.E., Godoi,
537 R.H.M., 2018. Assessing the impact of PM_{2.5} on respiratory disease using artificial
538 neural networks. *Environ. Pollut.* <https://doi.org/10.1016/j.envpol.2017.12.111>
- 539 Rodríguez-Urrego, D., Rodríguez-Urrego, L., 2020. Air quality during the COVID-19: PM_{2.5}
540 analysis in the 50 most polluted capital cities in the world. *Environ. Pollut.*
541 <https://doi.org/10.1016/j.envpol.2020.115042>
- 542 SEADE, 2020. Statistical Portal of São Paulo State (in Portuguese: Portal de Estatísticas do
543 Estado de São Paulo) <https://www.seade.gov.br/coronavirus/> (accessed 12 June 2020).
- 544 Seinfeld, J. H.; Pandis, S.N., 2016. Atmospheric chemistry and physics: from air pollution to
545 climate change. John Wiley & Sons.

546 Sharma, S., Zhang, M., Anshika, Gao, J., Zhang, H., Kota, S.H., 2020. Effect of restricted
547 emissions during COVID-19 on air quality in India. *Sci. Total Environ.*
548 <https://doi.org/10.1016/j.scitotenv.2020.138878>

549 Shehzad, K., Sarfraz, M., Shah, S.G.M., 2020. The impact of COVID-19 as a necessary evil
550 on air pollution in India during the lockdown. *Environ. Pollut.*
551 <https://doi.org/10.1016/j.envpol.2020.115080>

552 Sicard, P., De Marco, A., Agathokleous, E., Feng, Z., Xu, X., Paoletti, E., Rodriguez, J.J.D.,
553 Calatayud, V., 2020. Amplified ozone pollution in cities during the COVID-19 lockdown.
554 *Sci. Total Environ.* <https://doi.org/10.1016/j.scitotenv.2020.139542>

555 Siciliano, B., Dantas, G., da Silva, C.M., Arbilla, G., 2020. Increased ozone levels during the
556 COVID-19 lockdown: Analysis for the city of Rio de Janeiro, Brazil. *Sci. Total Environ.*
557 <https://doi.org/10.1016/j.scitotenv.2020.139765>

558 Sillman, S., 2003. Tropospheric Ozone and Photochemical Smog, in: *Treatise on*
559 *Geochemistry.* <https://doi.org/10.1016/B0-08-043751-6/09053-8>

560 Sillman, S., 1999. The relation between ozone, NO(x) and hydrocarbons in urban and
561 polluted rural environments. *Atmos. Environ.* [https://doi.org/10.1016/S1352-](https://doi.org/10.1016/S1352-2310(98)00345-8)
562 [2310\(98\)00345-8](https://doi.org/10.1016/S1352-2310(98)00345-8)

563 Siqueira, H., Boccato, L., Attux, R., Lyra, C., 2014. Unorganized machines for seasonal
564 streamflow series forecasting. *Int. J. Neural Syst.*
565 <https://doi.org/10.1142/S0129065714300095>

566 Siqueira, H., Boccato, L., Attux, R., Lyra, C., 2012a. Echo state networks and extreme
567 learning machines: A comparative study on seasonal streamflow series prediction, in:
568 *Lecture Notes in Computer Science (Including Subseries Lecture Notes in Artificial*
569 *Intelligence and Lecture Notes in Bioinformatics).* [https://doi.org/10.1007/978-3-642-](https://doi.org/10.1007/978-3-642-34481-7_60)
570 [34481-7_60](https://doi.org/10.1007/978-3-642-34481-7_60)

571 Siqueira, H., Boccato, L., Attux, R., Lyra, C., 2012b. Echo state networks for seasonal
572 streamflow series forecasting, in: *Lecture Notes in Computer Science (Including*
573 *Subseries Lecture Notes in Artificial Intelligence and Lecture Notes in Bioinformatics).*
574 https://doi.org/10.1007/978-3-642-32639-4_28

575 Siqueira, H., Boccato, L., Luna, I., Attux, R., Lyra, C., 2018. Performance analysis of
576 unorganized machines in streamflow forecasting of Brazilian plants. *Appl. Soft Comput.*
577 *J.* <https://doi.org/10.1016/j.asoc.2018.04.007>

578 Siqueira, H., Luna, I., 2019. Performance comparison of feedforward neural networks
579 applied to streamflow series forecasting. *Math. Eng. Sci. Aerosp.*

580 Siqueira, H., Macedo, M., Tadano, Y.S., Antonini Alves, T., Stevan Jr., S.L., Oliveira, D.S.,
581 Marinho, M.H.N., Neto, P.S.G. de M., Oliveira, João F. L. de, Luna, I., Filho, M. de A.L.,
582 Sarubbo, L.A., Converti, A., 2020. Selection of Temporal Lags for Predicting Riverflow
583 Series from Hydroelectric Plants Using Variable Selection Methods. *Energies*.
584 <https://doi.org/10.3390/en13164236>

585 Stedman, D.H., 2004. Photochemical ozone formation, simplified. *Environ. Chem.*
586 <https://doi.org/10.1071/EN04032>

587 The Guardian, 2020. Brazil scales back environmental enforcement amid coronavirus
588 outbreak. [https://www.theguardian.com/world/2020/mar/27/brazil-scales-back-](https://www.theguardian.com/world/2020/mar/27/brazil-scales-back-environmental-enforcement-coronavirus-outbreak-deforestation)
589 [environmental-enforcement-coronavirus-outbreak-deforestation](https://www.theguardian.com/world/2020/mar/27/brazil-scales-back-environmental-enforcement-coronavirus-outbreak-deforestation) (accessed 22 May
590 2020).

591 The Lancet, 2020. COVID-19 in Brazil: “So what?” *Lancet*. [https://doi.org/10.1016/S0140-](https://doi.org/10.1016/S0140-6736(20)31095-3)
592 [6736\(20\)31095-3](https://doi.org/10.1016/S0140-6736(20)31095-3)

593 Tobías, A., Carnerero, C., Reche, C., Massagué, J., Via, M., Minguillón, M.C., Alastuey, A.,
594 Querol, X., 2020. Changes in air quality during the lockdown in Barcelona (Spain) one
595 month into the SARS-CoV-2 epidemic. *Sci. Total Environ.*
596 <https://doi.org/10.1016/j.scitotenv.2020.138540>

597 Wu, X., Nethery, R.C., Sabath, B.M., Braun, D., Dominici, F., 2020. Exposure to air pollution
598 and COVID-19 mortality in the United States. *medRxiv*.
599 <https://doi.org/10.1101/2020.04.05.20054502>

600 Zambrano-Monserrate, M.A., Ruano, M.A., Sanchez-Alcalde, L., 2020. Indirect effects of
601 COVID-19 on the environment. *Sci. Total Environ.*
602 <https://doi.org/10.1016/j.scitotenv.2020.138813>

603

APPENDIX A

604

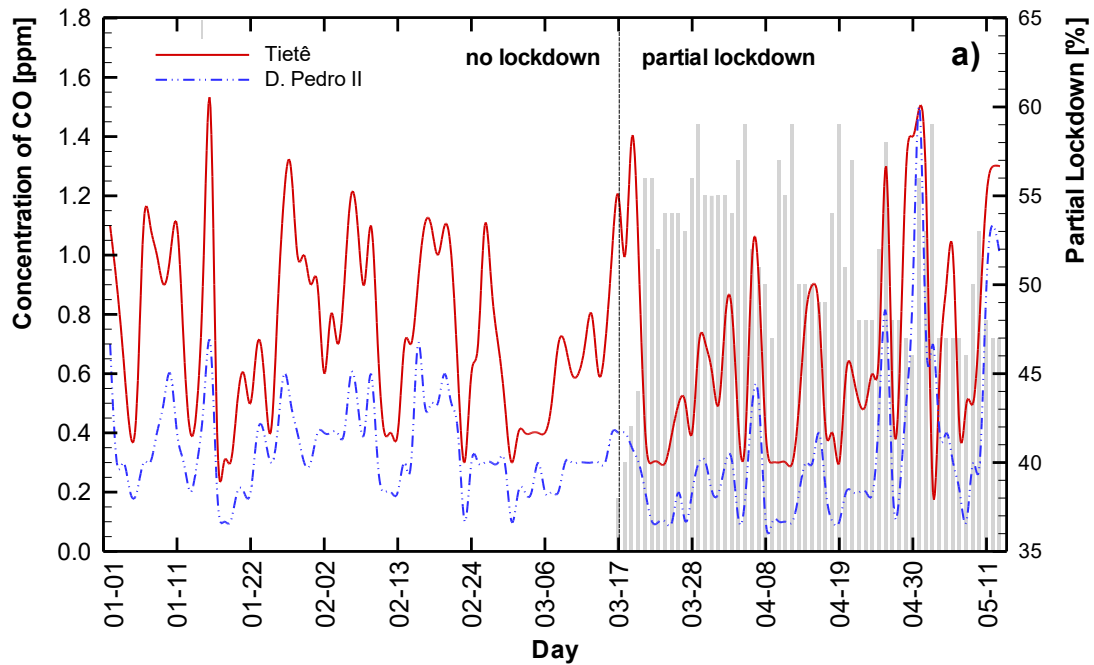
Figure A1 shows the behavior of each target (air pollutant concentration) before any

605

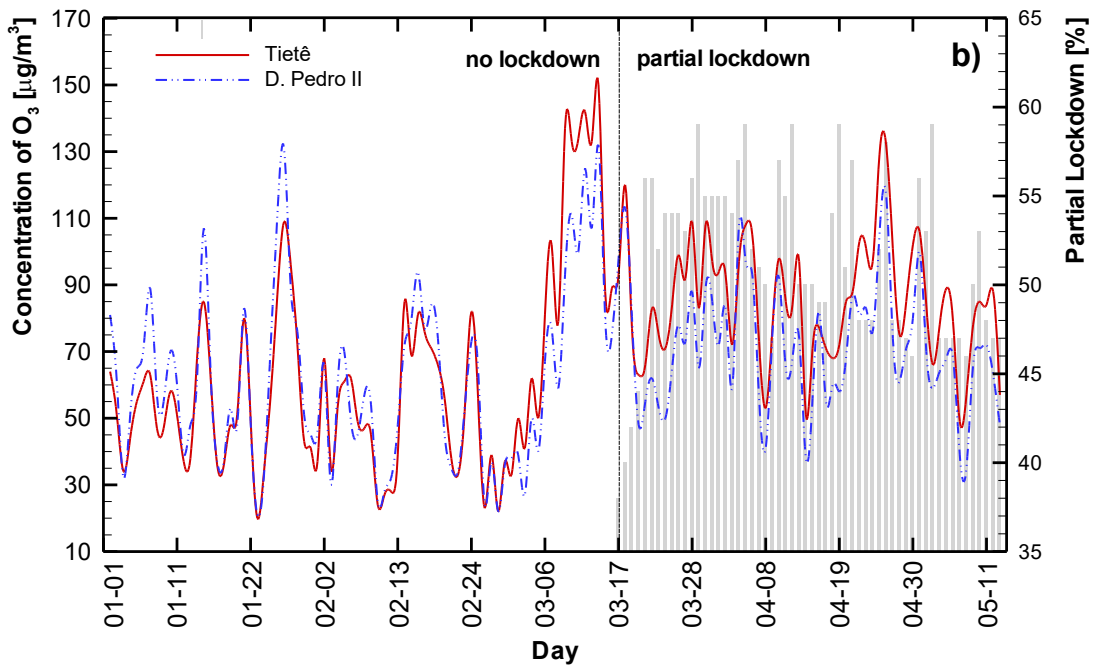
modeling. A black line was included on the first day of partial lockdown available data (March

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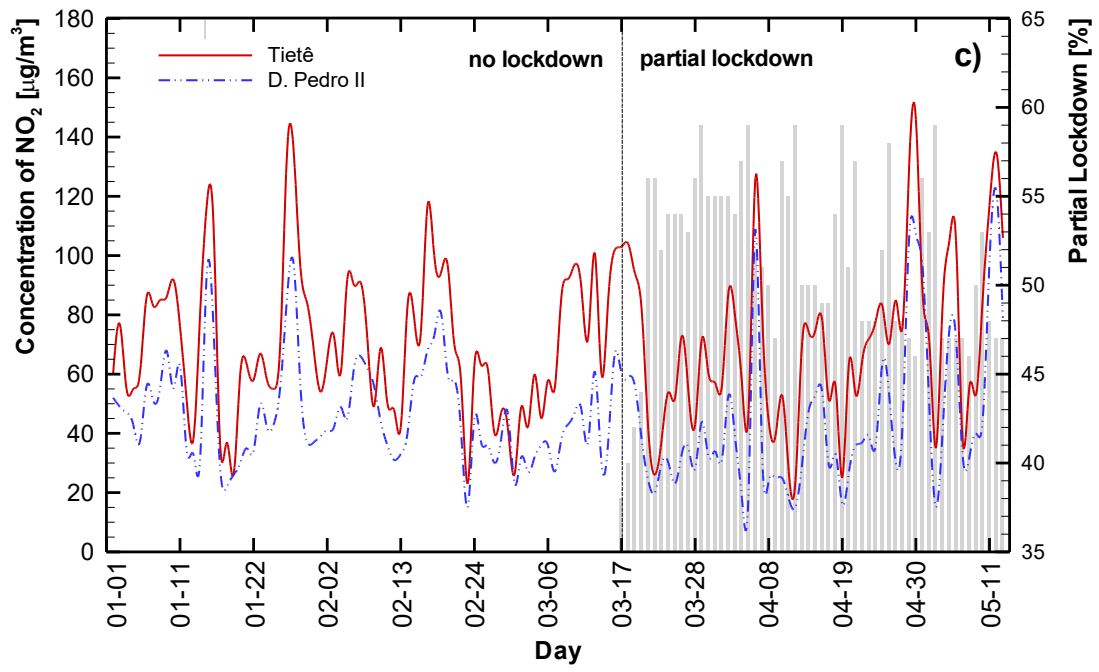
17, 2020). The shaded bars are the lockdown daily levels.



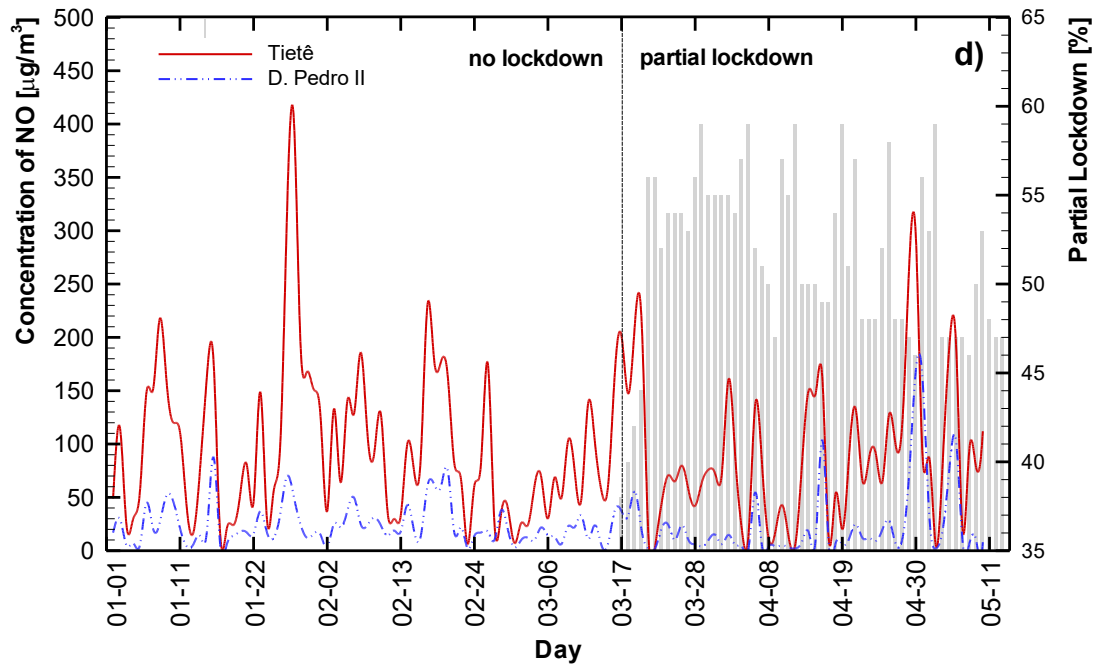
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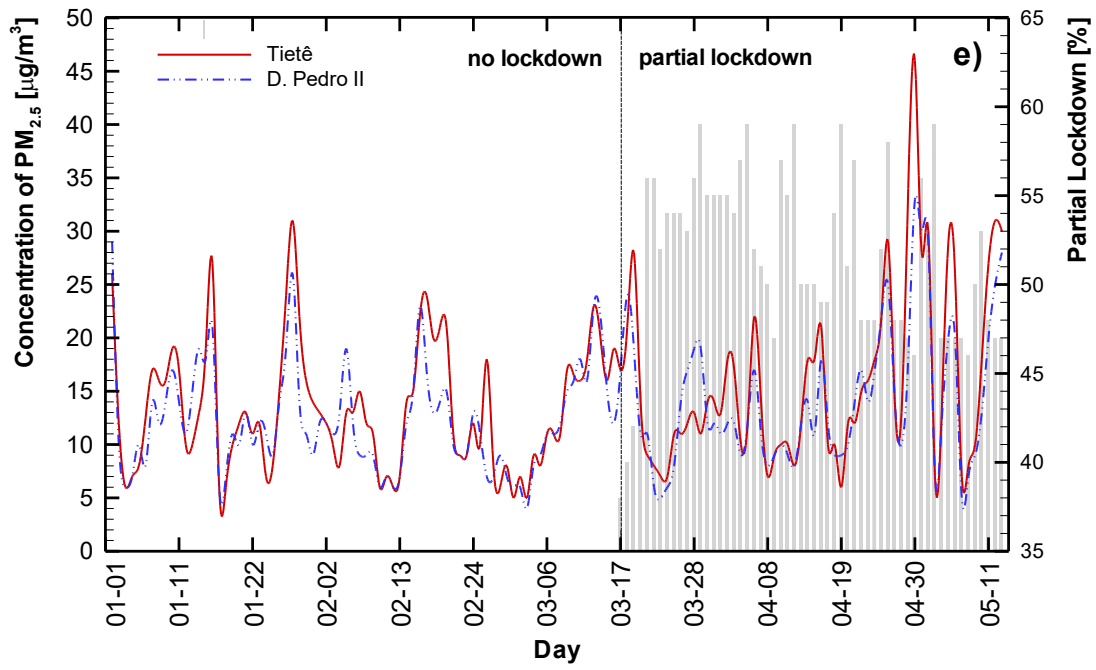
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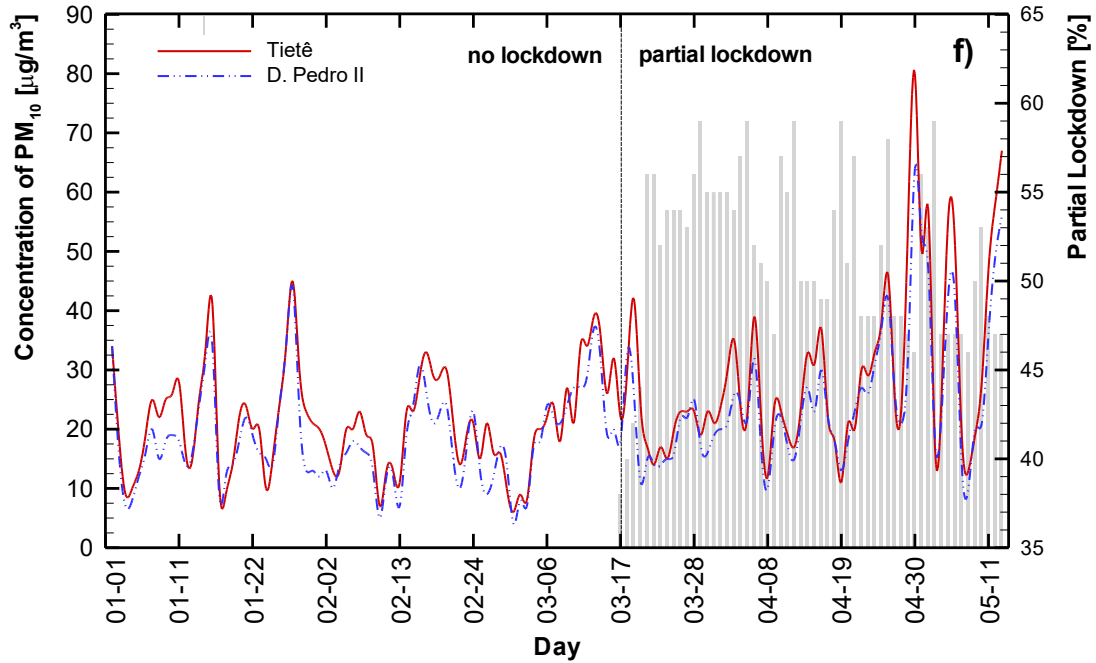
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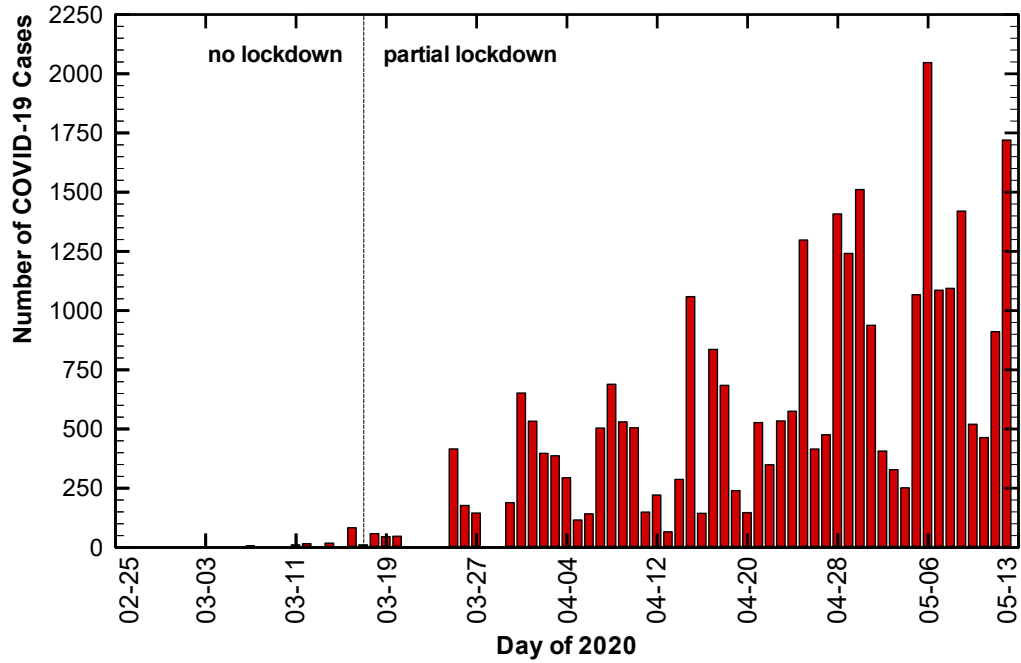


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Figure A1. Concentrations of CO [ppm] (a), O₃ [µg/m³] (b), NO₂ [µg/m³] (c), NO [µg/m³] (d), PM_{2.5} [µg/m³] (e), and PM₁₀ [µg/m³] (f) according to the date.



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Figure A2. Number of COVID-19 new cases by day.

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Tables A1 and A2 show the ANN computational results at the AQMS Tietê (ring road station) and AQMS D. Pedro II (downtown station), respectively. The addressed scenarios involves the errors based on the Mean Square Error (MSE); Mean Absolute Error (MAE); and Mean Absolute Percentage Error (MAPE); with all inputs, without the number of COVID-19 new cases (without COVID), and also without the number of COVID-19 new cases and the partial lockdown (without COVID and lockdown). Also, the results considering or not the use of Z-Score are presented. The shade of gray values are the results showing the best performances (lower MSE).

			CO				PM ₁₀				PM _{2.5}			
			NN	MSE	MAE	MAPE	NN	MSE	MAE	MAPE	NN	MSE	MAE	MAPE
Without Z-Score	All Inputs	ELM	3	0.082	0.241	27.336	55	302.983	14.792	41.053	40	118.040	9.379	68.917
		ESN	3	0.104	0.275	35.007	10	226.953	12.543	36.529	70	84.089	8.198	60.844
		MLP	5	0.056	0.189	21.054	35	125.516	89.970	22.795	7	58.597	6.129	32.043
		RBF	90	0.107	0.290	41.937	90	361.600	17.400	71.145	7	116.624	8.700	82.812
	Without COVID	ELM	3	0.074	0.229	29.136	25	363.393	16.372	47.747	25	121.795	9.633	71.572
		ESN	3	0.068	0.227	29.313	17	355.852	16.807	55.639	17	73.759	7.760	54.564
		MLP	7	0.054	0.189	20.939	3	228.692	12.619	32.008	5	54.020	5.728	25.911
		RBF	90	0.107	0.290	41.912	90	361.758	17.414	71.173	90	116.684	8.708	82.843
	Without COVID and Lockdown	ELM	15	0.196	0.355	41.284	20	441.157	18.462	59.690	80	142.555	10.548	76.858
		ESN	5	0.072	0.216	24.928	30	447.592	19.187	62.825	60	124.008	9.233	68.624
		MLP	5	0.106	0.248	24.881	5	274.828	12.145	29.001	50	73.181	5.655	22.787
		RBF	90	0.107	0.291	42.002	90	361.429	17.544	71.302	90	116.337	8.720	82.882
With Z-Score	All Inputs	ELM	20	0.139	0.334	37.919	50	332.813	16.294	55.894	45	98.053	8.778	59.417
		ESN	3	0.090	0.243	27.409	20	274.429	13.738	54.822	35	110.166	8.710	67.406
		MLP	5	0.039	0.135	16.132	50	172.991	11.027	30.508	3	56.984	6.054	32.115
		RBF	50	0.107	0.290	41.937	90	361.600	17.400	71.145	3	116.628	8.705	82.814
	Without COVID	ELM	3	0.084	0.242	26.932	35	472.051	19.277	65.402	30	118.146	9.505	69.580
		ESN	3	0.091	0.250	27.321	45	361.426	17.398	61.243	25	113.501	8.461	71.963
		MLP	5	0.072	0.214	23.736	3	229.299	12.624	32.513	5	43.484	5.494	23.216
		RBF	90	0.106	0.290	41.855	90	361.762	17.414	71.173	10	115.453	8.930	82.345
	Without COVID and Lockdown	ELM	55	0.226	0.389	45.342	25	445.110	18.727	58.367	35	140.325	9.701	72.964
		ESN	3	0.095	0.267	31.366	60	326.904	16.261	51.552	8	88.244	8.180	51.650
		MLP	5	0.096	0.237	24.660	5	268.293	12.734	30.227	12	69.210	5.750	24.045
		RBF	90	0.109	0.292	42.106	60	364.251	17.591	71.336	70	116.452	8.750	82.979

		NO ₂				NO				O ₃				
		NN	MSE	MAE	MAPE	NN	MSE	MAE	MAPE	NN	MSE	MAE	MAPE	
Without Z-Score	All Inputs	ELM	3	570.143	18.949	26.896	5	5078.089	59.080	50.701	3	152.666	9.651	14.731
		ESN	3	523.303	17.204	27.192	12	15628.355	107.259	111.761	3	175.935	8.487	14.828
		MLP	70	608.078	19.198	19.419	70	3433.167	42.272	27.680	3	99.301	8.259	11.462
		RBF	5	886.859	25.285	36.021	5	8659.767	76.705	156.927	3	302.236	11.996	20.629
	Without COVID	ELM	3	510.554	19.691	24.813	3	49881.516	137.964	1172.231	3	276.657	13.781	18.821
		ESN	3	627.172	21.815	32.491	3	78881.981	219.920	827.994	15	586.099	20.913	30.830
		MLP	55	353.000	16.116	18.310	25	9192.320	42.764	66.788	70	123.546	8.233	13.090
		RBF	7	867.116	24.994	35.427	3	111554.876	285.136	2637.350	80	301.881	11.907	20.519
	Without COVID and Lockdown	ELM	3	749.904	24.518	33.485	3	6961.111	72.554	73.985	3	131.299	9.394	12.071
		ESN	15	1861.376	36.211	58.479	45	25439.828	123.837	229.898	3	269.781	13.052	16.959
		MLP	3	851.311	25.370	26.479	50	4811.999	53.290	37.489	90	101.570	6.857	11.121
		RBF	90	898.970	25.540	36.594	3	9205.445	81.435	160.595	90	301.796	11.914	20.527
With Z-Score	All Inputs	ELM	3	487.354	18.486	21.649	3	6057.239	63.555	91.520	3	136.731	9.546	15.309
		ESN	3	495.247	17.895	26.537	8	13938.590	98.401	127.912	7	238.876	13.117	18.446
		MLP	40	582.091	18.977	18.885	80	4080.494	49.610	28.741	17	113.758	8.885	12.352
		RBF	10	889.713	25.358	36.396	3	8805.067	77.825	158.019	3	302.235	11.985	20.620
	Without COVID	ELM	3	474.979	15.936	20.734	3	60013.692	214.782	1084.430	8	264.538	13.597	18.703
		ESN	3	842.937	25.159	36.858	3	80788.855	203.875	1247.237	5	241.888	14.196	18.513
		MLP	45	533.281	16.090	15.524	3	7461.860	40.342	65.189	90	127.865	9.106	14.463
		RBF	12	804.605	24.276	35.628	3	109705.490	283.427	2652.255	90	302.142	11.933	20.552
	Without COVID and Lockdown	ELM	3	695.472	18.534	27.748	3	5391.007	57.438	58.166	3	131.299	9.394	12.071
		ESN	60	1887.730	36.051	58.700	40	25912.029	124.420	240.393	3	347.065	15.199	19.339
		MLP	5	763.582	23.187	22.621	80	4334.745	49.248	33.712	90	134.022	8.179	12.276
		RBF	90	898.803	25.535	36.590	90	9225.294	78.847	160.696	90	300.184	11.808	20.385

628 NN: Number of neurons; MSE: Mean Square Error; MAE: Mean Absolute Error; MAPE: Mean Absolute Percentage Error; *With COVID means including the
629 number of COVID-19 new cases and the partial lockdown.

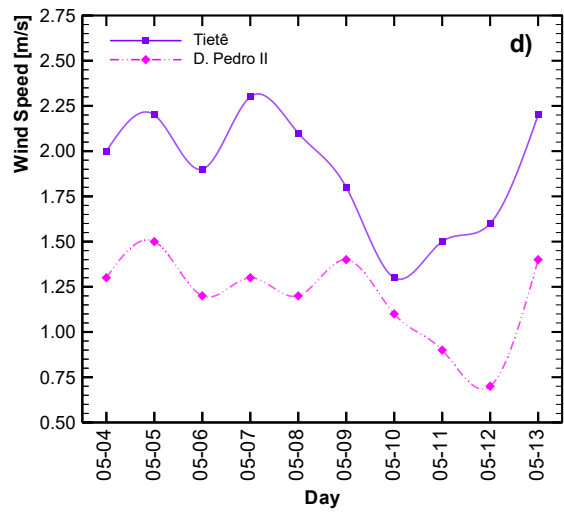
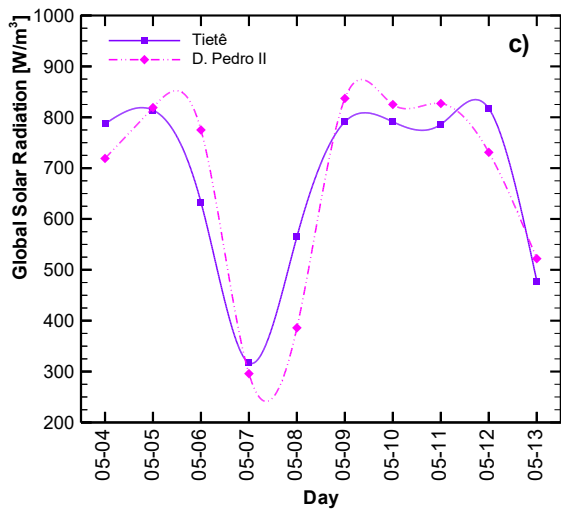
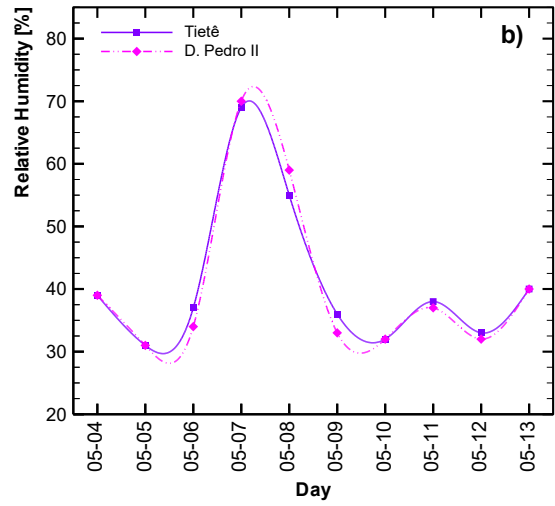
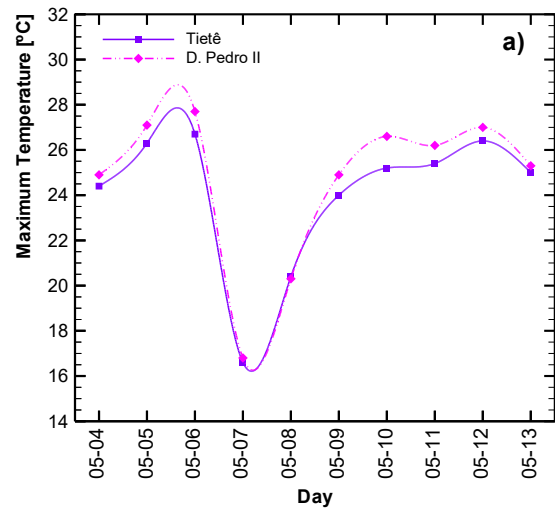
630 **Table A1.** Computational results for Tietê Station

		CO				PM ₁₀				PM _{2.5}				
		NN	MSE	MAE	MAPE	NN	MSE	MAE	MAPE	NN	MSE	MAE	MAPE	
Without Z-Score	All Inputs	ELM	5	0.145	0.274	46.448	3	201.221	11.965	49.041	3	51.303	5.935	39.177
		ESN	3	0.236	0.446	175.795	3	71.156	7.847	32.640	3	57.354	6.088	69.635
		MLP	3	0.133	0.257	45.423	5	62.832	6.936	28.881	3	18.305	3.323	19.678
		RBF	3	0.206	0.420	175.738	90	232.000	13.800	67.802	90	61.650	6.700	71.874
	Without COVID	ELM	3	0.088	0.220	60.253	3	167.635	10.469	34.639	3	67.776	6.738	42.641
		ESN	12	0.303	0.467	206.963	15	417.394	17.723	93.584	5	88.044	8.547	79.455
		MLP	45	0.101	0.242	42.666	15	73.638	7.079	26.922	60	18.683	3.579	23.942
		RBF	3	0.204	0.419	174.983	90	232.206	13.805	67.810	90	61.697	6.702	71.883
	Without COVID and Lockdown	ELM	3	0.119	0.251	59.499	3	378.428	15.714	44.049	3	66.176	6.592	41.157
		ESN	7	0.313	0.478	185.849	3	252.998	13.814	68.271	3	82.455	7.507	75.413
		MLP	45	0.111	0.276	55.548	40	79.912	8.113	33.120	40	27.414	4.675	25.319
		RBF	5	0.198	0.414	172.176	94	232.421	13.810	67.819	90	61.713	6.703	71.886
With Z-Score	All Inputs	ELM	3	0.096	0.227	39.350	3	127.320	9.337	33.624	3	33.077	5.200	48.258
		ESN	7	0.340	0.511	201.290	3	218.389	12.568	60.732	3	54.302	6.176	67.292
		MLP	3	0.069	0.200	48.480	80	76.843	7.220	24.859	3	15.806	3.582	28.000
		RBF	5	0.205	0.420	175.627	90	232.000	13.800	67.802	60	61.650	6.700	71.874
	Without COVID	ELM	5	0.123	0.268	67.011	3	101.629	8.641	35.393	3	54.781	6.227	42.818
		ESN	20	0.392	0.543	239.621	3	294.932	14.778	73.650	3	88.592	7.841	62.774
		MLP	25	0.117	0.240	48.976	10	100.519	7.814	30.358	80	18.872	3.395	22.804
		RBF	3	0.203	0.419	174.831	90	232.206	13.805	67.810	90	61.697	6.702	71.883
	Without COVID and Lockdown	ELM	3	0.106	0.281	47.589	3	280.543	13.438	54.251	3	42.747	5.443	38.441
		ESN	15	0.390	0.550	239.796	3	325.950	14.608	66.276	5	120.188	8.986	96.500
		MLP	12	0.101	0.269	53.239	80	79.421	8.156	33.803	12	26.396	4.350	28.352
		RBF	3	0.199	0.416	171.079	90	232.429	13.810	67.820	90	61.713	6.703	71.886

		NO ₂				NO				O ₃				
		NN	MSE	MAE	MAPE	NN	MSE	MAE	MAPE	NN	MSE	MAE	MAPE	
Without Z-Score	All Inputs	ELM	5	785.090	20.353	30.175	10	4839.146	50.450	153.333	7	467.375	15.322	34.491
		ESN	5	907.308	24.263	40.671	10	4772.463	58.378	373.840	3	353.385	15.163	32.252
		MLP	70	557.778	17.589	23.740	50	2657.192	38.072	81.015	7	138.893	9.422	16.985
		RBF	5	854.313	25.362	51.069	5	5416.426	68.009	486.657	17	454.931	17.020	36.846
	Without COVID	ELM	3	445.282	17.293	23.027	7	4340.598	57.558	348.301	5	63.055	5.997	12.454
		ESN	3	822.075	25.260	52.416	3	5996.855	71.571	445.188	3	260.122	12.669	26.255
		MLP	35	464.065	17.072	21.113	8	3924.373	45.396	86.404	45	120.824	9.143	16.909
		RBF	5	854.784	25.517	51.092	5	5461.894	68.337	488.802	3	391.845	15.011	33.241
	Without COVID and Lockdown	ELM	7	1064.275	27.726	57.589	3	3991.357	54.068	113.379	3	137.191	9.229	17.298
		ESN	10	1555.614	31.140	75.076	15	8592.139	80.821	675.729	40	1214.000	29.877	59.143
		MLP	40	295.496	14.269	22.207	12	4320.338	48.815	91.635	3	206.566	12.420	21.391
		RBF	90	868.487	25.612	51.959	90	5472.401	68.400	490.933	3	358.350	12.872	29.566
With Z-Score	All Inputs	ELM	3	663.132	16.812	19.980	7	4307.805	50.687	159.321	3	115.769	9.348	18.562
		ESN	3	1070.911	26.303	35.021	30	5849.480	67.123	420.655	3	489.524	18.494	37.389
		MLP	60	493.319	18.015	23.994	35	3274.156	46.066	84.661	40	86.199	7.354	13.290
		RBF	7	864.579	25.509	51.690	3	5402.146	67.809	477.883	20	456.534	17.058	36.916
	Without COVID	ELM	5	394.529	16.522	29.060	5	3768.178	54.428	120.683	3	43.501	5.673	10.074
		ESN	3	902.434	26.099	51.690	3	5501.005	68.838	471.550	3	260.471	11.801	25.214
		MLP	20	228.207	12.518	20.777	7	3911.060	47.170	76.890	5	89.321	7.703	14.095
		RBF	12	859.824	25.425	52.147	3	5401.730	67.974	478.234	3	391.500	14.998	33.216
	Without COVID and Lockdown	ELM	3	782.711	24.452	39.246	3	5041.661	61.826	202.719	3	107.910	7.838	16.488
		ESN	40	1871.582	36.537	83.638	12	8067.633	80.195	594.801	3	389.735	14.920	28.355
		MLP	55	346.735	16.076	26.940	7	4490.207	51.747	87.543	3	161.224	10.402	19.405
		RBF	90	868.484	25.612	51.959	40	5471.100	68.391	490.941	5	369.729	13.681	31.046

633 NN: Number of neurons; MSE: Mean Square Error; MAE: Mean Absolute Error; MAPE: Mean Absolute Percentage Error; *With COVID means including the
634 variables number of COVID-19 new cases and the partial lockdown.

635 **Table A2.** Computational results for D. Pedro II Station



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Figure A3. Meteorological variables raw data for the test set.