- **1** Community structure and functioning below the streambed across contrasting
- 2 geologies
- 3
- 4 Mark Dunscombe^{1, 2}, Anne Robertson^{1*}, Ignacio Peralta-Maraver¹, Peter Shaw¹
- 5 *corresponding author: <u>a.robertson@roehampton.ac.uk</u>
- ¹Department of Life Sciences, Roehampton University, Holybourne Avenue, London
 SW15 4JD U.K.
- 8
- 9 ²APEM Ltd, A17 Embankment Business Park,
- 10 Heaton Mersey, Stockport, Cheshire, SK4 3GN U.K.
- 11
- 12

13 ABSTRACT

There is little understanding of the variability in the structure and function of metazoan 14 hyporheic communities across streams draining geologies that weather to produce 15 different pore sizes and, by extension, different hydrological conditions. In this study 16 we selected two catchments in each of three geologies that had differing values of 17 hydraulic conductivity and porosity, and sampled four riffles in each catchment at 18 high and low water levels and at two depths. We found clear differences in the 19 physical template of streams draining different geologies and in the composition and 20 abundance of communities inhabiting the hyporheic zones of streams draining chalk/ 21 22 sandstone and limestone geologies. However, we did not detect any significant differences in body size, biomass or functional measures (diversity, richness, 23 redundancy) between the geologies. Our findings imply that ecosystem functioning in 24 25 streams draining geologies that produce fine grain sediments may be similar to those draining geologies that produce coarse grained sediments irrespective of differences 26 in the physical template and community structure. 27

Keywords: Hyporheic zone, hyporheos, functional diversity, biomass, streambed biotope, functional redundancy

32 **1. INTRODUCTION**

33 The central aim of ecological science is to understand the relationship between organisms and their environment, a relationship driven by the habitat template on 34 which evolution forges characteristic life history strategies (Southwood 1977, 1988). 35 Determining the nature of the physical habitat template in the benthic compartment of 36 river systems has a long history (see Townsend & Hildrew 1994). Here we focus on 37 another connected compartment of the same system, the interface between rivers and 38 aquifers, the 'hyporheic zone' (HZ), that hosts a continuous exchange of water, 39 nutrients and materials between the surface stream and the aquifer (e.g., Knapp et al. 40 2017). The complex interactions of these two sources of water produce a unique and 41 dynamic set of physicochemical conditions where there are strong gradients of pH and 42 oxygen (e.g., Krause et al. 2009, Williams et al. 2010, Bardini et al. 2012). These 43 gradients encourage the biogeochemical cycling of nutrients and attenuation of 44 contaminants (e.g., Gandy et al. 2007, Lewandowski et al. 2011, Hester et al. 2013, 45 Stegen et al. 2016). This function is mediated by microbial biofilms (e.g., Brunke & 46 Gonser 1997) that form part of a hyporheic community that also includes small 47 invertebrates (<1 mm in size) such as microcrustaceans, tardigrades, rotifers, small 48 oligochaetes, chironomids and nematodes as well as the early instars of 49 macroinvertebrates (Robertson et al. 2000, Peralta-Maraver et al. submitted). This 50 metazoan hyporheic community (the hyporheos; Orghidan 1959) is the focus of our 51 52 study.

53

.

54

55 The composition, abundance and biomass of metazoan hyporheos, and perhaps their ecological functioning, are strongly influenced by the physical and chemical 56 conditions experienced within the HZ (Robertson & Wood 2010). Sediment grain 57 size is an important factor because of its strong relationship to interstitial pore size, 58 porosity and permeability (Brunke and Gonser, 1997). In turn grain size is partly 59 dependent on the geology from which the sediments have been weathered although 60 other factors, such as topography and land use, also play a role. Easily weathered 61 geologies are associated with fine alluvial sediments whereas geologies resistant to 62 63 weathering are associated with coarse sediments (Morrice et al., 1997). Sediment grain size determines the dimensions of the habitat available to the hyporheos and 64 they are often most abundant in coarse-grained sediment where habitat availability is 65 66 high (Strayer et al. 1997). In fine grained sediments, where pore spaces are reduced, small sized organisms such as the meiofauna are disadvantaged because 67 there is little available habitat (Descloux et al 2014, Jones et al. 2015) whereas 68 69 larger bodied invertebrates (e.g., amphipods, isopods, coleopteran Elmidae larvae) are still able to move through these sediments (Boulton et al. 2008). Thus, although 70 no previous publications have presented data on body size across geologies with 71 varying pore sizes, we might expect body size to be larger in rivers with small sized 72 sediments (i.e., those draining chalk and sandstone geologies) than in rivers with 73 74 larger sized sediments (draining limestone). Sediment grain size (and interstitial velocity) is also closely coupled to the hyporheic exchange rate (the movement of 75 water between the HZ and the surface stream), which determines the supply of 76 77 organic matter and dissolved oxygen and thus hyporheos abundance and distribution (Hakenkamp & Palmer 2000). Depth in the HZ is another important 78 factor. In deeper layers the reduction in oxygen availability and pore space due to 79

sediment agglomeration limits the vertical distribution of larger organisms which are
replaced by a suite of numerous but small–bodied taxa (Schmid–Araya 1994,
Maridet & Philippe 1995, Strayer et al. 1997, Stead et al. 2004, Peralta-Maraver et
al. 2018, Peralta-Maraver et al. submitted).

84

Increased fine sediment loading (colmation), often resulting from agricultural 85 86 practices, alters the streambed physical habitat template and is a major and well recognised global threat to stream ecological integrity, leading to reduced hyporheos 87 88 diversity and modified community structure and functions (e.g., Jones et al. 2015, Descloux et al. 2014, Mathers et al 2017). In contrast, there is very limited information 89 on the underlying variability in the structure and function of metazoan HZ communities 90 91 (hyporheos) across streams draining different geologies, despite the centrality of such data to understanding how fine sediment ingress will affect hyporheos. 92

93

94 The range of species' functional traits (functional diversity) likely underpins a key mechanistic link between species richness and ecosystem functioning and a growing 95 body of evidence suggests that community functional diversity predicts ecosystem 96 functioning better than species based indices (e.g. Gagic et al. 2015, Griffin et al. 97 98 2009). Functional diversity can be decomposed into functional richness and 99 functional redundancy (Carmona et al 2016); communities with high functional redundancy, where two ecological units have the same trait values and therefore 100 occupy the same functional space, are thought to be more resilient to disturbances 101 102 (i.e., they can lose species without a great loss of function; Carmona et al 2016). Thus understanding how functional diversity, richness and redundancy vary across 103 the physical habitat template in the hyporheic zone is an important goal given the 104

likely role of the hyporheic community in riverine ecosystem functioning. However,
very few studies have used a trait-based approach for hyporheic communities and of
those that have, to our knowledge, none have included measures of functional
diversity, richness and redundancy (see Claret et al. 1999, Robertson & Wood 2010,
Descloux et al. 2014, Bona et al. 2016, Doretto et al. 2107, Mathers et al. 2017).

110

The objective of this study was to assess differences in hyporheos composition and function across streams draining differing geologies. We investigated eight UK rivers of three distinct geologies with differing values of hydraulic conductivity and porosity. We sampled multiple locations incorporating upwelling and downwelling sites and two depths on two occasions. We expected that the community composition, abundance, body size, biomass, and functional measures of the hyporheos in the rivers would differ in response to the variation in these features of the physical habitat template.

118 We hypothesised that:

119 1. Hyporheos structure (in terms of composition and abundance) will vary

between geologies resulting from the varying habitat template. We expect that
hyporheos will be more abundant and characterised by meiofauna in geologies
that weather to produce coarse-grained sediments.

Taxonomic richness, body size, biomass and diversity will also differ between
 geologies and with depth. We expect that body size will be greatest where
 pore size is small and the other measures will be highest where sediment size/
 pore space is greatest and decline with increasing depth.

Functional diversity, richness and redundancy will differ between geologies in
 response to differences in the physical habitat template. We also expect that
 these measures will decline with increasing depth.

130

131 **2. METHODS**

132 2.1 Survey design

Three geologies of differing hydraulic conductivity (k) and porosity (n) were studied 133 (i) cretaceous chalk k = 10^{-8} - 10^{-5} m/s, n= 0.15 – 0.45; (ii) permotriassic sandstone 134 $k=10^{-10}-10^{-5}$ m/s, n=0.05-0.35; (iii) carboniferous limestone k=10^{-6}-10^{0} m/s, n= 0.05-135 0.5. (Hiscock 2007). Two catchments each were selected in chalk, sandstone, 136 limestone (unglaciated) and limestone (glaciated) geologies (Table 1). In each 137 catchment four sites (permanently flowing riffles) were selected on the main stem 138 and tributaries (Table 1). We ensured that sites differed as little as possible apart 139 from geology; all sites had similar water quality, discharge, catchment area and 140 elevation (Table 1). The land use in all catchments was predominantly pasture. All 141 sites were located on 3rd to 4th order streams (stream order derived from Ordnance 142 Survey maps of the local area following Strahler 1964). Stream bank widths were 143 between 3-5 metres and streams were at least 5cm deep. 144

145 2.2. Sediment Grain Size

The assessment of sediment grain size at each site was undertaken using a variety of methods, i) broad scale site selection determined by the local geology and associated hydraulic properties of the aquifer (see above), ii) measurement of vertical hydraulic gradient (VHG) from both the head and tail of each riffle following

150 Baxter et al. (2003). Briefly a mini-piezometer was inserted to a depth of 30cm, purged and left to settle for 2 hours before measurement. VHG = $(h_s - h_p)/L$. h_s is 151 the distance the water in the mini-piezometer rises above the stream water surface. 152 h_p is the distance from the top of the mini-piezometer to the water level in the mini-153 piezometer and L is the depth of the mini-piezometer in the sediment, iii) collection 154 of sediment samples from the head of each riffle during late autumn when 155 groundwater levels and river flow velocities were reduced. Sediment was collected 156 using a spade (depth penetration 20cm, surface area 20cm²), emptied into a 100 µm 157 158 mesh net held just downstream and stored in a plastic bag. Sediment samples were air dried then weighed to the nearest gram to obtain the total sediment mass and fed 159 into a sediment shaker following the approach of Gordon et al. (2004). Sediment 160 161 fractions were expressed as a percentage of the total sediment mass. iv) at 50cm depth, granulometric conditions were assessed by recording the time taken to pump 162 5 litres of water because there is a strong positive relationship between hydraulic 163 conductivity and sediment grain size (Sharp 1988; Descloux et al. 2010). 164 The combined measures gave a weight of evidence approach to quantification of the 165 sediment proportions and associated properties for each catchment across the three 166

167 geologies.

168 2.3. Hyporheic community sampling:

Each riffle was sampled on two occasions 1) at high groundwater levels (summer
2008) and 2) at low groundwater levels (autumn 2008). In summer 2008, three
replicate hyporheos samples were taken from the deep (50 cm) HZ at both the head
(downwelling) and tail (upwelling) of each riffle (192 samples in total). In autumn
2008, the same survey design was followed except samples were additionally taken

from the shallow (20 cm) HZ (384 samples in total) because low numbers ofindividuals were collected at 50 cm depth.

Samples were collected using the Bou-Rouch method (Bou and Rouch 1967; Bou 176 1974). Piezometers were installed at least 1 metre apart and allowed to settle for at 177 least one hour before pumping 5 litres of water (after Kibichi et al 2009). This was 178 filtered through a 63 µm sieve and preserved in 70% IMS. Preserved samples were 179 sorted under a stereomicroscope and photographed at calibrated magnifications. 180 Subsequently the length and width of all organisms (except Oligochaeta, Nematoda 181 and Acari) were measured to the nearest micrometre using these images and an 182 image analyser. If large groups of similar sized instars were found (>50 µm, primarily 183 Bryocamptus sp. and Chironomidae larvae) all individuals were counted, then 20% 184 of the group was measured to obtain a mean value which was then applied to the 185 whole group. The prolate ellipsoid volume of each group was calculated following 186 Omesova et al. (2008). 187

188 V = $4/3\pi$ | w²

189 w = width, I = length, V = volume (mm³)

Invertebrates were identified to species level where possible using the following
keys: Copepoda (Einsle 1993, Janetzky et al. 1996); Cladocera (Scourfield 1994);
Macroinvertebrates (Holland 1972, Elliott and Humpesch 1983, Hynes 1993, Wallace
et al. 2003, Edington and Hildrew 2005) and Groundwater Crustacea (Gledhill 1993).
Oligochaeta, Diptera, Ostracoda and Nematoda were not identified further due to the
poor preservation of Oligochaeta and Nematoda and the predominance of early
instar Chironomidae larvae.

198 2.4. Water chemistry

Measurements were taken directly from each piezometer. Conductivity, dissolved 199 oxygen (DO), temperature, and pH were determined using portable meters (pH -200 201 Hanna HI-9025; electrical conductivity - Hanna HI-9635; dissolved oxygen temperature - Hanna HI-9145). Then, 0.1 litre of water was withdrawn from the 202 piezometer and frozen within 12 hours. Subsequently, on the day of analysis, water 203 204 samples were defrosted and passed through a 0.7 µm filter to remove particulate matter. Cations and anions were analysed using high performance anion-exchange 205 206 chromatography (DIONEX ED40 electrochemical detector (minimum reporting value =1.0 mg/L; Thermo Fisher Scientific, Waltham, Massachusetts). 207

208

209 2.5. Functional traits

Traits were characterised for 37 taxa using trait tables from Descloux et al. (2014) for 210 the permanent meiofauna and Tachet et al. (2010) for temporary meiofauna. Trait 211 profiles were described mainly at genus level except for some Diptera, Oligochaeta 212 (described as trait profiles of potential families in the corresponding biogeographic 213 area) and Copepoda (described as Harpacticoida or Cyclopoida). Nine traits with 55 214 modalities incorporating a range of life history, mobility, morphological, and 215 ecological traits were included following Descloux et al. (2014) but excluding the trait 216 217 'fecundity' due to lack of data (see Supplementary Information Table S25). A score was assigned to each taxon describing its affinity for that modality (Descloux et al. 218 2014; Tachet et al. 2010). The mean trait profile of each riffle assemblage was 219 220 obtained by weighting the individual trait profiles of taxa by their total abundance in the sample (replicate and upwelling/ downwelling samples were pooled to give a 221 single value of abundance per riffle and water level because we did not detect 222

significant differences in abundance depending on the direction of vertical flow).

Then, the sums of the weighted scores (one per trait category) were rescaled to sum

to one for each trait and riffle and depth, following Descloux et al. (2014).

226

227 2.6. Data analysis

2.6.1 Community structure - Composition and abundance of all identified taxa during 228 the study period was averaged by site and depth in order to merge all the recovered 229 information of the community structure per geology treatment. Then, a NMDS 230 231 analysis (Non-metric Multi-Dimensional Scaling) was performed to describe the dissimilarity among communities based on their structure (taxonomic composition 232 and abundance). This is an effective method for the ordinations of ecological data 233 234 that works with rank orders dissimilarity matrixes (rather than absolute values) and can handle nonlinear responses of the biological attributes of any shape and 235 effectively and robustly find the underlying gradients (Oksanen, 2015; Quinn & 236 Keough, 2002). In order to handle excessively large differences between abundance 237 values, these were transformed using Wisconsin double standardization (Bray & 238 Curtis, 1957). Dissimilarity matrices between communities used in the ordination 239 were calculated with the quantitative Bray-Curtis Index. An iterative process was 240 carried out to find the ordination with the best goodness of fit (Oksanen, 2015). 241

Subsequently, analysis of similarity (ANOSIM) was performed to determine whether there was a significant effect of the geology on the structure of the communities (*p-value* < 0.05) and the reliability of the test (ANOSIM R value). This analysis also uses a rank order dissimilarity matrix to determine whether there is a significant difference between two or more sample groups based on whether there are greater differences between these groups compared to within groups (Oksanen,

248 2015). In this manner, we compared the magnitude of differences intra-treatments 249 (ANOSIM within- and between-group rank dissimilarities are available as 250 Supplementary information: Fig S2). Finally, post-hoc comparisons between geologies 251 were conducted with PerMANOVA pairwise contrasts to identify those treatments that 252 differed significantly. All analyses were performed with the Vegan package (Oksanen 253 et al., 2013) within the R software platform (R Core Team 2014).

254

2.6.2 Taxonomic richness, body size, biomass, community diversity and functional 255 measures (diversity, richness, redundancy). Taxonomic richness, body size, 256 biomass, Shannon-Wiener diversity, functional diversity, functional richness, and 257 functional redundancy were compared between geologies (factor with four levels: 258 chalk, sandstone, unglaciated limestone and glaciated limestone) and depth 259 compartment (factor with two levels: 20 and 50 cm) using low water level data and 260 applying 2-way ANOVA tests. Then, taxonomic richness, body size, biomass, 261 262 Shannon-Wiener diversity, functional diversity, functional richness and functional redundancy were also compared between geologies during the whole study period 263 by using only data from 50 cm depth with a nested 1-way ANOVA test, correcting by 264 water level. In order to solve heterogeneity in the residuals, body size was log 265 transformed, but this was not necessary for the rest of responses. No differences 266 were found in the studied responses between downwelling and up-welling conditions 267 within riffles during data exploration. Accordingly, metrics were pooled by riffle (riffle 268 as a replicate per site). Dependency structure of the residuals with the study site 269 (Site) was incorporated in the previous ANOVA tests in the design of the study 270 (random factor). In this manner, we accounted for repeated measures (non-271 272 dependence of residuals with study site in the ANOVA tests). In order to fit the

273 random factor to the ANOVA tests we used the Imer function of the R package Ime4 (Bates et al. 2017, R Core Team 2016). Validation of underlying assumptions of 274 normality and homocedasticity of tests residuals was applied following Zuur et al. 275 276 (2009) (supplementary material: Fig S1). Finally, Wald chi-square tests were performed in order to assess the significance of the analysed factors on the 277 responses using the Anova function from the car R-package (Fox & Weisberg 2011) 278 and, subsequently, post-hoc Tukey tests were applied to compare which specific 279 treatments differ significantly using the *difflsmeans* function from the R-package 280 281 ImerTest (Kuznetsova et al. 2015).

282

283 3. RESULTS

The physical habitat template showed clear differences between geologies and 284 across water levels. Pump times were significantly higher in rivers draining chalk 285 286 than in other geologies (F_{3,176}=21.9, p<0.001, Table 2) and these findings were supported by VHG measurements (Table 2). Chalk rivers also had the highest 287 percentage of sediments in the 0.5 – 2mm size category. Conductivity (F_{3,563}=403.0, 288 p<0.001), pH (F_{3,563}=116.1, p<0.001), dissolved oxygen (F_{3,563}=7.52, p<0.001), 289 temperature (F_{3,563}=200.3, p<0.001), and nitrate (F_{3,563}=124.2, p<0.001), also 290 differed significantly between geologies. Conductivity (F1,563=12.3, p<0.001), pH 291 (F_{1,563}=11.9, p<0.001), dissolved oxygen (F_{1,563}=86.0, p<0.001), temperature 292 (F_{1,563}=2097.9, p<0.001), and nitrate (F_{1,563}=4.4, p<0.001), also differed significantly 293 294 between water levels. See Table 2 for means and standard errors of these parameters across geologies and water levels. 295

296

297 The NMDS ordination model based on taxonomic composition and abundance was run 20 times for the two-dimensional ordination with a very high goodness of fit 298 between the distances in the ordination against the original data (stress = 0.20, 299 linear fit $R^2 = 0.98$, nonmetric fit $R^2 = 0.86$). Accordingly, the Shepard plot of the final 300 model has small scatter around the fitted line (Supplementary Information Fig S1) 301 and the bi-dimensional ordination is representative of the original differences 302 303 between assemblages. The NMDS-plot clearly discriminated assemblages between geologies (Fig 1a) and showed which taxa contributed most to the observed 304 305 dissimilarities (Fig 1b). ANOSIM analysis found a significant effect of geology on the observed dissimilarities between assemblages (R = 0.63, p-value = 0.002), 306 supporting the NMDS plot. Thus our first hypothesis, community composition will 307 308 differ between geologies, was upheld. The post hoc PerMANOVA pairwise contrasts 309 detected significant differences between limestone assemblages (both glacial and un-glacial) with sandstone and chalk sites respectively. However, these differences 310 were not significant between sandstone and chalk assemblages, or between glacial 311 and un-glacial limestone assemblages (Fig 1a, Table S25) i.e., differences in 312 assemblage structure (composition and abundance) between sandstone and chalk 313 sites and between both limestones were too small to be able to characterise them as 314 different communities. The hyporheos of Limestone catchments was characterised 315 316 by a suite of stoneflies and caddisflies (e.g., *Leuctra* spp., Perlodidae, *Limnephilus* sp.) and microcrustacea (e.g., Paracyclops sp., Moraria sp., Elaphoidella sp.) (Fig. 317 1b). Chalk catchments were defined by large-bodied Crustacea (e.g., Gammarus 318 319 pulex, Niphargus spp., Crangonyx sp.) whereas indicative taxa of sandstone catchments included harpacticoid (e.g., Maraenobiotus sp., Canthocamptus sp.) and 320 cyclopoid (Eucyclops sp.) copepods (Fig. 1b). 321





Fig 1: (a) NMDS ordination model of different HZ assemblages by geology including ANOSIM statistic (R-value) and significance (p-value). Polygons group the assemblages by geology sites in the ordination. Assemblages whose structure is not significantly different after post-hoc PerMANOVA tests (P-value < 0.005) are indicated with the same letter. (b) Distribution of the most influential taxa on the ordination. In order to simplify interpretation of the plot, those taxa falling in the centre of the ordination (low influence in the ordination) have been removed. For explanations of species abbreviations see Supplementary Table 23.

Our second hypothesis; taxonomic richness, body size, biomass and Shannon-Wiener diversity will differ across geologies and depths was partly supported. Only taxonomic richness showed significant differences across some geologies with larger values in rivers draining unglaciated limestone (Fig. 2b). In contrast, depth showed a strong effect on all these responses. Taxonomic richness, body size, biomass and Shannon-Wiener diversity declined with increasing depth for all geologies (Fig. 2a, c, e, g) and in most cases these differences were significant.

3.4. Our third hypothesis; functional measures will differ with geology and decline
with increasing depth was only partially upheld. Functional measures did not differ
significantly across geologies either in the summer data or across the whole study
period (Fig. 2i – n). However, functional richness and redundancy declined
significantly with increasing depth for several geologies (Fig. 2 I,k). Numerical results
and p-values from the Wald chi-square and post-hoc Tukey tests are available as
Supplementary information (Tables S1 - S22).

- 352
- 353
- 354
- 355
- 356
- 357
- 358
- 359
- 360
- 361
- 362



Fig 2: Taxonomic richness, body size, biomass, Shannon-Wiener diversity and
functional richness, redundancy and diversity (mean ±1SE/ 5 L). Summer data
(a,c,e,g,i) are pooled across the four geologies (CC: Chalk, PTS: permo-triassic
sandstone, UGL: unglaciated limestone, GL: glaciated limestone) and two depths
(20cm and 50cm). While, data from 50 cm depth (b,d,f,h,j) of the whole study period
are just pooled across geologies. Geologies where total responses are not significantly
different are indicated with the same letter (95% CI).

4. DISCUSSION

371 Our study showed clear differences in the HZ physical habitat template in rivers draining different geologies at the regional scale and corresponding variations in the 372 community composition of these HZ's. Geologies that weather easily to produce fine 373 grained alluvial sediments (chalk and sandstone) possessed hyporheos that differed 374 significantly from those that were present in rivers where the geology did not weather 375 so easily resulting in coarser grained sediments (carboniferous limestones). The 376 hyporheos in coarse-grained limestone rivers included an abundance of small-377 bodied copepods, which were able to exploit the larger pore spaces available. In 378 379 contrast the hyporheos of chalk rivers was dominated by larger macroinvertebrates, presumably because they were able to burrow through these sediments more easily 380 (Boulton et al. 2008). Although copepods did occur in these rivers, they were much 381 382 less abundant and diverse.

383

We detected clear differences in taxonomic richness between geologies in 384 agreement with previous studies (e.g. Strayer et al. 1997; Ward et al. 1998; Bo et al. 385 2007, Descloux et al. 2014) despite the lower % of fine sediments in our study. 386 These authors noted that decreasing sediment size and interstitial space exert a 387 negative effect on abundance and taxon richness (although we did not compare 388 abundance between geologies as the taxonomic composition of the hyporheos 389 390 varied so markedly). To our knowledge only one publication has compared the biomass of metazoan hyporheos across sediment types in perennial streams 391 (Reynolds & Benke 2012). They found that the biomass of chironomid assemblages 392 393 was highest in the HZ's of gravel-cobble reaches, intermediate in sand reaches and lowest in mud-silt reaches. Our study is the first to compare HZ biomass of the 394 whole metazoan community retained on a 63µm sieve across geologies with 395

396 different pore sizes. Our finding, that biomass was similar across geologies, suggests that the HZ's of all these rivers support a similar quantity of hyporheos 397 although the structure and composition of these communities differ. Contrary to our 398 399 expectations, although rivers draining limestone included small bodied taxa and chalk rivers were characterised by taxa with larger body sizes, this did not translate 400 into significant body size differences between the rivers, perhaps because 401 macroinvertebrate taxa were represented by very early instar individuals. Our 402 findings suggest that individuals of any body size are equally likely to penetrate bed 403 404 sediments of the three geologies and that the HZ's may therefore have equal connectivity with the surface in terms of organism movement. 405

406

Our findings, that taxonomic richness declined with increasing depth and that body
size and biomass tended to decrease with depth, concur with those of other studies
(e.g., Strayer et al. 1997, Davy-Bowker et al. 2006, Pacioglu & Robertson 2017,
Peralta-Maraver et al. submitted). This likely results from a reduction in pore space
and oxygen with increasing depth (Maridet & Philippe 1995, Strayer et al. 1997,
Peralta-Maraver et al. 2018, Peralta-Maraver et al. submitted).

413

HZ's are increasingly recognised as a central component of the lotic ecosystem,
particularly regarding their functional role in the attenuation of nutrients and
pollutants (e.g., Lewandowski et al. 2011; Peralta-Maraver et al. 2018) and their
importance as a refuge from surface disturbances (e.g., Robertson & Wood 2010,
Maazouzi et al. 2017). Although biodiversity is understood to be a fundamental
driver of ecosystem functioning, there is a growing acceptance that the diversity of
species functional traits – functional diversity - within a community underpins and

421 better describes ecosystem functioning (Gagic et al. 2015). Our study is the first to determine the functional richness, redundancy and diversity of hyporheos in 422 contrasting geologies. Despite differences in the physical habitat template, which 423 424 resulted in distinctive hyporheos per geology, none of our measures of ecosystem function significantly differed between geologies implying equal functioning across 425 the HZ's. Several studies have demonstrated the impact of fine sediment deposition 426 (colmation) on individual functional traits of hyporheos (e.g., Descloux et al. 2014, 427 Bona et al. 2016, Doretto et al. 2017, Mathers et al. 2017). We therefore expected, 428 429 but did not find, that functional diversity would be lowest in the fine sediment rivers. As anticipated, we found functional measures declined with increasing depth, as did 430 taxonomic richness suggesting that ecosystem functioning decreases with increasing 431 432 depth. Thus, our results imply that maximal functioning of the metazoan hyporheos occurs in the shallow HZ in accordance with findings for microbial biofilms in the HZ 433 (Battin et al. 2003, O'Connor & Harvey 2008, Boano et al. 2014, Knapp et al. 2017). 434 Our findings suggest that, for these relatively unimpacted rivers draining three 435 different geologies, the HZ's all supported similar quantities of hyporheos, had 436 similar connectivity with the surface stream and had similar levels of functional 437 richness, redundancy and diversity, all of which implies relatively similar ecological 438 functioning irrespective of differences in the physical template and community 439 440 structure between geologies. However, our approach to characterising ecosystem functioning within the HZ was guite broad brush; we do not yet know the extent to 441 which their capacity to undertake these functions is resilient to environmental 442 perturbations i.e., the degree of response diversity within the functionally redundant 443 elements of the community (Mori et al. 2013) and whether this differs across 444 geologies. 445

5. CONCLUSION 447

The effects of fine sediment accumulation in streams is a global concern yet there is 448 449 limited understanding of the underlying variation in hyporheos structure and function across streams draining differing geologies despite their contribution to the delivery 450 of ecosystem services such as pollution attenuation. Such data is essential to 451 452 contextualise the effect of perturbations such as fine sediment ingress. We found that despite clear differences in metazoan community structure between streams 453 454 with fine sediment and streams with coarse sediment, there were no significant differences in biomass or functional measures between the geologies implying that 455 the HZ's in all geologies functioned equally. 456

457

458

6. ACKNOWLEDGEMENTS

MD acknowledges the support of a NERC industrial Case studentship. IPM 459 acknowledges funding from the European Union's Horizon2020 research and 460 innovation programme under Marie-Skłodowska-Curie grant agreement No. 461 641939. Thanks to the Environment Agency for assistance with site selection. We 462 acknowledge the detailed comments of six anonymous reviewers that improved the 463 manuscript. 464

465

466

7. COMPETING INTERESTS STATEMENT

The authors have no competing interests to declare. 467

468

8. REFERENCES 469

470	Bardini L., Boano F., Cardenas M.B., Revelli R., Ridolfi L. (2012). Nutrient cycling in							
471	bedform induced hyporheic zones. Geochimica et Cosmochimica Acta: 84. 47-61.							
472								
473	Bates, D., Maechler, M, Bolker, B. & Walker, S. (2016) Ime4: Linear Mixed-Effects							
474	Models using 'Eigen' and S4. R package version 1.1-12. Retrieved from							
475	https://cran.r-project.org/web/packages/Ime4/index.html							
476								
477	Battin, T. J., Kaplan, L. A., Newbold, J. D., & Hansen, C. M. (2003) Contributions of							
478	microbial biofilms to ecosystem processes in stream mesocosms, Nature 426, 439-							
479	442.							
480								
481	Baxter, C., F. R. Hauer, and W. W. Woessner. 2003. Measuring groundwater -							
482	streamwater exchange: New techniques for installing minipiezometers and							
483	estimating hydraulic conductivity. Transactions of the American Fisheries Society							
484	132 :493-502.							
485								
486	Bo, T., Fenoglio, S., Malacarne, G., Pessino, M., Sgariboldi, F. (2007) Effects of							
487	clogging on stream macroinvertebrates: An experimental Approach Limnologica 37:							
488	186–192							
489								
100	Boano E Harvey I.W. Marion & Packman & Revelli R. Ridolfi I. &							
490								
491	Wörman, A. (2014), Hyporheic flow and transport processes: Mechanisms, models,							
492	and biogeochemical implications, Reviews in Geophysics, 52, 603-679, doi:10.1002/							
493	2012RG000417.							

494	Bona, F., Doretto, A., Falasco, E., La Morgia, V., Piano, E., Ajassa, R., Fenoglio,								
495	S. (2016) Increased sediment loads in alpine streams: an integrated field study.								
496	River Research & Applications, 32: 1316-1326.								
497									
498	Bou C. (1974) Recherches sur les eaux souterraines -25- Les méthodes de récolte								
499	dans les eaux souterraines interstitielles. Annales de Spéléologie, 26, 611-619.								
500									
501	Bou, C. & Rouch, R. (1967) Un nouveau champ de recherches sur la faune								
502	aquatique souterraine. Comptes Rendus Hebdomadaires des Seances de I								
503	Academie des Sciences, 3, 369-370.								
504									
505	Boulton A. J., Fenwick G. D., Hancock P. J., Harvey M. S. (2008) Biodiversity,								
506	functional roles and ecosystem services of groundwater invertebrates. Invertebrate								
507	Systematics, 22, 103-116.https://doi.org/10.1071/IS07024.								
508									
509	Bray, J. R., & Curtis, J. T. (1957). An ordination of the upland forest communities of								
510	southern Wisconsin. Ecological Monographs, 27, 325–349.								
511									
512	Brunke, M., Gonser, T. (1997) The ecological significance of exchange processes								
513	between rivers and groundwater. Freshwater Biology, 37, 1-33.								
514									
515	Carmona, C.P., de Bello, F., Mason, N.W.H., Leps J (2016) Traits Without Borders:								
516	Integrating Functional Diversity Across Scales. Trends in Ecology and Evolution 31:								
517	382-394. https://doi.org/10.1016/j.tree.2016.02.003.								

519	Claret C., Marmonier P., Dole-Olivier M-J., Creuze des Chatelliers M., Boulton A.J.,
520	Castella E. (1999) A functional classification of interstitial invertebrates: supporting
521	measures of biodiversity using species traits and habitat affinities. Archiv fur
522	Hydrobiologie 145: 385-403.
523	
524	Davy-Bowker, J., Sweeting, W., Wright, N., Clarke, R. T. & Arnott, S. (2006) The
525	distribution of benthic and hyporheic macroinvertebrates from the heads and tails of
526	riffles. Hydrobiologia 563, 109-123.
527	
528	Descloux, S., Datry, T., Philippe, M. Marmonier, P. (2010) Comparison of Different
529	Techniques to Assess Surface and Subsurface Streambed Colmation with Fine
530	Sediments. International Review of Hydrobiology, 95, 520-540.
531	
532	Descloux, S., Datry, T. & Usseglio-Polatera, P. (2014) Trait-based structure of
533	invertebrates along a gradient of sediment colmation: Benthos versus hyporheos
534	responses. Science of the Total Environment, 466, 265-276.
535	
536	Doretto, A., Bona, F., Piano, E., Zanin, I., Eandi, A.C., Fenoglio, S. (2017) Trophic
537	availability buffers the detrimental effects of clogging in an alpine stream. Science of
538	the Total Environment. 592: 503-511.
539	
540	Edington, J.M. & Hildrew, A.G. (2005) A Revised Key to the Caseless Caddis Larvae
541	of the British Isles, with Notes on their Ecology, Windermere, Freshwater Biological
542	Association.

543	Einsle, U.K. (1993) Crustacea: Copepoda: Calanoida und Cyclopoida. IN
544	SCHWOERBEL, J. & ZWICK, P. (Eds.) Süsswasserfauna von Mitteleuropa. Gustav
545	Fischer Verlag.
546	
547	Elliott, J.M. & Humpersch, U.H. (1983) A Key to the Adults of the British
548	Ephemeroptera, with Notes on their Ecology, Windermere, Freshwater Biological
549	Association.
550	
551	Fox, J., & Weisberg, S. (2011) An R companion to applied regression. Sage
552	Publications. http://socserv.socsci.mcmaster.ca/jfox/Books/Companion
553	
554	
555	Gagic, V., Bartomeus, I., Jonsson, T., Taylor, A., Winqvist, C., Fischer, C., Slade,
556	E.M., Steffan-Dewenter, I., Emmerson, M., Potts, S.G., Tscharntke, T., Weisser, W.,
557	Bommarco, R. (2015) Functional identity and diversity of animals predict ecosystem
558	functioning better than species-based indices. Proceedings of the Royal Society
559	B.282:20142620. http://dx.doi.org/10.1098/rspb.2014.2620.
560	
561	Gandy, C.J., Smith, J.W.N., Jarvis, A.P. (2007) Attenuation of mining-derived
562	pollutants in the hyporheic zone: A review. Science of the Total Environment, 373,
563	435-446.
564	
565	Gelman, A., Su, Y.S., Yajima, M., Hill, J., Pittau, M.G., Kerman, J., Zheng, T., Dorie,
566	V., 2009. arm: data analysis using regression and multi–level/hierarchical models. R
567	package, version 9.01. <u>http://CRAN.R-project.org/package=arm</u> .

500	
569	Gledhill, T. (1993) British Freshwater Crustacea Malacostraca: A Key with
570	Ecological Notes, Windermere, Freshwater Biological Association.
571	
572	Gordon, N. D., McMahon, T., Finlayson, B. L., Gippel, C. J. & Nathan, R. J. (2004).
573	Stream Hydrology: An Introduction for Ecologists, 2nd Edition John Wiley and Sons,
574	Ltd, Chichester.
575	
576	Griffin, J. N., de la Haye, K. L., Hawkins, S. J., Thompson, R. C. and Jenkins, S. R.
577	(2008) Predator diversity and ecosystem functioning: Density modifies the effect of
578	resource partitioning. Ecology, 89: 298–305. doi:10.1890/07-1220.1.
579	
580	Hakenkamp, C.C Palmer, M. A. (2000). The ecology of hyporheic meiofauna. In
581	streams and Ground Waters, Jones J.B. & Mulholland P.J. Eds. Academic Press:
582	New York. Pp 307-336.
583	
584	Hester, E. T.; Young, K. I.; Widdowson, M. A. (2013) Mixing of surface and
585	groundwater induced by riverbed dunes: Implications for hyporheic zone definitions
586	and pollutant reactions. Water Resources Research. 49, 5221-5237.
587	
588	Hiscock, K. (2007) Hydrogeology: Principles and practice, Oxford, Blackwell
589	publishing.
590	

- Holland D.G. (1972) A key to the larvae, pupae and adults of the British species of *Elminthidae*, Windermere, Freshwater Biological Association, Scientific Publication
 No. 26.
- 594
- 595 Hynes, H.B.N. (1993) A Key to the Adults and Nymphs of the British Stoneflies
- 596 (Plecoptera), with Notes on their Ecology and Distribution, Far Sawrey, Cumbria,
- 597 Freshwater Biological Association.
- 598
- Jones, I., Growns, I., Arnold, A., McCall, S., Bowes, M. (2015) The effects of
- 600 increased flow and fine sediment onhyporheic invertebrates and nutrients in stream
- 601 mesocosms. Freshwater Biology 60: 813-826 doi:10.1111/fwb.12536.
- 602
- Knapp, J. L. A., R. González-Pinzón, J. D. Drummond, L. G. Larsen, O. A. Cirpka,
- and J. W. Harvey (2017), Tracer-based characterization of hyporheic exchange and
- 605 benthic biolayers in streams, Water Resources Research, 53, 1575–1594,
- 606 d<u>oi:10.1002/2016WR019393.</u>
- 607
- Krause, S., Hannah, D. M. & Fleckenstein, J. H. (2009) Hyporheic hydrology:
- interactions at the groundwater-surface water interface. Hydrological Processes, 23,
- 610 2103–2107.

- 612 Krause, S., Boano, F., Cuthbert, M.O., Fleckenstein, J.H. & Lewandowski,
- J. (2014), Understanding process dynamics at aquifer-surface water interfaces: An
- 614 introduction to the special section on new modeling approaches and novel

615	experimental technologies, Water Resources Research, 50, 1847–1855,
616	doi: <u>10.1002/2013WR014755</u> .
617	
618	Kuznetsova, A., Brockhoff, P. B., & Christensen, R. H. B. (2015). ImerTest: tests in
619	linear mixed effects models. R package version 2.0-20. Vienna: R Foundation for
620	Statistical Computing. https://cran.r-project.org/web/packages/ImerTest/index.html
621	
622 623 624 625 626 627	Lewandowski, J., Putschew, A., Schwesig, D., Neumann, C., Radke, M. (2011) Fate of organic micropollutants in the hyporheic zone of a eutrophic lowland stream: Results of a preliminary field study. Science of the Total Environment 409, 1824- 1835. <u>doi.org/10.1016/j.scitotenv.2011.01.028</u> .
628	Maridet, L., & Philippe, M. (1995) Influence of substrate characteristics on the
629	vertical distribution of stream macroinvertebrates in the hyporheic zone. Biologia,
630	91, 101–105.
631	
632	Mathers, K.L., Rice, S.P., Wood, P.J. (2017) Temporal effects of enhanced fine
633	sediment loading on macroinvertebrate community structure and functional traits.
634	Science of the Total Environment. 599-600: 513-52.
635	https://doi.org/10.1016/j.scitotenv.2017.04.096
636	
637•	Maazouzi C, Galassi D, Claret C, Cellot, B., Fiers, F., Martin, D., Marmonier, P.,
638	Dole-Olivier, M-J. (2017) Do benthic invertebrates use hyporheic refuges during

639 streambed drying? A manipulative field experiment in nested hyporheic

640 flowpaths. Ecohydrology. 2017;10:e1865. <u>https://doi.org/10.1002/eco.1865</u>

641

- Mori, A. S., Furukawa, T. and Sasaki, T. (2013), Response diversity determines the
 resilience of ecosystems to environmental change. Biological Reviews, 88: 349–
 364. doi:10.1111/brv.12004.
- 645
- Morrice, J. A., Valett, H. M., Dahm, C. N. and Campana, M. E. (1997), Alluvial
- 647 characteristics, groundwater-surface water exchange and hydrological retention in
- headwater streams. Hydrological. Processes, 11: 253–267.

649

- O'Connor, B. L. & Harvey, J. W. (2008) Scaling hyporheic exchange and its influence
- on biogeochemical reactions in aquatic ecosystems, Water Resources Research, 44,
- 652 W12423.
- 653
- Oksanen, J. (2015). Multivariate analysis of ecological communities in R: vegan
- 655 *tutorial*. Retrieved from <u>http://cc.oulu.fi/~jarioksa/opetus/metodi/vegantutor</u>.

656

- Oksanen, J., Guillaume Blanchet, F., Kindt, R., Legendre, P., Minchin, P.
- 658 R., & O'Hara, R. B., et al. (2013). Vegan: Community ecology package.
- 659 *R package version 2.0–10.* Retrieved from http://CRAN.R–project.org/
- 660 package=vegan.

662	Omesova, M., Horsak, M. & Helesic, J. (2008) Nested patterns in hyporheic meta-
663	communities: the role of body morphology and penetrability of sediment.
664	Naturwissenschaften, 95, 917-926.
665	
666	Orghidan, T. (1959) Ein neuer Lebensraum des unterirdischen wassers: des
667	hyporheische biotope. Arch. Hydrobiol. 55: 392-414.
668	
669	Pacioglu, O., & Robertson, A.L. (2017) The invertebrate community of the chalk
670	stream hyporheic zone:spatio-temporal distribution patterns Knowledge and
671	Management of Aquatic Ecosystems. 418, 10. DOI: 10.1051/kmae/2017001.
672	
673	Peralta-Maraver, I., Reiss, J. & Robertson, A. L. Interplay of hydrology, community
674	ecology and pollutant attenuation in the hyporheic zone. Science of the Total
675	Environment. 610, 267-275 (2018).
676	
677	Peralta-Maraver, I., Njeru, C., Reiss, J., Horn, M.A., Robertson A.L. submitted
678	Piecing together the puzzle: environmental filtering and community delineation in an
679	ecotone ecosystem.
680	
681	Quinn, G. P., & Keough, M. J. (2002). Experimental design and data analysis
682	for biologists. Cambridge: Cambridge University Press.
683	
684	R Core Team (2016). R: A language and environment for statistical computing.
685	R Foundation for Statistical Computing, Vienna, Austria. Retrieved from
686	http://www.R-project.org.

688	Reynolds Jr, S. K. & Benke, A. C. (2012) Chironomid production along a hyporheic
689	gradient in contrasting stream types. Freshwater Science, 31,167-168.
690	
691	Robertson A.L., Rundle, S.D. and Schmid-Araya, J.M. (2000) Introduction to Special
692	Issue on Lotic Meiofauna. Freshwater Biology, 44: 1-3.
693	
694	Robertson, A. L. & Wood, P. J. (2010) Ecology of the hyporheic zone: origins,
695	current knowledge and future directions. Fundamental & Applied Limnology/Archiv
696	fur Hydrobiologie. 176, 279–289.
697	
698	Schmid–Araya, J. M. (1994) Temporal and spatial distribution of benthic microfauna
699	in sediments of a gravel streambed. Limnology and Oceanography. 39, 1813–1821.
700	
701	Scourfield D.J & Harding J.P. (1966). A Key to the British Species of Freshwater
702	Cladocera, with Notes on their Ecology. Third Edition. Scientific Publication No. 5.
703	Freshwater Biological Association, Ambleside. 61pp.
704	
705	Sharp, J.M. (1988) Alluvial aquifers along major rivers. IN BACK, J. M.,
706	ROSENHEIM, J. M. & SEABER, P. R. (Eds.) Hydrogeology: The geology of North
707	America. Colorado, The geological society of North America.
708	
709	Southwood, T. (1977). Habitat, the Templet for Ecological Strategies? Journal of
710	Animal Ecology, 46(2), 337-365. doi:10.2307/3817.
711	

- Southwood, T. (1988). Tactics, Strategies and Templets. Oikos, 52(1), 3-18.
 doi:10.2307/3565974.
- 714
- Stead, T. K., Schmid–Araya, J. M. & Hildrew, A. G. (2004) The contribution of
- subsurface invertebrates to benthic density and biomass in a gravel stream. Archiv
- 717 fur Hydrobiologie 160, 171–191.
- 718
- Stegen, J.C., Fredrickson, J.K., Wilkins, M.J., Konopka, A.E., Nelson, W.C., Arntzen,
- E.V. Chrisler, W.B., Chu, R.K., Danczak, R.E., Fansler, S.J., Kennedy, D.W., Resch,
- 721 C.T., Tfaily, M., (2016). Groundwater–surface water mixing shifts ecological
- assembly processes and stimulates organic carbon turnover. Nature
- 723 Communications 7, 11237.
- 724
- Strahler, A. N. 1964. Quantitative geomorphology of drainage basins and channel
 networks. Pages 4–39 in V. T. Chow, editor. Handbook of Applied Hydrology.
- 727 McGraw-Hill New York.

- 729
- 730 Strayer, D. L., May, S. E., Nielsen, P., Wollheim, W. & Hausam, S. (1997) Oxygen,
- organic matter, and sediment granulometry as controls on hyporheic animal
- communities. Archiv fur Hydrobiologie. 140, 131–144.
- 733
- Tachet H, Richoux P, Bournaud M, Usseglio-Polatera P. (2010). Invertebres d'eau
- douce. Systematique, biologie, ecologie. CNRS Editions, Paris. 607 p.
- 736

- Townsend, C.R. & Hildrew, A.G. (1994) Species traits in relation to a habitat templet
 for river systems. Freshwater Biology, 31: 265–275. doi:10.1111/j.13652427.1994.tb01740.x.
- 740
- Wallace, I.D., Wallace, B, & Philipson, G.N. (2003) Keys to the Case-bearing Caddis
- 742 Larvae of Britain and Ireland, Windermere, Freshwater Biological Association

- Ward, J., Bretschko, G., Brunke, M., Danielopol, D., Gibert, J., Gonser, T. & Hildrew,
- A.G. (1998) The boundaries of river systems: the metazoan perspective. Freshwater
 Biology, 40, 531-569.

747

- 748 Williams, D. D., Febria, C. M. & Wong, J. C. (2010) Ecotonal and other properties of
- the hyporheic zone. Fundamental & Applied Limnology/Archiv fur Hydrobiologie,
- 750 176, 349-364.
- 751
- Zuur, A. F., Ieno, E. N., Walker, N. J., Savaliev, A. A. & Smith, G. M. Mixed Effects
- 753 Models and Extensions in Ecology with R. (Springer, New York, 2009).

Table 1.Physical and water quality characteristics of the studied catchments

756

Geology	Site location	Catchment area (Km²) ¹	Naturalized mean annual discharge (m ³ S ⁻¹)	Elevation (m)	RBMP C2 status MI ²
Chalk	RIVER FROME N50° 47' 23.0" W002° 34' 49.9" N50° 46' 36.3" W002° 28' 10.4" N50° 45' 07.3" W002° 30' 42.6" N50° 46' 35.4" W002° 34' 26.8" RIVER PIDDIE	30-40	0.31-1.25	78-95	High
	N50° 44' 20.7" W002° 15' 38.2" N50° 44' 42.5" W002° 19' 16.7" N50° 45' 22.3" W002° 13' 35" N50° 45' 17.4" W002° 21' 40.4"	48-70	0.31-0.62	44-60	High
Sandstone	RIVER TONE N51° 03' 12.6" W003° 14' 55.1" N51° 02' 11.4" W003° 16' 8.5" N50° 58' 24.7" W003° 18' 50.1" N50° 59' 14.1" W003° 15' 19.0"	42-65	0.31-1.25	16-62	Good- High
	N50° 54' 49.4" W003° 27' 21.7" N50° 48' 35.2" W003° 38' 53.8" N50° 46' 40.8" W003° 40' 01.8" N50° 46' 41.0" W003° 39' 46.3"	42-65	0.62-1.25	35-83	High
Glaciated Limestone	RIVER URE N54° 16' 11.1" W002° 08' 22.3" N54° 17' 10.5" W001° 58' 38.1" N54° 17' 12.2" W001° 58' 1.9" N54° 13' 23.2" W001° 55' 55.1" RIVER WHARFE	40	0.31-0.62	85-258	High
	N54° 11' 26.01" W002° 05' 43.4" N54° 11' 26.0" W002° 05' 44.4" N54° 08' 51.5" W002° 07' 42.4" N54° 08' 23.8" W002° 06' 42.1"	23-33	0.62-1.25	209-225	Good- High
Unglaciated Limestone	RIVER DERWENT N53° 20' 45.1" WO01° 46' 45.1" N53° 14' 58.0" W001° 51' 23.6" N53° 14' 23.2" W001° 44' 55.1" N53° 14' 17.5" W001° 44' 46.7"	20-70	0.31-2.5	128-255	Good- High
	RIVER DOVE N53° 05' 00.7" W001° 47' 30.0" N53° 03' 13.0" W001° 48' 01.9" N53° 03' 29.4" W001° 46' 37.7" N53° 03' 05.3" W001° 48' 30.1"	45	0.62-2.5	138-171	High

757

758 ¹ Data source <u>http://environment.data.gov.uk/catchment-planning/</u>

759 ²River Basin Management Plan Cycle 2 2016 <u>http://environment.data.gov.uk/catchment-planning/</u>

Table 2: Means (±1 SE) of selected physico-chemical parameters for three geologies and two water levels. Pump time = time (in

minutes) to pump 5 litres of water.

763

Geology	Vertical Hydraulic Gradient. H = riffle head, T = riffle tail	Water level	% sediment (0.5- 2mm)	% sediment (4-16mm)	Pump time	Conductivity (μS)	рН	Dissolved oxygen (mg/L)	Temperature (°C)	NO₃ (mg/L)
Chalk	0.008±0.003 (H)	High	30	48		542.2±7.39	7.7±0.03	5.8±0.17	18.2±0.15	17.2±0.75
	0.0006±0.002 (T)	Low			2.9±0.2	560.7±5.03	7.5±0.03	7.9±0.21	12.3±0.09	14.3±0.80
Sandstone	-0.02±0.005 (H)	High	17.5	62.5		404.2±10.7	7.1±0.03	5.2±0.18	15.3±0.1	15.2±1.14
	0.03±0.005 (T)	Low			1.6±0.1	317.2±11.5	7.3±0.04	7.2±0.13	11.8±0.08	11.5±1.31
Glaciated	-0.06±0.008 (H)	High	21	57		255.2±5.1	7.9±0.03	6.1±0.19	15.0±0.1	1.3±0.07
Limestone	0.08±0.011 (T)	Low			1.5±0.1	243.9±9.2	7.8±0.02	7.7±0.18	10.6±0.07	1.5±0.07
Unglaciated	-0.07±0.012 (H)	High	20	59.5		457.3±6.5	7.8±0.03	7.2±0.19	14.2±0.15	7.4±0.31
Limestone	0.08±0.009 (T)	Low			1.5±0.1	444.9±12.01	7.7±0.03	7.4±0.32	9.6±0.11	8.9±0.5