



Silicon, the silver bullet for mitigating biotic and abiotic stress, and improving grain quality, in rice?

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**Silicon, the silver bullet for mitigating biotic and abiotic stress, and
improving grain quality, in rice?**

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Highlights

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Title: Silicon, the silver bullet for mitigating biotic and abiotic stress, and
improving grain quality, in rice?

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Environmental and Experimental Botany

- Si nutrition is central mitigating abiotic and biotic stress in rice
- Structural Si maintains yield through shoot strength and defends the plant from insect attack
- Structural Si protects against, UV, water-stress, Na, Fe, Mn and Al toxicity

- and lodging
 - Si competes with As for uptake and Cd is bound to Si both in cell walls and phytoliths
 - Recycling straw phytolith Si is central to sustainable rice farming
- Abstract
-

Adequate silicon fertilization greatly boosts rice yield and mitigates biotic and abiotic stress, and improves grain quality through lowering the content of cadmium and inorganic arsenic. This review on silicon dynamics in rice considers recent advances in our understanding of the role of silicon in rice, and the challenges of maintaining adequate silicon fertility within rice paddy systems. Silicon is increasingly considered as an element required for optimal plant performance, particularly in rice. Plants can survive with very low silicon under laboratory/glasshouse conditions, but this is highly artificial and, thus, silicon can be considered as essential for proper plant function in its environment. Silicon is incorporated into structural components of rice cell walls where it increases cell and tissue rigidity in the plant. Structural silicon provides physical protection to plants against microbial infection and insect attack as well as reducing the quality of the tissue to the predating organisms. The abiotic benefits are due to silicon's effect on overall organ strength. This helps protect against lodging, drought stress, high temperature (through efficient maintenance of transpiration), and photosynthesis by protecting against high UV. Furthermore, silicon also protects the plant from saline stress and against a range of toxic metal stresses (arsenic, cadmium, chromium, copper, nickel and zinc). Added to this, silicon application decreases grain concentrations of various human carcinogens, in particular arsenic, antimony and cadmium. As rice is efficient at stripping bioavailable silicon from the soil, recycling of silicon rich

rice straw biomass or addition of inorganic silicon fertilizer, primarily obtained from iron and steel slag, needs careful management. Silicon in the soil may be lost if the silicon-cycle, traditionally achieved *via* composting of rice straw and returning it to the land, is being broken. As composting of rice straw and incorporation of composted or non-composted straw back to land are resource intensive activities, these activities are declining due to population shifts from the countryside to cities. Processes that accelerate rice straw composting, therefore, need to be identified to aid more efficient use of this resource. In addition, rice genetics may help address declining available silicon in paddy soils: for example by selecting for characteristics during breeding that lead to an increased ability of roots to access recalcitrant silicon sources from soil and/or *via* selection for traits that aid the maintenance of a high silicon status in shoots. Recent advances in understanding the genetic regulation of silicon uptake and transport by rice plants will aid these goals.

1. Introduction:

Rice can be considered as a silicon accumulator and can have shoot silicon concentrations above 10% of shoot dry weight (Yamamoto *et al.*, 2012), with typical ranges from low to high being 1.7 to >3.4% (Korndorfer *et al.*, 2001). Other members of the *Poaceae* share this characteristic (Epstein, 1999; Kido *et al.*, 2015; Ma and Takahashi, 2002; Van Bockhaven *et al.*, 2013), but silicon in rice is the most studied because of the economic importance of rice. Rice has specific mechanisms for assimilating silicon from soil as soluble silicic acid, and for

unloading silicic acid into the xylem, through the aquaglyceroporins *lsi1* and *lsi2* (*i.e.* Ma *et al.*, 2006, 2007a). Grasses, in general, seem to utilize silicon in their tissue for defense against biotic and abiotic stresses such as herbivory, leaf microbial pathogen resistance, lodging, salinity, high light intensity, toxic metal stress and drought tolerance (Goto *et al.*, 2003; Kim *et al.*, 2014; Khattab *et al.*, 2014; Ma and Takahashi, 2002; Van Bockhaven *et al.*, 2013). Plants fertilized with silicon, tend to have higher yields than non-fertilized plants. In fact, silicon fertilization has been shown to increase the number of grains per panicle by *circa.* 25-100% (Ma and Takahashi, 2002; Zhang *et al.*, 2013a). This can be attributed to the fact that silicon fertilization leads to better structural support and enhanced biomass and, consequently, higher yield bearing capacity, as well as increased resistance against various biotic and abiotic stresses that would otherwise cause yield decline.

Evidence, to date, suggests that silicon does not seem to be directly involved in regulation of cell functions besides its direct role in plant structural components and cell wall chemistry. Furthermore, silicon appears to interact with defense associated signaling pathways and silicon status seems to regulate a range of physiological activities (Ye *et al.*, 2013; Van Bockhaven *et al.*, 2013). Silicon appears to be fundamental to regulating grain nutrition by playing an important role with respect to assimilation of the problematic toxins arsenic, antimony and cadmium in rice (Liu *et al.*, 2014b; Li *et al.*, 2009).

This review aims to give an overview of the current state of knowledge of silicon in rice considering: soil-plant biogeochemistry; plants genetics involved in

silicon accumulation, transport and deposition in tissues; the role of silicon in maintaining yield by counteracting biotic and abiotic stress; and how silicon impacts on grain quality. The review, furthermore, provides an overview of topics that require further investigation as it appears that silicon nutrition of paddy soils is declining.

2. Silicon soil-plant biogeochemical cycle

2.1. Silicon soil chemistry

Total silicon is high in soils as it is the second most abundant element in the earth's crust (Sommer *et al.*, 2006). Soils contain >50% SiO₂ (Ma and Takahashi, 2002), predominantly in the form of silicates (aluminum, calcium, iron *etc.*), quartz, biogenic SiO₂ (phytoliths and diatoms) and silica gel (polymerized silicic acid). Soil silicon concentrations range between 1-45% dependent on soil type (Sommer *et al.*, 2006), and soil mineral composition is often the major determinant in silicon availability to the plants (Haynes, 2014). Solid phase SiO₂ is in equilibrium with soluble silicic acid in the soil solution, and it is silicic acid that is the plant available form of silicon (Ma and Takahashi, 2002; Marschner, 2012). Rice has a high demand for silicon, with straw having ~4-20% SiO₂. Consequently, for every 100 kg of brown rice produced, approximately 20kg of SiO₂ is being removed from the soil (Ma and Takahashi 2002).

Silicic acid soil chemistry is complex, regulated by soil pH affecting precipitation and polymerization of plant available silicic acid, with root uptake being greater at lower pH due to higher soil solution concentrations (Tavakkoli *et al.*, 2011).

Soil solution silicon ranges between 3-40 mg/L, averaging ~20 mg/L (Marschner, 2012). Higher concentrations, >56 mg/L silicon, are associated with

supersaturation or partial polymerization (Marschner, 2012). Anaerobism associated with paddy soils drives silicic acid mobilization, probably through dissolution of silicon associated with Fe(III) that is converted to Fe(II) under reduced conditions (Ma and Takahashi, 2002). Silicic acid's pH dependency means that silicon deficiencies are most notable at higher pHs >7 (Korndorfer *et al.*, 2001; Liang *et al.*, 1994). Plant available silicon is also low in soils high in sesquioxides such as highly weathered tropical soils where anion absorption is dominant (Marschner, 2012). Phytolithic biogenic silica dissolves more rapidly to silicic acid compared to other forms of soil silica, and is the dominant source of silicon to rice in paddy environments (Desplanques *et al.*, 2006).

2.2. Paddy soil silicon cycle

Inputs of silicon into paddy soils include silica mineral fertilizers, silica mineral particles deposited during alluvial floods, atmospheric inputs as silica and silicic acid, as well as silicic acid irrigation water inputs (Figure 1). Outputs are primarily leaching to groundwater, surface water run-off and removal of straw. Recycling of silicon is through direct reincorporation of straw, reincorporation after composting or reincorporation as manure after use as animal feed.

Phytoliths account for 90% of rice straw silicon (Ma and Takahashi, 2002). As phytoliths' dissolve relatively rapidly to silicic acid recycling of that straw back into soil is a key part of the paddy system silicon soil fertility. This was demonstrated by seasonal soluble silicon concentrations in paddy soil systems rising in winter after rice straw incorporation due to the dissolution of straw derived phytoliths (Seyfferth *et al.*, 2013). While phytoliths are mainly non-crystalline silicon-dioxide, they can also contain aluminum, iron and occluded

carbon (Li, *et al.* 2014). Evidence suggests that phytoliths with higher aluminum and iron contents break down faster in soils. Research also shows that these high aluminum and iron containing phytoliths are more prevalent in shoots compared to stems, and that the content of phytoliths differs between cultivars (Li *et al.*, 2014).

2.3. Atmospheric and irrigation water inputs

Sommer *et al.* (2006) reviewed the literature on silicon dynamics in terrestrial ecosystems and found that atmospheric inputs from rainwater were small, but solid deposition as dust could be significant in arid climates, while leaching could be an import route for silicon losses from soils. In general, atmospheric and irrigation water inputs supply only 1-10% of silicon above-ground off-take by rice (Desplanques *et al.*, 2006). While irrigation water inputs tend to be low, when silicic acid in the irrigation water is elevated crop uptake increases (Haynes, 2014). Furthermore, Ma and Takahashi (2002) found that rice straw silicon content was strongly correlated with that in irrigation water concentration: SiO₂ content of shoot was ~8% at 10 mg/kg in irrigation water, rising to 18% at concentrations above 30 mg/kg in the water.

2.4. Inorganic silicon fertilization

The main inorganic sources of silicon fertilizers are slags (Haynes, 2014). Metalliferous ores (chromium, iron, manganese and nickel) are heated with limestone and after metal extraction this leaves a calcium silicate waste, slag (Ma and Takahashi, 2002). The main components of slag besides calcium silicate are aluminum, iron and magnesium and traces of chromium, manganese and nickel.

Slag from iron and steel manufacture that contain primarily calcium silicate, can be used as a silicon fertilizer for rice, and as a liming agent, provided that a high proportion of the silicon is bioavailable, and that the slag is low in potentially toxic metals (Ma and Takahashi, 2002; Ning et al., 2014). To be used as a fertilizer in Japan slag must have >20% soluble SiO₂ (reduced to 10% in 1887) and >35% alkali component and <0.4% nickel and 4% chromium (Ma and Takahashi, 2002).

Other sources of silicon include Wollastonite, which is a naturally occurring calcium inosilicate (CaSiO₃) mineral deposit. Wollastonite may also contain small amounts of iron, magnesium, and manganese. Fused magnesium phosphate, formed by heating rock phosphate with serpentine, is rich in calcium, magnesium, phosphorus and silicon. Potassium silicate, a waste product from coal station fly ash, where the ash is mixed with either potassium carbonate, or potassium hydroxide and magnesium hydroxide and calcined at 900°C, is high in aluminum, iron, magnesium and potassium, along with silicon. Porous hydrate calcium silicate (tobamolite), produced from quick lime, quartz and cement, reacted at 180°C under 10 atm pressure, can also be used as a fertilizer.

Industrially produced silica gel is mainly used for nursery applications where there is a need to avoid alkalization. Comparing 12 sources of silicon fertilizer, Pereira *et al.* (2004) identified that phosphate slag was most effective, followed by Wollastonite and electrical furnace silicates.

2.5. Straw recycling of silicon

It was estimated that off-take of silicon by a rice crop was 270 kg/ha/year (Desplanques *et al.*, 2006). If this straw incorporated silicon is not recycled back into the paddy ecosystem, it is a serious loss with major implications for rice fertility. Ma and Takahashi (2002) estimated that, in Japan, historically the application of composted rice straw to the field was equivalent to adding between 300-1000 kg/ha/year SiO₂ and, therefore, provided an important buffer against silicon depletion in paddy soil. When rice straw is composted and added to paddy soil, 70% of the compost added silicon is taken up by rice (Ma and Takahashi, 2002). In a French study conducted by Desplanques *et al.* (2006), paddy fields without replenishment through fertilization were shown to be depleted in silicon at a rate of ~150 kg/ha/year. They found, alarmingly, natural silicon reserves in the soil were estimated to provide for only 5 years. These studies illustrates that continuous input of silicon, either through straw recycling or as mineral fertilizer, is essential to ensure long-term productivity of paddy soils. Maintaining adequate silicon fertility was traditionally achieved *via* addition of rice straw compost (organic silicon fertilization) or, more recently, inorganic forms of silica fertilization in form of industrially produced slag. In Japan, in the 1950s, composting practice was replaced with silicon fertilization through iron and steel slag, with slag eventually displacing compost in importance, but with the actual rate of silicon fertilization from both sources decreasing dramatically between the 1950s and the year 2000, dropping from over 600 kg/ha/year to less than 200 kg/ha/year CaSiO₂.

Silicon fertilization decreases the nutrient content of rice straw on a unit biomass basis, for calcium, iron, magnesium, manganese, nitrogen and

phosphorus, often dramatically, (Ma and Takahashi, 2002). Furthermore, the higher the silicon in the rice straw, the more difficult it will be to degrade and compost. This poses a paradox for silicon recycling as the most beneficial straws, *i.e.* those highest in silicon, are the most difficult to degrade and recycle. The high silicon content of rice straw, compared to other crop straws, makes it more poorly digestible to livestock (Song *et al.*, 2014). However, if manures produced from livestock, that have rice straw as a component of their diet, is incorporated into the compost/soil, the available silicon in soil increases, doubling from 130 to 270 mg/kg after 10 years in one Chinese example (Song *et al.*, 2014). Besides composting of rice straw, rice husks are also a rich source of silicon and if burnt or carbonized husk is, therefore, also a good silicon fertilizer (Ma and Takahashi, 2002). As well, biochars prepared from rice straw, apart from having a range of beneficial properties generally associated with biochars, could also be employed as a slow release soluble silicon fertilizer (Xiao *et al.*, 2014).

3. Silicon accumulation, transport and deposition in rice tissues

3.1. Silicon uptake and transport

Plant roots assimilate silicon through aquaglyceroporins (Meharg and Jardine, 2003). The mechanism behind this assimilation is known to involve two genes: *lsi1* (Ma *et al.*, 2006) and *lsi2* (Ma *et al.*, 2007). *Lsi1* is a silicic influx channel, loading silicon into cells and located on the distal cell wall membranes, while *lsi2* is a silicon effluxer, situated in the proximal cell wall, unloading silicic acid towards and into the xylem. *Lsi1* is primarily located in the basal zones of roots rather than at root tips (Yamaji and Ma, 2007). When the regulation of silicon uptake was studied in contrasting cultivars, a high silicon accumulator and a low

silicon accumulator, these plants were shown to differ in a number of important attributes (Ma *et al.*, 2007b). While the K_m for silicon uptake was similar in both plants, the high silicon accumulator had a higher V_{max} , and this taken along with higher *lsi1* and *lsi2* expression levels, suggested that it had a higher density of silicon transporters in its roots. Another study, again comparing cultivars that contrasted in their shoot silicon levels, also found that the cultivar that accumulated more silicon appeared to do this through a higher V_{max} (Wu *et al.*, 2006).

Silicon is exported from the root to the shoot and unloaded into stems and leaves. A silicic acid transporter *lsi6* was found to be responsible for redirection of silicon at stem nodes (Yamaji and Ma, 2009; Yamaji *et al.*, 2008). Knockout of *lsi6* decreased silicon in panicles and increased silicon in flag leaves, showing its physiological role in silicon distribution in the plant. *Lsi6* is primarily located in xylem transfer cells at the outer boundary of the enlarged vascular bundles (Yamaji and Ma, 2009).

3.2. Silicon deposition in tissues

Silicon is deposited in leaf blades, epidermal and vascular stem tissue, leaf sheath and hull (Ma and Takahashi, 2002), with silicon fertilization leading directly to enhanced thickness in the silicon levels of leaves (Ning *et al.*, 2014).

Silicon in rice is preferentially deposited in epidermal cell walls, where it polymerizes to form silicon-cuticle double layers (Kido *et al.*, 2015). In the leaves of rice, silicon is deposited in “dumbbell” type structures that increase in size as the plant leaf matures, preceded by lignification (Zhang *et al.*, 2013a). Silicon

deposits in the lumen of the cells through needle-like silica structures, molding the inner cell walls (Zhang *et al.*, 2013a).

The similarity between rice's silicon requirement and that of the Equisetums (Horsetails) lead to a hypothesis that a beta-d-glucan (1;3, 1;4) was common between the two and that this glucan may have a key role in silicon-cell wall chemistry (Kido *et al.*, 2015). In rice (1;3, 1;4) beta-d-glucan knockouts, silicon distribution in leaves was altered and reduced the mechanical properties of the leaf blades (Kido *et al.*, 2015). Increased silicon nutritional content of leaves induced lignin production, oxidative cross-linking in cell walls and phytoalexin production (Sun *et al.*, 2010). When rice cells grown in suspension-culture were fractionated, it was found that silicon was primarily associated, through covalent crosslinking, with hemicelluloses and, thereby, improved the mechanical properties of cells (He *et al.*, 2015a). *Lsi1* and *lsi2* knockdown resulted in higher lignin accumulation in cell walls to counteract the low silicon structural content in these mutants (Suzuki *et al.*, 2012). In plants starved of silicon, under hydroponic conditions, the leaf blades visibly drooped compared to those supplemented with silicic acid (1.5 mM) (Yamamoto *et al.*, 2012). In the starved plants, sugars, cellulose and lignin content increased, compensating for the lack of silicon with respect to structural integrity.

The structural deposition of germanic acid, a silicic acid analogue, introduced into phloem through a cut leaf, furthermore, illustrates where silicon is deposited in the developing seed (Carey *et al.*, 2012). Germanium is distributed throughout the husk, and husk hairs, achieves high rachilla concentrations, and is

localized in the ovular vacuolar trace of the edible grain, as is the silicic acid analogue arsenite (Carey *et al.*, 2012). The rice influx transporter has been shown to have a direct role in selenite, another silicic acid analogue (Zhao *et al.*, 2010). Selenite is the dominant form of selenium in soil solution, transported into roots, with *lsi1* knockout mutants having significantly reduced selenite accumulation (Zhao *et al.*, 2010). Antimonite is also a silicic acid analogue and, consequently, also assimilated through silicic acid uptake pathways in rice (Huang *et al.*, 2012, Tripathi *et al.* 2013).

Fleck *et al.* (2011) studied the role of silicon in root formation and function and found that silicon nutrition increased suberization of the exodermis and lignification of the sclerenchyma, as well as reduced the zone of radial oxygen loss by roots. This study also showed that genes involved in suberin and lignin formation were differentially regulated by silicon nutrition.

4. The role of silicon in maintaining yield by counteracting stress

Increased silicon content of leaves enhances their physical resistance as shown by the increased force required by a needle to penetrate the sheath (Schurt *et al.*, 2012) and the enhanced structural rigidity of leaf blades (Yamamoto *et al.*, 2012). Silicon tends to have positive influences in mitigating a range of biotic and abiotic stresses, and in increasing yield. High-yield rice cultivars are dependent on high mechanical strength of their stems (Liang *et al.*, 2013). Comparing *lsi1* knockdown mutants known to have significantly reduced selenite accumulation (Zhao *et al.*, 2010) and wild-type rice, under silicon stressed and silicon replete conditions, it was observed that silicon improved harvest index in parallel with

nitrogen use efficiency and with concurrent alterations of amino acid pools (Detmann *et al.*, 2012).

4.1. Biotic stress mitigation

A link of enhanced silicon content to increased defense against pathogens that are dependent on penetrating or consuming rice leaves was established as early as 1917, when it was found that rice with higher silicon content, had lower infection rates of blast (Ma and Takahashi, 2002). Early Japanese research also linked cultivar silicon content to blast resistance, and silicon fertilization to blast and brown spot mitigation. Susceptibility of rice to a range of diseases was identified to be related to genetic variability in silicon content, and to silicon deficiency in the plants. Disease amelioration was again achieved by silicon fertilization, with silicon fertilization also enhancing yield (Datnoff *et al.*, 1992; Deren *et al.*, 1994). Lower rates of infection due to increased physical strength of leaves in response to increased level of silicon will, most certainly, also indirectly affect molecular defense associated signaling responses, as these plants experience reduced levels of stress. While effects of silicon, leading to increased biotic stress resistance, *via* provision of increased leaf strength, cannot be disputed, it is harder to establish whether silicon also has a direct role in molecular defense associated signaling pathways. This would have to be independent of silicon's role in providing increased physical protection to the leaf tissue. Various studies aimed to further dissect the role of silicon in abiotic stress resistance including the direct impact on molecular defense, independent of silicon's structural role in in the plant (Van Bockhaven *et al.*, 2013).

Sheath blight, *Rhizoctonia solani* was decreased in two rice cultivars grown hydroponically with 2 mM Si compared to no silicon (Schurt *et al.*, 2014). Other studies into the effects of silicon fertilization and infection rates by sheath blight found that enhanced silicon in the shoot protected both a sensitive and insensitive cultivar from this fungus (Zhang *et al.*, 2013 b). It was hypothesized that silicon may exert this defensive action through interaction with phenolics production, but this was only found to be the case in the sensitive cultivar. A study on two rice cultivars grown hydroponically and fertilized with and without silicic acid found that enhanced silicon content of leaves decreased the damage done by leaf sheaf blight, but this did not appear to be through alteration of the sugar or lignin content of the leaves (Schurt *et al.*, 2013).

Fungal brown spot (*Bipolaris oryzae*) resistance of rice was improved through slag based silicon fertilization, and resistance was thought to be due to observed structural changes in leaves, with enhanced thickness of the silicon layer in mesophyll cells, and through enhanced cytoplasmic silicon concentrations (Ning *et al.*, 2014). Increased brown spot resistance in response to silicon fertilization was observed to be less in a *lsi1* knockout mutant compared to its wild-type counterpart cultivar *Oochikara* (Dallagnol *et al.*, 2011, Dallagonol *et al.*, 2013, Dallagnol *et al.*, 2014). Silicon fertilization (2mM in hydroponics) reduced infection by 81% in wild type and 50% in the mutant (Dallagnol *et al.*, 2011). These benefits of silicon, in protecting against *B. oryzae*, as measured by leaf sugar content and photosynthetic rate after infection, were identified to be leaf silicon concentration dose dependent (Dallagonol *et al.*, 2013). While high leaf silicon resulted in increase soluble sugar concentration, and to decreasing brown

spot severity, they did so in independent ways (Dallagonol *et al.*, 2013). In another study the relationship between brown spot sensitivity and leaf silicon was linear and negative across eight rice genotypes in soil grown plants fertilized with silicon (Prabhu *et al.*, 2012). While foliar application of silicon can decrease brown spot, it was not as effective as root-supplied silicon (Rezende *et al.*, 2009). A study on silicon induced rice resistance to brown spot fungus *Cachliobolus miyabeanus* (the sexual stage of *B. oryzae*), identified decreased infection to be due to silicon induced regulation of ethylene pathways, apparently through blocking fungal derived ethylene (Van Bockhaven *et al.*, 2015).

Fungal rice blast (*Magnaporthe oryzae*) resistance was enhanced by silicon fertilization (Abed-Ashtiani *et al.*, 2012; Liu *et al.*, 2014a), with *lsi1* deficient mutants found to be more susceptible to blast than their wild-type counterpart (Nakata *et al.*, 2008). Sun *et al.* (2010) attributes silicon induced resistance to blast to be due to physical changes in cell walls, lignification and oxidative cross-linking. A transcriptional expression study found that silicon affected the interaction between rice and rice-blast fungus at the molecular level (Brunings *et al.*, 2009). In two near isogenic lines, with differential blast resistance, silicon fertilization in nutrient solution was considerably increased in both genotypes (Cai *et al.*, 2008). This was not due to alteration in lignin or defense enzyme function on silicon fertilization *per se*, but because the silicon and lignin content of shoots increased in response to blast inoculation in both isogenic lines. The authors concluded that silicon acted through physical defense, primarily the deposition of silicon in leaf epidermal cells. Hayaska *et al.* (2008) also suggested

that the protection silicon affords to rice blast fungus is protection against its penetration into leaf tissue.

Screening found that rice varieties high in shoot silicon have enhanced resistance to rice leaf folder, *Cnaphalocrocis medinalis* Guenee, while silicon soil fertilization (at rates of 0.16 and 0.32 g Si/kg) also conferred protection to poorly resistant cultivars (Han *et al.*, 2015). The authors of this study concluded that it was due to the reduced quality/digestibility of the leaf, due to higher silicon content of the leaf, rather than the silicon acting as an active deterrent to leaf consumption.

This contrasts with a plant physiological investigation into the effects of rice leaf folder. Rice silicon concentrations and transcript levels, including *lsi1* and *lsi2* were measured to study the impacts of silicon on a wild-type rice and its jasmonate synthesis pathways RNAi knockdown (Ye *et al.*, 2013). The authors concluded that there was a strong interaction of silicon and the jasmonate signaling pathway. Silicon pre-treatment led to higher levels of jasmonate production and increased expression of a range of defense related genes in the wild-type. Reduced levels of silicon were observed in knockdown mutant leaves. Furthermore, reduced levels of *lsi1*, *lsi2* and *lsi6* transcripts were observed in this study in knockdowns after insect attack.

Stem borer resistance (*Chilo suppressalis*) was also enhanced by silicon addition to soil, with maximal effect seen at 20g silicon/kg (Kanew *et al.* 2012). This silicon induced resistance to the stem borer appears to be physical with enhanced silicon content causing decreased borer penetration, resulting in lower weight gain and stem damage (Hou and Han, 2010). This weight gain reduction

increased the length of the larval stage, making the borer more prone to predation.

Enhanced leaf silicon also provides resistance to planthoppers (Ma and Takahashi, 2002). Yoshihara *et al.* (1979) found that elevated silicic acid in the phloem inhibited brown planthopper (*Nilaparvata lugens*) feeding. Another study, that of He *et al.* (2015b), found that increased plant silicic acid content caused a decrease in planthopper residence time on the plant, a decrease in fertility and a decrease in honeydew production.

It appears that silicon defense against insects is threefold: **a.** enhanced physical protection of the leaf for attack/colonization, **b.** the leaf being a poorer quality substrate as silicon makes it less digestible and results in lower macro-nutrient content and **c.** the phenology of the insects' life-cycle is slowed down, also making it more prone to predation. Silicon content regulation of biochemical pathways may also play a role (Van Bockhaven *et al.*, 2013).

4.2. Abiotic stress mitigation

Silicon in outer walls of the stems and the silicon content correlates with bending strength and, thus, lodging resistance (Liang *et al.*, 2013). A study by Goto *et al.* (2003) showed that silicon fertilization leads to enhanced UV tolerance due the protective effect of deposition of silicon bodies in the leaf epidermis. Increased silicon nutrition enhanced photosynthesis, probably through enhanced mesophyll conductance (Detmann *et al.*, 2012).

Drought stress is mitigated by silicon (Khattab *et al.*, 2014; Nolla *et al.*, 2012). In an elegant soil based experiment conducted by Nolla *et al.* (2012), where upland rice was grown in two different savanna soils held at 60, 70 and 80% of water holding capacity and calcium silicate fertilized at 0, 200, 400 and 600 kg/ha/year, silicon fertilization enhanced yields under water stress. There is evidence that abscisic acid, known to mediate drought-induced growth arrest, targets support cells in mature tissue (Shobbar *et al.*, 2008). In this study it was shown that abscisic acid induced accumulation of transcripts in the cytosol of major support cells, including epidermal silicon cells, thereby linking abscisic acid signaling to silicon deposition. Furthermore, expression of *lsi1* was shown to be regulated by abscisic acid and by dehydration stress, suggesting a role in drought in regulating this proteins' expression (Yamaji and Ma, 2007).

Silicon fertilization enhances sodium tolerance in rice (Kim *et al.*, 2014). Silicon fertilization reduced transport of chloride to shoots in 3 rice cultivars (Shi *et al.*, 2013). Decreased chloride transport was correlated with a decrease in transpirational bypass flow (Shi *et al.*, 2013; Yeo *et al.*, 1999). Silicon fertilization increased stomatal conductance and net photosynthetic rate. The fact that silicon fertilization increased the net photosynthetic rate, stomata conductance, and transpiration of salt-stressed plants, showed that decrease in sodium and chloride transport to shoots in salt sensitive cultivars was not due to a decrease in the transpiration rate, and that the transpirational bypass flow change was the major factor in silicon decreasing chloride transport to shoots (Shi *et al.*, 2013). A mechanistic study of how silicon ameliorates sodium stress in rice provided evidence that enhanced silicon content in both root and shoot endodermis

reduced silicon uptake by the root. This was achieved by reducing apoplastic transport and, again, was not due to alteration of transpiration. (Gong *et al.*, 2006). It was specifically sodium uptake that was altered by silicon with no effects on potassium uptake and transport.

There is some evidence that silicon fertilization, at least in hydroponic solution, ameliorates iron(II) toxicity. Enhanced silicon in the growth solution decreased iron plaque formation on the rhizoplane and increased iron transport to shoot, though the mechanistic basis of these observations was not investigated (Fu *et al.*, 2012). Manganese toxicity is also ameliorated by silicon in hydroponics with enhanced plant silicon content restricting manganese transport to the shoot in one cultivar and decreasing root assimilation in another (Li *et al.* 2012). This decreased manganese uptake and transport led to decreased redox stress in the plant tissues. Similar findings have also been reported for zinc where the enhanced cell wall binding of zinc in silicon replete plants has been implicated in restricting zinc uptake into roots and transport to shoots (Gu *et al.*, 2012; Song *et al.*, 2011). Silicon fertilization in hydroponic studies was furthermore found to decrease aluminum (Hara *et al.*, 1999; Singh *et al.* 2011), and chromium (Zeng *et al.*, 2011) uptake by plants. Fertilizer, silicon applied as fly-ash or steel slag to multi-metal contaminated soil affected both root uptake and shoot translocation of cadmium, copper, lead and zinc (Gu *et al.*, 2011). The action of silicon appeared 2-fold. Toxic metals were precipitated in soil in the presences of silicon fertilizer to corresponding silicates, phosphates and hydroxides, while enhanced shoot silicon restricted transport of these metals into leaves from stems.

It has to be noted that the effect of silicon on abiotic resistance is not always positive. When soils fertilized with 600 kg/ha/year SiO₂, were subsequently subjected to 12 d submergence, silicon fertilization decreased rice survival (Ella *et al.*, 2011). Four genotypes were studied, including flooding sensitive and insensitive, but there was no silicon times genotype interaction. The mechanistic basis of this is unclear though survival was correlated negatively with chlorophyll degradation, carbohydrate depletion and shoot elongation during submergence.

5. Grain quality and silicon

Silicon fertilization appears to mitigate against accumulation of inorganic arsenic, antimonite and cadmium in rice grain. Arsenic, antimonite and cadmium are toxic at elevated concentrations and if accumulated in rice grain, present potential threat to human health. The interconnected pathways of silicon with arsenic and cadmium (and iron), from soil particles to shoot epidermal cells, are illustrated in Figure 2.

5.1 Silicon on inorganic arsenic, antimony and selenite uptake

Arsenite, antimonite and selenite are all silicic acid analogues with respect to their assimilation by rice roots. Antimonite and arsenite are potent carcinogens while selenium (as a range of species) is an essential element.

Inorganic arsenic in rice grain is a global problem as rice is the dominant source of this carcinogen to the human diet (Meharg *et al.*, 2009). Arsenite, the prevalent form of arsenic under anaerobic conditions in paddy soils, is a silicic

acid analogue and rice is efficient at assimilating arsenite. A range of studies has found that excess silicon, either in soil or hydroponic culture, reduces inorganic arsenic uptake and translocation to plant shoots and grain (*i.e.* Fleck *et al.*, 2013; Liu *et al.*, 2014b; Li *et al.*, 2009; Seyfferth *et al.*, 2012). When the availability of arsenic in soil solution was correlated to natural soils solution silicic acid across 6 soil types, it was observed that higher silicic acid in soil solution inhibited arsenic assimilation by rice (Bogdan and Schenk, 2008). Addition of sparingly soluble silicate gel to paddy soils decreased the accumulation of inorganic arsenic, but enhanced the uptake of dimethylarsinic acid (DMA) (Liu *et al.*, 2014b; Li *et al.*, 2009). It was postulated that this enhanced DMA assimilation was due to dissolved silicic acid derived from the silicate gel outcompeting DMA for soil exchange sites, enhancing its concentration in soil solution (Liu *et al.*, 2014b). Similar to Liu *et al.* (2014b), Fleck *et al.* (2013) found that wholegrain DMA was decreased by silicon addition, but that polished rice concentrations were identical in plants that had or had not received silicon supplementation to their soils. This silicon fertilization of paddy soils reduced arsenic concentrations in wholegrain rice by 22%. They found that silicon fertilization enhanced silicon, iron, phosphorus and arsenic in soil solutions, but postulate that the enhanced silicon in the soil solution, or in plant tissue out-competes arsenite transport, ultimately leading to lower grain concentrations. In a study comparing diatomaceous earth with silica-gel fertilized paddy soils, with respect to these amendments impact on silicon and arsenic uptake into the plant, it was found that diatomaceous earth was not efficacious at either releasing silicon into soil solution or in affecting plant arsenic levels, where silica-gel did both (Seyfferth *et*

al., 2012). It was postulated that silicon, as silicic acid, released from the gel outcompeted arsenite for uptake.

There is strong evidence that inorganic arsenic is remobilized from leaves during grain fill through phloem transport (Carey *et al.*, 2011). When the silicon content of rice leaves was correlated with grain arsenic in the field, a strong positive relationship was found, with the arsenic present predominantly in its inorganic form (Norton *et al.*, 2010). So while silicic acid may compete with arsenite for root uptake, plants that are genetically more efficient at assimilating silicon will have higher loadings of inorganic arsenic.

Rice is also a source of the environmental pollutant antimonite, a carcinogen, and toxicant causing diseases of the liver, skin, respiratory, cardiovascular systems (Wu *et al.*, 2011, Ren *et al.*, 2014). Antimonite is, like arsenite, also a silicic acid analogue and presence of silicic acid in solution culture was also shown to reduce uptake of antimonite in rice (Huang *et al.*, 2012, Tripathi *et al.* 2013).

There is also physiological evidence for a role in silicic acid transporters in selenite uptake, again with selenite acting as a silicic acid analogue (Zhao *et al.*, 2010), though the impact of this pathway on grain nutrition is not known. Given that selenium is an essential nutrient, often deficient in rice, and in humans, this elucidates that selenium species uptake pathways in rice are important.

5.2. Effect of silicon on cadmium uptake

Rice is also the dominant source of the carcinogen and renal toxicant cadmium into the human diet (Meharg *et al.*, 2013). Silicon fertilization is known to reduce cadmium uptake and transport by rice (Triaparthi *et al.*, 2012; Wang *et al.*, 2000), and to enhance cadmium tolerance (Nwugo and Huerta, 2008; Wang *et al.*, 2000), though the mechanisms is not well understood. Silicon was found to cross-link with cell wall hemi-celluloses (Ma *et al.*, 2015). It was postulated that the negative charge of these silicon complexes in the cell wall may lead to enhanced cadmium binding and, thereby, inhibit cadmium transport within the rice plant (Liu *et al.*, 2013; Ma *et al.*, 2015). In a time-course study by Zhang *et al.* (2008) silicon fertilization decreased cadmium assimilation by the plant and decreased cadmium toxicity in solution culture. Stimulation of plant biomass by silicon was a potential route of mitigation of cadmium, as was the deposition of cadmium in phytoliths, as observed by energy-dispersive X-ray micro-analysis. The experiments of Wang *et al.* (2000) also suggested that cadmium was precipitated by colloidal silicon in cells.

6. Future directions

Silicon appears to be a silver bullet for rice in counteracting both biotic and abiotic stress, leading to higher yields, as well as improving the quality of grain through decreasing cadmium, antimony and arsenic contents. Silicon may, furthermore, enhance grain selenium concentrations. The benefits of silicon nutrition to rice are outlined in Figure 3. However, as rice is efficient at stripping plant available silicon from soils the challenge is how to effectively cycle and sustain silicon levels in soil. The answer is two-fold: rice genetics and silicon fertilization.

While the specific mechanisms of silicon accumulation by rice roots are now well understood, *i.e.* *lsi1* and *lsi2*, the basis of genetic variance in shoot silicon content between cultivars is not. Quantitative trait loci (QTL) studies have identified variance in shoot and hull silicon content (Bryant *et al.* 2011, Dai *et al.*, 2005, 2008; Gong *et al.*, 2010). Twelve QTL markers were found for hull silica content with 6 of these also coinciding with markers for silicon content found in other studies (Bryant *et al.*, 2011). If the genes responsible for these QTLs can be identified, this will be the start of a mechanistic understanding of the regulation of the plants silicon status, plus aid in the accelerated breeding for desirable traits. Wider screening, and subsequent genetic analysis, of rice germplasm for high shoot silicon and enhanced efficiency of root assimilation from recalcitrant soil pools is another key strategy for breeding rice that is higher in silicon status.

Inorganic silicon fertilizers are finite and are related to industrial processing of iron and steel and, thus, availability of inorganic silicon fertilizers are related to local economy and, therefore, patchily distributed. Also, as issues regarding toxic elements in rice grain come to the fore, particularly cadmium (Meharg *et al.*, 2013) and arsenic (Meharg *et al.*, 2009), the trace metal contamination of slag, which has always been an issue regarding nickel and chromium (Ma and Takahashi 2002), needs further examination (Ganne *et al.* 2006; Houben *et al.*, 2013). However, slag has been suggested as remediation agent for both arsenic (Li *et al.*, 2006) and cadmium (Zhou *et al.*, 2012), so the balance of the risks and benefits of slag with respect to arsenic and cadmium rice grain concentrations needs exploring.

Perhaps a more sustainable strategy with respect to silicon fertilization is to increase the once common practice (Ma and Takahashi, 2002) of reincorporating composted rice straw back into paddy fields. This results in returning the vast bulk of silicon removed by the crop in a highly bioavailable form, by ensuring continuous cycling of phytoliths. Composting of rice straw, due to its high silicon content, is slow, needs valuable space in which to perform the compositing, and is time consuming, which explains the sharp decline in rice composting observed in countries such as Japan (Ma and Takahashi, 2002). The accelerated composting of rice straw needs some attention, such as *via* the use of fungal inoculants (Viji and Neelenarayanan, 2015), as well as the co-composting of straw alongside animal manures (Zhu, 2007). In particular use of manures produced from animals that consume rice straw (Song et al. 2014) as well as use of biochars derived from rice straw (Xiao et al. 2014) or husks (Ma and Takahashi, 2002) in composting or as soil amendment may provide a way forward.

Ultimately, if silicon can be more efficiently recycled, and plants bred for high silicon content, considerably improved yields of high quality grain will be achieved, along with decreased requirements for pesticide usage. High silicon rice is better adapted to the biotic and abiotic stresses that climate change will bring. Silicon can indeed be seen as the silver bullet for mitigating biotic and abiotic stress, and improving grain quality, in rice. Optimization of silicon's biogeochemical cycling in paddy environments is, therefore, on all accounts, a good long-term agricultural policy with great potential benefit to the global economy: It will help feed an expanding population, as will deliver increased

yield and higher quality rice crops (increase in essential nutrients *i.e.* selenium, decrease the carcinogens arsenic and cadmium), while at the same time leading to increased sustainability of the paddy soil system and decrease in agriculturally mediated pollution, *i.e.* tighter nutrient cycling, reduced requirements for pesticides.

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Legends to Figures

Figure 1. The paddy soil system silicon cycle.

Figure 3. Schematic model of silicon interactions with arsenic and cadmium (and iron) from the soil solid phase through to leaf epidermal cells.

Figure 3. The benefits of silicon in rice culture in addition to increased yield through enhanced structural integrity and biomass.

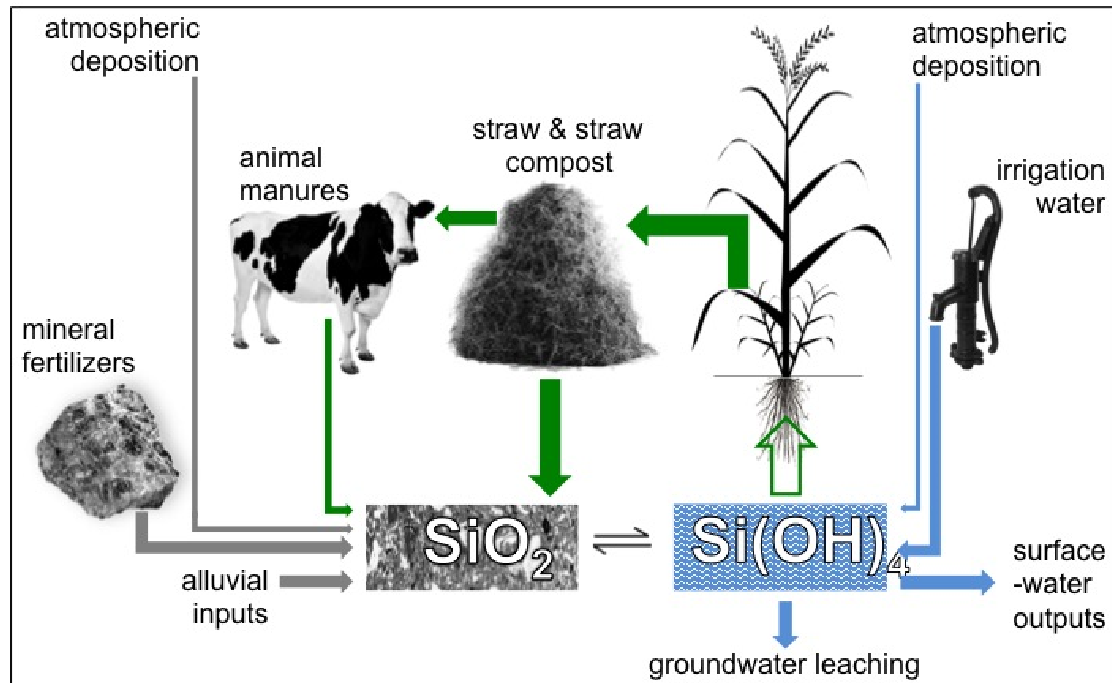


Fig 1 .

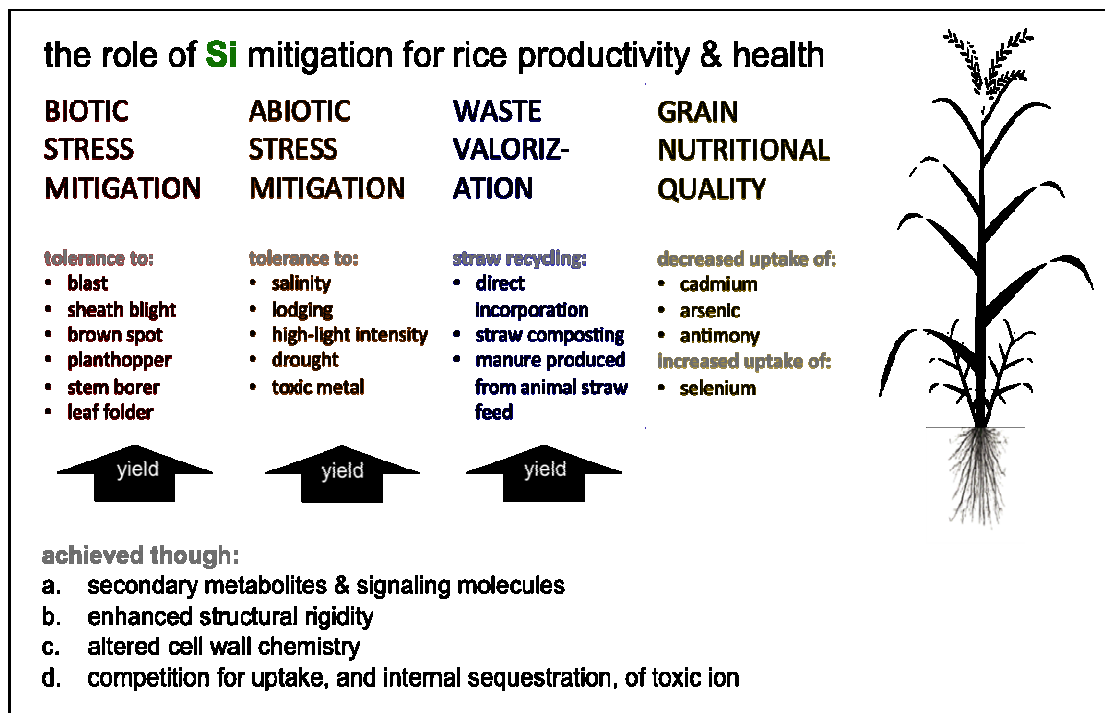


Fig 3 .