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Life cycle environmental and economic assessment of alumina recovery from secondary aluminum dross in China

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1 **Highlights**

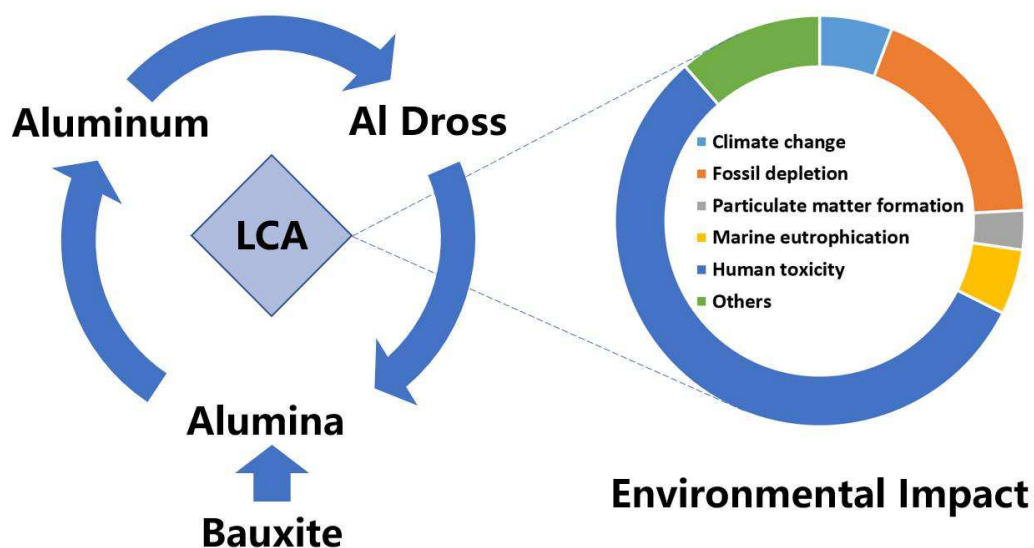
- Coupled LCA and LCC showed that producing alumina from secondary aluminum dross had lower environmental impacts and economic costs than dross process.
- Steam, sodium hydroxide and electricity contributed most to impact values.
- Suggestions for dross transportation and aluminum industry migration in China were proposed.

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3 Abstract

4 Secondary aluminum dross is regarded as a hazardous solid waste in many countries.
 5 A coupled life cycle assessment and life cycle costing method was used to evaluate
 6 the environmental impact and economic cost of two processes for producing alumina
 7 from bauxite and secondary aluminum dross. The results showed that the total
 8 normalized midpoint value of the dross process is 32.16% lower than that of the
 9 bauxite process. The cost of producing 1 t alumina by dross process is 130.01 \$,
 10 accounting for only 49.54% of that of bauxite process. Ammonium sulfate as a
 11 by-product also brought in a profit of 22.18 \$. These findings could be attributed to
 12 the decrease in energy and raw material consumption (i.e., steam, sodium hydroxide,
 13 electricity) and the relatively low cost of secondary aluminum dross. Adjusting raw
 14 materials for steam production and optimizing electricity structure could reduce the
 15 overall environmental impact of secondary aluminum dross recovery. Based on the
 16 forecast of environmental impact and policy adjustment in the future, inter-provincial
 17 dross transportation and southwest aluminum industry migration in China could be
 18 feasible solutions.



19

20 **Keywords** Life cycle assessment; Secondary aluminum dross; Recovery; Bauxite;

21 **1.Introduction**

22 In recent years, China has become the world's largest aluminum producer. During
23 the process of producing 1 t aluminum, more than 40 kg of aluminum dross is
24 generated (Meshram and Singh, 2018). In 2017, China produced 32.3 million tons of
25 primary aluminum, which means that the output of aluminum dross exceeded 1.29
26 million tons (National Bureau of Statistics of China, 2018). The component of
27 aluminum dross is determined by its source, usually containing aluminum metal,
28 aluminum oxide, iron oxide, silicon dioxide, aluminum nitride, aluminum carbide and
29 other metal oxides, chlorides and fluorides (Mahinroosta and Allahverdi, 2018b).
30 According to the number of times of aluminum dross recovery and the metal
31 aluminum content in dross, aluminum dross is usually divided into two categories:
32 primary aluminum dross and secondary aluminum dross. Secondary aluminum dross
33 has a low metal aluminum content of about 5-10 wt%, high oxide and salt content,
34 whereas primary aluminum dross contains a high metal fraction of about 30 wt% and
35 small amounts of oxide and salt compounds (Mahinroosta and Allahverdi, 2018).

36 Secondary aluminum dross is regarded as a hazardous waste in many countries.
37 In humid environment, secondary aluminum dross easily reacts with water, forming
38 flammable or toxic gases such as methane, hydrogen, and ammonia. Direct landfill
39 may let heavy metals such as Cd, Cr and Pb in the dross penetrate through soil and
40 water, causing harm to animals and plants (Mahinroosta and Allahverdi, 2018b).
41 Besides, the aluminum metal and aluminum oxide in the secondary aluminum dross
42 are both components with recovery value. The alumina resources contained in 1 t
43 secondary aluminum dross are generally equivalent to that from 1.8 t bauxite.
44 Therefore, the accumulation of aluminum dross not only causes environmental
45 pollution, but also causes loss of valuable materials.

46 At present, primary aluminum dross recovery technology is relatively mature in
47 various countries around the world. However, secondary aluminum dross recovery
48 technologies are still in the stage of laboratory exploration. Hiraki et al. (2005) used
49 aluminum dross to produce hydrogen but ignored its economic utility. Murayama et al.
50 used aluminum dross to make specific materials such as $AlPO_4-5$ type zeolitic
51 materials (Murayama et al., 2006) and Zn–Al type layered double hydroxides
52 (Murayama et al., 2012). Nevertheless, these methods cannot solve the practical
53 problem of massive dross accumulation. Mahinroosta et al. (2018a) successfully

54 extracted alumina from secondary aluminum dross in a low-energy and safe process.
55 However, the purity of alumina is hard to reach the standard. In a nutshell, these
56 existing secondary aluminum dross recovery processes have shortcomings such as
57 high cost, small scale, and low value. Fortunately, our research group successfully
58 recovered alumina from secondary aluminum dross by using sodium hydroxide,
59 which not only had good yield but also met the national quality standard (Li et al.,
60 2019; Jin et al., 2019; Song Ming, 2018).

61 In order to evaluate the environmental and economic superiority of our process,
62 life cycle assessment (LCA) coupled with life cycle costing (LCC) method is used,
63 which is an effective method to quantify the energy and materials invested in a
64 process and analyze the economic and environmental burdens caused by the process
65 (Hong et al., 2018). Unfortunately, there are very few LCA studies on aluminum dross.
66 Nakajima et al. (2007) conducted LCA of hydrogen production from aluminum dross.
67 However, the output only included waste and carbon dioxide emissions. Hong et al.
68 (2010) compared the resource consumption and waste discharge of aluminum-silicon
69 alloys production and alumina production from aluminum dross. Nevertheless, no
70 common LCA model was used to characterize the inventory, making their LCA results
71 lack systematicity and comparability. In conclusion, current LCA studies on
72 aluminum dross have serious limitations.

73 Since the process for recovering alumina from secondary aluminum dross is
74 promising and its environmental impact remains unknown, this study aims to use the
75 LCA coupled with LCC method evaluating the environmental impact and economic
76 cost of this innovative process. For comparison, the currently widely used process for
77 extracting alumina from bauxite through Bayer method was chosen as a baseline
78 scenario. The results are expected to provide data support for the industrialization of
79 secondary aluminum dross recovery. Based on the prediction of industrial adjustment
80 for China's aluminum industry, suggestions for process optimization and site selection
81 of dross recovery industry in China were proposed.

82 **2. Methodology**

83 **2.1 Goal and scope**

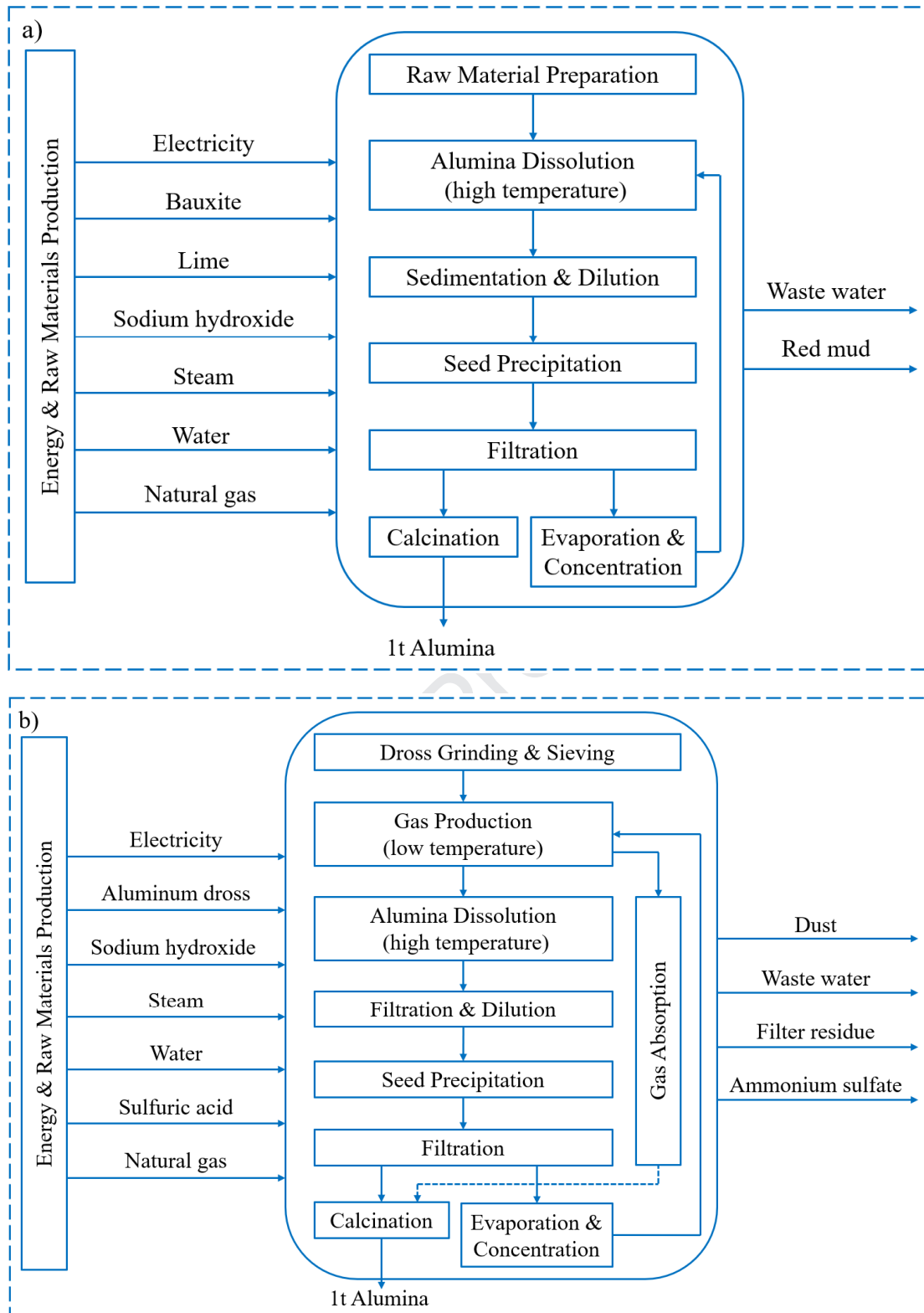
84 This study chose Bayer process as a baseline scenario to compare the
85 environmental impact and economic cost of recovering alumina from secondary
86 aluminum dross. The production of 1 t alumina was selected as the functional unit,
87 which was the base for life cycle inventory comparison. All materials, energy

88 consumption, emissions, waste disposal and economic costs were based on this
89 functional unit (International Organization for Standardization, 2006).

90 System boundaries were set by applying a cradle-to-gate approach, which only
91 focused on industrial production in the entire life cycle of alumina. Since the
92 composition of aluminum dross is close to bauxite from the perspective of resource
93 attribute, dross and bauxite were set as the start of the system boundary in two
94 scenarios. Figure 1 shows the system boundary for two processes. Since the
95 transportation of the two processes is similar and its impacts in alumina production
96 accounts for less than 1%, transportation was not included in this study.

97 For baseline scenario, bauxite is the raw material for producing alumina with
98 Bayer method. Under high temperature and high pressure, bauxite can be dissolved in
99 a solution of high concentration sodium hydroxide. After precipitation and filtration,
100 sodium aluminate solution and red mud are obtained respectively. Due to the chemical
101 nature, sodium aluminate will gradually transform into aluminum hydroxide crystals
102 through dilution and stirring. The remaining alkaline solution can be recycled after
103 evaporation and concentration. Aluminum hydroxide will be converted into alumina
104 after calcination.

105 For dross scenario, secondary aluminum dross is the raw material. The additional
106 process is 'gas production', in which metal aluminum, aluminum carbide and
107 aluminum nitride react with sodium hydroxide at 90 °C, producing hydrogen, methane
108 and ammonia. Ammonia is absorbed by sulfuric acid to obtain ammonium sulfate in
109 another unique 'acid absorption' unit. Hydrogen and methane are used as fuel for
110 'calcination' unit, achieving energy recovery. Other processes in these two scenarios
111 were similar, in which dosage or condition may be slightly different. The alumina
112 from dross is dissolved by sodium hydroxide at 250 °C. After filtration and dilution,
113 clear sodium aluminate solution would be obtained. After being diluted 2.5 times and
114 stirred at room temperature for 72 hours, sodium aluminate solution will precipitate
115 out aluminum hydroxide. The remaining alkaline solution can also be recycled after
116 evaporation and concentration. After calcination, aluminum hydroxide will be
117 converted to alumina. Under the above conditions, the recovery rate of alumina in
118 secondary aluminum dross can reach 88.20% (Li et al., 2019). The entire process will
119 produce dust, wastewater, and filter residue.



120

121

122 **Fig. 1.** System boundary of two processes: a) bauxite scenario, b) dross scenario.

123 2.2 Life cycle inventory

124 Life cycle inventory data for the bauxite process was from the average data of
 125 relative industries in China. Data for the secondary aluminum dross process was
 126 mainly from the average data of our experimental results (Li et al., 2019; Jin et al.,

127 2019; Song Ming, 2018). Since some processes of two scenarios were identical, some
 128 data for energy consumption and emissions also referred to bauxite scenario. The
 129 electricity type in this study was a hybrid electricity based on China's national
 130 conditions (70.99% thermal power, 18.59% hydropower, and 10.42% other forms of
 131 power) (Yu Chongde, 2018). The steam used in this study was coal based. In addition,
 132 the environmental impact data was from the latest version 3.6 Ecoinvent database
 133 integrated in GaBi 6.0 software. The cost of raw materials referred to market prices.

134 According to EN ISO 14040 standard, the life cycle inventory is an inventory of
 135 the input/output data of the processes (International Organization for Standardization,
 136 2006). Table 1 displays the inventory of two scenarios.

137 **Table 1**

138 Life cycle inventory

	Materials	Bauxite scenario	Dross scenario	Units
Resources (Input)	Bauxite	2.48×10^3	0	kg
	Secondary aluminum dross	0	1.42×10^3	kg
	Lime	3.20×10^1	0	kg
	Sodium hydroxide	6.15×10^1	5.70×10^1	kg
	Steam	2.43×10^3	2.21×10^3	kg
	Water	2.00×10^0	1.24×10^0	t
	Electricity	1.69×10^2	2.14×10^2	kwh
	Natural gas	7.25×10^1	2.34×10^1	m ³
	Sulfuric acid	0	3.01×10^2	kg
Emissions (Output)	Waste water	7.91×10^2	3.13×10^2	kg
	Filter residue	7.58×10^2	2.89×10^2	kg
	Dust	2.90×10^{-1}	1.58×10^{-2}	kg
	Ammonia	1.25×10^{-2}	8.93×10^{-2}	kg
	Chromium	1.32×10^{-3}	1.32×10^{-3}	kg
	Bromine	1.15×10^{-3}	1.14×10^{-3}	kg
	Carbon monoxide	1.16×10^0	8.60×10^{-1}	kg
	Hydrogen fluoride	1.57×10^{-3}	1.47×10^{-3}	kg
	Hydrogen sulfide	6.44×10^{-3}	3.57×10^{-3}	kg
	Sulfur dioxide	3.41×10^0	2.77×10^0	kg
Propane	2.52×10^{-2}	1.99×10^{-3}	kg	
Xylene	9.17×10^{-2}	9.16×10^{-2}	kg	

Particulate matters 10	1.61×10^{-1}	3.74×10^{-6}	kg
Particulate matters 2.5	1.03×10^0	9.90×10^{-1}	kg
Biochemical oxygen demand	9.52×10^{-3}	1.39×10^{-3}	kg
Chemical oxygen demand	2.60×10^{-1}	2.60×10^{-1}	kg

139 2.3 Impact assessment

140 On the one hand, since secondary aluminum dross is an emerging issue in recent
 141 years, there is no industrialized treatment process currently. The process of recovering
 142 alumina from secondary aluminum dross is promising, but it is still in the early stage.
 143 Therefore, the LCA of dross process in this study is an ex-ante type. On the other
 144 hand, the bauxite process is very mature and most units and materials of both
 145 processes are the same, which helps to estimate the data of dross process based on
 146 industrial scale. Overall, in order to evaluate the energy consumption and emissions of
 147 the two processes, the comparative LCA study of the two processes is an attributional
 148 type.

149 The LCA results were calculated at midpoint level using the version 1.08 of
 150 ReCiPe 2008 model, which is one of the most authoritative approaches in LCA
 151 analysis, including eighteen representative environmental impact categories. The
 152 characterization factors were based Ecoinvent database 3.6 integrated in GaBi 6.0
 153 software. The reference values for normalization were the global midpoint values for
 154 ReCiPe 2008 model, in which the normalization factors were updated in December
 155 2014 (Goedkoop et al., 2014; Sleeswijk et al., 2008). The costs of two processes were
 156 assessed through LCC method. The LCC method is similar to LCA, wherein the
 157 evaluation considers the cost of energy and raw materials listed in the inventory
 158 instead of the environmental impacts. Since the gap of costs of labor in the raw
 159 material, energy production and manufacture stages between two processes is small,
 160 those costs were not included in this paper. The LCC results were calculated based on
 161 the price and amount of materials.

162 2.4 Interpretation

163 Main contributing processes and key substances were identified through life
 164 cycle interpretation. In dross scenario, since the environmental impact and economic
 165 cost caused by alumina production accounted for more than 95% of the entire process,
 166 the allocation of environmental and economic burden from ammonium sulfate
 167 production was not considered for the convenience of this work. In section 4.3.1, the

168 benefit of using natural gas based steam was discussed. Two new scenarios were
169 assumed: (1) Using natural gas as raw material to produce steam at 85% efficiency; (2)
170 Using natural gas as raw material to produce steam at 95% efficiency. In section 4.3.2,
171 the adjustment of electricity structure was proposed. Since Henan and Shandong, the
172 two largest alumina production provinces in China, strongly depend on thermal power,
173 the power structure of Qinghai Province (24.8% thermal power, 53.9% hydro power,
174 18.3% solar power, 2.9% of wind power) was taken as an example to analyze the
175 reduction of environmental impacts. In section 4.4.1, inter-provincial dross
176 transportation was proposed and its benefits and impacts were discussed. In section
177 4.4.2, suggestion for the south and southwest migration of the aluminum industry in
178 China was discussed.

179 **3. Results**

180 **3.1 LCA midpoint results**

181 Table 2 shows the midpoint results of life cycle impact assessment pointing out
182 the contribution of most significant processes. For both scenarios, steam consumption
183 represented dominant contribution in most categories (usually over 50%), indicating
184 that alumina production is an energy-intensive industry. Furthermore, the dross
185 scenario had less potential impact in all categories and the improvements in most
186 categories was over 30%, which was mainly due to different alumina content in
187 bauxites and dross. Concretely speaking, the aluminum in bauxite is relatively less
188 than secondary aluminum dross, resulting in higher material input and energy
189 consumption during the extraction. In addition, the characteristic value of metal
190 depletion category in secondary aluminum dross process was -0.257. That's because
191 the dross process uses industrial residual as a raw material, avoiding the consumption
192 of bauxite. Another interesting point was that electricity was not found to dominate
193 the environmental impact, for example, its contribution to climate change was only
194 17.52%. That was because the amount of electricity used in both processes was much
195 lower than the amount of steam. Thus, steam replaced electricity as the main
196 contributor to the climate change category.

197 **Table 2** ReCiPe midpoint results for two processes

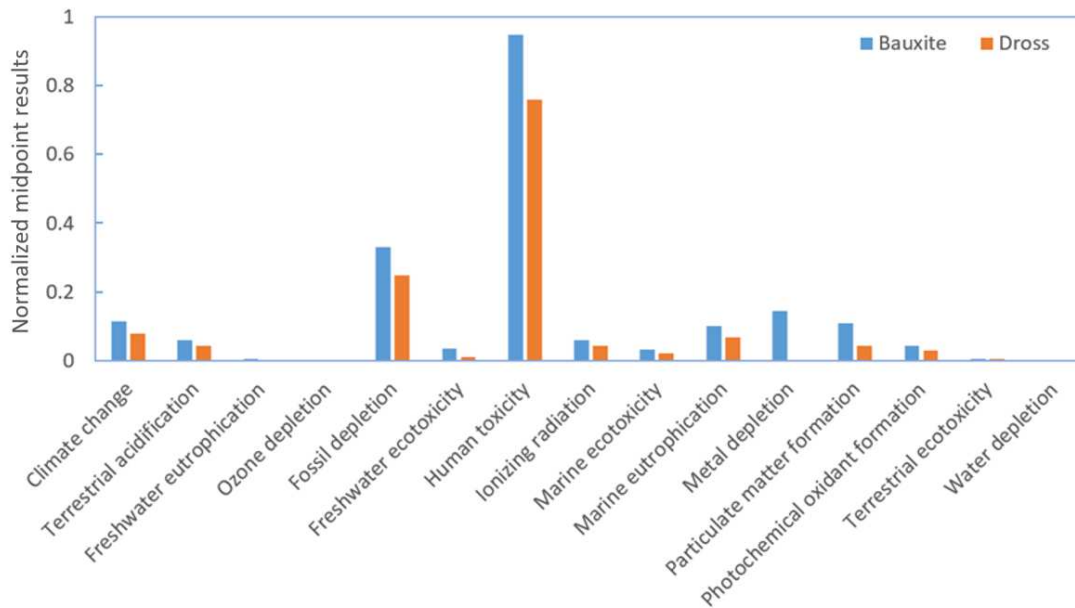
Categories	Units	Bauxite scenario		Dross scenario		Improve ment %
		Results	Main contributing process	Results	Main contributing process	
Climate change	kg CO ₂ eq	1.09×10 ⁵	Steam (55.5%) + NaOH (23.1%)	7.36×10 ²	Steam (68.1%)	32.48
Terrestrial acidification	kg SO ₂ eq	2.16×10 ⁰	Steam (47.8%) + NaOH (21.2%)	1.52×10 ⁰	Steam (52.4%) + H ₂ SO ₄ (31.2%)	29.63
Freshwater eutrophication	kg P eq	1.60×10 ⁻³	NaOH (88.1%)	2.86×10 ⁻⁴	NaOH (46.9%) + H ₂ SO ₄ (29.7%)	82.13
Ozone depletion	kg CFC-11 eq	3.13×10 ⁻¹⁰	NaOH (87.5%)	3.09×10 ⁻¹⁰	NaOH (39.6%) + H ₂ SO ₄ (37.9%)	1.28
Fossil depletion	kg oil eq	4.26×10 ²	Steam (55.0%)	3.21×10 ²	Steam (52.5%)	24.65
Freshwater ecotoxicity	kg 1,4-DB eq	1.55×10 ⁻¹	NaOH (71.0%)	4.87×10 ⁻²	Steam (34.2%) + NaOH (24.4%)	68.58
Human toxicity	kg 1,4-DB eq	7.86×10 ¹	Steam (48.1%)	6.29×10 ¹	Steam (77.8%)	19.97
Ionizing radiation	U235 eq	2.62×10 ¹	NaOH (92.0%)	1.92×10 ¹	H ₂ SO ₄ (47.9%) + NaOH (40.3%)	26.72
Marine ecotoxicity	kg 1,4-DB eq	6.74×10 ⁻²	Steam (34.6%) + NaOH (35.9%)	4.23×10 ⁻²	Steam (55.0%)	37.24
Marine eutrophication	kg N eq	7.40×10 ⁻¹	Steam (50.3%) + NaOH (27.1%)	5.10×10 ⁻¹	Steam (68.4%)	31.08
Metal depletion	kg Fe eq	6.46×10 ¹	Bauxite (91.3%)	-2.57×10 ⁻¹	Dross (100.0%)	100.40
Particulate matter formation	kg PM ₁₀ eq	1.53×10 ⁰	Steam (36.6%) + Bauxite (28.2%) + Electricity (22.8%)	5.90×10 ⁻¹	Steam (61.9%)	61.44
Photochemical oxidant formation	kg NMVOC eq	2.11×10 ¹	Steam (52.2%) + NaOH (23.8%)	1.43×10 ¹	Steam (65.1%)	32.23
Terrestrial ecotoxicity	kg 1,4-DB eq	4.03×10 ⁻²	Steam (54.7%)	3.39×10 ⁻²	Steam (79.7%)	15.88
Water depletion	m ³	6.11×10 ²	NaOH (78.6%)	5.24×10 ²	NaOH (35.7%) + Electricity (36.5%)	14.24

198 3.2 Normalized LCA midpoint results

199 In order to compare the differences between various categories, characteristic
200 values need to be normalized. As shown in Figure 2, human toxicity accounted for the
201 largest proportion in the normalized value of both processes, which was caused by a
202 variety of reasons: (1) Both processes produced particulate matters and harmful gases
203 such as hydrogen sulfide, sulfur dioxide, etc. (2) Harmful components such as
204 chromium and xylene were present in the waste water. (3) The amount of solid waste
205 was considerable, especially the red mud from bauxite process. (4) Alumina extraction
206 required high consumption of electricity, steam, and natural gas. During the
207 production of these energy, various toxic and hazardous substances were produced,
208 which was the hidden and main cause of high human toxicity value. Compared with
209 the proportion of 47.6% in bauxite scenario, the normalized human toxicity value of
210 dross process accounted for 63.0% of the total value. However, the actual human
211 toxicity normalization value was lower than that of the bauxite process, indicating that
212 the total values of all categories in dross scenario was significantly lower.

213 In addition, the normalized values of the two processes in the fossil depletion
214 category were apparently different. Fossil depletion was mainly caused by electricity
215 and steam consumption. Since the alumina that can be recovered from per unit mass
216 of dross was higher than that from per unit mass of bauxite, the energy requirement of
217 dross scenario was relatively low under the same output. Furthermore, according to
218 the dross process proposed by our group, hydrogen and methane generated by
219 aluminum and aluminum carbide was used as fuels in the calcination unit, which
220 made up part of the energy demand.

221 Overall, the total normalized values of bauxite process and dross process were
222 1.99 and 1.35 respectively, indicating that producing alumina by dross process had
223 better environmental benefits.



224

225

Fig. 2. Normalized midpoint results

226

3.3. Sensitivity analysis

227

Figure 3 presented the sensitivity analysis results of main contributors (steam, sodium hydroxide, and sulfuric acid) for both processes. The amounts of these main contributors were reduced by 5%.

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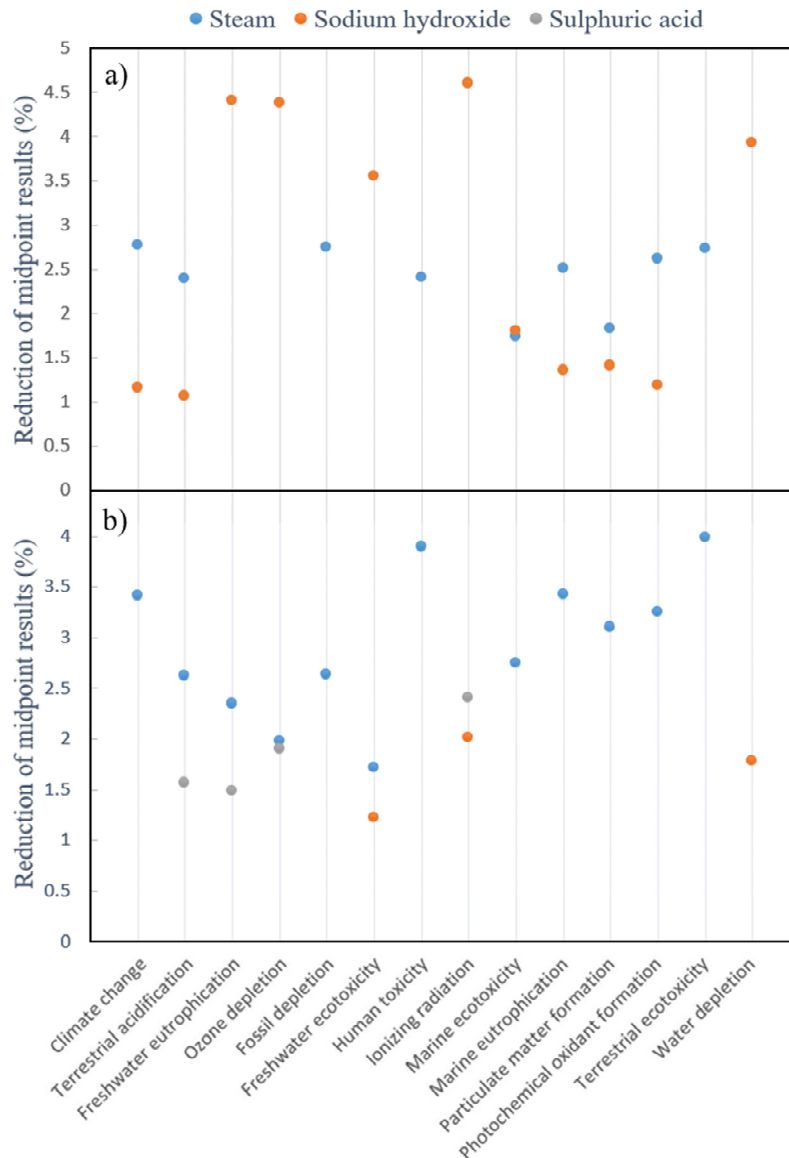
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For bauxite scenario, despite the high sensitivity of sodium hydroxide in freshwater eutrophication, ozone depletion, ionizing radiation, and water depletion, steam showed the highest sensitivity in other categories. For dross scenario, due to the reduced use of sodium hydroxide and the introduction of sulfuric acid, the sensitivity of sodium hydroxide to the above categories was reduced absolutely by 1% to 3%, partly replaced by sulfuric acid. Steam showed the highest sensitivity in 12 categories. However, this highlight was not due to the amount of steam but the decrease in the sensitivity of other materials. While the consumption of other resources reduced, steam consumption still accounted for 90.9% of the consumption in bauxite process, making its sensitivity more pronounced. As for electricity and water, neither of the two scenarios showed high sensitivity. Therefore, reducing steam demand or using renewable energy is the key to reducing the overall environmental impact of dross process.



243
244 **Fig. 3.** Sensitivity analysis: a) bauxite scenario, b) dross scenario.

245 3.4 LCC results

246 The LCA results proved that the environmental impact of dross process is less
247 than that of bauxite process. In order to find out whether dross process is also superior
248 in economic term, the raw material cost and energy cost of two processes were
249 compared in Table 3. In the calculation, the exchange rate of USD to RMB is set to 1
250 \$ equal to 7.136 ¥.

251 The total cost of producing 1 t alumina from bauxite was 262.46 \$, while the cost
252 of dross process was 130.01 \$, accounting for only 49.5% of the bauxite scenario.
253 This gap was mainly caused by the difference in raw material prices. Since secondary
254 aluminum dross was regarded as an industrial solid waste and the dross used in the

255 experiment was freely donated by the enterprise, the price of the secondary aluminum
 256 dross was set as 0 \$. In fact, due to the dangerous properties of secondary aluminum
 257 dross, ordinary companies do not have the processing qualifications (Meshram and
 258 Singh, 2018), and even need to spend money to ask other qualified organizations to
 259 properly handle the dross. In addition, when producing 1 t alumina, dross process
 260 obtained 406 kg ammonium sulfate as a by-product, with a profit of 22.18 \$, which
 261 was higher than the cost of sulfuric acid, water, and electricity. The absorption of
 262 ammonia by sulfuric acid not only reduced its pollution to the atmosphere, but also
 263 achieved economic benefits.

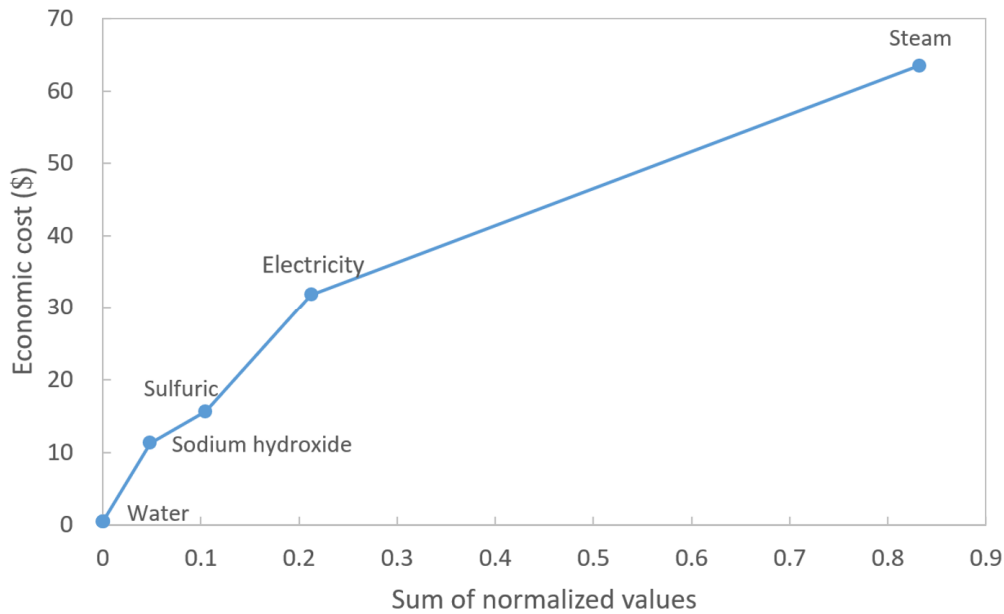
264 **Table 3**

265 LCC results of two processes

Materials	Price	Units	Cost [€]	
			Bauxite scenario	Dross scenario
Bauxite	52.971	\$/t	131.37	0
Secondary aluminum dross	0	\$/t	0	0
Lime	11.911	\$/t	0.38	0
Sodium hydroxide	197.590	\$/t	12.15	11.26
Steam	28.728	\$/t	69.81	63.49
Water	0.392	\$/t	0.78	0.49
Electricity	0.149	\$/kWh	25.10	31.79
Natural gas	0.315	\$/m ³	22.86	7.38
Sulfuric acid	51.850	\$/t	131.37	15.61
Total			262.46	130.01

266 Figure 4 shows the environmental impact coupled with economic costs of the
 267 key material used in dross process. The Y axis shows the economic cost of water,
 268 sodium hydroxide, sulfuric acid, electricity, and steam. The X axis represents the sum
 269 of normalized values of all environmental impact categories for each material. As can
 270 be seen, the polyline continues to extend to the upper right, meaning that for dross
 271 scenario, the environmental loads of materials are proportional with the economic
 272 costs. The use of steam brought the largest environmental impact and consumed
 273 highest economic costs. Environmental pollution and economic costs caused by
 274 electricity cannot be ignored neither. Therefore, reducing the use of steam and

275 electricity or using clean energy could be the key to reducing the environmental
 276 impacts as well as economic costs of secondary aluminum process.



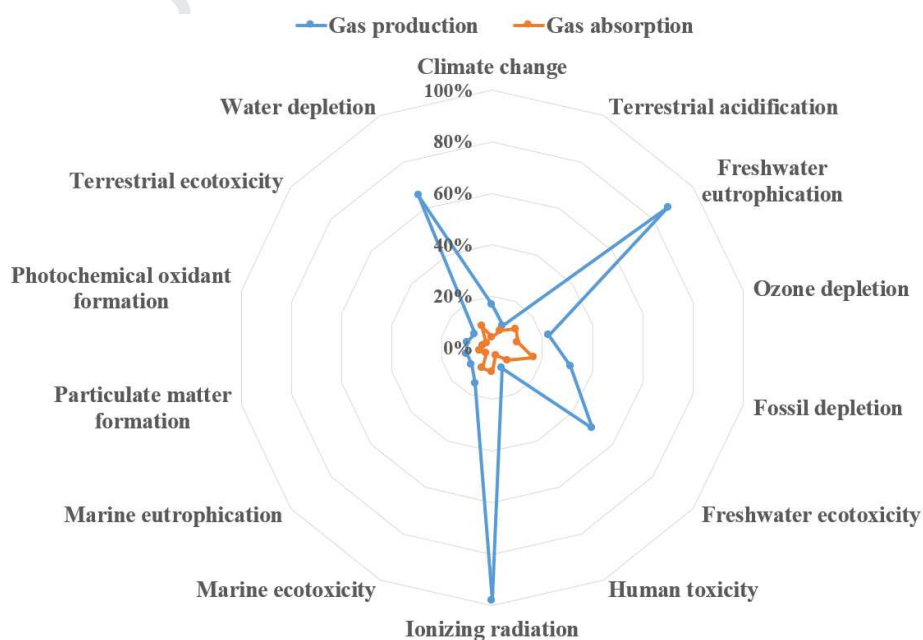
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278 **Fig. 4.** Environmental impact coupled with economic costs analysis for dross process

279 4. Discussion

280 4.1 Unit analysis

281 Due to the differences in the partial process of two scenarios, the gas production
 282 unit and the gas absorption unit are unique in dross process. To further optimize the
 283 environmental benefits of the dross process, it is necessary to identify the
 284 environmental impact of these two units. Figure 5 shows the ratio of the LCA
 285 midpoint values of these two units to the total values of the entire process.



286

287 **Fig. 5.** Proportion of the characteristic values of two unique units

288 In the gas absorption unit, the proportion was relatively small because of the low
289 material and energy input in this unit. Fossil depletion had the highest proportion of
290 16%. As the main input of this unit, sulfuric acid used in the process was produced
291 from pyrites. During the production of pyrites, high-temperature processing and
292 catalytic heating reaction were required, leading to high energy consumption (Yuan
293 and Wang, 2012). Together with the input of electricity and water, fossil depletion
294 accounted for the largest proportion.

295 In the gas production unit, freshwater eutrophication, freshwater ecotoxicity,
296 ionizing radiation, water depletion and fossil depletion accounted for a large
297 proportion of the total characteristic values. The characteristic value of ionizing
298 radiation in gas production unit accounted for 98.0% of that value of the whole
299 process, which was mainly caused by the consumption of sodium hydroxide for
300 dissolving secondary aluminum dross. In addition, sodium hydroxide also exacerbated
301 freshwater eutrophication and ecotoxicity to some extent. Furthermore, in order to
302 prepare the sodium hydroxide solution and reach the reaction temperature, this unit
303 also consumed a large amount of water as well as electricity, resulting in large
304 proportion of water and fossil depletion. However, despite of the environmental
305 hazards caused by sodium hydroxide in dross process, the bauxite process also
306 required large amounts of sodium hydroxide, which was even 8% higher (Zhang et al.,
307 2016).

308 **4.2 Key substances identification**

309 Identifying key substances emitted by the two processes is beneficial to better
310 analysis of their environmental impact. Figure 6 shows the key substances produced
311 by two processes that mostly affecting the climate change and fossil depletion. Since
312 carbon dioxide is the main cause of global warming, the key substances of climate
313 change for both processes were identified as carbon dioxide, with a proportion over
314 92%. Methane and Nitrous oxide accounted for about 2% of the contribution.
315 Although the distributions of key substances in two scenarios were similar, the
316 characteristic value of dross process was lower, therefore resulting less greenhouse
317 gas emissions. Additionally, the key substances of fossil depletion in two scenarios
318 were quite different. On the one hand, high energy consumption of bauxite process led
319 to the high proportion of natural gas. On the other hand, secondary aluminum dross
320 process consumed 44.86 kWh more electricity than the bauxite process when

321 producing 1 t alumina, leading to higher contribution of coal based on China's current
 322 power generation mode. As the backward small power plants gradually shutting down,
 323 the energy source of electricity will be much cleaner (Cui et al., 2012; Wang et al.,
 324 2019). Furthermore, it can be reasonably inferred that if dross process can be
 325 industrialized after years, electricity consumption and overall energy demand will be
 326 significantly reduced (Zhang et al., 2015). Technology development will bring 50%
 327 reduction in greenhouse gas emission factors after 10 years (Liu et al., 2016).

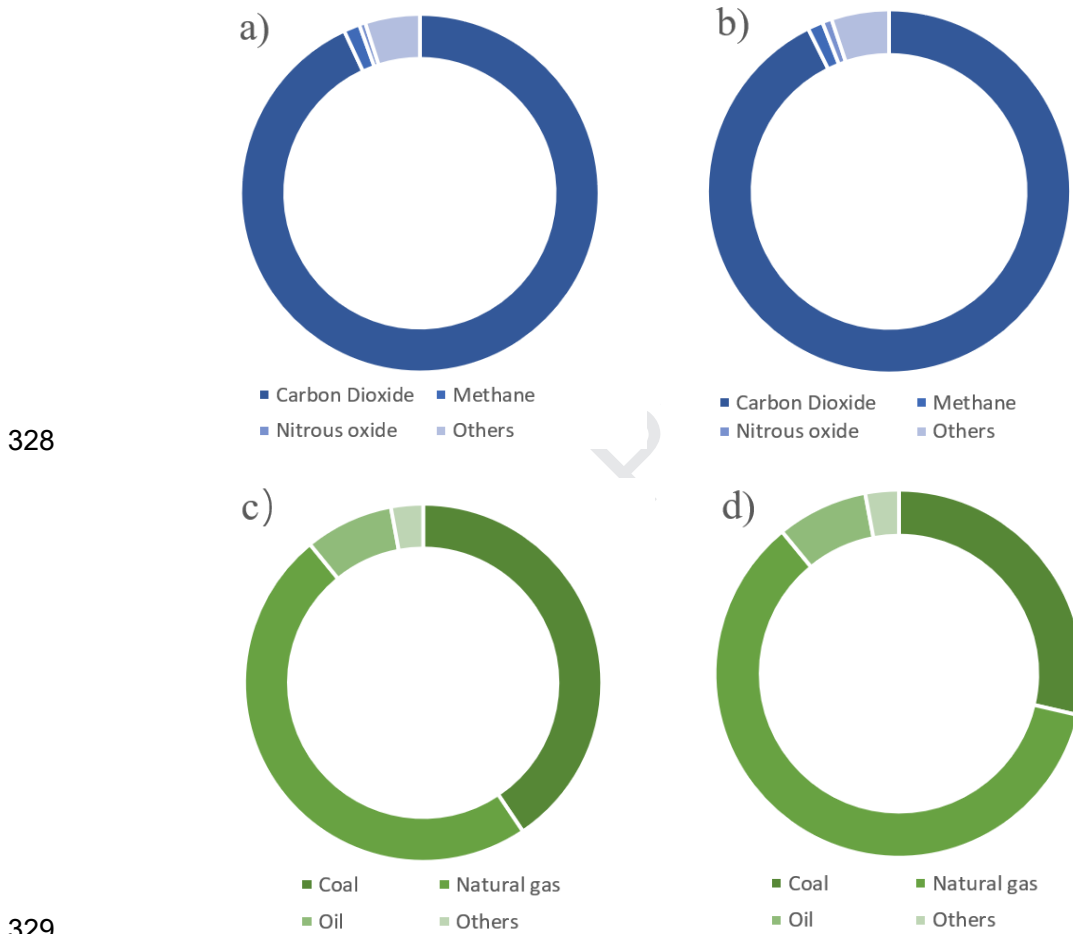


Fig. 6. Contribution of significant substances to the midpoint score: a) Climate change from dross scenario, b) Climate change from bauxite scenario, c) Fossil depletion from dross scenario, d) Fossil depletion from bauxite scenario.

4.3 Scenario analysis

Compared with the bauxite scenario, dross scenario represents lower economic costs and higher environmental benefit in all categories. According to the coupled LCC and LCA in Figure 4, the environmental and economic burden of dross scenario was mainly due to the strong dependence on steam and electricity. In order to further reduce the environmental impact caused by dross process, adjusting its energy

339 structure was proposed.

340 **4.3.1 Steam material replacement**

341 Table 2 identified the key processes in most categories as steam, which was
 342 produced from coal at 95% efficiency in the original calculation. As mentioned in
 343 section 2.4, two new scenarios using natural gas based steam were proposed. Table 4
 344 shows the changes of characteristic values of 10 categories that are most affected by
 345 steam.

346 When using natural gas as raw material to produce steam at 85% efficiency, the
 347 characteristic values of all environmental categories except fossil depletion
 348 significantly decreased. When utilization rate increased from 85% to 95%, the
 349 characteristic values were further decreased. Human toxicity, particulate matter
 350 formation and terrestrial ecotoxicity showed the largest decline, all of which exceeded
 351 50%. Since the characterization factor of natural gas in the fossil depletion category is
 352 slightly higher than that of coal (Steubing et al., 2016), fossil depletion showed a
 353 small increase of 9.91%. However, denying the use of natural gas based on only one
 354 indicator is unreasonable. From other evaluation indicators of the ReCiPe model
 355 (Goedkoop et al., 2014; Sleeswijk et al., 2008), the small increase in fossil depletion
 356 was followed by a tremendous improvement in the environment and human health.
 357 Obviously, the advantages of using natural gas as the steam material outweighed the
 358 disadvantages. Besides, if the utilization efficiency of steam is improved, the
 359 characteristic value of fossil depletion may also decrease. Furthermore, changes in
 360 steam production materials will also lead to changes in economic costs. Producing 1 t
 361 steam at 95% efficiency consumes about 75.82 m³ natural gas, 21.62 kWh electricity,
 362 and 0.17t water (Althaus et al., 2007). Based on the production of 1 t alumina,
 363 compared with the original steam cost of 63.49 \$, the cost of producing steam with
 364 natural gas would be 60.05 \$. Therefore, while the environmental impact was greatly
 365 reduced, the economic cost also reduced by 2.65%.

366 **Table 4**

367 Changes of characteristic values in new scenarios

Categories	Changes of Characteristic Values/%	
	85% Natural Gas	95% Natural Gas
Climate change	-14.65	-20.20
Terrestrial acidification	-42.00	-43.00

Fossil depletion	19.44	9.91
Freshwater ecotoxicity	-27.20	-27.92
Human toxicity	-76.70	-76.80
Marine ecotoxicity	-49.08	-49.69
Marine eutrophication	-31.34	-35.20
Particulate matter formation	-53.85	-54.95
Photochemical oxidant formation	-30.98	-34.51
Terrestrial ecotoxicity	-76.85	-76.95

368 4.3.2 Electricity structure adjustment

369 In dross scenario, the environmental impact of electricity ranked second only to
 370 steam. Henan, Shandong and Qinghai are the three provinces with the highest primary
 371 aluminum production in China (Hao et al., 2016). As mentioned in section 2.4,
 372 assuming the electricity composition of dross recovery follows the example of
 373 Qinghai, the reduction in environmental impacts is shown in Table 5.

374 **Table 5**

375 Improvement of characteristic values under electricity structure adjustment

Categories	Reduction /%
Climate change	12.58%
Terrestrial acidification	9.94%
Freshwater eutrophication	4.34%
Ozone depletion	14.34%
Fossil depletion	10.07%
Freshwater ecotoxicity	7.70%
Human toxicity	14.30%
Ionizing radiation	5.75%
Marine ecotoxicity	10.76%
Marine eutrophication	13.60%
Metal depletion	0.19%
Particulate matter formation	11.42%
Photochemical oxidant formation	12.83%
Terrestrial ecotoxicity	14.36%
Water depletion	10.60%

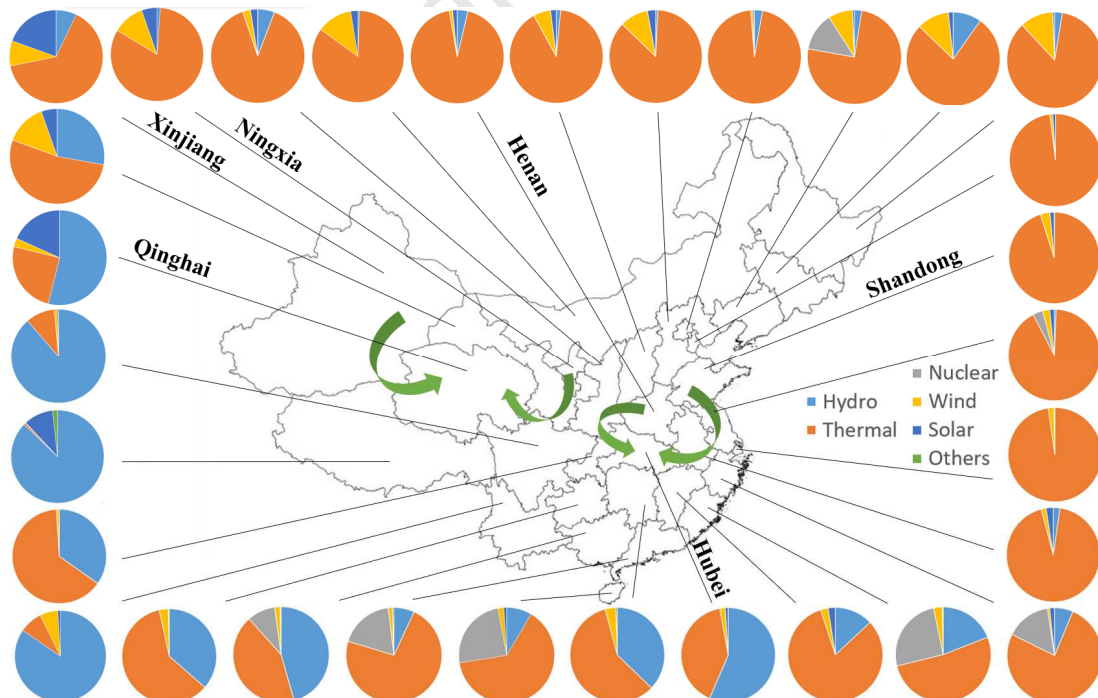
376 After adjusting the electricity structure, all environmental categories showed a

377 certain degree of improvement, mostly over 10%. The reason why the improvement
 378 of metal depletion is extremely slight was because the main process in this category is
 379 aluminum dross. Since the recovery of second aluminum dross has already achieved
 380 environmental benefits in metal depletion, the current result is acceptable. Therefore,
 381 if these high aluminum production provinces want to recover alumina from secondary
 382 aluminum dross, increasing the proportion of hydro power could be an effective way
 383 to solve environmental problems. However, since the aluminum industry is already
 384 mature, it is not easy to change the local electricity structure. In this case, dross
 385 transportation could be a feasible solution.

386 4.4 Industrial recommendations

387 4.4.1 Dross transportation

388 Transporting secondary aluminum dross from nearby provinces that heavily rely
 389 on thermal power to hydropower-type provinces for further recovery could also be a
 390 beneficial suggestion. Figure 7 shows the power structure of each province in China.
 391 As can be seen, Qinghai could be the transport destination for aluminum dross in
 392 Xinjiang and Ningxia, which are also two provinces with high aluminum production.
 393 Hubei could be the transport destination for aluminum dross in Henan and Shandong.

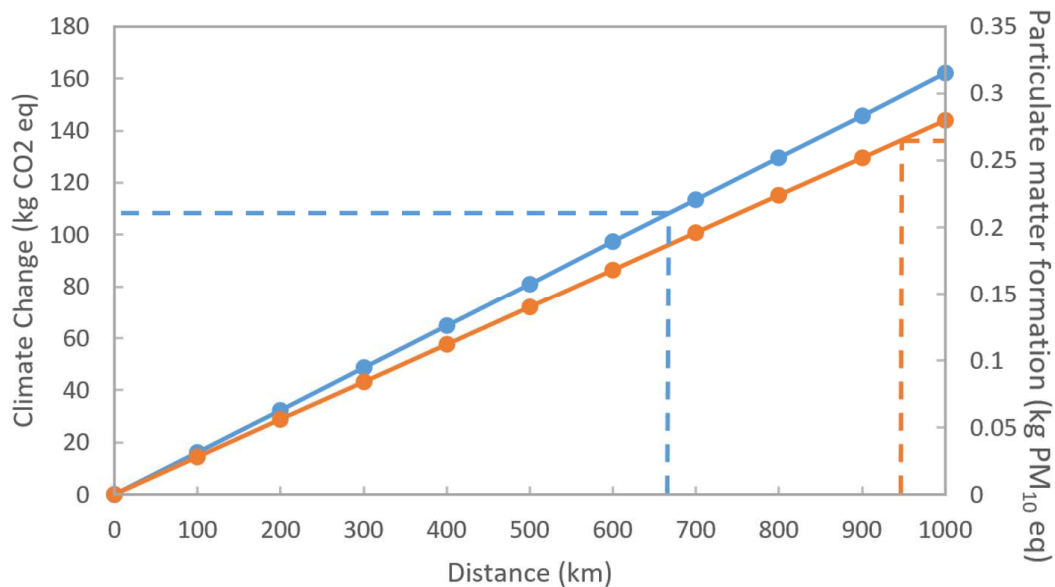


394
 395 **Fig. 7.** The power structure of each province in China

396 Long-distance transportation may cause additional environmental problems,
 397 especially climate change and particulate matter formation (Zhang et al., 2016).
 398 Transportation distance is the most important factor (Fan et al., 2018). For the case of

399 dross transportation, as the distance increases, the greenhouse gas emissions will
 400 gradually increase, eventually exceeding the environmental benefits brought by
 401 hydropower. According to the power structure of Qinghai and Ningxia (Yu Chongde,
 402 2018), hydropower replacement can reduce the characteristic value of climate change
 403 by 107.7 kg CO₂ eq when producing 1 t alumina. As shown in Figure 8, when dross is
 404 transported over 664 kilometers, electricity structure adjustment will no longer be a
 405 wise choice. Similarly, when the distance exceeds 945 kilometers, the particulate
 406 matter discharged from transportation exceeds the improvement from electricity
 407 structure adjustment (0.2646 kg PM₁₀ eq).

408 Furthermore, Qinghai Province has the lowest industrial electricity price in China
 409 at 0.053 \$/kWh. The industrial electricity price in Hubei is similar to that in Shandong
 410 and Henan (Yu Chongde, 2018). If using the average hydropower price in Hubei and
 411 Qinghai as the standard, the electricity cost can be reduced by 5.64 \$ when producing
 412 1 t alumina. According to trucks' general energy consumption and diesel prices in
 413 China (Ministry of Transport of the People's Republic of China, 2018), the cost of 100
 414 km transportation is about 0.717 \$/t. When the transportation distance exceeds 554
 415 kilometers, the economic cost of dross transportation will exceed the original plan.
 416 Therefore, from the perspective of economic cost, transporting the dross to Qinghai
 417 and Hubei has limited benefits but acceptable.



418

419 **Fig. 8.** Relationship between environmental impact and transportation distance

420

4.4.2 Industry migration

421

According to bauxite and alumina statistics information from United States

422 Geological Survey (United States Geological Survey, 2019), compared with other
423 countries, China is facing a dilemma of low reserves and high demand for bauxite.
424 Such situation requires China to increase the utilization of bauxite, which means that
425 the aluminum industry must pay attention to the recovery of million tons of tailings
426 such as secondary aluminum dross. Since dross recovery is the end industry of
427 aluminum production and it has not been industrialized, the site selection should
428 particularly focus on the future trends of aluminum industry.

429 Previous analysis showed that hydropower could bring significant environmental
430 benefits for dross recovery. Therefore, hydropower-type provinces will be suitable
431 construction sites. Similar perspectives were obtained according to the LCA of
432 China's aluminum industry (Guo et al., 2019; Hao et al., 2016). At the provincial level,
433 industry migration to south and southwest areas (hydropower-type provinces) was
434 reasonable from the perspective of environmental pollution.

435 According to the mineral resources report from Ministry of Natural Resources of
436 China (Ministry of Natural Resources of the People's Republic of China, 2019), as the
437 demand for bauxite increases year by year, Henan and Shanxi, which provides bauxite
438 for Shandong, is facing shortage of resources and decline in the bauxite quality.
439 Evaluation of China's bauxite potential indicated that more than 100 million tons high
440 quality laterite type bauxite was discovered in Guangxi Province in recent years.
441 Other southern provinces such as Yunnan, Guizhou and Guangdong were believed
442 having huge exploration potential. The distribution of bauxite shows that aluminum
443 industry will likely migrate from traditional industrial provinces such as Henan and
444 Shandong to southern provinces in the future.

445 Both policy and environmental factors suggest the south and southwest migration
446 of the aluminum industry. Based on a comprehensive forecast of environmental
447 impact and policy adjustment, Yunnan and Guangxi are the most suitable destinations
448 for the aluminum industry migration and secondary aluminum dross recovery.

449 **5. Conclusion**

450 In this paper, LCA method is used to compare the environmental impacts and
451 economic costs of two processes for producing alumina from secondary aluminum
452 dross as well as bauxite. Both characteristic values and normalized values of all
453 environmental impact categories of the secondary aluminum dross process are lower
454 than the bauxite process. LCC results showed that dross process could reduce the cost
455 of 132.45 \$ by producing 1 t alumina compared to bauxite process. The characteristic

456 value caused by sodium hydroxide accounts for the largest proportion of the unique
457 gas production unit in dross scenario, while it's still 8% lower than bauxite scenario.

458 The LCA coupled with LCC results showed that the use of steam and electricity
459 were the keys to reducing the environmental impacts as well as economic costs of
460 dross process. Using natural gas as raw material to produce steam instead of coal can
461 significantly reduce the environmental impact of the whole dross process. Increasing
462 the use of hydropower in China's high aluminum production provinces can generally
463 bring 10% environmental benefits for dross process. Dross transportation within 591
464 km or south and southwest migration of aluminum industry could be feasible
465 solutions in China.

466 This study provides data support for the industrialization of secondary aluminum
467 dross recovery. However, the current study has several limitations. Some data of the
468 dross process were based on experiments, which might be small changes in actual
469 industrial production. Some background data were selected from the European
470 database, which may have slight difference from the actual situation in China. Thus,
471 further research on secondary aluminum dross recovery is necessary.

472 **Author Contributions**

473 Jin.Q. and Zhu.X. performed the experiments and designed the life cycle assessment.
474 All authors collected and analyzed the data. Zhu.X. wrote the manuscript with
475 contributions of all the coauthors. All authors have given approval to the final version
476 of the manuscript.

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479 **Notes**

480 The authors declare no competing interests.

481 **REFERENCES**

482

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- 484 Cui, X., Hong, J., Gao, M., 2012. Environmental impact assessment of three
485 coal-based electricity generation scenarios in China. *Energy* 45(1), 952-959.
- 486 Fan, Y.V., Perry, S., Klemeš, J.J., Lee, C.T., 2018. A review on air emissions
487 assessment: Transportation. *J. Clean. Prod.* 194, 673-684.
- 488 Goedkoop M.J., Heijungs R., Huijbregts M., De Schryver A.; Struijs J., Van Zelm R.,
489 2014. ReCiPe 2008, A life cycle impact assessment method which comprises
490 harmonised category indicators at the midpoint and the endpoint level.
- 491 Guo, Y., Zhu, W., Yang, Y., Cheng, H., 2019. Carbon reduction potential based on life
492 cycle assessment of China's aluminium industry-a perspective at the province level. *J.*
493 *Clean. Prod.* 239.
- 494 Hao, H., Geng, Y., Hang, W., 2016. GHG emissions from primary aluminum
495 production in China: Regional disparity and policy implications. *Appl. Energy* 166,
496 264-272.
- 497 Hiraki, T., Takeuchi, M., Hisa, M., Akiyama, T., 2005. Hydrogen production from
498 waste 485 aluminum at different temperatures, with LCA. *Mater. Transactions* 46,
499 1052-1057.
- 500 Hong, J., Wang, J., Chen, H., Sun, B., Li, J., Chen, C., 2010. Process of aluminum
501 dross recycling and life cycle assessment for Al-Si alloys and brown fused alumina.
502 *Transactions Nonferrous Met. Soc. China* 20(11), 2155-2161.
- 503 Hong, J., Zhan, S., Yu, Z., Hong, J., Qi, C., 2018. Life-cycle environmental and
504 economic assessment of medical waste treatment. *J. Clean. Prod.* 174, 65-73.
- 505 International Organization for Standardization, 2006. Environmental Management -
506 Life Cycle Assessment - Principles and Framework.
507 <https://www.iso.org/standard/37456.html>
- 508 Jin, Q., Shan, A., 2019. A process for secondary aluminum dross recovery. Shanghai
509 Jiao Tong University, China. (in Chinese)
- 510 Liu, Z., Geng, Y., Adams, M., Dong, L., Sun, L., Zhao, J., Dong, H., Wu, J., Tian, X.,
511 2016. Uncovering driving forces on greenhouse gas emissions in China' aluminum
512 industry from the perspective of life cycle analysis. *Appl. Energy* 166, 253-263.
- 513 Li, L., Song, M., Jin, Q., 2019. Study on alkali dissolution of secondary aluminum

- 514 dross. *Inorg. Salt Industry* 51, 59-63. (in Chinese)
- 515 Mahinroosta, M., Allahverdi, A., 2018a. Enhanced alumina recovery from secondary
516 aluminum dross for high purity nanostructured gamma-alumina powder production:
517 Kinetic study. *J. Environ. Manage.* 212, 278-291.
- 518 Mahinroosta, M., Allahverdi, A., 2018b. Hazardous aluminum dross characterization
519 and recycling strategies: A critical review. *J. Environ. Manage.* 223, 452-468.
- 520 Mahinroosta, M., Allahverdi, A., 2018. A promising green process for synthesis of
521 high purity activated-alumina nanopowder from secondary aluminum dross. *J. Clean.
522 Prod.* 179, 93-102.
- 523 Ministry of Natural Resources of the People's Republic of China, 2019. China mineral
524 resources 2019.
525 [http://www.mnr.gov.cn/sj/sjfw/kc_19263/zgkczybg/201910/P02019102253891774952](http://www.mnr.gov.cn/sj/sjfw/kc_19263/zgkczybg/201910/P020191022538917749527.pdf)
526 [7.pdf](http://www.mnr.gov.cn/sj/sjfw/kc_19263/zgkczybg/201910/P020191022538917749527.pdf).
- 527 Ministry of Transport of the People's Republic of China, 2018. The transport
528 yearbook
529 of China 2018, Beijing.
- 530 Meshram, A., Singh, K.K., 2018. Recovery of valuable products from hazardous
531 aluminum dross: A review. *Resour. Conserv. Recycl.* 130, 95-108.
- 532 Murayama, N., Maekawa, I., Ushiro, H., Miyoshi, T., Shibata, J., Valix, M., 2012.
533 Synthesis of various layered double hydroxides using aluminum dross generated in
534 aluminum recycling process. *Int. J. Miner. Processing* 110-111, 46-52.
- 535 Murayama, N., Okajima, N., Yamaoka, S., Yamamoto, H., Shibata, J., 2006.
536 Hydrothermal synthesis of AlPO₄-5 type zeolitic materials by using aluminum dross
537 as a raw material. *J. European Ceramic Soc.* 26(4-5), 459-462.
- 538 Nakajima, K., Osuga, H., Yokoyama, K., Nagasaka, T., 2007. Material Flow Analysis
539 of Aluminum Dross and Environmental Assessment for Its Recycling Process. *Mater.
540 Transactions* 48(8), 2219-2224.
- 541 National Bureau of Statistics of China, 2018. China aluminum industry status and
542 development trend analysis report. <http://www.stats.gov.cn/>
- 543 Sleeswijk, A.W., van Oers, L.F., Guinee, J.B., Struijs, J., Huijbregts, M.A., 2008.
544 Normalisation in product life cycle assessment: an LCA of the global and European
545 economic systems in the year 2000. *Sci. Total Environ.* 390(1), 227-240.
- 546 Song, M., 2018. Study on the process of recovering alumina from secondary
547 aluminum dross. Qingdao University of Science and Technology, China. (in Chinese)
- 548 Steubing, B., Wernet, G., Reinhard, J., Bauer, C., Moreno-Ruiz, E., 2016. The
549 ecoinvent database version 3 (part II): analyzing LCA results and comparison to
550 version 2. *Int. J. Life Cycle Assess.* 21, 1269-1281.
- 551 United States Geological Survey, 2019. Bauxite and alumina statistics and
552 information 2019.
553 [https://www.usgs.gov/centers/nmic/bauxite-and-alumina-statistics-and-information?qt-](https://www.usgs.gov/centers/nmic/bauxite-and-alumina-statistics-and-information?qt-science_support_page_related_con=0#qt-science_support_page_related_con)
554 [-science_support_page_related_con=0#qt-science_support_page_related_con.](https://www.usgs.gov/centers/nmic/bauxite-and-alumina-statistics-and-information?qt-science_support_page_related_con=0#qt-science_support_page_related_con)
- 555 Wang, J., Qiu, Y., Ma, Y., He, S., Liu, N., Feng, Y., Dong, Z., Liu, L., 2019.
556 Quantifying the geographical distribution effect on decreasing aggregated nitrogen
557 oxides intensity in the Chinese electrical generation system. *J. Clean. Prod.* 222,
558 856-864.
- 559 Yu, C., 2018. China electric power yearbook 2017, Beijing, 764-765.
- 560 Yuan, X., Wang, B., 2012. Life cycle assessment of sulfuric acid from pyrite mixed
561 with ferrous slag. *Green Technol.* 6, 137-139. (in Chinese)
- 562 Zhang, W., Li, H., Chen, B., Li, Q., Hou, X., Zhang, H., 2015. CO₂ emission and
563 mitigation potential estimations of China's primary aluminum industry. *J. Clean. Prod.*

564 103, 863-872.

565 Zhang, Y., Sun, M., Hong, J., Han, X., He, J., Shi, W., Li, X., 2016. Environmental
566 footprint of aluminum production in China. *J. Clean. Prod.* 133, 1242-1251.

567 [dataset] Althaus, H., Chudacoff, M., Hischier, R., Jungbluth, N., Osses, M., Primas,
568 A., Hellweg, 474 S., 2007. Life Cycle Inventories of Chemicals. Final Report
569 Ecoinvent Data, v2.0.

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Declaration of interests

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

The authors declare the following financial interests/personal relationships which may be considered as potential competing interests:

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