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1 Rainfall interception modelling: is the wet bulb approach adequate to estimate mean  
2 evaporation rate from wet/saturated canopies in all forest types?

3

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15

16

17 Abstract

18

19 The Penman-Monteith equation has been widely used to estimate the maximum evaporation  
20 rate ( $E$ ) from wet/saturated forest canopies, regardless of canopy cover fraction. Forests are  
21 then represented as a big leaf and interception loss considered essentially as a one-  
22 dimensional process. With increasing forest sparseness the assumptions behind this big leaf  
23 approach become questionable. In sparse forests it might be better to model  $E$  and  
24 interception loss at the tree level assuming that the individual tree crowns behave as wet bulbs  
25 (“wet bulb approach”). In this study, and for five different forest types and climate conditions,  
26 interception loss measurements were compared to modelled values (Gash’s interception  
27 model) based on estimates of  $E$  by the Penman-Monteith and the wet bulb approaches.  
28 Results show that the wet bulb approach is a good, and less data demanding, alternative to  
29 estimate  $E$  when the forest canopy is fully ventilated (very sparse forests with a narrow  
30 canopy depth). When the canopy is not fully ventilated, the wet bulb approach requires a  
31 reduction of leaf area index to the upper, more ventilated parts of the canopy, needing data on  
32 the vertical leaf area distribution, which is seldom-available. In such cases, the Penman-  
33 Monteith approach seems preferable. Our data also show that canopy cover does not *per se*  
34 allow us to identify if a forest canopy is fully ventilated or not. New methodologies of  
35 sensitivity analyses applied to Gash’s model showed that a correct estimate of  $E$  is critical for  
36 the proper modelling of interception loss.

37

38

39 Keywords: interception loss; surface temperature; Gash model; sparse forest; Penman-  
40 Monteith

41

## 42 1. Introduction

43

44 A proportion of the rain falling on to a forest canopy is intercepted and evaporates back to the  
45 atmosphere (David et al., 2005). Several models of the process have been developed (see the  
46 review by Muzylo et al., 2009) and these have contributed to a good understanding of the  
47 underlying mechanisms of interception loss. Interception models are also important as a  
48 component of hydrological catchment models or continental-scale water balance models (e.g.  
49 Wallace et al., 2013), to assess global evaporation (e.g., Miralles et al., 2010; Zhang et al.,  
50 2016), and in the land surface schemes of Global Circulation Models (see Carlyle-Moses and  
51 Gash, 2011).

52 The most widely used interception models are those developed by Rutter (Rutter et al., 1972;  
53 Rutter et al., 1975) and Gash (Gash, 1979). The former was the first with a physically-based  
54 background where interception loss was explicitly driven by the rate of evaporation from the  
55 wet canopy. To calculate the dynamic water balance of the forest canopy and trunks, during  
56 each rainfall event, the Rutter model requires a continuous evaluation of the maximum  
57 evaporation rate under wet conditions. Based on the Rutter model, Gash (1979) proposed a  
58 simpler, storm-based analytical model to estimate interception loss, which needs only the  
59 average rainfall and evaporation rates ( $\bar{R}$ ,  $\bar{E}$ ) under fully saturated canopy conditions for the  
60 entire period of simulation.

61 In their original formulations, these models assume that forest canopy uniformly covers the  
62 entire ground area. Based on this assumption, they were successfully applied to closed canopy  
63 forests, but their application to sparse forests proved to be problematic, with interception loss  
64 being overestimated (Gash et al., 1995). To overcome this limitation, both the Rutter and  
65 Gash models have been reformulated to adapt to sparse forests (Gash et al., 1995; Valente et  
66 al., 1997) by treating the open and the covered areas separately. In these revised model

67 versions, the rate of evaporation is partitioned between the open area, where it is considered  
68 zero, and the covered area where it is modelled as a closed forest under the same  
69 environmental conditions.

70 Usually, the Penman-Monteith equation is adopted to estimate the maximum evaporation rate  
71 from the wet/saturated canopy (Carlyle-Moses and Gash, 2011), setting canopy resistance to  
72 zero. With the Penman-Monteith model the tree canopy is considered as a big leaf, and  
73 evaporation is treated as a one-dimensional vertical process, with the aerodynamic  
74 conductance estimated assuming a vertical logarithmic wind profile between the canopy level  
75 and some reference height above it (van Dijk et al., 2015). However, this assumption does not  
76 take into account the possible effect of forest sparseness on the enhancement of turbulence  
77 and evaporation rate – becoming increasingly questionable as the forest becomes more and  
78 more sparse.

79 Pereira et al. (2009b) suggested that, for very sparse stands, an approach based on the rate of  
80 evaporation from the individual, isolated wet (non-overlapping) tree-crowns would be more  
81 appropriate. These authors showed that the saturated crowns of isolated trees behave like wet  
82 bulbs, allowing the estimation of their evaporation rate through a simple diffusion equation.  
83 Knowing the tree density, the whole-stand evaporation could then be derived in this case as  
84 the sum of the contribution of the individual trees.

85 Like the Penman-Monteith model, this “wet bulb approach” is also physically based but,  
86 compared to the former, requires less data to estimate the maximum evaporation rate from  
87 saturated tree canopies.

88 By combining this approach with the Gash analytical model, Pereira et al. (2009a) estimated  
89 the interception loss from two savanna-type Mediterranean oak woodlands with a good  
90 accuracy (normalized mean error less than  $\pm 10\%$ ).

91 Being simpler and less data demanding than the Penman-Monteith equation, the wet bulb  
92 approach seems an attractive option. However, the need to check whether the assumption that  
93 tree crowns behave as fully ventilated wet bulbs remains. We need to answer the question: is  
94 the wet bulb approach applicable or adaptable to more-closed forests? For instance, Roberts et  
95 al. (1990; 1993) showed that the canopy of a closed Amazonian rainforest was much better  
96 ventilated in the upper crown strata (roughly the upper half of the canopy), where wind speed  
97 was higher and air temperature relatively uniform compared to the lower canopy layers.  
98 Furthermore, the results reported by Gash et al. (1999) show that better estimates of  
99 evaporation rate from a fully wet, sparse pine forest based on use of the Penman-Monteith  
100 model were obtained when the aerodynamic conductance for vapour flux was set equal to the  
101 measured conductance to momentum flux. This may be taken as an additional indication that  
102 in saturated canopies the lower boundary of the main source of water vapour flux is located at  
103 the same height where momentum is (apparently) absorbed.

104 Many forest structural characteristics may affect its aerodynamic behaviour, such as the  
105 canopy cover fraction, tree density, tree height, canopy depth and forest composition (type  
106 and number of species). Our aim is to determine how these structural features may interact,  
107 trying to distinguish in which types of forests interception loss can be best modelled using a  
108 one (Penman-Monteith) or a three-dimensional (wet bulb) approach.

109 The present study reanalyses data from several forest types and climate conditions where the  
110 measurement and modelling of interception loss has already been done previously: a  
111 eucalyptus plantation in central Portugal, two maritime pine stands (one in Portugal and  
112 another in Les Landes, France), an agroforestry system in Kenya and an Amazonian terra  
113 firme rainforest (see Table 1 for references).

114 The objectives of the work were: (1) to use the micrometeorological datasets obtained in the  
115 course of previous research to derive new estimates of the maximum evaporation rate from

116 fully wet canopies using the wet bulb approach ( $E_{WB}$ ); (2) to compare interception loss  
117 measurements with modelling results using these  $E_{WB}$  estimates, attempting to check the  
118 adequacy of the wet bulb approach in forests of different sparseness; (3) to quantify the  
119 impact of the method used to estimate  $E$  (Penman-Monteith or wet bulb) on the performance  
120 of Gash's interception model.

121

122

## 123 2. Methods

### 124 2.1. Sites

125 Two main criteria were used to select the forest sites: (1) they should cover a wide range of  
126 forest structure; and (2) availability of the necessary datasets. Four distinct forest types at five  
127 different locations were selected: two maritime pine stands with canopy covers of 45% and  
128 64%; a *Eucalyptus globulus* Labill. plantation with a canopy cover of 60%; an Amazonian  
129 tropical rainforest with a canopy cover of 92%; and an African agroforestry plantation  
130 consisting of a tree stratum of *Grevillea robusta* with a tree crown cover varying from 2 to 54  
131 %. Details of forest stands are given in Table 1. Besides differences in canopy cover, these  
132 forests also contrast in climate type and rainfall regime (maritime, Mediterranean, and tropical  
133 wet and semi-arid/sub-humid). Total annual rainfall and potential evapotranspiration varies  
134 between sites from 600 to 2400 mm and 741 to 1396 mm, respectively, while the ratio  
135 between them varies from 0.5 (in the Portuguese and Kenya sites) to 1.8 (in the Amazonian  
136 rainforest) (Table 1).

137 All the listed structural parameters (namely canopy cover, leaf area index, number of species,  
138 plant density, tree height and age) are liable to influence the rainfall interception process  
139 (Llorens and Domingo, 2007), either directly or indirectly.

140 As with most rainfall interception modelling studies, the contribution of undergrowth or of  
 141 lower vegetation strata to interception loss was not considered in the original studies.  
 142 Likewise, it is not considered in this study.

143

## 144 2.2. Mean evaporation rate

145 In all sites used in this study, the revised version of Gash's model has previously been applied  
 146 to predict interception loss, using the Penman-Monteith equation to estimate the average  
 147 maximum evaporation rate ( $\bar{E}_{PM}$ ) from the wet canopies assuming a one-dimensional  
 148 representation of the forests (see Table 2). The good modelling results obtained in all cases  
 149 (good fit between measured and modelled interception loss) suggest that those evaporation  
 150 rates were adequately estimated.

151 As an alternative and for comparison purposes, the wet bulb approach suggested by Pereira et  
 152 al. (2009b) is now used to estimate the average maximum evaporation rate ( $\bar{E}_{WB}$ ). According  
 153 to Pereira et al. (2009b), evaporation ( $E$ ,  $\text{kg m}^{-2} \text{s}^{-1}$ ) from a fully wet, isolated tree crown can  
 154 be estimated as:

155

$$156 \quad \lambda E = \frac{\rho_a c_p}{\gamma} g_{bv} [e_s(T_s) - e_a] \quad (1)$$

157

158 and the surface temperature  $T_s$  ( $^{\circ}\text{C}$ ) of a saturated tree crown as:

159

$$160 \quad T_s = \frac{1}{\rho_a c_p} \frac{\gamma}{\Delta + \gamma} \frac{A}{g_{bv}} + T_w \quad (2)$$

161

162 where  $\lambda$  ( $\text{J kg}^{-1}$ ) is the latent heat of vaporization,  $\rho_a$  ( $\text{kg m}^{-3}$ ) is air density,  $c_p$  ( $\text{J kg}^{-1} \text{ }^{\circ}\text{C}^{-1}$ ) is  
 163 air specific heat at constant pressure,  $\gamma$  ( $\text{Pa } ^{\circ}\text{C}^{-1}$ ) represents the psychrometric constant,  $g_{bv}$



164  $(\text{m s}^{-1})$  is the tree bulk aerodynamic conductance for water vapour,  $e_s(T_s)$  (Pa) is the saturation  
165 vapour pressure at surface temperature  $T_s$ ,  $e_a$  (Pa) represents the actual vapour pressure of the  
166 surrounding air,  $\Delta$  ( $\text{Pa } ^\circ\text{C}^{-1}$ ) is the slope of the saturation vapour pressure vs. temperature  
167 curve,  $A$  ( $\text{W m}^{-2}$ ) is the available energy per unit tree crown projected area and  $T_w$  ( $^\circ\text{C}$ ) is the  
168 wet bulb temperature of the air.

169 Since under typical rainfall conditions available energy tends to zero (e.g., Stewart, 1977;  
170 Teklehaimanot and Jarvis, 1991; Pereira et al., 2009b), it becomes apparent from Eq. (2) that  
171 the surface temperature of a wet tree crown should approach the wet bulb temperature of the  
172 surrounding air. Therefore, Eq. (1) was used to estimate evaporation from wet tree canopies  
173 considering  $T_s = T_w$ , an assumption consistent with the analysis made by van Dijk et al.  
174 (2015). The mean evaporation rate from a wet tree crown with a surface temperature identical  
175 to the air wet bulb temperature ( $\bar{E}_{WB}$ ), was then estimated, following Gash (1979), as the  
176 average evaporation rate for all hours when gross rainfall rate equalled or exceeded  $0.4 \text{ mm}$   
177  $\text{hr}^{-1}$  (two raingauge bucket tips for Gash's original study).

178 Although both the Penman-Monteith and the wet bulb approaches estimate the maximum  
179 evaporation rate at which intercepted rain may evaporate back to the atmosphere, hereafter we  
180 will refer to it simply as "evaporation rate".

181

### 182 2.3. Aerodynamic conductance

183 The use of Eq. (1) only requires the measurement of the air wet and dry bulb temperatures ( $T_w$   
184 and  $T_d$ , respectively) and knowledge of the bulk tree crown aerodynamic conductance.

185 In all forest sites used here, both air temperatures (dry and wet bulb) were measured in the  
186 original studies by aspirated psychrometers with an accuracy of  $0.2^\circ\text{C}$ .

187 Since those studies did not include any component dedicated to the evaluation of the bulk  
188 aerodynamic conductance ( $g_{bV}$ ) for a tree crown, we had to estimate it for all forest sites as a  
189 function of mean leaf dimensions, and leaf area index ( $L^*$ ) (Pereira et al., 2009b):

190

$$191 \quad g_{bV} = \overline{g_{lV}} L^* / c \quad (3)$$

192

193 where  $\overline{g_{lV}}$  ( $\text{m s}^{-1}$ ) is the mean leaf boundary layer conductance for water vapour,  $c$   
194 (dimensionless) the canopy cover fraction and  $L^*$  (dimensionless) the leaf area index  
195 expressed on a total ground area basis (according to the original measurements). The correct  
196 calculation of  $g_{bV}$  is critical for a proper application of the wet bulb approach (Eq. 1), but  
197 requires some somewhat subjective assumptions in the estimation of both  $\overline{g_{lV}}$  and  $L^*$ .

198 In all cases except for the Amazonian rain forest,  $\overline{g_{lV}}$  was derived using the so-called  
199 engineering formulae dependent upon average leaf characteristic dimensions and wind speed.  
200 For each forest type, the formulae used were derived from those given by Monteith and  
201 Unsworth (2008), assuming that eucalyptus and *Grevilea robusta* leaves could be represented  
202 as flat plates and pine needles as cylinders.

203 The characteristic dimension of the leaves ( $l$ ) was taken as the average leaf dimension (length  
204 or diameter) parallel to the direction of air flow (Grace, 1983). For *Eucalyptus globulus*  
205 leaves and from measurements made by J. Tomé (personal comm., 1994)  $l$  was taken as 18  
206 mm (most leaves are vertical). *G. robusta* has highly divided, bipinnate leaves, which cannot  
207 be easily represented by any typical geometric shape. Moreover, their orientation in the tree  
208 canopy is also variable. Hence, we assumed a characteristic dimension for these leaves given  
209 by the average of the length and width of the main leaflets ( $l = 28.2$  mm). In the case of *Pinus*  
210 *pinaster* needles,  $l$  was considered as 1.5 mm, corresponding to the mean value of the range of  
211 variation of needle diameters in this species (Castroviejo et al., 1993).

212 It has been noted that the values usually obtained by engineering formulae differ from the  
213 actual (experimentally measured) conductances, depending on the leaf type, i.e., leaves or  
214 needles. For broadleaf species, the engineering formulae tend to underestimate  $g_{IV}$ , with the  
215 ratio between observed and estimated conductance usually varying between 1.25 and 1.5  
216 (Schuepp, 1993) - although values as high as 2.5 have been reported (Monteith and Unsworth,  
217 2008). The opposite happens with needles, which are grouped in clusters that create a “shelter  
218 effect” (Monteith and Unsworth, 2008). Mutual sheltering between needles reduces needle  
219 conductance so that they tend to be lower than those estimated by the engineering formula.  
220 This reduction has been observed to be in the range of 0.33 to 0.50 (Tibbals et al., 1964;  
221 Monteith and Unsworth, 2008). As a result of these effects we need either an enhancement  
222 factor in conductance in the case of leaves, or a reduction factor in the case of needles. For  
223 both cases, we have assumed here values for these factors that represent the midpoints of the  
224 above reported intervals of variation, i.e., 1.38 and 0.40 for leaves and needles, respectively.  
225 These values can be used whenever no specific information is available.  
226 The formulae derived to estimate  $\overline{g_{IV}}$  as well as the enhancement/reduction factors adopted  
227 for each forest are presented in Table 3.  
228 The estimates of  $\overline{g_{IV}}$  were then combined with the leaf area index (expressed on a tree crown  
229 projected area basis,  $L^*/c$ ) to determine the bulk tree crown aerodynamic conductance  
230 according to Eq. (3).

231

#### 232 2.4. Evaporation rate and leaf area index

233 In the modelling of interception loss by the Gash model the Penman-Monteith equation has  
234 been widely and successfully used in canopies with variable cover fraction as was the case for  
235 all forests considered in the present study. On the other hand, the wet bulb approach was, so  
236 far, only tested (successfully) in the modelling of interception loss from a savannah-type

237 forest (Pereira et al., 2009a) and from a traditional olive grove - pasture system (Nóbrega et  
238 al., 2015). Therefore and to evaluate the adequacy of the wet bulb approach, the new  $\bar{E}_{WB}$   
239 estimates (Eq. 1) were compared to the already tested  $\bar{E}_{PM}$  ones and results were analysed  
240 considering that:

241 a) the matching of estimates of  $E$  by both methods could be taken as an indication that the tree  
242 canopies are fully ventilated and any of the approaches can be used to model interception loss  
243 with equally good accuracy;

244 b) whenever the two estimates failed to match ( $\bar{E}_{WB} > \bar{E}_{PM}$ ), this could be seen as indicative  
245 that the whole canopy is not fully ventilated. In those cases we hypothesized that the upper  
246 and more ventilated parts of the canopy were the main contributors to interception loss.  
247 Accordingly, when  $\bar{E}_{WB} > \bar{E}_{PM}$ , we reduced the canopy leaf area to that of the top layers to  
248 test if  $\bar{E}_{WB}$  converged to  $\bar{E}_{PM}$  and if it was still possible to model interception loss with a good  
249 accuracy through the wet bulb approach.

250

## 251 2.5. Rainfall interception - Gash's analytical model

252 Although the Gash analytical model was used to estimate interception loss in all of these  
253 forests, the versions adopted in each case were not the same and, thus, the meaning of the  
254 canopy structure parameters differs from case to case. Table 2 shows the values of those  
255 parameters for each forest as derived in the original studies and indicates, as well, the model  
256 version used. For further details on the model structure and formulation, Table 2 also includes  
257 the references to the papers where the different versions are described.

258 The model version proposed by Valente et al. (1997) was adopted in this study at the stand  
259 level since it has been shown to improve the estimation of total interception loss in sparse  
260 forests, while retaining the ability to accurately predict interception loss from closed canopies.

261

## 262 2.6. Sensitivity analysis

263 Considering that the objective of this paper was to test the impact of a different method of  
264 calculating the mean evaporation rate under wet/saturated conditions ( $\bar{E}$ ) on interception loss  
265 modelling results, a sensitivity analysis was done on the performance of Gash's model  
266 (considering  $\bar{E}$  and the other model parameters). Two different approaches were selected: the  
267 first consists of a local analysis on the impact of evaporation rate on model output; the second  
268 is a global analysis whereby the combined and simultaneous influence of the various model  
269 parameters is accounted for. Although local sensitivity analyses of Gash model parameters  
270 have been conducted previously (e.g., Limousin et al., 2008), it has never been done  
271 simultaneously for multiple datasets. The overall/combined sensitivity analysis technique  
272 used here has never been applied before in rainfall interception modelling.

273

### 274 2.6.1. Local approach

275 The local sensitivity analysis was performed for the  $\bar{E}$  parameter. As this type of analysis is  
276 data-dependent, only results from a set of studies can give a broad view on the influence of a  
277 given parameter on model performance. Therefore, the effect of the variation of  $\bar{E}$  when all  
278 the other parameters were kept constant at their derived value (Table 2) was assessed for the  
279 five forests under study. For the Kenya agroforestry stand,  $S$  and  $c$  were set to their maximum  
280 observed values, 0.93 mm and 0.54, respectively.

281

### 282 2.6.2. Morris screening

283 The global sensitivity analysis allows the evaluation of the combined and simultaneous  
284 effects of the various model parameters. The Morris method (Morris, 1991) is a global  
285 sensitivity analysis technique that aims to identify the parameters that have: negligible effects,  
286 linear or additive effects, non-linear effects and interaction with each other. The parameter

287 space is divided into  $p$  levels, transforming the experimental region ( $\Omega$ ) in a  $k$ -dimensional  $p$ -  
 288 level grid, where  $k$  is the number of parameters. Within  $\Omega$  a starting value for the parameter  
 289 vector  $\mathbf{X}$  is randomly selected. A succession of  $(k + 1)$  sampling points, called a trajectory, is  
 290 created varying one parameter at time by a quantity  $\delta$ , multiple of  $1/(p - 1)$ . Each sampling  
 291 point differs from the previous one in only one factor. Once a trajectory is constructed an  
 292 incremental ratio, called Elementary Effect ( $EE$ ), can be computed for each parameter. For a  
 293 given value  $\mathbf{x} = (x_1, x_2, \dots, x_k)$  of  $\mathbf{X}$ , the  $EE$  of the  $i$ th input factor is defined as

294

$$295 \quad EE_i(\mathbf{x}) = \frac{[y(x_1, \dots, x_{i-1}, x_i + \delta, x_{i+1}, \dots, x_k) - y(\mathbf{x})]}{\delta} \quad (4)$$

296

297 The experimental design consists of  $r$  trajectories independently generated, with each  
 298 trajectory having a different starting point randomly selected. Since each succession provides  
 299 one  $EE$  for each parameter,  $k$  finite distributions of  $r$  elementary effects are created. The mean  
 300 ( $\mu$ ) and the standard deviation ( $\sigma$ ), from the distributions represent the sensitivity measures:  $\mu$   
 301 gives the overall importance of an input parameter, while  $\sigma$  describes non-linear effects and  
 302 interactions between parameters. Campolongo et al. (2007) enhanced the Morris method by  
 303 improving the sampling strategy and proposed calculating the mean of the distribution of the  
 304 absolute values of the elementary effects ( $\mu^*$ ).  $\mu^*$  was introduced because the effects of  
 305 opposite signs of  $EE$  could mask the importance of a parameter. For instance, if  $\delta$  variations  
 306 of a  $i$ th parameter can cause positive as well as negative effects on  $EE$ ,  $\mu$  will assume lower  
 307 values than  $\mu^*$ . Therefore,  $\mu^*$  better expresses the importance of the parameters and is more  
 308 reliable in ranking them.

309 Campolongo et al. (2007) also suggested assigning even values to the number of levels  $p$ ,  
 310 while making  $\delta$  equal to  $p/[2(p-1)]$ . The number of trajectories,  $r$ , has to be large enough so  
 311 that if two subsequent Morris analyses are performed with the same  $r$ , similar values of  $\mu$ ,  $\mu^*$

312 and  $\sigma$  must be obtained for each parameter. In other words, the number of trajectories must  
313 ensure that the results are general and not sample-specific.

314 The Morris method was applied, for the first time, to the sparse version of the Gash analytical  
315 model with  $r = 1000$  different input trajectories. Each of the seven parameters of the model ( $c$ ,  
316  $S$ ,  $S_t$ ,  $p_a$ ,  $e$ ,  $\bar{R}$  and  $\bar{E}$ ) varied between minimum and maximum values pre-defined for each site.  
317 The ranges taken for parameter variation (Table 4) were based on published literature, trying  
318 to reflect the characteristics of the forests studied.

319

320

### 321 3. Results

#### 322 3.1. Estimation of average evaporation rates from wet canopies

323 The estimates of  $\bar{E}$  obtained according to the wet bulb approach ( $\bar{E}_{WB}$ ) for the forests under  
324 analysis are presented in Table 5, along with the values derived in the original studies through  
325 the Penman-Monteith equation ( $\bar{E}_{PM}$ ). In two of the studied forests (the Carrasqueira pine  
326 stand and the Kenya agroforestry system),  $\bar{E}_{WB}$  was almost identical to  $\bar{E}_{PM}$ , when  
327 considering the contribution of the whole canopy (full  $L^*$ ) to interception loss. In the other  
328 cases,  $\bar{E}_{WB}$  using the whole canopy  $L^*$  overestimated  $\bar{E}_{PM}$ , suggesting that the canopy was not  
329 entirely and fully ventilated. Therefore,  $L^*$  was reduced to the upper canopy layers to test if  
330  $\bar{E}_{WB}$  would reach a value that could still allow a reasonably good interception loss modelling  
331 (see Section 2.4.). In the end, the estimated  $\bar{E}_{WB}$  was closer to  $\bar{E}_{PM}$  for all studied forests.

332 Table 5 presents the estimates of  $\bar{E}_{WB}$  considering both the full and reduced  $L^*$  values. Table  
333 5 also presents interception loss results: the originally measured and modelled values and new  
334 simulations through the revised version of Gash's analytical model (Valente et al., 1997),  
335 based on the  $\bar{E}_{WB}$  estimates. For all interception loss estimates, the normalized mean errors  
336 are also provided in Table 5.

337

## 338 3.2. Impact of evaporation rate on interception loss modelling

339 Although Table 5 gives a perception of the impact of the different  $\bar{E}$  estimates ( $\bar{E}_{PM}$  and  $\bar{E}_{WB}$ )  
340 on interception loss, a deeper insight can be obtained by performing sensitivity analyses on  
341 the sparse version of Gash's analytical model. Two approaches were followed: a local one to  
342 assess the effect of variations in  $\bar{E}$ , while keeping all the other parameters constant; and a  
343 global approach – Morris screening – to identify the importance and nature of the influence of  
344 all model parameters on interception loss estimates. Results of the two sensitivity analyses are  
345 presented in Figs. 1 and 2, respectively. According to Fig. 1, the two Portuguese forests, the  
346 Espirra eucalyptus plantation and the Carrasqueira pine stand, show the most sensitivity of the  
347 sparse version of Gash's analytical model to the mean evaporation rate: a relative change of  
348 +50% in  $\bar{E}$  results in an increase of nearly 30% in the estimated interception loss. Though to a  
349 lesser extent, modelled interception loss in the other three forests is also still quite sensitive to  
350 the mean evaporation rate. The global sensitivity analysis by Morris screening (Fig. 2)  
351 confirmed the importance of  $\bar{E}$ , independently of the different values the other model  
352 parameters may take: for all datasets except the Kenyan one,  $\bar{E}$  has high values of mean ( $\mu^*$ )  
353 and standard deviation ( $\sigma$ ).

354

355

## 356 4. Discussion

## 357 4.1. Estimation of average evaporation rates from wet canopies

358 The estimates of  $\bar{E}$  obtained according to the wet bulb approach ( $\bar{E}_{WB}$ ), considering the  
359 contribution of the whole canopy, and those derived in the original studies using the Penman-  
360 Monteith equation ( $\bar{E}_{PM}$ ), matched very well in the Carrasqueira pine stand and the Kenya  
361 agroforestry system. These two forests have highly sparse canopies and narrow crown depths



362 which favours air circulation within the canopy, allowing the surface temperature of saturated  
363 tree crowns to approach the air wet bulb temperature under rainy conditions. In these cases  
364 both methods (Penman-Monteith or wet bulb) can be used - the choice depending on data  
365 availability. However, the wet bulb method may be preferable since it is less data demanding  
366 and it lacks the questionable underlying assumptions in applying the Penman-Monteith  
367 equation in sparse forests (Monteith, 1965; Pereira et al., 2009b).

368 In all the other forests,  $\bar{E}_{WB}$  overestimated the evaporation rate when  $L^*$  of the entire canopy  
369 was considered, limiting the chances of good interception loss modelling if these  $\bar{E}_{WB}$   
370 estimates were used directly. The evaporation estimates by the wet bulb approach were then  
371 recalculated only accounting for the contribution of the upper and better ventilated parts of the  
372 canopy. However, the scope of this analysis was somewhat constrained by the limited  
373 information available on the vertical leaf area distribution in these forests.

374 For the eucalyptus forest, the mean evaporation rate estimates given by the Penman-Monteith  
375 model and the wet bulb approach, when  $L^*$  of the upper third of the canopy is considered, are  
376 nearly identical (leaf area index in the eucalyptus stand was 0.83, 1.40 and 0.94, for the upper,  
377 middle and lower thirds of the canopy, respectively; J. Tomé, personal comm., 1994). This  
378 eucalyptus forest plantation is relatively sparse, but the canopy depth represents about 61% of  
379 the mean tree height (Valente, 1999). Therefore, the ventilation of the lower part of the  
380 canopy may be attenuated leading to a reduction in evaporation from this canopy region.  
381 These results seem to suggest that the upper third of the canopy constitutes the main effective  
382 source of evaporation during rainfall, when tree crowns are saturated.

383 In Les Landes pine forest, the whole canopy  $L^*$  (2.3) referred to by Gash et al. (1995) was  
384 estimated using remote sensing techniques during a special observation period, from May to  
385 July 1986 (André et al., 1990). Here, the leaf area and  $L^*$  for the top crown layers were  
386 estimated based on the leaf area vertical distribution models derived by Porté et al. (2000) for

387 three Les Landes maritime pine stands. Besides other identical characteristics, one of these  
388 stands (Bray 95) had a total leaf area index very similar to that of the forest studied by Gash et  
389 al. (1995) and, thus, its vertical leaf area distribution was used. When only accounting for the  
390 contribution of the higher canopy layers, corresponding to the top fourth or third of crown  
391 depth, the mean wet bulb evaporation rate was 0.142 or 0.223 mm hr<sup>-1</sup>, respectively, which is  
392 not much different from the rate originally reported by Gash et al. (1995) (see Table 5). By  
393 using  $\bar{E}_{WB}$  associated with the top third of crown depth, interception loss could be modelled  
394 as efficiently as in the original study, suggesting that the wet bulb approach can also be used  
395 in these conditions as long as only upper and well exposed parts of the canopy are considered.  
396 In the Amazonian rainforest, Roberts et al. (1993) divided the whole forest canopy in five  
397 strata, assigning to each of them the respective  $L^*$  and an average leaf boundary layer  
398 conductance. This allowed the evaporation rate to be modelled considering the contribution of  
399 the different strata, especially of the top three layers. According to Roberts et al. (1990;  
400 1993), and in relation to the lower strata, these top layers were characterized by a more  
401 homogeneous air temperature profile and higher values of leaf conductance, probably a  
402 consequence of higher wind speed and more effective turbulent mixing. The average  
403 evaporation rate estimated by the wet bulb approach considering the contribution of these  
404 upper three layers of the canopy was 0.178 mm hr<sup>-1</sup> which is about 15% less than the original  
405 Penman-Monteith estimate obtained by Lloyd et al. (1988). The difference between both  
406 estimates may be related with the more or less arbitrary choice of the canopy depth and with  
407 the use of constant values for leaf aerodynamic conductance irrespective of wind speed.  
408 Indeed, in a forest like this, with high species diversity and a complex spatial pattern of leaf  
409 area distribution, it is not simple to derive  $g_{IV}$  wind-dependent functions using engineering  
410 formulae which must then also be combined with  $L^*$  to estimate a bulk aerodynamic  
411 conductance.

412 In three of our sites where it was necessary to reduce  $L^*$  to the upper canopy layers (Les  
413 Landes, Amazonia and eucalyptus) it is questionable whether the wet bulb approach should be  
414 adopted, because it would require seldom-available information on the leaf area vertical  
415 distribution. This may be particularly problematic in mixed forests with a complex 3-D  
416 structure. In all these cases the application of the Penman-Monteith equation seems more  
417 appropriate, as long as its underlying assumptions remain valid (Monteith, 1965).  
418 Results also show that the canopy cover fraction ( $c$ ) is not, *per se*, an adequate sparseness  
419 indicator to define when the wet bulb is a good alternative to Penman-Monteith. The Espirra  
420 eucalyptus plantation and the Carrasqueira pine stand are an example of this: both have  
421 approximately the same  $c$  but the wet bulb approach can only be successfully used without  
422 further assumptions in the pine site, probably because canopy depth is smaller in the pine  
423 forest compared to that of eucalyptus. We believe that in moderately sparse forests their  
424 structure (e.g., tree density, tree crown height and radius) also play an important role in  
425 determining the depth of the fully ventilated part of the canopy. Les Landes pine forest is  
426 another example: it has a canopy cover which is about 20% lower than that of Carrasqueira  
427 forest but its tree density is 50% higher. This means that the structure of the stand and the  
428 characteristics of individual tree crowns should differ. For instance, Les Landes forest with  
429 smaller and younger trees is more likely to have a larger relative canopy depth with leaf area  
430 distributed predominantly in its lower half (e.g., Porte et al., 2000). With deeper tree crowns  
431 and smaller distances between trees than in the Carrasqueira stand, Les Landes pine forest  
432 may behave more like the closed canopy rainforest, with mainly the upper part of the crowns  
433 contributing to the evaporation from the saturated canopy. Thus, it is not surprising that, when  
434 using the whole canopy  $L^*$ , the wet bulb approach overestimates  $\bar{E}$  by a value that doubles the  
435 original Penman-Monteith estimate in Les Landes pine forest.

436 The previous discussion evidences that a wider application/validation of the wet bulb  
437 approach is limited by the lack of easily obtainable information on foliage profile, canopy  
438 structure and forest sparseness. Recent studies suggest that some remote sensing techniques  
439 such as LiDAR and InSAR (e.g. Lefsky et al., 2002; Treuhaft et al., 2009; Tang et al., 2015)  
440 may be extremely useful to get that information.  
441 Furthermore, in all situations, the use of the wet bulb approach also depends on the possibility  
442 of deriving wind functions for tree bulk aerodynamic conductance using engineering  
443 formulae. This will certainly be easier when there is only one tree species and leaves have a  
444 simple morphology.

445

#### 446 4.2. Impact of evaporation rate on interception loss modelling

447 For a better evaluation of the impact of  $\bar{E}$  ( $\bar{E}_{PM}$  and  $\bar{E}_{WB}$ ) on interception loss estimates, a  
448 sensitivity analysis was performed for the sparse version of Gash's analytical model.

449 In the context of rainfall interception modelling, sensitivity analysis is typically applied as a  
450 local measure of the effect of each parameter on the model output (usually the interception  
451 loss) (e.g., Llorens, 1997; Valente et al., 1997; Limousin et al., 2008). Commonly, the relative  
452 importance of the uncertainty of a parameter on the output of a model is computed  
453 numerically by perturbing each parameter around a base value, while holding all the other  
454 parameters constant: the so-called "one-factor-at-a-time" sensitivity analysis (Saltelli and  
455 Annoni, 2010). As shown by previous authors (Llorens, 1997; Limousin et al., 2008), the  
456 interception loss predicted by Gash's analytical model is positively and linearly related to  $\bar{E}$ .  
457 However, its sensitivity to errors in this parameter depends on the values taken by data inputs  
458 and other parameters (Fig. 1). According to the analysis presented in Fig. 1, interception loss  
459 estimated by the sparse version of Gash's analytical model was quite sensitive to the mean

460 evaporation rate in all studied forests, particularly in the Portuguese eucalyptus and pine  
461 plantations where a +50% change in  $\bar{E}$  results in a nearly 30% increase in interception loss.  
462 Although in the present study, the main concern is on the average evaporation rate during  
463 saturation conditions, the other parameters of the model are also subject to errors and  
464 uncertainties. The previous one-factor-at-a-time sensitivity analysis cannot detect interaction  
465 among parameters and does not answer relevant questions like “which of the uncertain input  
466 parameters is driving most of the uncertainty in the output of the model?” (Saltelli et al.,  
467 2004). What is the importance of  $\bar{E}$  in this context?

468 To address these issues a global sensitivity analysis (Morris screening) was performed to  
469 evaluate the effect of a factor while all the others are also varying and interacting. Fig. 2  
470 shows how model output, affected by changes in the parameters, depends on the dataset used  
471 to run the model. Except for results obtained with the Kenya dataset,  $\bar{E}$  is an important  
472 parameter (high values of  $\mu^*$  and  $\sigma$ ). On the other hand, factors that parameterize stemflow  
473 ( $S_t$ ,  $p_d$ , and  $e$ ) have a much smaller effect and, in general,  $\bar{R}$  has a moderate influence on the  
474 output. In Kenya as in the two pine forests and the eucalyptus plantation, the model is also  
475 highly sensitive to the ground cover fraction ( $c$ ) showing the importance of correctly  
476 assessing this parameter in sparse forests. In general, parameters with a high value for  $\mu^*$  are  
477 also associated with a high value for  $\sigma$ , indicating that these parameters have also relevant  
478 non-linear/interaction effects, i.e., none of them has a purely linear effect on the modelled  
479 output. The exception is the canopy storage capacity ( $S$ ), that in four of the sites (Les Landes,  
480 Espirra, Carrasqueira and Amazonia) has a high overall effect on the output of the model  
481 (high  $\mu^*$ ) but a low  $\sigma$ , indicating that the effect of  $S$  is almost independent of the values of the  
482 other parameters. Overall, Morris screening has shown that  $\bar{E}$  has a large influence on the  
483 interception loss modelled by the sparse version of the Gash analytical model but its relative  
484 importance to the other parameters can depend on the dataset used to run the model.

485

486

## 487 5. Conclusion

488 In two of the studied forests (Portuguese pine stand and Kenya agroforestry system), the wet  
489 bulb approach provided very good estimates of  $\bar{E}$  under canopy saturation using  $L^*$  of the  
490 whole canopy. These results together with the structural features of the forests (low canopy  
491 cover and a narrow canopy depth) suggest that in both these cases the whole canopy can be  
492 considered as fully ventilated. Under these circumstances either the wet bulb or the Penman-  
493 Monteith approach can be used to estimate  $\bar{E}$ , but the wet bulb approach is simpler and less  
494 data demanding. Furthermore and in contrast with the Penman-Monteith approach, it makes  
495 no assumptions about horizontal homogeneity, which becomes problematic when forest  
496 sparseness increases.

497 In the other three forests (Les Landes pine stand, eucalyptus plantation and Amazonian  
498 rainforest) the wet bulb approach required a reduction of  $L^*$  to the upper, more ventilated  
499 parts of the canopy, needing seldom-available data on the vertical leaf area distribution. In  
500 those cases, the Penman-Monteith approach seems preferable.

501 Therefore, the logical follow up to the present study would be the development of a way to  
502 identify whether, or not, the forest tree crowns are exposed to the same air temperature and  
503 humidity conditions, i.e., whether the canopy is fully ventilated. The data used here suggest  
504 that the aerodynamic canopy conductance and the wind speed vertical profiles may depend on  
505 several forest structural parameters, such as canopy cover fraction, canopy depth, tree height,  
506 crown radius, tree density and forest composition and heterogeneity. It would be interesting to  
507 find simple, easily applicable parameters and/or relationships between the structural and  
508 aerodynamic features of forests that might help to identify if the canopy is fully ventilated or

509 not. Additionally, this research could bring some new insights into the processes underlying  
510 the evaporation from wet forest canopies.

511 The sensitivity analysis on Gash's interception model confirmed that it is particularly  
512 sensitive to wet canopy evaporation rate and, therefore, choosing the correct estimation  
513 method is of critical importance. Developing techniques that might help make that choice is  
514 essential if we are to correctly represent interception loss across the range of sparseness  
515 encountered in real forests.

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**Table 1** Location and main characteristics of the forests and experimental sites considered in this study

Site name	Site				
	Les Landes	Espirra	Carrasqueira	Amazonia	Kenya
Local	Les Landes, France	Herdade da Espirra, Portugal	Pinhal da Carrasqueira, Portugal	Reserva Florestal Ducke, Manaus, Brazil	Machakos , Kenya
Forest type	44° 5' N, 0° 5' W Maritime pine forest	38° 38' N, 8° 36' W Eucalyptus plantation	38° 50' N, 8° 51' W Maritime pine forest	2° 57' S, 59° 57' W Amazonian rain forest	1° 33' S, 37° 8' E Agroforestry plantation
Tree species	Maritime pine ( <i>Pinus pinaster</i> Aiton)	Eucalyptus ( <i>Eucalyptus</i> <i>globulus</i> Labill.)	Maritime pine ( <i>Pinus pinaster</i> Aiton)	Many tree species (see Cuartas et al. (2007))	<i>Grevillea robusta</i> A. Cunn.
Elevation (m)	146	85	20	----	1560
Study period	Feb/1986 - Jan/1987	Jan/1992 - Jul/1994	Jan/1992 - Jul/1994	Sep/1983 - Aug/1985	Nov/1994 - Jun/1997
Age (year)	37	7 (1993; first rotation)	60 (1993)	----	3
Forest density (trees ha <sup>-1</sup> )	430	1010	312	3000	833
Canopy cover (c, %)	45.0	60.0	64.0	92.0	2.0 - 54.0
LAI (L*)	2.30	3.20	2.70	6.60 (Roberts et al., 1993)	0.25 - 2.75
Mean tree height (m)	20.3	16.5	23.9	35.0 aprox.	from 0.5 to 9.5
Climate	Maritime	Mediterranean	Mediterranean	Tropical wet	Semi-arid/sub- humid
Mean annual rainfall (mm)	942 (André et al., 1986)	600 aprox.	600 aprox.	2391	782
Total rainfall in the study period (mm)	613	1546	1366	4804	1583
Mean potential annual evaporation (mm)	741 (Habets et al., 1999)	aprox. 1300	aprox. 1300	aprox.1319 (Shuttleworth, 1988)	1450 (Ong et al., 2000)
Original study	(Gash et al., 1995)	(Valente et al., 1997)	(Valente et al., 1997)	(Lloyd et al., 1988; Lloyd and Marques, 1988)	(Jackson, 2000)



**Table 2** Parameters of the Gash analytical model derived for each forest in the original studies

		Site				
		Les Landes	Espirra	Carrasqueira	Amazonia	Kenya
Gash's analytical model (version adopted)		Revised (Gash et al., 1995)	Revised (Valente et al., 1997)	Revised (Valente et al., 1997)	Original (Gash, 1979)	Revised (Gash et al., 1995)
Average rainfall rate (mm hr <sup>-1</sup> )	$\bar{R}$	1.650	1.814	1.743	5.150	2.280 (monthly rates in the range 0.5 - 3.2)
Average evaporation rate (mm hr <sup>-1</sup> )	$\bar{E}_{PM}$	0.170	0.200	0.315	0.210	0.230
Canopy storage capacity (mm)	$S$	0.250	0.210	0.410	0.740	0.710 - 0.930
Trunk storage capacity (mm)	$S_t$	0.170	0.016	0.017	0.150	0.185
Drainage partitioning coefficient	$p_d$		0.0324	0.0076		
Stemflow partitioning coefficient	$p_t$	0.0275			0.0360	0.0260

**Table 3** Engineering formulae used to estimate mean leaf boundary layer conductance and values of the empirical “correction” factor adopted for each forest.

Site	Geometric shapes representing leaves	Leaf characteristic dimension (mm)	Leaf boundary layer conductance model ( $\text{m s}^{-1}$ )	Enhancement / reduction factor
Les Landes	cylinder	1.5	$\overline{g_{lv}} = 0.0778u^{0.47}$	0.40
Espirra	flat plate	18.0	$\overline{g_{lv}} = 0.0502u^{0.5}$	1.38
Carrasqueira	cylinder	1.5	$\overline{g_{lv}} = 0.0778u^{0.47}$	0.40
Kenya	flat plate	28.2	$\overline{g_{lv}} = 0.0623u^{0.5}$	1.38

**Table 4** Minimum and maximum values for Gash's analytical model parameters used in Morris screening for the different sites.

Parameter	Site					
	Les Landes			Les Landes		
	Espirra	Amazonia	Kenya	Espirra	Amazonia	Kenya
	Carrasqueira			Carrasqueira		
	Minimum values			Maximum values		
$c$	0.4	0.9	0.02	0.8	1	0.6
$S$	0.15	0.7	0.7	0.5	1	1
$\bar{R}$	1.5	4	0.5	2.2	6	3.2
	Minimum values common to all sites			Maximum values common to all sites		
$\bar{E}$		0.15			0.33	
$S_t$		0.01			0.2	
$p_d$		0.005			0.04	
$e$		0.01			0.03	

**Table 5** Mean evaporation rates determined in the original studies ( $\bar{E}_{PM}$ ) and using the wet bulb approach ( $\bar{E}_{WB}$ ). For the forests where the estimates are different, interception loss results are also presented (originally measured and modelled interception loss and new simulations based on  $\bar{E}_{WB}$  estimates through the revised version of Gash's analytical model (Valente et al., 1997)). For all the estimates of interception loss the respective normalized mean errors are between brackets.

	Site				
	Les Landes	Espirra	Carrasqueira	Amazonia	Kenya
Original studies					
$\bar{E}_{PM}$ (mm hr <sup>-1</sup> )	0.170	0.200	0.315	0.210	0.230
$I$ (mm) observed	73	101	154	428	161
$I$ (mm) modelled	70 (-0.041)	98 (-0.03)	157 (0.019)	543 (0.269)	128 (-0.205) (a) 154 (-0.043) (b)
$\bar{E}_{WB}$ (mm hr <sup>-1</sup> )	0.383	0.774	0.315	0.316	0.232
	$L^*$ value for the whole canopy at each site used for estimating $\bar{E}_{WB}$				
	2.3	3.2	2.7	6.6	variable
Actual study					
$\bar{E}_{WB}$ (mm hr <sup>-1</sup> )	0.223	0.203	0.315	0.178	0.232
	$L^*$ value for the canopy layer considered at each site for estimating $\bar{E}_{WB}$				
	1.34	0.83	2.7	2.52	variable
Canopy layer	1/3 top	1/3 top	whole canopy	1/2 top	whole canopy
$I$ (mm) modelled	76 (0.041) (c)		491 (0.147)		

(a) estimate obtained using the global  $\bar{R}$  and (b) estimate based on monthly  $\bar{R}$  values; (c) simulation for a slightly different (higher) gross rainfall total of 613 mm corresponding to the period 09 February 1986 – 03 January 1987, excluding the period of 13 March – 14 April 1986 when some data loss occurred.

1 Rainfall interception modelling: is the wet bulb approach adequate to estimate mean  
2 evaporation rate from wet/saturated canopies in all forest types?

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5 Figure Captions

6

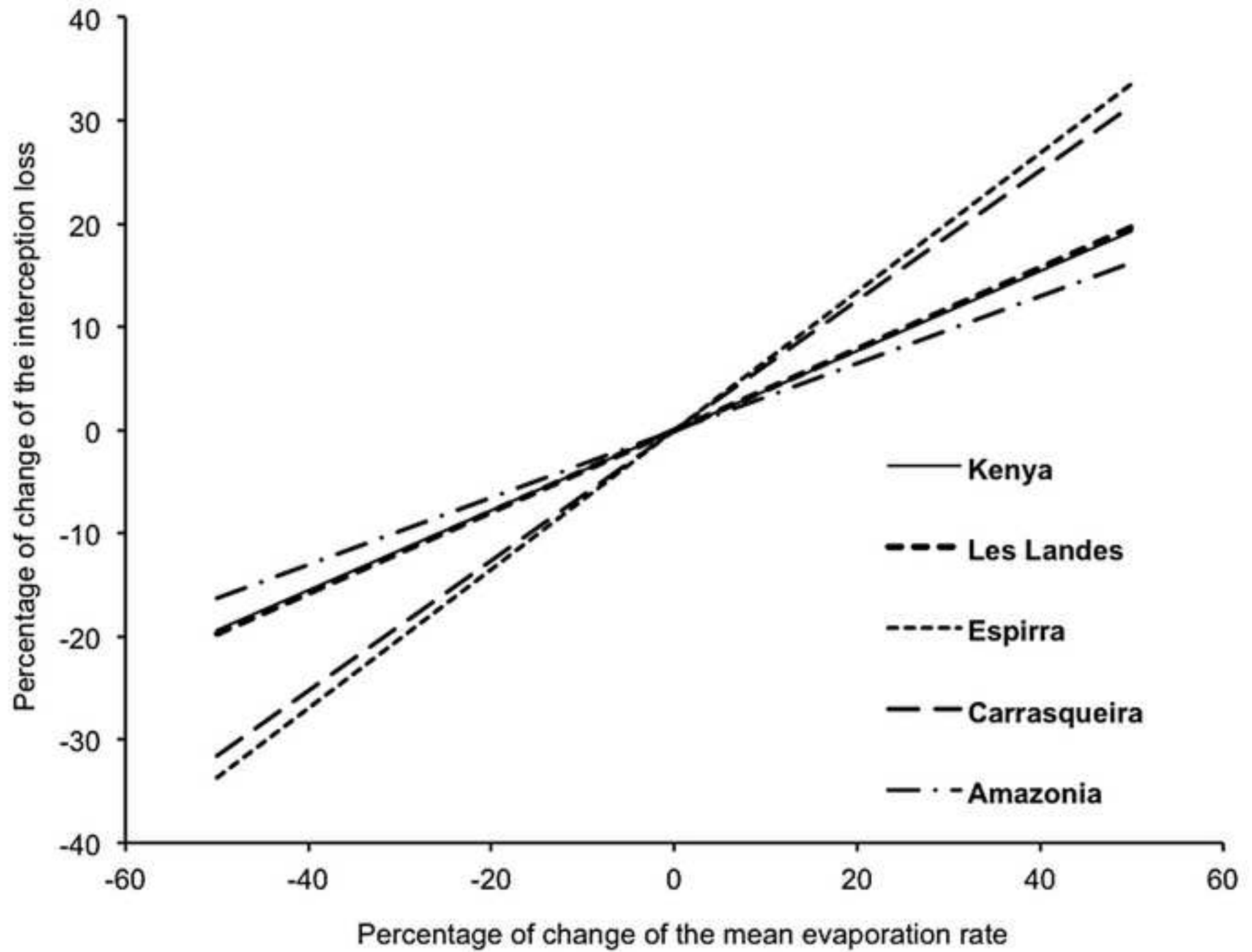
7 **Figure 1** Local sensitivity analysis for  $\bar{E}$  measured by the influence of the percentage change in this  
8 parameter on the percentage change in the interception loss simulated by the sparse version of  
9 Gash's analytical model, using the data sets of the five experiments.

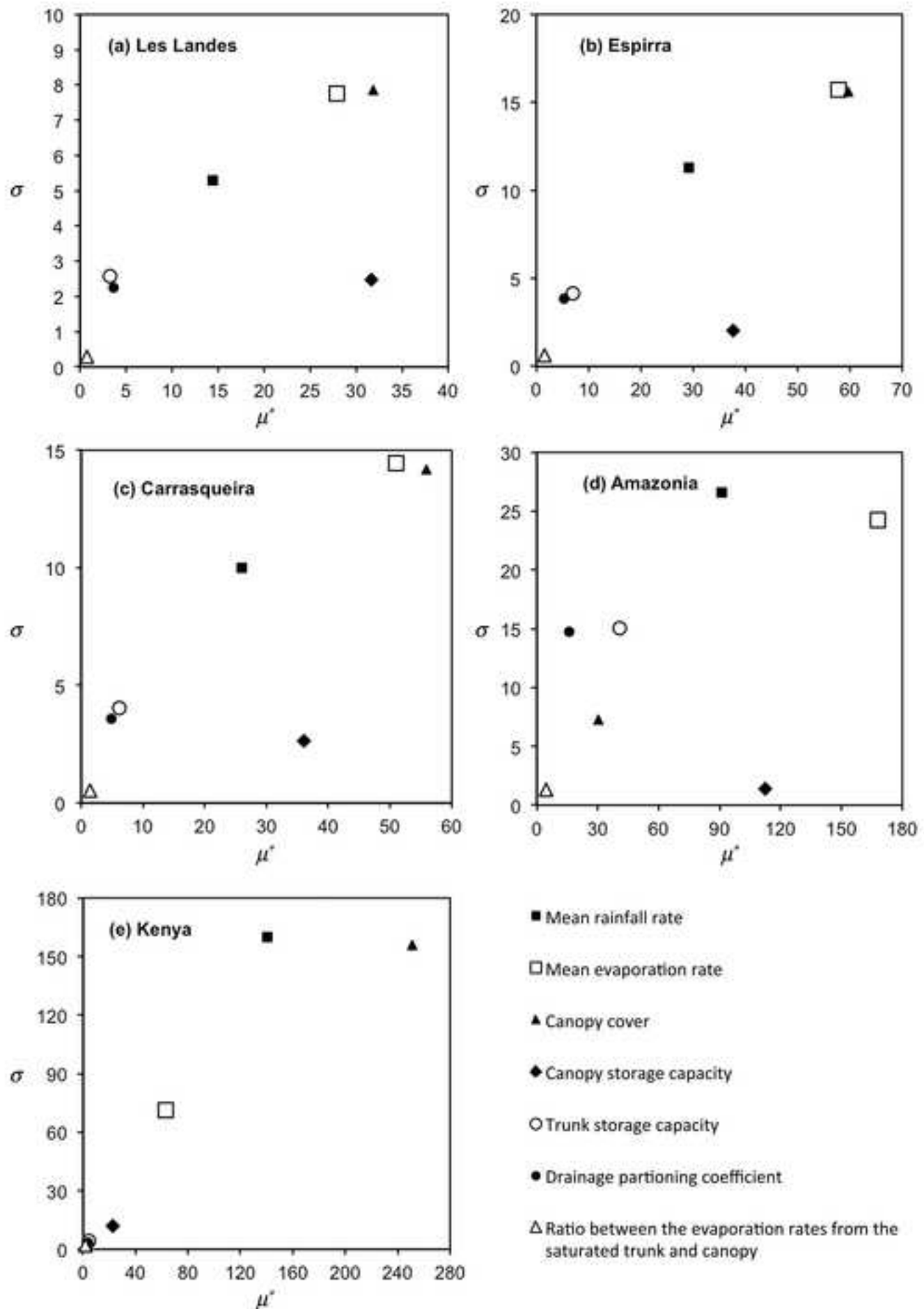
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12 **Figure 2** Plots of Morris sensitivity measures  $\mu^*$  and  $\sigma$  for the seven parameters of the sparse version  
13 of Gash's analytical model: mean rainfall rate ( $\bar{R}$ ), mean evaporation rate ( $\bar{E}$ ), canopy cover ( $c$ ),  
14 canopy storage capacity ( $S$ ), trunk storage capacity ( $S_t$ ), drainage partitioning coefficient ( $p_d$ ) and ratio  
15 between the evaporation rates from the saturated trunk and canopy ( $e$ ) Each graph was obtained with  
16 a different data set: (a) Les Landes (pine), (b) Espirra (eucalyptus), (c) Carrasqueira (pine), (d)  
17 Amazonia (rainforest) and (e) Kenya (agroforestry).

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Rainfall interception modelling: is the wet bulb approach adequate to estimate mean evaporation rate from wet/saturated canopies in all forest types?

#### Highlights

- Saturated crowns of individual sparse trees behave as wet bulbs
- Evaporation from fully ventilated canopies is well estimated by the wet bulb approach
- When applicable, this approach may be preferable to the Penman-Monteith model
- Fully ventilated canopy conditions do not depend solely on crown cover fraction
- Proper evaluation of wet canopy evaporation is critical to Gash's interception model