CORE


#### Abstract

In trawl surveys a cluster of fish are caught at each station, and fish caught together tend to have more similar characteristics, such as length, age, stomach contents etc., than those in the entire population. When this is the case, the effective sample size for estimates of the frequency distribution of a population characteristic can, therefore, be much smaller than the number of fish sampled during a survey. As examples, it is shown that the effective sample size for estimates of length-frequency distributions generated by trawl surveys conducted in the Barents Sea, off Namibia, and off South Africa is on average approximately one fish per tow. Thus many more fish than necessary are measured at each station (location). One way to increase the effective sample size for these surveys and, hence, increase the precision of the length-frequency estimates, is to reduce tow duration and use the time saved to collect samples at more stations.


# Assessing the precision of frequency distributions estimated from trawl-survey samples 

Michael Pennington

Institute of $M$ arine Research
Department of $M$ arine Resources
Nordnesgaten 33
N-5005 Bergen, Norway
E-mail address:michael@imr.no

## Liza-Maré Burmeister

M inistry of Fisheries and $M$ arine Resources of $N$ amibia, $N$ atM IRC PO Box 912
Swakopmund, Namibia

## Vidar Hjellvik

Institute of M arine Research
Department of $M$ arine Resources
N ordnesgaten 33
N-5005 Bergen, N orway

Survey-based assessments often appear to provide a more accurate prognosis of the status of a fish stock than catchbased assessments (Nakken, 1998; Pennington and Strømme, 1998; Korsbrekke et al., 2001). An advantage that survey-based assessments have over those based on commercial catch statistics is that the uncertainties associated with survey estimates can be studied and quantified, and based on such research, survey methods, and ultimately stock assessments, can be improved (Godø, 1994). In contrast, it is generally difficult to determine either the accuracy or the precision of estimates based on commercial catch data, and it is not clear how to improve, at a reasonable cost, the collection of catch data so that these data would more accurately reflect the mortality caused by fishing (Christensen, 1996).

Trawl surveys provide estimates of the abundance or relative abundance of a fish stock and estimates of the relative frequency of various population characteristics, such as length, age, and stomach contents. In our study we examined the precision of survey-based estimates of the length-frequency distributions of cod and haddock in the Barents Sea, hake off South Africa, and hake off

Namibia. The focus was on length, but the results are relevant for estimating the frequency distribution of other population characteristics.

## Materials and methods

## Survey length data

Bottom trawl survey length data for Northeast Artic cod (Gadus mohua) and Northeast Arctic haddock (Melanogrammus aeglefinus) ${ }^{1}$ were collected duringthel nstitute of M arineResearch (Norway) winter and summer surveys in the Barents Sea. The surveys were stratified systematic surveys and at each station the trawl was towed for 30 minutes. ${ }^{2}$

[^0]The data for the Namibian deepwater hake (Meluccius paradoxus) were collected during bottom trawl surveys off Namibia conducted by the Ministry of Fisheries and Marine Resources of Namibia in conjunction with the Norwegian Agency for Foreign Aid (NORAD). For these surveys, tows of 30-minute duration were made at stations along transects perpendicular to the coast. ${ }^{3}$

The data for the deepwater hake for South Africa, were collected from during bottom trawl surveys off the west coast of South Africa. The surveys were conducted by the M arine and Coastal M anagement Centre, South Africa, by using a stratified random design. Tows of 30-minute duration were made at each station (see Payne et al., 1985).

## Assessing the precision of length-frequency estimates

The sample of fish of a particular species measured during a survey is not a random sample of individual fish from theentire population but a sample of $n$ clusters, one cluster from each station. Becausefish caught together are usually more similar than those in the general population, a total of $M$ fish collected in $n$ clusters will contain less information about the population length distribution than $M$ fish sampled randomly. One way to measure the information contained in a sample of length measurements is to estimate the number of fish that one would need to sample at random (the effective sample size) to obtain the same information on length contained in the cluster samples.

The effective sample size for cluster sampling can be defined and calculated as follows (Pennington and Vølstad, 1994; Folmer and Pennington, 2000). First estimate the population mean fish length and its variance based on the clusters of fish caught at n stations. Because both the lengths and the number of fish at a station are random variables, a ratio estimator is appropriate (Cochran, 1977). The ratio estimator, R, of the mean length is given by

$$
\begin{equation*}
\hat{R}=\frac{\sum_{i=1}^{n} M_{i} \hat{\mu}_{i}}{\sum_{i=1}^{n} M_{i}} \tag{1}
\end{equation*}
$$

where $M_{i}$ = the number of fish caught (either actual or estimated) at station $i$; and
$\hat{\mu}_{\mathrm{i}}=$ an estimate of the average length of fish at station i.

For example, if the catch at a station is divided into strata and a random sample of fish are chosen in each stratum, then for that station $\hat{\mu}_{\mathrm{i}}$ would be the stratified estimate of mean length. The estimated variance of $R$ is approximately given by

[^1]\[

$$
\begin{equation*}
\operatorname{var}(\hat{R})=\sum_{i=1}^{n} \frac{\left(M_{i} / \bar{M}\right)^{2}\left(\hat{\mu}_{i}-\hat{R}\right)^{2}}{n(n-1)} \tag{2}
\end{equation*}
$$

\]

where $\bar{M}=\sum_{i=1}^{n} M_{i} / n$.
Next estimate the variance, $\sigma_{x}^{2}$, of the population length distribution. If $m_{i}$ fish are randomly selected at each station (or if all fish are measured), then

$$
\begin{equation*}
\hat{\sigma}_{x}^{2}=\frac{\sum_{i=1}^{n} \sum_{j=1}^{m_{1}}\left(M_{i} / m_{i}\right)\left(x_{i, j}-\hat{R}\right)^{2}}{M-1} \tag{3}
\end{equation*}
$$

is an estimator of $\sigma_{x}^{2}$,
where $M=\sum_{i=1}^{n} M_{i}$ is the total number of fish caught during ${ }_{i=1}$ the survey; and
$x_{i, j}=$ the length of the $j^{\text {th }}$ fish at station $i$.
For other sampling schemes at a station, first estimate the number of fish caught during the survey in each of $L$ length bins, then (e.g. Bhattacharyya and J ohnson, 1977)

$$
\begin{equation*}
\hat{\sigma}_{x}^{2}=\frac{\sum_{k=1}^{L} f_{k}\left(y_{k}^{\prime}-\hat{R}\right)^{2}}{M-1} \tag{4}
\end{equation*}
$$

is an estimator of $\sigma_{x^{\prime}}^{2}$
where $f_{k}=$ the frequency of fish in the $k^{\text {th }}$ length bin; and $y_{k}^{\prime}=$ the bin's midpoint.

Now if it were possible to sample $m$ fish at random from the population, then the variance of the sample mean would be equal to $\sigma_{x}^{2} / \mathrm{m}$. The effective sample, $\mathrm{m}_{\text {eff }}$, is defined as the number of fish that would need to be sampled at random so that the sample mean would have the same precision as an estimate based on a sample of $n$ dusters. An estimate of the effective sample size for a particular duster sample can be derived by substituting the estimates from Equation 2 and either Equation 3 or 4 into the equation

$$
\begin{equation*}
\frac{\hat{\sigma}_{x}^{2}}{\hat{m}_{\mathrm{eff}}}=\operatorname{var}(\hat{\mathrm{R}}) \tag{5}
\end{equation*}
$$

The effective sample size is related to Kish's design effect (deff), which for estimating the mean from cluster sampling is defined by (Cochran, 1977)

$$
\operatorname{deff}=\frac{\operatorname{var}(\hat{R})}{\sigma_{x}^{2} / m}
$$

Table 1
Summary statistics for assessing the precision of the estimated length distributions of Northeast Arctic cod based on the winter and summer bottom trawl surveys in the Barents Sea. The estimated effective sample size is denoted by $\hat{m}_{\text {eff }}, \mathrm{n}$ is the number of stations at which cod were caught, $M$ is the total number of cod caught, $m$ is the number measured, $R$ is the estimate of mean length, and $\operatorname{var}(\hat{R})$ is its variance.

| Year | n | M | m | $\hat{\mathrm{R}}(\mathrm{cm})$ | $\operatorname{var}(\hat{\mathrm{R}})$ | $\hat{\mathrm{m}}_{\text {eff }}$ | $\hat{\mathrm{m}}_{\text {eff }} / \mathrm{n}$ | $\left(\hat{\mathrm{m}}_{\text {eff }} / \mathrm{m}\right) \times 100 \%$ |
| :--- | :---: | ---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Winter |  |  |  |  |  |  |  |  |
| 1995 | 296 | 175006 | 47286 | 20.0 | 0.7 | 313 | 1.1 |  |
| 1996 | 314 | 209114 | 44021 | 18.0 | 0.3 | 511 | 1.6 | 0.7 |
| 1997 | 177 | 71418 | 25689 | 19.0 | 2.1 | 119 | 0.7 | 1.1 |
| 1998 | 197 | 60746 | 32536 | 22.1 | 0.7 | 394 | 2.0 | 0.7 |
| 1999 | 223 | 50192 | 21760 | 25.0 | 1.9 | 107 | 0.5 | 1.2 |
|  | Avg. | 113295 | 34258 |  |  | 289 | 1.2 | 0.5 |
| Summer |  |  |  |  |  |  |  | 0.8 |
| 1995 | 329 | 66643 | 46161 | 31.2 | 1.4 | 252 | 0.8 |  |
| 1996 | 341 | 115834 | 45286 | 24.4 | 0.6 | 478 | 1.4 | 0.6 |
| 1997 | 266 | 72093 | 26947 | 23.1 | 0.8 | 266 | 1.0 | 1.1 |
| 1998 | 218 | 72360 | 23461 | 25.1 | 1.1 | 184 | 0.8 | 1.0 |
| 1999 | 217 | 46593 | 23253 | 30.8 | 0.9 | 211 | 0.9 | 0.8 |
|  | Avg. | 74705 | 33022 |  |  | 278 | 1.0 | 0.9 |
|  |  |  |  |  |  |  | 0.9 |  |

and, therefore, $m_{\text {eff }}=m / d$ eff. Some statistical software packages for the analysis of complex survey data, such as SUDAAN (Research Triangle Institute, 2001), generate estimates of the design effect.

Simulation techniques were used to examine the effect that reducing the total number of fish measured during a particular survey would have on the estimates of the mean length and the effective sample size. Length measurements consist of one or more subsamples from the fish caught at each station. The simulated estimates of the distributions of $\hat{m}_{\text {eff }}$ and $R$ (given the actual fish measured during the survey) were generated by randomly selecting from every haul a maximum of $k$ fish without replacement from each subsample. If fewer than $k$ fish were in a subsample, then all were chosen. This was done 500 times for $k=10,30$, and 100, and each run produced values of $\hat{m}_{\text {eff }}$ and $R$.

To assess the precision of an estimated length distribution, bootstrapping (Efron, 1982) was used to generate $95 \%$ confidence intervals for the proportion of fish in each $5-\mathrm{cm}$ length bin. For each of 500 runs, $n$ stations (the number of tows made during the survey) were randomly sampled with replacement and the confidence interval for each $5-\mathrm{cm}$ length bin was based on the resulting 500 estimates of the proportion of fish in that bin. Finally, bootstrapping was used to examine how much the length of the $95 \%$ confidence intervals would increase if a maximum of 10 fish were selected from each subsample.

## Results

Estimates of the effective sample size and associated statistics for survey-based estimates of the length composi-
tion of cod in the Barents Sea are presented in Table 1. The results indicate that for cod the estimated effective sample size is small compared with the number of fish measured. For example during the 1995 winter survey, 175,006 cod were caught, 47,286 were measured, and the effective sample size was 313 fish or $0.7 \%$ of the total number measured (Table 1). The average effective sample size for the winter surveys was 1.2 cod per tow and for the summer surveys, 1.0 cod per tow. The estimated effective sample sizes for the N ortheast Arctic haddock survey data were, on average, approximately one fish per tow (Table 2).

The effective sample sizes for the survey estimates of the length distribution of deepwater hake off Namibia and off South Africa (Tables 3 and 4) fol lowed the same pattern as for cod and haddock in the Barents Sea. In particular, the average effective sample size was 0.5 hake per tow for the Namibian surveys and 1.3 hake per tow for the South African surveys.

The simulated distributions of $\hat{m}_{\text {eff }}$ and $\hat{R}$, which demonstrate the effects of reducing the total number of measured fish on estimates of mean length, for the 1995 and 1999 winter surveys of cod in the Barents Sea are shown in Figures 1 and 2. For example, if a maximum of 30 fish were selected from each subsample at each station, then a total of 11,123 fish would have been measured during the 1995 survey compared with 47,286 fish that were actually measured. For 1995, $R=19.96$ and the $95 \%$ confidence interval for $\hat{R}$ was (18.29, 21.63). As can be seen from Figure 1, all 500 simulated estimates of the mean based on the reduced sample size were well within the $95 \%$ confidence limits for R. When the number of fish measured was reduced to a maximum of 10 fish per subsample for a total sample of 2597 fish, the simulated estimates were also

Table 2
Summary statistics for assessing the precision of the estimated length distributions of Northeast Arctic haddock generated by the winter and summer bottom trawl surveys in the Barents Sea. The notation is the same as that shown in Table 1.

| Year | n | M | m | $\hat{\mathrm{R}}(\mathrm{cm})$ | $\operatorname{var}(\hat{\mathrm{R}})$ | $\hat{\mathrm{m}}_{\text {eff }}$ | $\hat{\mathrm{m}}_{\text {eff }} / \mathrm{n}$ | $\left(\hat{\mathrm{m}}_{\text {eff }} / \mathrm{m}\right) \times 100 \%$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Winter |  |  |  |  |  |  |  |  |
| 1995 | 199 | 66009 | 22938 | 25.0 | 1.0 | 168 | 0.8 |  |
| 1996 | 235 | 54892 | 25525 | 32.0 | 2.9 | 69 | 0.3 | 0.7 |
| 1997 | 140 | 37441 | 13273 | 22.0 | 0.8 | 185 | 0.8 | 0.3 |
| 1998 | 144 | 12704 | 9620 | 23.9 | 1.0 | 169 | 1.2 | 1.4 |
| 1999 | 182 | 41612 | 12152 | 13.4 | 0.4 | 188 | 1.0 | 1.8 |
|  | Avg. | 42532 | 16702 |  |  | 155 | 0.8 | 1.6 |
| Summer |  |  |  |  |  |  |  | 1.2 |
| 1995 | 208 | 25771 | 15763 | 27.0 | 0.95 | 147 | 0.7 |  |
| 1996 | 163 | 14139 | 7338 | 31.1 | 3.65 | 51 | 0.3 | 0.9 |
| 1997 | 114 | 13560 | 4314 | 23.1 | 1.72 | 56 | 0.5 | 0.7 |
| 1998 | 89 | 7432 | 2699 | 21.3 | 0.34 | 170 | 1.9 | 1.3 |
| 1999 | 140 | 11922 | 5489 | 20.1 | 0.36 | 197 | 1.4 | 6.3 |
|  | Avg. | 14565 | 7536 |  |  | 124 | 1.0 | 3.6 |
|  |  |  |  |  |  |  | 2.6 |  |

Table 3
Summary statistics for assessing the precision of the estimated length distribution of deepwater hake off Namibia based on bottom trawl surveys. The notation is the same as that shown in Table 1.

| Year | n | M | m | $\hat{\mathrm{R}}(\mathrm{cm})$ | $\operatorname{var}(\hat{\mathrm{R}})$ | $\hat{\mathrm{m}}_{\text {eff }}$ | $\hat{\mathrm{m}}_{\text {eff }} / \mathrm{n}$ | $\left(\hat{\mathrm{m}}_{\text {eff }} / \mathrm{m}\right) \times 100 \%$ |
| :--- | ---: | ---: | ---: | :---: | :---: | :---: | :---: | :---: |
| Sep 1990 | 37 | 6671 | 1837 | 24.6 | 1.8 | 35 | 1.0 | 1.9 |
| J an 1991 | 19 | 3887 | 1329 | 29.2 | 3.3 | 19 | 1.0 | 1.4 |
| Oct 1992 | 53 | 22369 | 5090 | 30.1 | 2.1 | 30 | 0.6 | 0.6 |
| Apr 1992 | 63 | 33107 | 5411 | 34.0 | 2.8 | 30 | 0.5 | 0.6 |
| Oct 1993 | 88 | 36814 | 8480 | 30.3 | 2.6 | 41 | 0.5 | 0.5 |
| J an 1993 | 70 | 36247 | 8208 | 37.3 | 2.9 | 33 | 0.5 | 0.4 |
| Apr 1993 | 84 | 25746 | 7023 | 32.7 | 4.2 | 35 | 0.4 | 0.5 |
| J an 1994 | 60 | 30134 | 7997 | 28.4 | 10.0 | 13 | 0.2 | 0.2 |
| Apr 1994 | 103 | 72012 | 17694 | 35.3 | 1.4 | 63 | 0.6 | 0.4 |
| Oct 1994 | 105 | 70817 | 17216 | 28.1 | 1.9 | 44 | 0.4 | 0.3 |
| Apr 1995 | 79 | 47585 | 14661 | 26.0 | 7.0 | 20 | 0.3 | 0.2 |
| J an 1996 | 105 | 57540 | 27834 | 30.3 | 2.9 | 45 | 0.4 | 0.2 |
| Sep 1996 | 105 | 78562 | 24975 | 28.7 | 1.6 | 50 | 0.5 | 0.2 |
| J an 1997 | 122 | 54995 | 27648 | 28.5 | 4.6 | 29 | 0.2 | 0.1 |
| J an 1998 | 104 | 52573 | 15717 | 34.5 | 2.3 | 44 | 0.4 | 0.4 |
| J an 1999 | 104 | 68419 | 19305 | 28.4 | 4.1 | 30 | 0.3 | 0.2 |
|  | Avg. | 43592 | 13152 |  |  | 35 | 0.5 | 0.5 |

well within the $95 \%$ confidence interval for $\hat{\mathrm{R}}$ (Fig. 1). The results of the simulations for the winter survey in 1999 were similar to those for 1995 (Fig. 2).

Bootstrapped estimates of the 95\% confidence intervals for the estimated proportion of fish in each $5-\mathrm{cm}$ length bin for the 1995 and 1999 Barents Sea winter survey of cod are shown in Figure 3. The inner brackets denote the confidence interval based on the total number of fish actually measured and the outer brackets denote the confi-
dence intervals if a maximum of 10 fish were measured per subsample.

## Discussion and conclusions

For all the surveys examined, the effective sample size for the survey length data was much smaller than the total number of fish measured. The average effective sample size

Table 4
Summary statistics for assessing the precision of the estimated length distribution of deepwater hake off South Africa based on bottom trawl surveys. The notation is the same as that shown in Table 1.

| Year | n | M | m | $\hat{\mathrm{R}}(\mathrm{cm})$ | $\operatorname{var}(\hat{\mathrm{R}})$ | $\hat{\mathrm{m}}_{\text {eff }}$ | $\hat{\mathrm{m}}_{\text {eff }} / \mathrm{n}$ | $\left(\hat{\mathrm{m}}_{\text {eff }} / \mathrm{m}\right) \times 100 \%$ |
| :--- | :---: | ---: | :---: | :---: | :---: | :---: | :---: | :---: |
| J an 1985 | 75 | 75883 | 13863 | 29.3 | 0.7 | 70 | 0.9 | 0.5 |
| J ul 1985 | 65 | 55704 | 10786 | 29.9 | 0.9 | 52 | 0.8 | 0.5 |
| J an 1986 | 86 | 82720 | 16216 | 28.2 | 0.3 | 132 | 1.5 | 0.8 |
| J an 1987 | 91 | 140685 | 18317 | 26.7 | 0.3 | 122 | 1.3 | 0.7 |
| J ul 1987 | 76 | 80476 | 14774 | 27.0 | 0.8 | 68 | 0.9 | 0.5 |
| Feb 1988 | 88 | 91828 | 17187 | 24.3 | 0.5 | 98 | 1.1 | 0.6 |
| Feb 1989 | 33 | 234796 | 10115 | 25.7 | 0.1 | 207 | 6.3 | 2.0 |
| J an 1990 | 75 | 150814 | 23093 | 26.0 | 1.6 | 43 | 0.6 | 0.2 |
| Jan 1991 | 73 | 226234 | 17115 | 27.2 | 0.9 | 38 | 0.5 | 0.2 |
| Feb 1992 | 83 | 174364 | 24334 | 25.2 | 0.8 | 58 | 0.7 | 0.2 |
| J an 1993 | 81 | 102395 | 24922 | 26.0 | 0.6 | 89 | 1.1 | 0.4 |
| J an 1994 | 63 | 139268 | 28621 | 27.1 | 0.7 | 68 | 1.1 | 0.2 |
| Jan 1995 | 81 | 137225 | 37481 | 26.4 | 0.5 | 97 | 1.2 | 0.3 |
| J an 1996 | 77 | 167765 | 31538 | 25.2 | 1.1 | 46 | 0.6 | 0.1 |
| J an 1997 | 88 | 219218 | 33537 | 24.2 | 0.7 | 70 | 0.8 | 0.2 |
| J an 1998 | 76 | 251050 | 26328 | 25.9 | 0.3 | 116 | 1.5 | 0.4 |
| J an 1999 | 74 | 218438 | 26144 | 25.4 | 0.5 | 91 | 1.2 | 0.3 |
|  | Avg. | 149933 | 22022 |  |  | 86 | 1.3 | 0.5 |



Figure 1
Simulated estimates of the distribution of the effective sample size, $\hat{m}_{\text {eff }}$, and of the mean length, R , when the total number of fish measured is reduced for the 1995 winter survey in the Barents Sea. The top panel is when a maximum of $k=100$ fish are selected per subsample for a total of $m=30,403$ fish in each run; the middle panel, $\mathrm{k}=30, \mathrm{~m}=11,123$; and the bottom panel, $\mathrm{k}=$ $10, m=3911$. The estimate of the population mean based on the entire sample ( $m=47,286$ ) is 19.96 and its $95 \%$ confidence interval is (18.29, 21.63).
was approximately one fish per tow and it seems to be typical that the effective sample size for estimating length distributions is relatively small for trawl surveys. For example, the effective sample size for trawl surveys of haddock on Georges Bank was on average less than 0.5 fish per tow (Pennington and Vølstad, 1994) and for shrimp in a small area off West Greenland, about 3 shrimp per tow (Folmer and Pennington, 2000).

The reason that the effective samplesizes were small, and, therefore, the estimates of the length distributions were rather imprecise, given the number of fish that were measured, is that the lengths of fish in a haul tend to be more similar than those in the entire population. An additional factor is that the density of fish in a survey region is usually quite variable. To see this, consider the equation for the expected value of $\operatorname{var}(R)$. If every fish is measured during a survey, then subject to some assumptions, the expected variance of R when n stations are sampled is given approximately by (Pennington and Vølstad, 1994)

$$
\operatorname{var}(\hat{R})=\frac{\sigma_{\times}^{2}\left\{1+\left(\overline{\mathrm{M}}-1+\sigma_{m}^{2} / \overline{\mathrm{M}}\right) \rho\right\}}{\mathrm{M}},
$$

where $\bar{M}=$ the expected mean catch per tow; $\sigma_{\mathrm{m}}^{2}=$ the tow-to-tow variance of catch;
$M(=n \bar{M})=$ the expected total number of fish caught; $\sigma_{x}^{2}=$ the population variance of length; and $\rho=$ the coefficient of intrahaul correlation (see Cochran, 1977, p. 209) for length.


Figure 2
Simulated estimates of the distribution of the effective sample size, $\hat{m}_{\text {eff }}$, and of the mean length, $\hat{R}$, when the total number of fish measured is reduced for the 1999 winter survey in the Barents Sea. The top panel is when a maximum of $k=100$ fish are selected per subsample for a total of $m=17,615$ fish in each run; the middle panel, $k=$ $30, m=7240$; and the bottom panel, $k=10, m=2597$. The estimate of the population mean based on the entire sample ( $m=21,769$ ) is 24.96 and its $95 \%$ confidence interval is ( $22.26,27.66$ ).

If $\rho=0$, then $\operatorname{var}(\hat{R})=\sigma_{X}^{2} / M$ and the effective sample size is equal to $M$. However if $\rho>0$ (i.e. fish of similar length tend to be caught together), then the terms in the parentheses can greatly increase the variance and thus drastically reduce the effective size. In particular, the term $\sigma_{x}^{2} / \overline{\mathrm{M}}$ is relatively large for trawl surveys. Finally, if $\rho<0$, which is rarely if ever the case for trawl surveys, then the effective sample size will be larger than $M$.

The precision of estimates of other population characteristics, such as age distribution, can also be relatively low compared with the number of fish sampled if the particular attribute or measurement is more similar for fish caught together than for those in the general population. For example, the precision of estimates of mean stomach contents (Bogstad et al., 1995) or diet composition (Tirasin and J ørgensen, 1999) can be relatively low because of intrahaul correlation.

An effective sample size of one fish per tow does not mean only one fish should be measured at each station, but it implies that the only way to improve survey precision significantly is to increase the number of stations, i.e. to sample fish from as many locations as possible. The bootstrapped estimates of precision and the sampling simulations showed that reducing or increasing the number of


Figure 3
Bootstrapped estimates of the $95 \%$ confidence intervals for the proportion of cod in the Barents Sea in each $5-\mathrm{cm}$ length bin, for winter 1995 and for winter 1999. The inner brackets denote the confidence intervals if the estimates are based on all the cod measured during the surveys and the outer brackets denote the confidence intervals if 10 fish are measured for each subsample.
fish measured (or caught and measured) at a station will not significantly affect the precision of length-distribution estimates. In general, if intracluster correlation is positive for an attribute, then it is usually best to take a small sample from as many locations as possible (e.g. Bogstad et al., 1995; McGarvey and Pennington, 2001).
It has been shown that tows of short duration are in general more efficient for estimating stock abundance than long tows (Godø et al., 1990; Pennington and Vølstad, 1991; Gunderson, 1993; Carlsson et al., 2000). Therefore one way to collect samples from more locations and improve overall survey efficiency without increasing survey cost is to reduce tow duration and use the time saved to increase the number of survey stations (Pennington and Vølstad, 1994). For example, if tow duration were reduced from 60 minutes to 15 minutes for a trawl survey of shrimp off West Greenland, then $44 \%$ more stations could be surveyed (Carlsson et al., 2000). Likewise, a reduction in tow duration from 30 minutes to 10 minutes for a trawl survey on Georges Bank would increase the number of survey stations by about 30\% (Pennington and Vølstad, 1994).

The total number of fish caught would be fewer, on average, if tow duration was reduced, but estimates of fish density would be more precise and the resulting sample of individuals would be more representative of the entire population (Pennington and Vølstad, 1994).

## Acknowledgments

We thank Rob Leslie (Marine and Coastal Management Centre, South Africa) for providing us with the South African survey data, and J on HelgeVølstad (Versar, Inc., USA) and two anonymous referees for their constructive comments and suggestions.

## Literature cited

Bhattacharyya, G. K., and R. A. J ohnson.
1977. Statistical concepts and methods. J ohn Wiley and Sons, New York, NY, 639 p.
Bogstad, B., M. Pennington, and J. H. Vølstad.
1995. Cost-efficient survey designs for estimating food consumption by fish. Fish. Res. 23:37-46.
Carlsson, D., P. Kanneworff, O. Folmer, M. Kingsley, and
M. Pennington.
2000. Improving the West Greenland trawl survey for shrimp (Pandalus borealis). J. Northwest AtI. Fish. Sci. 27:151-160.
Christensen, V.
1996. Virtual population reality. Rev. Fish Biol. Fish. 6: 243-247.
Cochran, W. G.
1977. Sampling techniques, $3^{\text {rd }}$ ed. J ohn Wiley and Sons, New York, NY, 428 p.
Efron, $B$.
1982. Thejackknife, the bootstrap, and other resampling plans. Society for Industrial and Applied Mathematicians (SIAM), Conference Board of the Mathematical Sciences (CBMS)National Science Foundation (NSF) regional conferenceseries in applied mathematics 38, Philadelphia, PA, 92 p.
Folmer O., and M. Pennington.
2000. A statistical evaluation of the design and precision of the shrimp survey off West Greenland. Fish. Res. 45: 165-178.
Godø, O. R.
1994. Factors affecting the reliability of groundfish abun-
dance estimates from bottom trawl surveys. In Marine fish behaviour in capture and abundance estimation (A. Fernø and S. Olsen, eds.), p. 166-199. Fishing News Books, Farnham, UK.
Godø, O. R., M. Pennington, and J.H. Vølstad.
1990. Effect of tow duration on length composition of trawl catches. Fish. Res. 9:165-179.
Gunderson, D. R.
1993. Surveys of fisheries resources. J ohn Wiley and Sons, New York, NY, 248 p.
Korsbrekke, K., S. Mehl, O. Nakken, and M. Pennington.
2001. A survey-based assessment of the Northeast Arctic cod stock. ICES J. Mar. Sci. 58:763-769.
McGarvey, R., and M. Pennington.
2001. Designing and evaluating length-frequency surveys for trap fisheries with application to the southern rock lobster. Can. J. Fish. Aquat. Sci. 58:254-261.
Nakken, O.
1998. Past, present and future exploitation and management of marine resources in the Barents Sea and adjacent areas. Fish. Res. 37:23-35.
Payne, A. I. L., C. J. Augustyn, and R. W. Leslie.
1985. Biomass index and catch of Cape hake from random stratified sampling cruises in division 1.6 during 1984. Colln. Scient. Pap. Int. Com. SE AtI. Fish. 12:99-123.
Pennington, $\mathrm{M}_{\mathrm{L}}$, and T . Strømme.
1998. Surveys as a research tool for managing dynamic stocks. Fish. Res. 37:97-106.
Pennington, M., and J. H. Vølstad.
1991. Optimum size of sampling unit for estimating the density of marine populations. Biometrics 47:717-723.
1994. Assessing the effect of intra-haul correlation and variable density on estimates of population characteristics from marine surveys. Biometrics 50:725-732.
Research Triangle Institute.
2001. SUDAAN user's manual, release 8.0. Research Triangle Institute, Research Triangle Park, NC, 886 p.
Tirasin, E. M., and T. J ørgensen.
1999. An evaluation of the precision of diet description. Mar. Ecol. Prog. Ser. 182:243-252.


[^0]:    ${ }^{1}$ Also known as "Atlantic cod" and "haddock," respectively, according to Common and scientific names of fishes from the United States and Canada. 1991. Am. Fish. Soc. Spec. publ. 20, Besthesda, MD, 183 p .
    ${ }^{2}$ Aglen, A. 1999. Report on the demersal fish surveys in the Barents Sea and Svalbard area during summer/autumn 1996 and 1997. Unpubl. manuscr., Fisket og Havet NR. 7-1999. Institute of Marine Research, PO Box 1870 Nordnes, N-5817 Bergen, Norway.

[^1]:    ${ }^{3}$ Anonymous. 1999. Survey of fish resources of Namibia. Preliminary cruise report 1/99. NORAD-FAO/UNDP Project GLO92/013, 52. Unpubl. manuscr. NatMIRC, PO Box 912, Swakopmund, Namibia.

