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1 **The UK Environmental Change Network after twenty years of integrated ecosystem**
2 **assessment: key findings and future perspectives**

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9 **Highlights**

- 10 • The UK Environmental Change Network has been in operation for 20 years
11 • The ECN addresses the causes and consequences of environmental change
12 • The range of papers in this issue illustrates wide applicability of the network
13 • Tighter integration with complementary programmes is highly desirable
14 • Evolving environmental challenges will require network development and adaptation

15

16 **Abstract**

17 The UK Environmental Change Network (ECN), the UK's Long-Term Ecosystem Research
18 (LTER) network, has now been operating for over twenty years. It was established in 1992 as
19 a set of terrestrial sites at which sustained observations relevant to a range of ecological
20 indicators and environmental parameters could be made. An additional ECN freshwater
21 network was launched in 1994. In this paper we provide a brief history of the network, and
22 describe its current structure and role within a complementary wider range of UK
23 environmental monitoring and observation programmes that are either more focussed on
24 specific parameters or habitats, or operate at different temporal and spatial scales. We then
25 provide a review of the other papers within this Special Issue, which exemplifies the broad
26 range of environmental concerns that ECN data and sites are helping to address. These
27 include network-wide summaries of environmental and biological trends over the first two
28 decades of monitoring, more site-specific assessment of the ecological impacts of local
29 pressures resulting from changes in management, biological and ecosystem service indicator
30 development, and the testing of new monitoring technologies. We go on to consider: (i)
31 future directions of network development and adaptation in light of recently emerging
32 environmental concerns, dwindling financial resources and the consequent need for greater
33 efficiency; (ii) the desire for tighter integration with other monitoring and observation
34 programmes both nationally and internationally; (iii) opportunities raised by recent
35 technological developments; and (iv) the need to process and make available data more

36 rapidly to increase the capacity of ECN sites as early warning systems. In its first two decades
37 of operation the ECN has accumulated a robust set of baseline data that describe
38 environmental and biological variability across a range of habitats in unprecedented detail.
39 With appropriate, informed development, these should prove invaluable in discerning the
40 causes and consequences of environmental change for decades to come.

41 Keywords: Long-term monitoring; ECN; LTER; Climate change; Air pollution; Indicators

42

43 **1. Introduction**

44 The diverse range of ecosystems that characterise the non-urban environment serve a host
45 of vital functions that underpin human health and wellbeing. A comprehensive overview of
46 the value and changing state of the UK's ecosystems, conducted by the UK Natural
47 Ecosystem Assessment in 2009-2011 (UK NEA, 2011), concluded that around 30% of these
48 'ecosystem services' were in a state of decline, while several others, including functions
49 provided by soils and wild species diversity, were in a reduced or degraded state. The UK
50 NEA identified a range of pressures, often acting in concert, contributing to this degradation,
51 including urbanisation, intensive agriculture, pollution and climate change. It argued that the
52 benefits derived from ecosystems were currently undervalued in an economic sense and that
53 future sustainable development would require their true worth to be taken into account in
54 any decision-making process. Among its conclusions the NEA states: 'In order to refine our
55 understanding of the fundamental ecosystem processes underpinning the delivery of
56 ecosystem services we need to both extend our observations and experimental
57 manipulations, and also improve our models of the key mechanisms.'

58 Of course, concern over the decline in the extent and quality of natural environments and
59 resources extends back well before the recent articulation of concern over ecosystem
60 services. Researchers, conservationists and decision-makers have long recognized the
61 importance of understanding processes that determine how ecosystems function, how and
62 why they may be changing, and how resilient they are to both short term and more
63 sustained disturbances. This need is served through a continuously evolving scientific
64 understanding founded on inter-dependent disciplines of repeated observation, controlled
65 experiments and process-based and mathematical modelling (Parr et al., 2003).

66 The first of these disciplines requires precise measurements to be made over much longer
67 periods of time than may be covered by standard scientific research grants (typically of 3-5
68 years duration). Indeed, strong natural variation in climate at a range of frequencies dictates
69 that several decades of data may be necessary to identify and quantify some key underlying
70 trends in the environment resulting, for example, from climate change or changing pollution
71 emission policy. The establishment and maintenance of systems of consistent, repeated
72 measurements over long periods, therefore, requires a clear long-term vision, sustained and
73 often substantial investment and the patience of funding bodies – particularly during the

74 early years of operation. Consequently, robust long-term environmental monitoring and
75 research programmes in the UK are relatively few and vary with respect to scope, structure
76 and funding models.

77 The most long standing and scientifically valuable UK monitoring or observation
78 programmes range from the series of occasional but measurement-rich GB Countryside
79 Surveys (CS) (Carey et al., 2008), to higher frequency – but often more measurement or
80 habitat-specific – programmes such as the UK Acidifying and Eutrophying Atmospheric
81 Pollutants network (UKEAP), the UK Upland Waters Monitoring Network (UWMN; Battarbee
82 et al., 2014), and the UK Butterfly Monitoring Scheme (UKBMS). These are complemented
83 further by extensive freshwater monitoring networks operated by national agencies to
84 ensure compliance of water quality standards, and a wealth of observations of species
85 occurrence, often made by amateur experts and managed by the Biological Records Centre
86 (Pocock et al., 2015), that inform particularly on changing species distributions. This
87 conceptual gradient of largely complementary programmes, from broad-scale, spatially
88 extensive, low frequency measurement to narrower, more site-focussed and measurement-
89 intensive observation can be represented in schematic form as a pyramid or triangle (Figure
90 1). The UK Environmental Change Network (ECN; www.ecn.ac.uk) fills a particular niche
91 toward the apex, with its emphasis on multi-disciplinary site-focussed monitoring.

92 Established in 1992 as a network of a relatively small number of instrumented sites, the ECN
93 spans a wide range of ecosystems at which sustained long-term observations of selected
94 physical, chemical and biological variables are made according to tightly defined protocols.
95 The resulting quality controlled data are lodged in a central database (Rennie, this issue)
96 which is openly available. The ECN originally comprised eight terrestrial sites, but four more
97 terrestrial sites joined the network during the period 1993-1999, bringing the total to 12
98 (Table 1). In 2014 ECN monitoring ceased at Drayton, reducing the number to 11. A
99 freshwater ECN network was established in 1994 and currently comprises 45 freshwater sites
100 (lakes, rivers and streams). The map (Figure 2) shows the locations of ECN sites.

101 The ECN is operated by a consortium of organisations (see ECN website, www.ecn.ac.uk, and
102 acknowledgements for details) who contribute variously to the provision of monitoring sites,
103 sampling/recording and analytical chemistry services. Separate partner organisations are
104 responsible for the management of individual ECN terrestrial sites. They provide data to a
105 small team at the NERC Centre for Ecology & Hydrology, who coordinate and promote the
106 network and manage the ECN's central database and website. Rennie (this issue) describes in
107 more detail the ECN's approach to – and recent developments in – field data capture, data
108 management and provision of data access. ECN datasets are freely available for use under
109 licence for research and educational purposes, and have supported a wide range of research
110 to date. A list of publications based on research using ECN data or taking place at ECN sites
111 is maintained on the ECN Data Centre website (ECN).

112 The ECN Data Centre team also manage data generated by related networks including the
113 UK Environmental Change Biodiversity Network (ECBN), the UK Lake Ecological Observatory
114 Network (UKLEON) and the UWMN (chemistry database). At each ECN site, measurements
115 are made and samples collected according to published protocols (Sykes & Lane, 1996;
116 Sykes et al., 1999). The indicators monitored (Table 2) were selected to provide an integrated
117 suite of driving, state and response variables, to enable relationships to be tested, whilst at
118 the same time taking into consideration issues such as cost, practicality and safety (Sykes &
119 Lane, 1996).

120 Several sites in the network were already important platforms for monitoring and research
121 before the onset of the ECN. For example: Rothamsted is the oldest continually functioning
122 agricultural research station in the world, and home to the Sir John Lawes's 'Classical
123 Experiments' (e.g. Silvertown et al., 2006), started between 1843 and 1856; Moor House-
124 Upper Teesdale and Wytham contributed to the Tiger Programme (Cummins et al., 1995);
125 Moor House, along with Snowdon and Cairngorms, also participated in the International
126 Biological Programme (IBP; Heal and Perkins, 1978); Oxford University owned Wytham
127 Woods have also been the focus for internationally recognised research over many decades
128 (Perrins et al., 2010), while Alice Holt is one of three Research Forests equipped with an eddy
129 covariance tower for studying carbon dynamics. Indeed, most terrestrial ECN sites were
130 selected primarily on the basis of their long histories of environmental monitoring and
131 research. Consequently, while ECN terrestrial sites cover a broad spectrum of UK habitats
132 (Dick et al., 2011), they are relatively few in number and have relatively limited power, in
133 isolation from the wider available evidence base, to inform on environmental change at a UK
134 scale.

135 Regardless of the spatial extent of the network, however, the ECN is nationally unique with
136 respect to the range of high frequency physical, biogeochemical and biological
137 measurements that are made in close proximity. This provides unparalleled opportunities to
138 directly link pressures and responses associated with long-term environmental change over
139 various timescales. For example, bi-weekly measurements of soil water chemistry can be
140 linked with weekly measurements of precipitation chemistry to assess the impact of
141 reductions in the emissions of acidic pollutants to the atmosphere on soil acidity, while the
142 wider impacts of emissions control on these ecosystems can be examined with reference, for
143 example, to vegetation assemblages. Similarly, the potential risks to biodiversity posed by an
144 increasingly erratic climate can be investigated using weather data specific to the location of
145 butterfly, moth and carabid beetle population assessments. The latter may be particularly
146 useful with respect to assessing the impact of changes in the frequency and intensity of
147 precipitation or drought, effects of which may be too localised to be represented by spatially
148 extrapolated meteorological data, and these in turn may shed light on drivers behind
149 national-scale temporal patterns.

150 Arguably, however, the full potential of the ECN in the assessment of the causes and
151 consequences of environmental change is realised only when data and observations from
152 this site-focussed network are integrated within a more spatially extensive 'network of
153 networks' comprising other national and international monitoring and observation systems.
154 Some elements required to facilitate this are already in place. Thus several of the monitoring
155 protocols are shared with other UK networks, including the Rothamsted Insect Survey
156 (Rothamsted Research), the UKBMS and the Ammonia Network (Defra), while six ECN
157 freshwater sites are drawn from the UWMN. ECN sites contribute to these other
158 programmes, while patterns of change at individual ECN sites are providing fine temporal-
159 scale context for observations made by other systems. For example high frequency ECN
160 vegetation measurements have been used by the CS to determine the extent to which
161 botanical measurements made in specific survey years may have been influenced by atypical
162 conditions such as very wet or very dry summers (Scott et al., 2010).

163 The ECN also has an integrative role internationally as the UK's official Long-Term Ecosystem
164 Research (Müller et al. (eds), 2010) network and as a member of the International Long-Term
165 Ecological Research network (ILTER; Kim, 2006) and its European regional component, LTER-
166 Europe (LTER-Europe; Mirtl, 2010). Several ECN sites are included in other international
167 networks. Cairngorms is in the European GLORIA network (alpine vegetation; Gottfried et al.,
168 2012; Pauli et al., 2012), whilst Alice Holt is part of the pan-European ICP Forest Level II
169 network. Among ECN freshwater sites, Windermere (Maberly & Elliott, 2012) and Loch Leven
170 (May & Spears, 2012) are particular focal points for long-term research into lake ecosystem
171 dynamics, while a number of lowland river sites, monitored by the Environment Agency of
172 England and Wales, the Scottish Environmental Protection Agency and the Northern Ireland
173 Environment Agency, are key sites within their Water Framework Directive monitoring
174 networks.

175 Observations and publications based on ECN data inform environmental policy development
176 across a range of disciplines. For example, the ECN soil solution chemistry records provided
177 the primary evidence, reported in the UK Review of Transboundary Air Pollution (RoTAP,
178 2012), for soil chemical responses to reductions in the deposition of acid air pollutants, and
179 evidence for links between the release of dissolved organic carbon from peatlands and both
180 acid deposition (Stutter et al., 2011) and droughts (Clark et al., 2006). The same data were
181 fundamental in recognising that hydrochloric acid deposition had made a significant
182 contribution to soil acidification at a UK scale (Evans et al., 2011) and this is now feeding into
183 a revision of dynamic models used to determine the sensitivity of soil biogeochemistry to
184 long-term changes in acid deposition (e.g. Rowe et al., 2014b). ECN bulk soil chemistry data
185 have also been used to determine long-term changes in woodland soil carbon storage as a
186 consequence of increasing soil horizon depth (Benham et al., 2012). These studies are of
187 particular interest to stakeholders concerned with carbon accounting, natural capital
188 assessments and water quality management.

189 Elsewhere, ECN vegetation data have been central to the development of a new indicator of
190 ecological impacts of nitrogen deposition, as requested by the Convention on Long Range
191 Transboundary Air Pollution (Rowe et al., 2014a), and ECN chemistry data from Rothamsted
192 have enabled Storkey et al. (2015) to identify some of the first indications of recovery in plant
193 biodiversity from the long-term impact of nitrogen deposition. Time series analysis of ECN
194 ecological data (Morecroft et al., 2009) contributed to the development of the UK
195 Biodiversity Report Card (Morecroft & Speakman, 2013) and a UK government report on
196 Biodiversity Indicators of Climate Change BICCO-Net (Pearce-Higgins et al., 2015). ECN
197 invertebrate data also contributed to an assessment of links between climate change and
198 phenological change (Thackeray et al., 2010) that was cited as evidence in the
199 Intergovernmental Panel on Climate Change's recent assessment of Terrestrial and Inland
200 Water Systems.

201 Two decades since the ECN was launched, 20 year time-series have been assembled for the
202 terrestrial network and published under a series of digital object identifiers (DOIs; ECN Data
203 Centre). This Special Issue provides an opportunity to exploit these datasets in order to
204 characterise and quantify some of the key changes that have occurred over this period both
205 at a network level and at individual sites. It also provides examples of how ECN datasets and
206 sites are contributing to the development of clearer process understanding, model testing
207 and development, and the trialling and calibration of new monitoring technologies. In the
208 following section we briefly review the other papers in the issue, focussing particularly on key
209 scientific findings of relevance to management and policy development. In Section 3, we go
210 on to consider, in the light of changing environmental pressures, evolving management and
211 policy needs and dwindling budgets, how the ECN might best continue to develop and
212 adapt in order to remain centrally relevant to the UK's evidence base serving the detection
213 and attribution of environmental change.

214

215 **2. Paper synopses**

216 The breadth of papers in this Special Issue illustrates the versatility of ECN data and ECN sites
217 in addressing a range of issues of concern surrounding the assessment and interpretation of
218 environmental change. They range in scope from the broad-scale quantification of change
219 across sites, to site-specific investigations aimed at developing clearer process
220 understanding, method testing and the development of the ECN data management system.

221 In order to support the wide range of research into environmental change exemplified by the
222 other papers in this issue, it is essential that data are collected, processed, stored and
223 distributed in a systematic and fully traceable manner. The ECN has pioneered the
224 development of an informatics approach to handling and integrating a diverse array of
225 environmental time-series data. The organisation, structure and function of the ECN
226 database, from initial data capture, through quality control and centralised data

227 management, to making data publicly available are presented by Rennie (this issue). The
228 paper also covers the importance of complying with current data and metadata standards,
229 such as the European INSPIRE Directive, and considers recent technological opportunities to
230 receive and manage ever larger volumes of telemetered data, and develop online data
231 exploration interfaces.

232 Background environmental context for several of the following papers is provided by
233 Monteith et al. (this issue) who present an analysis of change and variation in a range of
234 weather and atmospheric deposition indicators measured at terrestrial ECN sites between
235 1993 and 2012. Regional-scale influences of climate change and air pollution are likely to be
236 the most important drivers of environmental change at these sites, since local land
237 management tends not to vary much over time. Perhaps surprisingly, Monteith et al. (this
238 issue) found no evidence of significant change in monthly mean air temperatures over the
239 full time series, but noted a marked increase in precipitation over summer months, which
240 was linked to an unusually prolonged directional shift in the summer North Atlantic
241 Oscillation (NAO). The intensity of extreme precipitation events also increased. More
242 generally inter-annual variability in weather was strongly linked to variation in the NAO, thus
243 illustrating the challenge of separating signals of long-term climate change from shorter
244 term variability. Arguably the most ecologically influential changes in the environment of
245 ECN sites over the past two decades resulted from substantial reductions in acid deposition,
246 and much of the wider UK countryside is likely to have experienced similar trends over the
247 period. Sites experienced large declines in the concentration of sulphate in precipitation,
248 while concentrations of nitrogen species, i.e. nitrate (NO_3^+) and ammonium (NH_4^+), also fell
249 slightly at several sites. Regional scale drivers of environmental change are therefore likely to
250 be responsible for both widespread increases in soil solution pH, reflecting reductions in acid
251 deposition, and wetter conditions over the middle of the growing season.

252 One of the major strengths of the ECN is the co-located measurement of weather, pollutant
253 deposition and soil solution chemistry. The same suite of measurements is also made by the
254 Forest Level II monitoring network. Sawicka et al. (this issue) combined data from the two
255 networks to explore dynamic links between changes in the chemistry of deposition and
256 weather parameters, and soil solution chemistry. They focussed particularly on the dynamics
257 of Dissolved Organic Carbon (DOC), which has been increasing in upland surface waters
258 around the UK. Dissolved Organic Carbon poses a particular problem for the water industry
259 since it has to be removed using costly water treatment processes prior to disinfection by
260 chlorination to prevent the production of potentially toxic bi-products. Non-linear trend
261 analyses were used to characterise the timing of large reductions in sulphur deposition at
262 several sites and the resulting impact on soil chemistry. These changes included rising soil
263 pH and an increase in DOC concentrations in the surface organic layers particularly, thus
264 supporting the rarely tested hypothesis that the widely observed increase in DOC in surface
265 waters (Monteith et al., 2007) has its origins in soil processes. The findings help to clarify
266 process understanding and should benefit the parameterisation of process-based models

267 that are being developed to help water quality managers predict likely future changes in
268 DOC and develop appropriate adaptation and mitigation strategies.

269 Impacts on vegetation of these recent changes in weather, acid deposition and soil acidity, in
270 addition to land management, were considered by Rose et al. (this issue). They analysed
271 trends in vegetation data collected at the 12 terrestrial ECN sites for the period 1993-2012.
272 These data are unique nationally with respect to the unusually high frequency of
273 measurements (annual to tri-annual surveys) – and thus serve as particularly sensitive records
274 of long-term change. They found a network-wide increase in both plant species richness
275 and an ecological indicator characterising species associations with soil pH (Ellenberg R),
276 while species increasing in frequency tended to be characteristic of less acid soils. In some
277 lowland habitats, increasing species richness could also be linked to increased soil moisture
278 availability (perhaps reflecting the trend towards wetter summers (Monteith et al., this issue)),
279 and to a reduction in intensive farming practices (particularly with respect to reductions in
280 the application of nitrogen fertilisers). The apparent positive response of plant species
281 richness to declining soil acidity, which in turn is driven predominantly by reductions in
282 sulphur deposition, draws into question some current assumptions regarding the role of
283 nitrogen deposition in reducing plant diversity, which underpin the national and
284 international setting of critical loads for nitrogen. Plant species richness is known to be
285 strongly negatively correlated with nitrogen deposition and species richness, but as nitrogen
286 and sulphur deposition tend to co-vary spatially, it is feasible that part of this 'effect' is
287 actually attributable to the historical impacts of sulphur deposition that are now waning. The
288 study therefore highlights a clear policy need for a more thorough evaluation of the relative
289 impacts of atmospheric eutrophication and acidification on botanical biodiversity.

290 ECN vegetation data collected at particularly high frequency (i.e. annually) also enabled
291 Morecroft et al. (this issue) to assess the resilience of vegetation by quantifying the extent of
292 inter-annual variation in plant communities. Their findings challenge the commonly held
293 assumption that plant communities change relatively little from one year to the next. They
294 show that the extent of variability was dependent on habitat type. Plant communities
295 associated with low levels of disturbance and low agricultural inputs were the most stable,
296 and, they propose, are therefore more likely to be resilient to gradual environmental
297 changes. They conclude that plant monitoring scheme design needs to take into account the
298 extent of inter-annual variability, in order to correctly identify longer-term trends.

299 Assessment of the impact of changing farming practices on vegetation diversity and
300 productivity, a key issue in the development of food security policy, is investigated in further
301 detail at a single lowland agricultural ECN site by Pallett et al. (this issue). Using vegetation
302 data collected from a long-term experiment at Wytham, which switched from conventional
303 to organic agriculture mid-way through the ECN monitoring period, they demonstrate that
304 the withdrawal of nitrogen-based fertiliser applications resulted in an immediate reduction in
305 grassland productivity while species richness increased by 300%. They argue, therefore, that

306 the study illustrates a clear trade-off, whereby the increase in biodiversity occurs at the
307 expense of productivity, as a consequence of high-yielding nitrogen loving grass species
308 being replaced by a more diverse mix of less productive grasses and forbs. The study brings
309 into clear focus the challenges of meeting potentially conflicting policy agendas.

310 In a further assessment of impacts of lowland agricultural management on biodiversity, Eyre
311 et al. (this issue) investigated land use effects on the spatial and temporal variation in the
312 community structure of carabid beetles. These are an important group of crop pest
313 predators and have been the focus of previous studies using ECN data (e.g. Brooks et al.,
314 2012; Pozsgai & Littlewood, 2014). There is considerable interest in reducing dependence on
315 pesticides by boosting predatory insect abundance using non-crop field margins. The
316 authors focussed on carabid data from two English lowland agricultural research sites in
317 England: the ECN site Drayton in the west midlands and Nafferton Farm in Northumberland.
318 They found carabid activity in non-crop habitats, often used as 'beetle banks' in bio-control
319 management of crop pests, to be sub-optimal, and that the species composition of these
320 environments was influenced by surrounding management activity. They conclude that some
321 management is likely to be required to maximise the potential of these features to
322 contribute to pest control.

323 Milligan et al. (this issue) explored the implications of less intensive upland management on
324 biodiversity, drawing on data (over the period 1954 – 2000) generated by a long-term sheep
325 enclosure experiment at the Moor House ECN site in the northern Pennines. They found that
326 species diversity declined significantly in the sheep-grazed plots over this period, compared
327 with plots that were protected from grazing. The results contrast with those of Rose et al.
328 (this issue) who identified positive trends in vascular species richness across ECN sites in
329 more recent years, but there was little overlap between the two studies and grazing intensity
330 has declined at Moor House since 2000. Nevertheless the study emphasises the vulnerability
331 of upland systems to traditional farming practices and raises issues regarding what may be
332 the most appropriate targets for habitat management.

333 A holistic approach to upland management needs to consider marginal benefits and
334 potential trade-offs to ensure the optimal delivery of ecosystem services. At the Moor House
335 ECN site, in addition to sheep production and biodiversity conservation, this includes carbon
336 sequestration by the peat soils that also deliver high loads of dissolved organic carbon (DOC)
337 to drainage waters. Rising concentrations of DOC in surface waters might be expected to be
338 accompanied by greater losses of carbon from waters to the atmosphere in the form of CO₂
339 and methane following microbiological and physical degradation of dissolved organic
340 matter. The extent of these losses is difficult to quantify but is a potentially important policy
341 parameter with respect to national carbon accounting. Moody et al. (this issue) harnessed
342 the unique co-located long-term measurements of DOC and other solutes in atmospheric
343 deposition, soil water and stream water available for the ECN peatland site, Moor House. The
344 difference between a theoretical soil input flux of DOC (based on estimated DOC

345 contributions from various catchment sources), and the flux measured in stream runoff was
346 taken to represent either a carbon loss to the atmosphere or a gain by the stream,
347 depending on the method applied. Despite the conflicting results, the novel approach,
348 blending co-located soil and surface water time series, showed considerable potential for
349 contributing to our understanding of fluvial carbon dynamics.

350 This Moor House study illustrates the growing demand from policymakers for methods to
351 quantify how the UK's natural capital assets (such as soils, fresh water and biodiversity)
352 contribute to human health and wellbeing, in order to advise environmental decision-
353 making. Consequently, a range of tools has been developed to parameterise the delivery of
354 ecosystem services provided by the natural environment. In recent years, ECN data and the
355 expert knowledge of those responsible for managing individual sites have been used to
356 assess the validity and applicability of some of these methods (Dick et al. (2011)). In this
357 volume, Dick et al, (this issue) go on to explore how recent environmental change at ECN
358 sites translates into changes in ecosystem services using both qualitative and quantitative
359 approaches to ecosystem service assessment. The authors observed a gradual change in the
360 balance of ecosystem service delivery toward cultural services at most sites, associated, for
361 example, with the increased use of land for recreation and education. While the quantitative
362 method was more robust statistically, they concluded that a blend of qualitative and
363 quantitative approaches provided a more holistic picture of long-term trends in ecosystem
364 service delivery.

365 The availability of multiple ecological time series linked to both physical and chemical
366 supporting data provides various opportunities to test existing, and develop new, indicators
367 of environmental and ecological change and resilience. To date, climate change impacts on
368 species and ecosystems in various parts of the world have been clearest with respect to
369 changes in phenology and distribution ranges. However, population sizes and community
370 structure are also likely to be affected. Indicators are therefore required to quantify the
371 extent of these effects in order to inform climate change impact assessments. Martay et al.
372 (this issue) describe a new community-based climate change indicator approach to assess
373 climate impacts on moths and butterflies. Conventional climate change indicators are
374 calibrated using spatial relationships between species distributions and climate. However,
375 these can be very dependent on data being drawn from across wide geographical ranges,
376 and are also based on the assumption that species variations in climatic space can be applied
377 to predict change at individual locations over time. The authors have, therefore, explored an
378 alternative approach that exploits temporal relationships between species abundance and
379 climate. These were then tested on ECN sites where lepidopteran communities and weather
380 parameters are measured in close proximity. The authors found that the approach was
381 effective at predicting spatial and temporal variation in lepidopteran communities at ECN
382 sites but only when models were calibrated at a seasonal scale, thus emphasising the need,
383 in this case, to take seasonality into account.

384 The high frequency and taxonomic detail of several of the ECN datasets also provide
385 excellent opportunities to explore variation (between sites and over time) in community
386 structure, and its importance in underpinning ecological resilience. However, the
387 characterisation and application of ecological networks remains a great challenge for
388 ecologists. Pozsgai et al. (this issue), therefore, explored the potential of a Bayesian Network
389 approach to study interspecific relationships among carabid beetles at two ECN upland sites,
390 Glensaugh and Sourhope. They conclude that Bayesian networks are effective tools for
391 modelling interspecific relationships between carabid species and, given the relative ease by
392 which the necessary field data can be collected, propose that such methods could now start
393 to routinely inform ecological assessments and conservation plans.

394 High quality, long-term time series are essential not only for quantifying environmental
395 change but also to calibrate models that can then be used to predict likely future behaviour,
396 in response to changes in land use or climate change, for example. Elliott et al. (this issue)
397 investigated past changes in cyanobacteria and nutrient chemistry in the UK's largest lake,
398 Lough Neagh, in Northern Ireland. One of several lakes on the ECN freshwater network, the
399 water quality and biota of Lough Neagh has been severely affected by chronic nutrient
400 enrichment from agricultural and domestic sources. Potentially toxic cyanobacterial blooms,
401 which thrive when levels of phosphorus are high, present a particular societal threat, since
402 Lough Neagh provides drinking water to approximately one million people. Using the
403 PROTECH phytoplankton response model, calibrated with ECN input data, the authors
404 predicted how the lough's phytoplankton might respond to a potential increase in
405 temperature driven by climate change and to a gradual reduction in nutrient load as a
406 consequence of tighter controls on nutrient releases. The results suggest that future warming
407 could simply lead to the replacement of one cyanobacterial species by another, unless
408 phosphorus inputs are reduced more substantially.

409 As many of the studies in this issue demonstrate, accurate and repeatable measurements are
410 vital for the assessment of long-term trends in ecosystem structure or function. However,
411 some well-established monitoring techniques can be time consuming and therefore
412 expensive, and may also be prone to subjective variations arising, for example, from recorder
413 bias. There is an increasing expectation internationally for long-term environmental
414 monitoring programmes to become more efficient and thus less demanding on resources.
415 Recent technological developments provide a variety of opportunities to make
416 measurements at lower cost, and in some cases to higher levels of accuracy and provision.
417 Baxendale et al. (this issue), working at the Moor House ECN site, investigated the use of
418 several digital image techniques for recording vegetation cover at the plant functional type
419 level. Whilst not, in this case, a replacement for the current vegetation monitoring protocol,
420 this approach presents the potential to rapidly and accurately assess plant functional type
421 cover during spatial surveys and over time at fixed locations. ECN sites should not only be
422 able to benefit from such technologies in the longer term, but in the meantime they also

423 provide excellent research platforms at which new approaches can be tested alongside more
424 conventional measurements.

425

426 **3. Current and future development**

427 As the previous section demonstrates, the first two decades of consistent measurements for
428 the majority of the original parameters are serving to increase our understanding of the
429 changing state of the environment at ECN sites. The scientific potential of the network
430 should only increase as the datasets continue to lengthen. All long-term environmental
431 monitoring programmes, however, ultimately face conflicting pressures from the desire to
432 maintain uninterrupted records and the need to adapt to evolving environmental concerns
433 and policy priorities. Furthermore, new opportunities frequently arise from the development
434 of new monitoring and measurement technologies, while tightening financial constraints
435 often impose a need to increase the efficiency of data collection and management. In this
436 section, therefore, we consider the challenges and opportunities that the ECN, and the
437 terrestrial network in particular, needs to address to ensure optimal delivery of data of value
438 to science, land management and policy development over the next two decades and
439 beyond.

440 *3.1 Emerging environmental concerns*

441 At the time of initiation of the ECN, concerns over the possible impact of global climate
442 change had only recently begun to emerge (Huntingford and Friedlingstein, 2015), while
443 issues associated with air pollution, and acid rain particularly, dominated the environmental
444 policy agenda. Over the last two decades the balance has clearly shifted, particularly as a
445 consequence of large reductions in acid deposition as well as a heightened awareness of
446 changes in global climate. Trends in weather at ECN sites have been documented in some
447 detail (Morecroft et al., 2009; Monteith et al. this issue). While long-term change in air
448 temperature is arguably the element of climate change that receives most attention, the
449 strongest shifts in weather identified over the past two decades mostly involve changes in
450 hydrology, particularly increases in summer rainfall and the magnitude of extreme
451 precipitation events. Within the period covered by the ECN there have also been significant
452 periods of drought, e.g. 1995-97 and 2004-06. The likelihood that changes in the distribution
453 of precipitation extremes across the northern hemisphere are linked to climate change is
454 receiving increasing attention within the climate modelling community (see, for example, Min
455 et al., 2011). The co-location of high frequency meteorological, biogeochemical and
456 biological measurements places the ECN in a unique position to begin to assess the
457 implications of such effects on ecosystems. However, more detailed quantification of these
458 episodes, including higher frequency rainfall recording (until recently summarised at hourly
459 intervals only) and the deployment of more intelligent instrumentation, for example to assess
460 rain drop size and intensity, may be necessary to further enhance this capability.

461 While the UK NEA has highlighted threats to a range of ecosystem services it is also clear
462 that some of the potentially most important elements and processes contributing to these
463 services are rarely monitored in a consistent manner within a wider integrated framework of
464 measurements over long time scales. Indeed, several key components of nutrient cycling
465 (including soil microbial and animal communities, decomposition and net primary
466 productivity) and response variables such as gaseous fluxes of carbon dioxide, methane and
467 nitrous oxide are currently not assessed routinely at ECN sites; methods and instrumentation
468 in these areas were either underdeveloped and/or prohibitively expensive at the time of
469 initiation of the network. Recent methodological advances in, for example, gene sequencing
470 and microclimatological instrumentation, provide new opportunities to quantify processes at
471 various temporal and spatial scales, and link these to the broader measurements made at
472 ECN sites.

473 The ECN would clearly stand to gain from the incorporation of these new technologies as
474 they become more affordable and deployable. Alternatively, certain parameters of potential
475 importance could be modelled from data generated by other networks. For example, there is
476 growing concern that background ozone concentrations in the UK are rising to levels
477 potentially harmful to natural vegetation, which, when coupled with the occurrence of
478 drought, could lead to synergistic impacts (Mills et al., 2009). While ozone is not currently
479 measured routinely at ECN sites there is potential for ozone measurements made by the
480 Automatic Urban and Rural Network (Defra) to be extrapolated to ECN sites, while rotation
481 of a small number of ozone monitors might be sufficient to validate model predictions.

482 *3.2 Strengthening links with other programmes and initiatives*

483 As emphasised in Section 2, progress is being made in the development of stronger links
484 between the ECN and other networks and programmes. At a UK level the Environmental
485 Observation Framework (UKEOF) has been established to develop greater collaboration and
486 integration of observation systems and enhance the collective policy and scientific potential
487 of existing programmes. It also aims to secure benefits from the sharing of measurements,
488 data, equipment, skills and resources. One example of more collaborative working is the
489 ECN's recent adoption of the central data management role in Natural England's Long-Term
490 Monitoring Network (LTMN), the main contributor to the ECBN. Indeed, the ECBN was
491 conceived as a complementary network that would increase the spatial coverage of ECN-
492 compatible ecological measurements and thus provide greater capacity to assess
493 environmental change at a national level.

494 Further promising opportunities for integration at a national scale have arisen recently with
495 the development of the NERC funded Cosmic-ray soil moisture monitoring network
496 (COSMOS-UK). This focusses principally on measuring variations in soil moisture at the field-
497 scale, but monitoring also includes meteorological and spectrometric measurements, and
498 high frequency telemetry of data to a central database. Several COSMOS-UK stations have

499 been sited at or near ECN terrestrial sites and this should allow a close coupling of
500 observations.

501 As previously mentioned, the ECN is a member ofILTER and its European regional
502 component, LTER-Europe, but until recently the extent of international cooperation, in terms
503 of joint data analysis, has been limited (Vihervaara et al., 2013). However, LTER-Europe is now
504 receiving EU funding (from 2015-19) through the eLTER project (eLTER) to advance the
505 European network of Long-Term Ecosystem Research sites and socio-ecological research
506 platforms. This will include the design of a cost-efficient pan-European network, able to
507 address multiple ecosystem research issues. The ECN is represented in the project via the
508 NERC Centre for Ecology & Hydrology. One aim of eLTER is 'to develop the organisational
509 framework for data integration and enable virtual access to the LTER data'.

510 Data integration becomes much easier and more cost effective when core variables are
511 measured in comparable ways. The eLTER project will, therefore, build on earlier work by
512 LTER-Europe and the EnvEurope and ExpeER projects to develop a recommended set of
513 standard parameters and harmonised sampling methods (Firbank et al., 2014). One challenge
514 for the LTER community will be to ensure the widespread uptake of these methods, and
515 additional resources may be needed in order to safeguard the most important long-term
516 records of some well-established national LTER networks.

517 *3.3 Linking ECN observations to wider areas*

518 There is potential to both scale up observations made at ECN sites, and use ECN sites as
519 earth observation (EO) calibration platforms, particularly through (i) greater integration with
520 EO communities and (ii) other national and international cooperation. Earth observations
521 range from simple photographs made by fixed cameras deployed at a plot scale through to
522 more complex multi-spectral imaging and the deployment of Unmanned Aerial Vehicles
523 (UAVs; or drones) and satellites gathering data from landscape to international scales. Earth
524 observation data have only been used to a limited extent to date in assisting assessments of
525 ECN sites. For example, Dick et al. (2014) compared data on ecosystem service indicators
526 determined at 11 ECN sites with ecosystem service indicators obtained from pan-European
527 databases, based on sources including remote sensing, agricultural statistics and model
528 simulations. More recently, a series of UAV-derived images of the ECN Moor House site is
529 now enabling key site features, and the distribution of some elements of natural capital such
530 as peatland extent and the distribution of plant functional types, to be mapped in greater
531 detail than has been possible previously, and similar approaches are being considered for
532 the wider network.

533 The case for greater national cooperation between monitoring and survey programmes has
534 already been made with respect to the added value of bringing together measurements
535 made at different temporal and spatial scales. Broader environmental gradients can also
536 provide useful context for local observations. For example, assessment of the effects of

537 climate change or nitrogen deposition on an ecosystem will benefit from knowledge of
538 current environmental status and trends in regions that are warmer, cooler, or more or less
539 nitrogen impacted. In this respect even greater gains can be brought from effective
540 international cooperation. To this end the eLTER project will include two case studies. One
541 aims to demonstrate the ability of the network to gather – from selected European LTER sites
542 – a range of climate, soil and atmospheric deposition data with which to model climate and
543 pollution impacts on plant biodiversity and ecosystems (using LTER-derived vegetation data).
544 A second study will assess the potential of European Long-Term Socio-Ecological Research
545 (LTSER) platforms (Haberl et al., 2006) to address scientific and societally-relevant questions
546 concerning terrestrial and freshwater ecosystem services, natural capital (stock and change),
547 and related issues such as human wellbeing. The ECN will contribute to these case studies by
548 providing relevant data (in the case of LTSER platforms, from the Cairngorms National Park,
549 currently the UK's only LTSER platform). The ECN site within this LTSER platform has already
550 been used to study the ecological, economic and socio-cultural adaptive cycles at three
551 politically relevant spatial scales: National Nature Reserve, National Park and devolved
552 government (Dick et al 2011). Such studies can help land managers and policymakers to
553 evaluate risks to the delivery of ecosystem services posed by management practices, and can
554 be used to predict future service delivery.

555 *3.4 Increasing early warning capability*

556 There is a clear need to increase the speed of data capture and processing in order to
557 improve the capacity of environmental observation systems to provide early warnings of
558 change. The suggestion that ecological data may contain information indicative of
559 approaching 'tipping points' continues to be debated (Scheffer et al., 2009; Burthe et al.,
560 2015), but there is clearly merit in developing systems that are able to identify anomalous
561 behaviour rapidly, such as the sudden absence of a previously constant species. Furthermore,
562 the possibility that extreme climatic events are occurring more frequently heightens the need
563 to identify where and when these occur at as close to real time as possible. This would then
564 enable additional targeted sampling campaigns that might be required to quantify effects to
565 be conducted efficiently. Advances have already been made in accelerating the processing of
566 ECN field measurements, through the use of digital data entry templates, and with respect to
567 the telemetry of physical data (Rennie, this issue). Currently, the ECN is focussing on the
568 development of automated statistically-based analytical tools that should enable the most
569 recently collated biological and environmental data to be rapidly screened, allowing, for
570 example, variation in species community metrics to be assessed in the context of prevailing
571 weather conditions.

572 Clearly, early warning capacity could be further enhanced through wide scale deployment of
573 environmental sensors coupled with automated, digital data transfer, processing and
574 visualisation, as exemplified by the German Terrestrial Environmental Observatories network
575 (TERENO) and the US National Ecological Observatory Network (NEON), although this would

576 take considerable new investment. It is clearly vital, however, that the integrity of the ECN's
577 long-term records remains paramount, and that the use of more rapid, perhaps automated,
578 modes of data collection and transfer from the field to the end-user does not result in breaks
579 in key time series, poorer data quality or less rigorous scrutiny of data.

580

581 **4. Concluding remarks**

582 This Special Issue was produced to mark the first twenty years of monitoring at terrestrial
583 ECN sites, and illustrates the diverse ways in which long-term integrated environmental
584 monitoring is helping to quantify, and elucidate the causes and consequences of
585 environmental change across a broad range of UK habitats. The primary purpose of the
586 network is to provide long-term environmental data and physical platforms for
587 environmental research, but the information generated has considerably wider societal value
588 with respect to informing policy and management strategies, and providing an educational
589 resource. The extent of high frequency co-located physical, biogeochemical and biological
590 measurements is unique in a UK context and makes the network particularly valuable in the
591 development of clearer process understanding, and with respect to determining the impact
592 of relatively short term events, such as droughts and floods, on ecosystems. The true worth
593 of the network, however, may only be recognised through tighter integration of observations
594 with those generated by compatible monitoring and survey programmes that operate over
595 differing temporal and spatial scales, both nationally and internationally. Moreover, while
596 continuation of time series is paramount, there is a continuing need to review approaches to
597 data capture and management in order to improve efficiency, and to augment, or in some
598 cases replace, the current range of measurements with novel instrumentation and methods
599 that will accelerate rates of data transfer and processing. Such adaptation and development
600 is clearly necessary if the ECN is to continue to provide important insights into the nature of
601 environmental change over the next two decades and beyond.

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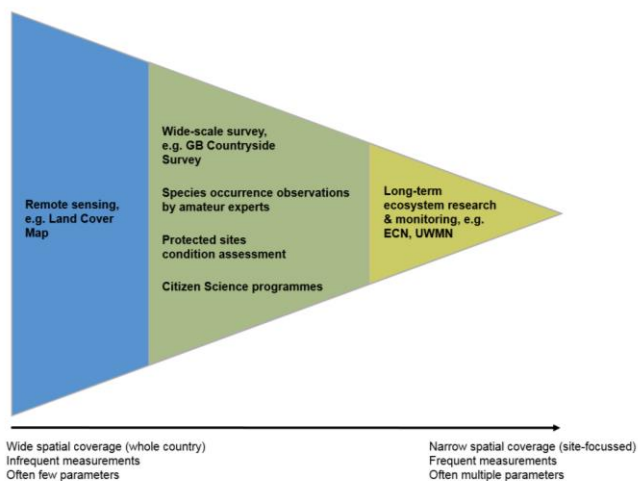
631 **Figure captions**

632 **Figure 1: Diagram illustrating the gradient of ecosystem observation programmes in**
633 **the United Kingdom, from spatially extensive, low frequency surveys to more site-**
634 **focussed, measurement-intensive observations, such as the ECN.**

635 Key

636 ECN: UK Environmental Change Network

637 UWMN: Upland Waters Monitoring Network



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640

641 **Figure 2: Map of ECN sites. The sites referred to in this Special Issue are labelled. Trout**
642 **Beck (a stream) is on the Moor House site.** Full details of all ECN sites can be found on the
643 ECN website, www.ecn.ac.uk/sites.

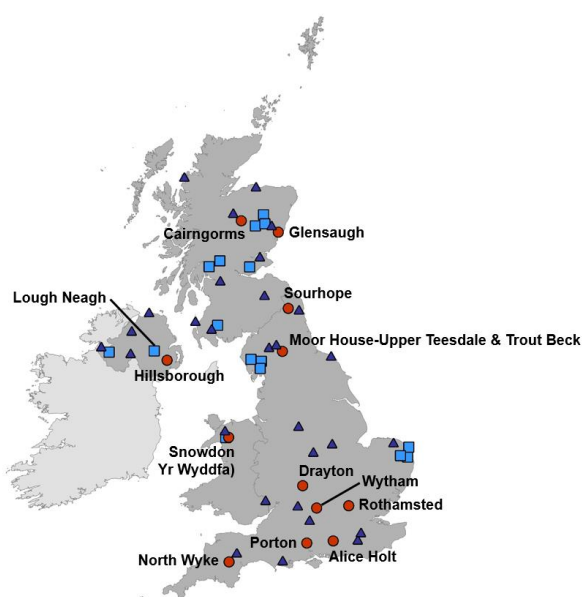
644 **Key**

645 Circles: Terrestrial sites

646 Squares: Lake sites

647 Triangles: Rivers and streams

648



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651 **Tables**

652 **Table 1: Selected characteristics of ECN terrestrial sites.** Mean annual temperature and rainfall were calculated using data from the ECN
 653 automatic weather station for the period 2000-2012, since this can be calculated for all sites. The weather stations are not necessarily sited at
 654 the mid-point of the site's altitude range

Site	Years as an ECN site	Altitude range (m a.s.l.)	Area (ha)	Mean annual temperature (°C)	Mean annual rainfall (mm)	Geology	Soils	Vegetation	Dominant land use
Alice Holt	1992 to present	76-125	850	10.8	833	Clay	Sandy brown forest soils and surface water Gleysols	Woodland (hard and softwood species)	Productive forest; amenity woodland
Cairngorms	1999 to present	320-1111	1000	4.7	900	Granite	Skeletal Peaty podsol; blanket peat	Montane heath; Grass/heather mosaic; woodland (Caledonian Pine)	Conservation; recreation
Drayton	1992 to 2014	40-80	190	10.3	630	Limestone and clay drift	Clay	Mixed arable and grassland; short-rotation coppice	Mixed farming; biocrops
Glensaugh	1992 to present	137-487	1125	7.5	1153	Old red sandstone; Schists	Peaty podsols; Brown forest soils	Grass/heather mosaic	Livestock grazing
Hillsborough	1992 to present	110-170	400	9.4	1119	Silurian Shales and Greywackes; Glacial and alluvial deposits	Dystric Stagnosol	Woodland; grassland; arable	Livestock grazing; amenity woodland
Moor House - Upper Teesdale	1992 to present	290-848	7500	5.8	2065	Limestone; Sandstone; Shale	Varied: Brown earths; podsols; gleys; peats	Blanket bog; acidic and calcareous grassland	Livestock grazing; grouse moor; water supply catchment
North Wyke	1992 to present	120-180	250	9.9	1048	Clay shales	Silty clay; clay	Grassland; woodland	Livestock grazing; Experimental research on grass
Porton Down	1994 to present	100-172	1227	9.7	803	Chalk	Rendzina	Semi-natural chalk grassland; small areas of woodland	Military testing (no agriculture since First World War)

Authors' accepted version

Rothamsted	1992 to present	94-134	330	10.2	705	Brown earths with gleying; Clay-with-flints; over chalk	Flinty silty clay loam (18-27 % clay)	Mixed arable and grassland; short-rotation coppice; woodland	Experimental research on grass, cereals, oilseeds and energy crops
Sourhope	1992 to present	200-601	1119	7.4	975	Old red sandstone	Peaty gleys; Brown forest soil	Grass/heather mosaic	Livestock grazing
Wytham	1992 to present	60-165	770	9.9	745	Limestone; Sand; Clay	Rendzinas; brown earths; alluvial deposits	Woodland; grassland; arable	Unmanaged woodland; livestock grazing; crop production
Yr Wyddfa / Snowdon	1995 to present	298-1085	700	7.4	3784	Rhyolite; Dolerite; Moraines	Varied: Brown earths; podsoles; gleys; peats	Acid/calcareous grassland; upland and montane heath; blanket bog	Livestock grazing; recreation; hydropower

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661 **Table 2: Principle variables measured at ECN terrestrial sites.** Additional measurements made only at some sites are shown in *italics*. Full
 662 protocols, including those used for monitoring freshwater sites, are available on the ECN website, www.ecn.ac.uk/measurements.

Variable measured	Frequency of measurement	Notes
Vertebrates		
Grazing animals, e.g. deer, sheep, rabbits	Two times per year	
Bats	Four times a year between June and September	
Frog spawning behaviour	Weekly phenological recording from adult congregation to full metamorphosis of tadpoles	
Birds	Two times per year	Using BTO Breeding Bird Survey methodology since 2000, prior to that the Common Bird Census and Moorland Bird survey were used
Invertebrates		
Moths	Each night or weekly at remote sites	Using Rothamsted light trap network methodology
Butterflies	Each week between April and September, provided weather conditions are favourable	Using UK Butterfly Monitoring Scheme methodology
Spittle bugs	Twice a year (nymphs in June, adults in August)	
Ground predators (beetles and spiders)	Every two weeks between May and October	
Vegetation		
Whole site baseline survey (species presence related to the National Vegetation Classification)	Once at establishment of the site	
Permanent plots monitored for species presence	Recording intervals for different plots are 1, 3 or 9 years	Some sites also record additional information on woodland plots, vegetation boundaries, grass yields and cereal field monitoring

Land use and site management		
Records of management activities		
Automatic Weather Station recording		
Solar radiation; net radiation; humidity; air temperature; wind speed; wind direction; rainfall; albedo (sky and ground); soil temperature at 10cm and 30cm; surface wetness; soil water content	Hourly summaries from 5-sec samplings	
Manual meteorological recording		
Dry bulb & wet bulb temperature; maximum & minimum temperature; grass minimum temperature; soil temperature; rainfall; wind run	Daily or weekly	Carried out for quality control purposes. Some sites have replaced manual meteorological recording with a second Automatic Weather Station
Atmospheric chemistry		
Nitrogen dioxide	Every two weeks	
Ammonia	Monthly	
Precipitation chemistry		
pH; conductivity; alkalinity; sodium; potassium; calcium; magnesium; iron; aluminium; phosphate; ammonium; nitrate; chloride; sulphate; total phosphorous; total nitrogen; dissolved organic carbon	Weekly	
Surface water discharge		
<i>Continuous discharge measurements</i>	Summarised every 15 minutes	
Surface water chemistry		
pH; conductivity; alkalinity; sodium; potassium; calcium; magnesium; iron; aluminium; <i>total</i>	Weekly at some sites	

<i>phosphorous</i> ; phosphate; <i>total nitrogen</i> ; ammonium; nitrate; chloride; sulphate; dissolved organic carbon		
Soil solution chemistry		
pH; conductivity; alkalinity; sodium; potassium; calcium; magnesium; iron; aluminium; <i>total phosphorous</i> ; phosphate; <i>total nitrogen</i> ; ammonium; nitrate; chloride; sulphate; dissolved organic carbon	Every two weeks at some sites	
5-yearly soil survey		
Horizon depth and thickness; soil moisture; pH; exchangeable acidity; exchangeable sodium, potassium, calcium, magnesium, manganese and aluminium; total nitrogen, phosphorus, sulphur, organic carbon and inorganic carbonate; particle size analysis and soil minerology	Once every 5 years	
20-year soil survey monitoring		
As fine-grain monitoring, with the addition of: Bulk density; total lead, zinc, cadmium, copper, mercury, cobalt, molybdenum, arsenic, chromium, and nickel; extractable iron, aluminium and phosphorus; particle size analysis and soil minerology	Once every 20 years	A baseline survey was made in the first year of monitoring, comprising a soil map at 1:10000 scale (or 1:25 000 for larger sites) and soil typologies derived from auger borings

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