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A DATA-DRIVEN REGIONAL MODEL OF STOMATAL CONDUCTANCE FOR

KRUGER NATIONAL PARK

by

Rebecca L. Tobin

A thesis submitted in partial fulfillment
of the requirements for the degree

of

MASTER OF SCIENCE

in

Ecology

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2020

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ABSTRACT

A Data-driven Regional Model of Stomatal Conductance for Kruger National Park

by

Rebecca Tobin, Master of Science

Utah State University, 2020

Major Professor: Dr. Andrew Kulmatiski
Program: Ecology

The basic drivers of stomatal conductance (g_s) are well understood at the leaf level under controlled conditions, but it has been difficult to extrapolate laboratory principals to plant communities. Here we estimate and model landscape-level g_s from a dataset with over 8,000 g_s measurements made over five years from four study sites in Kruger National Park, South Africa. Sites represented a wide range of precipitation (450-750 mm mean annual precipitation) and soil types (sand and clay). Measurements were used in a machine-learning (Random Forest) model to assess the effects of plant functional type (grass or woody), species, vapor pressure deficit, soil moisture, shortwave radiation, wind speed, atmospheric $[CO_2]$, time-of-season, soil type, and precipitation on g_s . Both plant functional type and species had large effects on g_s . Among environmental variables, shallow soil moisture had the greatest effect on g_s for both grasses and woody plants. Soil type had the smallest effect on g_s for both plant functional types. The effect of environment differed between grasses and woody plants. When the models were used with observed environmental data from several growing seasons, mean daytime g_s was

estimated as 67 and 158 $\text{mmol m}^{-2} \text{sec}^{-1}$ for grasses and woody plants, respectively. While laboratory-based models emphasize the role of leaf-level environmental parameters, this dataset highlights the role of species identity and soil moisture as major drivers of g_s at the landscape scale. Results also show a large amount of landscape-scale variability in g_s that remains to be explained.

(47 pages)

PUBLIC ABSTRACT

A Data-driven Regional Model of Stomatal Conductance for Kruger National Park

Rebecca Tobin

Stomata are the gateway between the lithosphere, the biosphere, and the atmosphere. Because of photosynthesis, plants inevitably lose water through their stomata. The rate at which water moves through stomata is stomatal conductance. As stomatal conductance increases, the rate of CO₂ assimilation increases, therefore, plants must reach a balance between acquiring CO₂ and losing H₂O. Plants achieve this balance by adjusting stomatal aperture. Therefore, modeling stomatal conductance is important to global circulation models and land surface models, as well as for predicting how changing climate conditions affect water use efficiency and plant productivity, and has implications for agriculture and natural resource management.

Here a large dataset of field measurements was used to describe stomatal conductance for Kruger National Park, South Africa and develop statistical models of landscape-level stomatal conductance. Then models were used to estimate stomatal conductance across the region over several growing seasons. Over 8,000 measurements of stomatal conductance were made in four sites that represented a range of precipitation regimes and soil types within Kruger National Park from 2007-2012. Known environmental drivers of stomatal conductance, such as soil moisture, temperature, and shortwave radiation, were also measured during this period.

Observed mean daytime stomatal conductance for the park was 75 ± 1 and 155 ± 2 mmol m⁻² sec⁻¹ for grasses and woody plants, respectively. When statistical models were

used to produce three years of continuous estimates of g_s from environmental data, average daytime stomatal conductance was estimated as 67 and 158 $\text{mmol m}^{-2} \text{sec}^{-1}$ for grasses and woody plants, respectively. The Random Forest statistical models that were used to produce continuous estimates of g_s indicated that soil moisture, particularly at shallow depths, and plant species identity are primary drivers of landscape-scale stomatal conductance for Kruger National Park. However, results indicate that there is still a large amount of landscape-scale variability in stomatal conductance that the environmental drivers investigated here were unable to explain.

Results provide a rare example of landscape-level estimates of stomatal conductance based on direct measurements. The models give insight into the relative importance of environmental drivers and the nature of their effect on stomatal conductance in savanna ecosystems. Because the measurements were collected over a range of species and soil conditions, the models should provide inference for many deciduous, sub-tropical savannas of southern Africa.

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This study would not have been possible without the cooperation of South African National Parks and Kruger National Park (project registration number 213896412). I would like to thank the Andrew Mellon Foundation, the Utah State University Agricultural Experiment Station, and the Office of Research and Graduate Studies for funding this project. Thank you to my major professor: Andrew Kulmatiski, my thesis committee: Karen H. Beard and Bruce Bugbee, and S. Durham and Thomas C. Edwards for assistance in statistical analyses and modeling. Thank you also to the field managers: M. Keretsetse, S. Heath, L. Hierl, M. Cooper, and M. Mazzacavallo and the field/laboratory assistants: W. Sibuyi, R. Mashele, V. Sibuyi, and M. Rogers.

Rebecca Tobin

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INTRODUCTION

Because stomatal conductance (g_s) is a measure of gas exchange between plants and the atmosphere, g_s is an important component of CO_2 and water cycles at both local and global scales¹⁻⁴. Therefore, understanding the factors that determine g_s is important for predicting small-scale processes such as plant productivity¹, species coexistence and crop water use⁴ as well as large-scale processes, such as global CO_2 and energy budgets^{2,5}.

Models of g_s are numerous, well-developed^{1,3,6} and fall into three general categories: empirical, mechanistic, and optimization³. Empirical, or data-based, models describe the response of g_s to environmental parameters, such as irradiance¹, temperature, vapor pressure deficit (VPD)^{1,7}, CO_2 concentration⁶⁻⁸, water stress¹, and interactions among these drivers. Because many factors can affect g_s , the majority of empirical approaches have been conducted in laboratory settings where the effect of individual factors can be tested. The empirical g_s models developed by Jarvis⁹ (including subsequent Jarvis-type models⁶) and Ball, Berry, and Woodrow¹⁰ provide reasonable estimates of g_s under laboratory conditions and some of the best estimates of g_s under field conditions^{3,11}. As a result, these models are widely-used in global circulation models, earth system models, and models of canopy-level processes^{2,3}. However, due to the difficulty of measuring g_s in the field¹, validation of model predictions remains limited^{3,9,10}. Mechanistic approaches rely on models and tests of the role of specific mechanisms, but are often difficult to apply to the landscape-level^{3,11}. Finally, optimization models seek to predict g_s behavior according to the premise that g_s is

regulated to maximize photosynthesis and minimize water loss^{2,3,5}. Although there can be computational difficulties in implementing optimization models³, there have been recent efforts to incorporate g_s optimization models into earth system models^{2,5,12}.

Developing g_s models applicable on a landscape or global scale has proven difficult^{1,2,13}. Due in part to technological limitations¹, g_s datasets are rarely large enough to capture the variability in g_s that occurs among species, within canopies, and over daily and seasonal time-scales in response to environmental drivers, such as soil moisture and VPD. As a result, response curves generated from limited observations may not be applicable across landscapes¹⁴. Savanna ecosystems pose a particular challenge because they include alternative dominant life forms: grasses and trees that can vary widely in both g_s and their g_s responses to environmental drivers. There remains, therefore, a need for both datasets and models of landscape-scale drivers of g_s across growing seasons for this region.

The overarching goal of this study was to describe g_s in the savanna ecosystems of the nearly 2 million ha Kruger National Park and surrounding ecosystems in South Africa. More specifically, the objectives were: 1) to develop a dataset large enough to describe g_s for Kruger National Park, 2) to use the dataset to build a landscape-scale model of g_s and 3) to use the model and observed environmental data to produce continuous estimates of g_s across Kruger Park for three growing seasons. To capture landscape-scale variability in g_s , measurements were collected over five years in four sites that represent a wide range of abiotic and biotic conditions. A machine-learning approach (Random Forest, hereafter RF) was used to describe the effect of the following environmental parameters on g_s : soil moisture, VPD, shortwave radiation, wind speed,

soil type, precipitation regime, time-of-season, time-of-day, atmospheric [CO₂], and species identity, on g_s . RF modeling has been shown to reveal nonlinear relationships and complex interactions in ecological data that may be missed by other statistical methods¹⁵. Because different species and functional groups are influenced by and respond differently to environmental conditions¹⁶, separate RF models were developed and conducted for grasses and woody plants. The RF models that explained the greatest variance in the g_s dataset were used with environmental data to produce continuous, three-year estimates of g_s for each study site and the entire study area.

METHODS

2.1 Study Site Information

Research was conducted between 2007 and 2012 in four deciduous, subtropical savanna sites in Kruger National Park, South Africa: Letaba (-23°46'49.00" S, 31°31'16.19" E), Phalaborwa (-23°51'25.27" S, 31°14'12.75"E), Pretoriuskop (-25°12'21.68" S, 31°17'9.92" E), and Lower Sabie (-25°12'2.09" S, 31°54'27.25" E). The four sites were selected to provide broad inference to conditions on the landscape, and represented a two-by-two factorial combination of precipitation (“wet” or “dry”) and soil texture (“sand” or “clay”)(Table 1)^{17,18}. Common grasses include *Bothriochloa radicans* (Lehm) A. Camus, *Setaria incrassate* (Hochst.) Hack. and *Urochloa mosambicensis* (Hack.) Dandy. Common woody plants include *Terminalia sericea* Burch. ex DC and the nitrogen-fixing *Colophospermum mopane* (Benth.) Leonard and *Dichrostachys cinerea* subsp. *africana* (Brenan & Brummitt)(Table 2).

Table 1. Precipitation regimes and soil types corresponding to the four study sites within Kruger National Park, South Africa^{17,18}.

Site Name	Soil Type	Precipitation Regime
Letaba	Clay (calcareous shallow clay)	Dry (450 MAP)
Phalaborwa	Sand (coarse fersiallitic sand)	Dry (475 MAP)
Lower Sabie	Clay (pedocutanic clay)	Wet (730 MAP)
Pretoriuskop	Sand (coarse fersiallitic sand)	Wet (750 MAP)

2.2 Study Design

Stomatal conductance measurements: At each site, g_s measurements were made across a roughly 4 ha sampling area that had been established for related research^{19–21}. g_s was measured using steady-state porometers (Decagon Devices, SC-1)²², which take g_s measurements in 30 seconds, allowing large sample sizes relative to null-balance porometers or dynamic porometers²³. Measurements were made during six sampling campaigns that represented early, mid- and late-season sampling during each of two growing seasons at each site. Each sampling campaign included 2-3 days of sampling. To prevent biased sampling of certain samples (*i.e.*, plant species), measurements were made during either consistent cloud cover or clear skies. Sampling was intended to be as representative of landscape-level g_s as possible, so samples were collected between sunrise and sunset, and were taken throughout the plant canopy^{20,21,24}. For grasses, g_s was measured from both abaxial and adaxial surfaces. For woody plants, g_s was not detectable on adaxial surfaces and was not measured. Forbs were also sampled, although their relative abundance was small compared to grasses and woody plants. Each of roughly 10 dominant target species at a site was measured within 15-minute increments to control for environmental variability. The species, plant functional type, soil type (clay or sand) and precipitation regime (wet or dry) and time-of-season [early (November – December), middle (January – February) or late (March - April)] in which measurements were taken were recorded. Tree and shrub species were classified together as “woody” (Table 2).

Table 2. Species, plant functional type and sample sizes from each study site. LT = Letaba, LS = Lower Sabie, PB = Phalaborwa, PK = Pretoriuskop. Numbers in parentheses indicate sample size.

Species	Code	Plant type	Sample Size by Site
<i>Acacia nigrescens</i>	ACAN	Woody	PB (33) LS (208)
<i>Lonchocarpus capassa</i>	APPL	Woody	PB (15) PK (3) LS (107)
<i>Bothriocloa radicans</i>	BRAR	Grass	LT (456) PB (12) LS (60)
<i>Combretum apiculatum</i>	COMA	Woody	LT (12) PB (102)
<i>Combretum imberbe</i>	COMI	Woody	PB (36) LS (68)
<i>Loudetia simplex</i>	CORU	Grass	PK (134)
<i>Dichrostachys cinerea</i>	DICH	Woody	PB (106) PK (749) LS (402)
<i>Grewia bicolor</i>	GREW	Woody	PB (12) LS (109)
<i>Sclerocarya birrea</i>	MARU	Woody	LT (2) PB (24) PK (148) LS (1)
<i>Colophospermum mopane</i>	MOPA	Woody	LT (843) PB (178)
<i>Panicum spp.</i>	PANI	Grass	LT (11) PB (71) LS (234)
<i>Terminalia sericea</i>	SCLE	Woody	PB (1) PK (797)
<i>Urochloa mosambicensis</i>	UROC	Grass	LT (25) PB (76) LS (245)
<i>Setaria incrassata</i>	VLEI	Grass	PK (458)
<i>Securinega virosa</i>	WHBE	Woody	PB (128) LS (298)
<i>Ximenia caffra</i>	XIME	Woody	PK (106)

Environmental parameters: Temperature, relative humidity (215L; Campbell Scientific, UT, USA), wind speed (014A cup anemometer; MetOne, OR, USA), total shortwave radiation (SP-110; Apogee Instruments, UT, USA), and precipitation (Texas Instruments TE-525; Texas Instruments, TX, USA) were recorded at each site on Campbell Scientific CR1000 dataloggers. Measurements were made at both “grass” (1 m) and “woody” (2 m) canopy heights, except at Pretoriuskop, where only 2 m heights were

measured. Air temperature and relative humidity were used to calculate VPD using the following equations²⁵:

$$e_s = 0.611 \frac{17.27(T)}{T+237.3}$$

$$e_a = \frac{RH}{100} \times e_s$$

$$VPD = e_a - e_s$$

where e_s is the saturation vapor pressure (kPa), e_a is the actual vapor pressure (kPa), T is the air temperature (°C), and RH is the relative humidity (%). Atmospheric [CO₂] measurements were provided by a flux tower near Skukuza²⁶. Heat dissipation sensors (229; Campbell Scientific, UT, USA) were used to produce a soil water potential “index” for 0-20 cm, 0-50 cm, 20-50 cm, and 50-150 cm depths for each site²¹. Each heat dissipation sensor was calibrated prior to installation by taking measurements from soil samples equilibrated to specific water potentials²¹. To preclude error associated with developing site-specific water potential curves, sensor-specific values of proportional temperature response were used as a soil moisture index^{18,21,27}.

2.3 Data Analyses and Statistics

Simple means and errors of observed daytime g_s by plant functional type for the entire dataset and by site are reported. Species with less than 100 measurements in the dataset were excluded (Table 2). One-way analysis of variance was used to test for differences in mean g_s values among sites for each plant functional type²⁸. To meet assumptions of normality, g_s values were log-transformed. Because sample sizes differed

among sites, Type III sum of squares were used. Pairwise comparisons were examined using the Tukey test²⁹.

Random Forest modeling: Random Forest modeling was used to describe the relationship between environmental parameters and g_s and to build a predictive model of landscape-scale g_s ¹⁵. In RF, a “forest” of regression trees is fit to a training dataset (approximately two-thirds of the sample data). The trees are then used to predict the out-of-bag data (*i.e.* the sample data not included in the training dataset) and the predictions from all trees are combined, giving a cross-validated measure of the accuracy of the model^{15,30}. The relative importance of predictors within the RF models was compared and g_s -predictor relationships were visualized. The RF models that explained the most variance were used to estimate g_s using environmental data from 3-4 growing seasons. Model estimates were generated by plant functional type. Statistical analyses were performed in RStudio³¹. All RF models and predictions were developed using the R package “randomForestSRC”³² and all model visualizations were created using the “ggRandomForests” package³³.

For RF modeling, g_s measurements were paired with meteorological and soil measurements from the closest recorded timestep. Missing meteorological and soil data were interpolated where possible by correlating and adjusting data from the nearest weather station using a simple linear equation ($y = mx + b$, where y is the adjusted measurement and x is the original measurement). To test for potential lag effects in the response of g_s to environmental conditions, the three-hour (3-hour) averages of air temperature, relative humidity, VPD, wind speed, and shortwave radiation, the 3-hour,

24-hour, and seven-day (7-day) averages of each soil moisture depth, and the 24-hour sum of precipitation were calculated.

Although highly-correlated predictors do not affect RF variable importance¹⁵, to simplify interpretation of variable relationships, correlation matrices for groups of related predictor variables were used to test multicollinearity¹⁵. Air temperature, relative humidity, VPD, shortwave radiation, wind speed, precipitation, and soil moisture depths were evaluated. Because data for numerical predictor variables were not all distributed normally, Spearman correlation was used³⁴. Where two predictors were highly correlated (correlation $> \pm 0.7$), the predictor with the greatest “adjusted squared deviance explained” by a generalized linear model (GLMs; linear + quadratic, family = Gaussian) was used in the RF model¹⁵. When highly-correlated predictors explained similar (difference of less than 2 %) amounts of variance in g_s , separate RF models were created to test the amount of variance explained with different combinations of predictors. The “best” RF models were selected based on the amount of variance in the dataset they explained. Categorical predictors were plant functional type, species, time-of-season, precipitation regime, and soil type.

Variable importance (VIMP) within the RF model was determined and visualized using the `ggRandomForests` package³³. Each variable was randomly permuted and the prediction error calculated using the out-of-bag data¹⁵. The VIMP value for each variable is the difference between the out-of-bag prediction error of the observed and permuted variables. Large VIMP values indicate that specifying the variables incorrectly increases prediction error; therefore, variables with large VIMP values are more important.

Negative VIMP values indicate that the randomly permuted variable was a better predictor than the observed variable³³.

Relationships between g_s and its environmental parameters were characterized with risk-adjusted partial dependence plots created using the “ggRandomForests” package³³. Partial dependence refers to the dependence of the response variable, in this case g_s , on one predictor variable¹⁵. The plots were created by averaging the effects of the other predictors and predicting how the response variable changes with the predictor of interest alone¹⁵. Partial dependence of categorical variables was analyzed by comparing the mean predicted g_s of each level of the variable. To avoid confusion with model predictions made with new data, “estimated” was used to describe partial dependence predictions.

g_s sampling was designed to produce a representative sample of g_s on the landscape. However, because sampling was difficult to perform at sunrise and sunset for safety reasons (dangerous animals occupy the areas) and during fluctuating cloud conditions and during rain, the models were used to produce continuous estimates of g_s across three growing seasons. This approach produced estimates that were not biased by a low number of samples at sunrise and sunset. Model predictions of g_s were generated using data from three growing seasons from each study site (2009 – 2012). The data were collected and prepared using the same instrumentation and methods as the data used to build the RF models. The data were then run through the RF models using the “predict” function in the “randomForestSRC” package³². Model predictions were generated separately for each study species. The modeled g_s values were averaged by plant functional type for each study site and for the entire park.

RESULTS

3.1 Observed g_s

Over the five years of the study, 8510 g_s measurements were made. Mean observed daytime g_s was $74 \pm 1 \text{ mmol m}^{-2} \text{ s}^{-1}$ for grasses, $155 \pm 2 \text{ mmol m}^{-2} \text{ s}^{-1}$ for woody plants, and $142 \pm 5 \text{ mmol m}^{-2} \text{ s}^{-1}$ for forbs. The total cover and sample size for forbs was small (580) relative to grasses and woody plants (2662 and 4962, respectively) and was not included in further analyses. Mean observed g_s was greater in the wet/clay site than the other sites for both grasses ($F = 28.399$, $p < 0.001$) and woody plants ($F = 77.298$, $p < 0.001$).

3.2 Random Forest

Across both plant functional types, the best RF model explained 58 % of variance and included, in descending order of importance: species, 24-hour shallow soil moisture, 24-hour deep soil moisture, 3-hour shortwave radiation, 3-hour VPD, 3-hour wind speed, atmospheric $[\text{CO}_2]$, time-of-season, time-of-day, precipitation regime, and soil type. When species was replaced with plant functional type as a predictor, the percent variance explained by the model decreased to 51 %. When neither species nor plant functional type was included in the model, percent variance explained decreased to 43 %. However, because it is reasonable to expect that grasses and woody plants may respond differently to environmental drivers¹⁶, and because savannas show wide variations in woody plant cover¹⁷, separate models were created for each plant functional type group. Percent variance explained for the grass dataset with and without species was 21 % and 20 %, respectively.

respectively. Percent variance explained for the woody g_s dataset with and without species was 54 % and 45 %, respectively.

For both plant functional types, shallow (0-20 cm) soil moisture and soil type were the most and least important predictors of g_s , respectively (Fig. 1). The remaining variables differed in importance between grasses and woody plants. For grasses, in descending order of importance: VPD, atmospheric $[CO_2]$, time-of-day, deep (50-150cm) soil moisture, time-of-season, wind speed, shortwave, radiation, species, precipitation regime, and soil type explained variance in g_s (Fig. 1). For woody plants, in descending order of importance: shortwave radiation, precipitation regime, species, atmospheric $[CO_2]$, VPD, wind speed, deep (50-150cm) soil moisture, time-of-season, time-of-day, and soil type explained variance in g_s (Fig. 1).

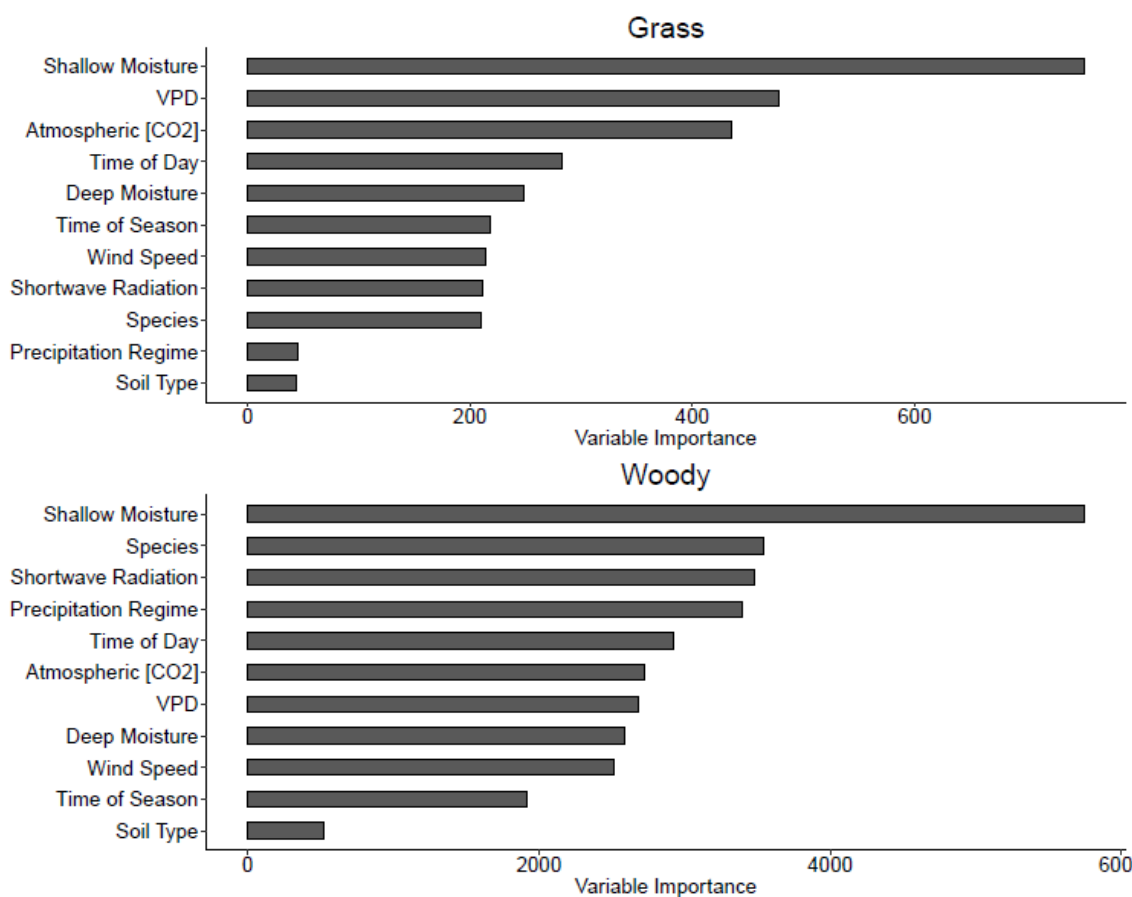


Fig 1. Variable importance in random forest models of stomatal conductance for grasses and woody plants. Variable importance is the difference in prediction error before and after a predictor variable is randomly permuted. Large variable importance values indicate that specifying the variables incorrectly increases prediction error.

Estimated grass g_s increased with shallow soil moisture, decreased with VPD, and decreased with shortwave radiation beyond $1250 \mu\text{mol m}^{-2} \text{s}^{-1}$ (Fig. 2). Estimated woody g_s increased with soil moisture, both shallow and deep, showed a hump-shaped response to VPD that peaked near 1 kPa, and showed a hump-shaped response to shortwave radiation, peaking near $500 \mu\text{mol m}^{-2} \text{s}^{-1}$. Both grass and woody plant estimated g_s increased with increasing atmospheric $[\text{CO}_2]$. Wind speed did not exhibit a clear relationship with grass or woody plant g_s . Estimated g_s also differed among categorical

variable levels. Mean estimated grass g_s decreased (3 %) over the growing season. Mean estimated g_s for woody plants peaked mid-season. Mean estimated grass g_s differed by less than 1 % between wet sites and dry sites and clay sites and sand sites (Fig. 3). Mean estimated woody plant g_s was 16 % greater in wet sites than dry sites and 1.5 % greater in clay sites than sand sites (Fig. 3).

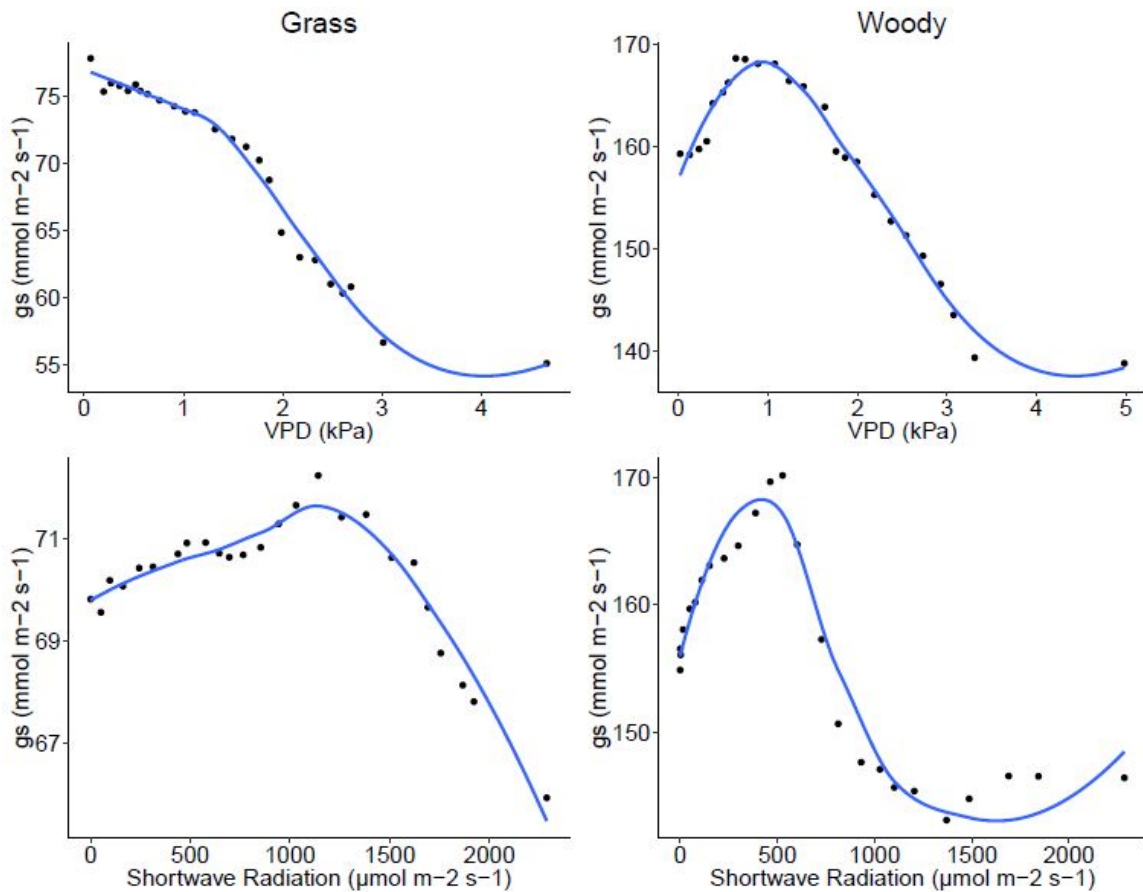


Fig 2. Partial dependence plots for VPD and shortwave radiation for grasses and woody plants. The top panels show estimated grass and woody plant g_s as a function of VPD. The bottom panels show estimated grass and woody plant g_s as a function of shortwave radiation. Partial dependence is determined by averaging the effects of the other predictors and predicting how the response variable changes with the predictor of interest alone.

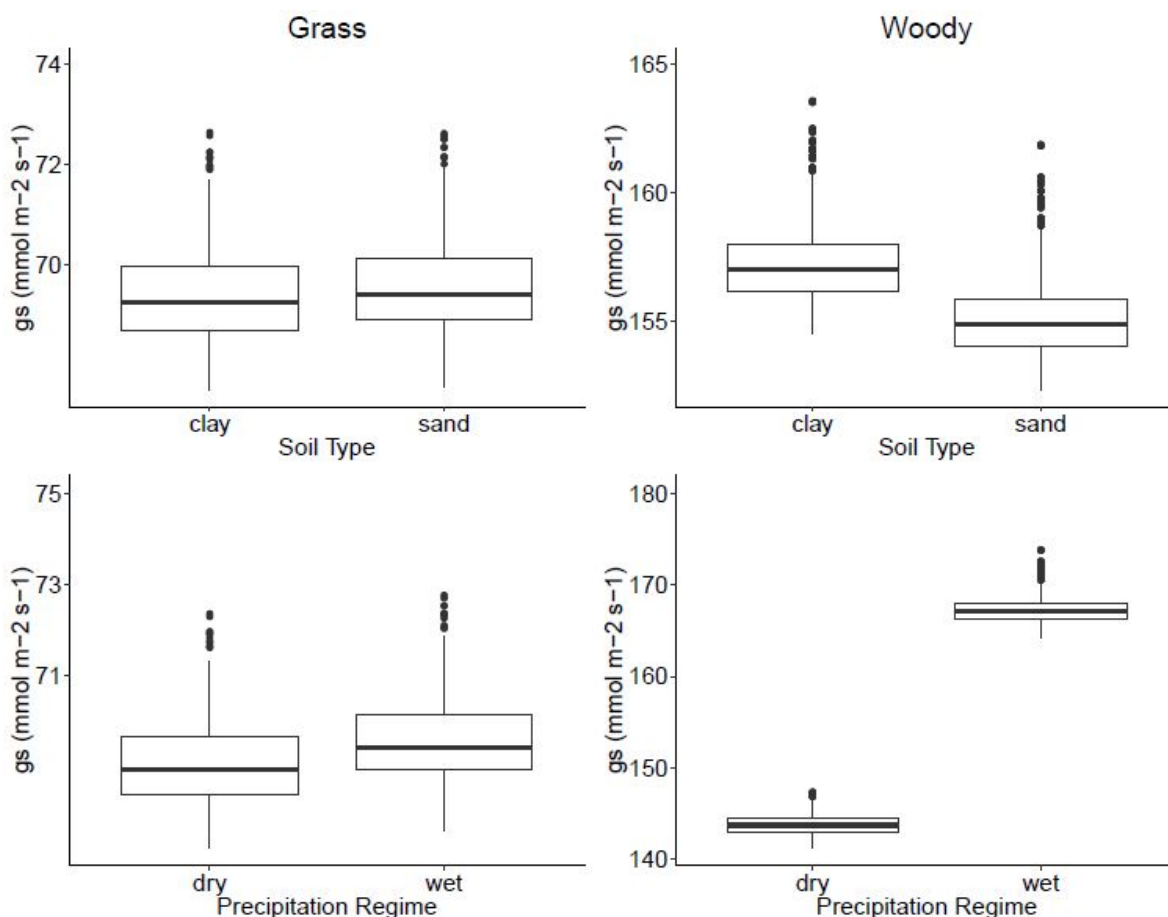


Fig 3. Partial dependence plots for soil type and precipitation regime. The top panels show estimated grass and woody plant g_s as a function of soil type. The bottom panels show estimated grass and woody plant g_s as a function of precipitation regime. Partial dependence is determined by averaging the effects of the other predictors and predicting how the response variable changes with the predictor of interest alone.

3.3 Model Predictions of g_s

When predicted by parameterizing our model with three years of observed environmental data, mean daytime g_s across the four study sites was 67 and 158 mmol m⁻² s⁻¹ for grasses and woody plants, respectively. The wet/sand site had the greatest predicted daytime g_s for both grasses and woody plants (Fig. 4). Mean predicted daytime g_s was 14 % - 24 % and 66 % - 92 % greater in the wet/sand site than other sites for

grasses and trees, respectively. Mean predicted daytime g_s differed by less than 10 % between the dry/clay and dry/sand sites for both grasses and woody plants. Mean predicted daytime g_s differed between species (Fig. 5), although it should be noted that species were not evenly distributed across the abiotic conditions of the park. Predicted g_s for grasses and woody plants decreased to a peak mid-morning before decreasing until sunset (Fig. 6). In general, predicted g_s peaked mid-growing season for both grasses and woody plants.

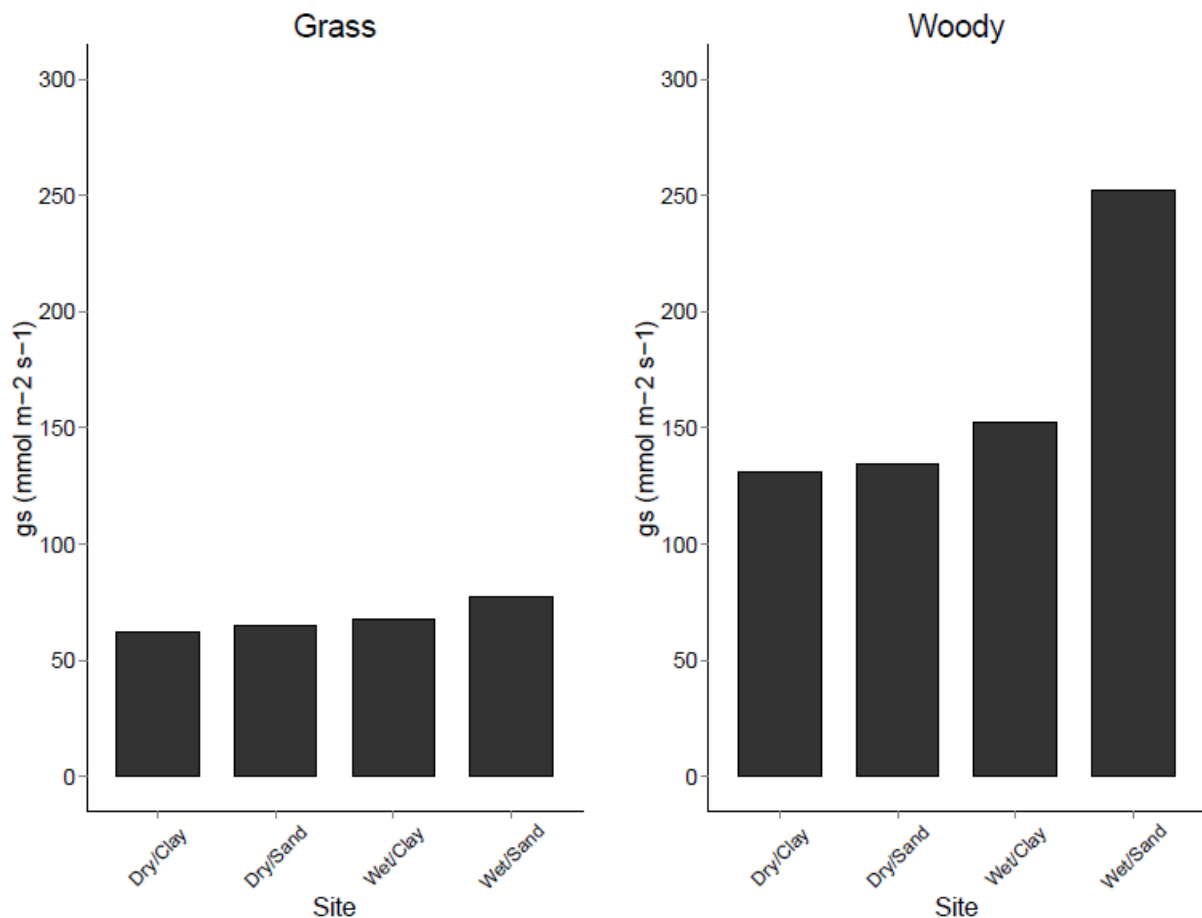


Fig 4. Mean modeled daily g_s for each study site. Modeled daytime g_s was greatest for the wet/sand site for both grasses and woody plants. g_s differed by less than 10 % between the dry/clay and dry/sand sites for both grasses and woody plants.

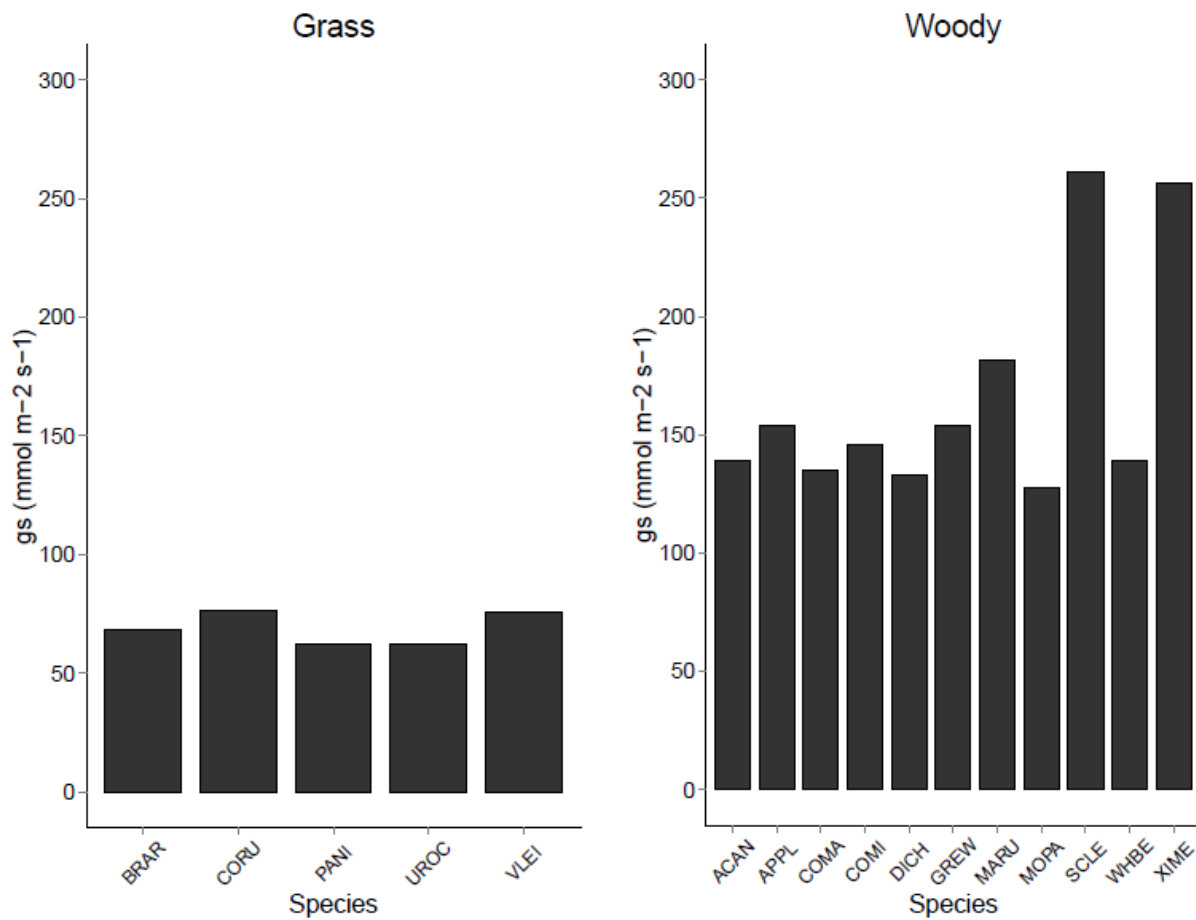


Fig 5. Mean modeled daytime g_s for each species. BRAR = *Bothriochloa radicans*, CORU = *Loudetia simplex*, PANI = *Panicum spp.*, UROC = *Urochloa mosambicensis*, VLEI = *Setaria incrassata*, ACAN = *Acacia nigrescens*, APPL = *Lonchocarpus capassa*, COMA = *Combretum apiculatum*, COMI = *Combretum imberbe*, DICH = *Dichrostachys cinerea*, GREW = *Grewia bicolor*, MARU = *Sclerocarya birrea*, SCLE = *Terminalia sericea*, WHBE = *Securinega virosa*, XIME = *Ximenia caffra*.

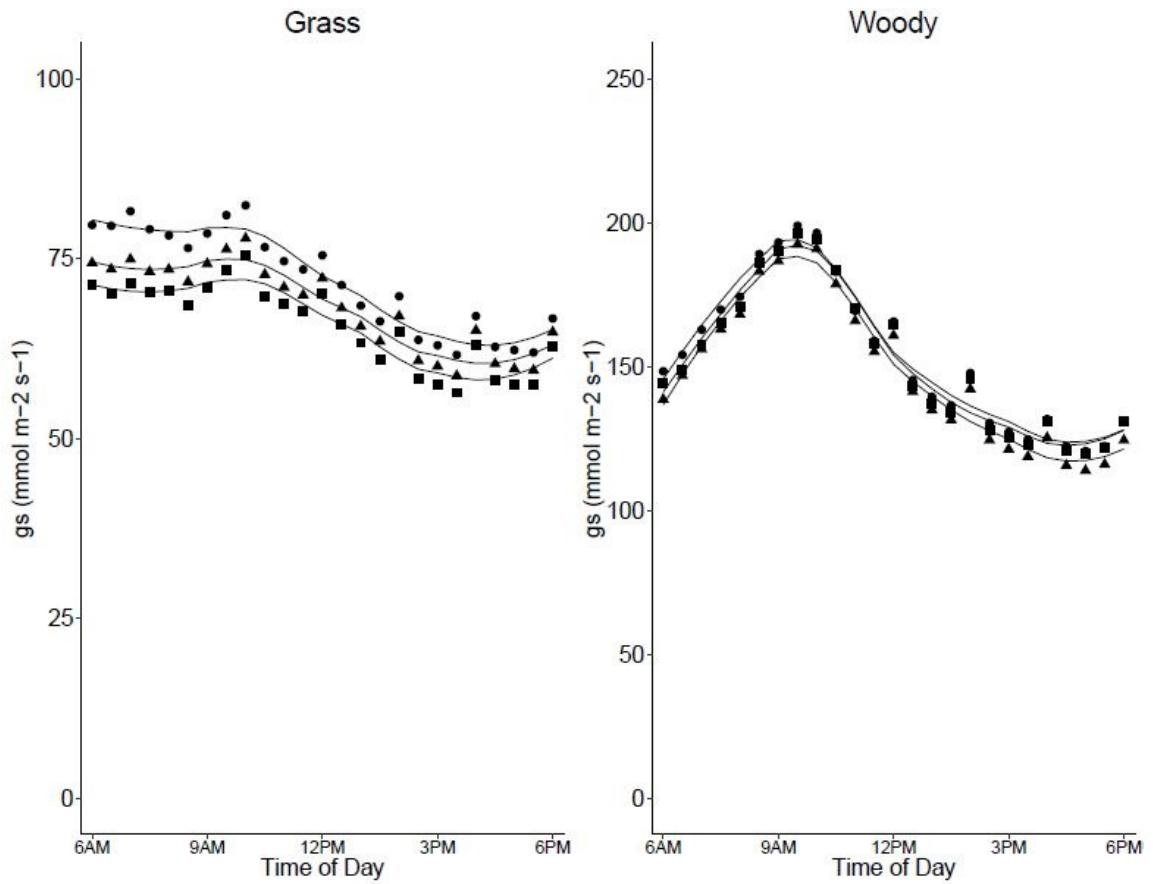


Fig 6. Modeled daily g_s for grasses and woody plants at the dry/sand site. The shape of the points indicates the species. In the grass panel: circles for *Bothriochloa radicans*, triangles for *Panicum* spp., and squares for *Urochloa mosambicensis*. In the woody panel: circles for *Acacia nigrescens*, triangles for *Dichrostachys cinerea*, and squares for *Securinega virosa*. Model predictions were averaged for each timestep and mid-season values are shown.

DISCUSSION

Using a large, field-based dataset, our results highlighted plant identity (functional type or species) and shallow soil moisture as primary drivers of g_s across the sub-tropical savanna landscape of Kruger National Park. These results stand in contrast to a large body of laboratory-based research that has emphasized the role of environmental variables such as temperature^{3,35,36}, VPD^{35,37,38}, solar radiation^{3,35,37}, and atmospheric $[CO_2]$ ^{35,39,40} as major drivers of g_s . Our analyses did detect an effect of VPD on g_s , but found that it was of secondary importance to plant functional type or species identity and shallow soil moisture. For several other meteorological variables, such as atmospheric $[CO_2]$ and shortwave radiation, however, the response of g_s found here was less consistent with previous research. The data and model reported here, therefore, provide a perspective on landscape scale values and drivers of g_s that differs from many laboratory-based approaches.

G_s has been shown to increase with radiation⁶ until a threshold of maximum g_s is reached³⁵. Here, both grasses and trees showed a hump-shaped pattern of g_s with increasing shortwave radiation. For grasses, g_s increased slightly with shortwave radiation until approximately $1100 \mu\text{mol m}^{-2} \text{s}^{-1}$ before decreasing. For woody plants, the threshold was lower at $500 \mu\text{mol m}^{-2} \text{s}^{-1}$. In this study, the effect of radiation on g_s was assessed by averaging all other observed variable values across a range of radiation values. This should have allowed the detection of an increasing relationship between radiation and g_s , unless, under natural conditions, it is the case that some variables limited g_s as radiation increased. It is likely, for example, that plants exhausted plant available

water immediately around their roots by midday so that no naturally occurring g_s values increase with radiation through the day⁴¹. This midday depression in g_s has been documented in the field^{39,42}. Where soils are consistently well-watered, g_s may continue increasing as shortwave radiation increases¹⁴ but consistently well-watered soils may only occur during heavy rains when shortwave radiation values do not reach high values. This hypothesis highlights both the strengths and weaknesses of the field data approach used in this study. Data from this study provide more realistic estimates of g_s during the years and conditions of this study, but do not provide inference to conditions unlike those observed during the study (*i.e.*, extreme conditions associated with climate change).

G_s also increased with atmospheric $[CO_2]$. This relationship is surprising, as previous studies have shown the opposite: g_s decreased as ambient $[CO_2]$ increased, presumably because plants could rapidly assimilate and close stomata to reduce water loss^{7,14,41,43}. Other variables, like VPD and soil moisture, might mask the response of g_s to atmospheric $[CO_2]$. Atmospheric $[CO_2]$ was highest in the morning and decreased throughout the day (data not shown). Thus, atmospheric $[CO_2]$ decreased as shortwave radiation and VPD were likely to increase and soil moisture likely to decrease. Because shallow soil moisture was a primary driver of g_s , it is likely that g_s decreased as a result of water stress rather than decreasing $[CO_2]$. Regardless of the mechanism, our results suggest that the laboratory-based observations of $[CO_2]$ effects on g_s were overwhelmed by the effects of other environmental conditions.

To estimate how environmental variables affect g_s on the landscape, g_s was modeled for four sites that represented a broad range of abiotic conditions from fairly mesic to fairly xeric savanna⁴⁴ and clay to sand soils. G_s was surprisingly similar among

most sites, with the exception of the wet/sand site. g_s was 14 % - 24 % and 66 % - 92 % greater in the wet/sand site than the other sites for grasses and woody plants, respectively. Also surprising was that soil type did not appear to have a consistent effect on g_s . The lack of a consistent response of g_s to soil type suggests that model results are applicable across edaphic gradients. Again, results emphasize the importance of understanding soil moisture effects, but suggest that these soil moisture effects are consistently important across a wide range of soil types and plant species. This is important because it suggests that our models of g_s may be applicable across a wide range of abiotic conditions.

This study highlights the difficulties of modeling g_s at the landscape scale. Despite large sample sizes, the dataset was highly variable, and our models explained only a modest proportion of variance (25 % for grasses and 54 % for woody plants, respectively). Plant-to-plant variation was anticipated to have explained a large portion of the variation in the dataset¹⁴; however, averaging measurements over 2-hour increments provided nominal improvements (*i.e.*, <2 % of error; data not shown). This suggested that plant-to-plant variation explained little of the residual variance. Plant age was not included as a parameter in this study, but may have accounted for some of the unexplained variation^{14,41}. A more likely source of variation is leaf-level environmental conditions. As stomatal aperture can change in response to leaf-level conditions, such as interstomatal [CO₂], leaf water potential, and leaf temperature, including these leaf-level parameters may be necessary to explain much of the unexplained variance in our dataset. Indeed, leaf-level models of g_s that incorporate these types of parameters often explain upwards of 80% of variation in g_s ¹¹.

Results provided novel insight into grass and woody plant rooting patterns in savannas. For nearly a century, Walter's two-layer hypothesis has suggested that grasses and trees can coexist because grasses use shallow soil moisture and trees use deep soil moisture^{24,37,45}. Our results are consistent with recent findings from hydrologic tracer experiments in Kruger Park that indicated that both grasses and trees rely on shallow soil water but that trees rely slightly more on deeper water than grasses^{21,27}. More specifically, shallow water was found to be about 3.7 times more important in explaining variance in g_s than deep water for grasses (Fig. 1). In contrast, shallow water was 2.1 times more important than deep water for woody plants. This suggested that both grasses and trees rely more on shallow than deep water but that grasses rely even more on shallow water than trees. Further, the fact that precipitation regime was more important to woody plants than grasses also supports the idea that woody plants rely more on deeper soil water than grasses. This is because wetter sites were more likely to realize deeper soil water penetration, which is likely to be more important to woody plants than grasses.

This study provides a prioritized list of variables important to landscape g_s in this region. Results indicate that plant identity and shallow soil moisture are of greater importance than atmospheric conditions and several environmental drivers that are commonly included in models of g_s . Model performance decreased markedly when species was replaced with plant functional type as a predictor, and decreased even more when neither was included in the model. While incorporating species into a global circulation or land surface model may not be practical, it is possible to include plant functional type data and species-level data may be useful for increasing accuracy in canopy or ecosystem-level modeling.

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APPENDICES

Appendix A. SUPPLEMENTARY INFORMATION

Table A.1 Studied species and their respective common names, families, and growth forms.

Species	Common/Alt. Names	Family	Growth Form
<i>Acacia nigrescenes</i>	Knobthorn	Fabaceae	Tree
<i>Acacia gerrardii</i>	Red thorn	Fabaceae	Tree
<i>Acacia tortilis</i>	Umbrella thorn	Fabaceae	Tree
<i>Albizia harveyi</i>	Common false thorn	Fabaceae	Tree
<i>Lonchocarpus capassa</i>	Apple-leaf	Fabaceae	Tree
<i>Aristida sp.</i>		Poaceae	Grass
<i>Euclea crispa</i>	Blue guarri	Ebenaceae	Tree
<i>Bothriochloa radicans</i>	Stinking grass	Poaceae	Grass
<i>Cenchrus ciliaris</i>	Buffelgrass, African foxtail grass	Poaceae	Grass
<i>Loudetia simplex</i>	Common russet grass	Poaceae	Grass
<i>Tragus berteronianus</i>	Carrot seed grass	Poaceae	Grass
<i>Combretum apiculatum</i>	Red bushwillow	Combretaceae	Tree
<i>Combretum hereroense</i>	Russet bushwillow	Combretaceae	Tree
<i>Combretum imberbe</i>	Leadwood	Combretaceae	Tree
<i>Gymnosporia buxifolia</i>	Common spike thorn	Celastraceae	Tree
<i>Dichrostachys cinerea</i>	Sickle bush	Fabaceae	Shrub
<i>Phoenix reclinata</i>	Wild date palm	Arecaceae	Tree
<i>Enneapogon conchroides</i>	Nine-awned grass	Poaceae	Grass
<i>Euclea divinorum</i>	Magic guarri	Ebenaceae	Shrub/tree
<i>Lannea schwinfurthii</i>	False marula	Anacardiaceae	Tree
<i>Digitaria erianthra</i>	Common finger grass	Poaceae	Grass
<i>Grewia bicolor</i>	White raisin	Malvaceae	Tree
<i>Hyperthelia dissolute</i>	Yellow thatching grass	Poaceae	Grass
<i>Hyparrhenia filipendula</i>		Poaceae	Grass
<i>Hyparrhenia hirta</i>	Common thatching grass	Poaceae	Grass
<i>Adenium multiflorum</i>	Impala lily	Apocynaceae	Forb
<i>Sclerocarya birrea</i>	Marula	Anacardiaceae	Tree
<i>Maerua angolensis</i>	Bead bean	Capparaceae	Tree
<i>Melinis repens</i>	Natal grass	Poaceae	Grass
<i>Strychnos madagascariensis</i>	Black monkey orange	Loganiaceae	Shrub/tree
<i>Colophospermum mopane</i>	Mopane	Fabaceae	Shrub/tree
<i>Panicum coloratum</i>	Small buffalo grass	Poaceae	Grass
<i>Panicum maximum</i>	Guinea grass	Poaceae	Grass
<i>Pogonarthria squarrosa</i>	Herringbone grass	Poaceae	Grass

<i>Ehretia rigida</i>	Sand paper bush	Boraginaceae	Tree
<i>Terminalia sericea</i>	Silver cluster-leaf	Combretaceae	Tree
<i>Heteropogon contortus</i>	Spear grass	Poaceae	Grass
<i>Setaria sphacelata</i>	Creeping bristle grass	Poaceae	Grass
<i>Themeda triandra</i>	Red oat grass	Poaceae	Grass
<i>Urochloa mosambicensis</i>	Bushveld signal grass	Poaceae	Grass
<i>Vangueria infausta</i>	Wild medlar	Rubiaceae	Shrub/tree
<i>Setaria incrassate</i>	Vlei bristle grass	Poaceae	Grass
<i>Securinega virosa</i>	White berry bush	Phyllanthaceae	Tree
<i>Ximenia caffra</i>	Sour plum	Olacaceae	Tree
<i>Dalbergia melanoxylon</i>	Zebra wood	Fabaceae	Tree
<i>Ziziphus mucronata</i>	Buffalo thorn	Rhamnaceae	Tree

Table A.2 Mean observed daytime grass g_s and summary statistics.

Site	Mean	SD	SE
Dry/Clay	66.92703	48.31501	2.178209
Dry/Sand	65.20629	38.89313	3.084426
Wet/Sand	62.27551	47.17955	1.939068
Wet/Clay	95.97528	79.05191	3.408170

Table A.3 Mean observed daytime woody plant g_s and summary statistics.

Site	Mean	SD	SE
Dry/Clay	123.4944	71.23182	2.433232
Dry/Sand	128.6277	78.24701	3.105138
Wet/Sand	150.4774	116.4686	2.742905
Wet/Clay	197.6815	136.1406	3.939902

Table A.4 Mean modeled daytime grass g_s and summary statistics.

Site	Mean	SD	SE
Dry/Clay	62.4447	17.33135	0.090633
Dry/Sand	64.82193	17.79558	0.092633
Wet/Sand	77.50184	15.41926	0.085978
Wet/Clay	67.82299	18.4539	0.134779

Table A.5 Mean modeled daytime woody plant g_s and summary statistics.

Site	Mean	SD	SE
Dry/Clay	131.2803	31.64087	0.202651
Dry/Sand	134.822	35.12521	0.111966
Wet/Sand	252.5363	56.23957	0.313591
Wet/Clay	152.4924	46.67333	0.241039

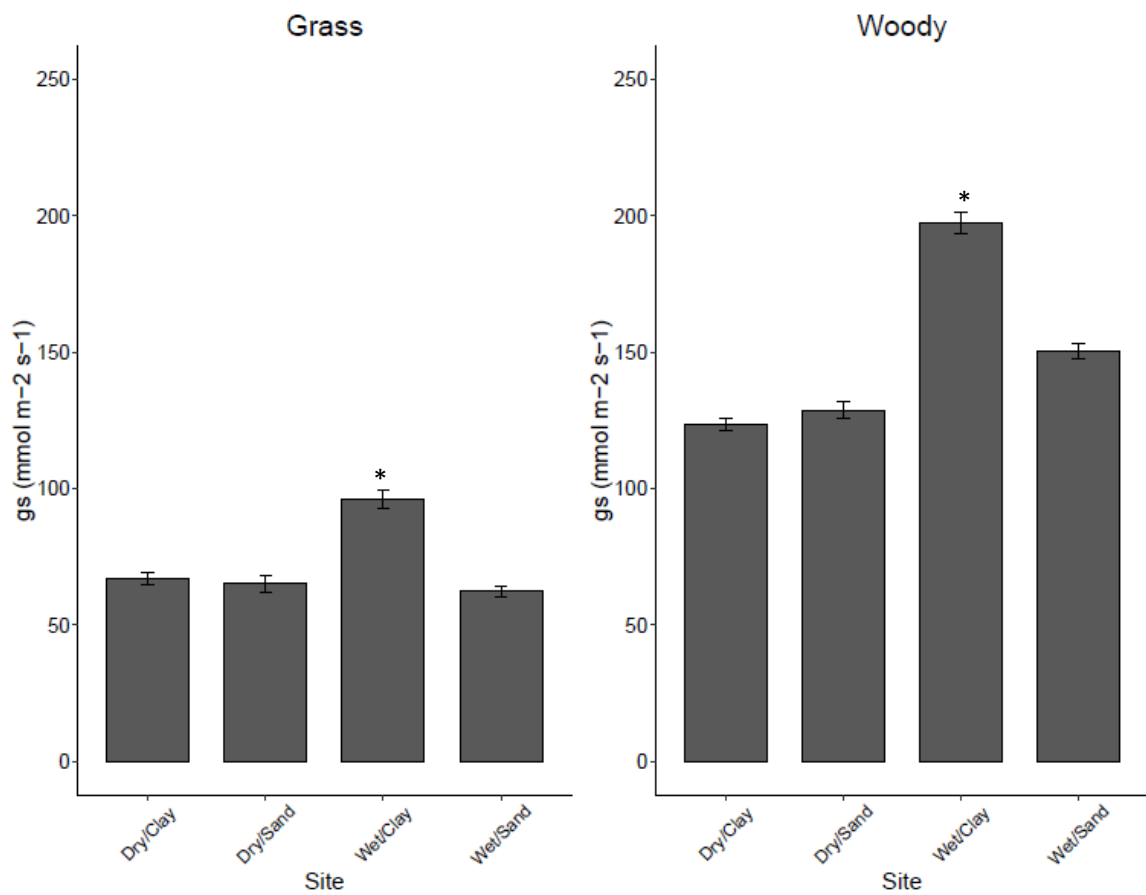


Fig. A1 Mean observed daytime g_s for Kruger National Park and each study site. Error bars represent standard error. Asterisks indicate significance ($p < 0.05$).

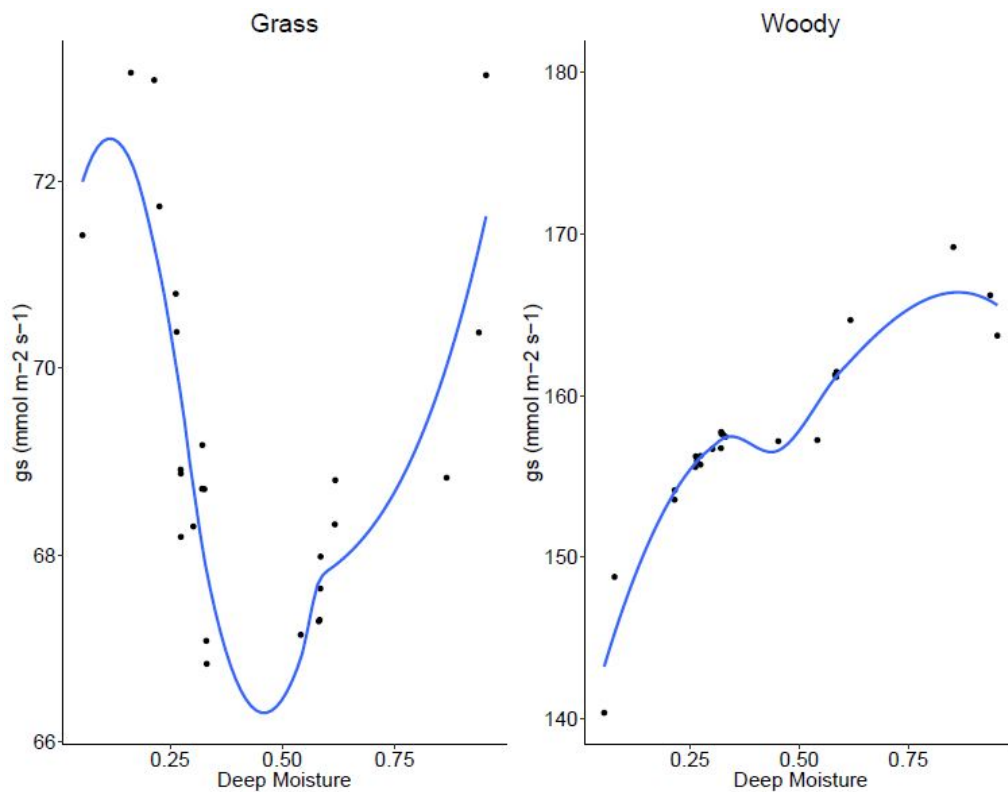


Fig A.2 Partial dependence plots for deep soil moisture for grasses and woody plants. The panels show predicted grass and woody plant g_s as a function of deep soil moisture.

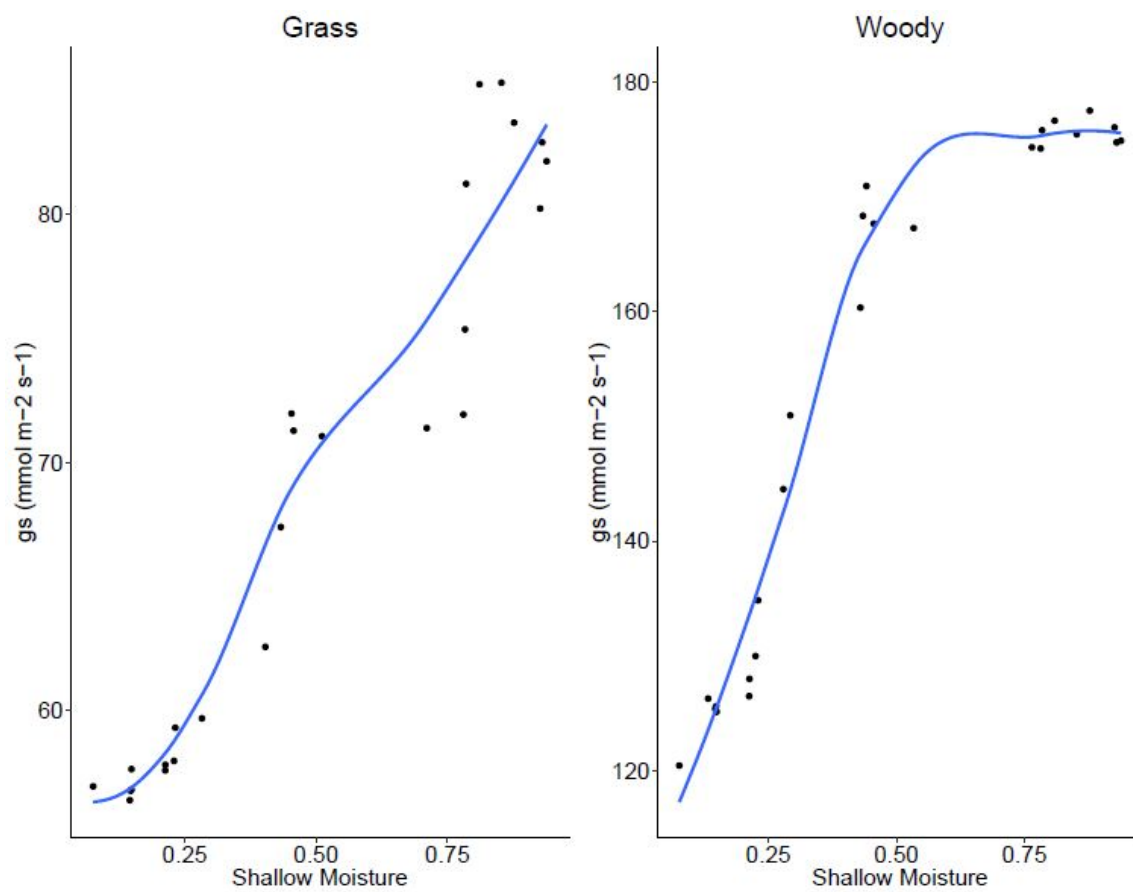


Fig A.3 Partial dependence plots for shallow soil moisture for grasses and woody plants. The panels show predicted grass and woody plant g_s as a function of shallow soil moisture.

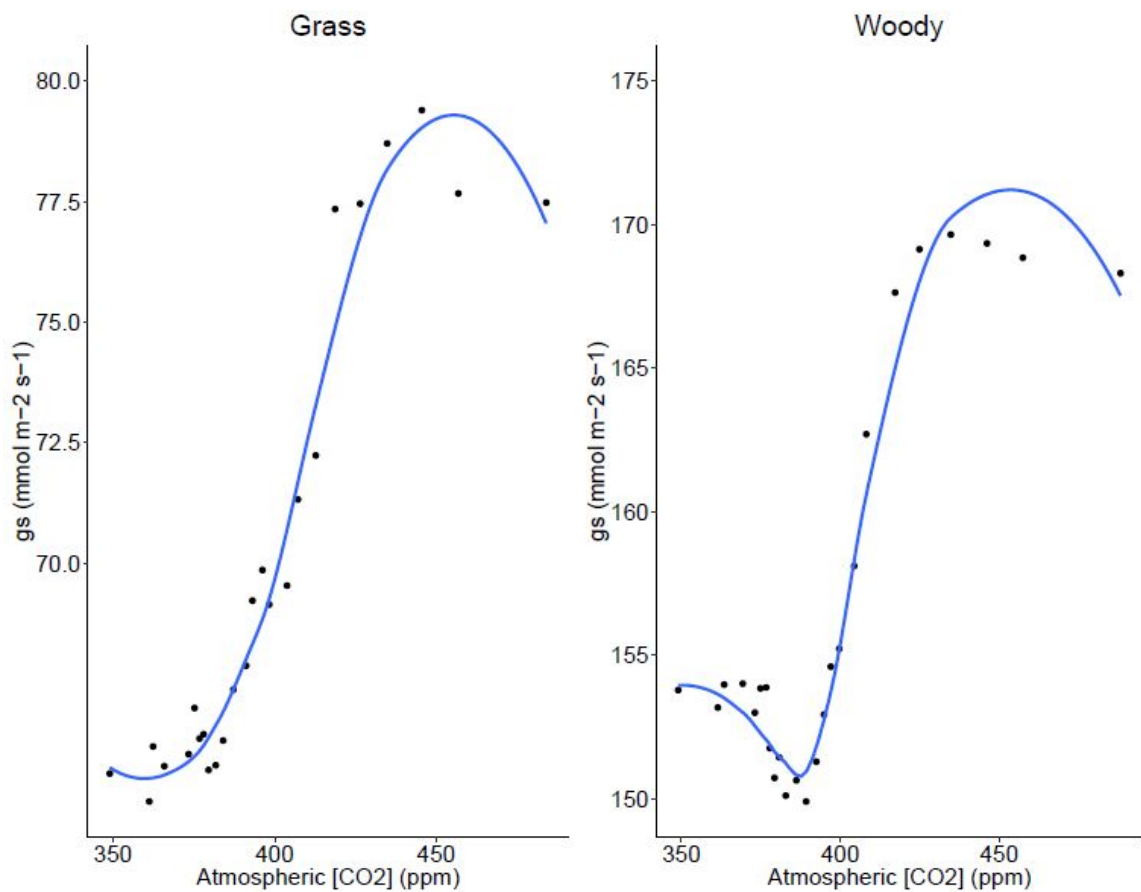


Fig A.4 Partial dependence plots for atmospheric [CO₂] for grasses and woody plants. The panels show predicted grass and woody plant g_s as a function of atmospheric [CO₂].

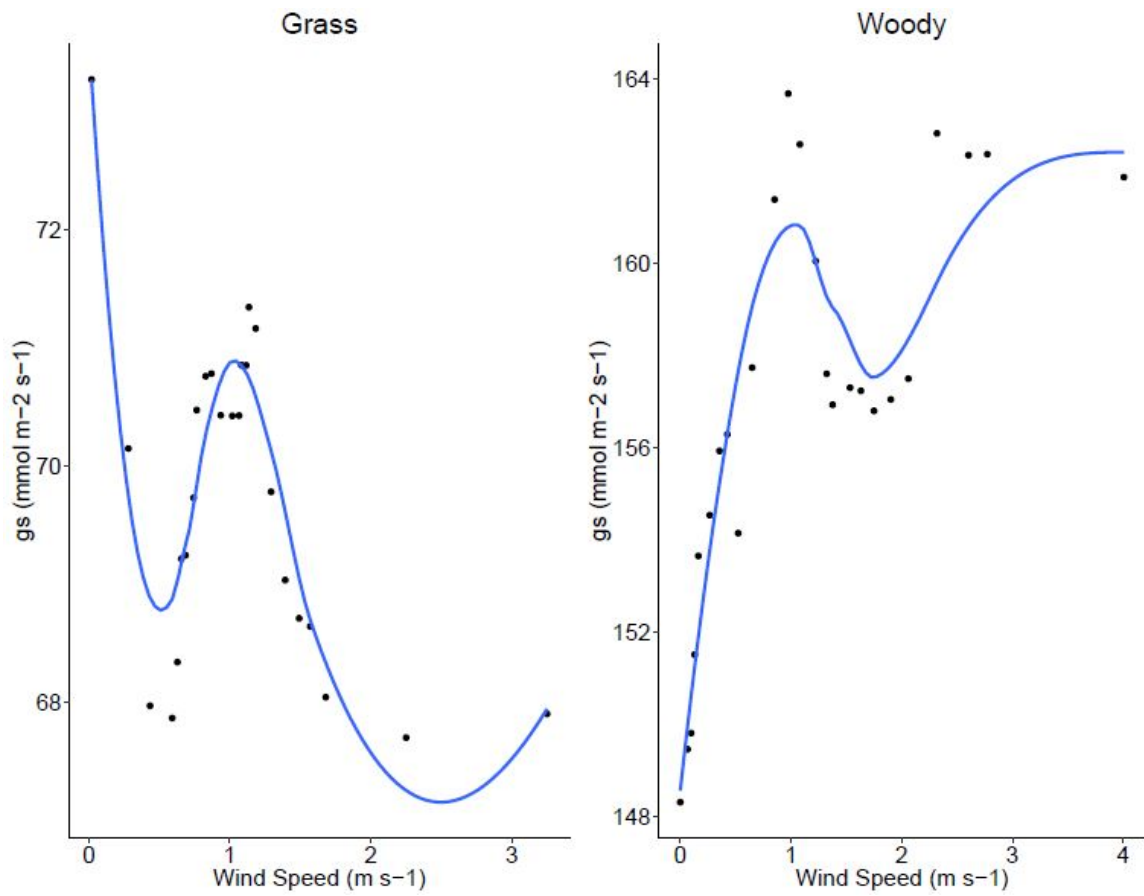


Fig A.5 Partial dependence plots for wind speed for grasses and woody plants. The panels show predicted grass and woody plant g_s as a function of wind speed.

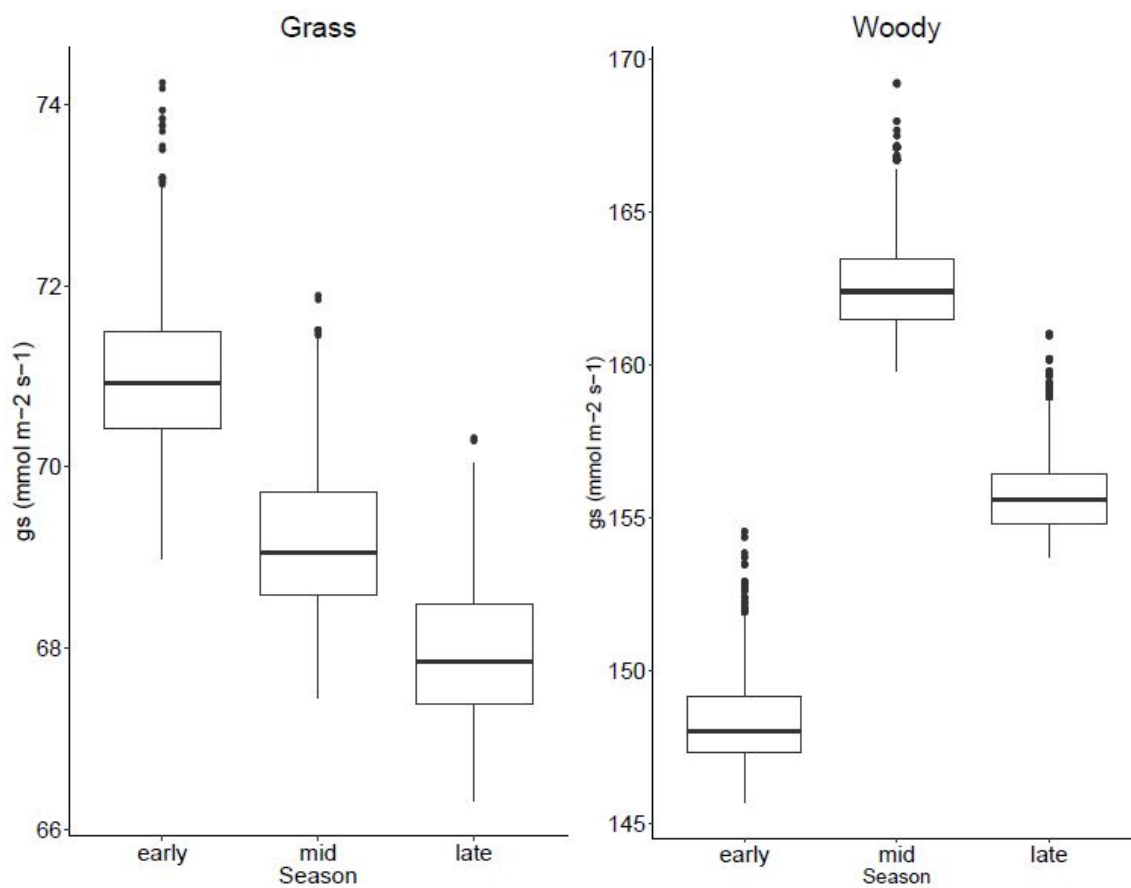


Fig A.6 Partial dependence plots for time of season for grasses and woody plants. The panels show predicted grass and woody plant g_s as a function of time of season.

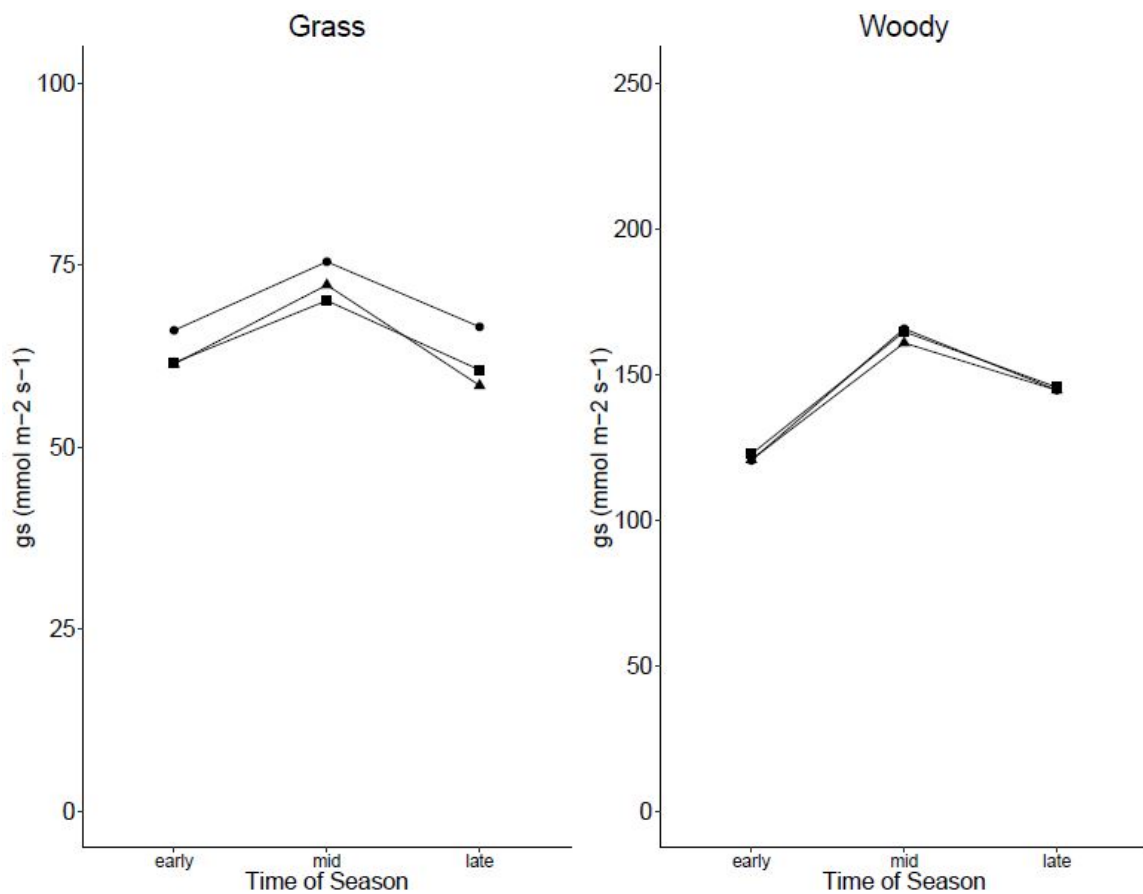


Fig A.7 Modeled seasonal g_s for grasses and woody plants at the dry/sand site. The shape of the points indicates the species. In the grass panel: circles for *Bothriochloa radicans*, triangles for *Panicum spp.*, and squares for *Urochloa mosambicensis*. In the woody panel: circles for *Acacia nigrescens*, triangles for *Dichrostachys cinerea*, and squares for *Securinega virosa*. Model predictions were averaged for each season and midday values are shown.

Appendix B. STATISTICAL RESULTS

Table B.1 One-way analysis of variance (Type III) of observed mean daytime grass gs values from each study site. See Table 1 for a description of each site.

Source	Sum of Squares	Mean Square	dF	F	p
Site	49.8	16.602	3	28.399	<0.001
Error	1104.3	0.585	1889		
Total	1154.1	17.187	1892		

Table B.2 One-way analysis of variance (Type III) of observed mean daytime woody plant gs values from each study site. See Table 1 for a description of each site.

Source	Sum of Squares	Mean Square	dF	F	p
Site	179.6	59.87	3	77.298	<0.001
Error	3472.8	0.77	4484		
Total	3652.4	60.64	4487		

Table B.3 Tukey test for pairwise comparisons of mean observed daytime gs values for grasses by site. See Table 1 for a description of wet, dry, sand, and clay.

Comparison	Estimate	SE	t-statistic	p
Dry/Sand – Dry/Clay	0.019	0.046	0.406	0.977
Wet/Sand – Dry/Clay	0.009	0.037	0.246	0.995
Wet/Clay – Dry/Clay	0.461	0.039	11.698	<0.001
Wet/Sand – Dry/Sand	-0.010	0.041	-0.239	0.995
Wet/Clay – Dry/Sand	0.442	0.043	10.234	<0.001
Wet/Clay – Wet/Sand	0.452	0.033	13.768	<0.001

Table B.4 Tukey test for pairwise comparisons of mean observed daytime gs values for woody plants by site. See Table 1 for a description of wet, dry, sand, and clay.

Comparison	Estimate	SE	t-statistic	p
Dry/Sand – Dry/Clay	0.075	0.070	1.074	0.699
Wet/Sand – Dry/Clay	-0.036	0.045	-0.802	0.849
Wet/Clay – Dry/Clay	0.344	0.048	7.209	<0.001
Wet/Sand – Dry/Sand	-0.111	0.067	-1.652	0.341
Wet/Clay – Dry/Sand	0.269	0.069	3.894	<0.001
Wet/Clay – Wet/Sand	0.380	0.044	8.669	<0.001