

1 Seagrass biofilm communities at a naturally CO<sub>2</sub>-rich vent

2

3 Hassenrück, C., Hofmann, L. C., Bischof, K., Ramette, A.

4 Author

5 Christiane Hassenrück

6 Max Planck Institute for Marine Microbiology

7 HGF MPG Group for Deep-Sea Ecology and Technology

8 Celsiusstraße 1

9 Bremen

10 Germany

11 28359

12 (E) [chassenr@mpi-bremen.de](mailto:chassenr@mpi-bremen.de)

13

14 Author

15 Laurie C Hofmann

16 University of Bremen

17 BreMarE, FB 02

18 Leobener Str. NW2

19 Bremen

20 Germany

21 28359

22 (E) [lhofmann@mpi-bremen.de](mailto:lhofmann@mpi-bremen.de)

---

This article has been accepted for publication and undergone full peer review but has not been through the copyediting, typesetting, pagination and proofreading process, which may lead to differences between this version and the Version of Record. Please cite this article as doi: 10.1111/1758-2229.12282

23

24 Author

25 Kai Bischof

26 University of Bremen

27 BreMarE, FB 02

28 Leobener Str. NW2

29 Bremen

30 Germany

31 28359

32 (E) kbischof@uni-bremen.de

33

34 Corresponding Author

35 Alban Ramette

36 Max Planck Institute for Marine Microbiology

37 HGF MPG Group for Deep-Sea Ecology and Technology

38 Celsiusstraße 1

39 Bremen

40 Germany

41 28359

42 (E) aramette@mpi-bremen.de

43 **Summary**

44 Seagrass meadows are a crucial component of tropical marine reef ecosystems.

45 Seagrass plants are colonized by a multitude of epiphytic organisms that contribute to

46 broadening the ecological role of seagrasses. To better understand how environmental  
47 changes like ocean acidification might affect epiphytic assemblages, the microbial  
48 community composition of the epiphytic biofilm of *Enhalus acroides* was investigated at  
49 a natural CO<sub>2</sub> vent in Papua New Guinea using molecular fingerprinting and next  
50 generation sequencing of 16S and 18S rRNA genes. Both bacterial and eukaryotic  
51 epiphytes formed distinct communities at the CO<sub>2</sub>-impacted site compared to the control  
52 site. This site-related CO<sub>2</sub> effect was also visible in the succession pattern of microbial  
53 epiphytes. We further found an increased abundance of bacterial types associated with  
54 coral diseases at the CO<sub>2</sub>-impacted site (*Fusobacteria*, *Thalassomonas*) whereas  
55 eukaryotes such as certain crustose coralline algae commonly related to healthy reefs  
56 were less diverse. These trends in the epiphytic community of *E. acroides* suggest a  
57 potential role of seagrasses as vectors of coral pathogens and may support previous  
58 predictions of a decrease in reef health and prevalence of diseases under future ocean  
59 acidification scenarios.

60

61 **Keywords:** ocean acidification, natural CO<sub>2</sub> vents, seagrass, epiphytes, microbial  
62 community composition, coral reef ecology

63

64 **Introduction**

65

66 Tropical marine reef ecosystems are hotspots of biodiversity and productivity in an  
67 otherwise desert-like marine system. Apart from corals, seagrass meadows are a crucial  
68 component of these reef ecosystems. As fish nurseries, nutrient cyclers, organic carbon

69 producers and sediment stabilizers, seagrass meadows contribute substantially to  
70 ecosystem functioning (Orth et al., 2006). Similar to corals (Mouchka et al., 2010),  
71 seagrasses are colonized by microorganisms that form epiphytic biofilms on the seagrass  
72 leaves (Michael et al., 2008). These biofilms have been shown to affect seagrass  
73 physiology as well as their interactions with other reef organisms by e.g. regulating light  
74 availability (Sand-Jensen, 1977), influencing the settlement of secondary epibionts and  
75 biofouling (Wahl, 1989) or the production of antimicrobial substances (Marhaeni et al.,  
76 2011). As such, a seagrass plant and its epiphytic biofilm can be referred to as a seagrass  
77 holobiont.

78 Ocean acidification (OA), defined as a decrease in ocean water pH caused by  
79 increased atmospheric CO<sub>2</sub> concentrations, is among the most worrisome threats to coral  
80 reef ecosystems (Hoegh-Guldberg et al., 2007). The impacts of OA on corals reach from  
81 a decrease of skeletal integrity (Hoegh-Guldberg et al., 2007) to changes in the  
82 composition of the microbial biofilm associated with the coral reducing larval settlement  
83 and probably coral health (Meron et al., 2011; Webster et al., 2013). Seagrasses, on the  
84 other hand, are generally thought to benefit from OA due to the increased availability of  
85 CO<sub>2</sub> and bicarbonate for photosynthesis (Koch et al., 2013; Brodie et al., 2014).  
86 However, data on how the epiphytic biofilm on seagrass leaves might respond to OA and  
87 on the behavior of the seagrass holobiont in future OA scenarios are still sparse.

88 Several studies have investigated the epiphytic community on seagrass leaves giving  
89 detailed information on the composition of bacterial or eukaryotic epiphytes (Uku et al.,  
90 2007; Medina-Pons et al., 2009; Hamisi et al., 2013). The effect of ocean acidification on  
91 epiphytic communities on seagrass leaves is far less well documented. Previous studies

92 reported a decrease of calcifying epiphytes such as crustose coralline algae (Martin et al.,  
93 2008; Donnarumma et al., 2014) as already seen elsewhere in coral reefs (Fabricius et al.,  
94 2011). Donnarumma et al. (2014) also highlighted the decrease in epiphyte diversity with  
95 decreasing pH. However, both studies only visually identified epiphytes by using light  
96 microscopy and did not address the multitude of cryptic epiphytes detectable only with  
97 the increased sensitivity and taxonomic resolution of molecular tools. Our study aims i)  
98 to provide a first overview of both bacterial and eukaryotic epiphytes at a molecular level,  
99 and ii) to estimate how the epiphytic community on seagrass leaves may change in  
100 response to OA. This may thus help increase our understanding of the part the seagrass  
101 holobiont may play in the reef ecosystem under future OA scenarios.

102 Recent research has turned to naturally CO<sub>2</sub>-rich systems as models for future OA  
103 scenarios (Hall-Spencer et al., 2008; Fabricius et al., 2011; Lidbury et al., 2012; Kerfahi  
104 et al., 2014). Unlike laboratory experiments, which are usually restricted to short-term  
105 studies, natural sites offer the opportunity to predict OA effects in long-term adapted  
106 systems that can be studied in their entirety without the need for experimental  
107 manipulation (Hall-Spencer et al., 2008). However, the inherent complexity of natural  
108 systems can also confound OA effects and caution is needed in selecting natural CO<sub>2</sub> rich  
109 sites for OA research (Vizzini et al., 2013).

110 Here, the epiphytic biofilm on the leaves of the seagrass *Enhalus acroides* was  
111 investigated at a natural CO<sub>2</sub> vent and a control site in Papua New Guinea (PNG; figure  
112 S1). The sites were previously described as potential sites to study long-term effects of  
113 OA on coral reef communities since the prevailing environmental conditions are assumed  
114 to have been stable for up to 100 years (Fabricius et al., 2011). The diversity and

115 composition of both bacterial and eukaryotic microbial epiphytic communities were  
116 assessed using molecular community fingerprinting and next generation sequencing of  
117 amplicon libraries. Besides the site-related CO<sub>2</sub> impact, the factor leaf age was included  
118 in the analysis to account for different developmental stages of the epiphytic biofilm as  
119 well as potential interactions of biofilm development with OA effects. To further  
120 characterize the seagrass leaves and their epiphytes, additional data was collected on total  
121 epiphyte cover and carbon and nitrogen content of the seagrass leaves.

122

## 123 **Results and Discussion**

124 Logger deployments over approx. 44h at the vent site at Dobu Island (Figure S1)  
125 recorded median pH values of 7.8 in the water column (Fabricius, pers. communication).  
126 At the control site pH values of 8.3 were measured (Hassenrück et al., unpublished).  
127 These values were consistent with previous data on the carbonate system at Dobu Island  
128 (Fabricius et al., 2011). Apart from the carbonate system, the physicochemical  
129 characteristics of the water at the two sampling sites were very similar suggesting that  
130 changes in the carbonate system between the vent and the control site were not  
131 confounded by any other of the observed parameters (Fabricius et al., 2011).

132 *E. acroides* shoots were collected at approximately 4 m water depth at each of the  
133 two sampling sites in May 2013. 18S ribosomal DNA sequence confirmed that the  
134 seagrass shoots belonged to one species and did not show any pattern by sampling site  
135 (data not shown). Each shoot consisted of 3 to 5 leaf pairs that were ranked by their order  
136 of budding, i.e. leaf age, with the youngest leaf pair being assigned the first rank. When  
137 possible we sampled ranks 1 to 4 (youngest to oldest). On average, *E. acroides* is

138 expected to produce a new pair of leaves approximately every month (Johnstone, 1979;  
139 Brouns and Heijns, 1986; Agawin et al., 2001). The time covered in this study would then  
140 amount to four to five months of settlement, although it is possible that growth rates were  
141 higher under low pH conditions (Koch et al., 2013). During that time, carbon (C) content  
142 of the seagrass leaves decreased with leaf age from approximately 33% to 26% dry  
143 weight (ANOVA,  $F_{1,38} = 25.986$ ,  $p < 0.001$ ) and nitrogen (N) content from 2% to almost  
144 1% (Kruskal-Wallis  $\chi^2 = 21.262$ ,  $df = 3$ ,  $p < 0.001$ ; Table 1). Carbon and nitrogen  
145 measurements matched previous measurements of leaves of *E. acroides* (Yamamuro et  
146 al., 2004) and were not affected by sampling site suggesting that the substrate type, i.e.  
147 the seagrass leaf, was not confounded between sampling sites.

148 Epiphyte cover increased with leaf age (Table 1). At the vent site, this increase  
149 reached only about three fold lower values than under control conditions, most likely due  
150 to a lower abundance of pH sensitive organisms such as crustose coralline algae (Corlett  
151 and Jones, 2007; Martin et al., 2008, Fabricius et al., unpublished). However, regardless  
152 of the trend in epiphyte cover, at the high taxonomic resolution provided by 16S and 18S  
153 amplicon sequencing, epiphyte communities seemed to be as diverse at the vent site than  
154 at the control site (Table 1).

155

156 *Molecular Fingerprinting using Automated Ribosomal Intergenic Spacer Analysis*  
157 (*ARISA*)

158 As a first step to assessing the composition of the epiphytic biofilm of *E. acroides*,  
159 the epiphytic community was screened using the molecular fingerprinting technique  
160 ARISA (Ramette, 2009; Wolf et al., 2013). ARISA identified 408 bacterial and 321

161 eukaryotic operational taxonomic units (OTUs). Non-metric multidimensional scaling  
162 (NMDS) plot based on Bray-Curtis dissimilarity coefficients revealed three prominent  
163 patterns in the bacterial and eukaryotic community structure (Figure 1).

164 First, there was a strong separation of the communities sampled at the vent and the  
165 control site, which tended to cluster away from each other (bacteria: ANOSIM,  $R =$   
166  $0.775$ ,  $p < 0.05$ ; eukaryotes:  $R = 0.692$ ,  $p < 0.05$ ; Table S1). Only about 30% of the  
167 bacterial and eukaryotic OTUs were shared between any two samples from the vent and  
168 the control site. Redundancy analysis (RDA) further confirmed that both sampling site  
169 and leaf age significantly explained part of the variation in the microbial community  
170 structure (Table S2). Of the observed parameters, sampling site was the dominant factor  
171 responsible for the patterns in epiphytic community structure (bacteria: adjusted  $R^2 =$   
172  $27.3\%$ ; eukaryotes: adjusted  $R^2 = 12.4\%$ ) with about four times more variation being  
173 explained by sampling site than leaf age (Table S2). This pronounced shift in the  
174 epiphytic community structure on seagrass leaves between vent and control site further  
175 supports previous results, which found a response of bacterial as well as eukaryotic  
176 microbes to OA in other habitats (Johnson et al., 2011; Lidbury et al., 2012; Kerfahi et  
177 al., 2014).

178 Second, at each site there appeared to be a successive shift in epiphyte communities  
179 from the youngest to older leaves (Table S1). Despite the differences in epiphytic  
180 community composition between the vent and the control site, a successional pattern in  
181 community composition from younger to older leaves was observed at both sites  
182 regardless of  $CO_2$  impact (Figure 1). Since organic matter has been shown to be  
183 transferred from the seagrass leaves to the epiphytes (Michael et al., 2008), changes in



184 carbon and nitrogen content with leaf age as documented here may contribute to the  
185 influence of leaf age in shaping epiphyte communities.

186 Third, apart from the general response to the factors sampling site and leave age,  
187 patterns in community structure between samples, i.e. the pairwise similarity between  
188 samples, correlated strongly between the bacterial and eukaryotic datasets (Mantel test,  $r$   
189 = 0.64,  $p < 0.05$ ). The strong correlation seemed unlikely to be caused exclusively by  
190 changes in abiotic parameters. A more likely explanation may be that both communities  
191 influence and shape each other as previously suggested by (Steele et al., 2011; Sawall et  
192 al., 2012).

193 *{insert figure 1}*

194

#### 195 *Amplicon sequencing of epiphytic communities*

196 To taxonomically classify the epiphytic communities on *E. acroides*, eight samples  
197 were selected for amplicon sequencing of 16S and 18S rRNA genes for bacterial and  
198 eukaryotic communities, respectively (ENA accession PRJEB7181). From each sampling  
199 site one sample was chosen for each leaf age. OTU clustering was performed at 97%  
200 sequence identity and SILVAngs was used for the taxonomic classification of the OTUs  
201 (Quast et al., 2013). A more detailed description of the sequence processing workflow  
202 can be found in SI text 1.

203 Amplicon sequencing of the V4-V6 variable region of the bacterial 16S rRNA gene  
204 recovered 2,179 OTUs with about 600 OTUs per sample. Approximately 62% of the  
205 OTUs were singletons (47%) or doubletons (15%), which accounted for 8-16% of the  
206 total sequence counts per sample. This percentage of rare bacterial types did not

207 significantly vary between sampling sites (Welch's t-test,  $t = -0.944$ ,  $df = 3.817$ ,  $p >$   
208  $0.05$ ). The Chao1 index of total OTU richness yielded estimates almost twice as high as  
209 the raw counts. There was no significant difference in OTU richness between the  
210 sampling sites (Welch's t-test,  $t = -0.819$ ,  $df = 2.204$ ,  $p > 0.05$ ; Table 1). Previous reports  
211 on bacterial richness and rare bacterial types using next generation sequencing  
212 technology showed inconsistent responses to OA (Kerfahi et al., 2014; Raulf et al., 2015;  
213 Hassenrück et al., unpublished), which might be explained by the difference in  
214 environments being investigated. As such, the lack of change in bacterial richness and  
215 rare bacterial types on seagrass leaves at the vent site should not be generalized beyond  
216 the scope of this study.

217 Amplicon sequencing of the V4 variable region of the eukaryotic 18S rRNA gene  
218 recovered 3,928 OTU. OTU number per sample ranged from 277 (C1) to 664 (C2; Table  
219 1). Similar to the bacterial OTU richness, there was no significant trend in the OTU  
220 number between sampling sites (Welch's t-test,  $t = 0.034$ ,  $df = 3.454$ ,  $p > 0.05$ ). This  
221 result was consistent with that of Lidbury et al. (2012) who did not detect a response of  
222 eukaryotic microbial richness on settlement tiles to OA using a molecular fingerprinting  
223 technique. However, the scarcity of OA studies on eukaryotic microbes applying next  
224 generation sequencing technology does not allow for a more comprehensive discussion  
225 on how eukaryotic epiphyte richness may respond to OA.

226

### 227 *Taxonomic composition of bacterial epiphytes*

228 Most of the bacterial sequences belonged to the phylum Proteobacteria (51%), with  
229 Gammaproteobacteria (38%) and Alphaproteobacteria (11%) constituting the majority.

230 The next most abundant phyla were Cyanobacteria (30%, chloroplast sequences 27%),  
231 Bacteroidetes (12%, Flavobacteria: 8%) and Fusobacteria (4%), which were especially  
232 abundant on older leaves at the vent site (Figure 2A). The high percentage of Gamma-  
233 and Alphaproteobacteria was consistent with previous observations on bacterial epiphytes  
234 of tropical seagrasses (Weidner et al., 2000; Uku et al., 2007). The high percentage of  
235 chloroplast sequences may be explained by the origin of the samples, which were taken in  
236 the photic zone from a chloroplast-containing substratum that was also colonized by  
237 algae. We identified several taxa that may potentially be influenced by sampling site  
238 and/or age of the seagrass leaves (Table S3). Notice that taxa that seemed to be  
239 predominantly affected by leaf age are not further discussed here, because the main  
240 objective of our study was to describe potential OA effects on epiphytic microbes.

241 Cyanobacteria appeared to have a higher relative abundance at the control site than  
242 at the CO<sub>2</sub>-impacted vent site. Predictions of OA effects on free living cyanobacteria are  
243 controversial and range from no effect on metabolic rates (Gradoville et al., 2014) to an  
244 increase in carbon and nitrogen fixation (Hutchins et al., 2007; Lomas et al., 2012). In  
245 microbial biofilms, OA seemed to decrease cyanobacterial abundance and diversity (Witt  
246 et al., 2011; Russell et al., 2013). In complex assemblages, cyanobacteria are supposed to  
247 benefit less from OA than other photosynthetic organisms such as chlorophytes, and may  
248 thus be outcompeted by them (Low-Décarie et al., 2014). In agreement with this  
249 hypothesis, cyanobacteria seemed to decrease in relative abundance with decreasing pH  
250 in this study: e.g. the two nitrogen fixing genera *Leptolyngbya* and *Lyngbya*, which are  
251 known epiphytes of seagrasses (Uku et al., 2007; Hamisi et al., 2013), were more  
252 abundant at the control site, the latter even being unique to the control site. In the case of

253 *Leptolyngbya*, this response has been documented before in a temperate system (Russell  
254 et al., 2013), whereas *Lyngbya* is expected to react more to changes in temperature and  
255 nutrient availability than to OA (Paerl and Huisman, 2009).

256 Contrarily to cyanobacteria, Deltaproteobacteria, Bacilli, Fusobacteria and Clostridia  
257 seemed to increase in relative abundance at the vent site. Within the Deltaproteobacteria  
258 this increase was mostly due to an increase in the relative abundance of OTUs of the  
259 order Bdellovibrionales at the vent site as also observed by Raulf et al. (2015) in  
260 sediments from PNG. The responses of Bacilli and Fusobacteria were mostly due to an  
261 increase in the relative abundance of only one OTU belonging to the genus *Paenibacillus*  
262 and to the family Leptotrichiaceae, respectively. For *Paenibacillus*, this response has  
263 previously been observed in sediments under elevated pCO<sub>2</sub> (Kerfahi et al., 2014). The  
264 fusobacterial OTU was among the most abundant OTUs in the dataset (3.5% of all  
265 sequences) and was further identified as a relative of *Propionigenium* sp. with a sequence  
266 identity of 93% to the latter (NCBI accession number KC918186). Fusobacteria are a  
267 group of strictly anaerobic bacteria, which have been associated with tidal flat sediments,  
268 where they contribute to organic matter degradation (Graue et al., 2012), and are present  
269 in the gut microflora of marine invertebrates (Li et al., 2012; Dishaw et al., 2014;  
270 Rungrasamee et al., 2014) and coral biofilm (Morrow et al., 2012). There is evidence  
271 that Fusobacteria associated with corals increase in abundance under OA (Vega Thurber  
272 et al., 2009), which might support our results, although the exceptionally high sequence  
273 abundance of Fusobacteria at the vent site was restricted to the two oldest leaves.  
274 Noticeably, Fusobacteria as well as Clostridia have further been implicated in coral  
275 diseases (Vega Thurber et al., 2009; Sweet et al., 2013).

276 Alphaproteo-, Gammaproteo- and Flavobacteria did not show a response to sampling  
277 site on class level. However, at a higher level of taxonomic resolution, several taxa  
278 appeared to be affected by sampling site (Table S3). Among the most abundant OTUs in  
279 the dataset, those potentially influenced by sampling site belonged to the  
280 Gammaproteobacteria, i.e. *Thalassomonas* (1.6%) and *Marinomonas* (3.8%), which were  
281 more abundant at the vent site, and *Reinekea* (7.2%) and *Melitea* (2.3%), which were  
282 more abundant at control site. Sequence comparison of the OTU belonging to  
283 *Thalassomonas* showed a high sequence identity (99%) to the sequence retrieved by  
284 Webster et al. (2013; NCBI accession number JQ178640), which was associated with the  
285 crustose coralline algae *Hydrolithon* at low pH. It was further closely related (96%  
286 sequence identity) to *Thalassomonas loyana* (NCBI accession number NR043066), the  
287 causative agent of white plague-like disease in corals (Thompson et al., 2006), suggesting  
288 a potentially pathogenic role. The OTU of *Marinomonas* was related to *Marinomonas*  
289 *poseidonica* (99% sequence identity, NCBI accession number NR074719), which has  
290 been reported to be beneficial to seagrass (Celdrán et al., 2012) and may contribute to  
291 increased growth rates at the vent site. *Reinekea* is a genus that might play an important  
292 role in the degradation of organic matter after phytoplankton blooms (Teeling et al.,  
293 2012). Its reduced abundance at the vent site may be caused by the decreased availability  
294 of degradable material presumably due to the lower percentage of epiphyte cover.  
295 However, it also belongs to the order Oceanospirillales, which are common in coral  
296 biofilms and expected to decrease in abundance in diseased corals (Mouchka et al.,  
297 2010). Hardly anything is known about the genus *Melitea* and, although it has been

298 mentioned before in OA research, its response to elevated pCO<sub>2</sub> still remains largely  
299 unknown (Meron et al., 2011).

300 The direction of potential changes (i.e. the increase or decrease) in relative OTU  
301 abundance from control to vent site or vice versa appeared to be related to total OTU  
302 abundance. Whereas approximately equal numbers of abundant OTUs (defined by more  
303 than 1% total sequence abundance) increased towards either the vent or control site, more  
304 OTUs of intermediate abundance level (defined by more than 2 sequence occurrences but  
305 less than 1% total sequence abundance) tended to increase towards the vent site than  
306 towards the control site (Table S4). Although not seen in the rare bacterial types as  
307 previously discussed, this trend might be comparable to the increase in rare types with  
308 decreasing pH observed in marine sediments at PNG (Raulf et al., 2015).

309 Among these increasing OTUs, sulfur oxidizers were overrepresented, some of  
310 which – but not all – were unique to the vent site. This suggests that a higher  
311 concentration of sulfur compounds that can be metabolized by bacteria might be present  
312 in the water column at the vent compared to the control site, although so far no direct  
313 evidence exists for that matter (Fabricius et al., 2011). H<sub>2</sub>S was detected in the sediment  
314 (Artur Fink, pers. communication) and gas, but H<sub>2</sub>S levels in the water column did not  
315 exceed values typically observed for seawater (Fabricius et al., 2011). On the other hand,  
316 sulfur-oxidizing bacteria might also constitute a contamination from the sediment and  
317 might not even be active on the seagrass leaves. Furthermore, apart from their  
318 biogeochemical function, sulfur-oxidizing bacteria have also been associated with coral  
319 diseases (Frias-Lopez et al., 2002, 2004; Bourne et al., 2013). Their increased relative  
320 abundance may therefore not only be attributable to sulfide seepage. Other OTUs of

321 intermediate abundance, which increased at the vent site, belonged to genera such as  
322 *Shewanella* and *Vibrio*, which again have been related to coral diseases (Mouchka et al.,  
323 2010; Meron et al., 2011; Garcia et al., 2013; Sweet et al., 2013). The general trend of an  
324 increase in disease-associated bacterial OTUs at the vent site has also been observed in  
325 corals in PNG (Morrow et al., 2014).

326

### 327 *Taxonomic composition of eukaryotic epiphytes*

328 The richness of eukaryotic OTUs was dominated by Florideophycidae, which mostly  
329 consisted of crustose coralline algae (Corallinophycidae, 2282 OTUs) and  
330 Rhodymeniophycidae (235 OTUs), followed by diatoms (695 OTUs), Ulvophyceae (171  
331 OTUs) and dinoflagellates (145 OTUs, figure 2B). This composition conforms with the  
332 findings of microscopy-based work on tropical seagrasses which also reported a  
333 prevalence of crustose coralline algae (Corlett and Jones, 2007; Martin et al., 2008).

334 Potential changes in OTU richness were related to genera of the taxa  
335 Corallinophycidae, Dinoflagellata and Diatomea (Table S5). Corallinophycidae were  
336 slightly less diverse at the vent site, especially on the older leaves where they only  
337 retained about 65% of their OTUs. As calcifying organisms, crustose coralline algae are  
338 likely to suffer from OA (Martin et al., 2008; Fabricius et al., 2011; Donnarumma et al.,  
339 2014). However, some genera appear to be more vulnerable to elevated pCO<sub>2</sub> than others.  
340 Here, *Hydrolithon* the most diverse genus of crustose coralline algae on the leaves of *E.*  
341 *acroides* lost about two thirds of its OTUs, and *Lithophyllum*, which disappeared  
342 completely at the vent site, seemed especially susceptible to acidified conditions. Severe  
343 declines in *Hydrolithon* have also been observed on settlement tiles in PNG (Fabricius et

344 al., unpublished). The calcite deposits of *Hydrolithon* and *Lithophyllum* contain a high  
345 percentage of magnesium, while e.g. *Spongites* which was the only crustose coralline  
346 algae unique to the vent site, deposits calcite with little magnesium content – a form that  
347 is less susceptible to reduced pH than high-Mg calcite (Smith et al., 2012). These  
348 differences in calcite composition may contribute to the resilience of crustose coralline  
349 algae under OA (Ries, 2011; Ragazzola et al., 2013).

350 The genus *Galeidinium* (Dinoflagellata) was more diverse at the vent compared to  
351 the control site. However, the impacts of OA on dinoflagellates, in general, and  
352 *Galeidinium*, in particular, are not very well studied, so that potential implications of an  
353 increased diversity of *Galeidinium* under elevated pCO<sub>2</sub> cannot yet be predicted.

354 Diatoms showed a variable response to sampling site with *Navicula* and  
355 *Grammatophora* being more diverse at the vent and *Cyclophora* and *Cylindrotheca* at the  
356 control site. These changes in the diversity of diatoms largely concurred with previous  
357 findings, which predicted an increase in the genera *Grammatophora* and *Navicula* under  
358 OA with a coinciding decrease of *Cyclophora* and *Cylindrotheca* (Johnson et al., 2011;  
359 Singh and Singh, 2014), which was also the case here. Although photosynthetic  
360 organisms in general are expected to benefit from OA, species-specific responses depend  
361 on the respective ability of each organism to utilize inorganic carbon during  
362 photosynthesis and their comparative competitiveness (Koch et al., 2013).

363 {insert figure 2}

364

365 *Conclusion: does epiphyte composition change due to OA?*



366 We detected a highly diverse bacterial and eukaryotic community on the leaves of *E.*  
367 *acroides*. Although OTU richness seemed unaffected, our results overall suggest a  
368 pronounced and interconnected shift in bacterial and eukaryotic community composition  
369 of the epiphytic biofilm of *E. acroides* with changes in the carbonate system of the  
370 surrounding water. Besides organisms well-known to respond to elevated pCO<sub>2</sub>, this shift  
371 may also include taxa that have not been identified in OA research before. In some cases,  
372 a potential response to elevated pCO<sub>2</sub> was only visible at a very high level of taxonomic  
373 resolution. We further detected an increased prevalence of microbial sequence types  
374 associated with coral diseases at the vent site under elevated pCO<sub>2</sub> conditions. This  
375 agrees with the hypothesis that coral reefs experiencing elevated pCO<sub>2</sub> levels will be  
376 more susceptible to diseases than reefs not yet exposed to OA (Hoegh-Guldberg et al.,  
377 2007). It further highlights the potential of seagrasses as vectors of coral pathogens  
378 (Sweet et al., 2013) and stresses the point that seagrasses should be viewed as a holobiont  
379 when making predictions about OA effects and ecological consequences in coral reefs.  
380 Given the high diversity of the epiphytic community on seagrass leaves, an accurate  
381 assessment of the interaction of seagrasses with other components of reef ecosystems will  
382 also require further knowledge of their epiphytic community composition.

383

#### 384 **Acknowledgements**

385 We thank the scientists and crew of the cruise to Papua New Guinea for their support  
386 during the sampling, especially Katharina Fabricius (AIMS, Australia), Dirk de Beer and  
387 Artur Fink (both MPI for Marine Microbiology, Bremen) for their continued advice. The  
388 research was funded by the German Federal Ministry of Education and Research (BMBF)

389 in the framework of the BIOACID II project, the Max Planck Society and the University  
390 of Bremen.

391

392 **Table 1:** Carbon (C) and nitrogen (N) content in percentage dry weight, C:N ratio and  
393 epiphyte cover of the leaves of *E. acroides*, the number of bacterial and eukaryotic OTUs  
394 obtained through ARISA and amplicon sequencing (Bacteria: 16S rRNA gene, Illumina  
395 sequencing; Eukarya: 18S rRNA gene, 454 sequencing); values constitute mean +/-  
396 standard error where applicable; for the bacterial sequencing dataset Chao1 richness  
397 estimates are given in italics with 95% confidence intervals in brackets.

398

399 **Figure 1:** Non-metric multidimensional scaling (NDMS) plot based on the Bray-Curtis  
400 dissimilarity matrix for bacteria (A) and on the Jaccard dissimilarity matrix for  
401 eukaryotes (B) on leaves of *E. acroides*; both bacterial and eukaryotic communities were  
402 assessed using ARISA; dashed hulls representing a minimum of 30% shared OTUs  
403 between samples within the hull; labeled points: samples selected for 16S/18S amplicon  
404 sequencing.

405

406 **Figure 2:** Taxonomic composition of the epiphytic biofilm on leaves of *E. acroides*, A:  
407 bacterial community based on the relative abundance of OTUs (16S rRNA gene  
408 sequences, 454 sequencing); B: eukaryotic community based on the presence/absence of  
409 OTUs (18S rRNA gene sequences, Illumina); bars colored by bacterial class or  
410 eukaryotic phylum, separated by genus; hatched areas: examples of genera potentially  
411 influenced by site and/or leaf age; bold: bacterial classes or eukaryotic phyla potentially

412 influenced by sampling site (Table S3 and S5); samples ordered by leaf age (left:  
413 youngest, right: oldest) within sampling site.

414

## 415 **Supporting information**

416 **SI text 1:** Sequence processing workflow.

417

418 **Table S1:** Analysis of similarity (ANOSIM) results to test the influence of sampling site  
419 and leaf age on the similarity of bacterial and eukaryotic communities based on ARISA;  
420 upper panel: ANOSIM R based on Bray-Curtis dissimilarity coefficient, lower panel:  
421 ANOSIM R based on Jaccard dissimilarity coefficient; 1-4 representing leaf ages from  
422 youngest to oldest; <sup>a</sup> only 2 samples for oldest leaves from the vent site; false discovery  
423 rate (fdr)-adjusted p-values < 0.1 (°), < 0.05 (\*), < 0.01 (\*\*), < 0.001 (\*\*\*), ANOSIM R  
424 of non-significant results are not reported.

425

426 **Table S2:** RDA results to identify factors significantly explaining the variation in  
427 bacterial and eukaryotic communities based on ARISA; results shown for RDA based on  
428 hellinger-transformed relative abundances and presence/absence (binary) data.

429

430 **Table S3:** Bacterial taxa potentially affected by sampling site based on relative OTU  
431 abundance; information is given on the abundance of the respective taxa (high: > 1% total  
432 abundance, rare: singletons and doubletons), the number of taxa of the next lower  
433 taxonomic level (subtaxa), taxa that cumulatively contributed to 70% of the dissimilarity  
434 between sampling sites (SIMPER analysis based on Bray-Curtis dissimilarity

435 coefficient), factors potentially affecting abundance based on an uncorrected significance  
436 threshold of 0.05 and 0.1 (effect), evaluation of the response of potentially impacted taxa  
437 (response to sampling site and/or leaf age, minor response), direction of potential change  
438 in abundance and whether taxa were exclusive to either sampling site.

439

440 **Table S4:** Number of bacterial (upper panel) and eukaryotic (lower panel) taxa  
441 potentially affected by sampling site and the direction of that effect (increase in  
442 abundance/diversity); in brackets: number of taxa unique to that site.

443

444 **Table S5:** Eukaryotic taxa potentially affected by sampling site based on the  
445 presence/absence of OTUs; information is given on the number of taxa of the next lower  
446 taxonomic level (subtaxa), factors potentially affecting OTU number (diversity) based on  
447 an uncorrected significance threshold of 0.15, evaluation of the response of potentially  
448 impacted taxa (response to sampling site and/or leaf age, minor response), direction of  
449 potential change in diversity and whether taxa were exclusive to either sampling site.

450

451 **Figure S1:** Sampling area in Papua New Guinea showing the two sampling sites, which  
452 were approximately 2 km apart (control site: S9.752, E150.854, vent site: S9.737,  
453 E150.869).

454

#### 455 **References**

456 Agawin, N.S.R., Duarte, C.M., Fortes, M.D., Uri, J.S., and Vermaat, J.E. (2001)  
457 Temporal changes in the abundance, leaf growth and photosynthesis of three co-  
458 occurring Philippine seagrasses. *J. Exp. Mar. Bio. Ecol.* **260**: 217–239.

- 459 Bourne, D.G., van der Zee, M.J.J., Botté, E.S., and Sato, Y. (2013) Sulfur-oxidizing  
460 bacterial populations within cyanobacterial dominated coral disease lesions.  
461 *Environ. Microbiol. Rep.* **5**: 518–524.
- 462 Brodie, J., Williamson, C.J., Smale, D. a., Kamenos, N. a., Mieszkowska, N., Santos, R.,  
463 et al. (2014) The future of the northeast Atlantic benthic flora in a high CO<sub>2</sub> world.  
464 *Ecol. Evol.* **4**: 2787–2798.
- 465 Brouns, J.J.W.M. and Heijs, F.M.L. (1986) Production and biomass of the seagrass  
466 *Enhalus acoroides* (L.f.) Royle and its epiphytes. *Aquat. Bot.* **25**: 21–45.
- 467 Celdrán, D., Espinosa, E., Sánchez-Amat, a, and Marín, a (2012) Effects of epibiotic  
468 bacteria on leaf growth and epiphytes of the seagrass *Posidonia oceanica*. *Mar. Ecol.*  
469 *Prog. Ser.* **456**: 21–27.
- 470 Corlett, H. and Jones, B. (2007) Epiphyte communities on *Thalassia testudinum* from  
471 Grand Cayman, British West Indies: Their composition, structure, and contribution  
472 to lagoonal sediments. *Sediment. Geol.* **194**: 245–262.
- 473 Dishaw, L.J., Flores-Torres, J., Lax, S., Gemayel, K., Leigh, B., Melillo, D., et al. (2014)  
474 The gut of geographically disparate *Ciona intestinalis* harbors a core microbiota.  
475 *PLoS One* **9**: e93386.
- 476 Donnarumma, L., Lombardi, C., Cocito, S., and Gambi, M.C. (2014) Settlement pattern  
477 of *Posidonia oceanica* epibionts along a gradient of ocean acidification : an approach  
478 with mimics. *Mediterr. Mar. Sci.* **15**: 498–509.
- 479 Fabricius, K.E., Langdon, C., Uthicke, S., Humphrey, C., Noonan, S., De'ath, G., et al.  
480 (2011) Losers and winners in coral reefs acclimatized to elevated carbon dioxide  
481 concentrations. *Nat. Clim. Chang.* **1**: 165–169.
- 482 Frias-Lopez, J., Klaus, J.S., Bonheyo, G.T., and Fouke, B.W. (2004) Bacterial  
483 Community Associated with Black Band Disease in Corals. *Appl. Environ.*  
484 *Microbiol.* **70**: 5955–5962.
- 485 Frias-Lopez, J., Zerkle, A.L., Bonheyo, G.T., and Fouke, B.W. (2002) Partitioning of  
486 Bacterial Communities between Seawater and Healthy, Black Band Diseased, and  
487 Dead Coral Surfaces. *Appl. Environ. Microbiol.* **68**: 2214–2228.
- 488 Garcia, G.D., Gregoracci, G.B., Santos, E.D.O., Meirelles, P.M., Silva, G.G.Z., Edwards,  
489 R., et al. (2013) Metagenomic analysis of healthy and white plague-affected  
490 *Mussismilia braziliensis* corals. *Microb. Ecol.* **65**: 1076–1086.
- 491 Gradoville, M.R., White, A.E., Bo, D., Church, M.J., and Letelier, R.M. (2014) Diversity  
492 trumps acidification : Lack of evidence for carbon dioxide enhancement of

- 493 Trichodesmium community nitrogen or carbon fixation at Station ALOHA. *Limnol.*  
494 *Oceanogr.* **59**: 645–659.
- 495 Graue, J., Engelen, B., and Cypionka, H. (2012) Degradation of cyanobacterial biomass  
496 in anoxic tidal-flat sediments: a microcosm study of metabolic processes and  
497 community changes. *ISME J.* **6**: 660–669.
- 498 Hall-Spencer, J.M., Rodolfo-Metalpa, R., Martin, S., Ransome, E., Fine, M., Turner,  
499 S.M., et al. (2008) Volcanic carbon dioxide vents show ecosystem effects of ocean  
500 acidification. *Nature* **454**: 96–99.
- 501 Hamisi, M., Díez, B., Lyimo, T., Ininbergs, K., and Bergman, B. (2013) Epiphytic  
502 cyanobacteria of the seagrass *Cymodocea rotundata*: diversity, diel *nifH* expression  
503 and nitrogenase activity. *Environ. Microbiol. Rep.* **5**: 367–376.
- 504 Hoegh-Guldberg, O., Mumby, P.J., Hooten, A.J., Steneck, R.S., Greenfield, P., Gomez,  
505 E., et al. (2007) Coral reefs under rapid climate change and ocean acidification.  
506 *Science* **318**: 1737–1742.
- 507 Hutchins, D.A., Fu, F.-X., Zhang, Y., Warner, M.E., Feng, Y., Portune, K., et al. (2007)  
508 CO<sub>2</sub> control of *Trichodesmium* N<sub>2</sub> fixation, photosynthesis, growth rates, and  
509 elemental ratios: Implications for past, present, and future ocean biogeochemistry.  
510 *Limnol. Oceanogr.* **52**: 1293–1304.
- 511 Johnson, V.R., Brownlee, C., Rickaby, R.E.M., Graziano, M., Milazzo, M., and Hall-  
512 Spencer, J.M. (2011) Responses of marine benthic microalgae to elevated CO<sub>2</sub>.  
513 *Mar. Biol.* **160**: 1813–1824.
- 514 Johnstone, I.M. (1979) Papua New Guinea seagrasses and aspects of the biology and  
515 growth of *Enhalus acoroides* (L.f.) Royle. *Aquat. Bot.* **7**: 197–208.
- 516 Kerfahi, D., Hall-Spencer, J.M., Tripathi, B.M., Milazzo, M., Lee, J., and Adams, J.M.  
517 (2014) Shallow Water Marine Sediment Bacterial Community Shifts Along a  
518 Natural CO<sub>2</sub> Gradient in the Mediterranean Sea Off Vulcano, Italy. *Microb. Ecol.*  
519 **67**: 819–828.
- 520 Koch, M., Bowes, G., Ross, C., and Zhang, X.-H. (2013) Climate change and ocean  
521 acidification effects on seagrasses and marine macroalgae. *Glob. Chang. Biol.* **19**:  
522 103–132.
- 523 Li, S., Sun, L., Wu, H., Hu, Z., Liu, W., Li, Y., and Wen, X. (2012) The intestinal  
524 microbial diversity in mud crab (*Scylla paramamosain*) as determined by PCR-  
525 DGGE and clone library analysis. *J. Appl. Microbiol.* **113**: 1341–1351.

- 526 Lidbury, I., Johnson, V., Hall-Spencer, J.M., Munn, C.B., and Cunliffe, M. (2012)  
527 Community-level response of coastal microbial biofilms to ocean acidification in a  
528 natural carbon dioxide vent ecosystem. *Mar. Pollut. Bull.* **64**: 1063–1066.
- 529 Lomas, M., Hopkinson, B., Losh, J., Ryan, D., Shi, D., Xu, Y., and Morel, F. (2012)  
530 Effect of ocean acidification on cyanobacteria in the subtropical North Atlantic.  
531 *Aquat. Microb. Ecol.* **66**: 211–222.
- 532 Low-Décarie, E., Fussmann, G.F., and Bell, G. (2014) Aquatic primary production in a  
533 high-CO<sub>2</sub> world. *Trends Ecol. Evol.* **29**: 223–232.
- 534 Marhaeni, B., Radjasa, O.K., Khoeri, M.M., Sabdono, A., Bengen, D.G., and Sudoyo, H.  
535 (2011) Antifouling Activity of Bacterial Symbionts of Seagrasses against Marine  
536 Biofilm-Forming Bacteria. *J. Environ. Prot. (Irvine, Calif.)* **02**: 1245–1249.
- 537 Martin, S., Rodolfo-Metalpa, R., Ransome, E., Rowley, S., Buia, M.-C., Gattuso, J.-P.,  
538 and Hall-Spencer, J. (2008) Effects of naturally acidified seawater on seagrass  
539 calcareous epibionts. *Biol. Lett.* **4**: 689–692.
- 540 Medina-Pons, F.J., Terrados, J., López-López, a., Yarza, P., and Rosselló-Móra, R.  
541 (2009) Evaluation of the 18S rRNA clone library approach to study the diversity of  
542 the macroeukaryotic leaf-epiphytic community of the seagrass *Posidonia oceanica*  
543 (L.) Delile. *Mar. Biol.* **156**: 1963–1976.
- 544 Meron, D., Atias, E., Iasur Kruh, L., Elifantz, H., Minz, D., Fine, M., and Banin, E.  
545 (2011) The impact of reduced pH on the microbial community of the coral *Acropora*  
546 *eurystoma*. *ISME J.* **5**: 51–60.
- 547 Michael, T.S., Shin, H.W., Hanna, R., and Spafford, D.C. (2008) A review of epiphyte  
548 community development: Surface interactions and settlement on seagrass. *J.*  
549 *Environ. Biol.* **29**: 629–638.
- 550 Morrow, K.M., Bourne, D.G., Humphrey, C., Botté, E.S., Laffy, P., Zaneveld, J., et al.  
551 (2014) Natural volcanic CO<sub>2</sub> seeps reveal future trajectories for host-microbial  
552 associations in corals and sponges. *ISME J.* **2**: 1–15.
- 553 Morrow, K.M., Moss, A.G., Chadwick, N.E., and Liles, M.R. (2012) Bacterial associates  
554 of two Caribbean coral species reveal species-specific distribution and geographic  
555 variability. *Appl. Environ. Microbiol.* **78**: 6438–6449.
- 556 Mouchka, M.E., Hewson, I., and Harvell, C.D. (2010) Coral-associated bacterial  
557 assemblages: current knowledge and the potential for climate-driven impacts. *Integr.*  
558 *Comp. Biol.* **50**: 662–674.

- 559 Orth, R.J., Carruthers, T.J.B., Dennison, W.C., Duarte, C.M., Fourqurean, J.W., Heck Jr,  
560 K.L., et al. (2006) A Global Crisis for Seagrass Ecosystems. *Bioscience* **56**: 987–  
561 996.
- 562 Paerl, H.W. and Huisman, J. (2009) Climate change: a catalyst for global expansion of  
563 harmful cyanobacterial blooms. *Environ. Microbiol. Rep.* **1**: 27–37.
- 564 Quast, C., Pruesse, E., Yilmaz, P., Gerken, J., Schweer, T., Yarza, P., et al. (2013) The  
565 SILVA ribosomal RNA gene database project: improved data processing and web-  
566 based tools. *Nucleic Acids Res.* **41**: D590–D596.
- 567 Ragazzola, F., Foster, L.C., Form, A.U., Büscher, J., Hansteen, T.H., and Fietzke, J.  
568 (2013) Phenotypic plasticity of coralline algae in a High CO<sub>2</sub> world. *Ecol. Evol.* **3**:  
569 3436–3446.
- 570 Ramette, A. (2009) Quantitative community fingerprinting methods for estimating the  
571 abundance of operational taxonomic units in natural microbial communities. *Appl.*  
572 *Environ. Microbiol.* **75**: 2495–2505.
- 573 Raulf, F.F., Fabricius, K.E., Uthicke, S., de Beer, D., Abed, R.M.M., and Ramette, A.  
574 (2015) Changes in microbial communities in coastal sediments along natural CO<sub>2</sub>  
575 gradients at a volcanic vent in Papua New Guinea. *Environ. Microbiol.* doi:  
576 10.1111/1462-2920.12729.
- 577 Ries, J.B. (2011) Skeletal mineralogy in a high-CO<sub>2</sub> world. *J. Exp. Mar. Bio. Ecol.* **403**:  
578 54–64.
- 579 Rungrassamee, W., Klanchui, A., Maibunkaew, S., Chaiyapechara, S., Jiravanichpaisal,  
580 P., and Karoonuthaisiri, N. (2014) Characterization of intestinal bacteria in wild and  
581 domesticated adult black tiger shrimp (*Penaeus monodon*). *PLoS One* **9**: e91853.
- 582 Russell, B.D., Connell, S.D., Findlay, H.S., Tait, K., Widdicombe, S., and Mieszkowska,  
583 N. (2013) Ocean acidification and rising temperatures may increase biofilm primary  
584 productivity but decrease grazer consumption. *Philos. Trans. R. Soc. Lond. B. Biol.*  
585 *Sci.* **368**: 20120438.
- 586 Sand-Jensen, K. (1977) Effect of epiphytes on eelgrass photosynthesis. *Aquat. Bot.* **3**: 55–  
587 63.
- 588 Sawall, Y., Richter, C., and Ramette, A. (2012) Effects of eutrophication, seasonality and  
589 macrofouling on the diversity of bacterial biofilms in equatorial coral reefs. *PLoS*  
590 *One* **7**: e39951.
- 591 Singh, S.P. and Singh, P. (2014) Effect of CO<sub>2</sub> concentration on algal growth: A review.  
592 *Renew. Sustain. Energy Rev.* **38**: 172–179.



- 593 Smith, A.M., Sutherland, J.E., Kregting, L., Farr, T.J., and Winter, D.J. (2012)  
594 Phylomineralogy of the coralline red algae: correlation of skeletal mineralogy with  
595 molecular phylogeny. *Phytochemistry* **81**: 97–108.
- 596 Steele, J.A., Countway, P.D., Xia, L., Vigil, P.D., Beman, J.M., Kim, D.Y., et al. (2011)  
597 Marine bacterial, archaeal and protistan association networks reveal ecological  
598 linkages. *ISME J.* **5**: 1414–1425.
- 599 Sweet, M.J., Bythell, J.C., and Nugues, M.M. (2013) Algae as reservoirs for coral  
600 pathogens. *PLoS One* **8**: e69717.
- 601 Teeling, H., Fuchs, B.M., Becher, D., Klockow, C., Gardebrecht, A., Bennke, C.M., et al.  
602 (2012) Substrate-controlled succession of marine bacterioplankton populations  
603 induced by a phytoplankton bloom. *Science* **336**: 608–611.
- 604 Thompson, F.L., Barash, Y., Sawabe, T., Sharon, G., Swings, J., and Rosenberg, E.  
605 (2006) *Thalassomonas loyana* sp. nov., a causative agent of the white plague-like  
606 disease of corals on the Eilat coral reef. *Int. J. Syst. Evol. Microbiol.* **56**: 365–368.
- 607 Uku, J., Björk, M., Bergman, B., and Díez, B. (2007) Characterization and Comparison  
608 of Prokaryotic Epiphytes Associated With Three East African Seagrasses. *J. Phycol.*  
609 **43**: 768–779.
- 610 Vega Thurber, R., Willner-Hall, D., Rodriguez-Mueller, B., Desnues, C., Edwards, R.A.,  
611 Angly, F., et al. (2009) Metagenomic analysis of stressed coral holobionts. *Environ.*  
612 *Microbiol.* **11**: 2148–2163.
- 613 Vizzini, S., Di Leonardo, R., Costa, V., Tramati, C.D., Luzzu, F., and Mazzola, A. (2013)  
614 Trace element bias in the use of CO<sub>2</sub>-vents as analogues for low-pH environments:  
615 Implications for contamination levels in acidified oceans. *Estuar. Coast. Shelf Sci.*  
616 **134**: 19–30.
- 617 Wahl, M. (1989) Marine epibiosis. I. Fouling and antifouling: some basic aspects. *Mar.*  
618 *Ecol. Prog. Ser.* **58**: 175–189.
- 619 Webster, N.S., Uthicke, S., Botté, E.S., Flores, F., and Negri, A.P. (2013) Ocean  
620 acidification reduces induction of coral settlement by crustose coralline algae. *Glob.*  
621 *Chang. Biol.* **19**: 303–315.
- 622 Weidner, S., Arnold, W., Stackebrandt, E., and Pu, A. (2000) Phylogenetic Analysis of  
623 Bacterial Communities Associated with Leaves of the Seagrass *Halophila stipulacea*  
624 by a Culture-Independent Small-Subunit rRNA Gene Approach. *Microb. Ecol.* **39**:  
625 22–31.

626 Witt, V., Wild, C., Anthony, K.R.N., Diaz-Pulido, G., and Uthicke, S. (2011) Effects of  
627 ocean acidification on microbial community composition of, and oxygen fluxes  
628 through, biofilms from the Great Barrier Reef. *Environ. Microbiol.* **13**: 2976–2989.

629 Wolf, C., Frickenhaus, S., Kiliyas, E.S., Peeken, I., and Metfies, K. (2013) Regional  
630 variability in eukaryotic protist communities in the Amundsen Sea. *Antarct. Sci.* **11**:  
631 1–11.

632 Yamamuro, M., Umezawa, Y., and Koike, I. (2004) Internal variations in nutrient  
633 concentrations and the C and N stable isotope ratios in leaves of the seagrass  
634 *Enhalus acoroides*. *Aquat. Bot.* **79**: 95–102.

Table 1: Carbon (C) and nitrogen (N) content in percentage dry weight, C:N ratio and epiphyte cover of the leaves of *E. acroides*, the number of bacterial and eukaryotic OTUs obtained through ARISA (SI text 1) and amplicon sequencing (Bacteria: 16S rRNA gene, Eukarya: 18S rRNA gene, SI text 1); values constitute mean +/- standard error where applicable; for the bacterial sequencing dataset Chao1 richness estimates are given in italics with lower confidence interval/upper confidence interval in brackets.

		N [%]	C [%]	C:N ratio	Epiphyte cover [%]	ARISA		Amplicon sequencing	
						Bacteria	Eukarya	Bacteria (16S)	Eukarya (18S)
Control	All ages	1.42 ± 0.08	28.87 ± 0.8	20.98 ± 0.71	18.08 ± 2.45	96.52 ± 3.27	86.14 ± 2.56	507.5 ± 31.45 <i>917.31 ± 49.13</i>	520.75 ± 84.25
	Youngest	1.98 ± 0.18	34.63 ± 2.21	17.89 ± 2.5	3.25 ± 3.25	111.25 ± 6.02	72.5 ± 3.8	585 <i>913.76 (822.52/1040.05)</i>	277
	2nd youngest	1.54 ± 0.1	29.68 ± 0.73	19.56 ± 0.88	14.3 ± 5.5	96.83 ± 7.49	86.67 ± 4.97	532 <i>991.53(863.52/1168.96)</i>	664
	3rd youngest	1.2 ± 0.04	28.07 ± 0.8	23.47 ± 0.42	22.33 ± 2.72	94.83 ± 2.4	93.17 ± 3.36	462 <i>984.34 (827.90/1207.66)</i>	561
	Oldest	1.19 ± 0.04	25.83 ± 0.89	21.87 ± 1.31	22.54 ± 3.56	86.4 ± 6.31	88 ± 4.69	451 <i>779.63 (680.69/921.19)</i>	581
Vent	All ages	1.41 ± 0.12	26.93 ± 1.8	19.96 ± 0.79	6.61 ± 1.6	135.95 ± 2.82	89.68 ± 2.71	619.67 ± 48.22 <i>1044.09 ± 187.94</i>	517.75 ± 23.24
	Youngest	2.05 ± 0.05	32.02 ± 0.42	15.68 ± 0.42	2.5 ± 1.5	139.6 ± 7.64	97.4 ± 7.06	594 <i>933.61 (839.61/1063.58)</i>	459
	2nd youngest	1.25 ± 0.21	24.87 ± 3.83	20.15 ± 0.74	8.7 ± 2.2	133.5 ± 5.94	87.33 ± 4.36	NA	553
	3rd youngest	1.07 ± 0.18	24.26 ± 3.91	22.65 ± 1.3	8.55 ± 3.84	133.83 ± 2.7	84.33 ± 3.19	552 <i>788.18 (718.60/886.80)</i>	502
	Oldest	1.29 ± 0.04	28.36 ± 0.1	22.04 ± 0.76	3.75 ± 0.75	140.5 ± 7.5	93.5 ± 7.5	713 <i>1410.49 (1240.43/1635.38)</i>	557
Total		1.41 ± 0.07	27.92 ± 0.97	20.48 ± 0.53	12.81 ± 1.77	408	329	2179 <i>3811.43 ± 142.83</i>	3928



