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# Bubble Transport Mechanism: Indications for a gas bubble-mediated inoculation of benthic methanotrophs into the water column

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## Abstract

The importance of methanotrophic microorganisms in the sediment and water column for balancing marine methane budgets is well accepted. However, whether methanotrophic populations are distinct for benthic and pelagic environments or are the result of exchange processes between the two, remains an area of active research. We conducted a field pilot

study at the Rostocker Seep site (Coal Oil Point seep field, offshore California, USA) to test the hypothesis that bubble-mediated transport of methane-oxidizing microorganisms from the sediment into the water column is quantifiable. Measurements included dissolved methane concentration and showed a strong influence of methane seepage on the water-column methane distribution with strongly elevated sea surface concentrations with respect to atmospheric equilibrium (saturation ratio ~17,000%).

Using Catalyzed Reporter Deposition Fluorescence In Situ Hybridization (CARD FISH) analysis, aerobic methane oxidizing bacteria (MOB) were detected in the sediment and the water column, whereas anaerobic methanotrophs (ANME-2) were detected exclusively in the sediment. Critical data for testing the hypothesis were collected using a novel bubble catcher that trapped naturally emanating seep gas bubbles and any attached particles approximately 15 cm above the seafloor. Bubble catcher experiments were carried out directly above a natural bubble seep vent and at a nearby reference site, for which an “engineered” nitrogen bubble vent without sediment contact was created. Our experiments indicate the existence of a “Bubble Transport Mechanism”, which transports MOB from the sediment into the water column. In contrast, ANME-2 were not detected in the bubble catcher. The Bubble Transport Mechanism could have important implications for the connectivity between benthic and pelagic methanotrophic communities at methane seep sites.

Keywords: gas vent, methane, benthic-pelagic bubble transport, methanotrophs

## 1. Introduction

After water vapor and carbon dioxide (CO<sub>2</sub>), methane is the most relevant greenhouse gas on earth, contributing about 20% to radiative forcing by well-mixed greenhouse gases on a century timescale (IPCC, 2013). Significant natural marine methane sources are mud volcanoes (Etiope and Klusman, 2002), hydrothermal systems (Schmale et al., 2012b), and

cold hydrocarbon seeps (Judd and Hovland, 2007). Due to global warming, flooded permafrost (Shakhova et al., 2010) and submarine Arctic gas hydrates (Berndt et al., 2014; Westbrook et al., 2009) are currently under debate as additional drivers for increasing marine methane emissions to the atmosphere.

Due to their high organic matter content, continental slope and marginal sea sediments have huge methane generation potential by biogenic or thermogenic processes (Judd and Hovland, 2007; Reeburgh, 2007). In areas where methane is present in sediment pore fluids or in the water column, it serves as an important energy and carbon source for a complex microbial community (Knittel and Boetius, 2009; Valentine, 2011). Depending on the availability of methane and either oxygen or sulfate, microbial methane oxidation in the marine environment is conducted by aerobic methane-oxidizing bacteria (MOB) and anaerobic methanotrophic archaea (ANME), respectively. Although ANME activity is more relevant in the sediment, MOB represent the dominant microbial methane sink in the water column. MOB can be subdivided into the three main groups, Type I, II and X (Hanson and Hanson, 1996). There are three phylogenetically distinct groups of ANME (ANME-1, 2, and 3), which are commonly associated with sulfate-reducing bacteria (Knittel et al., 2005). This so called “benthic methane filter” (Sommer et al., 2006) is very efficient and removes a major fraction of upward-migrating, dissolved methane in sediments by converting it into carbonates (Peckmann et al., 2001) and biomass (Nauhaus et al., 2007). In addition to sediment processes, the pelagic methane filter is dominated by MOB that decrease dissolved water column methane before it can enter the atmosphere (Reeburgh, 2007). In the water column, time-scales of microbial methane degradation below the wave-mixed layer are relatively short compared to ventilation to the atmosphere, preventing contribution from deep methane to atmospheric methane budgets (Rehder et al., 1999). However, methane bubbles largely bypass the benthic and pelagic microbial filter, enabling highly efficient transport of methane from the sediment towards the sea surface. Dissolution of bubbles rising from the seabed elevates

methane concentrations in the overlying water column (McGinnis et al., 2006; Schmale et al., 2010). The vertical distribution of such dissolved methane from gas plumes and hence the potential for atmospheric impact depends on (a) water depth, (b) initial bubble size and gas fraction therein, (c) bubble dissolution kinetics, (d) local oceanographic conditions such as currents and density stratification (Leifer and Patro, 2002; Schmale et al., 2010), and (e) microbial methane oxidation (Valentine et al., 2001). Recent studies in the water column surrounding hydrocarbon seeps indicated an elevated abundance of methanotrophic microorganism in the near field of gas bubble plumes (Schubert et al., 2006). The enhanced methane concentration in the seep-affected water column stimulates the activity of methane oxidizers (Valentine et al., 2001) and might lead to a rise in the abundance of methane-oxidizing microbes in the aging plume water as was shown by Crespo-Medina et al. (2014) in the Deepwater Horizon plume and by de Angelis et al. (1993) in an aged hydrothermal plume.

Laboratory (Zhou and Mopper, 1998), modeling (Sadhil and Johnson, 1983), and field studies (Leifer and Clark, 2002) show that bubbles vertically transport surface-active substances (surfactants). Surfactants are compounds with hydrophobic and hydrophilic characteristics, which readily attach to gas/water interfaces (Schaefer et al., 1998; Wan and Wilson, 1994). To this group of substances belong colloidal, surface active particles which can consist of mineral particles (clay), organic macromolecules (proteins, lipids, and hydrocarbons), and even bacteria and viruses (Wan and Wilson, 1994).

Laboratory studies have shown that the accumulation of microorganisms on bubble surfaces facilitates their transport through a sandy matrix (Powelson and Mills, 1998; Wan et al., 1994). In contrast to studies regarding the transport of colloids on gas/water interfaces within groundwater environments (e.g., Powelson and Mills, 1998; Wan and Wilson, 1994; Wan et al., 1994), the relevance of this transport mechanism for the marine environment is discussed only rudimentarily in the literature (Leifer and Judd, 2002; Schubert et al., 2006). The accumulation of particles at the sea surface/atmosphere interface and the controlling

mechanisms have long been the subject of different scientific investigations (e.g., Cunliffe et al., 2013; and references therein). In this interface, wind-driven surface-waves break and form bubble clouds. This well-studied mechanism sparges particles and surface active substances in the upper water column and transports them to the sea-surface microlayer (Bezdek and Carlucci, 1972; Wallace and Duce, 1978). This transport mechanism is hypothesized to influence microorganism abundance and microbial community diversity in the sea-surface microlayer (Cunliffe et al., 2013). Laboratory studies also show that bacterial enrichment at the sea-surface (the so-called bacterioneuston) affects air-sea gas exchange (Upstill-Goddard et al., 2003).

Studies at the sediment/water interface showed that resuspension (Shimeta et al., 2002) and submarine fluid discharge (Campbell et al., 2013) can transport microorganisms from the sediment into the water column and that these transport mechanisms have implications on the microbial community structure within the water column. Here, we hypothesize that a bubble-mediated transport mechanisms between the benthic and pelagic habitats represents an additional exchange process. More specifically, for an environment influenced by gas seepage, we hypothesize that gas bubbles transport methanotrophic microorganisms from the sediment into the water column, a process we termed the “Bubble Transport Mechanism”. This hypothesized mechanism would influence the water column methanotrophic community, thereby indirectly providing feedback mechanisms for dissolved methane concentrations in the water column with the potential to impact the sea/atmosphere methane flux. To test our bubble transport hypothesis we conducted a pilot field study at a natural gas bubble seepage area (informally named Rostocker Seep) in the Coal Oil Point seep field, offshore Southern California, USA. Key to this pilot study was the development of a bubble collection device (bubble catcher) that collected gas bubbles after they emerged from the seafloor together with any attached surfactant particles.

## 2. Material and Methods

### 2.1 Study site

The Santa Barbara basin (California, USA) contains one of the world's largest and best investigated hydrocarbon seep fields, the Coal Oil Point seep field (Allen et al., 1970; Clark et al., 2003; Hornafius et al., 1999; Fig. 1). The seep field is located in water depths between a few meters to 80 m, extending from near the beach to 3 km offshore (Hornafius et al., 1999; Leifer et al., 2010). Seepage is concentrated above the Coal Oil Point and South Ellwood Anticlines, which harbors Miocene age oil in the Monterey Formation (Fischer, 1977).

A wide diversity of seepage has been documented in the Coal Oil Point seep field from sporadic single bubble vents releasing gentle bubble emanations to intense high-flow vents that span a broad range of bubble sizes, although generally radius  $r > 1000 \mu\text{m}$  (Leifer, 2010). Mean seep gas composition is 87.5% methane, 5.1% ethane, 3.1% propane and 1.3% carbon dioxide along with trace gases of heavier hydrocarbons and hydrogen sulfide (Clark et al., 2000). Some emission sites are constant, others eruptive. A further dimension of complexity is the oiliness of most vents, leading to persistent and extensive sea surface oil slicks (Leifer et al., 2006).

Using biomolecular methods and biogeochemical techniques, the process of microbial anaerobic and aerobic  $\text{CH}_4$  oxidation has been identified clearly in sediments and in the overlying water column of the Coal Oil Point seep field (e.g., Kinnaman et al., 2010; Mau et al., 2012; Treude and Ziebis, 2010). Seep gas composition varies between different regions of the Coal Oil Point seep field. At the Brian Seep site (10 m water depth,  $34^\circ 24.109' \text{ N}$ ;  $119^\circ 49.917' \text{ W}$ ) captured gas consisted of 91% methane, 7%  $\text{CO}_2$ , 1% ethane, and 0.2% propane (Kinnaman et al., 2010).

Field measurements were made between 15 and 17 February 2012 at the Rostocker Seep site ( $34^\circ 24.230' \text{ N}$ ;  $119^\circ 50.438' \text{ W}$ ), which is located northeast of Campus Point and ~30 m northeast of Brian Seep in 10 m water depth (Fig. 1). The site is situated slightly north of the

Coal Oil Point Anticline, which strikes approximately east-west through the campus point (Fig. 1; Leifer et al., 2010). Rostocker Seep releases non-oily gas bubbles from primarily sandy sediments. This sediment type is distinct from most other deeper seep sites in the Coal Oil Point seep field, whose sediments are comprised of a mixture of sand and tar. Marine conditions were ideal for sampling, with wave heights of less than 30 cm and calm winds during the study. However, winds and waves during previous days were significant and led to limited water visibility of approximately 0.5-1 m during the study.

## 2.2 Bubble catching experiment

The key tool developed for this study was a bubble catcher (Fig. 2), whose design was different from previous instruments that have been used to collect gas bubbles for gas composition analysis (Huttunen et al., 2001; Pape et al., 2010). The main difference of this adapted catcher is the sample cylinder, which is pre-filled with sterile-filtered artificial seawater. This setup allows after recovery the collection of microorganisms that were transported into the bubble catcher sample cylinder over the runtime of the experiment.

The bubble catcher was constructed to collect gas bubbles close to the sediment-water interface, i.e., approximately 15 cm above the gas vent, which is the distance between seafloor and the inlet of the stop-cock 2, Fig. 2), and to transfer the bubbles together with attached surface-active substances (surfactants) into an acrylic glass sample cylinder (Fig. 3A). The cylinder was filled with sterile-filtered artificial seawater (filter size 0.2  $\mu\text{m}$ , salinity 35, total volume of the bubble catcher sample cylinder 8.8 L) prior to deployment. The transported surfactants (including bacterial cells) then accumulated in the bubble catcher at the water/gas interface upon bubble bursting (Fig. 3B).

The bubble catcher was lowered to the seafloor using a davit, and positioned above a gas vent by scuba divers. After deployment, the divers waited for about 2 minutes before opening the sample cylinder using stop-cock 2 (Fig. 2), to allow the settlement of particles



resuspended from the seafloor. The scuba divers and the bubble catcher were monitored continuously onboard during the experiments with a video system attached to the frame of the bubble catcher.

In total two studies were conducted. The first study was dedicated to catch gas bubbles and attached microbes from a natural gas vent (hereafter termed "BC vent"; Fig. 3A). A second study, conducted about 1 hour after the "BC vent" experiment at the same location, was directed to obtain a water-column background value of microbes collected and transported by bubbles that had not been in contact with sediment (hereafter termed "BC blank"). The setup of the "BC blank" experiments was developed to establish experimental condition similar to the "BC vent" experiment (Fig. 2). However, different to the "BC vent" experiment, the bubble catcher was positioned at about 50 cm distance from the gas vent and bubbles were created from a pressurized nitrogen tank through a 1/8" diameter tube whose outlet was placed between the seafloor and the funnel of the bubble catcher (Fig. 2). Pressure was calibrated in the laboratory and monitored during sampling to provide an artificial bubble stream that appeared qualitatively similar to the nearby natural vents. The experimental runtimes and gas volumes collected with the bubble catcher were quite similar (4.15 L gas in 95 min for "BC vent", 4.00 L gas in 75 min for "BC blank"). The longer runtime of the "BC vent" experiment can be explained by the transient bubble emission from the "BC vent" site that could not be considered in the "BC blank" experiment (see 3.1). On board the vessel, water samples from these two experiments were subsampled into sterile bottles as described below and the bubble catcher was rinsed carefully with sterile-filtered artificial seawater (see above) to collect all microbial cells out of the bubble catcher for quantitative analyses.

### 2.3 Gas bubble measurements

The gas bubble measurements were conducted about 2 hours after the "BC vent" experiment at the same gas vent. For bubble size and gas flux analyses we used the bubble

measurement system described in detail by Leifer (2010). The bubble measurement system consisted of a video camera (Super Seacam 6500, Deepsea Power and Light, San Diego, CA), back illumination facilitated by underwater LED lights, an aluminum framework, and blocking components that constrained rising bubbles to a known distance from the camera to minimize parallax errors. Scuba divers positioned the bubble measurement system on the seafloor, which was cabled to the boat for recording, and camera control by a custom video controller. Video analysis to derive the bubble size distribution was carried out according to (Leifer, 2010). Parallax size uncertainty was approximately  $\pm 5\%$  (Leifer et al., 2003).

#### 2.4 Water column and sediment sampling

Different types of water-column samples were taken directly above the bubble catcher sampling site. The sea-surface microlayer was sampled using a glass plate sampling technique with a framed wiper (Stolle et al., 2009). This technique collects water samples and particles therein of up to a 50  $\mu\text{m}$  thick film of the sea-surface microlayer. Approximately 100 single glass plate dips were needed to yield a total volume of 1 L. Mid-water column samples were collected with a 2-L sample bottle (LIMNOS, Finland) that was deployed manually on a rope to specific water depths. The bottle was closed with a drop weight at the respective depths during each haul. Divers collected near-seabed water samples using an empty 1-L PET bottle that was opened and filled a few centimeters above the sediment-water interface.

After water column work was finished, sediment samples were recovered close to "BC vent" spot (distance 10 cm) using steel plates (dimension 20 x 30 cm) that were pushed into the sediment by scuba divers. The plates were prepared with drilled holes (closed with plastic tape, 18 vertical holes, hole distance 1 cm) to allow underwater sediment sampling with plastic syringes (volume 5 ml). From the far side of the gas vent spot, the scuba divers removed the sediment, opened the plate holes, and sampled the sediment with the syringes

horizontally. Sediments to depths of up to only 12 cm were sampled because the sandy sediment was underlain by a gravel bed.

After collection in the field, samples for gas analyses and molecular biological studies were transferred directly into sample-specific containers onboard the research vessel (see details below).

## 2.5 Methane analyses

A 600 ml subsample from each water-column sample bottle was transferred into pre-evacuated 1100 ml glass bottles. The dissolved methane was extracted by the use of a vacuum degassing line and a subsample of the gas was transferred into 10 ml pre-evacuated crimp-top glass vials containing 4 ml of supersaturated salt solution (Schmale et al., 2012a). The vials were sealed with a butyl rubber septum and stored “upside-down” to prevent gas exchange with the atmosphere until analysis in the home laboratory. The methane concentration was determined using a gas chromatograph (Trace GC, Thermo Fisher Scientific Inc.) equipped with a flame ionization detector (average precision of  $\pm 3\%$ ).

For respective determinations in sediments, a 4-ml subsample of each sediment sample was transferred into 20-ml crimp-top glass vials containing 10 ml of 2.5% NaOH solution to stop microbial activity. Vials were sealed with a butyl rubber septum and stored inverted until analyzed in the laboratory. The methane concentration were determined as described for the water samples above.

## 2.6 Quantification of MOB and ANME

From each sediment sample,  $0.5 \text{ cm}^3$  were fixed following the protocol of Treude and Ziebis (2010). In the home laboratory, fixed samples were diluted (1:3000) with phosphate-buffered saline (PBS) and treated by mild sonication for 25 s with a Sonotrode probe (Bandelin electronic, Type UW200) at an amplitude of 20% (40W), cycle at 15% (0.15 s). An

aliquot of 1000  $\mu\text{l}$  was filtered on 0.2- $\mu\text{m}$  GTTP polycarbonate filters (Millipore). The filters were embedded in low-gelling-point agarose and frozen at  $-20^{\circ}\text{C}$  until further processing.

Water samples (including bubble catcher samples) were collected in 100-ml glass bottles. After fixation with formaldehyde (2% , max. 30 min.), a 50-ml sample from the water column and 100-ml sample from the "BC vent" and "BC blank" experiments were filtered on Nucleopore filters (Whatman, pore size 0.2  $\mu\text{m}$ ). Filters were stored at  $-20^{\circ}\text{C}$  until further processing.

For the quantification of active MOB and ANME, Catalysed Reporter Deposition Fluorescence In Situ Hybridization (CARD-FISH) followed by tyramide signal amplification was performed with filters from sediment and water samples according to the protocol of Pernthaler et al. (2002) except for the cell wall permeabilization step. For permeabilization of archaeal cell walls, agarose-embedded filters were incubated in PBS (pH 7.6) containing 0.5% (w/v) sodium dodecyl sulfate (SDS) for 10-15 min at room temperature, washed in MilliQ water, and then dehydrated in absolute ethanol. For permeabilization of bacterial cells, agarose-embedded filters were incubated with lysozyme solution (10 mg  $\text{ml}^{-1}$ ) and incubated for 60 min at  $37^{\circ}\text{C}$ . Binding of streptavidin-HRP to biotin-labeled oligonucleotide probes and tyramid signal amplification was realized using the TSA<sup>TM</sup> Kit #22 with HRP-streptavidin and Alexa Flour<sup>®</sup> 488 tyramide (Invitrogen, Karlsruhe) according to the manufacturer's instructions. The following probe/probe mixtures for ANME and aerobic MOB were used: (i) ANME 2-538 (Treude et al., 2007) specific for ANME-2 archaea (40% formamide, hybridization temperature of  $46^{\circ}\text{C}$ ), (ii) a mixture (1:1:1) of M( $\alpha$ )450 specific for type II methanotrophs (Eller et al., 2001), and M( $\gamma$ )84 and M( $\gamma$ )705 (Eller et al., 2001) specific for type I methanotrophs (20% formamide, hybridization temperature of  $46^{\circ}\text{C}$ ). ANME-2 probes were selected, because this type of ANME has been detected nearby to the sampling site in a previous study (Treude and Ziebis, 2010). MOB probes were pooled to ensure maximum detection of aerobic methanotrophs. All hybridized filters were counter stained with 4',6'-

diamidino-2-phenylindole (DAPI). Hybridized and DAPI-stained cells were counted with an epifluorescence microscope (x1000, Leitz Aristoplan). The relative standard deviation for DAPI and CARD-FISH analysis obtained by repetitive cell counting on the same filter were below 5% and 12%, respectively.

### 3. Results

#### 3.1 Bubble size measurements

The determination of bubble sizes distribution for the "BC vent" site was an important basis to calculate the bubble-mediated flux of microorganisms. Herein we present analysis of ~15 minutes imaging of gas ebullition from the "BC vent" site. Bubble emissions were transient, lasting for tens of seconds - far longer than the wave period (~6 sec) - followed by a period of inactivity lasting tens of seconds before re-activating. Scuba divers reported this emission pattern was widespread at Rostocker Seep during fieldwork. The bubble emission rate was ~20 bubbles per second. The bubble size distribution (Fig. 4) had two dominant emission modes centered at  $r = 2890$  and  $1810 \mu\text{m}$ , where  $r$  is the equivalent spherical radius. Both sizes were well fit by Gaussian functions ( $R^2 = 0.98$  and  $0.92$  for the larger, and smaller emission modes, respectively). The larger mode dominated the volume flux,  $V$ , which was well described ( $R^2=0.9952$ ) by a Gaussian function (not shown) with a peak at  $r = 3040 \mu\text{m}$ , while the surface area flux for this mode peaked at  $3000 \mu\text{m}$  ( $R^2=0.9915$ ).

#### 3.2 Methane concentrations

Methane concentration analyses were carried out to obtain information about methane distribution in the sediment and the seep-surrounding water column (Fig. 5 and 6).

The methane sediment vertical concentration profiles displayed a clear trend from elevated concentrations in the deep sediment towards relatively low concentrations at the sediment surface (Fig. 5A). Water-column methane concentrations surrounding Rostocker Seep

showed strong methane enrichment in the lower water column, decreasing towards the sea surface (Fig. 6A). Despite the decrease, surficial water concentrations were elevated strongly with respect to atmospheric equilibrium (saturation ratio ~17,000%) indicating a non-negligible dissolved methane flux into the atmosphere. Background water-column methane profiles upstream of Rostocker Seep (34°24.177' N; 119°50.445' W) were homogeneous throughout the vertical depth profile with concentrations similar to those for Rostocker Seep lower water column with a mean methane concentration of ~2000 nmol L<sup>-1</sup> (data not shown).

### 3.3 Abundance of methanotrophs

DAPI and CARD-FISH analyses were performed on water, sediment, and bubble catcher samples to quantify total microbial cell numbers as well as organism type (ANME-2 and MOB) responsible for methane consumption, and to test the existence of bubble-mediated transport of organisms from the sediment to the water column exists (Fig. 7).

Total cell abundance in the sediment (DAPI staining) increased with decreasing sediment depth showing a maximum of  $5 \times 10^{10}$  cells cm<sup>-3</sup> sediment at ~1 cm below the sediment surface (Fig. 5B). MOB (CARD-FISH) were most abundant in the sediment surface ( $1.46 \times 10^9$  cells cm<sup>-3</sup>, 2.9% of DAPI stained cells, Fig. 5B). ANME-2 archaea (CARD-FISH) were present only between 1-3 and at 8 cm sediment depths at numbers of between 1 and  $5 \times 10^8$  cells cm<sup>-3</sup> (0.1 to 1.7% of DAPI stained cells, Fig. 5B). Total cell numbers in the water column were on average 5 orders of magnitude lower compared to the sediment. The highest cell numbers ( $10^6$  cells cm<sup>-3</sup>) were found between 1 and 3 m water depth (Fig. 6B). MOB in the bottom water, in intermediate (5 m) water depths, and in the sea surface microlayer surrounding the Rostocker Seep, revealed similar abundances (~ $10^3$  cells cm<sup>-3</sup>, 0.2 to 0.4% of DAPI stained cells, Fig. 6). In contrast to these results, ANME-2 were not found in the water column.

In the BC experiments, MOB cells were found in the "BC vent" ( $1.23 \times 10^3 (\pm 7\%)$  cells  $\text{cm}^{-3}$ ) (0.4% of DAPI stained cells) but not the "BC blank" sample. No ANME-2 cells were detected in either of the two samples. DAPI analysis showed that the abundance of total cells collected during the "BC vent" experiment ( $3.42 \times 10^5 (\pm 4\%)$  cells  $\text{cm}^{-3}$ ) was about three times higher than the "BC blank" experiment ( $1.09 \times 10^5 (\pm 3\%)$  cells  $\text{cm}^{-3}$ ).

#### 4. Discussion

In this study, we confirmed the presence of ANME-2 and MOB in Rostocker Seep sediments, which is in accordance with findings by Treude and Ziebis (2010) for the nearby Brian Seep area. In their study the authors suggested that aerobic methane oxidation is active in the sediment surface layer and anaerobic methane oxidation in sediment layers deeper than 5 cm. Anaerobic methane oxidation was confirmed further by the presence of ANME-2 organisms, whose cell abundance was very similar to our study. In the present study, MOB showed a decreasing trend of abundance with depth, while ANME-2 organisms were present only between 1-3 and 8 cm sediment depth forming a distinct layer where anaerobic methane oxidation probably is occurring (Fig. 5B). The presence of MOB within and below the depth of ANME is surprising, because those two groups should be mutually exclusive given their opposite relationship with oxygen: whereas ANME are inactivated permanently by oxygen (Treude et al., 2005b), MOB depend on oxygen for metabolism. As described by Treude and Ziebis (2010) for Brian Seep, gas bubble migration through permeable sediments can drive pore-water convection that discharges sediment pore water into the ocean (Dando et al., 2000; Haeckel et al., 2007). By continuity, discharged sediment fluid must be replaced, driving a water flow into the sediment. Thus, complex pore water circulation may explain some of the unusual distribution of MOB. Alternatively, or in addition, strong winds and water turbulence at the seafloor, directly preceding the field sampling, could have disturbed previously established zonations of methanotrophic groups along with the sediments.

The bubble catcher experiment above the gas vent revealed the transport of MOB from the sediment into the water column in contrast to the blank experiment, which did not. The complete lack of ANME-2 organisms in the bubble catcher was also in good agreement with the water column findings (Fig. 6B). We hypothesize that the strong oxygen sensitivity of ANME organisms (Treude et al., 2005b) may not allow prolonged water-column survival time even if their exposure to oxygen during the runtime of the experiment and recovery of the water samples was relatively short with a maximum of 120 min. Consequently, the applied CARD-FISH technique would have failed to detect dead ANME-2 cells if present, because the targeted RNA quickly degenerates after cell death (Schippers et al., 2005).

Another aspect that might influence the bubble transport of microorganisms is their specific surface activity (Schaefer et al., 1998). If ANME have a poor surface activity, they might detach from bubbles within a few centimeters after exiting the sediment. Leifer and Culling (2010) observed in tank experiments how non-surface active sediment particles were lofted by bubbles just a few centimeters, before falling from the bubble's trailing edge. Lastly, differences in cell size, shape, or density and position in the sediment may impact the benthopelagic transport of microorganisms by gas bubbles. Notably, in the upper 1 cm of the sediment the abundance of MOB was one order of magnitude higher compared to ANME-2 (Fig. 5B).

The transport of water column-born microorganisms into the bubble catcher was evident from the "BC blank" experiment, which showed that about one third of the microorganisms were transported into the sample cylinder without having contact with the sediment. As fluid movement at the Bubble Catcher inlet was not monitored (e.g. by tinting the water within the funnel), the advection of microorganisms from fluids discharged from the seep sediment or by advection of bottom water into the bubble catcher sample cylinder cannot be excluded. We are confident, however, that bubble-mediated transport of cells should be far more efficient (because faster) compared to fluid movement. Therefore we conclude that microorganisms



found in the “BC blank” sample cylinder mostly were stripped from the water column by the artificial bubble stream. The fact that no MOB were detected in the "BC blank", despite their presence in the water column surrounding the Rostocker Seep area, indicated that MOB-catch from the short water column between the sediment and the bubble catcher (ca. 15 cm) was of minor relevance during the runtime of the experiments, especially given their low abundance in the water column (Fig. 6) compared to the sediment (Fig. 5).

With the following calculations, we estimated the MOB and the total cell transport rate from the sediment into the water column by the Bubble Transport Mechanism at the studied gas vent. A total gas volume of 4.15 ( $\pm 1\%$ ) L was collected during the "BC vent" experiment. According to our bubble measurements we assume that the total volume bubble flux was represented by bubbles with volumes of 0.12 ( $\pm 16\%$ )  $\text{cm}^3$  equivalent to  $r = 3040$  ( $\pm 5\%$ )  $\mu\text{m}$ , derived from the peak of the volume flux function, resulting in approximately 35,800 ( $\pm 17\%$ ) bubbles trapped during the "BC vent" experiment. The water sample volume of the "BC vent" experiment was 4.65 ( $\pm 1\%$ ) L (i.e. water volume that was left in the bubble catcher after the experiment). Based on the concentration of MOB in the "BC vent" water sample ( $1.23 \times 10^3$  ( $\pm 7\%$ ) MOB  $\text{cm}^{-3}$ ), we estimated that 160 ( $\pm 25\%$ ) MOB cells were transported per bubble from the sediment into the water. Using our DAPI cell counts from the “BC vent” experiment ( $3.42 \times 10^5$  ( $\pm 4\%$ ) cells  $\text{cm}^{-3}$ ), which were corrected by the counts of DAPI cells stripped from the water column ( $1.09 \times 10^5$  ( $\pm 3\%$ ) cells  $\text{cm}^{-3}$ , "BC blank") and applying the same approach as used in the calculation above, we calculated a bubble transport rate of 30,000 ( $\pm 25\%$ ) total cells per bubble. The experimental runtime was 95 min, leading to a transport rate of  $60 \times 10^3$  ( $\pm 25\%$ ) MOB cells  $\text{min}^{-1}$  and  $11 \times 10^6$  ( $\pm 25\%$ ) total cells  $\text{min}^{-1}$ . Given a continuous gas release from numerous vents at the Rostocker Seep site and a water column cell abundance between  $10^3$  and  $10^6$  cells  $\text{cm}^{-3}$  we propose that these bubble transport rates are sufficient to alter the pelagic microbial community composition.

However, independent from the existence of such Bubble Transport Mechanism, it should be kept in mind that our pilot study provides no specific information on the survival rate of benthic MOB in the water column. Molecular studies of benthic and pelagic MOB diversity at seeps along the California continental margin indicated rather distinct communities (Tavormina et al., 2008), suggesting that benthic MOB transported by the Bubble Transport Mechanism may not necessarily survive or remain active in the water column over extended time periods.

## 5. Conclusion

Our pilot field study on a shallow cold seep site with bubble emissions indicated the existence of a bubble-mediated mechanism that transports microorganisms (including methanotrophic bacteria) from the benthic into the pelagic environment. Assuming that the benthic MOB remain active after their transport into the water column, we hypothesize that this Bubble Transport Mechanism contributes to the pelagic methane sink by continuously inoculating the water column with methane-oxidizing microorganisms, adding to the local pelagic MOB population.

Our study was, however, insufficient to evaluate the relevance of the Bubble Transport Mechanism in relation to the abundance of pelagic methanotrophs or other microorganisms in the vicinity of a seep area. Thus, further studies are required to (1) obtain a statistically robust dataset, (2) to parameterize the sediment-water transport efficiency, (3) compare the benthic and pelagic methanotrophic communities on a phylogenetic level, and (4) test the ability of the benthic microorganisms to adapt to the water column environmental conditions. Moreover, the density and distribution of gas vents, the current strength, and the abundance of MOB in the upstream water are important factors that need elucidation to allow a budget approach assessment of the impact of the transported methanotrophic bacteria on the pelagic methane sink.

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#### Figure captions

Figure 1. Map of sonar-derived seepage activity in the Coal Oil Point (COP) seep field. Adapted from (Leifer et al., 2010a). Rostocker Seep (red dot) is at 34°24.230' N; 119°50.428' W.

Figure 2. The sketch shows the concept of the Bubble Transport Mechanism and the technical drawing of the bubble catcher. The bubble catcher is positioned on the seafloor above a gas

vent. Gas bubbles (gray circles) rise within sediment voids, collect microbial cells (red circles), and transport microbial cells from the sediment, through the water, into the bubble catcher. During the "BC vent" experiment (see text) the gas bubble stream was focused by a funnel (1, positioned ca. 2 cm above the seafloor) and transferred into the sample cylinder (3) through the stop-cock (2). The gas bubbles moved further upwards through the sterile-filtered artificial seawater (4) and released the attached surfactants at the gas-water interface within the cylinder. After the gas bubble sampling the stop-cock (2) was closed. The pressure compensation during the recovery of the bubble catcher was controlled by a relief pressure valve (6). Through stop-cock 2 and 5, the bubble catcher sample water was released and the inner part of the sample cylinder was rinsed. During seafloor deployments, three frame legs stabilized the sampling cylinder. The insert shows the set-up of the "BC blank" experiment, where gas bubbles were released from a pressurized nitrogen tank that was situated above the seafloor (7).

Figure 3. (A) Funnel and stop-cock of the bubble catcher during gas bubble sampling in the "BC vent" experiment. (B) Sample cylinder with gas-water interface. The scale bar is 100 mm.

Figure 4. Bubble emission size distribution,  $\Phi$ , and volume flux,  $V$ , as a function of radius,  $r$ , and Gaussian fits to emission modes of  $\Phi$ . Data key on figure.  $\Phi$  is the number of bubbles (#) that cross an arbitrary horizontal plane per second in each (logarithmically-spaced)  $r$ -bin.

Figure 5. (A) Sediment methane ( $\text{CH}_4$ ) concentrations profiles from the Rostocker Seep. (B) Total cell abundance (DAPI, black circles), as well as aerobic methane oxidation bacteria (MOB, red diamonds) and anaerobic methanotrophs (ANME-2, blue triangles) cell abundance (CARD-FISH) in sediment samples.

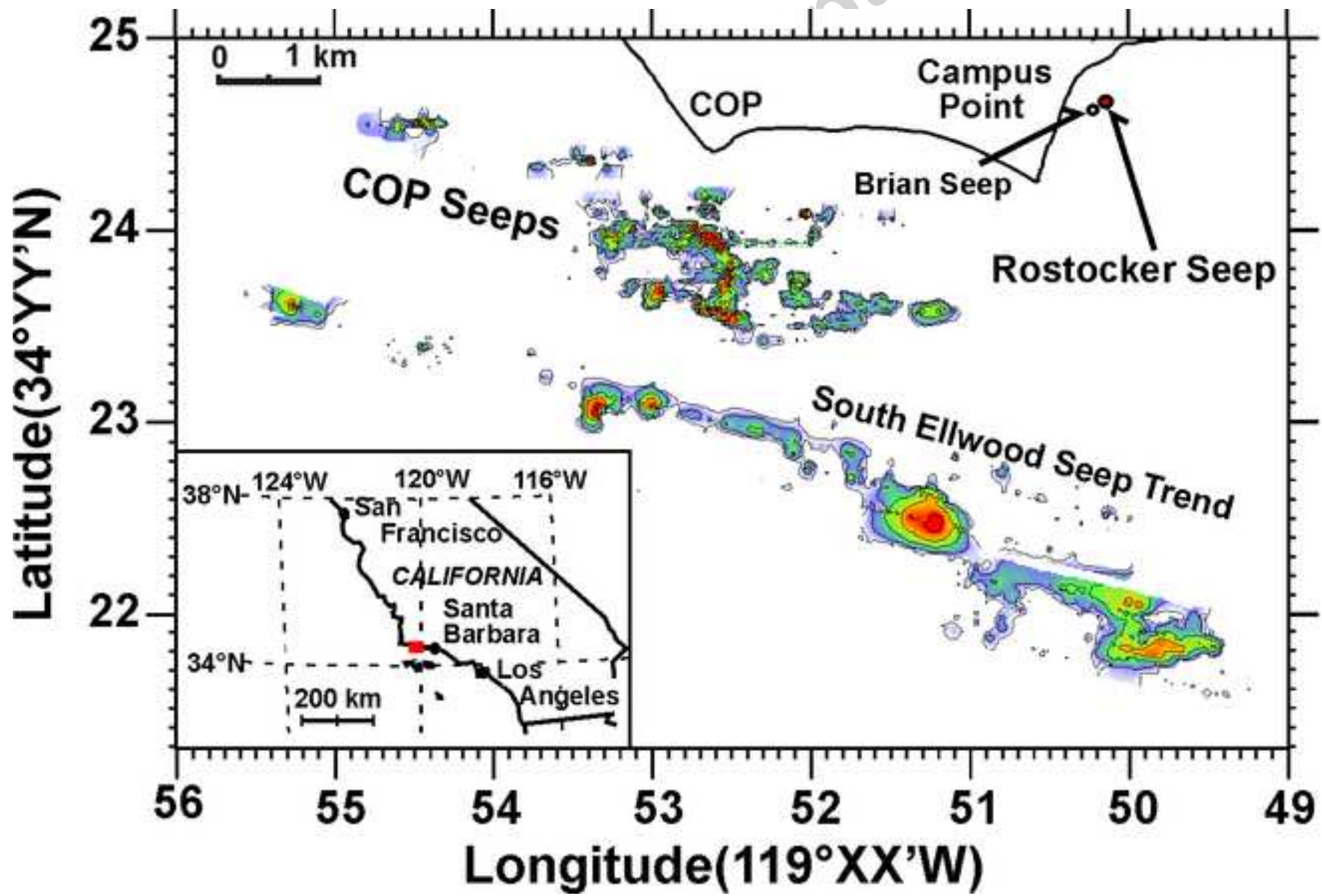
Figure 6. (A) Water column methane ( $\text{CH}_4$ ) concentrations at Rostocker Seep. (B) Total cell abundance (DAPI, black circles), as well as aerobic methane oxidizing bacteria (MOB, red diamonds, CARD-FISH) in water column samples. Anaerobic methanotrophs (ANME-2, CARD-FISH) were not detected in any water samples.

Figure 7. Epifluorescence images of MOB and ANME-2 cells visualized by 4', 6-Diamidin-2-phenylindol (DAPI, blue = A, C, E) and Catalyzed Reporter Deposition Fluorescence in Situ Hybridization (CARD-FISH, Alexa 488, green = B, D, F). MOB cells were identified applying CARD-FISH with the combined probes  $M(\alpha)450$ ,  $M(\gamma)84$ , and  $M(\gamma)705$ . ANME-2 cells were identified using probe ANME 2-538. (A, B) Aggregate of MOB cells obtained from the "BC vent" experiment. (C, D) MOB cells present in 1 cm sediment depth at the seep gas vent. (E, F) ANME-2 aggregate present in 3 cm sediment depth at the seep gas vent. The scale bar is 10  $\mu\text{m}$ .

#### Highlights

1. A new benthic-pelagic transport mechanism of microorganisms is hypothesized
2. A bubble transport hypothesis was tested using a new gas bubble-collecting device
3. Bubble-mediated transport rate of methanotrophs was quantified at a gas vent
4. The Bubble Transport Mechanism may influence the pelagic methane sink

Figure 1



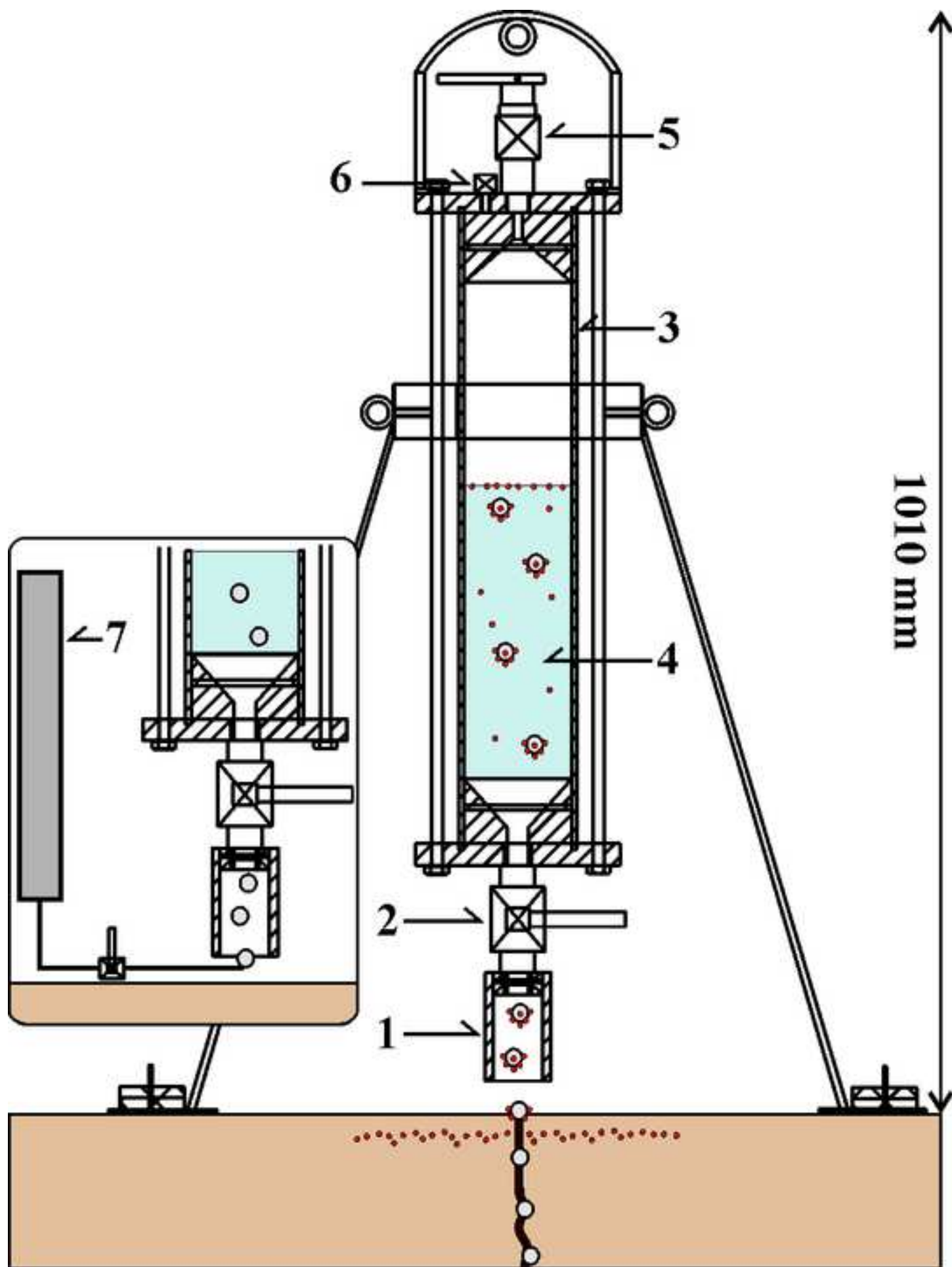


Figure 3

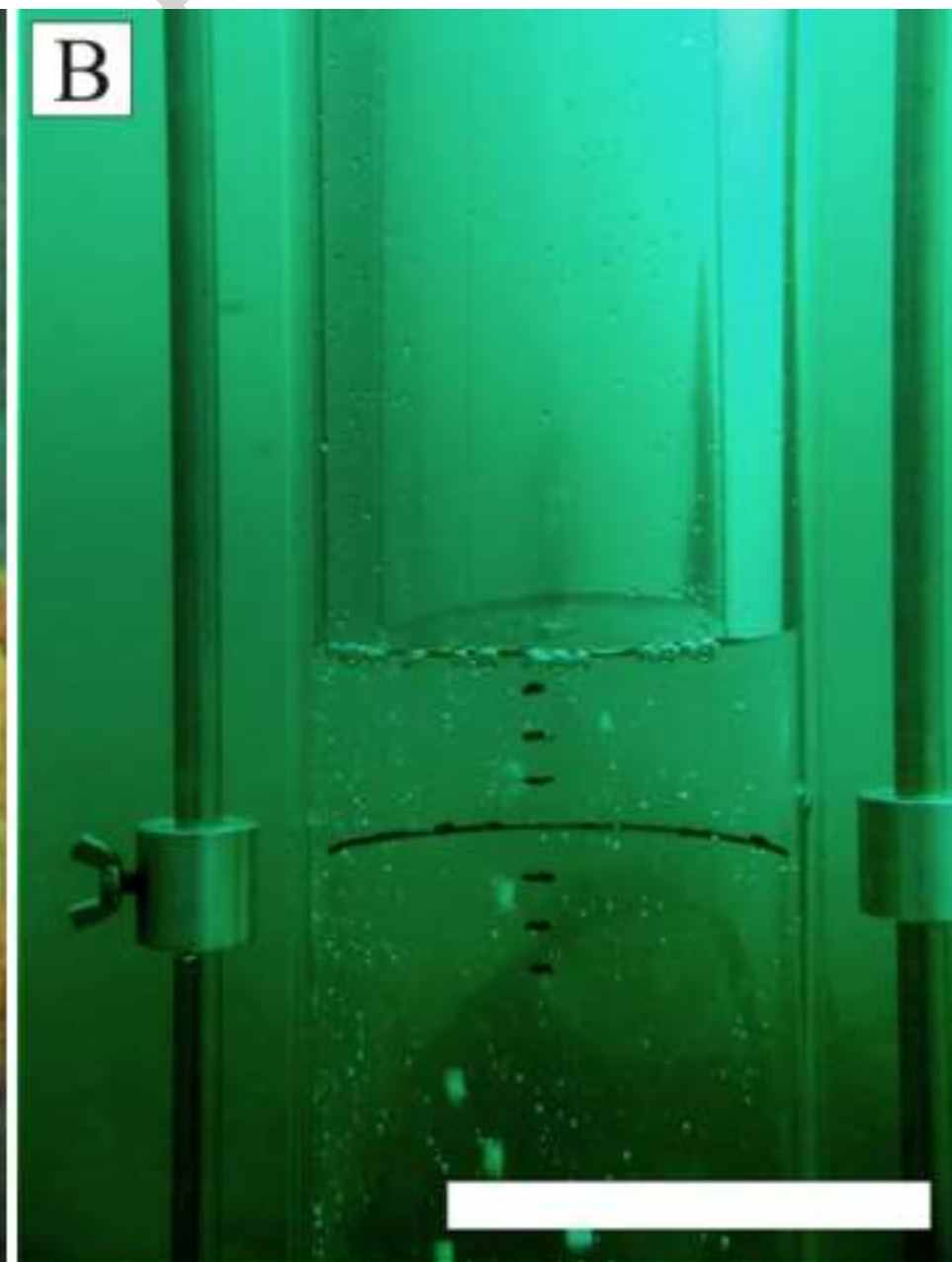
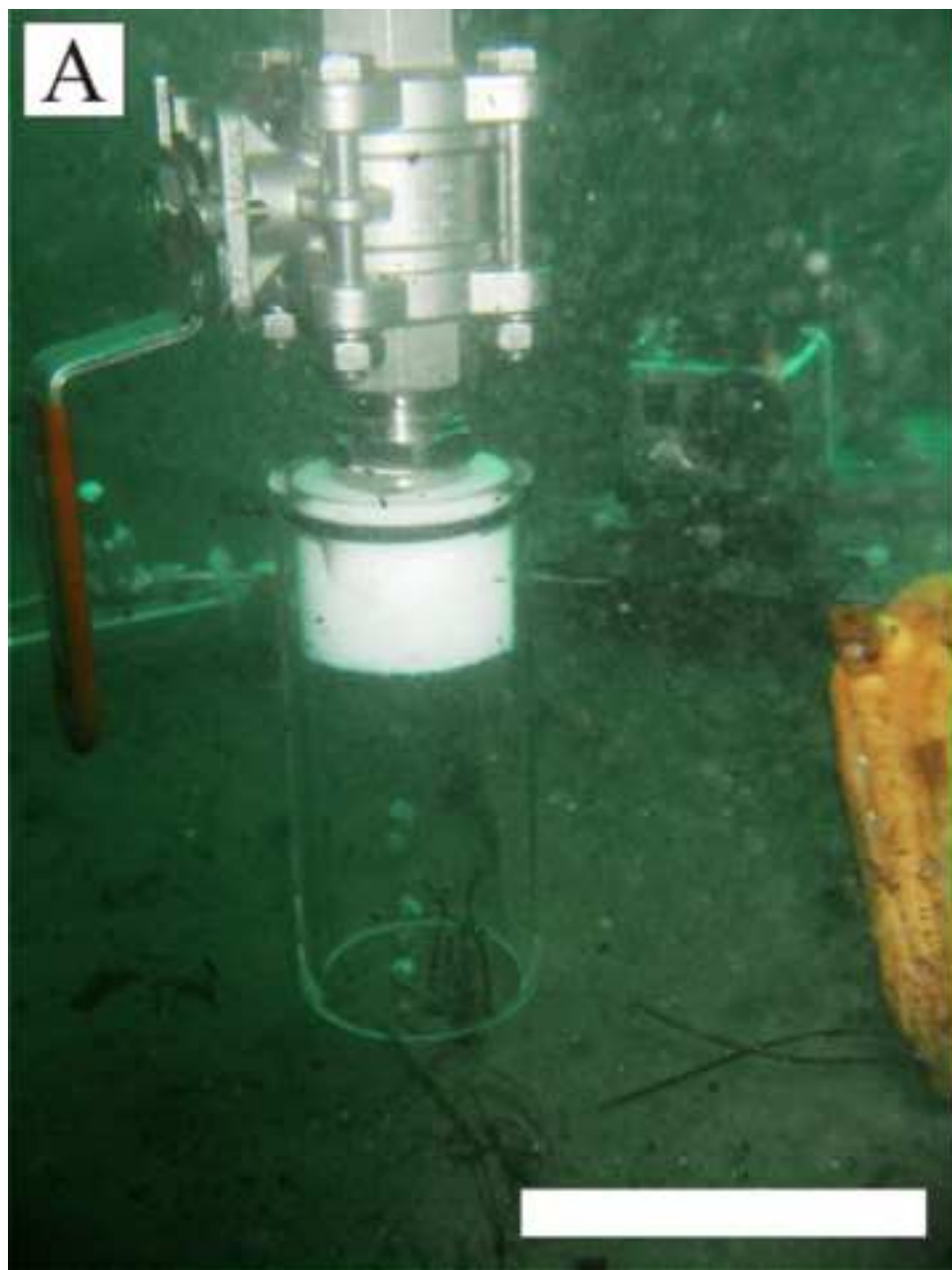


Figure 4

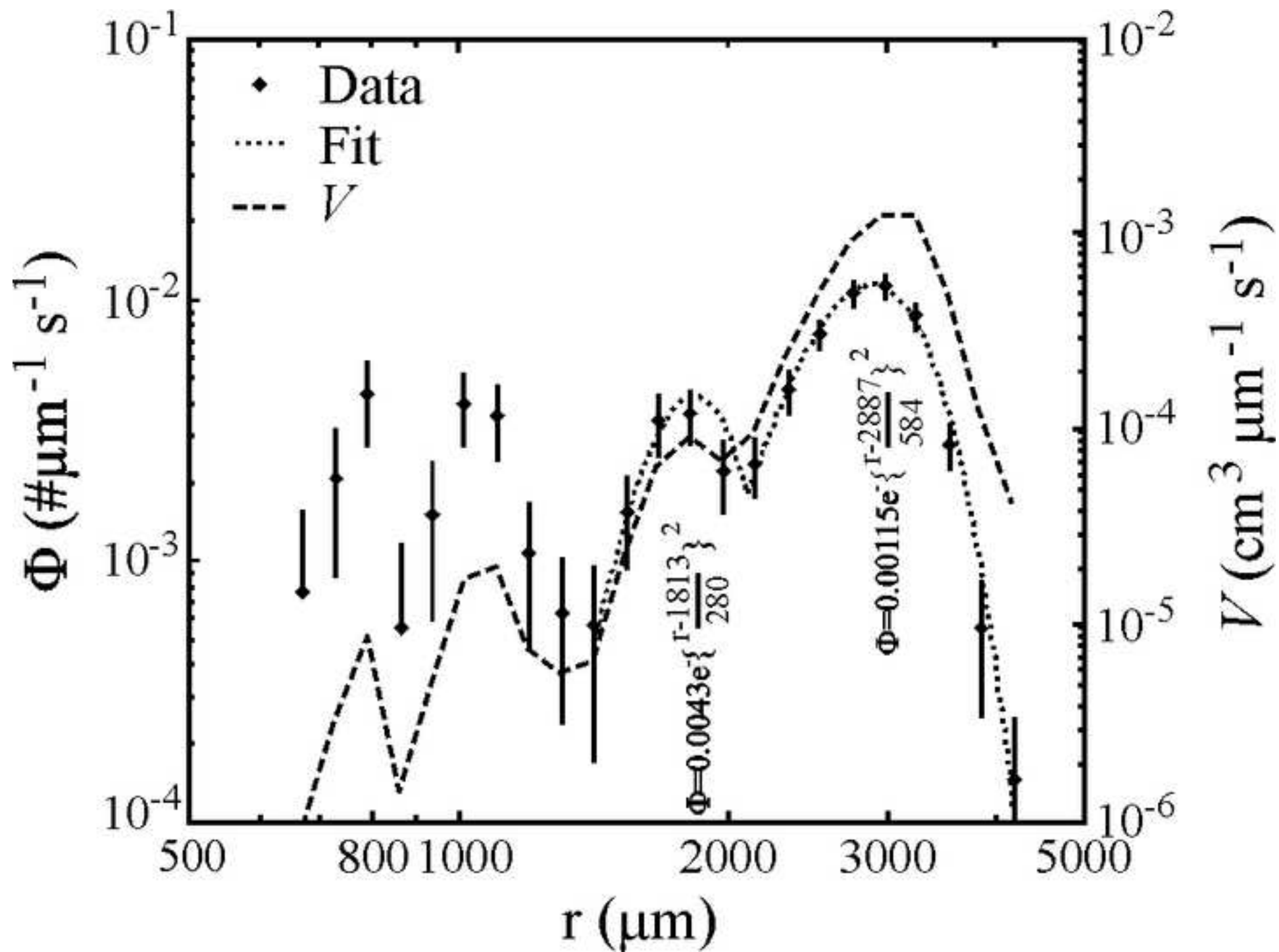




Figure 5

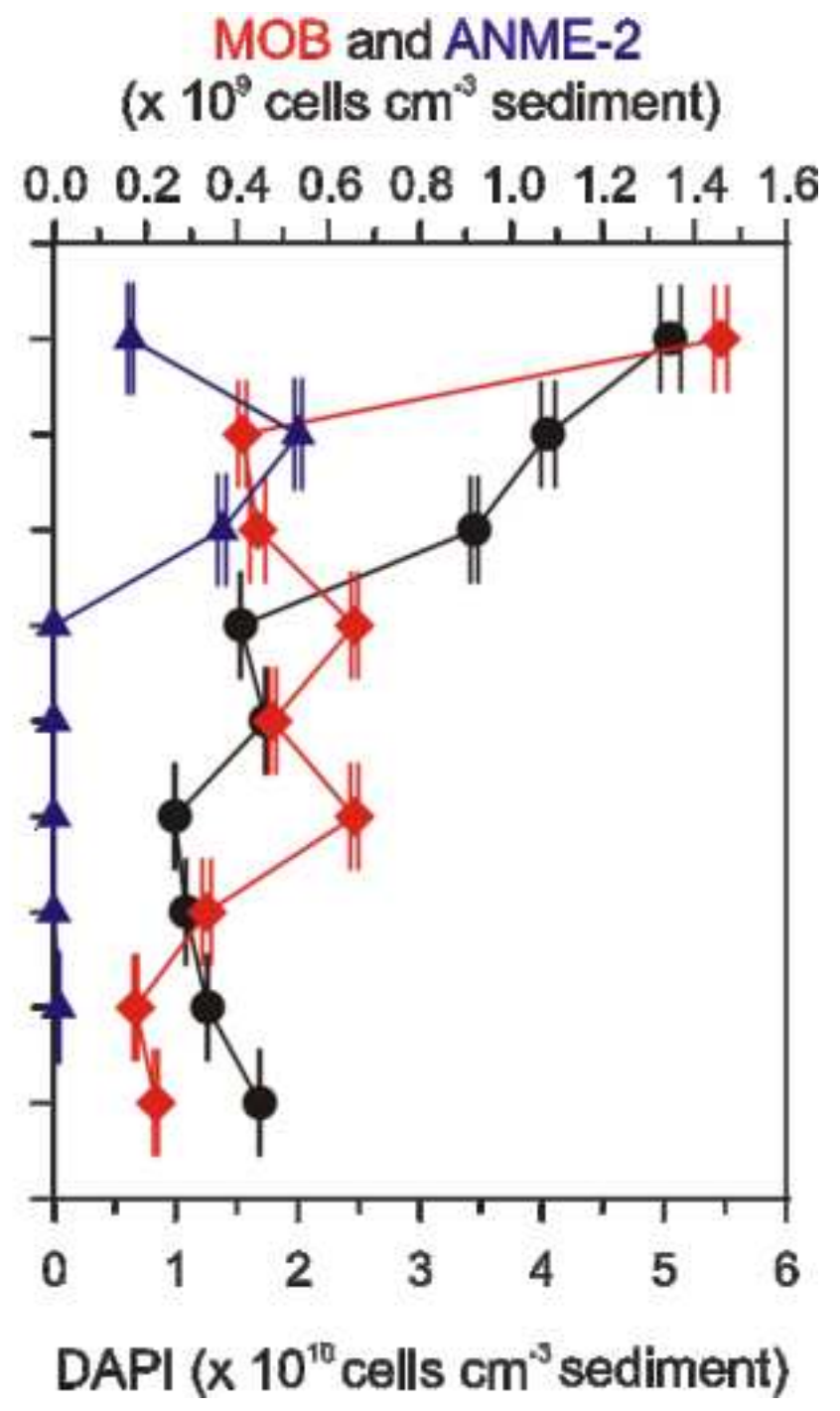
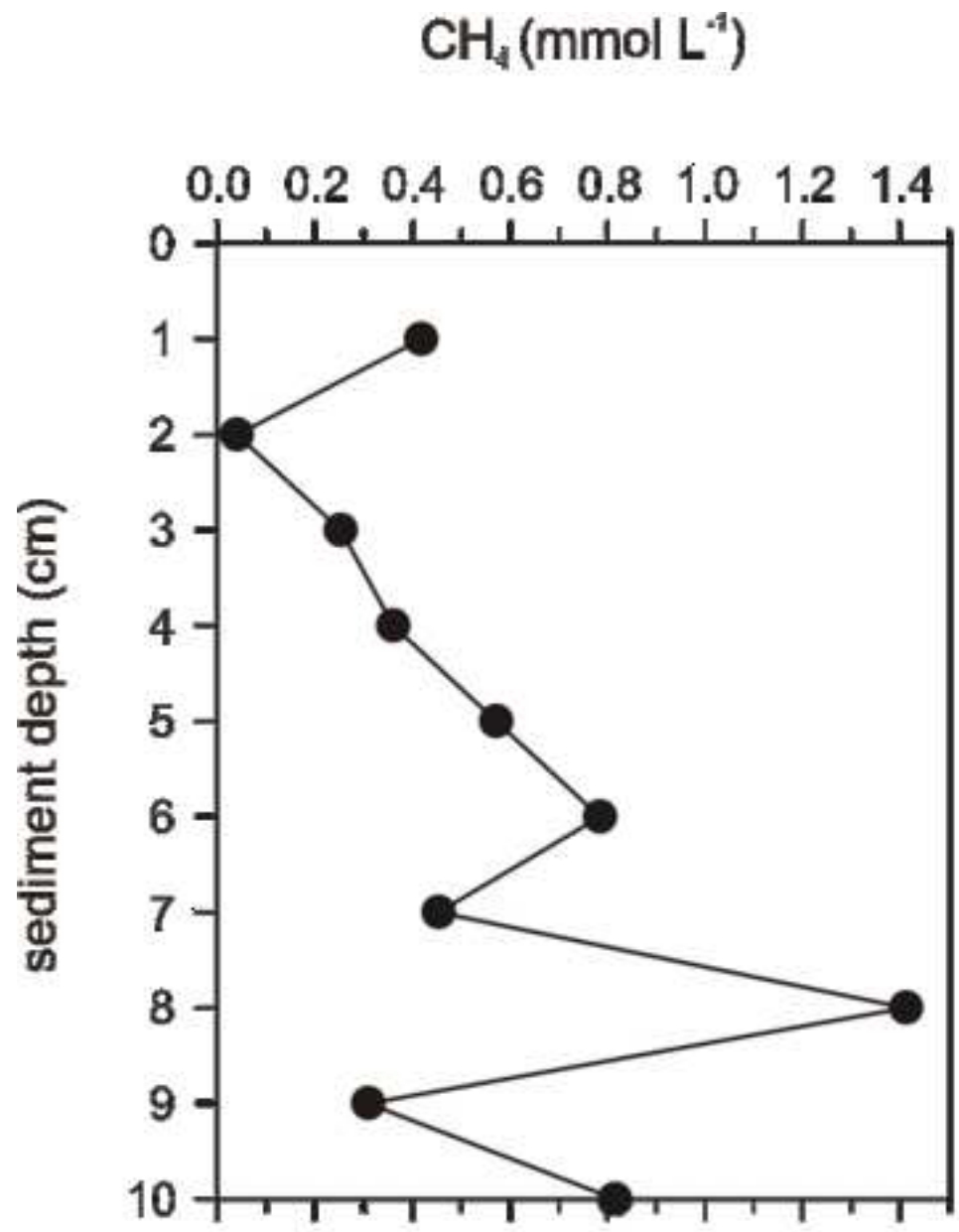


Figure 6

