

1 Modelling groundwater rebound in recently abandoned  
2 coalfields using DInSAR

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14

15 **Abstract**

16 Advances in differential interferometric synthetic aperture radar (DInSAR) processing  
17 algorithms, such as the Intermittent Small Baseline Subset (ISBAS), and increased data  
18 availability from SAR systems, such as Sentinel-1, provide the opportunity to increase the  
19 spatial and temporal density of ground deformation measurements. Such measurements,

20 when combined with modelling, have the potential to make a significant cost-effective  
21 contribution to the progressive abandonment strategy of recently closed coalfields.  
22 Applications of DInSAR over coalfields have observed heave in coal measures rocks and  
23 temporal correlations between the rise of mine water and deformation time-series. The  
24 cessation of systematic dewatering can have a variety of detrimental impacts and knowledge  
25 of the time-scales (i.e. the rate of rebound) and structure of the mine system are crucial to the  
26 remediation strategy. Although mine plans and borehole measurements provide vital  
27 information in this regard, mine plans are often incomplete or inaccurate, whereas  
28 monitoring boreholes are spatially sparse. Consequently, groundwater can flow in  
29 unanticipated directions via goaf, mine shafts and roadways, making it difficult to  
30 determine where the impacts of rebound are likely to occur. In this study, ground  
31 deformation data obtained using ISBAS DInSAR on ENVISAT (2002 – 2009) and Sentinel-1  
32 (2015 – 2019) data are used to develop a simple method to model groundwater rebound in  
33 abandoned coalfields. A forward analytical model based upon the principle of effective  
34 stress and mine water ponds is first implemented to estimate surface heave in response to  
35 changes in groundwater levels measured in monitoring boreholes. The forward model is  
36 then calibrated and validated using the ground deformation data. The DInSAR data were  
37 subsequently inverted to map the change in groundwater levels in greater detail across the  
38 coalfield and forecast surface discharges in order to support mitigation strategies.

39

40 **Keywords:**

41 Surface Deformation; DInSAR; Intermittent SBAS; Groundwater Modelling; Coal Mining;  
42 Dewatering; Heave; Hydrogeology

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44 **1. Introduction**

45 Deep underground coal mining almost always extends below the natural water table  
46 (Younger, 2016). Accordingly, extensive dewatering of coal measures rocks is required to  
47 ensure dry and safe working conditions in collieries, with strategic pumping stations located  
48 around the complex network of mine workings (Dumpleton *et al.*, 2001). After colliery  
49 closure, the cessation of dewatering results in the progressive rise of groundwater and  
50 flooding of old workings and surrounding rock, as the rock matrix begins to revert back to  
51 saturated conditions. Acid mine water drainage (AMD) is water that becomes polluted  
52 through contact with former workings. AMD can have many adverse consequences, which  
53 are listed in order of environmental significance and frequency (Younger & Adams, 1999):

- 54 (i) surface water pollution;
- 55 (ii) localised flooding of agricultural, industrial or residential areas;
- 56 (iii) loss of dilution for other pollutants in surface waters where former pumped  
57 discharges have ceased;
- 58 (iv) overloading and clogging of drains and sewers;
- 59 (v) pollution of overlying aquifers by upward movement of mine water (principally  
60 from sulphate and chloride);
- 61 (vi) temporarily increased emissions of mine gases, driven ahead of rising mine water;
- 62 (vii) ground deformation due to renewed mining subsidence and reactivation of faults  
63 (Donnelly, 2006);
- 64 (vii) adverse effects on landfills – possible damage to lining, leakage of leachate and  
65 increased gas emissions.

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67 Methods to estimate the spatial extent and timescales (i.e. the rate of rebound) of mine water  
68 rise are vital to generate rational policies and a cost-benefit remediation strategy for  
69 progressive mine abandonment (Younger & Adams, 1999; Younger, 2016). Mine plans  
70 provide crucial information in the prediction of the directional and structural control of  
71 groundwater flow. When above the water table, groundwater flows down-dip through  
72 permeable stratigraphic horizons, such as the fractured or jointed sandstones, goaf,  
73 collapsed strata and mining roadways. Down-dip flow will stop and groundwater will pond  
74 once it reaches a hydraulic barrier such as an intact coal pillar or fault. Water levels will rise  
75 until a new overflow point is reached in the form of a drainage adit at the surface or a  
76 connection to a neighbouring colliery. However, the prediction of flow is often challenging  
77 due to incomplete and inaccurate information regarding inter-seam connections within  
78 collieries, as well as connections between adjacent collieries from mine plans that might  
79 have been drawn more than a century ago. Flow can therefore occur in unanticipated  
80 directions which can lead to surface discharges and pollution in areas not thought to be  
81 vulnerable (Younger & Adams, 1999). In addition, monitoring boreholes are expensive and  
82 often spatially sparse, providing only a limited overview of mine water levels in complex  
83 and often vast mine systems.

84 Many studies have recognised the impact of changing groundwater levels on the  
85 compaction of strata and deformation at the surface (e.g. Hoffman *et al.*, 2001; Galloway &  
86 Hoffmann, 2007; Bell *et al.*, 2008, Chaussard *et al.*, 2017, Motagh *et al.*, 2017). As groundwater  
87 levels change, compaction or expansion of the subsurface strata occurs due to a change in  
88 hydrostatic pressure (Bekendam & Pöttgens, 1995). Conventional monitoring methods for

89 mapping surface deformation, such as precise levelling, close range photogrammetry and  
90 GPS, can be laborious, expensive and offer limited spatial coverage. Remote differential  
91 interferometric synthetic aperture radar (DInSAR) is a cost-effective wide-area method to  
92 measure sub-centimetre rates of surface deformation. In particular, the emergence of time-  
93 series methods in the late 1990s has increased the integration of surface deformation  
94 measurements into hydrogeological studies. DInSAR data facilitate the identification of  
95 areas of groundwater depletion or recharge, and so provide the means to calibrate  
96 groundwater models, delineate lithological boundaries and map aquifer storage variations  
97 and assist characterisation (e.g. Chaussard *et al.*, 2014; Castellazzi *et al.*, 2016; Bejar-Pizarro *et*  
98 *al.*, 2017; Castellazzi *et al.*, 2018). Additionally, DInSAR measurements have previously been  
99 utilized to calibrate and/or validate models of surface deformation associated with  
100 anthropogenic fluid injection or extraction (e.g. Rutqvist *et al.*, 2010; Pearse *et al.*, 2014; Gee *et*  
101 *al.*, 2016). Initial applications over mining areas also confirmed the capability of DInSAR to  
102 measure mining induced subsidence (e.g. Carnec *et al.*, 1996; Wright & Stow, 1999). More  
103 recently, spatial distributions of heave in coal measures rocks (e.g. Sowter *et al.*, 2013;  
104 Bateson *et al.*, 2015; Sowter *et al.*, 2018; Gee *et al.*, 2019) and temporal correlations between the  
105 rise of mine water and deformation time-series (e.g. Cuenca *et al.*, 2013; Gee *et al.*, 2017) have  
106 been observed over abandoned coalfields.

107 The monitoring of groundwater rebound in abandoned coalfields requires a dense, regular  
108 spatiotemporal sampling of ground deformation measurements. This is because the  
109 response to hydrostatic pressure changes caused by pumping can be highly spatially  
110 variable due to differences in the compressibility, thickness and confinement of the aquifer  
111 volume (Castellazzi *et al.*, 2018). However, the obtained spatial distribution and density of

112 DInSAR measurements can be severely affected by phase decorrelation or incoherence  
113 (Crosetto *et al.*, 2010). This effect is most prevalent in higher frequency radar bands (e.g. C-  
114 band) and incoherence is pervasive over agricultural land, forests, semi-natural areas and  
115 wetlands. This is particularly constraining for both persistent scatterer interferometry (PSI)  
116 and small baseline techniques. Small baseline methods have proven to achieve meaningful  
117 results in areas typically unfavourable for DInSAR analysis (e.g. Lu & Kwoun, 2008),  
118 however, the density and distribution of measurements is still limited (Osmanoğlu *et al.*,  
119 2016). Furthermore, data availability has been an additional historical limitation where it is  
120 crucial to have sufficient data acquisitions spanning a deformation event at an appropriate  
121 revisit time for a reliable analysis.

122 The potential of DInSAR has continued to improve over time with many advancements in  
123 processing algorithms and increasing data availability from SAR orbital systems. In this  
124 regard, two notable advances are the development of the Intermittent Small Baseline Subset  
125 (ISBAS) (Sowter *et al.*, 2013; Sowter *et al.*, 2016) processing method and the launch of the  
126 Sentinel-1 satellite. The ISBAS method is an adapted version of the SBAS method (Berardino  
127 *et al.*, 2002) that is capable of computing velocities over unfavourable land cover types to  
128 return a spatially distributed set of deformation measurements with near complete ground  
129 coverage (Cigna & Sowter, 2017). Sentinel-1 is a two-satellite constellation carrying a C-Band  
130 (5.5 cm wavelength — 5.4 GHz) SAR payload which provides conflict-free data every 6 days  
131 in Europe and at least 12 days for the majority of the global landmasses (Torres *et al.*, 2012).  
132 Characterising and predicting the rise of AMD is a global challenge facing parts of Europe  
133 (Gee *et al.*, 2017), Africa (van Tonder *et al.*, 2007), Asia (Liu *et al.*, 2020), Australia (Wright *et*  
134 *al.*, 2010) and the Americas (Gammons *et al.*, 2010), particularly as the demand for coal

135 declines in favour of more sustainable energy sources. With this in mind, this study aims to  
136 propose a cost-effective method to model groundwater rise in recently abandoned coalfields  
137 using DInSAR for the purpose of identifying where problems associated with mine water  
138 rebound may manifest. The specific objectives are to:

- 139 (i) generate a forward analytical model to predict the change in thickness of the  
140 strata and, hence, determine surface movement, based upon measurements of  
141 groundwater change from monitoring boreholes;
- 142 (ii) calibrate and validate the model using ISBAS DInSAR;
- 143 (iii) invert the DInSAR measurements to provide an estimate of the change in  
144 groundwater levels with detailed and near-complete spatial coverage across the  
145 coalfield;
- 146 (iv) utilize the inverted rates to estimate the time it will take for groundwater to  
147 discharge out of the Coal Measures rock.

148 These objectives provide the structural sub-headings used in the following Methods and  
149 Results and Discussions sections.

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## 152 **2. Study Site & Materials**

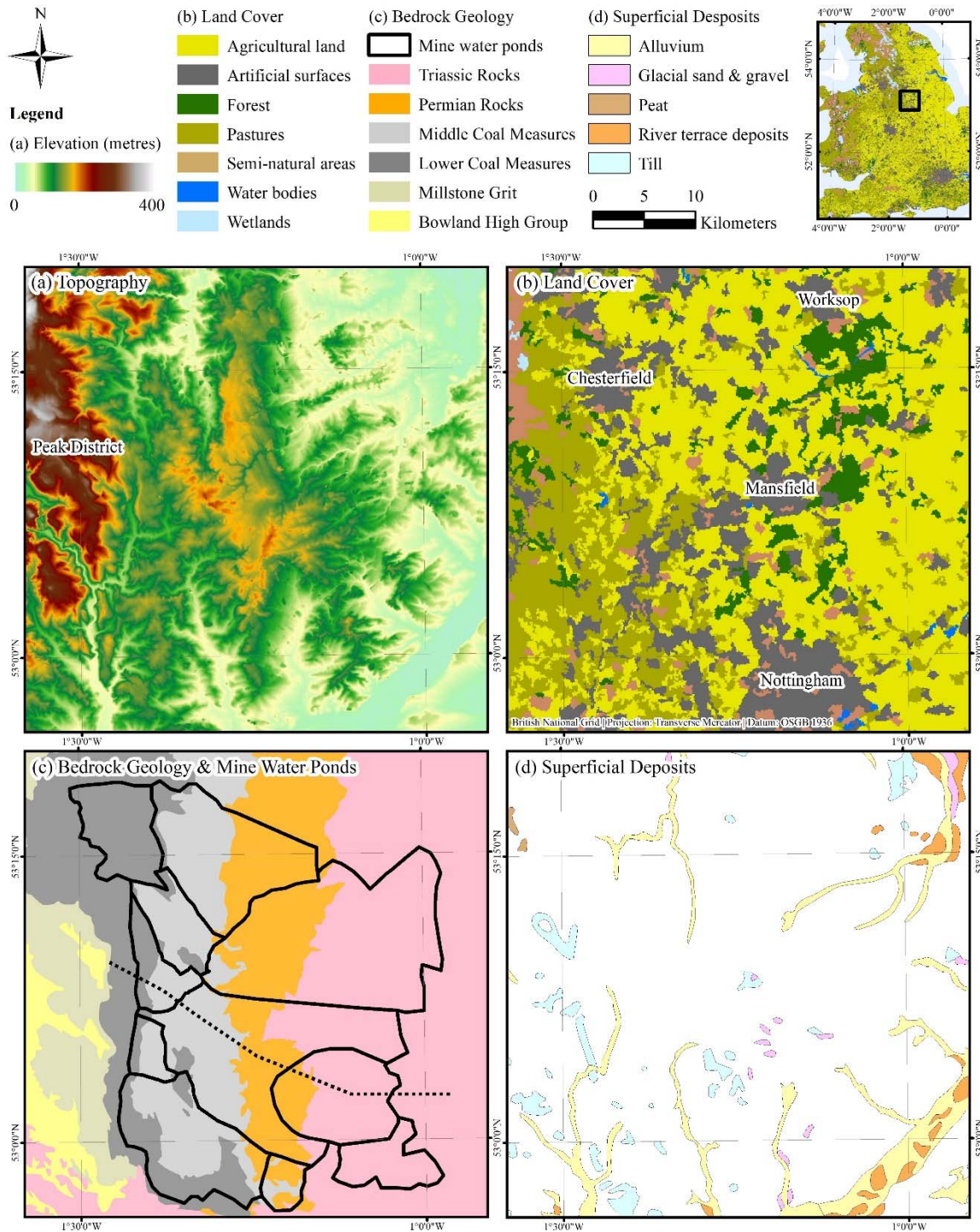
### 153 **2.1. Land Cover**

154 The study site comprises ~2000 km<sup>2</sup> and covers the counties of Derbyshire and  
155 Nottinghamshire, UK (Figure 1). Land cover is largely rural, dominated by agricultural land  
156 (50%) with patches of pastures (21%), forests (7%), semi-natural areas (<2%), and wetlands

157 (<1%) (European Environment Agency, 2012). Artificial surfaces (20%) predominantly  
158 correspond to the city of Nottingham in the south and towns of Mansfield and Chesterfield  
159 in the centre and north-west, respectively. Water bodies make up less than 1% of the total  
160 land cover. Topographically, the study site is located between the Peak District at the  
161 southern end of Pennines hill and mountain range in the west, where the elevation reaches a  
162 maximum of 380 m above ordnance datum (AOD), and the midlands plains to the east at  
163 mean sea level.

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**Figure 1.** (a) Topography – NEXTMap® DTM at 10 m resolution; (b) CORINE land cover inventory (European Environment

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Agency, 2012); (c) Bedrock geology at 1:625,000 scale, from BGS Geology 625k (DiGMapGB-625) data; and (d) Superficial

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deposits at 1:625,000 scale, from BGS Geology 625k (DiGMapGB-625) data. The dashed black line in (c) shows the location of

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the cross-section in Figure 2. NEXTMap® Britain © 2003, Intermap Technologies Inc. European Environment Agency © 2012.

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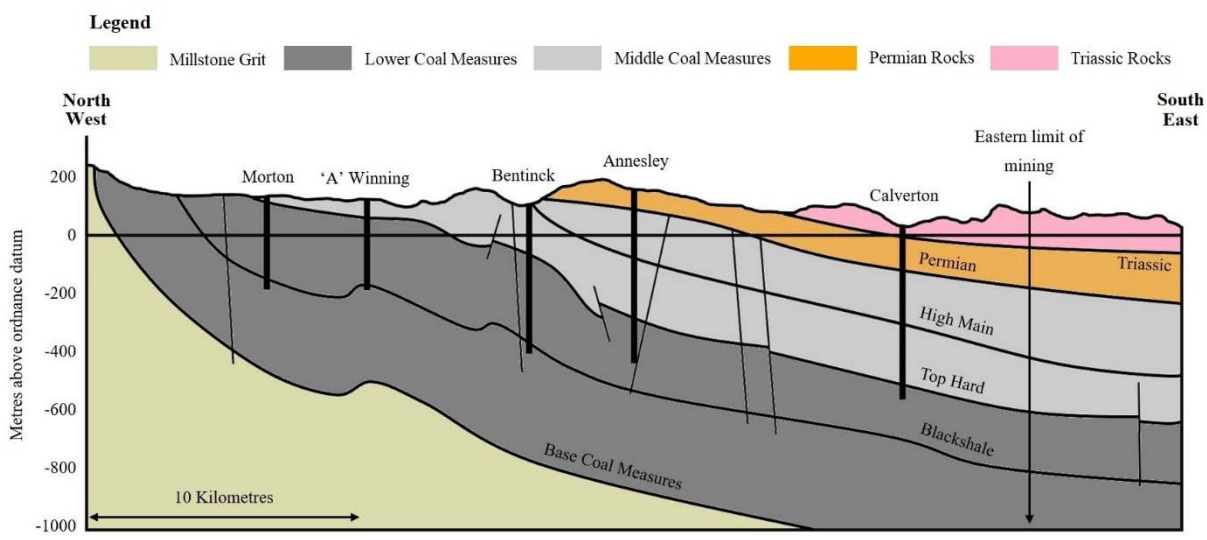
## 174 **2.2. Geology**

175 The study area lies to the east of the Pennine Dinantian-Namurian anticlinal ridge (Figures 1  
176 and 2). The Coal Measures Group (CMG) are dominated by argillaceous strata with  
177 alternations of sandstone, grey siltstone and grey mudstone with frequent coal seams and  
178 seatearth horizons. Groundwater movement largely occurs via secondary permeability  
179 through fractures, given that the carboniferous strata are extensively faulted which act to  
180 either increase permeability or prevent hydraulic connectivity (Holliday, 1986). The  
181 Nottinghamshire Coalfields have been extensively worked since the early 1800s until the last  
182 deep mine, Thoresby, was closed in July 2015. Over thirty separate seams have been mined,  
183 and the extraction of large volumes of coal and adjacent rock, associated fracturing and  
184 collapse of *in situ* strata, generation of roadways and entry shafts, and areas of goaf generate  
185 a complex combination of disturbed and undisturbed strata. The most exploited horizon is  
186 the Top Hard seam and neighbouring collieries are frequently linked via the working of this  
187 seam from different directions (Dumpleton *et al.*, 2001). The CMG dip regionally east under  
188 the East Midlands shelf, which is formed of Permian and Triassic strata and ranges between  
189 50 m to 200 m in thickness from west to east. The bedrock is overlain with superficial  
190 deposits that are only up to a few metres thick (Charsley *et al.*, 1990).

191 The geological parameters required for establishing a groundwater model are the porosity  
192 and compression index of the CMG. The porosity was determined as 0.134, calculated as the

193 mean of 236 measurements of the Pennine CMG from the National Geotechnical Properties  
 194 Database of the British Geological Survey (BGS). No laboratory samples were available to  
 195 determine the compression index, therefore a value for dense sand of 0.005 was defined  
 196 according to the literature (Jain *et al.* 2015). In addition, a digital elevation model  
 197 (NEXTMap® DTM) and the 3D extent of the top of the Middle Coal Measures formation  
 198 (Figure 2) was required, with the latter taken from the BGS 1:250,000 3D geological models  
 199 of the Pennine Basin (Hulbert & Terrington, 2014a; b).

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203 **Figure 2.** Cross-section of the Nottinghamshire Coalfield. The location is marked on Figure 1c. The thick vertical  
 204 black lines represent shaft locations and thinner off vertical black lines represent faults. The high main seam  
 205 demarcates the upper limit of mining. The base of the Top Hard and Blackshale seams for the areas east and west  
 206 of Annesley-Bentinck, respectively, indicates the base of the zone of enhanced permeability due to mining  
 207 induced fractures (adapted from Dumpleton *et al.*, 2001).

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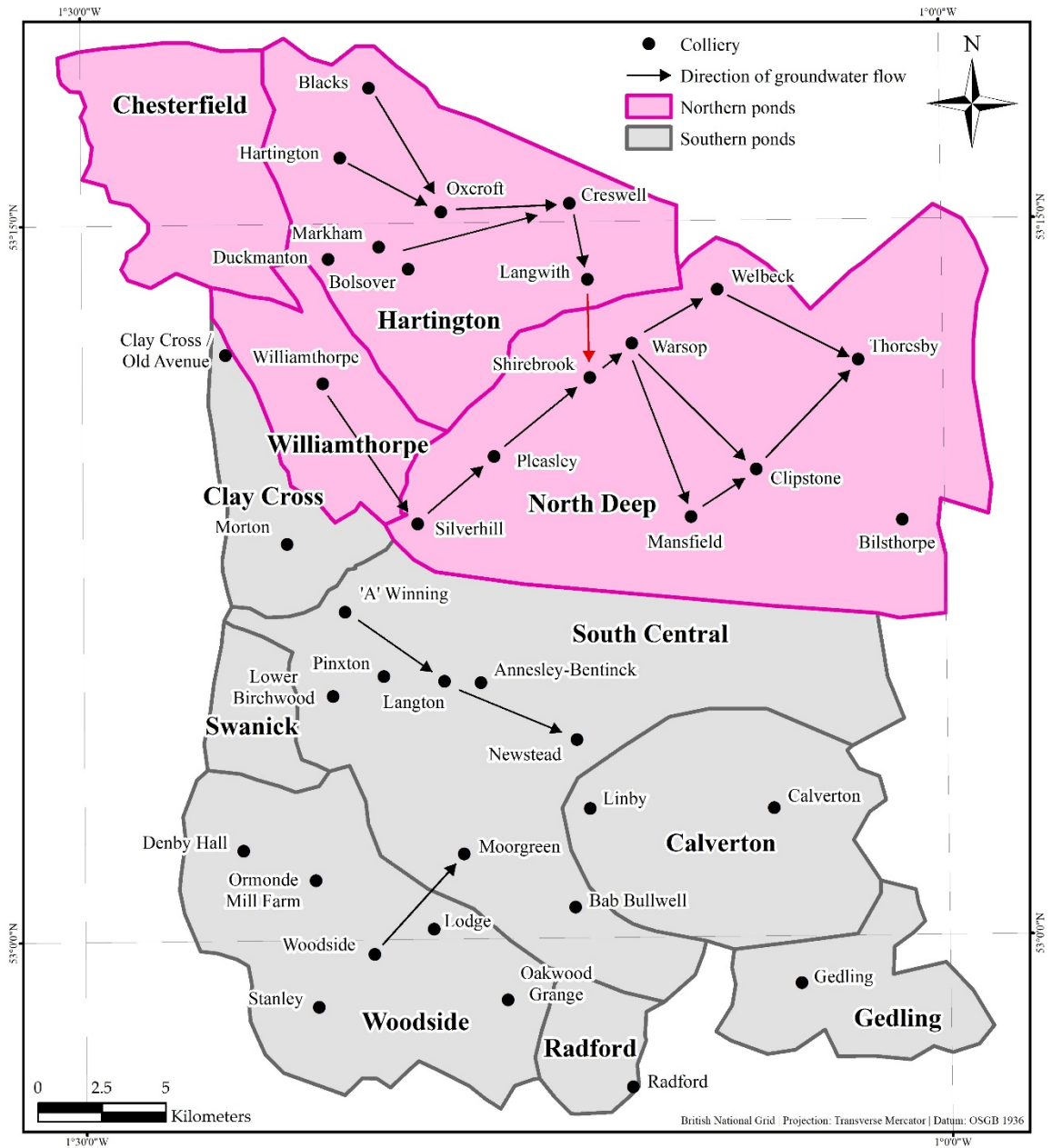
209 **2.3. Hydrogeology**

210 Mining activities have substantially altered the permeability of the CMG, whereby hydraulic  
211 connectivity has been generated between previously isolated horizons and abandoned  
212 saturated workings (Banks, 1997). The Nottinghamshire Coalfield is divided by the NW–SE  
213 Hardsoft-Mansfield anticline into northern and southern fault-bound sections, which can be  
214 further sub-divided into mine water ponds (Figure 3). Younger & Adams (1999) introduced  
215 the Groundwater Rebound in Abandoned Mineworkings (GRAM) modelling concept,  
216 within which the coalfield is simplified into discrete ponds that represent groups of mine  
217 workings that are – with respect to rebound – hydraulically analogous. Connectivity within  
218 the ponds can be extensive and, hence, these are expected to subside or uplift uniformly at  
219 the surface dependant on the associated groundwater regime. Some of the ponds are known  
220 to contain sub-ponds which have not been resolved due to a lack of mine water information.  
221 Hydraulic connectivity between ponds occurs at different levels between permeable  
222 geological features, mining roadways, existing boreholes and areas where adjoining goaf  
223 panels collapse. As the groundwater rises, the ponds fill until a connection to another pond  
224 is reached and water decants into the neighbouring pond. Connections between ponds alter  
225 at depth and therefore ponds can become isolated or connected at different stages of  
226 abandonment. Figure 3 shows the expected schematic of groundwater flow. The workings of  
227 the deepest mines in the east were very dry, however, they are connected via an extensive  
228 network of workings up-dip which receive considerable ingress from rainfall and surface  
229 runoff (Rae, 1978). Little is known about the amount of flow, if any, between the deepest  
230 mines, since mine water accumulating in shallower workings was pumped to prevent the  
231 build-up of water against hydraulic barriers and therefore reduce the risk of inrushes.

232 Hydrogeological data was provided by the Coal Authority and was available for 21 and 24  
233 boreholes coinciding with the ENVISAT and Sentinel-1 periods, respectively (Table 1). The  
234 number and frequency of groundwater measurements vary and where borehole data was  
235 incomplete across the epoch, it was linearly interpolated to fit the time-period provided that  
236 it covered more than three quarters of the epoch; otherwise it was discarded. Spatially,  
237 groundwater measurements are relatively sparse, with no information on levels in the  
238 Chesterfield, North Deep, Swanick or Gedling ponds for either time epochs.

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242 **Figure 3.** Mine water ponds and expected schematic of flow of groundwater for the Nottinghamshire Coalfield.

243 The red arrow between Langwith and Shirebrook indicates that at present there is very little or no flow between  
 244 this barrier.

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251 **Table 1.** Groundwater data for the ENVISAT (November 2002 – November 2009) and Sentinel-1 (May 2015 – April 2019) periods.

Coalfield	Pond	Borehole	ENVISAT (2002 – 2009)								Sentinel-1 (2015 – 2019)							
			No. Measurements	Start Height (Metres AOD)	End Height (Metres AOD)	Head Change (Metres)	Coefficient of Determination (R <sup>2</sup> )	Mean Start (Metres AOD)	Mean End (Metres AOD)	Mean Head Change (Metres)	No. Measurements	Start Height (Metres AOD)	End Height (Metres AOD)	Head Change (Metres)	Coefficient of Determination (R <sup>2</sup> )	Mean Start (Metres AOD)	Mean End (Metres AOD)	Mean Head Change (Metres)
North	Chesterfield	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	
	Hartington	Blacks	77	2.6	2.8	0.2	0.09	-139.4	-120.3	19.1	12	2.7	15.9	13.2	0.65	-82.3	-29.4	52.9
		Hartington	288	-140.9	-142.6	-1.7	0.16				30	-58.5	14.5	72.9	0.98			
		Duckmanton	39	69.7	69.8	0.1	0.24				-	-	-	-	-			
		Markham	3	-341.4	201.1	140.3	0.98				-	-	-	-	-			
		Oxcroft No.2	161	1.5	-11.8	-13.2	0.82				5	-10.2	15.9	26.1	0.97			
		Creswell	162	-300.3	-301.4	-1.1	0.26				4	-172.5	-95.0	77.6	0.97			
		Langwith	158	-266.9	-257.9	9.0	0.31				23	-172.9	-98.4	74.5	0.99			
Williamthorpe	Williamthorpe	280	-143.7	-147.4	-3.7	0.00	-143.7	-147.4	-3.7	11	-49.0	20.3	69.3	0.91	-49.0	20.3	69.3	
North Deep	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-		
South	Clay Cross	Morton	183	33.8	78.3	44.5	0.95	57.9	78.5	20.6	25	88.4	84.8	-3.6	0.16	82.6	82.2	-0.4
		Clay Cross 1	168	56.5	77.4	20.9	0.94				12	79.9	81.7	1.8	0.23			
		Clay Cross 2	161	56.5	76.9	20.4	0.94				12	79.5	79.5	0.0	0.20			
		Old Avenue 3	939	65.2	78.1	12.9	0.74				19	80.3	80.2	-0.1	0.02			
		Old Avenue 4	166	77.4	81.8	4.4	0.55				21	84.9	85.0	0.1	0.18			
	South Central	'A' Winning	162	-41.8	44.3	86.1	0.99	-175.5	-21.1	154.4	162	47.3	36.1	-11.2	0.02	23.5	35.6	12.1
		Lower Birchwood	-	-	-	-	-				10	55.6	55.1	-0.5	0.15			
		Pinxton	-	-	-	-	-				142	27.8	44.7	16.9	0.11			
		Langton	82	-309.2	-86.5	222.7	0.98				155	25.7	44.5	18.8	0.97			
		Newstead	-	-	-	-	-				146	22.5	41.1	18.5	0.96			
		Moorgreen	-	-	-	-	-				141	10.4	34.6	24.2	0.84			
	Woodside	Babington Bulwell	-	-	-	-	-	46.0	50.9	4.9	132	-25.1	-7.0	18.1	0.83	67.4	66.7	-0.7
		Denby Hall	788	78.2	85.5	7.2	0.22				13	88.8	94.1	6.1	0.33			
		Ormonde (Mill Farm)	565	64.4	68.7	4.3	0.64				9	70.9	76.3	5.4	0.69			
		Woodside	301	16.5	0.3	-16.2	0.00				82	38.5	10.9	-27.6	0.34			
		Stanley	1098	53.5	58.9	5.4	0.27				17	91.7	101.6	9.9	0.20			
	Oakwood Grange	1231	17.3	41.0	23.7	0.74	14	47.9	50.6	2.7	0.17							
	Radford	Radford	-	-	-	-	-	-	-	-	14	25.1	25.1	0.0	0.00	25.1	25.1	0.0
Calverton	Calverton	38	-408.9	-377.9	30.0	0.94	-408.9	-377.9	31.0	-	-	-	-	-	-	-	-	
Gedling	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	

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259 **Table 2.** Earth Observation data and ISBAS processing parameters.

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Satellite	Geometry	Track	N° of Images	Date Range	Median Revisit (Days)	Max. Orbital Baseline (Metres)	Min. Temp. Baseline (Days)	Max. Temp. Baseline (Days)	Mean Temp. Baseline (Days)	Multilooking (Azimuth: Range)	Coherence Threshold	N° of Interferograms	Interferogram Threshold	Reference Point
ENVISAT	Desc.	366	40	30/Nov/2002 – 28/Nov/2009	35	250	0		618	20 : 4		240	70	Nottingham City Centre
Sentinel-1	Desc.	154	185	08/May/2015 – 17/Apr/2019	6	45	365	1460	673	6 : 21	0.25	3603	370	
	Asc.	132	175	06/May/2015 – 15/Apr/2019		50			674			3558		

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265 **2.4. Earth Observation Data**

266 Thirty-nine ENVISAT C-Band Stripmap Level-1 single look complex (SLC) Image Mode  
267 products in VV polarisation from descending track 366 were available over the study site  
268 (Table 2). There were insufficient acquisitions from the ascending geometry for a reliable  
269 analysis. The descending data has a median revisit time of 35 days over a seven-year period  
270 (November 2002 – November 2009) regularly distributed throughout the epoch. The  
271 incidence angle ranges between  $20.1^{\circ}$  and  $25.9^{\circ}$  from near to far range with respect to the  
272 surface normal. The products have a pixel spacing of 8 m in slant range and 4 m in azimuth,  
273 corresponding to an approximate ground spatial resolution of 25 m in range and 5 m in  
274 azimuth. In addition, 185 Sentinel-1 C-Band Terrain Observation with Progressive Scans  
275 SAR Interferometric Wide (IW) SLC VV polarized images from descending track 154 and  
276 176 images from ascending track 132 (May 2015 – April 2019) were utilized (Table 2). The  
277 median revisit time is much reduced with respect to the ENVISAT data at 6 days. The  
278 Sentinel-1 IW products have a pixel spacing 2.3 m in slant range and 13.9 m in azimuth,  
279 corresponding to a spatial resolution 5 m in ground range and 20 m in azimuth at scene  
280 centre. The incidence angle varies from  $29.1^{\circ}$  to  $46^{\circ}$  across the swaths from near to far range.

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285 **3. Methods**

286 **3.1. Generation of Forward Model using Borehole Measurements**

287 The forward model applies the principle of effective stress (Terzaghi, 1925) and mine water  
288 ponds (Younger & Adams, 1999) to calculate the increase in bed thickness (i.e. heave) that  
289 occurs for a given rise in groundwater level. Separate models were set up for the ENVISAT  
290 (2002 – 2009) and Sentinel-1 (2015 – 2019) epochs since the two stacks of SAR data were  
291 processed separately. The model is treated as a homogeneous matrix, where the initial bed  
292 thickness ( $b_0$ ) (m) was calculated as the depth from the surface to the groundwater level at  
293 the start of the modelling epoch (Table 1). The bed thickness was interpolated for each pond  
294 based on the borehole measurements. The structure of mines or any of the interseam  
295 connections were not considered as their inclusion would introduce an undesired level of  
296 complexity to the model that would exceed the scope of the simple approach presented here.  
297 As groundwater levels fluctuate over the time-epochs but remain confined to the CMG unit,  
298 the model calculates the change in bed thickness of only the CMG unit. It does not consider  
299 any compaction of the overlying Permo-Triassic formation. For each pond to determine the  
300 rise in groundwater, or change in piezometric head ( $\Delta h$ ) (m), across the coalfield over the  
301 modelling epoch the borehole measurements were interpolated (Table 1). For ponds where  
302 only one measurement was available, a uniform surface was utilized.

303 The strata is subject to a level of geostatic pressure ( $p$ ) (kPa), which increases as more  
304 material is overlain over time. Geostatic pressure is resisted by the intergranular (effective)  
305 stress ( $p_{s0}$ ) of the rock matrix and the fluid pressure of pore water ( $p_{w0}$ ) (Poland, 1984):

$$306 \quad p = p_{s0} + p_{w0} \quad (\text{Eqn. 1})$$

307 Equilibrium must be maintained in Eqn. 1, thus, an increase in piezometric head increases  
308 the pore fluid pressure and decreases the effective stress on the strata. This results in  
309 expansion of the strata until equilibrium is again reached. The stress transfer from fluid to

310 rock matrix per unit change in piezometric head was calculated at 10 kPa/m by Poland

311 (1984). The geostatic pressure was calculated from the initial bed thickness ( $b_0$ ) as:

$$312 \quad p = 10 \cdot b_0 \quad (\text{Eqn. 2})$$

313 Poland (1984) estimated for an unconfined aquifer that geostatic pressure is divided as 60%  
314 effective stress and 40% pore fluid pressure, whereas for a confined aquifer this is divided as  
315 75% effective stress and 25% pore fluid pressure. In the western exposed coalfield only a  
316 single formation exists (the CMG), whilst in the east the CMG is overlain by the Permo-  
317 Triassic strata. To account for this, in the western exposed coalfield the CMG is considered  
318 unconfined, whereas it is treated as a confined formation where the coalfield is overlain by  
319 Permo-Triassic rocks. In the confined formation in the thinnest area of cover the pore fluid  
320 pressure is at 25%, which decreased linearly with cover thickness to 10% in the east where  
321 cover is thickest.

322 Following the change in head ( $\Delta h$ ), the new pore fluid pressure ( $p_w$ ) is calculated as:

$$323 \quad p_w = p_{w0} + 10 \cdot \Delta h \quad (\text{Eqn. 3})$$

324 and by maintaining the equilibrium in Eqn. 1, the new effective stress ( $p_s$ ) is:

$$325 \quad p_s = p - p_w \quad (\text{Eqn. 4})$$

326 hence, the change in effective stress ( $\Delta p_s$ ) can be expressed as a function of the initial

327 geostatic pressure (Eqn. 1) and change in piezometric head:

$$328 \quad \Delta p_s = p_s - p_{s0}$$

$$329 \quad \Delta p_s = p - p_w - p_{s0}$$

$$330 \quad \Delta p_s = p - p_{w0} - 10 \cdot \Delta h - p_{s0} \quad (\text{Eqn. 5})$$

331 The initial void ratio ( $e_0$ ) is calculated from the initial porosity ( $n_0$ ):

$$332 \quad e_0 = \frac{n_0}{1 - n_0} \quad (\text{Eqn. 6})$$

333 and after a change in effective stress a new void ratio ( $e$ ) is calculated:

$$334 \quad e = e_0 - c_c \cdot \log\left(\frac{p_s}{p_{s0}}\right) \quad (\text{Eqn. 7})$$

335 as expressed as a function of the initial void ratio ( $e_0$ ), the compression index ( $c_c$ ) and the  
336 initial ( $p_{s0}$ ) and new effective stress ( $p_s$ ). The compression index is a dimensionless parameter  
337 that determines the compressibility of the stratigraphic bed and considers the elastic  
338 properties of the unit.

339 The coefficient of volume compressibility ( $m_v$ ) relates the coefficient of compressibility ( $a_v$ )  
340 and the initial void ratio ( $e_0$ ):

$$341 \quad m_v = \frac{a_v}{1 + e_0} \quad (\text{Eqn. 8})$$

342 where,

$$343 \quad a_v = \frac{\Delta e}{\Delta p_s} \quad (\text{Eqn. 9})$$

344 and  $\Delta e$  is the difference in void ratio.

345 The change in bed thickness ( $\Delta b$ ) is caused by the change in effective stress ( $\Delta p_s$ ) and is  
346 calculated as a function of the coefficient of volume compressibility ( $m_v$ ) and the initial  
347 thickness of the unit ( $b_0$ ):

$$348 \quad \Delta b = s \cdot \Delta p_s \cdot m_v \cdot b_0 \quad (\text{Eqn. 10})$$

349 where  $s$  is a scaling factor to account for predicted inelastic (non-recoverable) deformation.

350 The response of the strata to changes in piezometric head are dependent on historical

351 pressure changes. Small-scale variations in head (e.g. seasonal effects) are elastic and  
352 recoverable so the strata expand and contract in equal measure. When variations in head are  
353 greater, the expansion and contraction is bigger which results in inelastic, and non-  
354 recoverable deformation, therefore limiting future expansion and contraction. Coarse  
355 grained strata (e.g. sand, gravel) are more likely to maintain equilibrium under increased  
356 effective stress due to their rigid skeletal matrix, however, fine-grained material (e.g. clays)  
357 are susceptible to high rates of potential compaction due to their plastic nature (Hiscock,  
358 2009). Given the extensive and often variable dewatering regime that has occurred to great  
359 depths over many decades in the Nottinghamshire Coalfields and dominance of argillaceous  
360 strata within the CMG, it is assumed that the majority of previous compaction is inelastic.

361

362

### 363 **3.2. Calibration & Validation of Forward Model using ISBAS DInSAR Observations**

364 To calibrate the models and calculate the scaling factors, two sets of ground deformation  
365 measurements were generated from the ENVISAT and Sentinel-1 data (Table 2). The three  
366 stacks, processed separately, were co-registered to a common slant-range coordinate system  
367 using an amplitude-based Fast Fourier Transform method (Guizar-Sicairos *et al.*, 2008). In  
368 the case of Sentinel-1, each image was deburst and merged prior to co-registration. Short  
369 orbital perpendicular baseline differential interferograms were generated following the  
370 method proposed by Berardino *et al.* (2002). For ENVISAT, employing a maximum  
371 perpendicular baseline of 250 m and maximum temporal baseline of 4 years generated 270  
372 interferograms with a mean temporal baseline of 618 days (Table 2). For the Sentinel-1

373 analysis, a maximum perpendicular baseline of 45 m and 50 m were applied to the  
374 descending and ascending data sets. The baseline was more stringent for the descending  
375 dataset to ensure a consistent number of interferograms were generated for both stacks  
376 despite the difference in the number of available images. A minimum temporal baseline of 1  
377 year and no maximum was applied to achieve a similar mean temporal baseline as the  
378 ENVISAT stack. A total of 3603 and 3558 interferograms were generated, which have a mean  
379 temporal baseline of 673 and 674 days for the descending and ascending stacks, respectively.

380 An Intermittent Small Baseline Subset (ISBAS) (Sowter *et al.*, 2013; Sowter *et al.*, 2016)  
381 analysis was employed due to the dominance of the non-urban land cover. ISBAS is based  
382 upon the method of Berardino *et al.* (2002), with a modification in the selection of pixels for  
383 analysis. The modification facilitates the retention of pixels which are coherent for only a  
384 subset of the total number of interferograms, which is common outside of urban areas.

385 Pixels that fulfil the selection criteria exhibit a defined level of coherence in a minimum  
386 number of interferograms. The choice of interferogram threshold determines both the spatial  
387 coverage and the precision of the measurement – the standard error. Utilizing an  
388 interferogram threshold equal to the total number of interferograms is the equivalent to the  
389 method of Berardino *et al.* (2002) and minimizes the standard error of velocity, but at the  
390 expense of the density of measurements. The choice of interferogram threshold is a trade-off  
391 between coverage and quality, hence, the inversely proportional relationship between the  
392 interferogram threshold and the standard error (Cigna & Sowter, 2017). A coherence  
393 analysis was employed on the small baseline interferograms, considering all of the image  
394 permutations meeting the baseline criteria. No spatial filtering was applied, and coherence  
395 was calculated over a multi-looked window of 4 in range and 20 in azimuth for ENVISAT

396 and 21 in range and 6 in azimuth for Sentinel-1. In ground range, the resultant pixels have  
397 an approximate pixel spacing of 90 m. A minimum of 70 and 370 interferograms that  
398 exhibited coherence  $>0.25$  were utilized for the ENVISAT and Sentinel-1 analyses,  
399 respectively.

400 Phase ramps attributed to orbital errors were subtracted and phase associated with  
401 topography was removed using a 90 m Shuttle Radar Topography Mission (SRTM) DEM  
402 (Farr *et al.*, 2007). Coherent and intermittently coherent pixels satisfying the criteria were  
403 unwrapped from modulo- $2\pi$  phase to relative deformation using a statistical-cost network-  
404 flow algorithm (Chen & Zebker, 2001) with respect to a reference point located in  
405 Nottingham City centre. The same reference point was used for all data sets and was  
406 assumed to be stable due to its urban characteristics and abundant cluster of coherent pixels.

407 After unwrapping, the method is in accordance with that of Berardino *et al.* (2002). The  
408 linear velocities are derived from a least squares covariance analysis of the unwrapped  
409 phase, which also determines DEM height errors. Standard errors associated with both  
410 measurements are calculated from the standard deviation of residuals after fitting. To  
411 determine the time-series for each pixel, phase associated with the linear velocities and  
412 height errors were removed from the differential interferograms, before being added back  
413 after the residual phase components had been unwrapped. The phase velocities between  
414 adjacent images were inverted through Singular Value Decomposition. Phase was  
415 subsequently integrated to derive the phase at each acquisition interval before temporal  
416 high-pass and spatial low-pass filters compute and remove atmospheric components.  
417 Modulo- $2\pi$  phase time-series are then converted to deformation time-series.



418 Sufficient ENVISAT data was only available from the descending geometry so, on the  
 419 assumption that horizontal motion is negligible, the ENVISAT deformation data were  
 420 projected into the vertical by means of dividing by the cosine of the incidence angle. For  
 421 Sentinel-1, the descending time-series were linearly interpolated to the dates of the  
 422 ascending data before the vertical ( $d_{Vert}$ ) and horizontal ( $d_{Hor}$ ) components of motion were  
 423 resolved for both the average velocities and time-series, on the assumption that there is  
 424 negligible motion in the north-south direction:

$$425 \quad d_{Vert} = \frac{(-\sin\theta_{Desc} \cdot \cos\phi_{Desc} \cdot LOS_{Asc}) + (-\sin\theta_{Asc} \cdot \cos\phi_{Asc} \cdot LOS_{Desc})}{(-\sin\theta_{Asc} \cdot \cos\phi_{Asc} \cdot \cos\theta_{Desc}) - (\cos\theta_{Asc} \cdot \cos\phi_{Desc} \cdot \sin\theta_{Desc})} \quad (\text{Eqn. 11})$$

$$426 \quad d_{Hor} = \frac{(\cos\theta_{Desc} \cdot LOS_{Asc}) + (-\cos\theta_{Asc} \cdot LOS_{Desc})}{(-\sin\theta_{Asc} \cdot \cos\phi_{Asc} \cdot \cos\theta_{Desc}) - (\cos\theta_{Asc} \cdot \cos\phi_{Desc} \cdot \sin\theta_{Desc})} \quad (\text{Eqn. 12})$$

427 where  $\theta$  is the incidence angle from surface normal,  $\phi$  is the azimuth track angle and  $LOS$   
 428 are the line-of-sight velocity measurements. The superscript indicates whether the  
 429 parameter corresponds to the ascending (*Asc*) or descending (*Desc*) geometry.

430 The DInSAR average velocities were utilized to calibrate the forward model. Where there is  
 431 a notable rise in groundwater, the rise is observed to occur linearly over time. For example,  
 432 over the ENVISAT and Sentinel-1 epochs the coefficient of determination ( $R^2$ ) is on average  
 433 0.93 and 0.94, respectively, for cases where groundwater rises more than 20 m. Similarly, the  
 434 time-series data showed that in areas of heave the DInSAR measurements were principally  
 435 linear which also provides an indication that the observed heave is attributable to  
 436 groundwater rise. Accordingly, it is appropriate to utilize the average velocities for  
 437 comparison and calibration (Table 1). In this study, an attempt was not made to utilize the  
 438 time-series to account for variable pumping rates, however, this could be implemented on a  
 439 pointwise basis at the borehole locations (e.g. Bateson *et al.*, 2009). To perform calibration, a

440 1D forward model was generated at each borehole – excluding those located in areas of  
 441 mining subsidence. The scaling factor(s) that minimised the root mean square error (RMSE)  
 442 between the modelled measurements and 95<sup>th</sup> percentile of the DInSAR average velocities  
 443 within the surrounding 1 km of the borehole were calculated. The mean scaling factor was  
 444 taken for each pond. Whilst this approach achieved realistic results in most ponds, some  
 445 scaling factors were manually adjusted to account for the fact that heave caused by a change  
 446 in groundwater does not always directly manifest at the borehole location. Instead, it can  
 447 arise elsewhere within the mine complex and failure to account for this could lead to  
 448 erroneous calibration. It was clear that when the scaling factor was too low, the average  
 449 velocities generated by the forward model were too small and, consequently, the inverse  
 450 map of groundwater rise was unrealistically high. Conversely, when the scaling factor was  
 451 too high, the velocities in the forward model were similarly too high and the inverse map  
 452 improbably small. Finally, to provide a quantitative comparison between the forward model  
 453 and DInSAR average velocities the residuals were calculated by subtracting the modelled  
 454 deformation from the DInSAR measurements.

455

### 456 3.3. Inversion of ISBAS DInSAR Observations to Estimate Groundwater Rise

457 To provide a quantitative estimate of groundwater rise ( $\Delta h$ ) across the entire coalfield, an  
 458 inversion of both the average DInSAR velocities and the displacement time-series was  
 459 implemented as:

$$460 \quad \Delta h = \frac{1}{10.s} (p - p_{so} - p_{wo} - \left(\frac{\Delta b}{m_v.b_0}\right)) \quad (\text{Eqn. 13})$$

461 where the DInSAR measurements are utilized to determine the change in bed thickness ( $\Delta b$ ).  
462 The change in bed thickness was calculated by multiplying the average velocities by the  
463 length of the DInSAR time epoch to determine the cumulative deformation, whereas for the  
464 time-series each relative height change was inverted. In addition, the cumulative standard  
465 errors of the average velocities were inverted to provide an estimate of how the  
466 measurement error of the DInSAR translates into error in the estimate of groundwater rise.

467

#### 468 **3.4. Estimate of Time Until Discharge using Inverted Map of Groundwater Rise**

469 The inverted average rate of rise of the Sentinel-1 data was utilized to calculate the time it  
470 would take for groundwater to reach the top of the CMG and either discharge at the surface,  
471 where the CMG are unconfined, or infiltrate the Permo-Triassic formation. Groundwater  
472 recovery curves in most hydrogeological environments, including studies of mine water  
473 rebound, follow a shape that conforms to an exponential function. The recovery of mine  
474 water is a non-linear process because as the head difference between the formerly dewatered  
475 and now recovering strata and surrounding aquifers reduces, the head-dependant inflow  
476 also reduces. Hence, the rate of rebound decreases exponentially with time (Younger &  
477 Adams, 1999).

478 It is possible to predict mine water rebound by fitting an exponential curve to early rebound  
479 data, however, this is particularly challenging as it assumes that the reduction in head-  
480 dependant inflow into the voids of the recovering strata occur at the same rate for entire  
481 rebound period. This assumption rarely holds – for instance, the curve might be fit to data  
482 for a period when an extensively worked lower seam of relatively high specific yield was

483 flooding. In this case, once the mine water reaches the roof and proceeds into the strata of  
484 lower permeability, rebound would accelerate and thus deviate from the assumed  
485 exponential function. Rebound might take on an exponential function again once it reaches a  
486 higher seam, however, rebound will ultimately occur sooner than predicted. Alternatively, if  
487 the curve is fit to data related to the rapid filling of a seam interval, then rebound will occur  
488 later than predicted (Younger & Adams, 1999). Accounting for such processes requires  
489 detailed geological, mining and hydrogeological data which are often unavailable.

490 In Britain, linear projections are utilized by the Coal Authority for the management of mine  
491 water, as this provides a scenario towards the worst-case; one in which rebound occurs more  
492 rapidly than in reality, which affords more time to implement pumping and treatment  
493 schemes. Such an estimate is preferable over an exponential projection that might  
494 erroneously indicate an area is not at risk. In line with this, the estimate of time until  
495 discharge was calculated by dividing the depth of the groundwater from the top of the CMG  
496 at the end of the Sentinel-1 period by the inverted yearly rate of groundwater rise. The depth  
497 of groundwater at the end of the Sentinel-1 period was interpolated from the boreholes for  
498 each pond.

499

## 500 **4. Results & Discussion**

### 501 **4.1. Generation of Forward Model using Borehole Measurements**

502 The results from the initial models showed good spatial agreement with areas of heave  
503 identified in the DInSAR data, however, large differences occurred in the magnitude of  
504 deformation. Such differences occur due to the choice of parameters; for example, whilst the  
505 value of porosity has been taken from real samples, the extraction of vast areas of strata and

506 resultant void space will act to likely increase this value. The compression index was taken  
507 from literature and even had actual measurements been utilized, the CMG are a highly  
508 variable complex multilayer aquifer so samples may not be representative of the entire  
509 formation. Additionally, the response of the strata to pressure changes is also dependant on  
510 historical elastic and inelastic deformation, the degree of which is unknown, hence, leading  
511 to uncertainties.

512

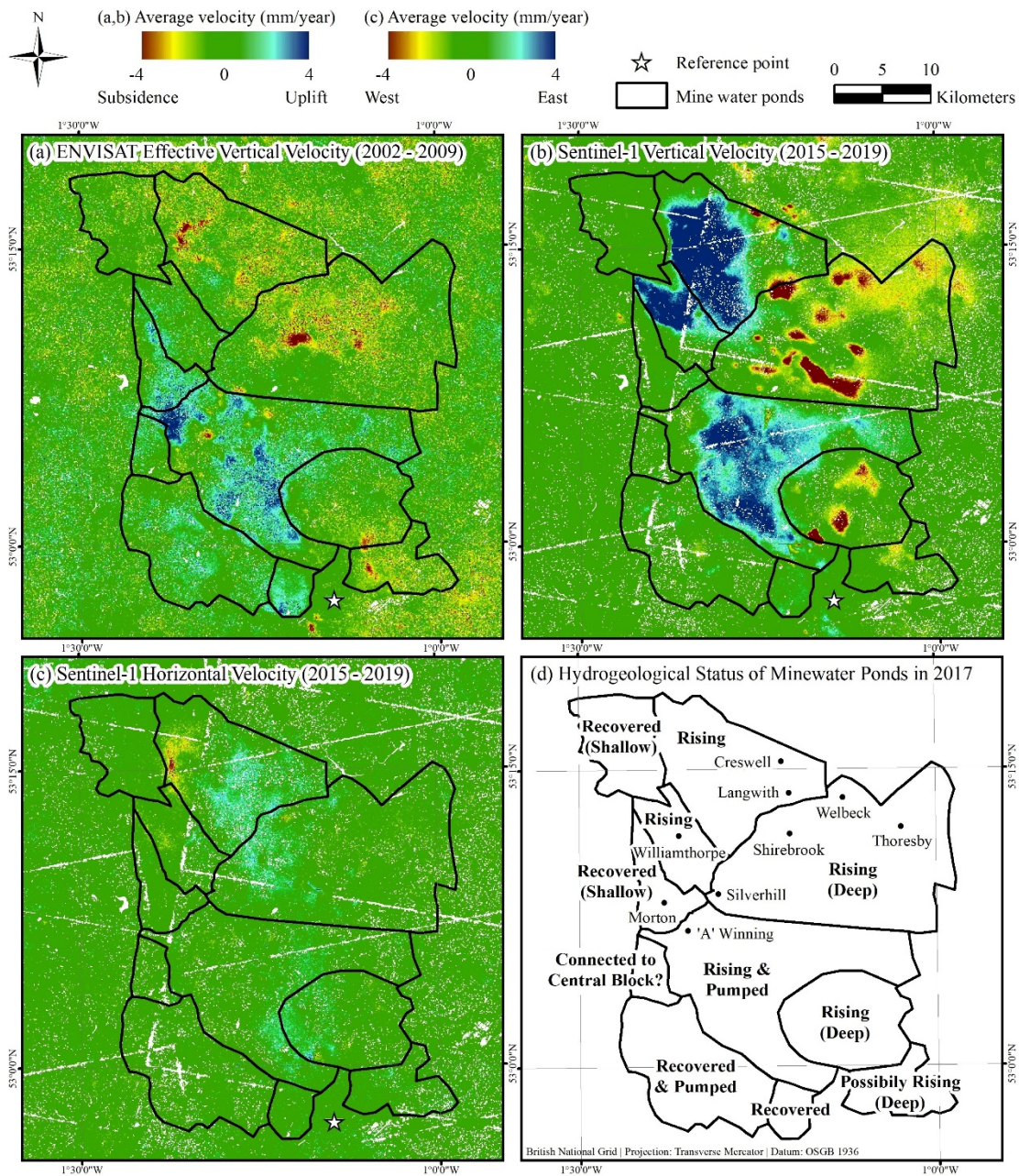
#### 513 **4.2. Calibration & Validation of Forward Model using ISBAS DInSAR Observations**

514 The average annual velocities identified over the ENVISAT and Sentinel-1 periods are  
515 shown in Figure 4. The near complete spatial coverage (~94%) of the ISBAS measurements  
516 facilitates a comprehensive characterisation of the deformation occurring over the coalfield.  
517 Areas of spatially correlated heave are measured in both data sets and occur because of  
518 rising groundwater. The temporal evolution of deformation between the ENVISAT and  
519 Sentinel-1 data concur with the timeframes of coalfield abandonment. For example, rebound  
520 in the northern coalfield has only recently commenced. Over the ENVISAT period the area is  
521 relatively stable, and pumping was performed in the west of the Hartington and  
522 Williamthorpe ponds to prevent infiltrated groundwater migrating downdip. Following the  
523 closure of Welbeck in May 2010, pumping has gradually been reduced allowing the deeper  
524 easterly mine workings to flood. At present there is very little flow across the barrier  
525 between the Langwith and Shirebrook, located in the North Deep pond. As a result, mine  
526 water is rising at Creswell and Langwith and heave is identified in the Sentinel-1 DInSAR.  
527 The barrier between Williamthorpe and Silverhill is known to leak and is allowing some  
528 flow into the North Deep pond, as evidenced from modelling by the Coal Authority in

529 2016/17. When Thoresby was closed in 2015, aside from the expected flow and flooding from  
530 the make of water from the mine workings, there was no evidence for notable large or  
531 increased flows coming from Williamthorpe, suggesting that pumping at Williamthorpe had  
532 prevented significant down-dip decant. Rebound is soon expected in the North Deep pond,  
533 however, no heave is observed over the period 2015 – 2019. This suggests that the  
534 connections with the westerly ponds are not present as anticipated, or that the groundwater  
535 levels have not yet reached the depth of the connections. Future DInSAR monitoring of this  
536 area will help to identify such a change, where groundwater has not yet reached the bottom  
537 of the monitoring borehole in the North Deep pond. In the southern coalfield, the most  
538 easterly up-dip measures around Morton and 'A' Winning recovered during the 2000s  
539 where heave is identified, and are subsequently stable once recovered and pumped in the  
540 2010s.

541

542



543

544 **Figure 4.** (a) ENVISAT effective vertical velocities (mm/year); (b) Sentinel-1 vertical velocity (mm/year); (c)

545 Sentinel-1 horizontal velocity (mm/year); and (d) Hydrogeological status of mine water ponds in 2017.

546

547 At the regional scale of the entire coalfield, the calibrated forward models of heave show

548 good agreement with the DInSAR surface deformation measurements (i.e. heave is observed

549 in ponds where groundwater is rising) (Figure 5). Over the ENVISAT epoch (2002 – 2009),

550 the modelled heave primarily occurs in the southern coalfield, within the South Central  
551 pond, driven by rebound at 'A' Winning and Langton. The DInSAR similarly identifies  
552 heave in South Central, although the rates are spatially variable in this relatively large pond  
553 suggesting South Central could contain sub-ponds. This evidences the simplification that the  
554 concept of ponds are relative to the actual geological setting. In the north the modelled  
555 heave is small. The DInSAR indicate that this pond is relatively stable, although there is a  
556 region of subsidence over the Markham borehole with an area of heave further to the south.

557 Over the Sentinel-1 epoch (2015 – 2019), heave occurs in the north coalfield within the  
558 Hartington and Williamthorpe ponds, in areas experiencing some of the largest recovery of  
559 groundwater (Figure 5). Hydraulic connectivity within the Hartington and Williamthorpe  
560 ponds are extensive. The DInSAR measurements concur with the modelled deformation,  
561 showing that heave is geologically bound by the outcropping of Pennine Middle Coal  
562 Measures. Most significantly, within the Hartington pond both the DInSAR and modelled  
563 deformation show that surface heave is of a lower velocity where the coal measures dip  
564 under the Permo-Triassic horizon. Where the CMG is confined, the DInSAR identifies lesser  
565 movement, mostly in close proximity to the boreholes at Creswell and Langwith where  
566 groundwater is rising. Crucial to this verification was the near-complete coverage (~94%) of  
567 the ISBAS DInSAR measurements in an area dominated by agricultural land where  
568 coherence is intermittent. If a conventional SBAS analysis had been utilized, the ground  
569 measurement coverage would not have been sufficient to validate the change in rate of  
570 heave of the forward model where the CMG become confined (Figure 6). Furthermore,  
571 calibration of the model would have been more challenging because 70% of the boreholes  
572 are situated within intermittently coherent pixels.

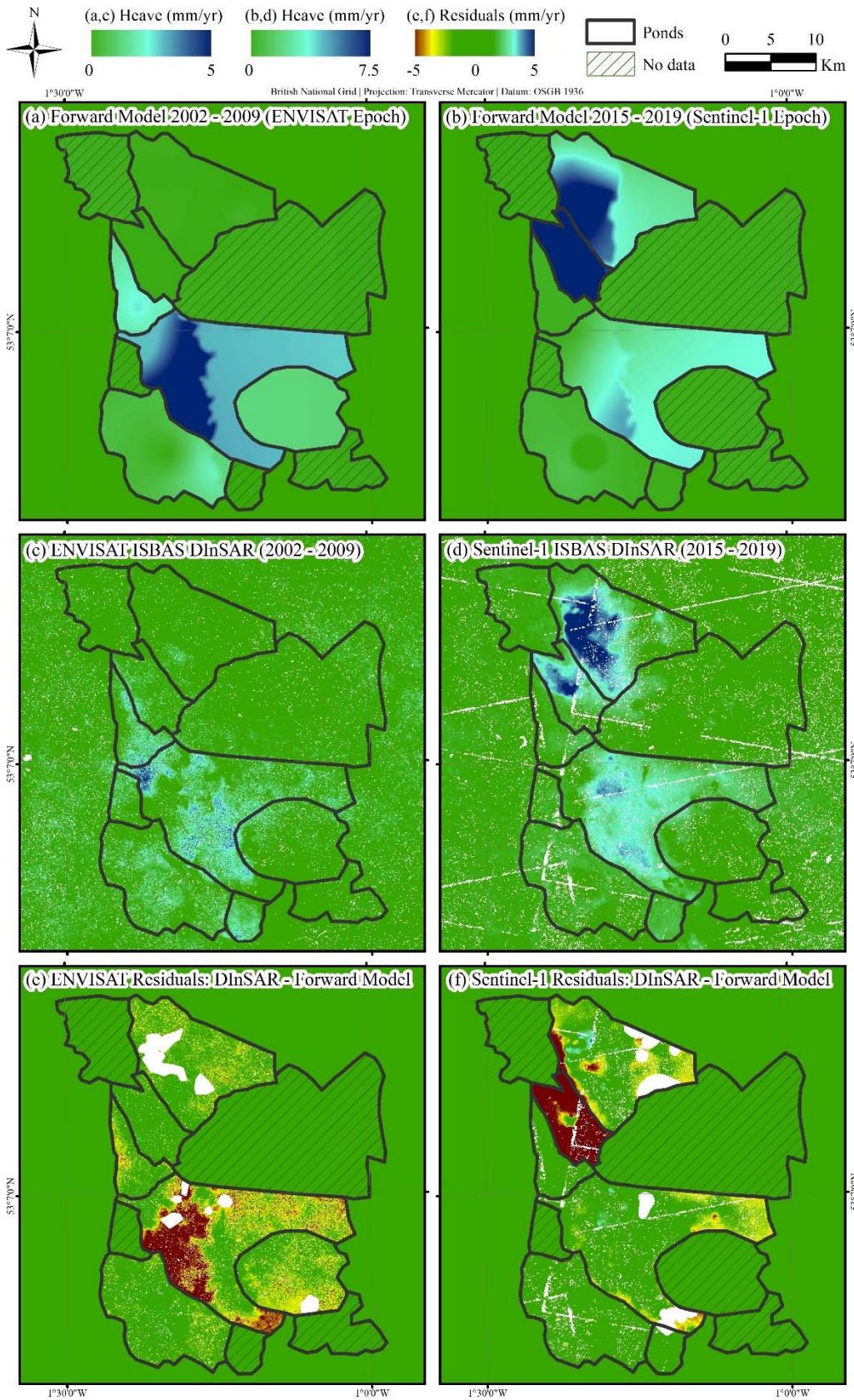


573 Heave also occurs in the DInSAR and the model in the South Central pond, in which  
574 groundwater levels are rising. To the west of the pond heave is not measured in the DInSAR  
575 in proximity of 'A' Winning where pumping occurs up-dip to control the levels at Langton  
576 and Newstead. In South Central heave is bound to the north by Hardsoft-Mansfield anticline  
577 which divides the northern and southern coalfields, to the south by a narrow but complex  
578 anticlinal zone which divides the South Central pond from the Woodside and Radford  
579 ponds (Dumpleton *et al.*, 2001), and to the east by the Calverton pond, despite significant  
580 mine water pressure on the barriers (>200 m). As expected, the recovered ponds of Clay  
581 Cross, Radford and Woodside are characterised as stable by the DInSAR and the model.  
582 Some heave is identified by the DInSAR in close proximity to the boreholes at Denby Hall,  
583 Ormonde Mill Farm and Lodge Erewash within the Woodside pond, which is indicative that  
584 whilst Woodside has recovered, it is still pumped to prevent mine water discharging at the  
585 surface and/or flowing into South Central.

586 Unfortunately, no ground truth with sufficient spatial and temporal sampling at the  
587 required accuracy and precision is available to verify the DInSAR results. However, one of  
588 the great advantages of DInSAR is the ability to provide historical and current ground  
589 deformation measurements in areas where there is limited or no ground truth data. Prior  
590 ISBAS measurements have been validated in urban and rural environments, providing  
591 confidence in the method (e.g. Gee *et al.*, 2016; Alshammari *et al.*, 2019; Gee *et al.*, 2019;  
592 Grebby *et al.*, 2019). Over the Nottinghamshire Coalfields, the temporal evolution of  
593 deformation between the ENVISAT and Sentinel-1 data and the quantitative comparison  
594 between the deformation measured by the forward models and DInSAR confirm that the  
595 heave is caused by the recovery of mine water. No groundwater data is available to generate

596 the forward model in several ponds, however, given the agreement between the DInSAR  
597 and model, the DInSAR can be utilized to infer the status of rebound. For instance, during  
598 the ENVISAT epoch heave is measured in Radford although no borehole measurements are  
599 available for this period (Figure 5). The pond is known to have rebounded by the Sentinel-1  
600 epoch (Table 1; Figure 4d) so the DInSAR indicate that recovery occurred partly or wholly  
601 between 2002–2009.

602



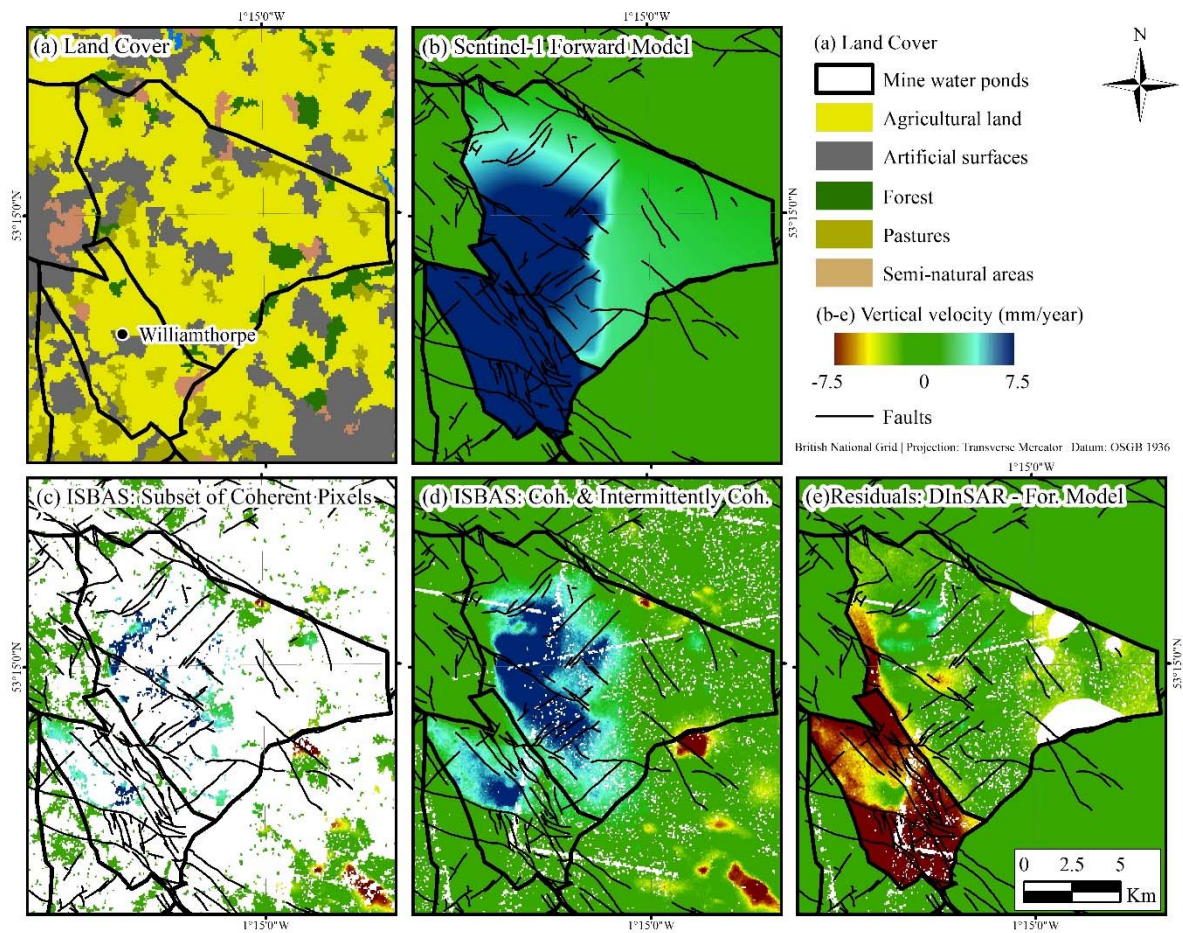
603

604 **Figure 5.** Calibrated forward models for the: (a) ENVISAT and (b) Sentinel-1 time epochs. Intermittent SBAS

605 DInSAR analysis for the: (c) ENVISAT and (d) Sentinel-1 time epochs. Residuals between ISBAS DInSAR and

606 forward models for the: (e) ENVISAT and (f) Sentinel-1 time epochs. Red and blue indicate respective  
 607 underestimation and overestimation of heave by the DInSAR relative to the model. Areas of notable subsidence  
 608 identified by the DInSAR have been masked from the residuals.

609



610

611 **Figure 6.** Fault-bound motion and variable deformation within the Hartington and Williamthorpe ponds,  
 612 measured over the Sentinel-1 period: (a) CORINE land cover inventory (European Environment Agency, 2012);  
 613 (b) Forward model (mm/year); (c) Sentinel-1 average vertical velocities (mm/year) for the subset of coherent  
 614 pixels only; (d) Sentinel-1 average velocities (mm/year) for all pixels (coherent and intermittently coherent).  
 615 Residuals between ISBAS DInSAR and forward models. Red indicates underestimation of heave by the DInSAR  
 616 relative to the model and blue indicates and overestimation by the DInSAR relative to the model. Areas of  
 617 notable subsidence identified by the DInSAR have been masked from the residuals. European Environment  
 618 Agency © 2012.

619

620

621 **4.3. Inversion of ISBAS DInSAR Observations to Estimate Groundwater Rise**

622 The adoption of ponds to characterise rebound in the coalfield is a practical, albeit necessary,  
623 simplification of the actual hydrogeological setting for the forward modelling. Whilst at the  
624 regional coalfield scale the forward models show a good level of agreement with the  
625 DInSAR measurements, the variability in groundwater rise has to be interpolated, or a single  
626 value input for an entire pond where only a single borehole exists. The complex and  
627 extensive network of mine workings and poor spatial sampling of the boreholes means that  
628 groundwater rise between boreholes is not likely to be realistically characterised. For  
629 example, over the Sentinel-1 period within the Williamthorpe pond there is variability in the  
630 groundwater levels where heave is confined due to the presence of impermeable faults  
631 (Figure 6d,e). The residuals between the forward model and DInSAR observations tend  
632 towards zero in the area surrounding the borehole, however, because of the variability in  
633 groundwater levels large residuals, up to -10 mm/year, occur in the south of the pond.  
634 Similarly, due to the variability of rebound within South Central over the ENVISAT epoch  
635 residuals of up to -8mm/year are measured (Figure 5). The near-complete spatial sampling  
636 of the ISBAS DInSAR surface measurements can capture the groundwater variability, hence,  
637 once inverted the change in groundwater levels are characterised in greater detail than the  
638 boreholes can solely achieve across the coalfield.

639 The inverted DInSAR measurements assume that heave occurs solely as a result of rising  
640 groundwater, which may not always be the only cause of the observed ground motions.

641 However, it is a reasonable assumption given that the spatially correlated heave is  
642 delimited by the structural geology and has been validated by the forward models (Section  
643 4.2). The model can be theoretically utilized to predict a drop in groundwater levels.  
644 Notably, there are areas where subsidence occurs due to groundwater depletion at boreholes  
645 where groundwater has been pumped, for example, in close proximity to Woodside and 'A'  
646 Winning over the Sentinel-1 period (~ -2 mm/year). However, groundwater depletion does  
647 not cause many of the detrimental problems associated with rebound (e.g. surface water  
648 pollution, localised flooding, pollution of overlying aquifers), hence, the purpose of the  
649 study is to identify where the problems associated with rebound may manifest. There are  
650 also localised hotspots of subsidence that are either known to be or likely to be related to  
651 mining induced collapses (e.g. the North Deep and Calverton ponds). Where such  
652 subsidence processes are transpiring an estimate of groundwater rise cannot be generated  
653 and so these areas are not considered and were masked from the inverse model.

654 The average inverted rates are shown in Figure 7 and the average of all the inverted time-  
655 series in the surrounding 1 km of the boreholes of Hartington, Williamthorpe and Morton  
656 are shown in Figure 8. The inverted measurements provide a more realistic estimate of the  
657 rise in groundwater than previous simple correlations that universally determine a value of  
658 groundwater rise per unit of surface deformation (e.g. Banton *et al.*, 2013). A change in bed  
659 thickness within the CMG would plausibly have less of an effect at the surface when the  
660 CMG is overlain by successive strata and, similarly, a rise in groundwater closer to the  
661 surface would be expected to have a greater control on surface deformation than a rise that  
662 occurs at depth. The value of groundwater rise per unit of surface deformation is different  
663 for every pixel since the inverted estimate considers both the geology and the depth at

664 which mine water is rising. For example, over the Sentinel-1 period heave measures up to  
665 ~10 mm/year around the Williamthorpe borehole which is in a shallow unconfined area,  
666 whereas at Creswell, in a deeper confined area, heave is ~3 mm/year (Figure 9a). If only the  
667 surface deformation data was considered, it might be erroneously interpreted that  
668 groundwater levels are rising ~3-times as fast at Williamthorpe than Creswell. The inverted  
669 measurements consider the geology and depth of groundwater and, hence, correctly  
670 determines that over the course of the Sentinel-1 epoch groundwater levels rose faster at  
671 Creswell (77.6 m) than Williamthorpe (69.3 m) (Figure 9b).

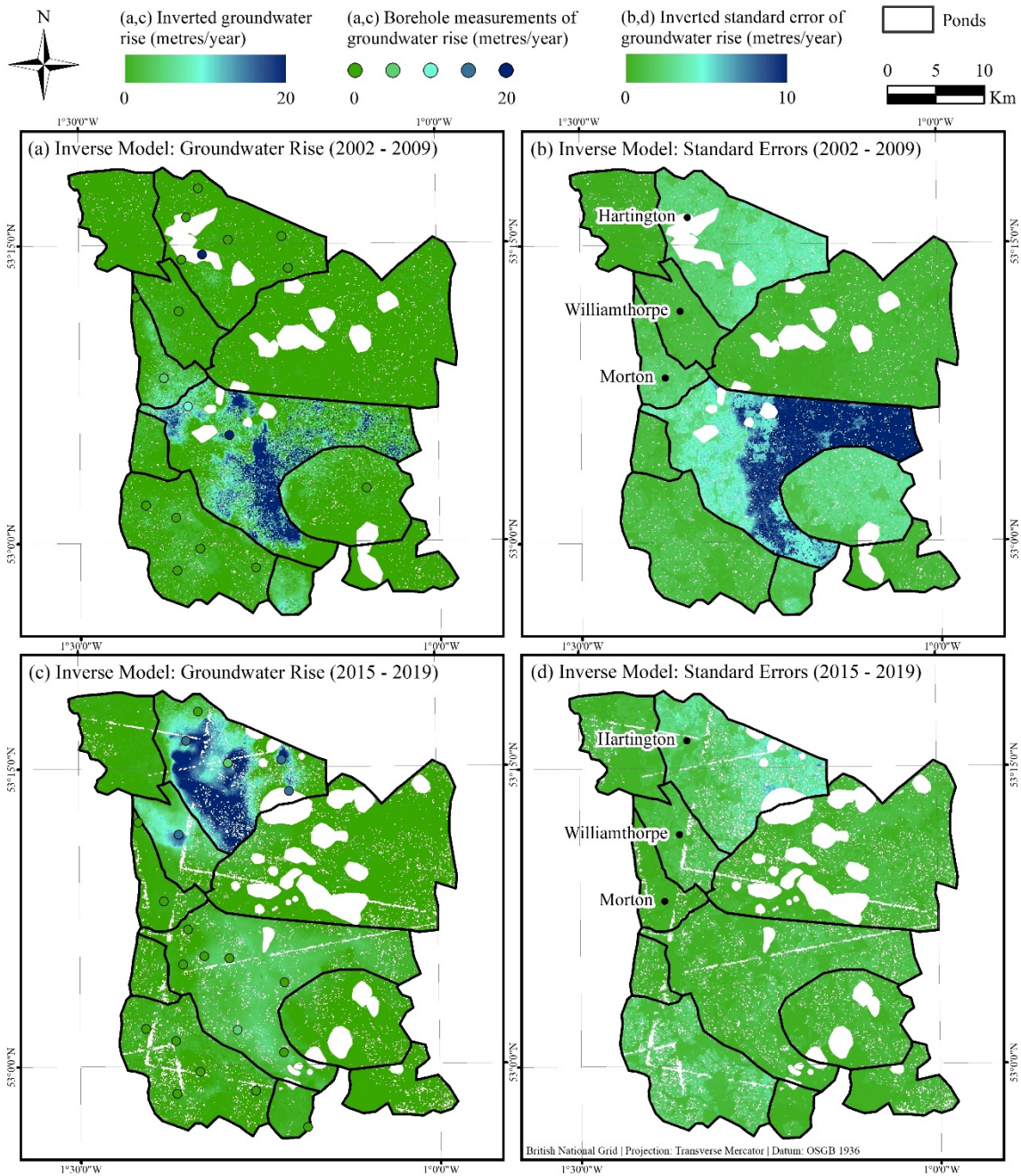
672 The standard errors associated with the inverted rates are shown in Figure 7b,d. The  
673 inverted standard errors are largest when a relatively small rate of heave translates into a  
674 large rise in groundwater (i.e. in deep confined areas such as the east of South Central pond  
675 over the ENVISAT period and east of the Hartington pond over the Sentinel-1 period). The  
676 standard error of the ISBAS velocities are controlled by the number of coherent  
677 interferograms, which is in turn predominantly controlled by the land cover. Coherence is  
678 less likely to be maintained outside of urban areas, hence, there are less coherent  
679 interferograms and the standard errors of velocities are higher. The standard errors of the  
680 Sentinel-1 velocities are on average less than the ENVISAT velocities, as a result of the  
681 greater number of images available and smaller perpendicular orbital baselines between  
682 images. Hence, the inverted standard errors of groundwater rise are controlled by the  
683 quantity and quality of SAR data, land cover, the confinement of the CMG, and depth and  
684 rate of change of the groundwater.

685 Again, the near complete ground coverage of the DInSAR measurements afforded by the  
686 ISBAS method has been crucial to the inverse mapping. In previous studies where DInSAR

687 has been utilized to map groundwater level variations, decorrelation over non-urban land  
688 cover meant that measurements had to be interpolated to achieve complete coverage before  
689 inversion (e.g. Bejar-Pizarro *et al.*, 2017). Whilst the density of measurements is important for  
690 modelling, for spatially correlated deformation, achieving a balanced spatial sampling can  
691 be more important to successfully characterise the signal of interest (Hanssen *et al.*, 2008).  
692 Spatial interpolation is not required to achieve both a high density and regular spatial  
693 sampling with the ISBAS method, ultimately leading to a more accurate characterisation of  
694 the changes in groundwater levels.

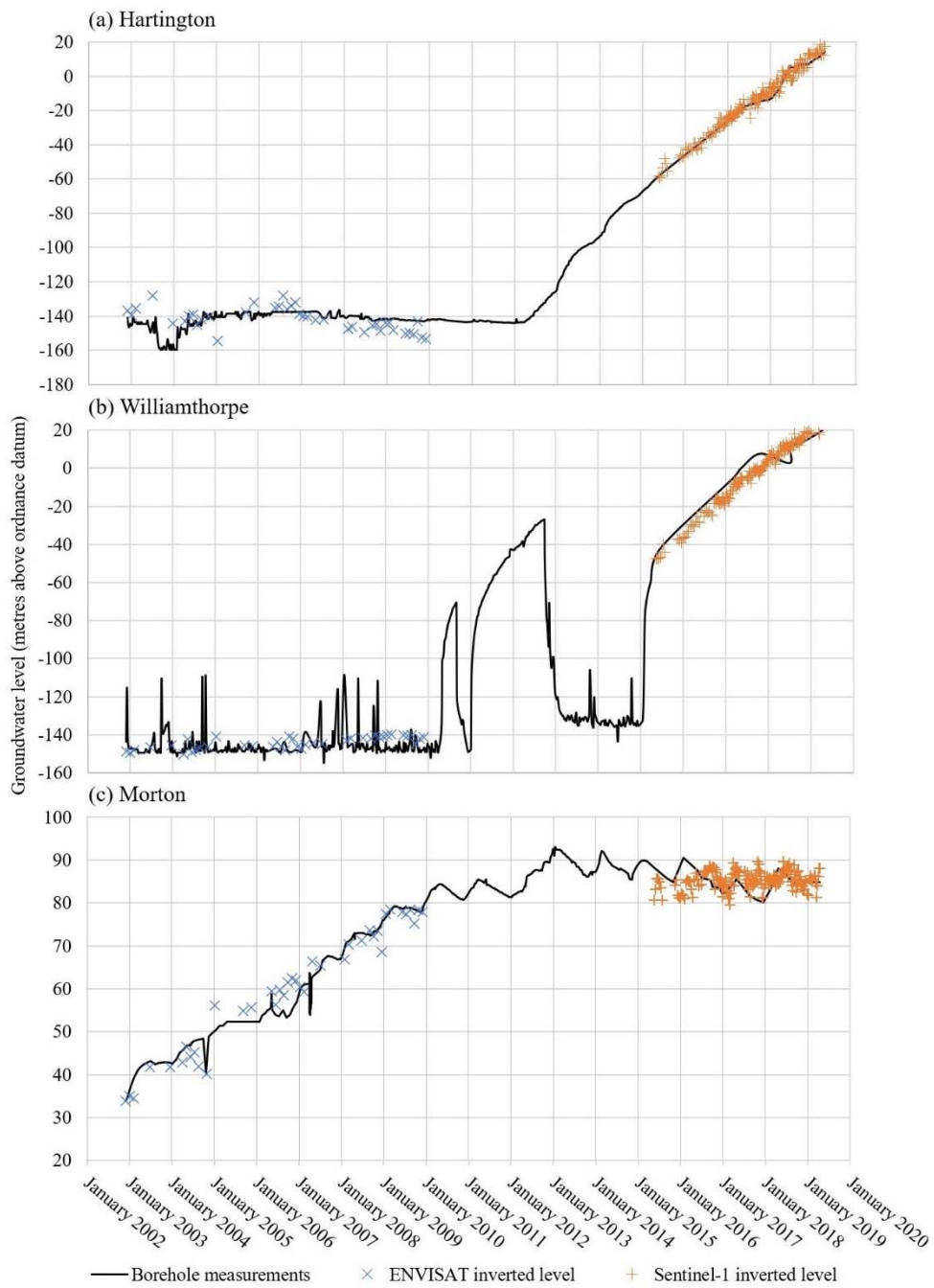
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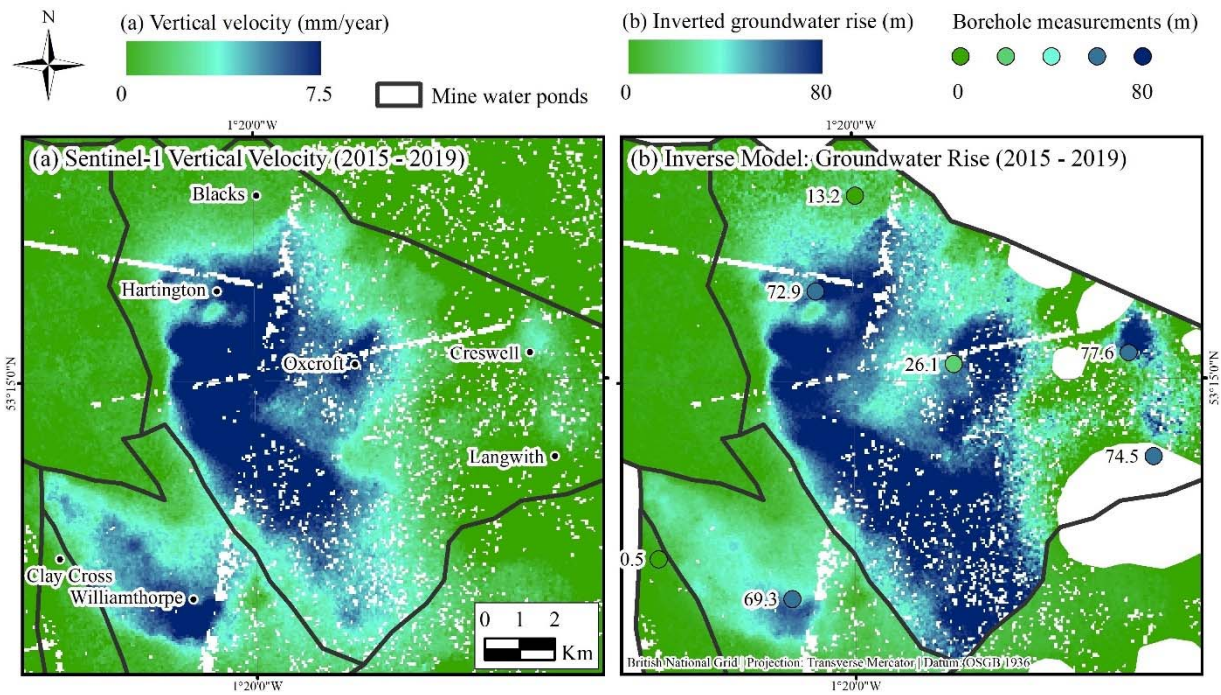
697 **Figure 7.** Average change in groundwater levels per year as derived via the inverse model from: (a) ENVISAT  
 698 ISBAS DInSAR; (b) ENVISAT standard errors; (c) Sentinel-1 ISBAS DInSAR and (d) Sentinel-1 standard errors.



699

700 **Figure 8.** Measured groundwater rise from monitoring boreholes and calibrated inverted groundwater rise from  
 701 ENVISAT and Sentinel-1 data: (a) Hartington; (b) Williamthorpe; and (c) Morton. The locations of the boreholes  
 702 are shown in Figure 7.

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710 **4.4. Estimate of Time Until Discharge using Inverted Map of Groundwater Rise**

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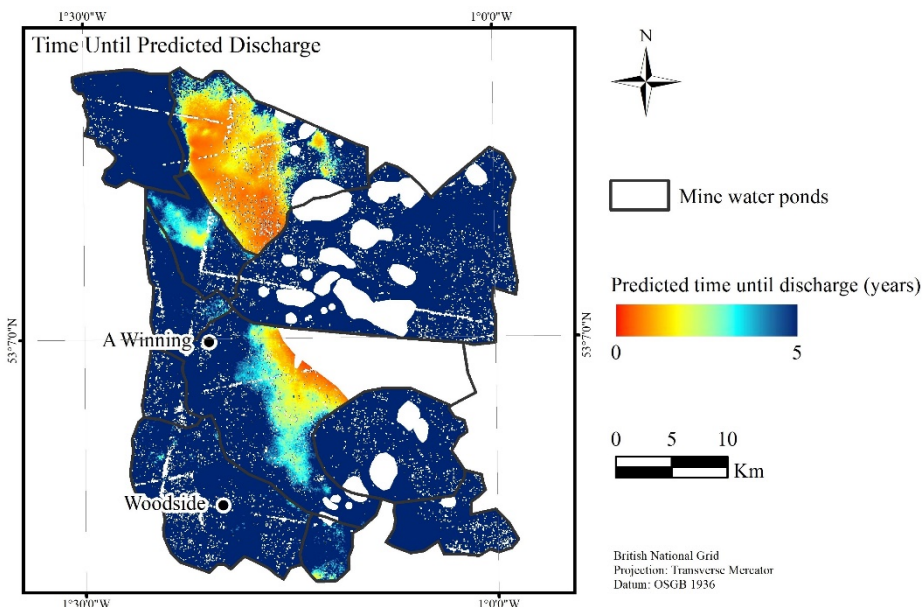
717

**Figure 9.** (a) Sentinel-1 vertical velocity (mm/year); and (b) cumulative inverted groundwater rise (m) and borehole measurements over the Sentinel-1 period. Areas of notable subsidence identified by the DInSAR have been masked from the inverse map.

Figure 10 displays the estimated time until discharge out of the CMG for the worst-case scenario. This assumes that the groundwater regime of the previous four years continues (i.e. the Sentinel-1 period), which might not be the case if a connection to a neighbouring mine is reached or if the Coal Authority decide to pump or stop pumping at selected boreholes. Nevertheless, it will be indicative of where mine water discharges will occur under the current management and can be used to help prevent pollution at the surface or of the overlying Triassic Sherwood Sandstone aquifer, which is a major source of public water

718 supply for the region. Such an approach could only be implemented effectively for areas  
 719 where the depth of the groundwater is known at the end of the Sentinel-1 period (April  
 720 2019) and requires a regular sampling of borehole measurements, particularly in areas  
 721 where the CMG dip relatively steeply. This is not the case in the eastern portion of the South  
 722 Central pond which was, therefore, excluded. The results show that the west of the  
 723 Hartington and centre of the South Central ponds are predicted to discharge first, in as little  
 724 six months under the current groundwater regime. In the Williamthorpe pond, discharge is  
 725 predicted in approximately 18 months at the centre of the pond. Conversely, the effects of  
 726 pumping are clear in the areas surrounding the 'A' Winning borehole and the entire  
 727 Woodside pond which are not at risk of discharge.

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 732 **Figure 10.** Predicted time until discharge out of the Coal Measures Group, either at the surface or into the Permo-  
 733 Triassic strata.

734

735 **5. Conclusion**

736 Characterisation and prediction of hydraulic flow in recently abandoned coalfields is  
737 particularly challenging due to inaccurate or incomplete mine plans and sparse monitoring  
738 boreholes. In this study, a method based upon the principle of effective stress and concept of  
739 mine water ponds was implemented to map variations in groundwater levels using  
740 DInSAR. The near complete coverage of the ISBAS deformation measurements was critical  
741 to calibrate the model, verify the outputs and provide a spatial quantitative estimate of the  
742 rate of groundwater rebound and an estimate of the time it will take for groundwater to  
743 discharge out of the Coal Measures rock. Given that many shafts are backfilled following  
744 coalfield closure and drilling new deep monitoring boreholes is expensive, the method  
745 proposed can provide a valuable cost-effective input into the abandonment strategy to help  
746 identify where pumping might be required. It is relatively straightforward to apply and  
747 requires few parameters to be estimated, and so, with the aid of the Sentinel-1 archive, has  
748 the potential to be readily applied over other recently abandoned coalfields.

749

750

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760 Director of BGS.

761

## 762 **Conflicts of Interest**

763 The authors declare no conflicts of interest.

764

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962

### 963 **Figure Captions**

964 Figure 1. (a) Topography – NEXTMap® DTM at 10 m resolution; (b) CORINE land cover  
965 inventory (European Environment Agency, 2012); (c) Bedrock geology at 1:625,000 scale,  
966 from BGS Geology 625k (DiGMapGB-625) data; and (d) Superficial deposits at 1:625,000  
967 scale, from BGS Geology 625k (DiGMapGB-625) data. The dashed black line in (c) shows the  
968 location of the cross-section in Figure 2. NEXTMap® Britain © 2003, Intermap Technologies  
969 Inc. European Environment Agency © 2012. Reproduced with the permission of the British  
970 Geological Survey © NERC. All rights Reserved.

971

972 Figure 2. Cross-section of the Nottinghamshire Coalfield. The location is marked on Figure  
973 1c. The thick vertical black lines represent shaft locations and thinner off vertical black lines  
974 represent faults. The high main seam represents the upper limit of mining. The base of the  
975 Top Hard and Blackshale seams for the areas east and west of Annesley-Bentinck,  
976 respectively, indicates the base of the zone of enhanced permeability due to mining induced  
977 fractures (adapted from Dumbleton et al., 2001).

978

979 Figure 3. Mine water ponds and expected schematic of flow of groundwater for the  
980 Nottinghamshire Coalfield. The red arrow between Langwith and Shirebrook indicates that  
981 at present there is very little or no flow between this barrier.

982

983 Figure 4. (a) ENVISAT effective vertical velocities (mm/year); (b) Sentinel-1 vertical velocity  
984 (mm/year); (c) Sentinel-1 horizontal velocity (mm/year); and (d) Hydrogeological status of  
985 mine water ponds in 2017.

986

987 Figure 5. Calibrated forward models for the: (a) ENVISAT and (b) Sentinel-1 time epochs.  
988 Intermittent SBAS DInSAR analysis for the: (c) ENVISAT and (d) Sentinel-1 time epochs.  
989 Residuals between ISBAS DInSAR and forward models for the: (e) ENVISAT and (f)  
990 Sentinel-1 time epochs. Red and blue indicate respective underestimation and  
991 overestimation of heave by the DInSAR relative to the model. Areas of notable subsidence  
992 identified by the DInSAR have been masked from the residuals.



993

994 Figure 6. Fault-bound motion and variable deformation within the Hartington and  
995 Williamthorpe ponds, measured over the Sentinel-1 period: (a) CORINE land cover  
996 inventory (European Environment Agency, 2012); (b) Forward model (mm/year); (c)  
997 Sentinel-1 average vertical velocities (mm/year) for the subset of coherent pixels only; (d)  
998 Sentinel-1 average velocities (mm/year) for all pixels (coherent and intermittently coherent).  
999 Residuals between ISBAS DInSAR and forward models. Red indicates underestimation of  
1000 heave by the DInSAR relative to the model and blue indicates overestimation by the  
1001 DInSAR relative to the model. Areas of notable subsidence identified by the DInSAR have  
1002 been masked from the residuals. European Environment Agency © 2012.

1003

1004 Figure 7. Average change in groundwater levels per year as derived via the inverse model  
1005 from: (a) ENVISAT ISBAS DInSAR; (b) ENVISAT standard errors; (c) Sentinel-1 ISBAS  
1006 DInSAR and (d) Sentinel-1 standard errors.

1007

1008 Figure 8. Measured groundwater rise from monitoring boreholes and calibrated inverted  
1009 groundwater rise from ENVISAT and Sentinel-1 data: (a) Hartington; (b) Williamthorpe; and  
1010 (c) Morton. The locations of the boreholes are shown in Figure 7.

1011

1012 Figure 9. (a) Sentinel-1 vertical velocity (mm/year) and (b) cumulative inverted groundwater  
1013 rise (m) and borehole measurements over the Sentinel-1 period. Areas of notable subsidence  
1014 identified by the DInSAR have been masked from the inverse map.

1015

1016 Figure 10. Predicted time until discharge out of the Coal Measures Group, either at the  
1017 surface or into the Permo-Triassic strata.