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DOI:

[10.1016/j.ijbiomac.2020.07.113](https://doi.org/10.1016/j.ijbiomac.2020.07.113)

Document Version

Accepted author manuscript

[Link to publication record in Manchester Research Explorer](#)

Citation for published version (APA):

Baruah, J., Chaliha, C., Kalita, E., Nath, B. K., Field, R. A., & Deb, P. (2020). Modelling and optimization of factors influencing adsorptive performance of agrowaste-derived Nanocellulose Iron Oxide Nanobiocomposites during remediation of Arsenic contaminated groundwater. *International Journal of Biological Macromolecules*, 164, 53-65. <https://doi.org/10.1016/j.ijbiomac.2020.07.113>

Published in:

International Journal of Biological Macromolecules

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Modelling and optimization of factors influencing adsorptive performance of agrowaste-derived Nanocellulose/Iron Oxide Nanobiocomposites during remediation of Arsenic contaminated groundwater

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Highlights:

- Nanocellulose based superparamagnetic adsorbents developed for ~99% As removal
- CCD-based RSM and MLP-based ANN models used for modelling and optimizing As removal
- Kinetic studies suggest chemisorptive removal of As
- Adsorbents are recyclable and better than commercial adsorbents
- Adsorbents effective for removal of As from contaminated groundwater

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Abstract

Nanocellulose/Iron Oxide Nanobiocomposites (NIONs) were synthesized from rice husk and sugarcane bagasse derived nanocelluloses for adsorptive removal of arsenic and associated contaminants present in groundwater samples. These NIONs were superparamagnetic, hence magnetically recoverable and demonstrated promising recyclability. Synthesis of NIONs was confirmed by Transmission electron microscopy (TEM), X-Ray Diffraction (XRD), Fourier transform infrared spectroscopy (FTIR) and X-ray photoelectron spectroscopic (XPS). FTIR and XPS data together with adsorption kinetics provide insights into probable adsorption mechanism of Arsenic by NIONs. The experimental conditions for 10 different variants were modelled using response surface methodology (RSM) based on central composite design (CCD), considering the

parameters; adsorbate dosage, adsorbent dosage, pH and contact time. The results identified the best performing variants and the optimal conditions for maximal absorption (~99%). These results were validated using a three-layer feed-forward Multilayer Perceptron (MLP) based Artificial Neural Network (ANN) model. Both RSM and ANN chemometric models were in close conformity for optimized conditions of highest adsorption by specific variants. The standardized conditions were used to expand the study to field-based arsenic contaminated groundwater samples and their performance to commercial adsorbents. NIONs show promising commercial potential for water remediation applications due to their high adsorptive performance, magnetic recoverability and recyclability.

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25 **14 Abstract**

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27 **15** Nanocellulose/Iron Oxide Nanobiocomposites (NIONs) were synthesized from rice husk and sugarcane bagasse derived nanocelluloses for
28 **16** adsorptive removal of arsenic and associated contaminants present in groundwater samples. These NIONs were superparamagnetic, hence
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30 **18** (TEM), X-Ray Diffraction (XRD), Fourier transform infrared spectroscopy (FTIR), and X-ray photoelectron spectroscopic (XPS). FTIR and XPS
31 **19** data together with adsorption kinetics provide insights into probable adsorption mechanism of Arsenic by NIONs. The experimental conditions for
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35 **23** Network (ANN) model. Both RSM and ANN chemometric models were in close conformity for optimized conditions of highest adsorption by
36 **24** specific variants. The standardized conditions were used to expand the study to field-based arsenic contaminated groundwater samples and their
37 **25** performance to commercial adsorbents. NIONs show promising commercial potential for water remediation applications due to their high
38 **26** adsorptive performance, magnetic recoverability and recyclability.

39 **27** **Keywords:** Nanobiocomposite; Remediation; Chemometric model
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47 **28 1. Introduction**

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51 **29** Groundwater contamination with arsenic (As) represents emergent health and environmental issue worldwide due to the high toxicity of As that
52 **30** have been found at chronic levels in most countries (e.g., China, India, Indonesia, Pakistan, Bangladesh, USA, Hungary, Canada, Brazil, and
53 **31** Taiwan) [1, 2]. World Health Organization (WHO) has strictly limited the Maximum Contaminant Level (MCL) of As in drinking water to 10 µg/L
54 **32** [3]. Therefore, removal of As from contaminated water to meet the drinking water standards is being pursued fervently. A variety of treatment
55 **33** processes such as precipitation [4], membrane filtration [5], electro-coagulation [6], adsorption [7], and ion-exchange [8] are available to reduce As
56 **34** concentration in water. However, their high operational costs along with the production of toxic sludge are important factors that should be
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4 35 considered before the section of treatment method. In this context, biosorption using low-cost biosorbents derived from abundant renewable
5 36 resources has seen promising advancement to remove As from water thus proving a sustainable and eco-friendly solution [9, 10].
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8 37 Over the past few years, the use of biomaterials such as cellulose, chitosan, chitin, starch, gelatin and alginate for water remediation has the
9 38 advantages of biocompatibility and non-toxicity compared with non-biological adsorbents [11]. Among such biomaterials, cellulose has superior
10 39 advantages of being renewable, biodegradable, biophilic, large abundance in nature, and low cost [11, 12]. Since it has a large number of hydroxyl
11 40 groups, natural cellulose has excellent adsorption properties for harmful metal ions [13]. However, cellulose also forms an unstable complex
12 41 between the β -D-glucose unit and heavy metal ions, reducing the adsorption capacity and limiting its potential. To overcome this limitation different
13 42 approaches have been reported to date [14, 15, 16]. Among such approaches, converting cellulose into a nano-sized material exhibits excellent
14 43 properties like high surface area, high mechanical resistance, and stability that has been reported to show improved adsorption efficiency than the
15 44 natural cellulose [17, 18, 19].
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22 45 In recent years, nanoscale inorganic metal oxides and hydroxides have shown significant outcomes in the removal of As and other groundwater
23 46 contaminants [20]. The hydroxyl groups in the cellulose structure complex well with metal ions and therefore there is homogeneous dispersion of
24 47 inorganic nanomaterials onto the cellulose matrix with improved functions. Such chemically modified nanocelluloses have shown significant
25 48 potential in the remediation of heavy metals, such as cadmium, lead, arsenic, and even radioactive pollutants like uranium [21, 22, 23, 24]. For
26 49 instance, Wei et al. 2019 have reported the preparation of magnetic hybrid aerogel by integrating nanocellulose and ferroferric oxide (Fe_3O_4)
27 50 nanoparticles for effective removal of chromium (Cr)(VI) ions. This hybrid aerogel also exhibits similar adsorption behavior on lead (Pb)(II) and
28 51 copper (Cu)(II) ions suggesting extended hybrid aerogel applications for removing heavy metal ions [25]. Likewise, Zarei et al. 2018 reported the
29 52 synthesis of Fe_3O_4 -nanocellulose compounds by acid hydrolysis and co-precipitation as well as sol-gel methods for the removal of mercury from
30 53 wastewaters in the coastal cities of the Persian Gulf [26]. In another study, Taleb et al. 2016 reported the performance of two nanocellulose based
31 54 adsorbents: (i) nanocellulose (NC) modified by PEG-6-arm amino polyethylene glycol, PEG-NH₂ (NC-PEG) for Cd^{2+} and Ni^{2+} removal and (ii)
32 55 NC-PEG subsequently modified by precipitation iron oxide (FO) from goethite (NC-PEG/FO) has been found applicable for As(V) and As(III)
33 56 removal [27].
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41 57 The present study is oriented towards developing an effective and environmentally benign solution using nanocellulose based iron-oxide
42 58 nanoparticles for arsenic remediation from arsenic contaminated groundwater. Although various nanocomposites containing nanocellulose and
43 59 iron-oxide nanoparticles had been developed earlier [25, 26, 27], the bottleneck lies in their synthesis procedures that involve the use of toxic
44 60 chlorine-based compounds as bleaching agents during nanocellulose extraction which are established as potent pollutants and carcinogens. Besides,
45 61 the acid concentration for the isolation processes range $\sim 64\%$ (W/W) H_2SO_4 , and is often criticized for the environmental damage it causes and
46 62 the incremental costs involved. Therefore, the necessity for developing a simple and appropriate method for synthesis of effective nanosorbents
47 63 retaining high adsorption capacities and field applicability still exists. In the current study, we have reported two magnetically recoverable adsorbent
48 64 systems for the sorptive removal of total dissolved As. The synthesis process involves the extraction of the nanocelluloses from agrowastes via
49 65 total-chlorine free bleaching along with low concentrated acid hydrolysis step followed by single-step incorporation of iron oxide nanoparticles
50 66 into the nanocellulose polymer matrix that makes the process more economical and environment-friendly. The adsorbent derived from rice husk
51 67 bioresource is termed as Rice Husk-Nanocellulose Iron Oxide Nanobiocomposites (RH-NIONs) and the sugarcane bagasse derived adsorbent is
52 68 termed as Sugarcane Bagasse-Nanocellulose Iron Oxide Nanobiocomposites (SB-NIONs). In order to understand the effects of experimental
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4 69 parameters (pH adsorbent dosage, total arsenic concentration, and contact time) on the adsorption of As, an empirical tool named Response surface
5 70 methodology (RSM) was employed. The findings of RSM were then confirmed using Artificial Neural Networks (ANN) as a modeling tool and
6 71 the optimal conditions for total arsenic adsorption by the NIONs were derived. The adsorption process was further analyzed through isotherm and
7 72 kinetic modeling and a comparative assessment of the adsorbents with other metals and field samples was carried. Comparative performance with
8 73 respect to commercially available adsorbents has also been studied.

12 74 2. Materials and methods

14 75 2.1. Materials and chemicals

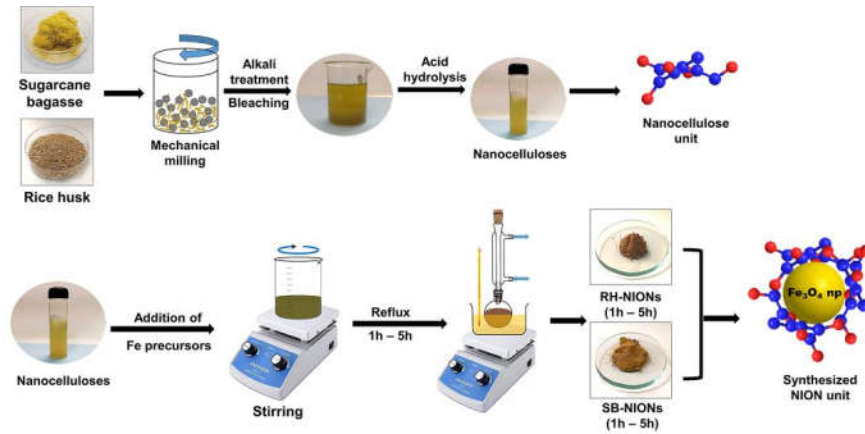
16 76 Rice husk (RH) and Sugarcane bagasse (SB) were collected as process wastes from local rice mills and jaggery production units, located in the
17 77 district of Sonitpur, Assam, India. Sulfuric acid (H_2SO_4), Sodium hydroxide (NaOH), Hydrogen peroxide (H_2O_2), ferrous sulfate heptahydrate
18 78 ($FeSO_4 \cdot 7H_2O$), sodium citrate ($Na_3C_6H_5O_7$), and urea (NH_2CONH_2) of analytical grade were purchased from Merck India Pvt. Ltd. and used without
19 79 further purification for the synthesis of the NIONs.

24 80 2.2. Isolation of nanocelluloses from RH and SB

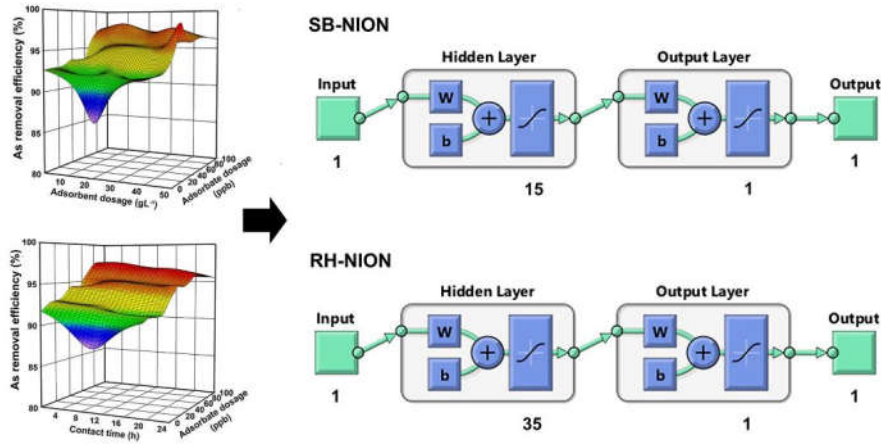
26 81 The isolation of nanocellulose from RH and SB was carried out based on our previous experiences, [28] with appropriate modifications. The SB
27 82 and RH agro-waste precursors were milled separately in an SS316 stainless steel grinding container (50 ml) and processed in a planetary ball mill
28 83 (EGOMA, India) at a frequency of 8.33 Hz until finely milled powders were obtained. The finely milled SB and RH powders (5g) were then
29 84 subjected to chemical processing by soaking in NaOH (3%, w/w) solution which was successively autoclaved in an SS-316 Mechomine Teflon
30 85 lined-autoclave (India) under 137895 Pa (20 Psi) at a temperature of $210 \pm 5^\circ C$, for a period of 45 min. Afterward, chlorine-free bleaching treatment
31 86 was performed to the steam-exploded samples with a solution containing a predetermined amount of H_2O_2 and NaOH. The as-obtained cellulose
32 87 was isolated from RH and SB were then acidified with 5% H_2SO_4 and sonicated for 2 h at room temperature in a Labsonic M Ultrasonic
33 88 Homogenizer from Sartorius (Germany) to obtain the nanocelluloses.

41 89 2.3. Synthesis of RH-NIONs and SB-NIONs

42 90 The nanocelluloses obtained from RH and SB (5g) were added separately into a solution of ferrous sulfate heptahydrate (0.3 mol), urea (1 mol),
43 91 and sodium citrate (0.1 mol) and subjected to solvothermal treatment for five different time intervals viz. 1h, 2h, 3h, 4h, and 5h. The synthesized
44 92 RH-NIONs and SB-NIONs were separated from the solution by a magnet and rinsed with deionized water three times, followed by absolute alcohol
45 93 and acetone. The NIONs were finally vacuum dried at $\sim 40^\circ C$ for a period of ~ 12 h prior to further investigations. A schematic representation of
46 94 their synthesis and subsequent optimization using chemometric tools is shown in Fig. 1.



Solvothermal synthesis of SB and RH-NION variants



Influence of parametric interactions on As adsorption using RSM and ANN

Fig. 1. Schematic representation of the synthesis and optimization of SB and RH-NIONs for the removal of As.

2.4. Experimental design

2.4.1. RSM Modeling

Response surface methodology (RSM) is defined as a statistical tool that simultaneously optimizes levels of independent variables to achieve the best system performance [29]. RSM was employed to study the influence of univariate and multi-variate interactions between the adsorption parameters on As removal efficiency and determine the optimal conditions. In this context, adsorbate dosage, adsorbent dosage, pH, and contact time were chosen as the variables of concern. Central Composite Design (CCD) was adopted to produce a minimum number of variations for the aforementioned four factors [30]. A total of 30 experiments were designed for the quadratic model and artificial neural network study, involving four factors at five levels (Table S6). The adsorption data were modeled using the following second-order polynomial function to express the correlation between the predictive response and the experimental factors [31]:

$$Y = b_0 + \sum_{i=1}^n b_i X_i + \sum_{i=1}^n b_{ii} X_i^2 + \sum_{i=1}^n \sum_{i>j}^n b_{ij} X_i X_j$$

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4 107 Where Y is the predictive response, b_0 is the model constant, b_i is the linear coefficient, b_{ii} is the interaction coefficients, b_{ij} is the quadratic
5 108 coefficient, and X_i and X_j are the coded values and n is the total number of variables.

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8 109 In this study, the Design-Expert 7.0 statistical software package (Stat-Ease Inc. USA) was used for the regression analysis, graphical analysis, and
9 110 analysis of variation (ANOVA).

11 111 2.4.2. ANN Modelling

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14 112 ANN is a parallel distribution processing system composed of neurons and weights and is based on the principle of a highly interconnected system
15 113 of simple processing elements capable of learning complex interrelations between dependent and independent variables. The network contains an
16 114 input layer, a hidden layer, and an output layer (Fig.1). The most popular ANN is Multilayer Perception (MLP) based ANN which consists of
17 115 comparing responses of output units to desired responses through an iterative process where an error term is calculated to readjust the weights in
18 116 the network so that the network responses are obtained close to the desired responses. In this study, the findings from the RSM studies were
19 117 validated with MLP based ANN study using the MATLAB R2013a computing suite (MathWorks USA) [32, 33]. The performance of RSM and
20 118 ANN models was compared using several statistical parameters viz. coefficient of determination (R^2), mean squared error (MSE), root mean square
21 119 error (RMSE), mean absolute deviation (MAD) and mean absolute percent error (MAPE) (Table S2).

22 120 2.5. Adsorption isotherms and kinetics

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25 121 The adsorption capacity was determined against standard curves of total As using an Atomic Emission Spectrophotometer (Agilent 4200 Microwave
26 122 Plasma-Atomic Emission Spectrophotometer, USA). The adsorption of As by the NIONs was modeled using pseudo-first-order (PFO) and pseudo-
27 123 second-order (PSO) kinetic equations (Table S3) using the OriginPro 8[®] graphing suite. Further, to gain an insight into the underlying mechanism
28 124 of As removal by the NIONs, the quantitative adsorption of As by the optimal SB and RH-NION solvothermal variants were measured as a function
29 125 of non-linear Langmuir, Freundlich, and Sips isotherm equations [34]. In addition, the adsorption capacity of the NIONs was determined using the
30 126 equation [35, 36]:

$$31 \quad q_e = \frac{(C_0 - C_e)V}{M}$$

32 127 Where, q_e is the equilibrium adsorption (mgg^{-1}) C_0 and C_e are the initial and equilibrium As concentration in solution (ppb), respectively. V and M
33 128 are the volumes of studied solution (ml) and the mass of NIONs for each solution (mg), respectively.

34 129 2.6. Analysis and characterization

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38 131 The morphology and particle size of the synthesized NIONs were studied using Transmission Electron Microscopy (TEM) (JEOL JEM-2100,
39 132 Japan) and their average dimensions were determined using the Image J (NIH) processing suite. The magnetic properties of the NIONs were
40 133 characterized using a Dynacool Model-9 physical property measurement system (PPMS, Quantum Design, USA). An X-ray powder diffractometer
41 134 (RIGAKU-Miniflex Benchtop, Japan) was used to analyze the crystalline structures of the adsorbents. Fourier-transform Infrared (FTIR) spectra
42 135 of the two adsorbents before and after adsorption were recorded on a Perkin Elmer-Spectrum 100 Optica FT-IR Spectrometer, USA. To determine
43 136 the amount of iron oxide impregnated within these nanocellulosic matrices, the adsorbents were digested in the presence of a strong acidic medium

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4 137 and the Fe (%) content was estimated using Atomic Emission Spectroscopy (Agilent Technologies-Agilent 4200 MP-AES, USA) against standard
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6 138 curves of Fe. The molecular weight of cellulose and the degree of polymerization (DOP) of cellulose within the nanocellulosic NION matrices was
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8 139 estimated using Photon Correlation Spectroscopy (Malvern P analytical-Zetasizer Nano ZS-90, UK). To detect the binding energies and atomic
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10 140 ratio of adsorbents surface, X-ray photoelectron spectroscopic (XPS, Thermo Fisher Scientific Pvt. Ltd., ESCALAB Xi+, UK) analysis was carried
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12 141 out.

13 142 3. Results and discussion

14 143 3.1. Characterization of the NIONs

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17 144 The TEM analysis provided insights on the topological features of the SB and RH-NIONs (**Fig. 2a-f**). At low magnifications, the SB and RH-
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19 145 NIONs show the presence of an irregular matrix suggesting the presence of nanocellulose fibers (**Fig. 2a, d**). This is attributed to the fibrillation of
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21 146 the milled SB and RH during the high-temperature solvothermal synthesis of NIONs [37]. At higher magnifications, the growth of quasi-spherical
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23 147 iron oxide nanoparticles within these fibrillated SB and RH nanocellulose matrices is clearly observed in both the NIONs (**Fig. 2b, e**). As estimated
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25 148 from the TEM micrographs, the average radius of these iron oxide nanoparticles was found to be ~4.8 nm for the SB-NIONs and ~7.6 nm for the
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27 149 RH-NIONs respectively. The iron oxide particles also show the presence of lattice fringes, which suggests the crystalline nanostructure of Fe₃O₄
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29 150 (**Fig. 2b inset, e inset**). This is further reflected in the SAED patterns of the NIONs (**Fig. 2c, f**), which show the characteristic diffraction signatures
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31 151 for Fe₃O₄, corresponding to the 220, 311, and 440 diffraction planes. This is also indicative of the inverse spinel Fe₃O₄ phase [38].

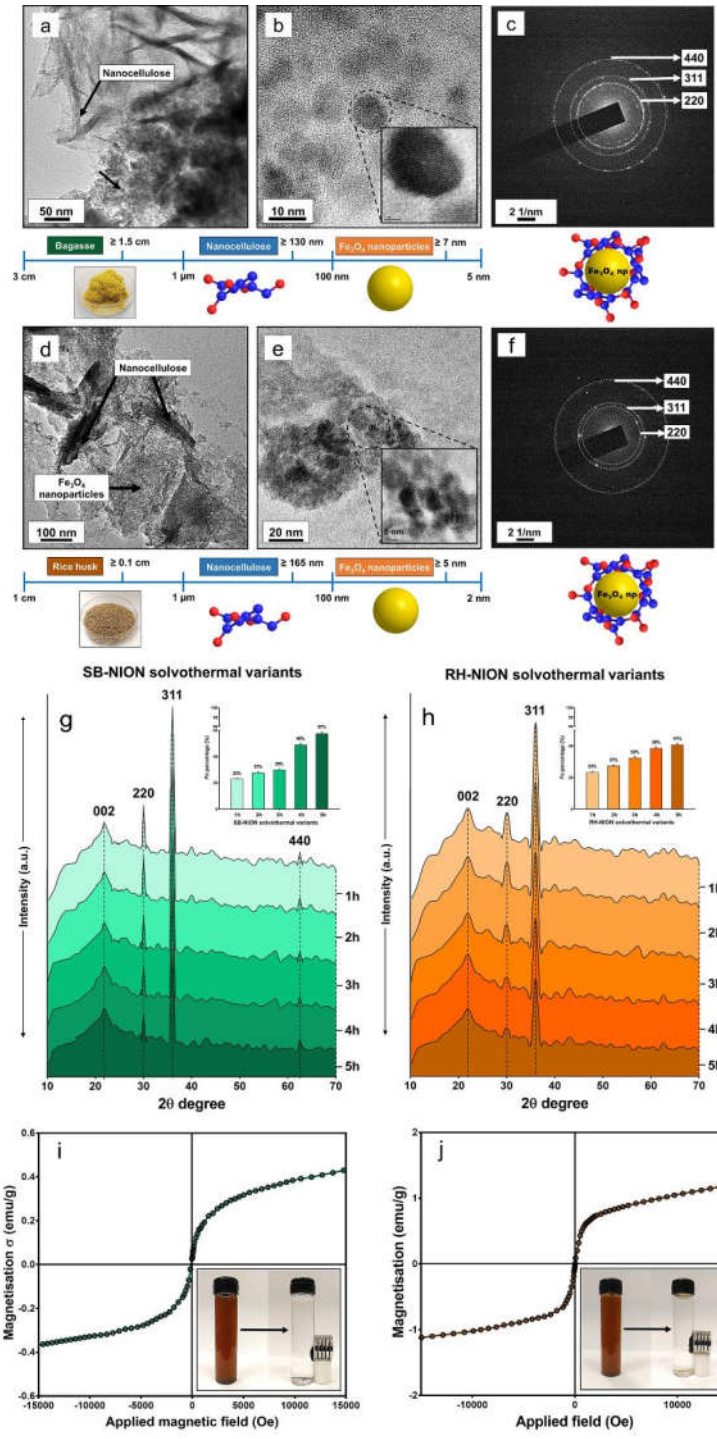
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33 152 The XRD analysis confirms the presence of inverse spinel cubic-phase Fe₃O₄ in both the NIONs. All the XRD spectra of SB and RH-NIONs (**Fig.**
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35 153 **2g, h**) display three well-defined peaks at $2\theta = 23^\circ$, 30° and 35° which corresponds to the crystal planes (002) for crystalline nanocellulose and
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37 154 (220) and (311) for inverse spinel cubic-phase Fe₃O₄ (JCPDS No: 19-6029) [38, 39, 40]. A minor peak at $2\theta = 65^\circ$ was also observed in the XRD
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39 155 spectra of the SB-NIONs, which is attributed to the 440 diffraction plane of cubic phase Fe₃O₄ [40]. Furthermore, characteristic infrared signatures
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41 156 relating to nanocelluloses and Fe₃O₄ bound nanocelluloses were also observed in the FT-IR spectra of the SB and RH-NIONs which are discussed
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43 157 in detail in Section 3.5.

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45 158 The percentage Fe composition in the SB and RH-NIONs was determined to recognize a possible correlation between the progressive increase in
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47 159 solvothermal residence time and the impregnation of the Fe₃O₄ nanoparticles into the SB and RH nanocellulosic matrices. The spectroscopic
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49 160 measurements of the NION variants reveal a progressive increase in the Fe content from an initial 23% for the 1h SB and RH-NION up to 57% and
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51 161 41% for the 5h SB-NION and 5h RH-NION respectively (**Fig. 2g inset, h inset**). Thus, the degree of Fe₃O₄ impregnation in the nanocellulosic
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53 162 matrix progressively increased, as the solvothermal residence of the NIONs increased from 1h to 5h. Further, estimations of the cellulosic degree
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55 163 of polymerization of SB and RH-NIONs were carried out based on their molecular weight and found to be 1900 ± 20 DP and 1650 ± 25 DP
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57 164 respectively. This high cellulosic degree of polymerization suggests that the SB and RH-NIONs matrices are rich in long-chain nanocellulosic
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59 165 structures consisting of numerous surface hydroxyl (-OH) groups. The availability of these surface -OH groups is likely to enhance the As
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61 166 remediation properties of the NIONs because of their affinity towards dissolved As [41]. This also confirms the high structural stability of the
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63 167 NIONs under various solvent conditions.

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65 168 The presence of magnetic characteristics of the synthesized SB and RH-NIONs was measured using a physical property measurement system
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67 169 (PPMS). The results showed that the remnant magnetization and coercive field values of specific magnetization curves for both the NIONs were
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69 170 negligible which confirms their superparamagnetic nature [42]. The magnetization curve of the SB-NION saturated at ~ 0.5 emug⁻¹ while that of the

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171 RH-NION saturated at $\sim 1 \text{ emu g}^{-1}$ (Fig. 2i, j). The superparamagnetic character is a valuable attribute for facilitating the magnetic recovery of
 172 NIONs from water samples to enable its recyclability and reuse. To experimentally validate this attribute, the superparamagnetic SB and RH-
 173 NIONs were recovered under the influence of an external magnetic field. It was observed that the well-dispersed SB and RH-NIONs (Fig. 2i inset,
 174 j inset) rapidly coalesce and accumulate in the presence of a strong magnetic field (214 Gauss).



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Fig. 2. TEM micrographs of the SB-NION (a, b) and RH-NION (d, e) with scale diagrams of precursor materials, derived nanocellulose and Fe₃O₄ NPs, selected area electron diffraction (SAED) patterns of the SB-NION (c) and RH-NION (f); comparative X-ray diffractograms of the synthesized SB-NION solvothermal variants (g) and the RH-NION solvothermal variants (h); Fe content (%) of the synthesized SB-NIONs (g inset) and RH-NIONs (h inset); magnetic hysteresis (M-H) curve for the synthesized SB-NION (i) and RH-NION (j) and magnetic recovery of SB-NION (i inset) and RH-NION (j inset) in As solution under the influence of external magnetic field (214 Gauss).

3.2. Modelling and optimization using RSM

In our study, the experimental parameters considered were adsorbate dosage, adsorbent dosage, pH and contact time within defined boundary levels with 95% confidence limit (**Table S1**). Thus, a five level-four factor (5⁴) full factorial CCD, leading to a total of 30 experimental runs was employed for optimization of SB-NION and RH-NION. The results obtained for each combination are given in **Table S6**. The empirical relations between the four variables and the removal efficiency of As by the two adsorbents are described by the equations as follows:

$$\text{As removal efficiency of RH-NION (\%)} = +101.54093 - 0.90081 \times A + 0.12301 \times B + 0.33039 \times C - 0.27567 \times D + 0.021798 \times A \times B + 0.041568 \times A \times D - 0.028235 \times B \times C - 0.018187 \times B \times D - 0.071884 \times C \times D + 0.010394 \times A^2 - 0.011267 \times B^2 + 0.036321 \times C^2 - 7.93116E-003 \times D^2 + 2.52070E-003 \times B \times C \times D - 3.02532E-004 \times A^2 \times B - 3.76187E-004 \times A^2 \times D + 1.95595E-004 \times A \times B^2$$

$$\text{As removal efficiency of SB-NION (\%)} = +79.95382 + 0.50252 \times A + 0.084804 \times B - 0.34624 \times C - 0.20229 \times D - 2.88335E-004 \times A \times B - 4.02971E-003 \times A \times C + 1.52974E-003 \times A \times D + 3.81898E-003 \times B \times C - 1.03299E-003 \times B \times D - 3.49242E-003 \times C \times D - 3.19711E-003 \times A^2 - 6.15267E-004 \times B^2 + 0.038551 \times C^2 + 5.91028E-003 \times D^2$$

Where, A = Adsorbate dosage (ppb), B = Adsorbent dosage (gL⁻¹), C = Solvent pH and D = contact time (h) with the adsorbent.

The predicted results showed good agreement with the experimental data. Analysis of the variance (ANOVA) was used to assess the statistical significance, the adequacy of the fit, and the competency of the predicted models (**Table S4**). Variables with a p-value less than 0.05 have a significant effect on the response and show the accuracy of the model. However, for higher accuracy, the p-value for lack of fit should be higher than 0.05. **Table S4** shows that the lack of fit values for both SB and RH-NIONs are higher than 0.05 which confirms the accuracy of the predicted models. **Table S4** also presents the value of R² for both SB and RH-NIONs and were found to be 0.96 and 0.97, which suggests that 96% and 97% of the total variations in the experimental parameters could be elucidated by the RSM predictive models. The Pareto plots illustrated in **Fig. 3a, b** explains the impact of the variables on the response. The Pareto chart for SB-NIONs (**Fig. 3a**) indicates that the influence of univariate deviations in the process parameters was found to be the most significant on the adsorption of As which also corroborates with results obtained from **Table S4**. On the other hand, for RH-NIONs the RSM investigations demonstrate univariate, bi-variate, and multi-variate interactions between the process parameters to have a significant influence on the adsorption of As. Thus, in comparison with the RH-NIONs, the SB-NIONs were found to be more resistant to changes in parametric variability, suggesting their better suitability as an effective adsorbent for dissolved As under various parametric conditions.

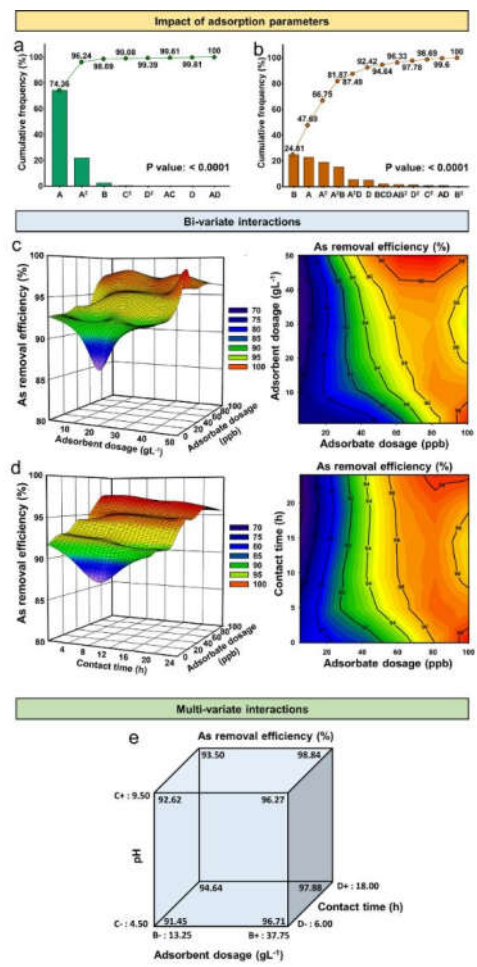
Fig. 3c and **3d** demonstrate the response surface 3D plots for the effect of bi-variate interactions of RH-NIONs for As removal. As evident from the response surface plot (**Fig.3c**) the interactions between adsorbent dosage and adsorbate dosage was found to have a significant influence on the adsorption of As and a maximum As removal efficiency of ~97% was observed with adsorbate dosage of ~76 ppb and an adsorbent dosage of ~38 gL⁻¹. Interactions between adsorbate dosage and contact time were also found to be significant where a maximum As removal efficiency of ~96%

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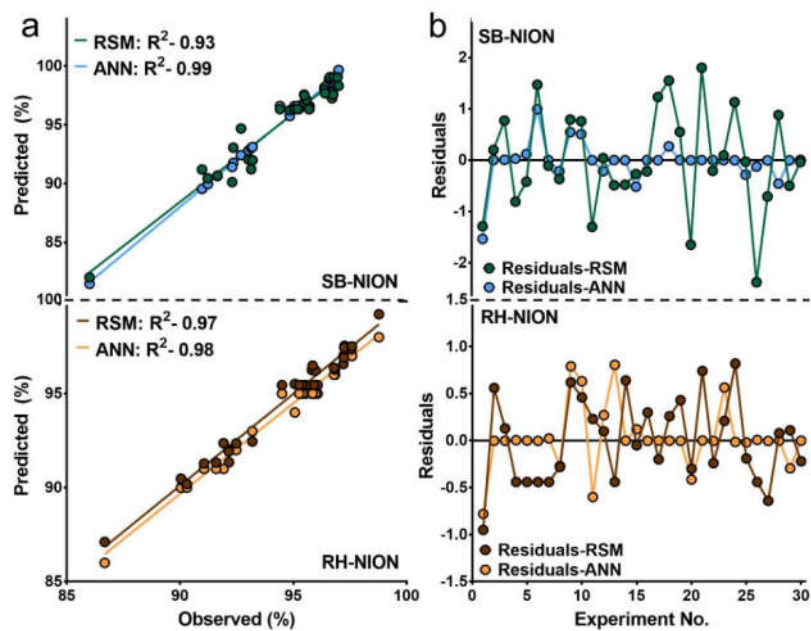
209 was observed, using a ~76 ppb adsorbate dosage for a contact time of ~18h (Fig. 3d). Furthermore, the RSM investigations for RH-NIONs also
 210 showed multi-variate interactions between adsorbent dosage, contact time and pH to have a significant influence on the adsorption of As (Fig. 3e).
 211 A maximum As removal efficiency of 99% was observed for this multi-variate interaction using an adsorbent dosage of ~38 gL⁻¹ for a contact time
 212 of ~18h at pH ~10.

213 **3.3. Modeling and optimization using ANN**

214 To validate the findings of RSM studies with MLP base AAN study the experimental data was divided into three portions for the purpose of training
 215 (70%), testing (15%), and validation (15%) to keep the network away from overtraining and over parameterization [43]. The parameters adsorbate
 216 dosage, adsorbent dosage, pH, and contact time were taken as input together at a time and the As removal efficiency was set as output values from the
 217 experimental results. The number of neurons in the input layer (4) was assigned based on the number of input parameters while the number of neurons
 218 in the output layer (1) was assigned based on the desired output. The network was optimized by varying the number of neurons in the hidden neuronal
 219 layer in the range 5-100 where a minimum of 10 training runs was carried out for each ANN topology [44]. The rigorous iteration was performed in
 220 each case to obtain an optimized combination of training. The performance indicators were selected as mean squared error (MSE) values and
 221 determination coefficients (R²). The ANN topologies of 4-15-1 and 4-35-1 showed the lowest MSE values of 0.003 and 0.03 respectively and were
 222 used for the predictive modeling of



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4 **224** Fig. 3. Pareto chart showing the impact of various experimental parameters in the adsorption of As by the 5h SB-NION (a) and RH-NION (b);
5 **225** effect of bi-variate interactions on the adsorption of As by the 5h RH-NION (c, d); cubic plot showing the effect of multi-variate interactions on
6 **226** the adsorption of As by the 5h RH-NION (e) as described by the CCD matrices of RSM.
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9 **227** As adsorption using SB and RH-NIONs (Fig. S2 a, b; Table S5). Various error functions were used to assess the accuracy of the model which are
10 **228** given in Table S2. The determination coefficients for both of the ANN topologies were found to be 0.99, which suggests that 99% of the total
11 **229** variations in the experimental parameters could be elucidated by these ANN predictive models (Table S6).
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14 **230** **3.4. Comparison of the predictive chemometric models**

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16 **231** The predictive outputs of the RSM and ANN-based chemometric models of SB and RH-NIONs were compared to that of the experimental results
17 **232** and are depicted in Table S6. Their performance was evaluated based on the linear correlation plots and the distribution of residuals is shown in
18 **233** Fig. 4. The high R^2 values (>0.9) for the linear correlation plots suggest that both RSM and ANN chemometric models were in good agreement
19 **234** with that of the experimental results (Table S7, Fig. 5a). This confirms the suitability of both RSM and ANN predictive models for the optimization
20 **235** of As adsorption. In addition, the better R^2 values of the ANN models (Fig. 4a) and the localization of their residual values, close towards the
21 **236** central line (Fig. 4b) suggest that the predictions of the ANN models were closer to the experimental values in comparison to the RSM models.
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23 **237** This is further supported by the MSE, RMSE, MAD, and MAPE based statistical values measured for both RSM and ANN models, where the
24 **238** values are significantly better for ANN models in comparison to RSM (Table S6). The results thus confirm the superior nature of the ANN models
25 **239** over RSM for predictive optimization of experimental parameters and are consistent with similar studies carried out earlier. However, ANN-based
26 **240** models lack the ability to describe the impact of experimental parameters on the experimental output based on relative significance/insignificance
27 **241** using mathematical frameworks, which is a remarkable attribute of RSM-based predictive models [33]. Thus, in the current study, the combined
28 **242** use of RSM and ANN predictive models was found to be crucial in understanding the significant impact of the experimental parameters on the
29 **243** adsorption of As with improved approximation and generalization capabilities for predictive assessments.
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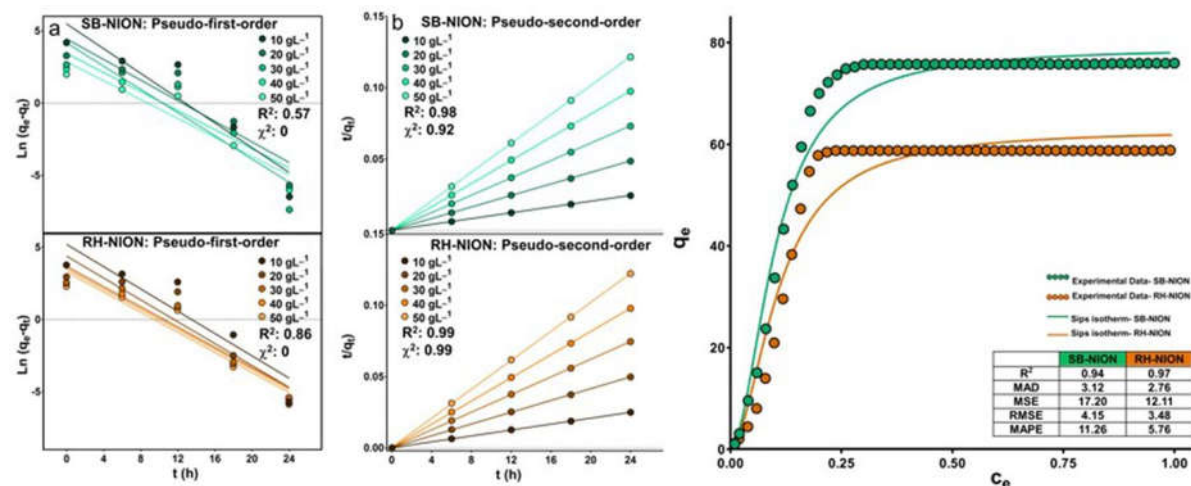
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4 **246** Fig. 4. Comparison of predictive outputs of RSM and ANN-based chemometric models based on determination coefficient (a) and distribution of
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6 **247** residuals (b) for the adsorption of As by the 5h SB-NION and 5h RH-NION.
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8 **248**

9 **249 3.5. Study of adsorption isotherms, adsorption kinetics, and mechanism of adsorption**

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11 **250** The rate-limiting step behind the adsorption of As by NIONs under the optimal parametric conditions was determined by the adsorption equilibrium
12 **251** data. The adsorption equilibrium data were fitted using linear forms of two conventional kinetic models *viz.* Lagergren's pseudo-first-order model
13 **252** and Ho's pseudo-second-order model (Fig. 5a, b). Based on the R^2 values and χ^2 measures of precision, the pseudo-second-order plots from Fig.
14 **253** 5b indicates that the adsorption of As by the SB and RH-NIONs follow pseudo-second-order kinetics. This is further confirmed from the calculated
15 **254** adsorption capacities (q_e calc) of the pseudo-second-order plots for the NIONs, which were in good agreement with that of the experimental (q_e)
16 **255** adsorption capacities (Table S7). The rate-limiting step for the adsorptive process governed by pseudo-second-order kinetics has primarily been
17 **256** attributed to a chemisorptive mechanism involving covalent binding and/or ionic interactions between the adsorbent and the adsorbate [45, 46].

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23 **257** Langmuir, Freundlich, and Sips isotherms were also used to investigate the adsorption of As by SB and RH-NIONs under the optimal experimental
24 **258** conditions. The Sips adsorption isotherm describes best the As adsorption processes for both SB and RH- NIONs which is further confirmed by
25 **259** the R^2 values (0.97-0.94) together with the derived MAD, MSE, RMSE and MAPE values (Fig. 5c, Fig. S3). The Sips heterogeneity factor (n_s)
26 **260** which denotes the nature of the adsorption process, was derived from the non-linear plots and was found to be 1.9 and 1.8 for SB and RH-NIONs
27 **261** respectively (Fig. 5c, Fig. S3). This along with the findings of the kinetics studies suggests the adsorption of As to be mediated through covalent
28 **262** interactions between the NIONs and the dissolved As. Therefore, the findings confirm chemisorption to be a dominant mechanism governing the
29 **263** adsorption of dissolved As by the NIONs [46]. The maximum As adsorption capacities of the NIONs were also derived from the fitted isotherms
30 **264** and found to be >550 mgg^{-1} for SB-NIONs which was observed to be appreciably better than that of RH-NIONs (~ 100 mgg^{-1}) under optimal
31 **265** adsorption conditions.
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58 **268** Fig. 5. Pseudo-first-order (a), pseudo-second-order (b) adsorption kinetics curves, and non-linear fits of Sips adsorption isotherm plot for the
59 **269** adsorption of As by the SB-NION and the RH-NION (c).
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To further gain insight into the mechanistic interaction between the dissolved As and the active sorption sites in the adsorbents, the SB-NION was assessed using X-ray photoelectron spectroscopy (XPS) and Fourier transform infrared spectroscopy (FT-IR), both prior to and after As adsorption under optimal conditions. The FT-IR spectroscopic studies of both SB-nanocellulose and SB-NION composite shows the presence of several bonds formed by oxygen functional groups. These include a broad absorption band at 3415 cm^{-1} region (Fig. 6a) corresponding to the O-H stretching vibration modes for nanocelluloses. The band at 2903 cm^{-1} can be attributed to the aliphatic C-H stretching vibration modes of nanocellulose and those bands at 1654 cm^{-1} and 1115 cm^{-1} may be attributed to O-H bending and C-O stretching vibrations of nanocellulose [47, 39]. The peaks at 611 cm^{-1} and 422 cm^{-1} in the spectrum of SB-NION prior to As adsorption corresponds to C-O stretching vibration and Fe-O bending vibration indicating that Fe_3O_4 nanoparticles were successfully attached to the nanocellulose surfaces. The spectrum of SB-NION after As adsorption shows that the bands at $3415, 2903, 1654,$ and 1115 cm^{-1} have decreased in the intensity which is the outcome of the complexation of the oxygen functional groups with As (III) ions. In addition, the oxygen atoms present in hydroxyl groups possess free pair of electrons that are likely to interact with the empty orbital of As (III) and form complexes through coordinative bonds [41, 48]. On the other hand, the decreased in intensity of the peaks at 611 and 422 cm^{-1} is indicative of the prominent interaction of C-O and Fe-O bonds with As through bidentate binuclear corner-sharing (2C) complexes between As(III) pyramids and adjacent edge-sharing FeO_6 octahedra at alternate vertices [40, 49, 50] (Fig.6a).

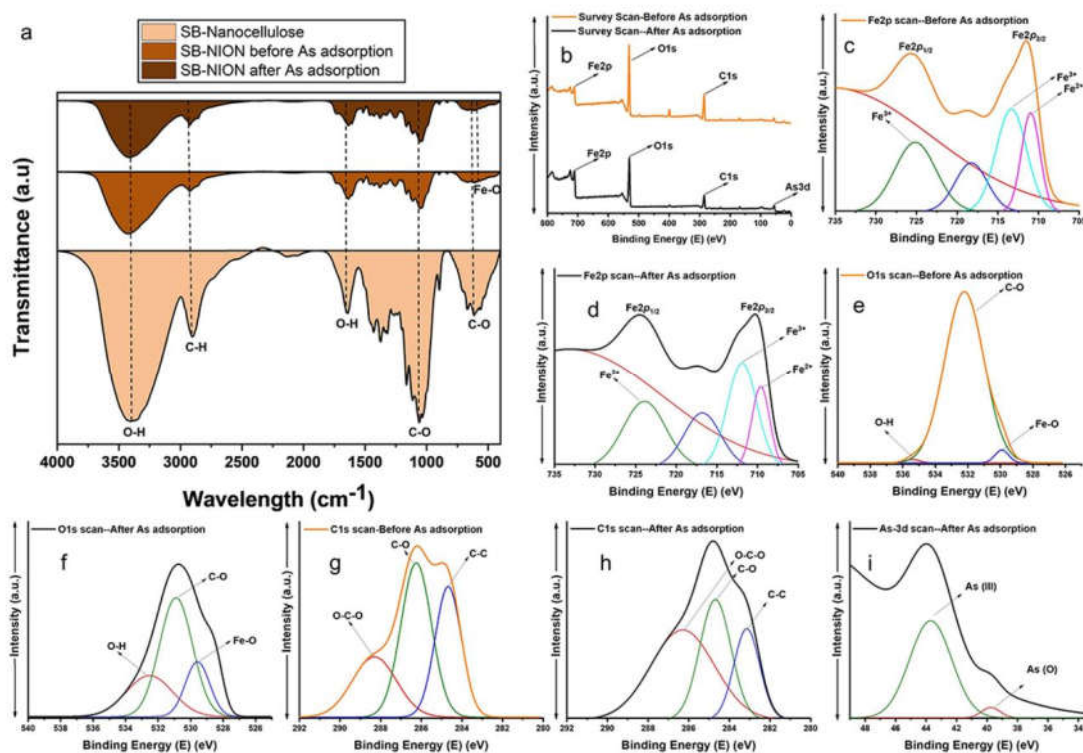
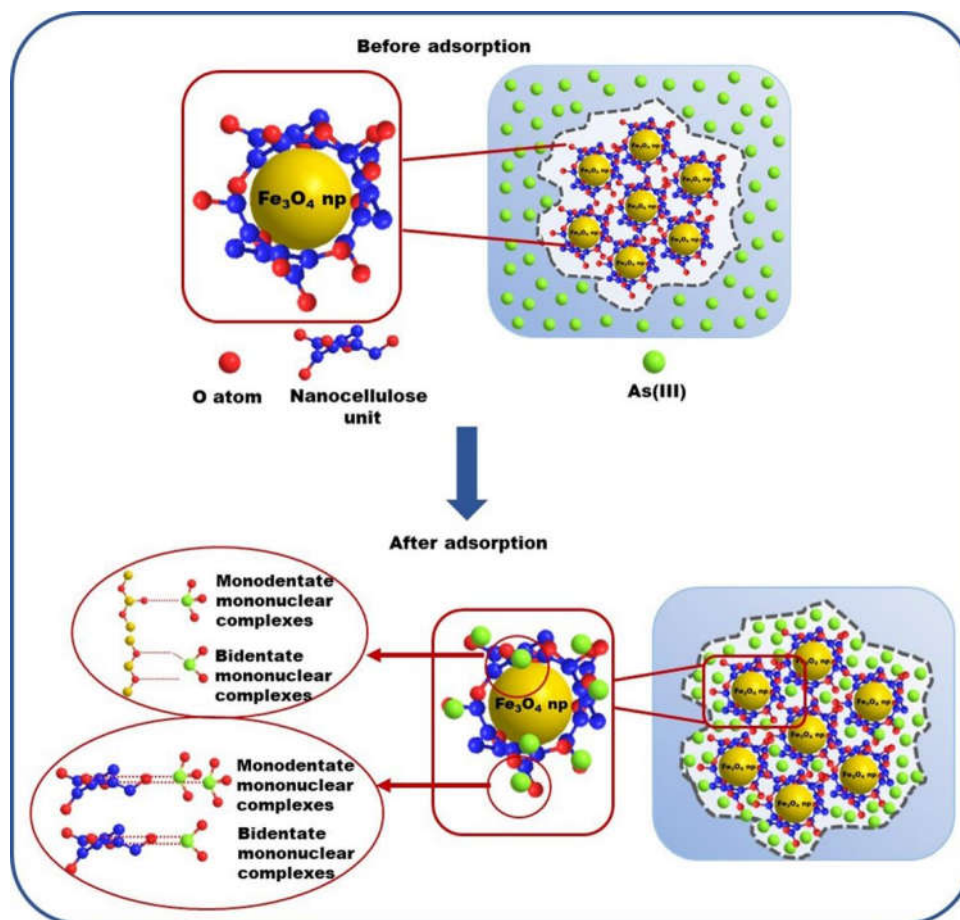


Fig. 6. Comparative FTIR spectra of the SB- nanocellulose, SB-NION before adsorption and SB-NION after adsorption (a); XPS-survey scan of the SB-NION before and after As adsorption (b); Fe (2p) XPS spectra (c, d); O (1s) XPS spectra (e, f); C (1s) XPS spectra (g, h); before and after As adsorption and As (3d) XPS spectra (i) after the adsorption of As by the SB-NION.

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 4 287 The wide-scan survey XPS spectra of SB-NION (**Fig. 6b**), show prominent photoelectron lines at binding energies of 711.3 eV, 530.5 eV, and 284.0
 5 288 eV both before and after the adsorption of As that may be attributed to Fe(2p), O(1s) and C(1s) respectively. Additionally, an apparent As (III) peak
 6 289 at a binding energy of 44.5 eV is observed in the post-adsorption wide-scan survey spectrum thus indicating that the adsorbent has taken up As (III)
 7 290 efficiently. High-resolution XPS scans for the Fe (2p) photoelectron signature (**Fig. 6c**) further shows the presence of Fe $2p_{1/2}$ and Fe $2p_{3/2}$ photoelectron
 8 291 peaks, which upon de-convolution revealed individual peaks at 725.0 eV (Fe 3^{+}), 711.5 eV (Fe $^{3+}$) and 710.5 eV (Fe $^{2+}$), characteristic of mixed-phase
 9 292 Fe $_3$ O $_4$ nanoparticles [51]. These Fe (2p) signatures appear to shift towards lower binding intensities (**Fig. 6d**), which suggests covalent bonding
 10 293 between Fe $_3$ O $_4$ and As though multi-dentate nuclear complexes [52]. The O 1s peaks appear to consists of three components at 533 eV, 531 eV, and
 11 294 530 eV (**Fig. 6e, f**) that suggest the presence of oxygen functional groups; O-H, C-O, and Fe-O respectively, which is consistent with the IR spectra.
 12 295 Further, the C 1s peak fits of O-C-O (288 eV), C-O (286 eV), and C-C (284 eV) also attribute to the presence of oxygen functional groups that undergo
 13 296 a prominent decrease in intensities after adsorption (**Fig. 6g**). These decrease in intensities are likely to occur due to the complexation of oxygen
 14 297 functional groups of SB-NION with As (III) which was further revealed by the dominance of As (III) peak at a binding energy of 44 eV (**Fig. 6h**).
 15 298 The findings are also in agreement with the trends observed in the FTIR spectroscopic studies, which suggests that the underlying mechanism of As
 16 299 adsorption may be attributed to the formation of nanocellulose: As and Fe $_3$ O $_4$: As multidentate covalent complexes, during the adsorption of As by
 17 300 the adsorbents as shown in **Fig. 7**.



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 302 **Fig. 7.** Probable mechanism for As adsorption by the NIONs, based on XPS and FT-IR spectroscopic studies.

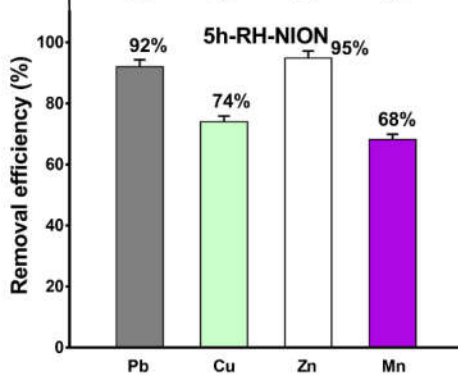
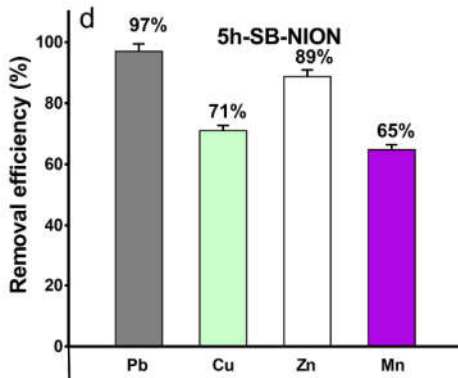
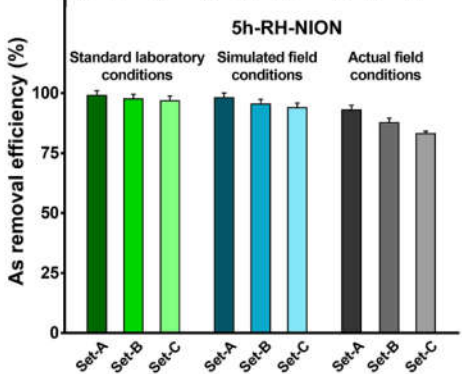
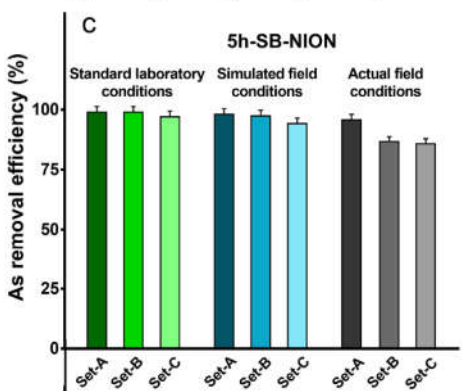
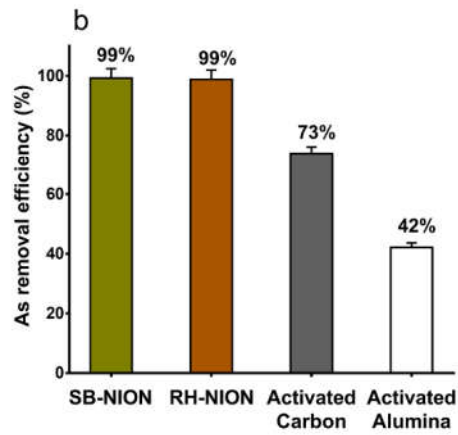
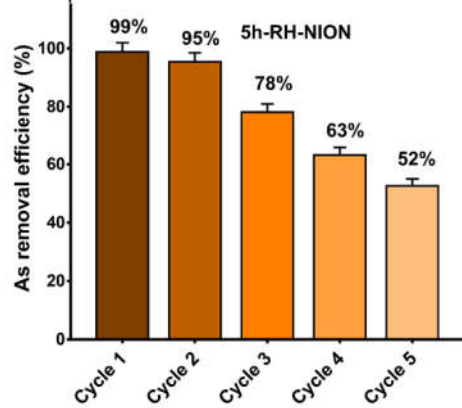
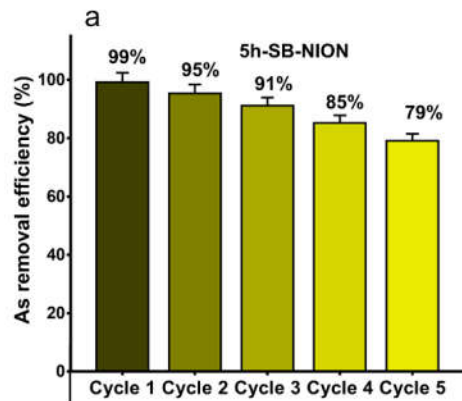
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4 303 **3.6. Assessment of NION performances with contaminated groundwater samples (other heavy metals), recyclability, and**
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6 304 **comparison with commercial adsorbents**
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8 305 The recyclability performance of an adsorbent is a key concern in terms of cost-effectivity and applicability for an actual water treatment system.
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10 306 Hence, the recyclability assessments for both SB and RH-NIONs were performed under the optimal conditions derived from RSM and ANN-based
11 307 chemometric investigations. Desorption experiment was conducted to examine the reusability of the adsorbents. Desorption may occur either by
12 308 thermal treatment or through suitable desorbing agents like acid or alkali solutions. In the present study, however, a chemical method of regeneration
13 309 was employed in a neutral solution. The arsenic-loaded adsorbents obtained after the adsorption process was added into a 50 ml of Phosphate Buffer
14 310 Saline (PBS) solution (pH 7.4) as the desorbing agent. The solution was shaken at 150 rpm and then agitated at specific time intervals for up to 24
15 311 h. The solution was then centrifuged and the desorbed NIONs were re-incubated in 100 mL of 100 mgL⁻¹ As solutions. Adsorption-desorption
16 312 experiments were repeated up to 5 successive cycles under the optimum condition to determine their reusability. Fig. 8a demonstrates that the SB-
17 313 NIONs have shown better results compared to RH-NIONs retaining an appreciable As removal efficiency of ~79% at the end of five successive
18 314 cycles. The adsorption capacity of the individual components viz. native, milled SB and RH powders, SB and RH derived nanocelluloses, and the
19 315 native magnetic Fe₃O₄ nanoparticles formed during NION synthesis were also evaluated (Table S8). The native SB and RH precursors showed the
20 316 lowest As removal efficiencies (4% and 2%, respectively) compared to the milled SB and RH powders (20% and 18%, respectively), while the
21 317 derived nanocelluloses had an efficiency of 43% and 47% respectively (Table S8). Likewise, the native Fe₃O₄ nanoparticles had an As removal
22 318 efficiency of 69% (Table S9). Thus, it can be concluded that the enhanced adsorption capacity of the synthesized NIONs may have resulted from
23 319 the synergistic effect of nanocelluloses and Fe₃O₄ nanoparticles in the synthesized nanocomposite.

24 320 For the comparative assessment of As adsorption efficiency by the NIONs, it was evaluated against commercial adsorbents like activated carbon
25 321 (Sigma Aldrich) and activated alumina (Sigma Aldrich) (Fig. 8b). Under the optimized experimental conditions employing 50ppb As, 52.5 gL⁻¹
26 322 adsorbent, pH:7 and a contact time of 12H, both SB and RH-NIONs showed ~73% and ~42% removal efficiencies of activated carbon and activated
27 323 alumina which is lower than As removal efficiency (~99%).

28 324 To validate the potential of the NIONs for practical applications, the As removal efficiency of the NIONs was also investigated under *in vitro*
29 325 simulated-field conditions and groundwater samples collected from actual contamination sites. Since potable groundwater sources polluted with
30 326 As are reported to suffer from co-contamination with other heavy metals, salts, anions and cations that interfere competitively with As removal, the
31 327 *in vitro* simulated-field conditions were derived using standard solutions of prevailing heavy metals at concentrations recorded in earlier studies
32 328 (Table S9) [53]. Both SB and RH-NIONs show only a minimal decline of 5% to 9% in their As removal efficiencies when contaminated field
33 329 samples were used (Fig. 8c). These findings suggest that the co-contaminants of As that are usually present in naturally contaminated groundwater,
34 330 do not exert a significant influence on the adsorption capabilities of the NIONs. The adsorbents were also evaluated for their remediation abilities
35 331 against the other commonly prevalent groundwater contaminants such as lead (Pb), copper (Cu), zinc (Zn) and manganese (Mn) under *in vitro*
36 332 simulated-field conditions. The RH-NIONs showed a removal efficiency of 92% for Pb, 95% for Zn, 74% for Cu and 68% for Mn, while SB-
37 333 NIONs showed removal efficiencies of 97% for Pb, 89% for Zn, 71% for Cu and 65% for Mn (Fig. 8d).

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Fig. 8. Recyclability of the 5h SB-NION and 5h RH-NION for five successive readsorption cycles (a), comparison assessment of 5h SB-NION and 5h RH-NION of the As removal efficiencies with commercial adsorbents, activated carbon and activated alumina (b), As removal efficiencies of the 5h SB-NION and the 5h RH-NION from standard As laboratory solution, simulated field condition of As in solution with other prevalent metals and actual contaminated groundwater samples (c), removal efficiencies of the 5h SB -NION and the 5h RH -NION for other heavy metals viz. Pb, Cu, Zn, and Mn commonly prevalent in groundwater (d).

4. Conclusions

Nanoscale iron-oxide impregnated agro-waste derived nanocellulose were synthesized, characterized, and analyzed for aqueous As (III) adsorption. The adsorptive performance of the synthesized NIONs were studied using CCD-based RSM models and MLP-based ANN models that have led to the finding of maximum As removal efficiency of ~99% under the conditions of ~52.5 ppb As concentration, ~50 gL⁻¹ adsorbent dosage at pH for a contact time of 12h for both SB and RH-NIONs. Adsorption on both the adsorbents meet Sips isothermal equation following pseudo-second-order kinetics that suggested chemisorption as the underlying mechanism. The high adsorptive capacity together with their economic synthesis indicated the potential of the adsorbents to be used for real-world applications.

Declaration of competing interest

The authors declare that they have no competing interests.

Acknowledgments

The authors wish to acknowledge DBT, Govt. of India, for the Research Grant (Grant No. BT/258/NE/TBP/2011), UGC for the Research Grant (TU/ Fin/MBBT/ 116/ 05/ 11-12/ 64), and DST-FIST. Author Chayanika Chaliha, would like to acknowledge DST, Govt. of India for her DST INSPIRE Junior Research Fellowship (IF-19064). The authors also acknowledge SAIC-Tezpur University, Tezpur, Assam and SAIF-North-Eastern Hill University, Shillong, Meghalaya, for the analytical assistance.

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Modelling and optimization of factors influencing adsorptive performance of agrowaste-derived Nanocellulose/Iron Oxide Nanobiocomposites during remediation of Arsenic contaminated groundwater

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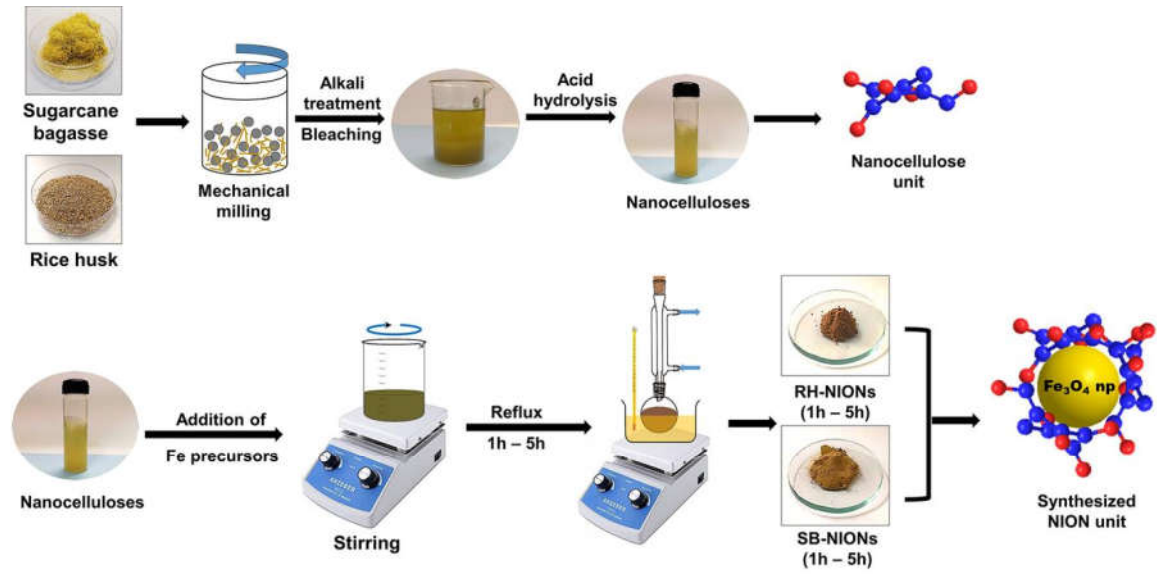
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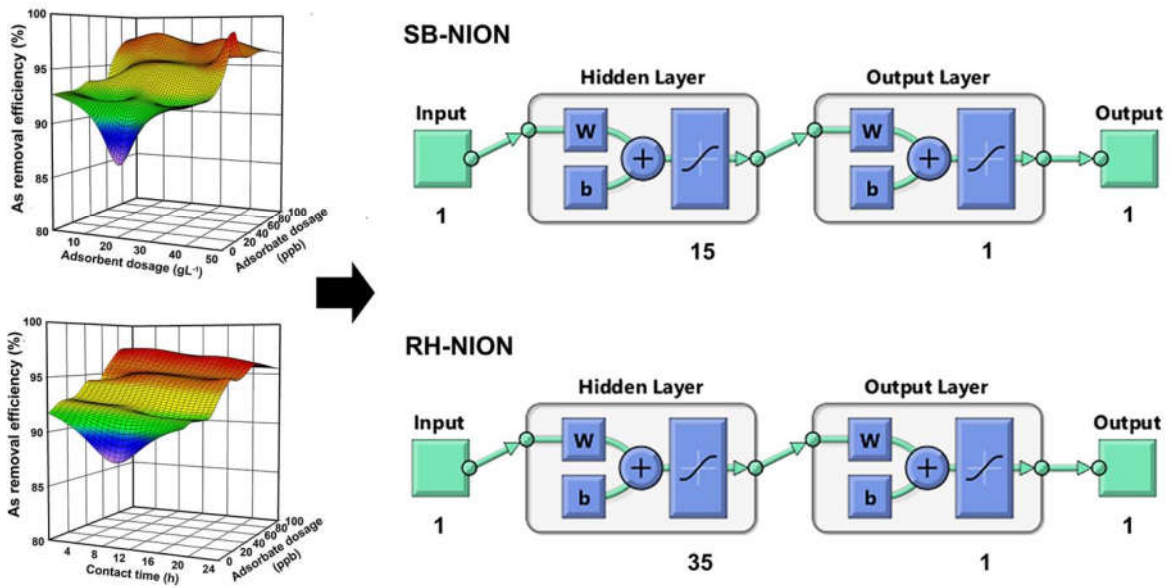
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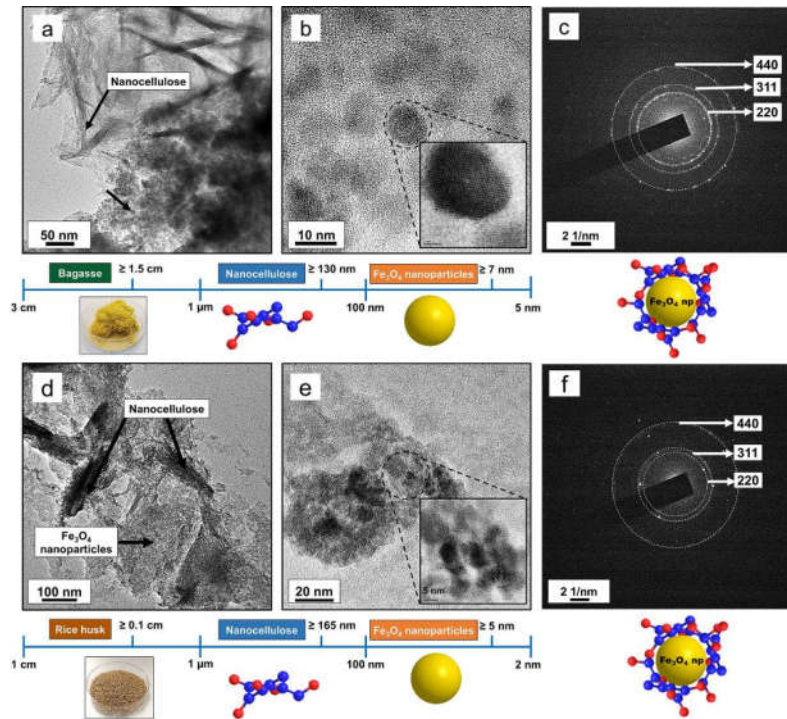


Solvothermal synthesis of SB and RH-NION variants

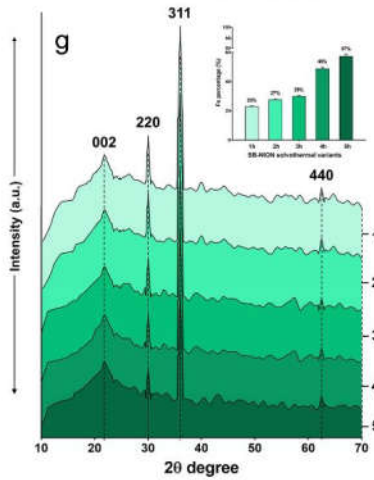


Influence of parametric interactions on As adsorption using RSM and ANN

Fig. 1. Schematic representation of the synthesis and optimization of SB and RH-NIONs for the removal of As.



SB-NION solvothermal variants



RH-NION solvothermal variants

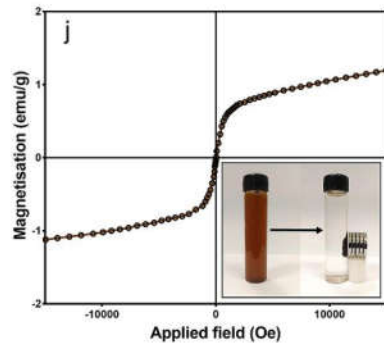
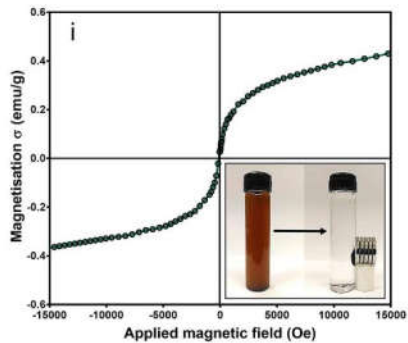
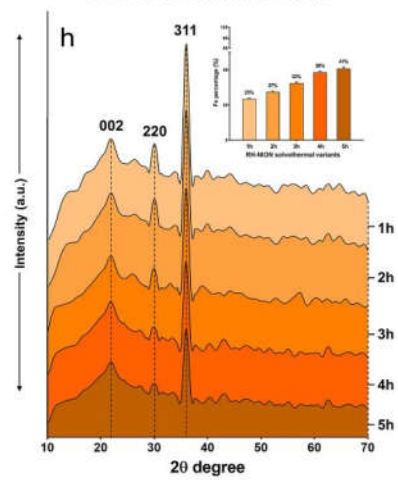


Fig. 2. TEM micrographs of the SB-NION (a, b) and RH-NION (d, e) with scale diagrams of precursor materials, derived nanocellulose and Fe₃O₄ NPs, selected area electron diffraction (SAED) patterns of the SB-NION (c) and RH-NION (f); comparative X-ray diffractograms of the synthesized SB-NION solvothermal variants (g) and the RH-NION solvothermal variants (h); Fe content (%) of the synthesized SB-NIONS (g inset) and RH-NIONS (h inset); magnetic hysteresis (M-H) curve for the synthesized SB-NION (i) and RH-NION (j) and magnetic recovery of SB-NION (i inset) and RH-NION (j inset) in As solution under the influence of external magnetic field (214 Gauss).

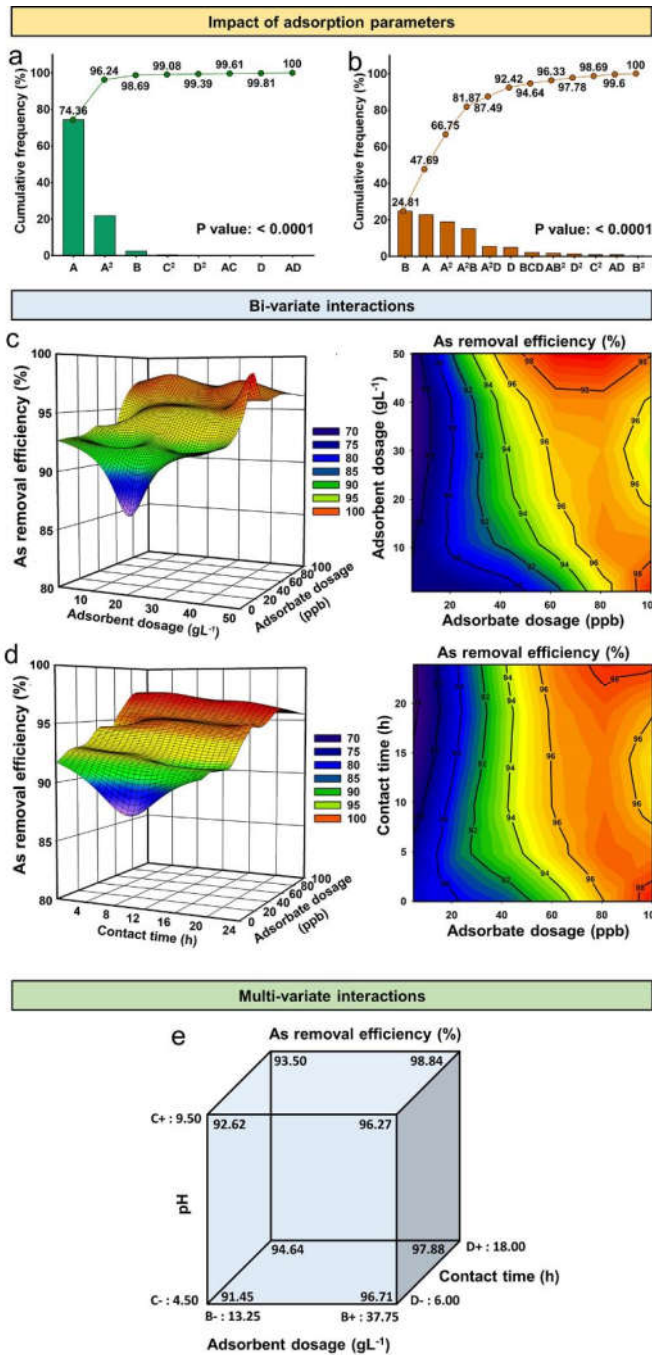


Fig. 3. Pareto chart showing the impact of various experimental parameters in the adsorption of As by the 5h SB-NION (a) and RH-NION (b); effect of bi-variate interactions on the adsorption of As by the 5h RH-NION (c, d); cubic plot showing the effect of multi-variate interactions on the adsorption of As by the 5h RH-NION (e) as described by the CCD matrices of RSM.

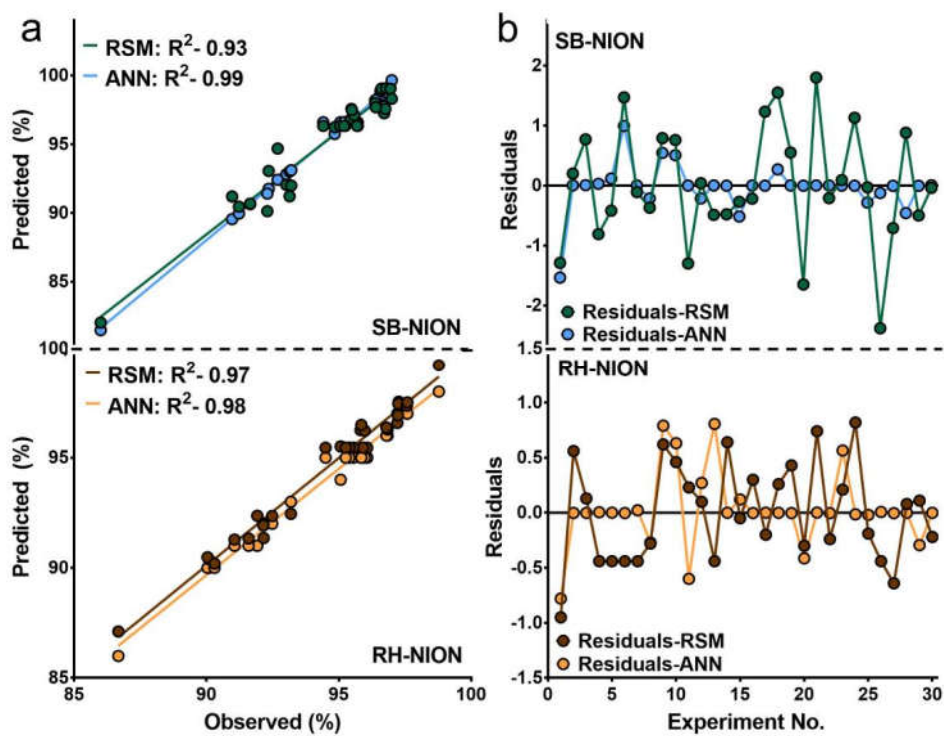


Fig. 4. Comparison of predictive outputs of RSM and ANN-based chemometric models on the basis of determination coefficient (a) and distribution of residuals (b) for the adsorption of As by the 5h SB-NION and 5h RH-NION.

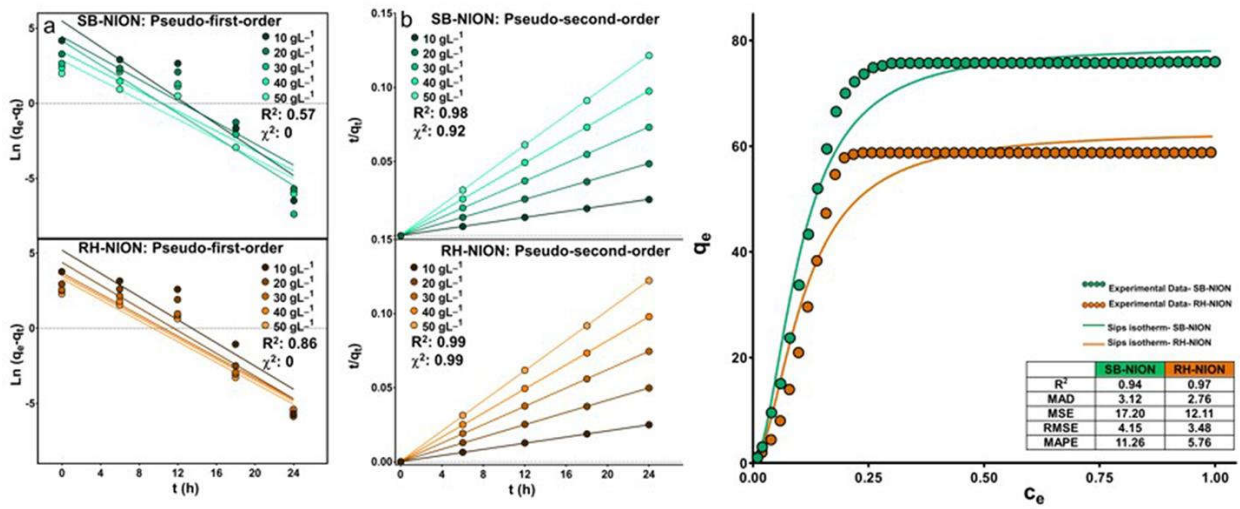


Fig. 5. Pseudo-first-order (a), pseudo-second-order (b) adsorption kinetics curves and non-linear fits of Sips adsorption isotherm plot for the adsorption of As by the SB-NION and the RH-NION (c).

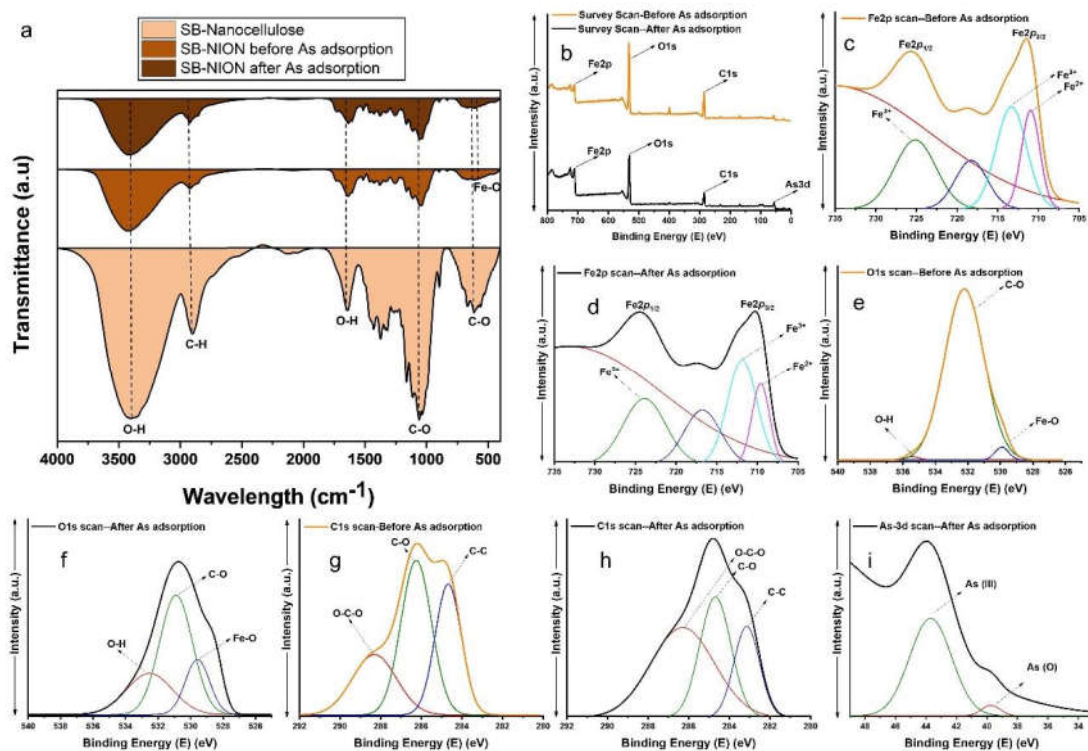


Fig. 6. Comparative FTIR spectra of the SB- nanocellulose, SB-NION before adsorption and SB-NION after adsorption (a); XPS-survey scan of the SB-NION before and after As adsorption (b); Fe (2p) XPS spectra (c, d); O (1s) XPS spectra (e, f); C (1s) XPS spectra (g, h); before and after As adsorption and As (3d) XPS spectra (i) after the adsorption of As by the SB-NION.

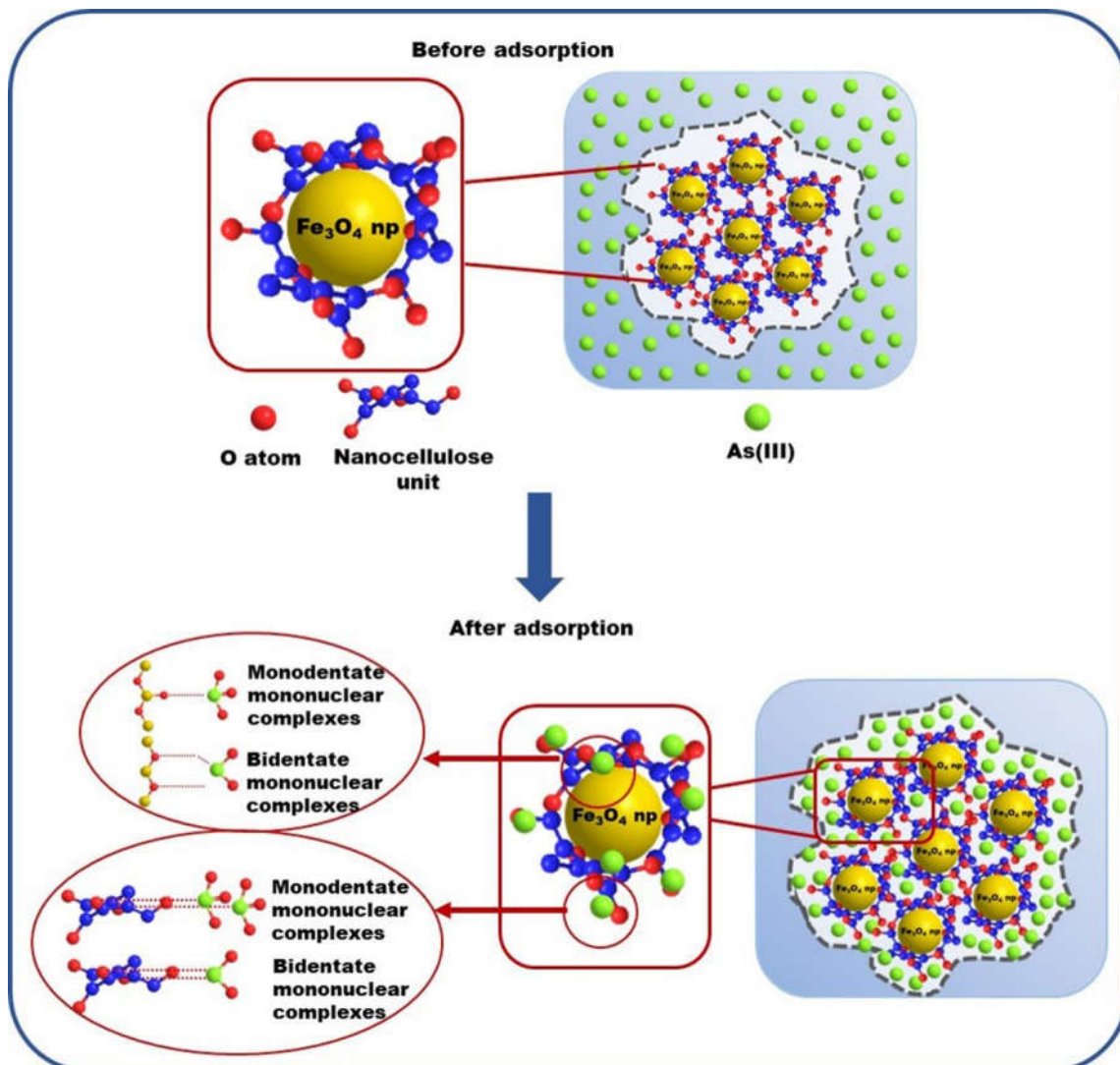


Fig. 7. Probable mechanism for As adsorption by the NIONs, based on XPS and FT-IR spectroscopic studies.

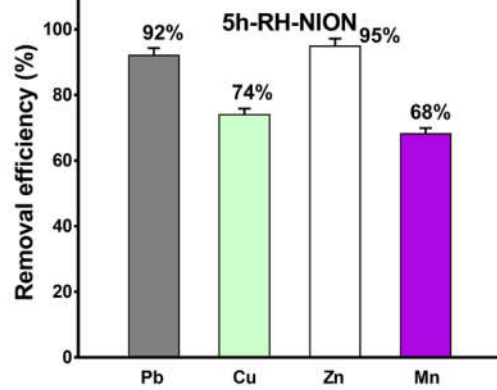
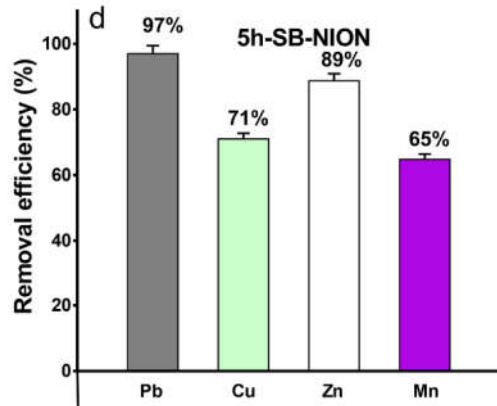
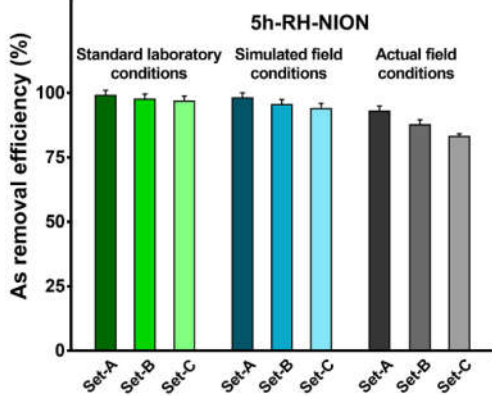
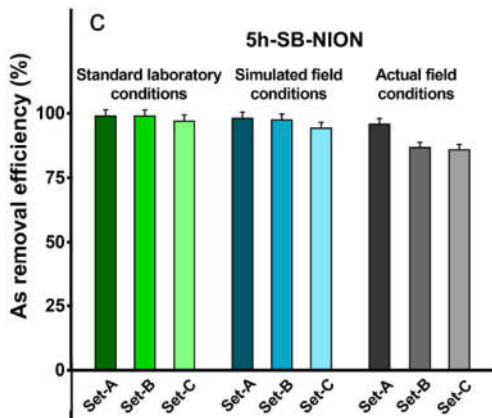
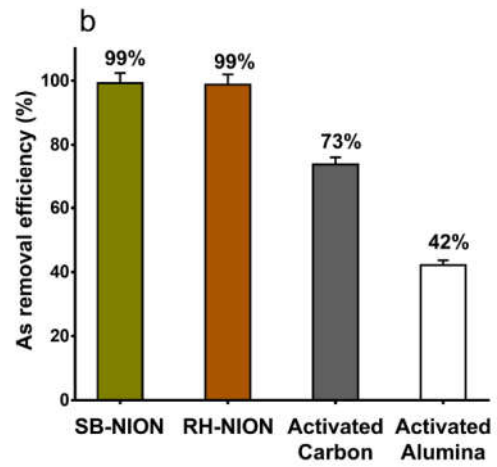
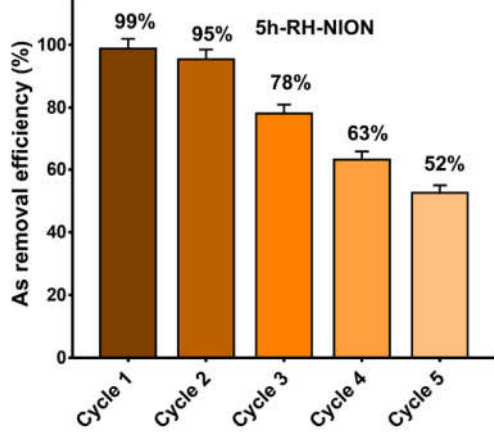
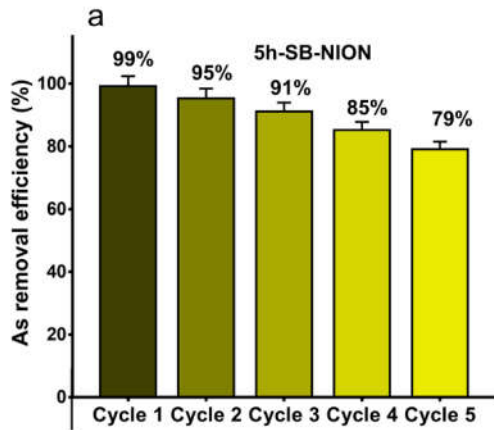


Fig. 8. Recyclability of the 5h SB-NION and 5h RH-NION for five successive readsorption cycles (a), comparison assessment of 5h SB-NION and 5h RH-NION of the As removal efficiencies with commercial adsorbents, activated carbon and activated alumina (b), As removal efficiencies of the 5h SB-NION and the 5h RH-NION from standard As laboratory solution, simulated field condition of As in solution with other prevalent metals and actual contaminated groundwater samples (c), removal efficiencies of the 5h SB -NION and the 5h RH -NION for other heavy metals viz. Pb, Cu, Zn and Mn commonly prevalent in groundwater (d).

Modelling and optimization of factors influencing adsorptive performance of agrowaste-derived Nanocellulose/Iron Oxide Nanobiocomposites during remediation of Arsenic contaminated groundwater

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CRedit authorship contribution statement

Julie Baruah: Investigation, Methodology, Formal analysis, Data curation, Writing - original draft. **Chayanika Chaliha:** Software, Formal Analysis, Data curation. **Eeshan Kalita:** Conceptualization, Methodology, Supervision, Writing - review & editing, Project administration, Funding acquisition. **Bikash Kar Nath:** Investigation, Formal analysis. **Robert A. Field:** Conceptualization, Supervision. **Pritam Deb:** Validation, Resources, Writing - review & editing

