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Assessing reference conditions and ecological status for lakes using subfossil diatoms

Draft report to the Environment Agency
Contract No. SC030103

A. Burgess, H. Bennion, G. Clarke

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Department of Geography
University College London
26 Bedford Way
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WC1H 0AP

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(Cover photos, from left to right: Llyn Bodlyn, Ben Goldsmith/ECRC © 2004; Lake Windermere, Ewan Shilland/ECRC © 2004; Fonthill Lake, Mike Hughes/ECRC © 2004

Executive Summary

This is the final report to the Environment Agency under an extension to the project, '*Development of a phytobenthos classification tool for lakes and lochs of UK (DALES– Diatom Assessment of lake and loch ecological status)*' (contract no. SC030103), funded by the Environment Agency (EA).This project forms part of the strategy for the implementation of the European Council Water Framework Directive (WFD), which requires reference conditions to be determined for all water body types including lakes.

This is a palaeoecological study, focusing on the analysis of diatom remains in sediment core samples from a set of lakes across England, Wales and Scotland. Diatoms are sensitive to water quality changes and shifts in the diatom community often correspond closely to changes in water quality and other biological elements. In summer 2004, short sediment cores were collected by the Environmental Change Research Centre (ECRC) on behalf of the EA from ~100 lakes in England and Wales as part of the WFD Lake Monitoring Phase 2 programme, and by SEPA from 27 Scottish lochs. Surface sediment samples from many of these cores are already being analysed as part of DALES, however the current project extension provides the necessary funding to enable core bottom material from ~50 lakes to be analysed. In selecting a subset of sites, preference was given to natural lakes and those sites not having previously been the subject of downcore diatom analysis. In addition, selected lakes were required to span a range of GB Lake Types (Phillips, 2004). For each site, the core bottom diatom assemblages are compared with those of the surface sediment diatom assemblages, to provide an estimate of floristic change from the reference condition. Diatoms were not preserved in all core bottom samples from the selected lakes, therefore mid-core samples were analysed from those sites where screening revealed some degree of change in diatom floristic composition between core top and bottom. Following screening for diatom preservation and the subsequent elimination of sites displaying diatom dissolution, core bottoms were analysed from 30 lakes (10 England, 7 Wales and 13 Scotland). Mid-core samples were also analysed from a subset of 20 of these sites (7 England, 5 Wales and 8 Scotland), bringing the total number of mid and bottom samples analysed to 50.

This project builds on existing palaeoecological work in the UK, in particular the '*Identification of reference lakes and evaluation of palaeoecological approaches to define reference conditions for UK (England, Wales, Scotland and Northern Ireland) ecotypes, WFD08*' (Bennion, 2004). Furthermore, this study complements the additional palaeoecological work recently undertaken on thirteen English lakes of conservation interest (Bennion *et al*., 2005) funded by English Nature (contract no. 13063) and the ongoing Site Condition Assessments of nineteen lakes in Welsh Special Areas of Conservation funded by Countryside Council for Wales (Goldsmith *et al*., in prep.) The data provide information on reference conditions that can subsequently be added to the large database of diatom reference samples analysed as part of the recently completed SNIFFER funded WFD08 Project (Bennion, 2004). The output of this project provides valuable information regarding reference and impact sites and enables broad assessment of ecological change at impacted sites.

The results of this study indicate that 17 of the 30 sites appear to have experienced significant floristic change in diatom species assemblages and 13 sites show minimal floristic change. The majority of changes appear to relate to increases in trophic status, although at some sites, floristic changes suggest increasing acidity (e.g. Llyn Bodlyn, Llyn Egnant and Llyn Ogwen) or decreasing acidity (Lochs Tormasad and Shnathaid).

A total of 10 out of the 30 lakes are thought to be good examples of reference lakes, as highlighted by the minimal change in their diatom species assemblages and low squared chord distance dissimilarity distances between core bottom and top samples. Tal-y-llyn Lake (Low alkalinity, Deep) is the only non-Scottish example of a reference lake in the current study. All other potential reference sites are Scottish lochs and examples are present for each lake type with the exception of Marl lakes, for which there are no Scottish examples in the current study. Loch Lagain and Loch Ascaig are examples of Peat lake reference sites, Lochs Skerrols, Ailsh and nan Gabhar are examples of Low alkalinity, Shallow reference lakes, Lochs Craggie and Loch Hope are examples

of Medium alkalinity, Deep reference lakes, Lochan Lùnn Dà-Bhrà is an example of a Medium alkalinity, Shallow reference lake and Loch Kinnabus is the only High alkalinity, Deep lake reference site. At some reference lakes e.g. Loch Kinnabus, Loch Hope, Loch Craggie and Lochan Lùnn Dà-Bhrà, there may be early warning signs of slight increases in trophic status. It is suggested that the water chemistry and ecology of these sites is monitored closely over the coming years, to ascertain whether a shift from reference conditions is occurring.

A further eight lakes may be potential reference lakes based on the relatively low degree of floristic change observed in the cores- Loch na Moracha (Peat, Deep), Loch Shnathaid (Peat, Deep), Thirlmere Reservoir (Low alkalinity, Deep), Loch Borralan (Low alkalinity, Shallow), Loch Tormasad (High alkalinity, Deep), The Mere, Ellesmere (High alkalinity, Deep), Shear Water (High alkalinity, Shallow) and Llyn Coron (High alkalinity, Shallow), although further investigation is necessary to confirm their status. One limitation of the current study is that none of the cores are dated and for some lakes where sediment cores are short, core bottom samples may not represent true ‘reference’ samples. This is of particular concern at the High alkalinity sites which lie in productive catchments (e.g. The Mere, Ellesmere, Shear Water and Llyn Coron) where sediment accumulation rates are expected to be high. The lack of a chronology is of less concern where the diatom assemblages remain stable throughout the core but it becomes a greater limitation when interpreting the data from sites that exhibit floristic change, as we have no estimate of the time at which the changes occurred. Sediment accumulation rates are site specific and it is unlikely that all cores cover comparable time periods. Longer cores would need to be collected and radiometric or spheroidal carbonaceous particle (SCP) dating carried out to provide a more detailed assessment of the nature and timing of ecological changes at these sites over longer timescales. In addition, sites such as Loch Borralan, where no change between top and bottom samples was seen but where the assemblages were comprised of many nutrient-tolerant diatom taxa, may represent naturally meso-eutrophic reference sites and further analysis is advised. Analysis of remains of other biological elements preserved in lake sediment cores, such as plant macrofossils, cladocera and chironomids, would enable ecological reference conditions to be defined in a more holistic way than can be achieved using the diatom record alone, and would provide valuable information on changes in ecological structure and function (e.g. Sayer *et al*., 1999; Bennion, 2001).

In conclusion, this project illustrates that relatively low-resolution analysis of lake sediment cores and the application of simple techniques such as dissimilarity scores and ordination analyses to palaeoecological data can provide valuable information for defining ecological reference conditions and assessing deviation from the reference state at impacted sites. This information aids implementation of the Water Framework Directive at the national level.

Key words: diatoms, lakes, palaeolimnology, palaeoecology, ecological status, reference conditions, Water Framework Directive.

List of Contributors

Amy Burgess, Helen Bennion, Gina Clarke
Environmental Change Research Centre
Department of Geography
University College London
26 Bedford Way
London
WC1H 0AP

Report written and edited by Amy Burgess and Helen Bennion

aburgess@geog.ucl.ac.uk
hbennion@geog.ucl.ac.uk

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Table of Contents

3. RESULTS AND DISCUSSION

APPENDICES

List of Figures

Page

List of Tables

Page

1. INTRODUCTION

1.1 Study rationale and objectives

There is a pressing need for simultaneous progress along several fronts in gaining understanding of lakes in the United Kingdom (UK). Specifically Annex II of the European Council Water Framework Directive (WFD) requires the identification of candidate reference lakes; and for Annex V, the development of tools for determining reference condition and classifying status (European Union, 2000). Furthermore, the European Council Habitats and Species Directive require the setting of conservation objectives, which may in turn lead to, the need for restoration targets for lakes. For lakes, the use of palaeolimnological techniques has the potential to contribute to the delivery of these requirements. Consequently, a number of recently completed and ongoing projects employ palaeolimnological methods to identify reference lakes, describe reference conditions and assess ecological status of UK lakes. For example, the project, '*Identification of reference lakes and evaluation of palaeoecological approaches to define reference conditions for UK (England, Wales, Scotland and Northern Ireland) ecotypes, WFD08*' (Bennion, 2004) involved the analysis of 219 core top and bottoms to identify reference lakes. The project, '*Lake Monitoring to support Environment Agency Water Framework Directive intercalibration exercise and classification tool development, and CCW Site Condition Assessment - Phase 2*' programme involves the analysis of the tops and bottoms of cores from 19 lakes in Wales.

Sediment cores were taken by the ECRC on behalf of the EA from ~100 lakes in England and Wales in 2004 as part of the '*Lake Monitoring to support Environment Agency Water Framework Directive intercalibration exercise and classification tool development, and CCW Site Condition Assessment - Phase 2*' programme. Approximately 50 of these lakes had not previously been the subject of down-core diatom analysis. Short cores were also collected by SEPA from 27 Scottish lochs in summer 2004, approximately 20 of which have never been subjected to palaeoecological analysis. The ongoing project '*Development of a phytobenthos classification tool for lakes and lochs of UK, DALES'* includes the analysis of only the surface sediment samples from a number of these sediment cores. The current project extension provides funding to enable analysis of core bottoms from a subset of ~50 sites across England, Wales and Scotland.

This project utilises the top and bottom approach (see Background). This is a low-resolution technique useful for broadly assessing the degree of change at large numbers of sites. Where floristic change was exhibited between the core top and bottom samples, an additional mid-core sample was analysed, thus enhancing the top and bottom approach. In the absence of core chronologies, the timing and rate of change cannot be established, although the additional mid-core samples do provide further insight into the point of change.

The objectives of the project were to:

1. Analyse subfossil diatoms in core bottoms from ~50 sites across England, Wales and Scotland
2. Compare the bottom and mid-core diatom assemblages to those of the surface sediment diatom assemblages already being analysed as part of DALES, to provide an estimate of floristic change from reference conditions
3. Perform detrended correspondence analysis (DCA) and a dissimilarity measure to assess the nature and degree of change in the diatom assemblages
4. Provide information on reference conditions, verify choice of reference lakes and assess shifts in ecological status at impacted sites.

1.2 Background

The European Council Water Framework Directive (WFD) 2000/60/EC came into force in 2000 (European Union, 2000) and aims to achieve good ecological quality in all relevant waters within 15 years. It requires that biological, hydromorphological and chemical elements of water quality should be assessed by the degree to which present day conditions deviate from those expected in the absence of significant anthropogenic influence, termed reference conditions. The WFD specifically requires the determination of reference conditions for different waterbody types in order to identify sites of High status, i.e. where the various elements correspond totally or almost totally to undisturbed conditions. The four categories of Good, Moderate, Poor and Bad status are defined according to the degree of deviation from the reference state. In the absence of long-term data, the WFD states that reference conditions based on modelling may be derived using hindcasting methods, and palaeolimnology is given as one such technique (Pollard & Huxham, 1998; European Union, 2000).

The study of the sediment accumulated in a waterbody can provide a record of its past biology and chemistry, a science known as palaeoecology. Diatoms (*Bacillariophyceae*: single-celled, siliceous algae) are commonly used in such studies because they are sensitive to water quality changes and are, therefore, good indicators of past lake conditions such as lake pH (Battarbee *et al.,* 1999; 2001) and total phosphorus (TP) concentrations (Hall & Smol, 1999). Of the biological elements relevant to the WFD, diatoms represent components of both the phytoplankton and phytobenthos, but importantly shifts in the diatom community often correspond closely to changes in other biological groups (e.g. Kingston *et al*., 1992). The diatom record is a potentially useful tool, therefore, for assessing water quality and defining lake reference conditions, both chemical and ecological (e.g. Kauppila *et al*., 2002; Bennion *et al*., 2004). The use of palaeoecological techniques for determining reference conditions and assessing ecological change in lakes is well established with many examples of their application to aquatic management and conservation (e.g. Battarbee, 1999; Stoermer & Smol, 1999; Bennion *et al*., 1996, 2004).

The top and bottom approach involves the analysis of only two samples per site from a sediment core (Cumming *et al*., 1992) and has been successfully applied by the US Environmental Protection Agency's (USEPA) Environmental Monitoring and Assessment Program for Surface Waters (EMAP-SW) (Dixit *et al*., 1999) and in Canada to infer changes in southeastern Ontario lakes (Reavie *et al*., 2002). The approach makes the assumption that the top and bottom samples represent the present day and reference conditions, respectively. The analysis of reference samples in this way removes the problem inherent in spatial-state schemes in which the lakes have been subject to different pressures and varying degrees of impact. For the UK, it is generally agreed that approximately 1850 AD is a suitable date against which to assess impacts for lakes as this represents a period prior to major industrialisation and agricultural intensification (Battarbee, 1999; Fozzard *et al*., 1999). However, it is accepted that aquatic systems have been subjected to anthropogenic impacts over much longer time-scales than simply the last 100-150 years and, therefore, our reference conditions are unlikely to equate to the natural or pristine state. The diatom data from the surface (uppermost 0.5 or 1 cm) sample of each core is used to provide information on the current diatom assemblages of the lakes since this represents the last few years' accumulation of diatoms deposited from a variety of habitats within the lake. The bottom sample of the core is taken to represent the reference conditions, although in the absence of core chronologies, there are uncertainties as to whether the bottom samples represent true reference samples.

1.3 Report structure

The report is comprised of four main sections – 1) introduction, study rationale and background; 2) methodology; 3) presentation and discussion of the results; and 4) a summary of the findings and suggestions for further work.

2 Methods

2.1 Site selection

Short cores were collected by SEPA from 27 Scottish lochs in summer 2004, approximately 20 of which had never been subjected to palaeoecological analysis. Short cores were also collected by the ECRC on behalf of the Environment Agency from ~100 lakes in England and Wales in summer 2004, of which approximately 50 lakes had not previously been the subject of downcore diatom analysis. Following elimination of most artificial reservoir sites, a total of 44 lakes in England, Wales and Scotland were short-listed for inclusion in this project. Cores from five lakes were deemed too short and were excluded immediately, and following further screening, a subset of 30 sites (10 English, 7 Welsh and 13 Scottish) underwent analysis. The site characteristics are given in Table 1 and details of the cores and analysis undertaken on each site are given in Table 2. The sites represent a range of lake types in the GB Lake Typology (Phillips, 2004), including Peat (P) (4 sites), Low alkalinity (LA) (11 sites), Medium alkalinity (MA) (6 sites), High alkalinity (HA) (8 sites) and Marl (1 site) systems, and with examples of both shallow (Sh) and deep (D) waters (11 and 19 sites, respectively) (see Tables 1 and 2). Lakes of low alkalinity are best represented. Most of the lakes are circumneutral to alkaline and even the low alkalinity waters are only mildly acid. With the exception of Llyn Bodlyn, Llyn Egnant and Llyn Ogwen which are at altitudes >200 m, the study lakes lie in relatively productive, lowland catchments and hence eutrophication is likely to be the key pressure. Further details of the sites and their chemistry are given in Table 1.

2.2 Field and laboratory methods

A sediment core was taken in the summer of 2004 from the deepest part of each lake using a Glew gravity corer which collects short cores of typically 20-40 cm in length. It might be expected, therefore, that at very productive sites, the short cores would not extend back as far as 100 years. Based on previous palaeolimnological research of dated sediment cores (e.g. Bennion, 2004), it is estimated that for Low alkalinity lakes, where production is lower and in turn, sediment accumulation rates slower, a sediment depth of 20-30 cm dates back to ~1850. In higher alkalinity sites, where sediment accumulation rates are faster due to higher productivity, a sediment depth of ~50 cm dates back to 1850, and in the highly productive Cheshire-Shropshire Meres, the depth required to extend back to approximately 1850 is typically ~60-70 cm. All cores were extruded in the field at either 0.5 cm or 1.0 cm intervals (depending on expected sediment accumulation rate).

Cores from five of the 44 lakes were deemed too short to provide reasonable core bottom samples to approximate reference conditions. This resulted in a suite of 39 short-listed sites (see Table 2). Surface sediment samples had previously been prepared as part of '*Development of a phytobenthos classification tool for lakes and lochs, DALES*', hence slides for diatom analysis were prepared from the bottom of each of the 39 cores (see Table 2), using standard methods (Battarbee *et al*., 2001). Screening of the slides revealed that nine sites had extremely poor diatom preservation in the bottom samples, thus excluding these sites from analysis. Therefore, diatom analysis was carried out at only 30 lakes. Screening also revealed that some sites exhibited greater shifts in floristic composition than others. Therefore, for a subset of 20 of the 30 lakes, mid-core samples were also prepared for diatom analysis to allow the general trend in water quality to be determined. This resulted in a total of 50 samples for diatom analysis as part of this project, comprising 20 mid-core samples and 30 core bottom samples. At least 300 valves (siliceous component of the cell wall bearing the taxonomic features) were counted from each sample using a Leitz research microscope with a 100x oil immersion objective and phase contrast. Principal floras used in identification were Krammer & Lange-Bertalot (1986-1991). All slides are archived at the ECRC and the data are stored in the Amphora database.

2.3 Data analysis

All diatom data were expressed as percentage relative abundance, and were screened and harmonised prior to data analysis. The 50 samples analysed as part of the current project were harmonised with the surface samples of the 30 study lakes. The full dataset of 80 samples (30 cores; with 2 samples from 10 cores and 3 samples from 20 cores) comprised 398 diatom taxa. The most common 203 taxa (occurring at >1% in >2 samples) are listed in Appendix 1. Summary diagrams of the diatom assemblages in the cores from each site (showing only those taxa present with a relative abundance of >5% in at least one sample) were produced for each lake type (the one Marl site is included in the HA diagram) using C^2 (Juggins, 2003) – see Figures 1 to 4.

The degree of floristic change between the bottom (reference) sample and the surface (and mid) sample analysed in each core was assessed using the squared chord distance dissimilarity coefficient (Overpeck *et al.*, 1985) implemented in the statistical software R (R Development Core Team, 2004). This is preferred to other dissimilarity measures as it maximises the signal to noise ratio, it performs well with percentage data and has sound mathematical properties (Overpeck *et al.*, 1985). The scores range from 0 to 2 whereby 0 indicates that two samples are exactly the same and 2 that they are completely different. Scores less than 0.29, 0.39, 0.48 and 0.58 indicate insignificant floristic change at the 1st, 2.5th, 5th and 10th percentile, respectively (Simpson, 2003). The 5th percentile (score <0.475) is used here to define sites with low floristic change between the bottom (reference) sample and surface (and mid) sample. The scores are plotted for each lake in the form of bar graphs in Figures 1-4, to show how dissimilar the surface and mid samples are from the bottom (reference) sample. The actual values are shown in Appendix 2. The vertical line in Figures 1-4 is drawn at a squared chord distance dissimilarity score of 0.475 to illustrate which samples fall above and below this critical value.

For each lake type [P, LA, MA and HA (including 1 Marl site)], detrended correspondence analysis (DCA) (Hill & Gauch, 1980) was performed using CANOCO version 4.5 (ter Braak & Smilauer, 2002) to assess the direction and magnitude of floristic change at each site. Only those 77, 127, 90 and 97 diatom taxa present with a maximum relative abundance of > 1% in > 2 samples, for P, LA, MA and HA lake types respectively, were included in the ordination analyses. The results are presented as biplots of axis 1 and 2 sample scores and species scores in Figures 5-8. Samples with similar scores on the two axes lie in close proximity, reflecting similar diatom composition. For each core, lines connect the samples in a series from core bottom to core top (see Appendix 2 for sample codes). The direction of the line indicates the direction of floristic change and its length is a measure of species turnover in Hill's standards deviation units (Hill & Gauch, 1980). For species codes see Appendix 1.

The results are detailed below for each lake type. For each site, the major species shifts are described, the degree of floristic change is presented and an interpretation of the floristic changes is given.

Table 1: Summary site characteristics of the 44 lakes short-listed for analysis (N.B. SEPA data missing)

GB Lakes WBID	Name	Altitude (m.a.s.l.)	Surface area (ha)	[1]Max depth (m)	[2]GB Lake type		pH	Cond (µS/cm)	Alk (mg/L)	TP (µg/L)	SRP (µg/L)	TN (mg/L)	Chla (µg/L)	Si (mg/L)
12578	Loch an Lagain	136	27.7		P	D								
15316	Loch na Moracha	4	36.4		P	D								
18113	Loch Shnathaid	4	23.2		P	D								
8945	Loch Ascaig	135	27.1		P	D								
11611	Loch Brora	25	66.5		P	D								
29184	Grasmere	61	60.7	20.8	LA	D	6.99	45	12.74	20	3	0.41	7	1.31
35561	Llyn Bodlyn	385	16.5	20.0	LA	D	6.40	30	2.25	14	8	0.55	1	0.55
38409	Llyn Egnant	420	13.9	14.2	LA	D	6.12	32	1.88	23	9	4.13	3	0.43
33803	Llyn Ogwen	300	38.5	2.7	LA	D	7.03	25	3.65	11	6	0.36	2	0.58
33730	Llyn Padarn	105	97.6	27.0	LA	D	7.46	45	8.70	19	8	0.38	6	1.28
36405	Tal-y-llyn Lake	85	50.7	3.5	LA	D	7.34	42	8.44	29	9	0.55	6	1.20
29021	Thirlmere Reservoir	178	313.3	24.0	LA	D	7.15	32	5.00	19	1	37.01	3	1.46
11355	Loch Borralan	142	47.0		LA	Sh								
26257	Loch Skerrols	25	26.1		LA	Sh								
1138	Loch Ailsh	154	105.2		LA	Sh								
2257	Loch nan Gabhar	1	16.6		LA	Sh								
5714	Loch Rangag	117	31.6		LA	Sh								
16530	Loch Gowan	156	18.2		LA	Sh								
29321	Coniston Water	46	470.5	36.0	MA	D	7.19	61	13.54	22	1	0.49	5	0.69
29233	Windermere	37	1435.9	32.8	MA	D	7.96	66	16.76	23	6	0.47	9	0.76
11642	Loch Craggie	166	54.2		MA	D								
2490	Loch Hope	4	638.3		MA	D								
18682	Loch Druidbeag	7	256.5		MA	D								
29222	Elter Water	53	18.2	7.7	MA	Sh	6.98	48	12.46	18	2	0.43	-	1.85
22395	Lochan Lùnn Dà-Bhrà	156	26.0		MA	Sh								
4974	Loch Syre	122	44.0		MA	Sh								
32538	Llyn Alaw	42	308.4	3.8	MA	Sh	7.72	186	41.77	37	14	1.52	11	1.43

GB Lakes WBID	Name	Altitude (m.a.s.l.)	Surface area (ha)	[1]Max depth (m)	[2]GB Lake type		pH	Cond (µS/cm)	Alk (mg/L)	TP (µg/L)	SRP (µg/L)	TN (mg/L)	Chla (µg/L)	Si (mg/L)
28386	Talkin Tarn	128	25.5	12.5	Marl	D	8.12	122	43.97	51	9	1.71	25	1.51
15551	Loch Tormasad	8	21.1		HA	D								
26944	Loch Kinnabus	77	43.7		HA	D								
34990	The Mere, Ellesmere	98	43.4	18.0	HA	D	7.86	282	113.80	954	766	1.48	16	1.93
25899	Ardnave Loch	18	11.1		HA	D								
26178	Loch Ballygrant	77	26.6		HA	D								
43135	Blagdon Lake	45	164.6	9.0	HA	D	8.30	406	167.25	236	34	2.95	62	5.90
6405	Loch Meadie	116	39		HA	D								
2499	Loch Scarmclate	25	75.9		HA	D								
44518	Fonthill Lake	94	2.5	5.1	HA	Sh	8.06	453	200.67	31	16	5.20	30	6.44
43909	Shear Water	139	13.7	7.2	HA	Sh	8.75	235	70.63	31	8	2.42	78	10.56
32948	Llyn Dinam	8	9.7	1.5	HA	Sh	7.76	355	76.81	73	35	1.68	12	2.91
30244	Hornsea Mere	8	133.3	2.0	HA	Sh	8.22	509	170.42	500	-	2.01	26	4.26
33337	Llyn Coron	9	28.0	3.8	HA	Sh	8.05	311	97.00	106	56	3.45	9	6.75
32968	Llyn Penrhyn	8	22.3	2.6	HA	Sh	7.41	403	93.00	426	332	1.11	19	3.74
33627	Llyn Rhos-ddu	8	2.4	1.0	HA	Sh	7.43	348	134.37	54	28	1.22	22	8.93
2088	Loch of Mey	15	23		HA	Sh								

[1] Maximum depths given are those measured at the coring location. Note that these may not always be the absolute deepest point.

[2] GB Lake type follows the scheme of Phillips (2004); LA, MA, HA = low, medium & high alkalinity, respectively; Sh = shallow, D = deep.

Chemical data are given as annual means calculated from the Environment Agency WFD 2003-2004 dataset.

Table 2: Details of the cores and analysis undertaken at the 44 short-listed lakes – only 30 sites are included in the final analysis

GB Lakes WBID	Name	Grid Reference	[1]GB Lake type		[2]Core Code	Coring Date	Core length (cm)	Sample intervals analysed for diatoms (cm)
12578	Loch an Lagain	NH658955	P	D	LAGN1	07.09.04	20	0, 10, 20
15316	Loch na Moracha	NF846663	P	D	MORA1	27.09.04	20	0, 10, 20
18113	Loch Shnathaid	NF826426	P	D	SHNA1	29.09.04	20	0, 10, 20
8945	Loch Ascaig	NC849255	P	D	ASCA1	24.08.04	20	0, 20
11611	Loch Brora	NC852078	P	D	BROR1	24.08.04	1	None**
29184	Grasmere	NY338065	LA	D	GRAS1	25.07.04	28	0, 10, 27
35561	Llyn Bodlyn	SH648239	LA	D	BODL1	14.09.04	23	0, 10, 22
38409	Llyn Egnant	SN792671	LA	D	EGNA1	24.09.04	31	0, 15, 30
33803	Llyn Ogwen	SH659604	LA	D	OGWE1	20.08.04	25	0, 10, 25
33730	Llyn Padarn	SH569614	LA	D	PADA1	21.08.04	25	0, 10, 25
36405	Tal-y-llyn Lake	ST850421	LA	D	TALY1	18.09.04	21	0, 20
29021	Thirlmere Reservoir	NY313162	LA	D	THIR1	22.07.04	31	0, 30
11355	Loch Borralan	NC262108	LA	Sh	BORL1	29.06.04	15	0, 10, 15
26257	Loch Skerrols	NR341638	LA	Sh	SKEL1	14.07.04	20	0, 10, 20
1138	Loch Ailsh	NC315109	LA	Sh	AILS1	30.06.04	40	0, 40
2257	Loch nan Gabhar	NM968632	LA	Sh	GABH1	18.08.04	15	0, 15
5714	Loch Rangag	ND177415	LA	Sh	RANG1	25.08.04	25	None*
16530	Loch Gowan	NH152564	LA	Sh	GOWA1	10.08.04	30	None*
29321	Coniston Water	SD301940	MA	D	CONI1	25.07.04	36	0, 20, 35
29233	Windermere	SD392958	MA	D	WIND1	26.07.04	31	0, 20, 30
11642	Loch Craggie	NC624074	MA	D	CRA4	11.07.04	30	0, 10, 30
2490	Loch Hope	NC463548	MA	D	HOPL1	07.07.04	20	0, 20
18682	Loch Druidbeag	NF789376	MA	D	DRUI1	28.09.04	10	None**
29222	Elter Water	NY333041	MA	Sh	ELTW1	26.07.04	21	0, 10, 20
22395	Lochan Lùnn Dà-Bhrà	NN087659	MA	Sh	LUNN1	17.08.04	20	0, 10, 20
4974	Loch Syre	NC661448	MA	Sh	SYRE1	08.07.04	25	None*
32538	Llyn Alaw	SH392866	MA	Sh	ALAW1	16.08.04	20	None*

GB Lakes WBID	Name	Grid Reference	[1]GB Lake type		[2]Core Code	Coring Date	Core length (cm)	Sample intervals analysed for diatoms (cm)
28386	Talkin Tarn	NY545587	Marl	D	CZNYSSB1 (TALK)	22.07.04	31	0, 20, 30
15551	Loch Tormasad	NF820651	HA	D	TORM1	26.09.04	20	0, 10, 20
26944	Loch Kinnabus	NR301422	HA	D	KINB1	21.07.04	20	0, 20
34990	The Mere, Ellesmere	SJ406349	HA	D	SCM04B (ELLE)	11.08.04	35	0, 35
25899	Ardnave Loch	NR284727	HA	D	ARDN2	13.07.04	20	None*
26178	Loch Ballygrant	NR405662	HA	D	BALG3	22.07.04	20	None*
43135	Blagdon Lake	ST515596	HA	D	BLAG1	17.09.04	23	None*
6405	Loch Meadie	NC502410	HA	D	MEAD1	15.09.04	10	None**
2499	Loch Scarmclate	ND189596	HA	D	SCAM1	26.08.04	10	None**
44518	Fonthill Lake	ST937311	HA	Sh	FONT1	30.09.04	21	0, 10, 20
43909	Shear Water	ST850421	HA	Sh	SST84_1 (SHEA)	14.09.04	26	0, 15, 25
32948	Llyn Dinam	SH310775	HA	Sh	DINA2	16.08.04	20	0, 10, 20
30244	Hornsea Mere	TA190469	HA	Sh	HORN2	15.07.04	27	0, 26
33337	Llyn Coron	SH378700	HA	Sh	CORO2	16.08.04	20	0, 20
32968	Llyn Penrhyn	SH313768	HA	Sh	PERH2	16.08.04	30	None*
33627	Llyn Rhos-ddu	SH424648	HA	Sh	RHSD1	17.08.04	25	None*
2088	Loch of Mey	ND271736	HA	Sh	MEY1	14.09.04	10	None**

* Diatom preservation too poor for analysis of core bottom sample (core tops analysed within DALES)
** Cores too short for top and bottom analysis
[1] GB Lake type follows the scheme of Phillips (2004); LA, MA, HA = low, medium & high alkalinity, respectively; Sh = shallow, D = deep.
[2] Core code is the AMPHORA core code. For some sites, simpler, alternative codes (noted in parentheses) have been used in all figures.

3 Results and discussion

3.1 Peat (P) lakes

A summary diagram of the common diatom taxa found in the samples from the Peat lake type (occurring at >5% relative abundance in >1 sample) is illustrated in Figure 1. DCA biplots (axis 1 and 2) of the sample scores and species scores for the Peat lakes are displayed in Figure 5.

The four Peat lakes are all in Scotland and are deep. The sites all have fairly similar circumneutral to acidophilous diatom assemblages, dominated by *Fragilaria exigua*. Other diatom taxa common in the P lakes include *Achnanthes minutissima*, *Brachysira vi*trea, *Cymbella graciis*, *Tabellaria flocculosa*, *Eunotia incisa*, *Frustulia rhomboides*, small *Fragilaria* spp. and small *Navicula spp*. (*vitiosa* and *arvensis*). Overall, the diatom assemblages of the four Peat lakes do not show much change from core bottoms to core tops. Loch Ascaig has the lowest squared chord distance dissimilarity score at 0.318, with the distances of the other sites all being relatively low, lying close to the critical value of 0.475 and supporting the observed floristic stability.

Loch na Moracha shows the greatest change from core bottom to top, with the planktonic diatom, *Cyclotella comensis* decreasing from ~20% relative abundance in the bottom and mid-core samples, to <1% in the top. In Loch an Lagain's surface sediment sample, there is an increase in relative abundance of small benthic (sediment-dwelling) *Fragilaria* spp. at the expense of the epiphytic diatoms, *A. minutissima* and *B. vitrea*. This could indicate a decrease in aquatic macrophyte abundance. Loch Shnathaid may have experienced a slight reduction in acidity from core bottom to top. *F. exig*ua and *E. incisa* (acidophilous taxa) decrease in relative abundance, whilst *A. minutissima*, *B. vitrea* and *T. flocculosa* (circumneutral-acidophilous taxa) increase in relative abundance. Caution should be taken in over-interpreting the data from the Peat lakes because all four cores were only 20 cm long and therefore may be too short to represent reference conditions at the core base.

3.2 Low alkalinity (LA) lakes, deep and shallow

A summary diagram of the common diatom taxa found in the samples from the Low alkalinity lake types (occurring at >5% relative abundance in >1 sample) is illustrated in Figure 2. DCA biplots (axis 1 and 2) of the sample scores and species scores for the Low alkalinity lakes are displayed in Figure 6. The majority of lakes of the LA type are dominated by circumneutral, non-planktonic diatom taxa, although planktonic taxa are also common.

Grasmere and Llyn Padarn would appear to have experienced significant change from core bottom to top. In Grasmere, the dominant taxon in the core bottom is *A. minutissima*, whereas in the mid and surface samples, *A. formosa* constitutes a higher relative abundance. *C. comensis* only occurs in the bottom sample. An increasing abundance of *A. formosa* at the expense of *C. comensis* and *A. minutissima* is usually interpreted as indicating increased nutrient status, since *A. formosa* frequently appears in formerly oligotrophic lakes as a sign of enrichment. The changes also indicate a shift from a largely periphytic diatom community to a plankton-dominated community. The relatively high squared chord dissimilarity distance (0.65) (Figure 2) and the shift in sample scores (Figure 6) between the top and bottom samples of Grasmere support the inference of significant change in diatom species assemblages. In Llyn Padarn, the diatom assemblage appears to have shifted from dominance of *C. comensis*, towards co-dominance of other planktonic diatom taxa such as *T. flocculosa*, *Aulacoseira subarctica*, *A. formosa* and *Cyclotella pseudostelligera*. Thirlmere Reservoir may tentatively be considered a reference site, since the diatom assemblage of *A. minutissima* and *C. comensis / rossii* has remained relatively constant from core bottom to top. In common with Grasmere, the appearance of *A. formosa* in the surface sediment sample may indicate a slight shift towards more mesotrophic conditions.

Three of the LA deep Welsh lakes, Llyn Bodlyn, Llyn Egnant and Llyn Ogwen have experienced significant changes in species composition from core bottom to top. These changes would appear to be indicative of increasing acidity (i.e. a decrease in pH). For Llyn Bodlyn and Llyn Egnant, Figure 6 clearly illustrates the shift in sample scores from the right to the left of the diagram, as one moves from core bottom to top. This corresponds to a shift from predominantly circumneutral, non-planktonic taxa to an increase in relative abundance of circumneutral acidophilous, non-planktonic and planktonic taxa. In terms of diatom species, the community has changed from an *A. minutissima*-dominated community, to one dominated by *E. incisa* and *T. flocculosa* (Llyn Egnant), with *Cymbella perpusilla* (Llyn Egnant) and *Peronia fibula* (Llyn Bodlyn) also increasing in relative abundance. At Llyn Bodlyn and Llyn Egnant there has also been a loss of *C. comensis* and *Cyclotella rossii,* respectively. It is noted that care should be taken in the interpretation of the shift towards *T. flocculosa* in the surface sediments of both Llyn Bodlyn and Llyn Egnant, because *T. flocculosa* is a bloom-forming planktonic taxon and is a frequent component of autumn diatom blooms. Since cores from these sites were taken in September, it is likely that *T. flocculosa* is over-represented in the surface sediments of the cores from these sites. It may be worth examining a sample from a depth of 1-2 cm to determine whether the increase in *T. flocculosa* in the core tops is merely a seasonal artefact.

Lochs Ailsh (LA, Sh), nan Gabhar (LA, Sh) and Tal-y-llyn Lake (LA, D) do not appear to have experienced significant diatom assemblage changes from core bottom to top and it is therefore suggested that these sites are good examples of LA reference sites. These sites are dominated by predominantly non-planktonic, circumneutral to acidic diatom taxa. Species diversity is high, with a broad range of periphytic taxa (e.g. *A. minutissima* usually dominating, with *F. exigua*, *B. vitrea*, *Synedra rumpens*, *Cymbella minuta* and *E. incisa* co-occurring in differing proportions in the different sites) and the presence of the planktonic diatom taxon, *T. flocculosa*. The high diversity of the diatom assemblages, likely brought about by the availability of a wide range of habitats, may explain the relative stability of these sites, in much the same way as high macrophyte species diversity plays a key structuring role in lakes (Carpenter & Lodge, 1986; Jeppesen *et al*., 1997). The squared chord dissimilarity distances between the top and bottom samples of these sites all lie below 0.475, supporting the inference of little change in diatom species assemblages. Furthermore, the top and bottom sample scores (displayed in Figure 6) lie in close proximity, indicating similar species assemblages.

Visual interpretation of the diatom profile from Loch Skerrols indicates that the core is dominated throughout by a high diversity of periphytic taxa (Figure 2). The record appears to be relatively stable, with *A. minutissima* occurring at a consistent relative abundance throughout the core. The squared chord dissimilarity distance is highest (0.647) in the mid-core sample, probably due to a slight shift towards increased relative abundances of small *Fragilaria* spp. and *Gomphonema pumilum*. In the surface sediment sample, the epiphytic diatom, *Cocconeis placentula* is dominant. These shifts may merely indicate subtle changes in the availability of different periphytic diatom habitats within a clear-water lake, as opposed to a significant ecological shift towards a different ecological state. The ongoing work of DALES may elucidate this observation through the analysis of seasonal periphytic diatom samples. The data suggest that Loch Skerrols is an example of a LA, Sh reference site.

Although Loch Borralan displays relatively little change in its diatom species assemblage and has low squared chord dissimilarity distances between its top and bottom samples, *A. formosa*, *Fragilaria crotonensis*, *Aulacoseira granulata* var. *angustissima* and *Stephanodiscus parvus* are consistent components of the diatom community in all core samples. These planktonic taxa are considered to be indicators of mesotrophic to eutrophic conditions and it is therefore unlikely that this site can be considered an example of a LA, Sh reference site. Furthermore, the apparent stability in the diatom assemblage of this site may arise because the core is only 15 cm long, and is probably insufficient to extend back to baseline conditions unless the sedimentation rate is extremely low. Alternatively, Loch Borralan may be a naturally meso-eutrophic LA, Sh lake.

3.3 Medium alkalinity (MA) lakes, deep and shallow

A summary diagram of the common diatom taxa found in the samples from the Medium alkalinity lake types (occurring at >5% relative abundance in >1 sample) is illustrated in Figure 3. DCA biplots (axis 1 and 2) of the sample scores and species scores for the Medium alkalinity lakes are displayed in Figure 7.

Lake Windermere (MA, D) has experienced the greatest change of all MA sites in this study. The diatom assemblage of this lake has shifted from dominance of the oligotrophic, planktonic taxon, *C. comensis* (and *C. krammeri*) in the bottom and mid-core samples, to the co-occurrence of the mesotrophic indicators, *A. formosa*, *A. subarctica*, *F. crotonensis*, *Aulacoseira islandica* and *S. parvus* in the surface sediment sample. In addition, the relative abundance of *A. minutissima* has decreased and *Cyclotella radiosa* has been eliminated. These species shifts are indicative of nutrient enrichment; a phenomenon that is well documented for Windermere based on both long-term monitoring data (Reynolds & Irish, 2000) and detailed palaeolimnological studies (Sabater & Haworth, 1995). The dissimilarity score between the bottom (30 cm) and mid-core (20 cm) sample is <0.475 and the DCA biplot sample scores are similar, indicating little change in the lower part of the profile. In contrast, the high squared chord distance dissimilarity score (1.17) between the bottom and top samples and the position of the top sample at a distance from the mid and bottom core samples in the DCA biplot (Figure 7) both indicate significant floristic change in recent times. Coniston Water (MA, D) has experienced similar floristic changes indicative of an increase in productivity but to a lesser extent than Windermere, with dissimilarity scores between the bottom sample and upper samples of ~ 0.7.

Elter Water (MA, Sh), appears to have experienced a shift from periphyton dominance to plankton dominance. *A. minutissima* has decreased in relative abundance and *A. formosa*, *C. stelligera / pseudostelligera* and *Cyclostephanos invisitatus* have increased in relative abundance (Figure 3). Since Elter Water is shallow, the change from periphyton to plankton may be indicative of a shift from the clear-water to turbid water state (cf. the alternative stable states theory of Scheffer *et al*., 1993). The species shifts are reflected in the squared chord distance dissimilarity scores. The dissimilarity scores between the bottom sample and the 20 cm and 10 cm samples are both in excess of the critical value. The sample scores in the DCA biplot (Figure 7) move from right to left, reflecting the compositional change in the sediment record and the large distances between data points reflect the large degree of change. Lochan Lùnn Dà-Bhrà is the other MA, Sh site included in this project and it would appear to be a good example of a MA, Sh reference site. *A. minutissima* is dominant from core bottom to top, with a stable community of small *Fragilaria* spp., *B. vitrea*, *F. exigua* and *Synedra tenera / nana* occurring at low relative abundances (Figure 3). Although *A. formosa* does not appear in the surface sample, *C. comensis* is absent. The loch should be monitored in future years to determine whether the loss of *C. comensis* continues and whether this change represents an ecological shift.

Upon visual inspection of Figure 3, Loch Hope (MA, D) appears to be a potential reference site. This site's diatom assemblage has changed little between core bottom and top, displaying co-dominance of *A. minutissima* and *C. comensis / rossii* in both samples and co-occurrence of *T. flocculosa*, *B. vitrea* and *S. tenera / nana*. Loch Hope also displays a squared chord distance dissimilarity score of <0.475 and its sample scores are very similar, with the top and bottom samples lying in close proximity on the DCA biplot (Figure 7). However, the appearance of *A. formosa* in the surface sediment sample may provide evidence of a slight increase in nutrients. Loch Craggie (MA, D) also shows minimal change between core top and bottom (30 cm) samples, with *Fragilaria* spp. dominating throughout, alongside the small *Navicula* species, *Navicula vitiosa*. Any shifts in species abundances occur between different small *Fragilaria* spp. As in the case of Loch Hope, planktonic diatom taxa (including *A. formosa*, *C. pseudostelligera* and *A. subarctica*) are found at very low relative abundances in the upper core samples (10 cm and surface sediment), perhaps providing a subtle warning of the potential for nutrient enrichment at this site. Overall, however, both Loch Hope and Loch Craggie provide the best examples of MA, D reference lakes in this study.

3.4 High alkalinity (HA), deep and shallow lakes, and Marl lakes

A summary diagram of the common diatom taxa found in the samples from the High alkalinity and Marl lake types (occurring at >5% relative abundance in >1 sample) are illustrated in Figure 4. DCA biplots (axis 1 and 2) of the sample scores and species scores for the High alkalinity and Marl lakes are displayed in Figure 8.

The one Marl lake included in the current study, Talkin Tarn, appears to have experienced moderate change from core bottom to top, with nutrient enrichment being the likely cause. The shift in the diatom species assemblage is from *C. radiosa*, a mesotrophic planktonic taxon dominant in the core bottom (30 cm) and mid (20 cm) samples, to the more nutrient-tolerant *S. parvus*, *Fragilaria capucina* var. *mesolepta*, *Cyclostephanos dubius*, *Stephanodiscus hantzschii* and *A. granulata* var. *angustissima* in the surface sediment sample (Figure 4 and Figure 8). The species shifts are reflected in the squared chord distance dissimilarity scores. Although there is little change between the bottom and mid-core samples (0.330), the dissimilarity between the bottom and top samples is considerably higher (0.784). Furthermore, there is little difference in the sample scores in the DCA biplot (Figure 8) between 30 and 20 cm, but between 30 and 0 cm, the sample scores move from lower left to upper right reflecting the compositional change in the diatom species assemblage. These results indicate that the lake's diatom community was stable until recent times.

Of the eight high alkalinity lakes, three are deep (Loch Tormasad, Loch Kinnabus and The Mere, Ellesmere) and five are shallow (Fonthill Lake, Shear Water, Llyn Dinam, Hornsea Mere and Llyn Coron). The majority of the HA sites appear to have experienced significant change from core bottom to top. Fonthill Lake has seen a shift from a small *Fragilaria* spp. community to one in which the planktonic diatoms *Aulacoseira granulata* (and var. *angustissima*) and *F. crotonensis* all occur at increasing relative abundance. This species shift indicates nutrient enrichment and a switch from clear-water to turbid water conditions (cf. the alternative stable states theory of Scheffer *et al*., 1993). Planktonic diatoms have dominated Shear Water throughout the core with *C. dubius* being dominant and *S. hantzschii* being sub-dominant. The squared chord distance score of 0.848 between core bottom and top samples, reflects changes in the planktonic diatom taxa with a shift from *F. crotonensis* and *A. formosa* to *C. radiosa* and *A. subarctica* (Figure 4 and Figure 8). These taxa are all considered mesotrophic indicators and therefore the shifts may simply reflect inter-annual variation. It is recommended that a longer core is taken from Shear Water for a higher resolution investigation to determine ecological state, and in particular to determine whether the lake is a potential reference site. Llyn Dinam appears to have experienced an increase in the relative abundance of epiphytic (plant-dwelling) diatom taxa (e.g. *C. placentula* and *Rhoicosphenia abbreviata*) alongside a decreasing relative abundance of small benthic (sediment-dwelling) *Fragilaria* spp. and a corresponding increase in the relative abundance of small planktonic centric diatom taxa (e.g. *S. parvus*, *Cyclostephanos invisitatus / tholiformis*). Llyn Dinam may therefore have experienced a slight shift away from the clear-water state, increasing the available habitat for planktonic diatoms. In addition this site may have seen an increase in the available habitat for the growth of *C. placentula* and *R. abbreviata*, namely an increase in the infestation of aquatic macrophytes or filamentous algae. Work being undertaken as part of DALES may shed light on the habitat preferences of *C. placentula* and *R. abbreviata*, thus enabling better interpretation of the observed shifts in Llyn Dinam's diatom species composition.

Loch Tormasad's diatom profile is difficult to interpret, since there are no major shifts in floristic composition, despite significant squared chord distance dissimilarity scores of 0.891 (core bottom to top) and 0.686 (core bottom to middle). There is a possibility that this site has seen a slight decrease in acidity from core bottom to top, since the relative abundance of *F. exigua* decreases as that of small *Fragilaria* spp. typically associated with more circumneutral to alkaline waters, and *A. minutissima* increase (Figure 4). To investigate the possibility of pH shifts at this site, it may be worth applying a pH transfer function to the diatom samples of Loch Tormasad. In terms of nutrients, this site could potentially be an example of a reference site, although further investigation is recommended using a longer sediment core to provide a deeper bottom sample.

The short distances between core bottom and top samples in the DCA biplot (Figures 8) illustrate that the diatom communities of four of the HA lakes - Loch Kinnabus, Hornsea Mere, Ellesmere and Llyn Coron - remain relatively stable from core bottom to top, perhaps indicating that these lakes can be classified as 'reference' sites. However, the data must be interpreted with caution because all cores were short (20 to 35 cm). For productive, HA lowland systems, where sediment accumulation rates are relatively rapid, it is unlikely that cores of only 20 to 35 cm in length are sufficient to extend back to 1850. Although Ellesmere has similar top and bottom diatom assemblages (Figure 4), it is recommended that bottom samples from longer sediment cores are investigated prior to making a decision as to the status of this lake. Hornsea Mere has been the subject of a previous palaeoecological study of a longer core. Nutrient-related shifts in diatom species assemblages were observed and the squared chord distance between core top and bottom (40 cm) was 0.893 (Bennion, 2004). Therefore, Hornsea Mere cannot be considered a reference site. Poor diatom preservation in Llyn Coron hindered data analysis and the results must be interpreted with caution. It is suggested that at this site, separate studies are conducted using indicators other than diatoms (e.g. cladocera, plant macrofossils), to further assess ecological status.

Given the concerns raised above, there appears to be only one potential reference lake within the HA group – Loch Kinnabus, where the diatom assemblage shows no significant change between top and bottom samples (squared chord distance = 0.418). The diatom assemblage of this site is diverse, dominated by *A. minutissima* throughout, with small *Fragilaria* spp. occurring alongside (Figure 4). These non-planktonic taxa are typical of moderate nutrient status and clear-water conditions. It is worth noting that the planktonic diatoms *C. radiosa*, *A. formosa*, *C. dubius* and *Stephanodiscus medius* occur at low relative abundance in the core top sample. This species shift is indicative of mild eutrophication and may well provide an early warning that an ecologically important threshold has been crossed at Loch Kinnabus and that it is experiencing the early stages of nutrient enrichment. Nevertheless, this site provides the best example of a HA, D reference lake within this study, but monitoring and nutrient reduction measures should be put in place to ensure that this site remains in good condition.

Figure 1: Summary diagram of diatom changes (% relative abundance) in cores from the Peat (P), Deep lake types. 'Distance' is the squared chord distance dissimilarity score between the core bottom and top (and mid) samples.

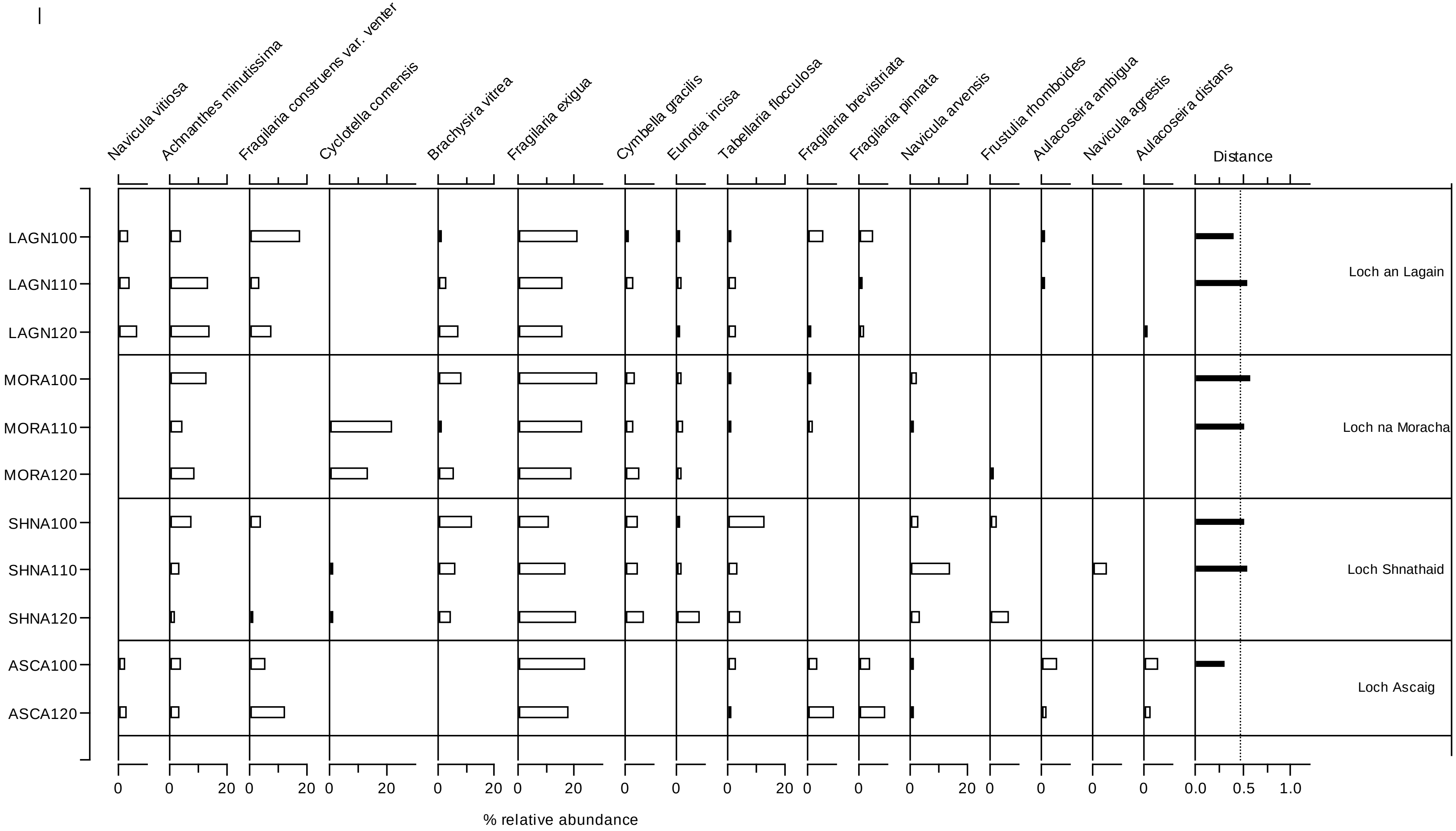


Figure 2: Summary diagram of diatom changes (% relative abundance) in cores from the Low Alkalinity (LA), Deep and Shallow lake types. 'Distance' is the squared chord distance dissimilarity score between the core bottom and top (and mid) samples.

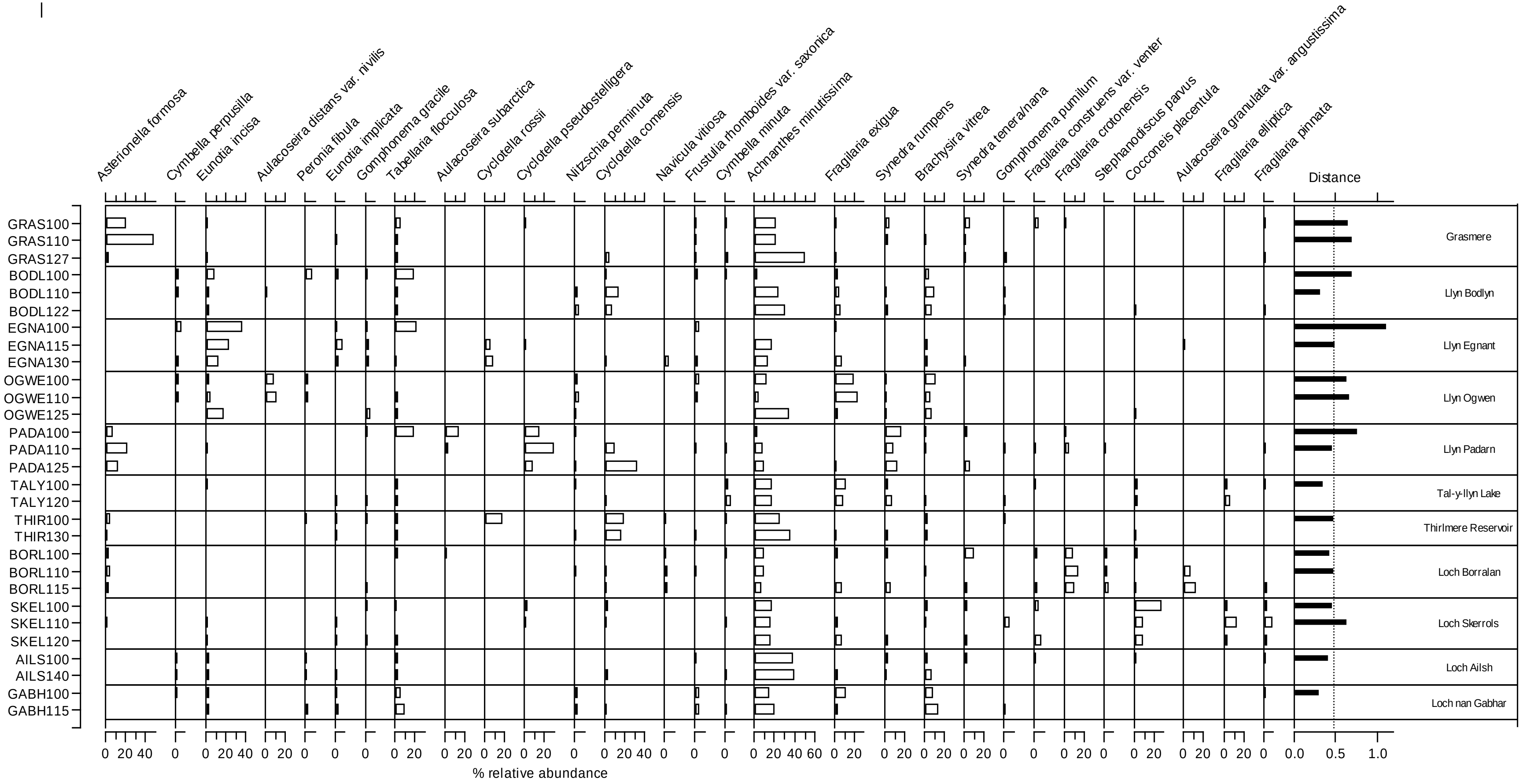


Figure 3: Summary diagram of diatom changes (% relative abundance) in cores from the Medium Alkalinity (MA), Deep and Shallow lake types. 'Distance' is the squared chord distance dissimilarity score between the core bottom and top (and mid) samples.

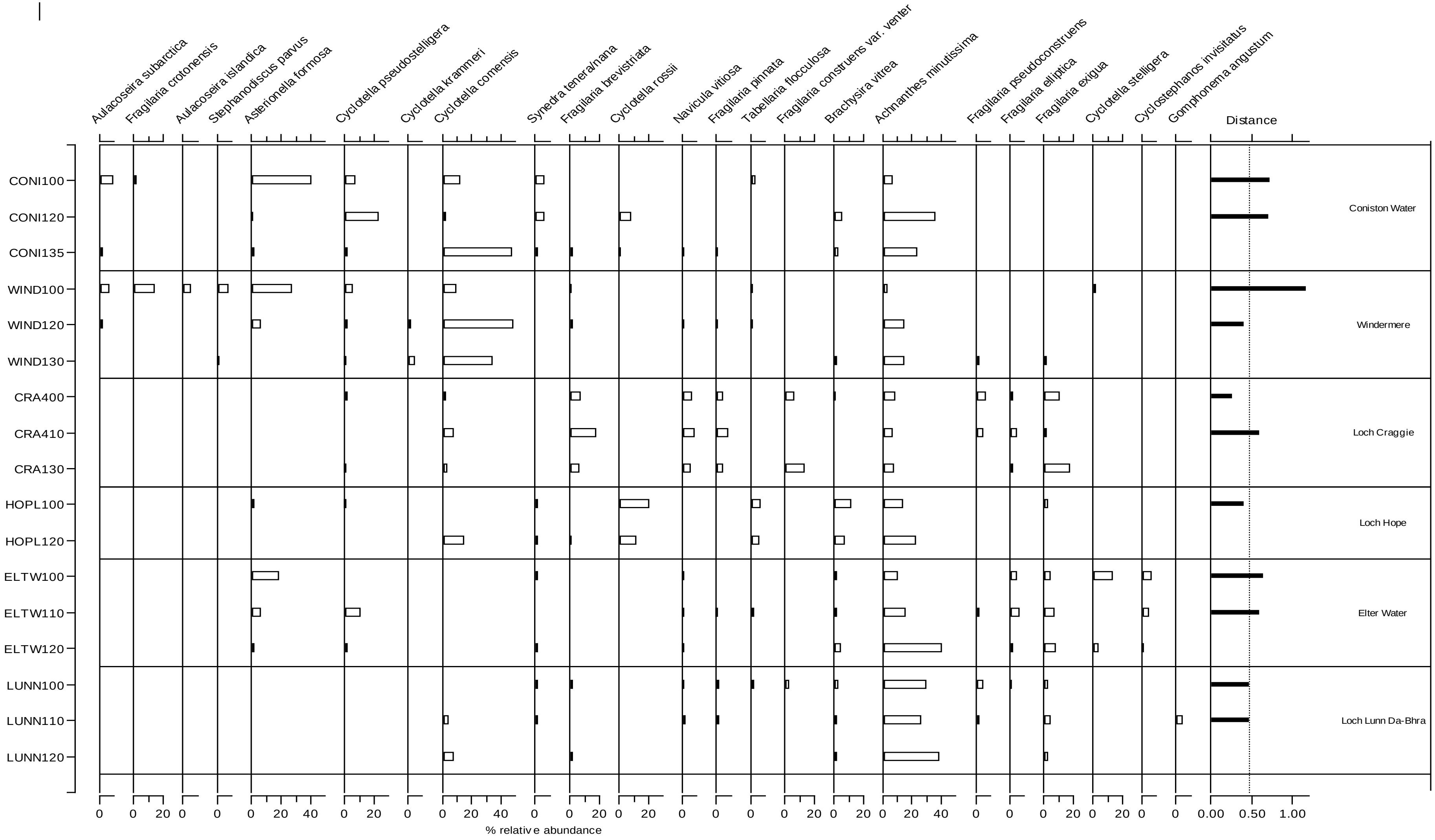


Figure 4: Summary diagram of diatom changes (% relative abundance) in cores from the High Alkalinity (HA), Deep and Shallow and Marl (Marl) lake types. 'Distance' is the squared chord distance dissimilarity score between the core bottom and top (and mid) samples.

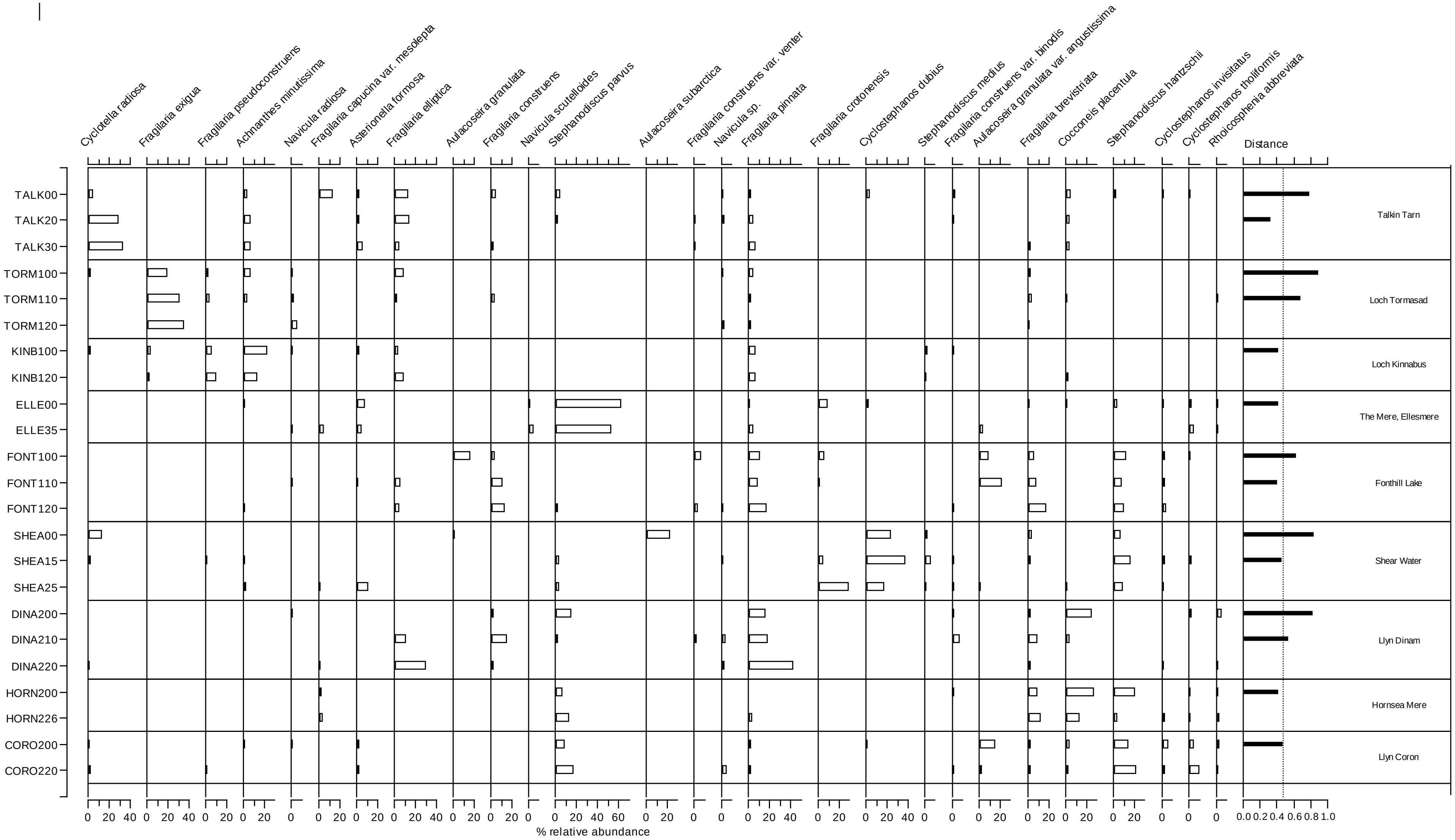
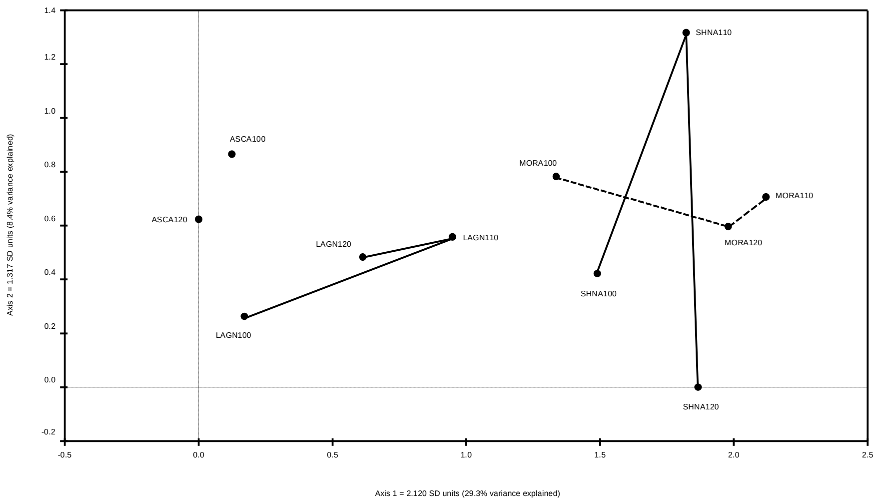


Figure 5 DCA biplots (axis 1 and 2) of the sample scores and species scores for Peat lakes



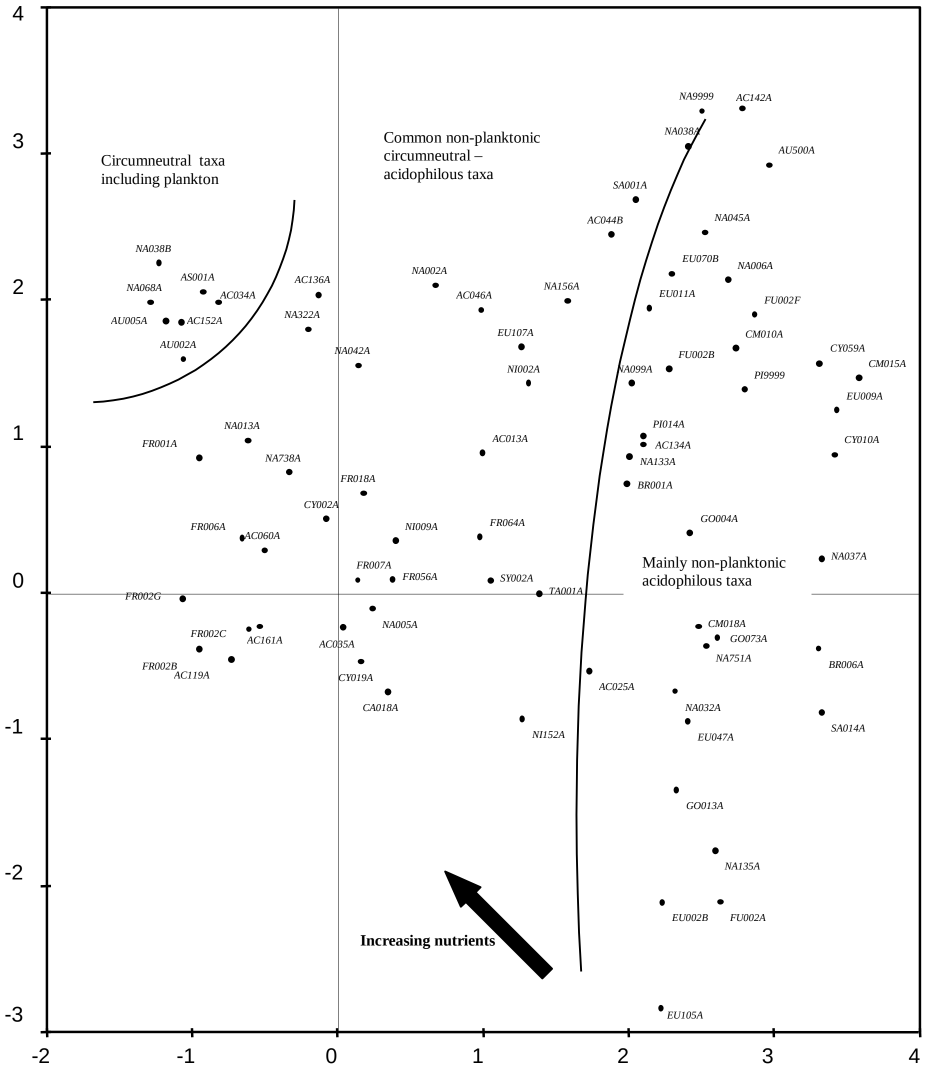


Figure 6 DCA biplots (axis 1 and 2) of the sample scores and species scores for LA lakes

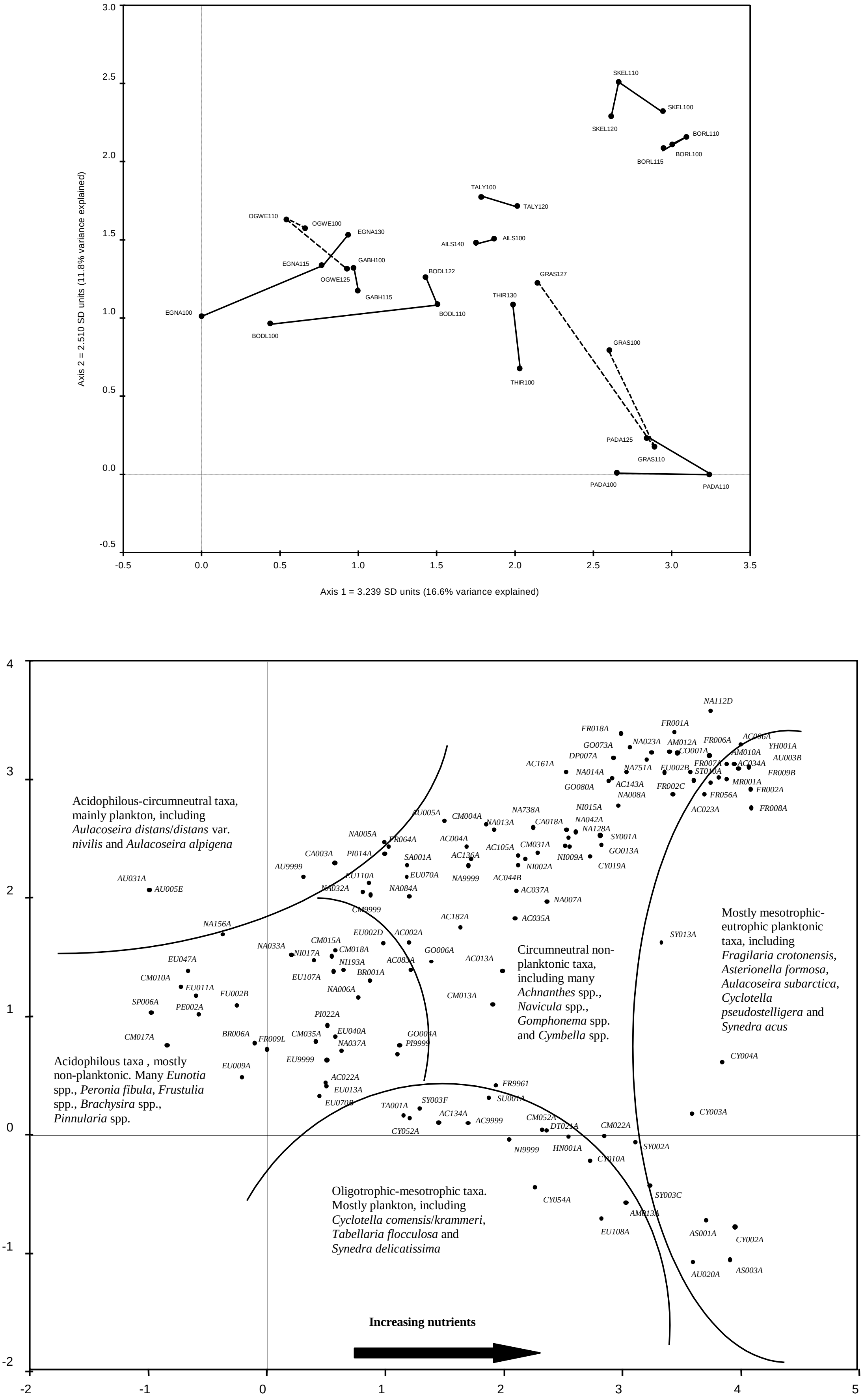
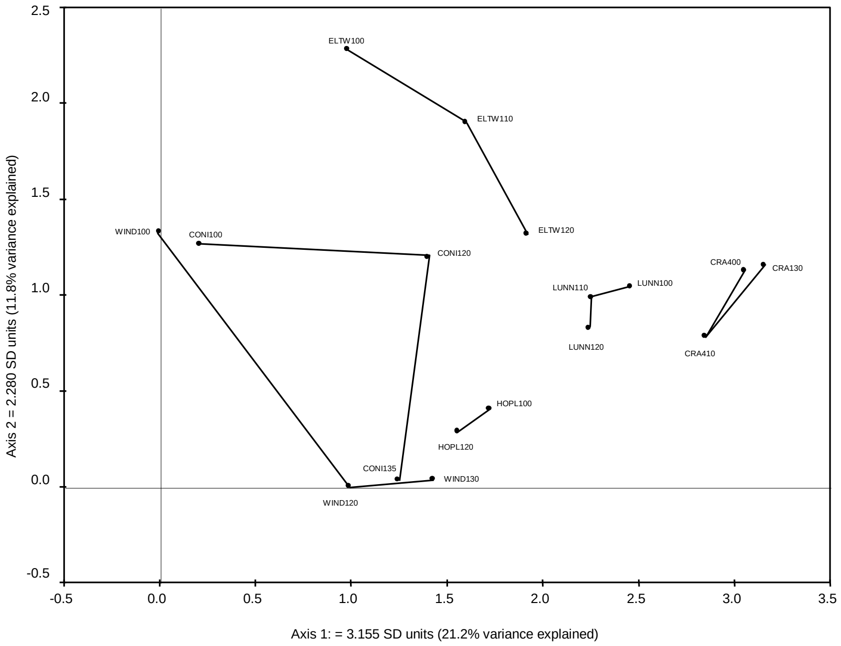


Figure 7 DCA biplots (axis 1 and 2) of the sample scores and species scores for MA lakes



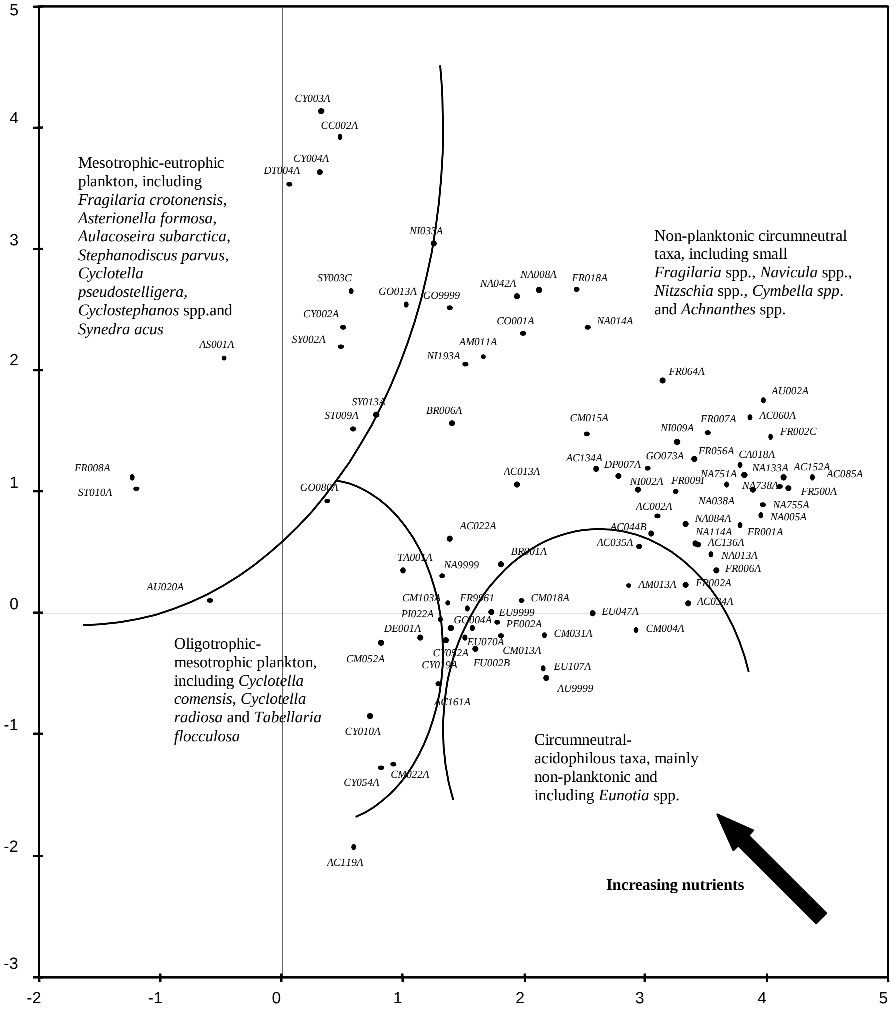
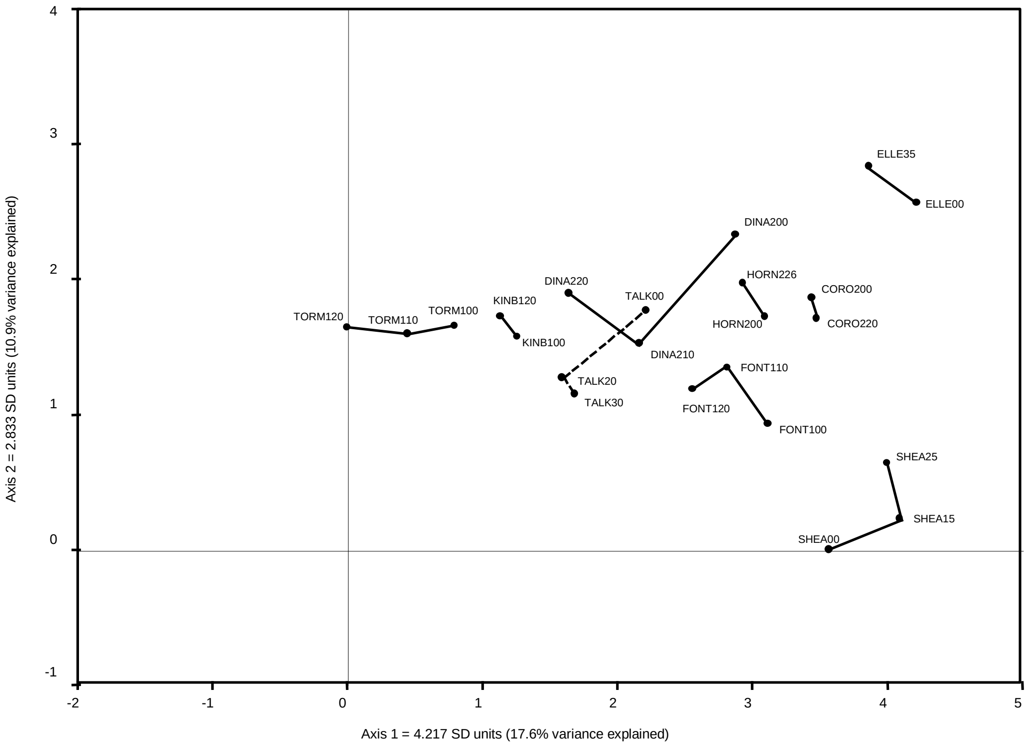
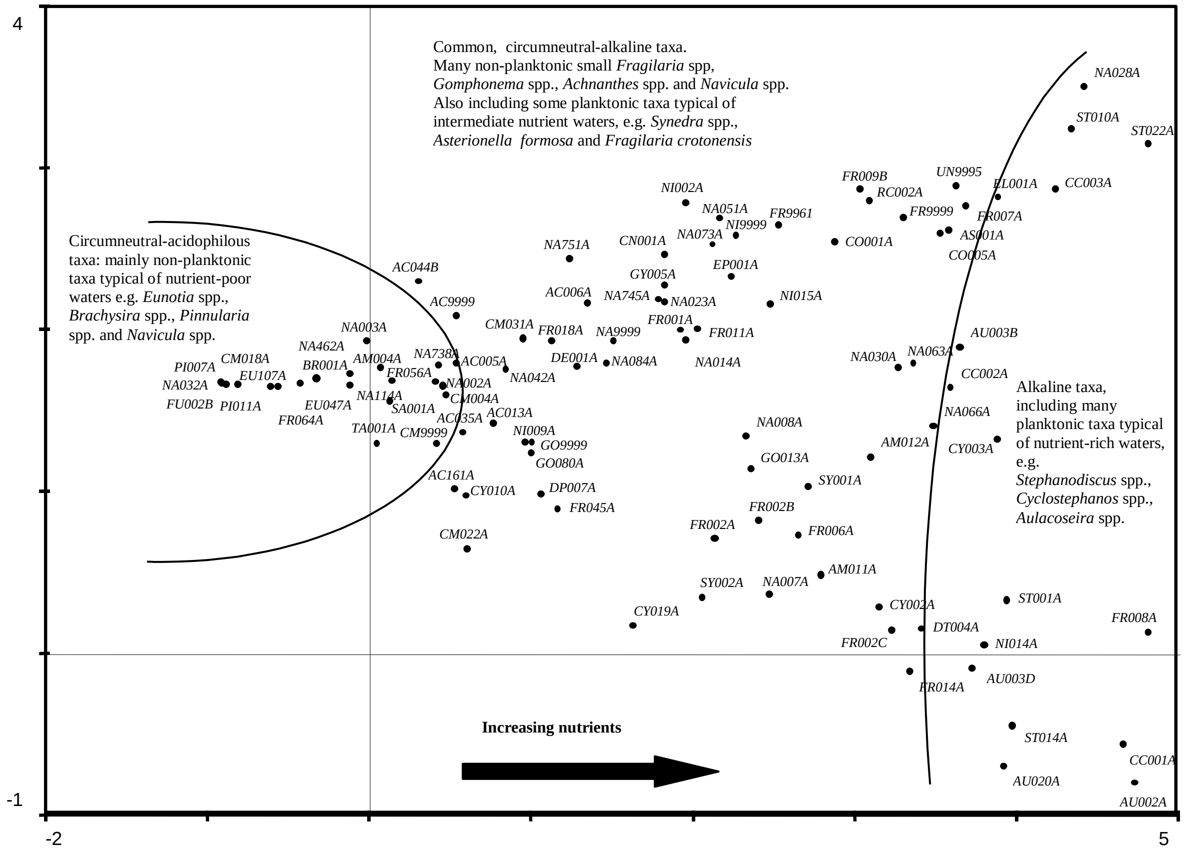


Figure 8 DCA biplots (axis 1 and 2) of the sample scores and species scores for HA and Marl lakes





Common, circumneutral-alkaline taxa. Many non-planktonic small *Fragilaria* spp, *Gomphonema* spp., *Achnanthes* spp. and *Navicula* spp. Also including some planktonic taxa typical of intermediate nutrient waters, e.g. *Synedra* spp., *Asterionella formosa* and *Fragilaria crotonensis*

Circumneutral-acidophilous taxa: mainly non-planktonic taxa typical of nutrient-poor waters e.g. *Eunotia* spp., *Brachysira* spp., *Pinnularia* spp. and *Navicula* spp.

Alkaline taxa, including many planktonic taxa typical of nutrient-rich waters, e.g. *Stephanodiscus* spp., *Cyclostephanos* spp., *Aulacoseira* spp.

4 Summary of findings

Following screening for diatom preservation and the subsequent elimination of sites displaying diatom dissolution, core bottoms were analysed from 30 lakes (10 England, 7 Wales and 13 Scotland). Mid-core samples were analysed from a sub-set of 20 of these lakes (7 England, 5 Wales and 8 Scotland), resulting in the analysis of a total of 50 mid and bottom core samples. The degree of floristic change between the bottom and top/mid samples analysed in each core was estimated using the squared chord distance dissimilarity coefficient, and detrended correspondence analysis (DCA) was performed to assess the direction and magnitude of floristic change at each site. The results are summarised in Table 3 and an overview of the findings is discussed below.

Table 3: Summary results from the diatom analysis of 30 cores

Lake name	[1]GB Lake type		[1]Significant floristic change?	[2]Trophic change?	Depth of reference sample (cm)	[3]Potential reference site?
Loch an Lagain	P	D	No (20-0) … Yes (20-10)	No	20	Yes
Loch na Moracha	P	D	Yes	No	20	Possibly
Loch Shnathaid	P	D	Yes but relatively low score (~0.5)	↑ pH?	20	Possibly
Loch Ascaig	P	D	No	No	20	Yes
Grasmere	LA	D	Yes	↑	28	No
Llyn Bodlyn	LA	D	Yes (22-0) … No (22-10)	↓ pH?	23	No
Llyn Egnant	LA	D	Yes	↓ pH?	31	No
Llyn Ogwen	LA	D	Yes	↓ pH?	25	No
Llyn Padarn	LA	D	Yes (25-0) … No (25-10)	↑	25	No
Tal-y-llyn Lake	LA	D	No	No	21	Yes
Thirlmere Reservoir	LA	D	Yes but relatively low score (0.48)	?	31	Possibly
Loch Borralan	LA	Sh	No (15-0) … Yes (15-10)	?	15	Possibly
Loch Skerrols	LA	Sh	No (20-0) … Yes (20-10)	No	20	Yes
Loch Ailsh	LA	Sh	No	No	40	Yes
Loch nan Gabhar	LA	Sh	No	No	15	Yes
Coniston Water	MA	D	Yes	↑	25	No
Windermere	MA	D	Yes (30-0) … No (30-20)	↑	30	No
Loch Craggie	MA	D	No (30-0) … Yes (30-10)	No	30	Yes
Loch Hope	MA	D	No	No	20	Yes
Elter Water	MA	Sh	Yes	↑	21	No
Lochan Lùnn Dà-Bhrà	MA	Sh	No	No	20	Yes
Talkin Tarn	Marl	D	Yes (30-0) … No (30-20)	↑	31	No
Loch Tormasad	HA	D	Yes	↑ pH?	20	Possibly
Loch Kinnabus	HA	D	No	No	20	Yes
The Mere, Ellesmere	HA	D	No	?	35	Possibly
Fonthill Lake	HA	Sh	Yes (20-0) …. No (20-10)	↑	21	No
Shear Water	HA	Sh	Yes (25-0)… No (25-15)	↑	26	Possibly
Llyn Dinam	HA	Sh	Yes	↑	20	No
Hornsea Mere	HA	Sh	No	?	27	No
Llyn Coron	HA	Sh	Yes but relatively low score (0.48)	?	20	Possibly

[1] Change was deemed significant where the squared chord distance dissimilarity scores between the core bottom and the mid and/or top sample exceeded the critical value of 0.475. The numbers in parentheses indicate the depth (cm) of the two samples being compared.

[2] ↑ increase in trophic status; ↓ decrease in trophic status; ? uncertain. Assessment of trophic change is based on the diatom species shifts. Where floristic change is indicative of a shift in pH rather than trophic status, the symbol 'pH' is shown.

[3] For sites classed as 'Possibly', please see text.

In summary, 17 of the 30 sites appear to have experienced significant floristic change in diatom species assemblages and 13 sites displayed minimal floristic change. The majority of changes appear to relate to increases in trophic status, although at some sites, floristic changes suggest increasing acidity (e.g. Llyn Bodlyn, Llyn Egnant and Llyn Ogwen) or decreasing acidity (Lochs Tormasad and Shnathaid).

A total of 10 out of the 30 lakes are thought to be good examples of reference lakes, as highlighted by the minimal change in their diatom species assemblages and low squared chord distance dissimilarity distances between core bottom and top samples. Tal-y-llyn Lake (LA, D) is the only non-Scottish example of a reference lake in the current study. All other potential reference sites are Scottish lochs and examples for each lake type are present with the exception of Marl lakes, for which there are no Scottish examples in the current study. Loch Lagain and Loch Ascaig are examples of Peat lake reference sites, Lochs Skerrols, Ailsh and nan Gabhar are examples of LA, Sh reference lakes, Lochs Craggie and Loch Hope are examples of MA, D reference lakes, Lochan Lùnn Dà-Bhrà is an example of a MA, Sh reference lake and Loch Kinnabus is the only example of a HA, D reference lake. At some reference lakes e.g. Loch Kinnabus, Loch Hope, Loch Craggie and Lochan Lùnn Dà-Bhrà, there may be early warning signs of slight increases in trophic status based on the appearance of nutrient-tolerant diatom taxa in the surface samples. It is suggested that the water chemistry and ecology of these sites is monitored closely over the coming years, to ascertain whether a shift from reference conditions is occurring. Can the reference status of these lakes be further supported by current measured water chemistry? (NEED DATA FROM SEPA).

A further eight lakes experienced relatively low floristic change over the period represented by the cores and may therefore be potential reference lakes - Loch na Moracha (P, D), Loch Shnathaid (P, D), Thirlmere Reservoir (LA, D), Loch Borralan (LA, Sh), Loch Tormasad (HA, D), The Mere, Ellesmere (HA, D), Shear Water (HA, Sh) and Llyn Coron (HA, Sh). However, further investigation is required to confirm their status.

One limitation of the current study is that none of the cores are dated and for some lakes where sediment cores are short, core bottom samples may not represent true ‘reference’ samples. This is of particular concern for the HA lakes since sediment accumulation rates can be rapid in these productive systems. The lack of a chronology is of less concern where the diatom assemblages remain stable throughout the core but it becomes a greater limitation when interpreting the data from sites that exhibit floristic change, as we have no estimate of the time at which the changes occurred. Sediment accumulation rates are site specific and it is unlikely that all cores cover comparable time periods. Longer cores would need to be collected and radiometric or spheriodal carbonaceous particle (SCP) dating carried out to provide a more detailed assessment of the nature and timing of ecological changes at these sites over longer timescales. Analysis of remains of other biological elements preserved in lake sediment cores, such as plant macrofossils, cladocera and chironomids, would enable ecological reference conditions to be defined in a more holistic way than can be achieved using the diatom record alone, and would provide valuable information on changes in ecological structure and function (e.g. Sayer *et al*., 1999; Bennion, 2001).

It is recommended that further palaeoecological work be carried out at those sites highlighted as potential reference lakes. In cases where sediment cores were short and core bottom samples are thought not to extend back far enough to represent reference conditions, it is suggested that longer sediment cores are taken. Diatom analysis of core bottom samples and dating using either radiometric or SCP methods to confirm suitability as reference samples is advised. In addition, sites such as Loch Borralan, where no change between top and bottom samples was seen but where the assemblages were comprised of many nutrient-tolerant diatom taxa, may represent naturally meso-eutrophic lakes and further analysis is recommended.

Nevertheless, this study illustrates that even low resolution analysis of sediment cores can produce valuable information for the determination of reference conditions and to assist with reference lake selection, and to enable the assessment of ecological status and extent of ecological change at impacted sites. Simple techniques such as dissimilarity scores and ordination analyses applied to palaeoecological data can act as powerful tools for characterising lake types, defining ecological reference conditions and assessing deviation from the reference state.

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Appendix 1: List of the 203 common diatom taxa (>1% in >2 samples) in the 30 cores with Diatcode and full name

Diatcode	Name
AC002A	Achnanthes linearis
AC004A	Achnanthes pseudoswazii
AC005A	Achnanthes calcar
AC006A	Achnanthes clevei
AC013A	Achnanthes minutissima
AC022A	Achnanthes marginulata
AC023A	Achnanthes conspicua
AC025A	Achnanthes flexella
AC034A	Achnanthes suchlandtii
AC035A	Achnanthes pusilla
AC037A	Achnanthes biasolettiana
AC044B	Achnanthes levanderi
AC046A	Achnanthes altaica
AC060A	Achnanthes curtissima
AC083A	Achnanthes laevis
AC085A	Achnanthes lauenbergiana
AC105A	Achnanthes petersenii
AC119A	Achnanthes saccula
AC134A	Achnanthes helvetica
AC136A	Achnanthes subatomoides
AC142A	Achnanthes kuelbsii
AC143A	Anchnanthes oblongella
AC152A	Achnanthes carrisima
AC161A	Achnanthes ventralis
AC182A	Achnanthes rosenstockii
AC9999	Achnanthes sp.
AM004A	Amphora veneta
AM010A	Amphora fogediana
AM011A	Amphora libyca
AM012A	Amphora pediculus
AM013A	Amphora inariensis
AS001A	Asterionella formosa
AS003A	Asterionella ralfsii
AU002A	Aulacoseira ambigua
AU003B	Aulacoseira granulata var. angustissima
AU003D	Aulacoseira granulata
AU005A	Aulacoseira distans
AU005E	Aulacoseira distans var. nivilis
AU020A	Aulacoseira subarctica
AU031A	Aulacoseira alpigena
AU500A	Aulacoseira crassipuncta
AU9999	Aulacoseira sp.
BR001A	Brachysira vitrea
BR006A	Brachysira brebbisonnii
CA003A	Caloneis silicula
CA018A	Caloneis tenius
CC001A	Cyclostephanos dubius
CC002A	Cyclostephanos invisitatus
CC003A	Cyclostephanos tholiformis
CM004A	Cymbella microcephala
CM010A	Cymbella perpusilla
CM013A	Cymbella helvetica
CM015A	Cymbella cesattii
CM017A	Cymbella hebridica
CM018A	Cymbella gracilis
CM020A	Cymbella gaeumannii
CM022A	Cymbella affinis
CM031A	Cymbella minuta
CM035A	Cymbella angustata
CM052A	Cymbella descripta
CM085A	Cymbella lapponica
CM103A	Cymbella silesiaca
CM9999	Cymbella sp.
CN001A	Cymbellonitzschia diluviana
CO001A	Cocconeis placentula
CO005A	Cocconeis pediculus
CY002A	Cyclotella pseudostelligera
CY003A	Cyclotella meneghiniana
CY004A	Cyclotella stelligera
CY010A	Cyclotella comensis
CY019A	Cyclotella radiosa
FR008A	Fragilaria crotonensis
FR009B	Fragilaria capucina var. mesolepta
FR009I	Fragilaria capucina var. austriaca
FR009L	Fragilaria capucina var. amphicephala
FR011A	Fragilaria lapponica
FR014A	Fragilaria leptostauron
FR018A	Fragilaria elliptica
FR045A	Fragilaria parasitica
FR056A	Fragilaria pseudoconstruens
FR064A	Fragilaria exigua
FR500A	Fragilaria suboldenburgiana
FR9961	Fragilaria vaucheriae (fine)
FR9999	Fragilaria sp.
FU002A	Frustulia rhomboides
FU002B	Frustulia rhomboides var. saxonica
FU002F	Frustulia rhomboides var. viridula
GO004A	Gomphonema gracile
GO006A	Gomphonema acuminatum
GO013A	Gomphonema parvulum
GO073A	Gomphonema angustum
GO080A	Gomphonema pumilum
GO9999	Gomphonema sp.
GY005A	Gyrosigma accuminatum
HN001A	Hannaea arcus
MR001A	Meridion circulare
NA002A	Navicula jaernefeltii
NA003A	Navicula radiosa
NA005A	Navicula seminulum
NA006A	Navicula mediocris
NA007A	Navicula crytopcephala
NA008A	Navicula rhynocephala
NA013A	Navicula pseudoscutiformis
NA014A	Navicula pupula
NA023A	Navicula gregaria
NA028A	Navicula scutelloides
NA030A	Navicula menisculus
NA032A	Navicula cocconeiformis
NA033A	Navicula subtilissima
NA037A	Navicula angusta
NA038A	Navicula arvensis
NA038B	Navicula arvensis var. major
NA042A	Navicula minima
NA045A	Navicula bryophila
NA051A	Navicula cari
NA063A	Navicula trivialis
NA066A	Navicula capitata
NA068A	Navicula impexa
NA073A	Navicula placentula
NA084A	Navicula atomus
NA099A	Navicula bremensis
NA112D	Navicula minuscula var muralis
NA114A	Navicula subrotundata
NA128A	Navicula schoenfeldtii
NA133A	Navicula schassmannii
NA135A	Navicula tenuicephala
NA156A	Navicula leptostriata
NA190A	Navicula agrestis
NA322A	Navicula detenta
NA462A	Navicula joubardii
NA738A	Navicula vitiosa
NA745A	Navicula capitoradiata
NA751A	Navicula cryptotenella
NA755A	Navicula kuelbsii
NA9999	Navicula sp.
NE004A	Nedium bisulcatum
NI002A	Nitzschia fonticola
NI009A	Nitzschia palea
NI014A	Nitzschia amphibia
NI015A	Nitzschia dissipata
NI017A	Nitzschia gracilis
NI033A	Nitzschia paleacea

CY052A	Cyclotella rossii
CY054A	Cyclotella krammeri
CY059A	Cyclotella cyclopuncta
DE001A	Denticula tenuis
DP007A	Diploneis oblongella
DT004A	Diatoma tenuis
DT021A	Diatoma mesodon
EL001A	Ellerbeckia arenaria
EP001A	Epithemia sorex
EU002B	Eunotia pectinalis var. minor
EU002D	Eunotia pectinalis var. undulata
EU009A	Eunotia exigua
EU011A	Eunotia rhomboidea
EU013A	Eunotia arcus
EU025A	Eunotia fallax
EU040A	Eunotia paludosa
EU047A	Eunotia incisa
EU070A	Eunotia bilunaris
EU070B	Eunotia bilunaris var. mucophila
EU105A	Eunotia subarcuatoides
EU107A	Eunotia implicata
EU108A	Eunotia intermedia
EU110A	Eunotia minor
EU9999	Eunotia sp.
FR001A	Fragilaria pinnata
FR002A	Fragilaria construens
FR002B	Fragilaria construens var. binodis
FR002C	Fragilaria construens var. venter
FR002G	Fragilaria construens var. pumilla
FR006A	Fragilaria brevistriata
FR007A	Fragilaria vaucheriae
NI152A	Nitzschia pusilla
NI193A	Nitzschia perminuta
NI9999	Nitzschia sp.
PE002A	Peronia fibula
PI007A	Pinnularia viridis
PI011A	Pinnularia microstauron
PI014A	Pinnularia appendiculata
PI022A	Pinnularia subcapitata
PI9999	Pinnularia sp.
RC002A	Rhoicosphenia abbreviata
RE001A	Reimeria sinuata
RH003E	Rhopalodia rupestris
SA001A	Stauroneis anceps
SA014A	Stauroneis gracilis
SP006A	Stenopterobia curvula
ST001A	Stephanodiscus hantzschii
ST009A	Stephanodiscus alpinus
ST010A	Stephanodiscus parvus
ST014A	Stephanodiscus medius
ST022A	Stephanodiscus neoastraea
ST9999	Stephanodiscus sp.
SU001A	Surirella angusta
SY001A	Synedra ulna
SY002A	Synedra rumpens
SY003C	Synedra acus var. angustissima
SY003F	Synedra delicatissima
SY013A	Synedra tenera/nana
TA001A	Tabellaria flocculosa
UN9995	Unknown centric
YH001A	Ctenophora pulchella

Appendix 2: Squared chord distance dissimilarity scores for the mid and surface samples in the 30 cores

Lake name	**[1]Sample code**	**[2]Squared chord distance dissimilarity score**
Hornsea Mere	HORN200	0.420
	HORN226	
The Mere, Ellesmere	ELLE00	0.419
	ELLE35	
Thirlmere	THIR100	0.482
	THIR130	
Tal-y-llyn Lake	TALY100	0.359
	TALY120	
Llyn Coron	CORO200	0.479
	CORO220	
Loch Ailsh	AILS100	0.415
	AILS140	
Loch Hope	HOPL100	0.403
	HOPL120	
Loch nan Gabhar	GABH100	0.307
	GABH115	
Loch Kinnabus	KINB100	0.418
	KINB120	
Loch Ascaig	ASCA100	0.318
	ASCA120	
Grasmere	GRAS100	0.650
	GRAS110	0.698
	GRAS127	
Llyn Egnant	EGNA100	1.113
	EGNA115	0.503
	EGNA130	
Llyn Dinam	DINA200	0.830
	DINA210	0.544
	DINA220	
Coniston Water	CONI100	0.724
	CONI120	0.712
	CONI135	
Elterwater	ELTW100	0.649
	ELTW110	0.600
	ELTW120	
Shearwater	SHEA00	0.848
	SHEA15	0.466
	SHEA25	
Fonthill Lake	FONT100	0.636
	FONT110	0.404
	FONT120	
Talkin Tarn	TALK00	0.784
	TALK20	0.330
	TALK30	
Windermere	WIND100	1.170
	WIND120	0.408
	WIND130	
Llyn Bodlyn	BODL100	0.707
	BODL110	0.316
	BODL122	
Llyn Padarn	PADA100	0.761
	PADA110	0.465
	PADA125	
Llyn Ogwen	OGWE100	0.641
	OGWE110	0.679
	OGWE125	

Lake name	[1]Sample code	[2]Squared chord distance dissimilarity score
Loch Skerrols	SKEL100	0.465
	SKEL110	0.647
	SKEL120	
Loch Craggie	CRA400	0.269
	CRA410	0.597
	CRA130	
Loch an Lagain	LAGN100	0.418
	LAGN110	0.559
	LAGN120	
Loch Shnathaid	SHNA100	0.516
	SHNA110	0.555
	SHNA120	
Loch na Moracha	MORA100	0.583
	MORA110	0.528
	MORA120	
Loch Borralan	BORL100	0.434
	BORL110	0.480
	BORL115	
Lochan Lùnn Dà-Bhrà	LUNN100	0.474
	LUNN110	0.466
	LUNN120	
Loch Tormasad	TORM100	0.891
	TORM110	0.686
	TORM120	

[1] Last 2 digits of sample code indicate sample depth (cm)

[2] Squared chord distance dissimilarity score between the core bottom sample and each other sample in that core