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# Lineage Functional Types (LFTs): Characterizing functional diversity to enhance the representation of ecological behavior in Land Surface Models

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5	
6	Authors: Daniel M. Griffith <sup>1,2,3*</sup> , Colin Osborne <sup>4</sup> , Erika J. Edwards <sup>5</sup> , Seton Bachle <sup>6</sup> , David J. Beerling <sup>4</sup> ,
7	William J. Bond <sup>7,8</sup> , Timothy J. Gallaher <sup>9,15</sup> , Brent R. Helliker <sup>10</sup> , Caroline E.R. Lehmann <sup>11</sup> , Lila
8	Leatherman <sup>1</sup> , Jesse B. Nippert <sup>6</sup> , Stephanie Pau <sup>12</sup> , Fan Qiu <sup>6</sup> , William J. Riley <sup>13</sup> , Melinda D. Smith <sup>14</sup> ,
9	Caroline A. E. Strömberg <sup>9</sup> , Lyla Taylor <sup>4</sup> , Mark Ungerer <sup>6</sup> , and Christopher J. Still <sup>1</sup>
10	
11	<sup>1</sup> Forest Ecosystems and Society, Oregon State University, OR, U.S.A.
12	<sup>2</sup> US Geological Survey Western Geographic Science Center, Moffett Field, CA, 94035
13	<sup>3</sup> NASA Ames Research Center, Moffett Field, CA, 94035
14	<sup>4</sup> Department of Animal and Plant Sciences, University of Sheffield, U.K.
15	<sup>5</sup> Department of Ecology and Evolutionary Biology, Brown University, RI, U.S.A.
16	<sup>6</sup> Division of Biology, Kansas State University, KS, U.S.A.
17	<sup>7</sup> South African Environmental Observation Network, National Research Foundation, Claremont, South
18	Africa
19	<sup>8</sup> Department of Biological Sciences, University of Cape Town, Rondebosch, South Africa
20	<sup>9</sup> Department of Biology and the Burke Museum of Natural History and Culture, University of
21	Washington, Seattle, WA, U.S.A.
22	<sup>10</sup> Department of Biology, University of Pennsylvania, PA, U.S.A.
23	<sup>11</sup> School of GeoSciences, University of Edinburgh, Edinburgh, U.K.
24	<sup>12</sup> Department of Geography, Florida State University, FL, U.S.A.
25	<sup>13</sup> Lawrence Berkeley National Laboratory, CA, U.S.A.
26	<sup>14</sup> Department of Biology, Colorado State University, CO, U.S.A.
27	<sup>15</sup> Bishop Museum, Honolulu, HI, U.S.A
28	*Corresponding author: (T) +19105450632, (E) griffith.dan@gmail.com
29	
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### 35 Summary (200/200):

- 36 Process-based vegetation models attempt to represent the wide range of trait variation in biomes by
- 37 grouping ecologically similar species into plant functional types (PFTs). This approach has been
- 38 successful in representing many aspects of plant physiology and biophysics, but struggles to capture
- 39 biogeographic history and ecological dynamics that determine biome boundaries and plant distributions.
- 40 Grass dominated ecosystems are broadly distributed across all vegetated continents and harbor large
- 41 functional diversity, yet most Land Surface Models (LSMs) summarize grasses into two generic PFTs
- 42 based primarily on differences between temperate C<sub>3</sub> grasses and (sub)tropical C<sub>4</sub> grasses. Incorporation
- 43 of species-level trait variation is an active area of research to enhance the ecological realism of PFTs,
- 44 which form the basis for vegetation processes and dynamics in LSMs. Using reported measurements, we
- 45 developed grass functional trait values (physiological, structural, biochemical, anatomical, phenological,
- 46 and disturbance-related) of dominant lineages to improve LSM representations. Our method is
- 47 fundamentally different from previous efforts, as it uses phylogenetic relatedness to create lineage-based
- 48 functional types (LFTs), situated between species-level trait data and PFT-level abstractions, thus
- 49 providing a realistic representation of functional diversity and opening the door to the development of
- 50 new vegetation models.
- 51

52 Keywords: C<sub>4</sub> photosynthesis, earth system models, land surface models, evolution, grass biogeography,
 53 plant functional types, vegetation models

54

### 55 Main body:

56

#### 57 Introduction

58 Functional trait variation within biomes arises from evolutionary histories that vary biogeographically,

59 leading to plant taxa with differing ecological behavior and differences in ecosystem structure and

function across continents (Lehmann *et al.*, 2014; Higgins *et al.*, 2016; Griffith *et al.* 2019). Land

61 Surface Models (LSMs), fundamental components of Earth System Models, typically apply abstracted

62 plant functional types (PFTs; but see Pavlick *et al.*, 2013; Scheiter *et al.*, 2013; Medlyn *et al.*, 2016) to

63 represent physical, biological, and chemical processes crucial for soil and climate-related decision making

- 64 and policy. However, PFTs must generalize across species, and inevitably encapsulate a wide range of
- 65 plant strategies and vegetation dynamics, a demand that contrasts with efforts to investigate nuanced and
- species specific ecological behavior (Cramer *et al.*, 2001; Bonan, 2008; Sitch *et al.*, 2008; Kattge *et al.*,
- 67 2011). Furthermore, PFTs account for only a modest degree of variation in a wide array of functional
- traits, ranging from seed mass to leaf lifespan (LL), in the TRY database (Kattge *et al.*, 2011). For

69 example, standard PFTs may not generally capture key drought responses in tree species (Anderegg,

- 70 2015), although models with a hydraulics module can be specifically applied for this purpose (e.g.,
- 71 ecosys; Grant et al., 1995). Oversimplification of the physiognomic characteristics of PFTs can have
- 72 major unintended consequences when simulating ecosystem function (Griffith et al., 2017), such as highly
- 73 biodiverse savanna ecosystems (Searchinger *et al.*, 2015). However, studies that explicitly incorporate
- recies-level trait variation into vegetation models (e.g., Grant et al., 1995; Sakschewski et al., 2016; Lu
- *et al.*, 2017; Grant *el al.*, 2019; Mekonnen *et al.*, 2019) have demonstrated improvements in model
- 76 performance. Selecting trait data from multi-variate trait distributions for model parameterization (Wang
- *et al.*, 2012; Pappas *et al.*, 2016) is very challenging for global modeling applications, particularly in
- 78 hyper-diverse regions like the tropics, and may not be feasible for areas with biased or limited data. Until
- 79 these data-gaps are filled, a finer-grained representation of the functional diversity among species might
- 80 be achieved by reorganizing PFTs based on tradeoffs and evolutionary relatedness.
- 81

82 Importantly, in seeking approaches to restructure PFTs, numerous observations over the last decade have 83 shown that both plant traits and biome-occupancy are commonly phylogenetically conserved, with closely 84 related species having similar traits and niches (e.g., Cavender-Bares et al., 2009, 2016; Crisp et al., 2009; 85 Liu et al., 2012; Donoghue & Edwards, 2014; Coelho de Souza et al., 2016). The existence of strong 86 evolutionary constraints on plant functioning and distribution suggests that, as an alternative, vegetation 87 types should be organized in a manner consistent with phylogeny. Eco-evolutionary models have 88 increased our mechanistic understanding of ecological patterns in fields ranging from community ecology 89 (e.g., Webb et al., 2002; Cavender-Bares et al., 2009) to global biogeography (e.g., the Latitudinal 90 Diversity Gradient; Pontarp et al., 2019; Visser et al. 2014). We advocate for explicit inclusion of 91 evolutionary history and a consistent framework for integrating traits into global vegetation models. This 92 approach brings a testable method for defining vegetation types, enables the functional traits of 93 uncharacterized species to be inferred from relatives, and allows evolutionary history to be explicitly 94 considered in studies of biome history. Here, we illustrate this approach for grasses and grass-dominated 95 ecosystems, where we use our framework to aggregate species into Lineage-based Functional Types 96 (LFTs) to capture the species-level trait diversity in a tractable manner for large-scale vegetation process 97 models used in LSMs. Capturing the evolutionary history of woody plants is also critical to understanding 98 variation in ecosystems function in savannas (Lehmann et al., 2014; Osborne et al., 2018), and in general 99 we are advocating for the development of LFTs in other vegetation types and in other ecosystems. 100 Grasses provide a tractable demonstration for the utility of LFTs; we also discuss the potential to 101 significantly improve ecological and biogeographical representations of other plants in LSMs. 102

103 Grasses are one of the most ecologically successful plant types on earth (Linder *et al.*, 2018) and provide 104 great opportunity for increasing understanding of plant functional diversity. Ecosystems containing or 105 dominated by grasses (i.e., temperate, tropical, and subtropical grasslands and savannas) account for 106 a>40% of global land area and productivity, and are a staple for humanity's sustenance (Tilman *et al.*, 107 2002; Still *et al.*, 2003; Asner *et al.*, 2004; Gibson, 2009). The photosynthetic pathway composition ( $C_3$  or 108  $C_4$ ) of grass species is a fundamental aspect of grassland and savanna function, ecology, and 109 biogeography. Of the  $\sim 11,000$  grass species on Earth, some  $\sim 4,500$  use the C<sub>4</sub> photosynthetic pathway 110 (Osborne *et al.*, 2014). Although they account for less than 2% of all vascular plant species (Kellogg, 111 2001), C<sub>4</sub> grasses are estimated to account for 20-25% of terrestrial productivity (Still et al., 2003), 112 having risen to such prominence only in the last 8 million years (Edwards et al., 2010). Dominance by 113  $C_4$  versus  $C_3$  grasses has major influences on gross primary productivity and ecosystem structure and 114 function (Still et al., 2003) and strongly influences interannual variability of the global carbon cycle, due 115 to a combination of ecological and climatic factors (Poulter et al., 2014; Griffith et al., 2015). Dynamic 116 vegetation models largely fail to reproduce spatial patterns of grass cover —both past and present— 117 and productivity at regional to continental scales, limiting ability to predict future plant community 118 changes (Fox et al., 2018; Still et al., 2018). As a consequence, LSMs require significant improvement to 119 adequately represent vegetation responses to increasing CO<sub>2</sub> (Smith et al. 2016; De Kauwe et al., 2016). 120 Many models also miss key transitions between biome states (e.g., Still et al., 2018) that exist as a result 121 of disturbance or biogeographic history (e.g., Staver *et al.*, 2011; Dexter *et al.*, 2018). 122

123 Most LSMs classify grasses into two PFTs based on differences between temperate C<sub>3</sub> grasses and sub-124 tropical and tropical  $C_4$  grasses. However, grass ecological adaptations and physiological properties are 125 highly diverse, ranging from cold-specialized to fire- and herbivore-dependent species. While grasses are 126 often equated functionally, in reality they exhibit a high degree of variation in hydraulic, leaf economic, 127 and phenological traits (Taylor et al., 2010; Liu et al., 2012) that likely explains their broad geographic 128 dominance in different regions (Edwards et al., 2010; Visser et al., 2014). These differences include 129 economically important forest-forming grasses such as bamboos, although here we focus on globally 130 dominant herbaceous lineages. Grasses exhibit strong phylogenetic diversity in leaf economics variation 131 and associations with disturbance (Taylor et al., 2010; Liu et al., 2012; Simpson et al., 2016). 132 Disturbances such as fire and herbivory have large impacts on ecosystem function and distributions, and 133 PFT based approaches are unlikely to capture these differences among lineages. At broad phylogenetic 134 and spatial scales, niche and biome conservatism of major plant lineages is common (Crisp *et al.*, 2009; 135 Cornwell et al., 2014; Donoghue & Edwards, 2014), and we therefore argue that evolution and 136 biogeography provide a framework for aggregating species (across ecosystems and strata) into LFTs that

- 137 capture species-level trait diversity in a way that can be feasibly incorporated for use in global vegetation
- 138 models, and that will improve PFT-based modeling approaches. Focusing on grasses, we developed this
- approach by collecting grass traits from databases (e.g., Osborne *et al.*, 2011) and literature (e.g.,
- 140 Atkinson et al., 2016; Supplemental Appendix S1), for five key categories (physiology, structure,
- 141 biochemistry, phenology, and disturbance). We summarize these species traits at the lineage level and
- relate these functional types to their observed global distributions.
- 143

# 144 Methods for establishing lineage-based functional types (LFTs) for grasses

145 There are 26 monophyletic C<sub>4</sub> lineages described in the Poaceae family, yet only two (the Andropogoneae 146 and Chloridoideae) account for most of the areal abundance of  $C_4$  grasses globally (Lehmann *et al.*, 2019) 147 ; Fig 1.) (Edwards & Still, 2008; Edwards et al., 2010; Grass Phylogeny Working Group II, 2012). 148 Among C<sub>3</sub> grasses, only the Pooideae are globally dominant today. The Pooideae occupy cooler climates 149 than the C<sub>4</sub> Andropogoneae and Chloridoideae, which dominate in warm and wetter and drier climates, 150 respectively. Therefore, we focused on collecting species-level trait data from the literature and from 151 databases for grass species from these three lineages. The term 'trait' is defined differently across 152 research disciplines (Violle et al., 2007). Our aims necessitate a collection of broad trait space beyond 153 that typically used for the leaf economic spectrum to include morphological and physiological 154 determinants of plant hydraulics, physicochemical controls of photosynthesis, allocation to reproduction, 155 and spectral reflectance. Many traits are highly correlated, reflecting plant functional strategies. Further, a 156 single trait can relate to multiple forms of plant fitness. Here, traits were assigned to groups (Table 1) 157 based on their use in models and how they might be used in future applications (e.g., hyperspectral remote 158 sensing of LFTs, or modeling of fire). We present median and variation in trait values among-species for 159 three major grass lineages (LFTs) as per Figure 1, and compare these with commonly used values for  $C_3$ 160 and C<sub>4</sub> PFTs (Table 1).

161

# 162 *LFTs for grasses differ drastically in key functional traits*

163 Our LFTs demonstrate both the importance of considering lineage to explain ecological patterning, and 164 the need for modification of current LSM PFT approaches. For instance,  $C_4$  plants typically have lower 165 RuBisCO activity ( $V_{cmax}$ ) but higher electron transport capacity ( $J_{max}$ ) than  $C_3$  plants, reflecting both the 166 additional energetic cost of  $C_4$  physiology and the greater efficiency of RuBisCO in higher CO<sub>2</sub> 167 environments (Collatz *et al.*, 1998). The Chloridoideae ( $C_4$ ) grasses have intermediate Vcmax and  $J_{max}$ 

- 168 compared to the Andropogoneae  $(C_4)$  and the Pooideae  $(C_3)$  (Table 1). Furthermore, the Pooideae have
- 169 evolved to tolerate much colder conditions (reflected in Trange; Sandve & Fjellheim, 2010; Vigeland et
- 170 *al.*, 2013; McKeown *et al.*, 2016), and our results suggest that C<sub>4</sub> lineages may differ in their thermal

171 tolerances (Watcharamongkol *et al.*, 2018). These differences suggest that macroecological synthesis

172 studies with global implications (e.g., Walker et al., 2014; Heskel et al., 2016) should, at minimum,

include more grass species in their datasets, ideally organized as LFTs.

174

175 Trade-offs among adaptations and tolerances in natural systems promote coexistence among plant species 176 (Tilman, 1988; Tilman & Pacala, 1993; Kneitel & Chase, 2004). Specific leaf area (SLA) measures the 177 cost of constructing a leaf, which represents a tradeoff between acquisitive (high relative growth rate) and 178 conservative (high leaf lifespan) plant strategies (Westoby, 1998; Westoby et al., 2002; Wright et al., 179 2004). Model simulations of growth are highly dependent on the value of SLA used (Korner, 1991; Sitch 180 et al., 2003; Bonan, 2008). However, in most of these LSMs, C<sub>3</sub> grass PFTs have higher or similar SLA 181 values as  $C_4$  PFTs likely biasing predictions. In contrast, we found that the  $C_4$  LFTs had higher SLA than 182 the C<sub>3</sub>LFT, but SLA did not differ between the two dominant C<sub>4</sub> grass lineages (Atkinson et al. 2016). 183 SLA can be highly variable within lineages in grasses, likely due to the importance of herbivore pressure 184 as a competing demand on leaf economics (Anderson et al., 2011; Griffith et al., 2017) as well as 185 intraspecific variation. As a result, SLA highlights that some traits are harder to generalize than others 186 using the LFT approach, and suggests that a range of values may be appropriate than a single value for 187 constraining LFT parameters. The phylogenetic signal among grass lineages is stronger for stature (Taylor 188 et al., 2010; Liu et al., 2012), with the Andropogoneae being considerably taller on average than the 189 Chloridoideae. This difference suggests that not all traits are oriented along a fast-slow axis at broad 190 taxonomic scales across C<sub>3</sub> and C<sub>4</sub> grass lineages (Reich, 2014; Díaz et al., 2016; Archibald et al. 2019). 191 Furthermore, the  $C_3$ - and eudicot-centric approach in the current leaf economics framework suggests that 192 a higher SLA should also correlate with a higher specific leaf nitrogen content, yet the evolution of  $C_4$ 193 photosynthesis allows for a significant reduction in RuBisCO content, and hence plant nitrogen 194 requirements (Taylor et al., 2010). Thus, grass lineages differ in numerous leaf traits which have 195 consequences that extend from palatability and flammability to hydrological differences. 196

197 Physiological and morphological leaf vascular traits underlie variation in SLA, constrain the hydrology of 198 plants (e.g., Blonder *et al.*, 2014; Sack *et al.*, 2014), and are key traits related to the evolution of  $C_4$ 

199 photosynthesis (Sage, 2004; Ueno, 2006). We describe next key hydraulic differences between the two

dominant  $C_4$  lineages, which correspond to the  $C_4$  biochemical subtypes (Ueno, 2006; Liu & Osborne,

201 2015). The Chloridoideae have low conductance and high embolism resistance hydraulic traits (Table 1),

and tend to inhabit drier sites (Fig. 1). Some Andropogoneae have been described as "water spenders"

203 (Williams et al., 1998), and their hydraulic traits help to explain their affinity with higher rainfall habitats

where they rapidly expend available soil water (Taub, 2000) and promote fire after curing. These

hydraulic differences should have large effects in models, especially those that consider tree-grass
coexistence (Higgins *et al.*, 2000) and explicit representation of plant hydraulics (Grant *et al.*, 1995;

207 Mekonnen *et al.*, 2019).

208

209 Lineages also differ in biogeochemical traits that influence nutrient turnover rates and the reflectance and 210 absorbance properties of vegetation. For example, Andropogoneae have higher C:N than Chlordoideae 211 grasses, likely a result of growth rate differences and the frequent association of Andropogoneae grasses 212 with fire. Similarly, a greater proportion of N in Chloridoideae leaves is allocated to RuBisCO, which is 213 related to Vcmax (Ghannoum et al. 2012). Finally, C<sub>3</sub> and C<sub>4</sub> grasses are distinguishable spectrally at the 214 leaf, canopy, and landscape level based on differences between the functional types in chlorophyll a/b 215 ratio, canopy structure, and seasonality (Foody & Dash, 2007; Siebke & Ball, 2009; Irisarri et al., 2009). 216  $C_3$  and  $C_4$  grasses are typically given many of the same optical properties in vegetation models, but we 217 show here that Chloridoideae might have considerably higher near infra-red (NIR) reflectance than other 218 lineages, possibly producing interesting optical variation and affecting the surface energy balance and 219 albedo (Ustin & Gamon, 2010)(Table 1). Foliar spectral traits are also correlated with morphological and 220 chemical traits related to nutrient cycling and plant physiology (Dahlin et al., 2013; Serbin et al., 2014).

221

222 Grass lineages also show key differences in reproductive traits and the timing of related biological events 223 (e.g., leaf-out times) that should be captured in models, especially those that include demographic 224 predictions (Davis et al., 2010). Chloridoideae grasses have seeds with lower mass than other lineages 225 (Liu et al., 2012; Bergmann et al., 2017), and this may represent a life-history trade-off with higher seed 226 production and other 'fast' growth strategies (Adler *et al.*, 2014). Wind versus animal dispersal strategies 227 might also affect diaspore size in a way not directly related to disturbance (e.g., Westoby 1998; Bergmann 228 et al., 2017), whereas some reproductive traits may also indicate fire and disturbance-related adaptations. 229 Phenological traits, such as flowering and leaf-out times and their cues (which can include disturbance 230 factors) exhibit conservatism across many plant lineages (Davies et al., 2013). Fire and herbivory are two 231 globally important and contrasting disturbances for grass-dominated vegetation (Archibald & Hempson, 232 2016; Archibald et al., 2019) and adaptations to both can be characterized by phenological and 233 reproductive traits in addition to physiological and leaf traits. It is less clear how herbivory effects can be 234 captured in such models, given that many herbivore-related traits vary greatly in grasses (Anderson *et al.*, 235 2011). Many fire-related traits show patterns of phylogenetic conservatism, with high flammability 236 clustering into particular lineages such as the Andropogoneae (Simpson et al., 2016). Large-scale 237 vegetation models that have simulated grass fires in Africa have attributed faster curing (becoming dry

fuel) rates to C<sub>4</sub> vegetation (Scheiter *et al.*, 2012), and this behavior appears to be due largely to dominant
Andropogoneae grasses.

240

241 We have identified large differences among LFTs, across six trait categories, that are not captured by the 242 standard PFT approach. Many of these trait data have very low sample sizes (from 1 to 1365) and come 243 from non-overlapping species, highlighting the need for systematic data collection for grasses. Such a 244 data collection effort would be an excellent opportunity to test for coordination among trait axes in a 245 phylogenetic context, which has rarely been done in other systems despite the likelihood that relatedness 246 drives patterns of trait covariation (e.g., Salguero-Gómez et al., 2016; Griffith et al., 2016). Furthermore, 247 intra-group (whether PFTs or LFTs) trait variation deserves to be properly estimated (only some traits in 248 Table 1 have enough data to estimate variability) as convergence and adaptation produce meaningful trait 249 variation that should be incorporated into models.

250

# 251 Potential for lineage-based functional types in other vegetation types

252 Many current PFTs implicitly represent groupings of closely related lineages (e.g., pinaceous conifers, 253 grasses). However, even in these cases biogeographic distributions, and the coarseness of the 254 phylogenetic unit, generates a lack of useful resolution. Currently, there are efforts to incorporate species-255 level trait data and methods such as those proposed by Cornwell et al., (2014) could be employed to 256 cluster species into prominent lineage-based groupings representing unique trait combinations. 257 Phylogenies are hierarchical by nature and allow the LFT approach to be scalable and adjustable to the 258 research question being addressed. While many technical challenges still remain, the ability to remotely 259 sense plant lineages adds potential for rapidly developing LFTs from spectral data (e.g., Cavender-Bares 260 et al., 2016). LFTs would be valuable for a wide range of systems. For example, trees in Eurasian boreal 261 forests suppress canopy fires through the structure of their canopies, whereas North American boreal trees 262 enable greater intensity canopy fires (Rogers et al., 2015). These distinctions lead to major differences in 263  $CO_2$  emissions and function (Rogers *et al.*, 2015) that might be captured in an LFT framework. The 264 boreal tree example is challenging because these communities are comprised of closely related species 265 that are ecologically different, potentially requiring species level parameterization or being better 266 represented by fire-based PFTs. Secondly, LFTs for savanna tree communities could better represent 267 differing climatic responses that are driven by unique evolutionary and biogeographic histories (Lehmann 268 et al., 2014; Osborne et al., 2018). Finally, tropical ecosystems such as the dipterocarp forests in 269 Southeast Asia would be well suited to LFTs which might better represent carbon storage (Brearley *et al.*, 270 2016).

- 272 Potential challenges with a lineage-based functional approach include the fact that many plant traits do 273 not show strong phylogenetic conservatism (Cadotte *et al.*, 2017), with several being labile. There are 274 likely spatial and phylogenetic scales at which the LFT approach will be most appropriate; for example, at 275 large scales (regional to continental), lineage conservatism is common (Crisp et al., 2009). In contrast, at 276 the scale of local communities, we might expect character displacement and limiting similarity (processes 277 that lead to reduced trait similarity of coexisting species) could obscure phylogenetic patterns and limit 278 the utility of LFTs as proposed here (Webb et al., 2002; Cavender-Bares et al., 2009; HilleRisLambers et 279 al., 2012). However, in grassy ecosystems, there is evidence that the patterns of spatial ecological sorting 280 of lineages would be captured with LFTs also at landscape scales (e.g., within Serengeti National Park, 281 Anderson *et al.*, 2011; Forrestel *et al.*, 2017). Finally, we focus on extant lineages that are functionally
- important today, but their past interactions with other clades may have shaped the biomes they inhabit(Edwards *et al.*, 2010).
- 284

# 285 Conclusions

- 286 We conclude that an LFT perspective captures important variation in functional diversity for grasses 287 (Table 1). Our analysis of current knowledge of grass functional diversity (in terms of physiology, 288 structure, biochemistry, phenology, and disturbance), distributions, and phylogeny indicates that to 289 represent grass ecological behavior, division of today's ecologically dominant grasses into at least two  $C_4$ 290 and at least one  $C_3$  LFT could potentially improve representation in LSM and models. These proposed 291 LFTs capture key evolutionary differences in physiological, structural, biogeochemical, anatomical, 292 phenological, and disturbance-related traits. We also highlight the need for systematic trait data collection 293 for grasses, which we show are vastly underrepresented in trait databases, despite their ecological and 294 economic importance. More broadly, we outline the LFT framework which is highly flexible and has the 295 potential for use in a wide range of applications. Here, we speak to incorporating LFTs as groupings in 296 vegetation models, but we also suggest that trait-based models might capture important biogeographic 297 variation (e.g., due to historical contingency) through the inclusion of phylogenetic conservatism. We 298 advocate for the use of phylogeny as a way to help guide and constrain the inclusion of burgeoning plant 299 trait data to expand the range of functional types considered by global vegetation models.
- 300

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**Table 1.** Common PFT parameters from LSM models, and median LFT parameters (IQR; interquartile
range in parentheses, where calculable) for three dominant grass lineages, taken from the literature and
trait databases. Lineage assignments are based on Osborne *et al.* (2014). The table shows a subset of

596 common parameters, with up to five parameters from each of six major categories. Blank values in the

597 PFT/LFT columns signify parameters that are not typically included in LSM models but are

598 potentially important for accounting for the ecological behavior of grasses. Bolded numbers with letters

599 (i.e., a compact letter display; sharing a letter [a, b, c] indicates no difference) indicate significant

600 differences with a Tukey's test from simple linear model fits when all three lineages had at least three

data points. Sources are in table footer.

602

		PFT		_	LFT*		
Category	Parameter	C <sub>4</sub>	C <sub>3</sub>	Source	Andropogoneae	Chloridoideae	Pooideae
Physiological	Vcmax ( $\mu mol \ m^{-2} \ s^{-1}$ )	39	90	1, 2	38	45.6 (4.4)	63.6 (28)
	Jmax ( $\mu mol \ m^{-2} \ s^{-1}$ )	400	100	3	180	108.1 (43)	128.8 (45)
	Rd ( $\mu mol\ m^{-2}\ s^{-1}$ )	0.8	1.1	1, 2	0.9a (0.2)	2a (1.4)	0.9a (0.7)
	Phi ( $\mu mol \ \mu mol \ ^{-1})$	0.06	0.085	4, 5	0.06	0.06	0.09
	Trange (°C)	< 15.5	> 15.5	6	> 5**	> - 5**	> -30 and < 5**
Structural	SLA ( $m^2 kg^{-1}$ )	16	33	7	33b (11)	29b (14)	25a (12)
	LDMC $(gg^{-1})$	-	-		-	-	-
	SRL ( $m g^{-1}$ )	-	-		-	-	-
	Culm Height (cm)	-	-		150c (150)	80b (50)	60a (60)
	R:S ( <i>99</i> <sup>-1</sup> )	2	2	7	0.4b (0.07)	0.3a (0.07)	0.4b (0.1)
Biogeochemical/	C:N ( <i>99</i> <sup>-1</sup> )	10	17	7	66.1b (14.7)	39.9a (22.2)	55.7ab (10)
spectral	Nrubisco (proportion)	0.09	0.137	7	0.05 (0.01)	0.08 (0.03)	0.2
	Reflectance (300–2500 nm)	-	-		-	-	-
	e.g., rNIR (reflectance)	0.35	0.35	7	0.4	0.5	0.3
Anatomical/	IVD ( <i>µm</i> )	-	-		85.7a (25.2)	136.8b (40)	242.1c (58)
Hydraulic	Xylem dia. (µm)	-	-	7	21.4b (12.2)	16.8a (10.7)	19.3a (6.7)
	K <i>leaf</i>	-	-		-	-	-
	Vein Hierarchy ( $Primary \ vein \ secondary \ vein \ n^{-1}$ )	-	-		0.11a (0.09)	0.29b (0.2)	0.58c (0.6)
	Leaf Width:Length	-	-		0.04b (0.05)	0.03a (0.04)	0.03a (0.02)
Life History	LL (months)	1.68	12	7	2 (0.4)	1.1	2 (1.8)
	1000-seed mass $(g)$	-	-	7	1.4b (2.4)	0.2a (0.4)	1.4b (3.8)
	Life History (% annual)	-	-	7	0.25	0.28	0.15
Disturbance	Curing rate (%)	80	20	8	80	50**	20
	Bud Bank	-	-		-	-	-
	Flammability ( $gs^{-1}$ )	-	-		-	-	-

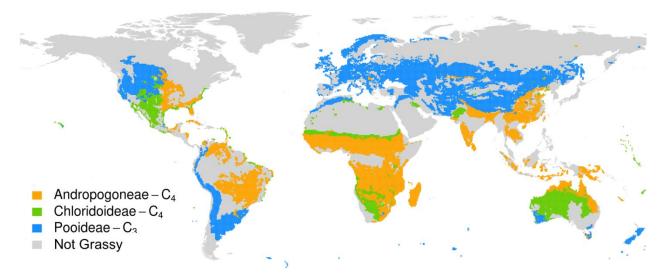
603

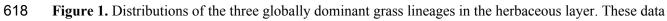
604

605 1 - Farquhar *et al.* (1980), 2 - Collatz *et al.* (1992), 3 - von Caemmerer

- 606 (2000), 4 Ehleringer et al. (1997), 5 Collatz et al. (1998), 6 Sitch et al. (2003), 7 Oleson et al.
- 607 (2013), 8 Scheiter *et al.* (2012); Curing rate is the % cured 30 d after the end of the growing season as
- described in Scheiter *et al.* (2012); \*Published citations for LFT values can be found in Appendix S1.
- 609 Anatomical data come from Gallaher et al. in prep. \*\*Estimated value. Abbreviations: Vcmax (maximum
- 610 carboxylation rate), Jmax (light saturated rate of electron transport), Rd (dark Respiration), Phi (quantum
- 611 efficiency), SLA (Specific Leaf Area), LDMC (Leaf Dry Matter Content), SRL (Specific Root Length),
- 612 R:S (root to shoot ratio), C:N (Carbon to Nitrogen ratio), IVD (InterVeinal Distance), Kleaf (leaf
- 613 hydraulic conductance), LL (Leaf Lifespan).
- 614

# 616 Figures:





- 619 come from Lehmann et al (2019), and show where each lineage is more abundant than the other two
- 620 lineages on a 0.5 degree grid.