| 1 | Stocking strategies for coldwater fish populations under temperature stress | |
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29 Summary

Coldwater fish stocks are increasingly affected by steadily increasing water temperatures.
 The question arises whether stock management can be adapted to mitigate the conse quences of this climatic change.

Here, we estimate the effects of increasing water temperatures and different stocking strat-2. 33 egies on fisheries yield by recreational anglers. Using a process-based population model 34 based on an empirical long-term data set for the whitefish population (Coregonus lavaretus 35 (L.) species complex) of Lake Irrsee, Austria, we project density-dependent and tempera-36 ture-dependent population growth and compare established stock enhancement strategies 37 to alternative stocking strategies under the aspect of increasing habitat temperatures and 38 cost neutrality. Additionally, we contrast the results obtained from the process-based model 39 40 to the results from simple regression models and argue that the latter show qualitative inadequacies in projecting catch with rising temperatures. 41

42 3. Our results indicate that increasing habitat temperatures reduce population biomass and43 catch by the fishery through their effect on growth and survival.

44 4. Regarding stocking strategies, we find that stocking mostly small fish produces higher pop-45 ulation biomass than stocking mostly larger fish, while catch remains almost constant.

5. *Synthesis and applications*. Stocking larger fish is more beneficial for the angling fishery
under the aspect of increasing temperatures. Adaption to climate change by changing stocking strategies cannot, however, prevent an overall reduction in catch and population size of
coldwater fish.

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Keywords: Alpine lake, angling, density dependence, growth probability, matrix model, natural
mortality, temperature dependence

53 Introduction

Compared to lakes in lowland areas, lakes in Alpine areas are typically characterized by great 54 depth and low water temperatures (Dokulil et al. 2010). Mean temperatures of surface and deep-55 water layers in Alpine lakes of Central Europe have, however, increased between 0.5°C and 56 1°C over the last 40 years and further warming is expected because of ongoing climatic changes 57 (Dokulil et al. 2006; IPCC 2007; Dokulil et al. 2010; Dokulil 2014). This change in the thermal 58 regime is very likely to affect population dynamics of fish species that are living in Alpine lake 59 ecosystems, and consequently also the related fishery could be affected (Ficke, Myrick & Han-60 sen 2007; Jeppesen et al. 2012). 61

Whitefish (Salmoniformes: *Coregonus* spp.) are typical coldwater fish that grow optimally at low water temperatures (Casselman et al. 2002; Siikavuopio et al. 2013). They are very important for freshwater fisheries in northern temperate regions (Berka 1990; Petr 1999; Ebener et al. 2008; Jeppesen et al. 2012). The planktivorous European whitefish (*Coregonus lavaretus* (L. 1758) species complex) lives in the cold-water layers of Alpine lakes and was exploited mainly by commercial fisheries before the 1970s. With improving angling techniques over the last decades, whitefish has become also very important for recreational fisheries.

To compensate for harvesting by fisheries, managers of exploited whitefish populations commonly conduct stocking programs. In general, stocking strategies comprise introductions of small (e.g., larvae) and large (e.g., one-summer-old) fish in various proportions. Stocking small fish is common, although many authors argue that stocking larger fish is more profitable for whitefish fisheries compared to stocking smaller fish (Salojärvi & Huusko 1990; Wanzenböck & Jagsch 1998; Lasenby, Kerr & Hooper 2001; Gerdeaux 2004).

Stocking strategies are almost never systematically evaluated in small fisheries (Arlinghaus, Mehner & Cowx 2002; Cowx & Gerdeaux 2004). Fisheries managers often do not pay enough attention to the cost-effectiveness of the applied stocking program and to possible negative impacts of stocking due to, e.g., density-dependent effects on growth and mortality (Salojärvi 1991; Arlinghaus, Mehner & Cowx 2002). Moreover, in the context of climate change, the
question arises how stocking strategies can be adapted to ensure sustainable fisheries management of coldwater fish under increasing habitat temperatures.

In general, fish are poikilothermic animals and live in specific temperature ranges, preferring water temperatures that promote optimal growth (Jobling 1981; Ohlberger et al. 2008; Mehner et al. 2010). Growth in turn is related to natural mortality (Pauly 1980; Jensen 1996; Lorenzen 1996). Fishery yield depends on how well the fish grow and survive. Therefore, a correlation between water temperatures and catches often exist (Sutcliffe Jr. Drinkwater & Muir 1977; Scarnecchia 1984; Sakuramoto, Hasegawa & Suzuki 2005; Biswas et al. 2009).

Mathematical models are very helpful to estimate how increasing temperatures and various stocking strategies will affect population dynamics and the related catch by the fishery. Simple regression models, fitted to observed water temperatures and catches, can be used to extrapolate catches under higher temperatures. This model approach, however, does not account for the relevant life-history processes and the resulting population dynamics.

In contrast, a process-based model approach provides additional opportunities for analyzing
population dynamics and can readily be extended to account for relevant mechanisms, such as
fishing, stocking, and density dependence. Models based on life-history processes are differential equations, matrix models (MMs), and individual-based models (IBMs).

Differential equations can be analytically solved for unstructured populations, while only 97 numerical solutions are feasible (and effectively become matrix models) for structured popula-98 99 tions. In contrast, IBMs provide great flexibility and detailed insights into population dynamics, primarily because they explicitly account for individual variation (Grimm 1999; DeAngelis & 100 101 Mooij 2005). Although IBMs and MMs often produce similar results, particularly when the MMs account for aspects of variation, IBMs require substantially higher computational effort 102 (Pfister & Stevens 2003; Sable & Rose 2008). Therefore, matrix models provide a good com-103 promise and allow studying structured populations with reasonable computational effort. 104

Conventional matrix models used for studying fish populations, also known as Leslie matrix
models (Leslie 1945; Caswell 2001), consider only age classes. Although age is a natural demographic property in whitefish life history, vital parameters and management interventions
often depend on body size (Lorenzen & Enberg 2002; DeRoos, Persson & Cauley 2003; Lewin,
Arlinghaus & Mehner 2006; Ficker et al. 2014). A length-based model may therefore be more
suitable for whitefish populations.

111 Here, we use a length-structured matrix model with temperature dependence and density dependence in growth and mortality to evaluate the effects of increasing habitat temperatures on 112 the total biomass and catch by recreational anglers of a European whitefish population. A long-113 114 term (10 years) dataset of experimental gillnet catches was used to derive model parameters for 115 the whitefish population of Lake Irrsee (Gassner, Hassan & Wanzenböck 2004; Gassner & Wanzenböck 2007). We further compare our modeling results to projections by simple regres-116 sion models describing the correlation between catch and habitat temperature. We additionally 117 assess the cost-effectiveness of the applied stocking strategy on the Lake Irrsee population and 118 compare it to various other strategies with consideration of the fraction of invested money on 119 small (i.e., 1 cm total length) and large (i.e., 10 cm total length) fish under constant and under 120 continuously increasing temperature scenarios. Finally, we offer policy recommendations for 121 stocking strategies of European whitefish under the aspect of climate change. 122

124 Material and Methods

We develop a process-based model to project the whitefish population of Lake Irrsee under different stocking and temperature scenarios. The resulting length-structured matrix model augmented with stochastic elements includes all relevant processes for population dynamics of whitefish, which are: temperature-dependent and density-dependent growth, survival, and reproduction.

Stocking strategies and catch by anglers are incorporated into the model through vectors of stocked and caught whitefish, respectively. Assuming different temperature scenarios, we project annual biomass and catches over a period of 50 years with different stocking strategies. Below, we briefly discuss selected points specifically. Details can be found in the supplementary material.

135

136 *Sampling data*

The pre-alpine Lake Irrsee, Austria (N 47° 53′, E 13° 18′) is classified as an oligo-mesotrophic lake with a holomictic-dimictic mixing regime. Its maximum depth is 32 m and its surface area stretches over 3.6 km². European whitefish is the dominant fish species in Lake Irrsee and important for the local recreational fishery.

Since the year 2000, the whitefish population of Lake Irrsee is studied by means of gillnetting 141 carried out annually in October (pre-spawning census; Gassner, Hassan & Wanzenböck 2004; 142 Gassner & Wanzenböck 2007). The overall catch amounted to 2,013 individual whitefish be-143 144 tween years 2000 and 2009. Gillnet fleets with different randomized mesh sizes between 15 mm and 70 mm were assembled and set over night in part of the lake in 12 to 15m depth. 145 146 Individual length (\pm 0.5 cm), weight (\pm 5 g), age, sex and ripeness of gonads were determined for all caught whitefish. Age identification was achieved by scale reading according to 147 the method used by DeVries & Frie (1996) and Gassner, Hassan & Wanzenböck (2004). 148

The examination of sex and ripeness stages according to Nikolsky (in: Ricker 1970) was done after dissection by classifying individuals into male, female, or juvenile and as spawners or non-spawners. Fresh eggs of mature female individuals were counted per unit weight in the year 2010 according to the gravimetric sub-sampling method described by Bagenal (1978).

Total fish biomass in Lake Irrsee was estimated through simultaneously performed hydroacoustic surveys in the open water area with two split-beam echo sounders in the year 2000 (Wanzenböck et al. 2003). The population biomass of European whitefish was assumed to account for 60% of the total observed biomass.

Temperatures and oxygen concentration were available from water samples collected in 0, 2, 5, 8, 10, 12, 15, 20, 25, and 30m depth at the deepest site of the lake on a monthly basis. Temperatures were measured in the field with a mercury thermometer and oxygen concentrations were determined in the laboratory according to the Winkler procedure (Winkler 1889). Annual mean growth temperatures for European whitefish during the growth period from May to October were derived from temperature measurements in the suitable oxythermal habitat for coldwater fish (i.e., $O_2 > 3mgl^{-1}$ and T < 21.2 °C; Stefan et al. 1995)

164

165 Spawning, eggs, and larvae

European whitefish reproduce in early winter and spawned eggs develop over the winter months 166 till larvae hatch in spring (Fuller, Scott & Fraser 1976; Wahl & Löffler 2009). We calculated 167 the biomass of female spawners using the observed sex ratio, a sigmoid maturity function 168 (Ficker et al. 2014), and an allometric length-weight relationship. The average fecundity, that 169 is, the average number of eggs per unit weight female fish, is estimated from our data and 170 modeled as a stochastic variable. Finally, the number of hatching larvae, and thus the success 171 of natural reproduction, is obtained from the effective fecundity, which is defined as the number 172 of produced offspring that survives till hatching from the egg. 173

Survival is usually much lower for early development stages compared to larger fish, like in eggs and freshly hatched larvae (Salojärvi 1982; Fuiman & Werner 2002). We assume egg mortality over the developmental period and larval mortality over the first four weeks of life to be much higher compared to mortality rates of larger whitefish (see supplementary material).

178

179 Density-dependent and temperature-dependent growth

180 Growth of a fish is depends primarily on size and is also affected by population density and environmental temperature. Small fish grow almost linearly and large fish grow according to a 181 von Bertalanffy model toward an asymptotic length (Quince et al. 2008). The asymptotic length 182 183 depends on total biomass and therefore on population density via a Maynard Smith-Slatkintype functional response (Smith & Slatkin 1973; Beverton & Holt 1993; Lorenzen & Enberg 184 2002; Ylikarjula et al. 2002), while the von Bertalanffy growth coefficient depends on environ-185 186 mental temperature (Ricker 1979; Fontoura & Agostinho 1996; Jensen 1996; see supplementary material for details). Asymptotic length and growth coefficient are related (Pauly 1980; Jensen 187 1996), which makes the asymptotic length also indirectly dependent on temperature. We as-188 sume a lognormal distribution of monthly growth increments and allow growth to vary among 189 individuals of the same length. 190

191

192 Natural and fishing mortality

Natural mortality of a fish is related to growth and environmental temperature (Pauly 1980; Quinn & Deriso 1999; Kenchington 2013) and therefore indirectly depends on population density. We estimated natural mortality through two different methods (Pauly 1980; Jensen 1996; see supplementary material) from density-dependent and temperature-dependent growth parameters. Additionally, we consider fishing mortality. Fisheries impose certain size limits which leads to selective removal of fish ofcertain lengths. We model this size-selective removal as a stochastic process. We assume a constant angling effort per unit time, which implies that 200 the total catch is limited, and that total catch drops faster than linearly as abundance in the 201 catchable size range decreases towards 0. We used catch statistics of the local angler association 202 for parameterization of stochastic fish removal by anglers.

203

204 Stocking strategies

Currently, fisheries stock small whitefish (around 630,000 individuals of ~1 cm length with an individual price of $\notin 0.014$) in March and larger whitefish (around 6,000 individuals of ~10 cm length with an individual price of $\notin 0.30$) in September. This means that about 83% of the money invested into stocking is used for stocking small fish and the remainder for stocking large fish. To compare the cost-effectiveness, we investigate stocking strategies that allocate the same total amount of money in different ratio (thus, a stocking ration of 0.1 means 10% of the money is invested into stocking small fish etc.).

212

213 *Temperature scenarios*

We consider three different temperature scenarios (i.e., constant temperature, $\pm 1^{\circ}$ C, and $\pm 2^{\circ}$ C over 50 years) The two scenarios with increasing temperatures are based on the observed temperature increase in surface waters of Lake Irrsee over the last decades (i.e., annual average with $\pm 0.9^{\circ}$ C and average of spring and summer temperatures with $\pm 1.9^{\circ}$ C; Dokulil et al. 2010) and we also consider deep water warming and projected future temperature development of Austrian lakes described in Dokulil et al. (2006) and Dokulil (2014).

220

222 **Results**

We projected population biomass and anglers catch under changing annual habitat temperatures, investigating three basic temperature scenarios. We compared the predictions from simple regression models to our process-based model; we investigated the effects of increasing temperatures on biomass and catch; we analyzed the mechanism underlying the temperature effect; and finally assessed stocking strategies comprising introductions of small and large whitefish in different ratios.

229

230 Process-based model vs. regression models

Projections with the process-based model are shown for two different estimates of natural mor-231 tality (Pauly 1980; Jensen 1996), both resulting in qualitatively very similar predictions. We 232 233 project annual catches (with a three year delay) as a function of growth temperature with our process-based model and extrapolate catches with simple regression models fitted to observa-234 tions. The quadratic regression model agrees with the process-based model in that both project 235 236 saturating catch at low growth temperatures. The exponential regression model agrees with the 237 process-based model in that both project decreasing catches with increasing growth temperatures showing a non-linear pattern (although projected catches differ substantially). Quadratic 238 239 and linear regression models project a complete collapse in catches for a relatively modest increase in growth temperatures similar to the collapse projected by the process-based model. In 240 241 contrast, the linear and the exponential regression model also project high catch without saturation for low growth temperatures. No regression model shows qualitative agreement with the 242 process-based model over the whole range of growth temperatures considered (Fig. 1). 243

244

245 *Temperature effects*

Using our process-based model we project changes in population biomass and catch by anglersover a period of 50 years under three temperature scenarios (Fig. 2.a). We find that population

biomass and catch by anglers decrease with increasing temperatures. The effect is stronger 248 249 when the temperature increase is larger. Our projections with Jensen's estimate of natural morality show that increasing habitat temperature reduce biomass by about 2.6% (i.e., 250 -0.9 kg ha^{-1}) and by about 4.4% (i.e., -1.6 kg ha^{-1}), respectively (Fig. 2.b), while catch de-251 creases by about 24% (i.e., -1.2 kg ha^{-1}) and 45% (i.e., -2.3 kg ha^{-1}), respectively (Fig. 252 2.c). Our projections with Pauly's estimate show that increasing habitat temperatures reduce 253 biomass by about 4.3% (i.e., -1.7 kg ha^{-1}) and by about 7.9% (i.e., -3.1 kg ha^{-1}), respec-254 tively, and that catch decreases by about 26% (i.e., -1.4 kg ha^{-1}) and 48% (i.e., 255 -2.6 kg ha^{-1}), respectively (not shown). 256

257

258 Underlying mechanism

Temperature has direct and indirect effects in our process-based model. The growth coefficient 259 260 depends directly on temperature (Fig. 3.a) via a simple relation (see material and methods section and supplementary material). Since population dynamics in the model depends on growth, 261 also the density-dependent parameters asymptotic length and survival probability are indirectly 262 dependent on temperature. Increasing temperature increases the growth coefficient (Fig. 3.a) 263 and decreases asymptotic length (Fig. 3.b) and annual survival (Fig. 3.c). Our projections show 264 265 that increasing habitat temperature increase the growth coefficient by about 6.7% (i.e., $+0.02 \text{ y}^{-1}$) and 12.4% (i.e., $+0.02 \text{ y}^{-1}$), respectively, while asymptotic length decreases by 266 about 2.9% (i.e., -1.3 cm) and 5.2% (i.e., -2.3 cm), respectively, and natural annual survival 267 decreases by about 3.7% (i.e., -0.02%) and 6.7% (i.e., -0.04%), respectively. Our projections 268 using Pauly's estimate show that increasing habitat temperature increase the growth coefficient 269 by about 6.7% (i.e., $+0.02 \text{ y}^{-1}$) and 12.4% (i.e., $+0.05 \text{ y}^{-1}$), respectively, while asymptotic 270 length decreases by about 2.7% (i.e., -1.2 cm) and 4.8% (i.e., -2.1 cm), respectively, and 271 natural annual survival decreases by about 4.6% (i.e., -0.03%) and 8.6% (i.e., -0.05%), re-272 spectively. 273

274

275 Stocking strategies

Stocking strategies, in our case, are expressed by the ratio of money invested into stocking small 276 277 fish to the total amount of money invested for stocking. This includes the extreme cases where the money is invested either only into stocking small fish (corresponding to a stocking ration 278 of 1) or only into stocking large fish (corresponding to a stocking ratio of 0). To assess the cost-279 effectiveness of stocking strategies for constant temperatures, we project population biomass 280 and catch by anglers for different stocking ratios with a fixed investment budget. Different 281 stocking ratios result in very different numbers of introduced fish, because large fish are sub-282 stantially more expensive than small fish (e.g., in Lake Irrsee10 cm fish cost 21.4 times more 283 than 1 cm fish). Our projections reveal that increasing the current stocking ratio of 0.83 in-284 creases population biomass after 10 years, and decreasing the current stocking ratio decreases 285 biomass, while the catch remains nearly the same with a very inconspicuous peak at a stocking 286 ratio of about 0.6 (Fig. 4). 287

288

289 *Mitigation of climate change*

To evaluate how stocking strategies can be adapted to mitigate the effects of climate change, we project population biomass and catch by anglers over a period of 10 and 25 years for increasing habitat temperatures (+2°C over 50 years; Scenario 3 in Fig. 2 and 3) and different stocking ratios. Compared to the projection with constant temperature (Fig. 4), population biomass and catch by anglers is generally lower. The catch, however, is now clearly maximized at lower stocking ratios of about 0.3 (Fig. 5).

296 **Discussion**

Whitefish stocks in cold Alpine lake ecosystems are affected through increasing temperatures 297 due to climatic changes. Fisheries management of coldwater fishes commonly uses stocking to 298 maintain available catches for recreational and commercial fisheries. To evaluate the often un-299 known effects of stocking on population dynamics as well on the fishery itself, we have devel-300 oped a process-based model of density-dependent and temperature-dependent population 301 302 growth. Density dependence has been introduced in the growth parameter asymptotic length: higher population densities reduce asymptotic length (Jensen 1997). Additionally, the effect of 303 temperature has been integrated into the growth coefficient: higher temperatures lead to higher 304 growth coefficients (depending on the temperature optimum for coldwater fish; Jobling 1981; 305 Stefan et al. 1995; Casselman 2002). 306

Natural mortality of whitefish has been derived from growth parameters and temperatures through two different methods (Pauly 1980; Jensen 1996). Both are considered to produce useful estimates when the growth coefficient can be derived accurately from population data and when adult life span is not exceptionally long (Kenchington 2013). We found that the simpler method proposed by Jensen (Jensen 1996), generally leads to higher estimates of natural mortality than the regression based model of Pauly (1980). Still, both methods produce qualitatively and quantitatively similar results in our model projections.

The parameterization of the process-based model is based on an empirical long-term data set 314 of Lake Irrsee collected by annual gillnet samples and catch statistics. We have estimated initial 315 316 biomass, growth parameters, fecundity, maturity and sex ratio directly from the data. Because of the importance of predation mortality in early life stages, we have modeled early life-stage 317 318 mortality separately as a density-independent process. Nevertheless, reproduction is temperature- and density-dependent because of the relationship between adult size and reproduction 319 320 efficiency (i.e., size-dependent maturation and size-dependent egg production). The optimal 321 temperature range for whitefish growth, as well as egg and larval mortality, which were not available from field sampling, have been taken from literature. The sensitivity of our model to
egg and larval mortality is high, which is in accordance to theoretical expectations that early
life stages have a strong influence on population growth and consequently on recruitment to the
fishery (Ricker 1975; Chambers & Trippel 1997; Fuiman & Werner 2002).

The assumed optimal growth temperature range (i.e., $T_{min} = 2$ °C, $T_{max} = 22$ °C) had also a 326 great effect on the quantity of projected catches, whereas the decreasing trend with increasing 327 temperature was robust. The minimal temperature for growth that we used in our model was 328 very precisely evaluated by Siikavuopio et al. (2010) who showed that whitefish grows at 3°C 329 330 but not at 1°C water temperature. In contrast, the maximum temperature for growth is characterized only vaguely in literature and ranges from 13.5 °C to 22 °C (Jobling 1981; EIFAC 1994; 331 Casselman 2002; Siikavuopio, Knudsen & Amundsen 2010; Szczepkowski, Szczepkowska & 332 Krzywosz 2006) and it is also very likely that the temperature-dependence in growth is species-333 specific as proposed by Ohlberger et al (2012). Consequently, the temperature at which a col-334 lapse of an actual fishery occurs may be different from the 13°C at which it was observed in 335 our model projections. To refine the prediction, the maximum temperature for growth needs to 336 337 be assessed more accurately.

The strength of our model is the consideration of important life-history processes with respect to body size. Although simple statistical models showed similar trends of catches under a changing climate, the underlying mechanisms in population dynamics remain unclear, and consequently a process-based model is advantageous.

Our results clearly demonstrate that lower catches must be expected in cold-water fisheries with continuously increasing temperatures in the future. Additionally, the process-based model reveals that lower catches are mainly due to accelerated growth of juveniles resulting in smaller sizes of adults and consequently lower recruitment into the established size-limit of the recreational fishery. We further found that population biomass decreases as a consequence of higher natural mortality. Modeling results for different stocking strategies indicate that this trend could

be partly mitigated through stocking higher ratios of small fish. While changing stocking strat-348 egies cannot prevent a reduction in catch with increasing temperatures, stocking larger white-349 fish nevertheless seem to be more advantageous for the recreational angling fishery, insofar as 350 it maximizes catch under the circumstances and thus angler satisfaction. 351

352

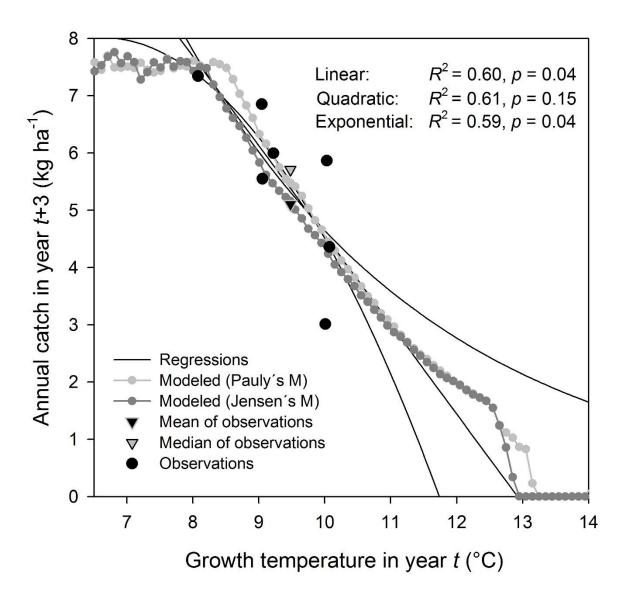
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361 **Figures**

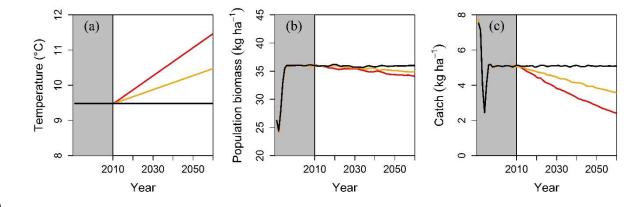
362 Figure 1:

Catch predictions of our process-based model compared to simple regression models. Black 363 solid lines show predictions of three regression models (linear, quadratic, and exponential) fit-364 ted to observational data of growth temperature and anglers catch, with a time lag of three years 365 (black points; see text). Grey points and interpolation lines show predictions of our process-366 367 based models using two different mortality estimation procedures. All models capture the decrease of anglers catch with increasing temperatures. They differ in whether they allow a satu-368 ration of the catch towards low temperatures, and in whether they allow a collapse towards high 369 370 temperatures and in how this collapse is approached.



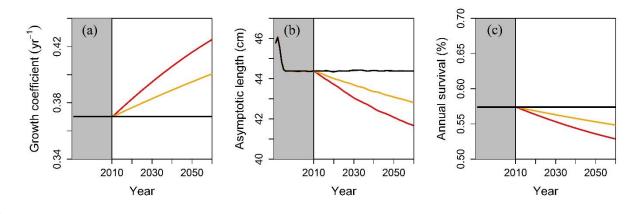
372 Figure 2:

Increasing growth temperatures decrease population biomass and catch. Projections for three different temperature scenarios (a): constant temperature (black line), +1°C increase over 50 years (orange line) and +2°C increase over 50 years (red line). Population biomass of whitefish decreases only slightly with increasing temperature (b), while catch by recreational angling decreases substantially with increasing temperature (c). Grey shading indicates the initial stabilization period (see text).



380 Figure 3:

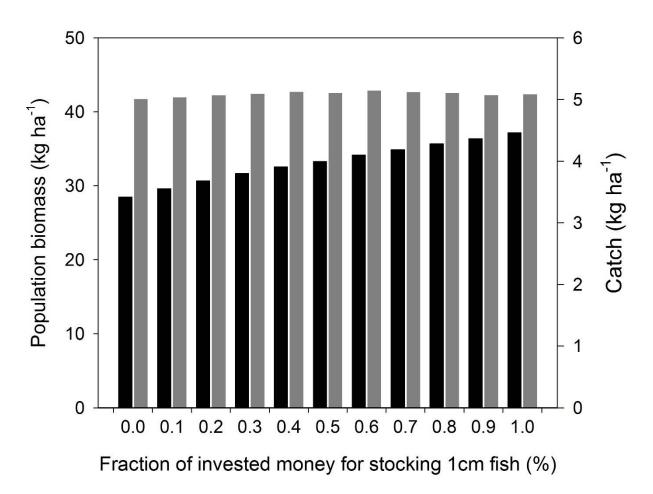
Higher temperatures affect growth and survival. Increasing temperatures (a) increase growth
coefficients, (b) decrease asymptotic lengths and (c) consequently also reduce annual survival.
Colors as in Fig.2.



384

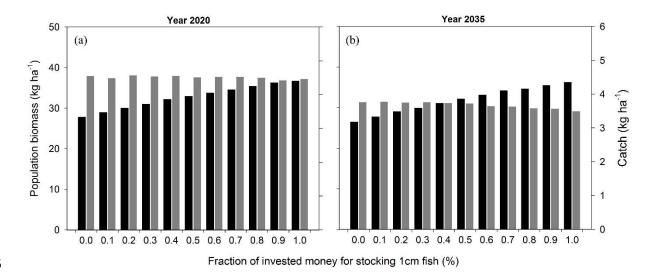
385 Figure 4:

Stocking ratio affects population biomass more strongly than catch. For constant temperatures,
solid bars show projected population biomass (black) and catch by anglers (grey) ten years after
changing the stocking ratio (i.e., fraction of money invested in small fish) from the current
stocking ratio in Lake Irrsee of 0.83.



391 Figure 5:

With increasing temperatures catch is maximized at lower stocking ratios. For increasing temperatures (+2°C over 50 years; scenario 3 in figure 2 and 3), panels show projections of population biomass and catch by anglers after (a) 10 years and (b) 25 years after changing the stocking ratio from the current stocking ratio (see Fig.4).



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