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1 Overstated carbon emission reductions from voluntary REDD+ projects in the

2 Brazilian Amazon

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12 **ABSTRACT**

Reduced Emissions from Deforestation and forest Degradation (REDD+) has gained 13 international attention over the past decade, as manifested in both United Nations policy 14 discussions and hundreds of voluntary projects launched to earn carbon-offset credits. There 15 are on-going discussions about whether and how projects should be integrated into national 16 17 efforts under the Paris Agreement. One consideration is whether these projects have generated additional impacts over and above national policies and other conservation 18 19 measures. To help inform these discussions, we compare the crediting baselines established ex-ante by voluntary REDD+ projects in the Brazilian Amazon to counterfactuals constructed 20 ex-post based on the quasi-experimental synthetic control method. We find that the crediting 21 baselines assume consistently higher deforestation than counterfactual forest loss in synthetic 22 control sites. This gap is partially due to decreased deforestation in the Brazilian Amazon 23 during the early implementation phase of the REDD+ projects considered here. This suggests 24 that forest carbon finance must strike a balance between controlling conservation investment 25 risk and ensuring the environmental integrity of carbon emission offsets. Relatedly, our 26 results point to the need to better align project- and national-level carbon accounting. 27

Keywords: Impact evaluation; synthetic control; payment for environmental services; carboncredit; deforestation.

30 Significance

There are efforts to integrate the reduced carbon emissions from avoided deforestation claimed by voluntary REDD+ projects into national greenhouse gas (GHG) emission inventories. This requires careful consideration of whether and how much of the reduced 34 carbon emissions can be attributed to projects. However, credible evidence on the 35 effectiveness of such voluntary activities is limited. We adopted a quasi-experimental, 36 synthetic control method to examine the causal effects of 12 voluntary REDD+ projects in 37 the Brazilian Amazon. We compared these ex-post estimates of impacts with the reductions 38 in forest loss claimed by those projects based on ex-ante baselines. Results suggest that the 39 accepted methodologies for quantifying carbon credits overstate impacts on avoided 40 deforestation and climate change mitigation.

41 Introduction

42 Concerns over global warming have led both the public and private sectors to promote climate change mitigation through the reduction of carbon (CO₂) emissions from 43 deforestation and forest degradation in tropical countries—a concept known as REDD+ (1). 44 This strategy gained international attention after 2005 as a voluntary, performance-based 45 46 payment mechanism for reduced carbon emissions (2). While the regulations and capacity for national REDD+ programs are still under development in many countries, hundreds of 47 48 voluntary, subnational REDD+ projects are operational worldwide (3). These projects intend to preserve forests through a variety of activities, e.g., improved monitoring and control, 49 promotion of sustainable land uses, and engagement of local communities (4), either as 50 proof-of-concept or to profit from the commercialization of "carbon-offset credits" (i.e., Mg 51 CO₂ removed from or not emitted to the atmosphere) in a variety of markets. While these 52 markets do not provide the level of funding originally envisioned for national REDD+ 53 programs, they are substantial: in 2018 alone, the volume of carbon offsets traded totaled 54 98.4 million Mg CO₂, with a market value of USD 295.7 million; a third of those credits 55 (30.5 million Mg CO₂) were generated by REDD+ projects (5). The Paris Agreement has 56 57 raised thorny questions about how the carbon emission reductions claimed by these projects relate to Nationally Determined Contributions (NDCs) and national greenhouse gas (GHG) 58 59 emission inventories reported to the United Nations Framework Convention on Climate Change (6-8). 60

61 Carbon credits from REDD+ (at both the project and national levels [1]) are issued 62 based on performance, as defined by the comparison of realized forest cover to a baseline 63 scenario constructed by projecting the forest cover expected in the absence of REDD+ (9). 64 These baseline scenarios typically assume a continuation of historical deforestation trends 65 (10), and thus eventually become unrealistic counterfactuals as the regional economic and 66 political context change. Notably, these types of changes were observed in the Brazilian

Amazon during 2004–2012, a period of sharply declining rates of forest loss (11), and also 67 during 2019, when deforestation soared again (12) (Fig. 1). Consequently, credits for reduced 68 deforestation (or lack thereof) claimed by voluntary REDD+ projects in the Brazilian 69 Amazon may have been artifacts of external factors rather than REDD+ activities. Further, 70 71 critics of voluntary REDD+ projects have raised concerns that deforestation baselines might 72 be intentionally inflated by profiteers seeking to financially benefit from the 73 commercialization of superfluous credits, or "hot air" (13–15). In addition to the direct cost of not effectively off-setting GHG emissions, the excess credits generated by these projects 74 75 impose an indirect cost on legitimate climate change mitigation efforts by undercutting the 76 price of their credits.

Early efforts to address these concerns included the establishment of standards and registries for voluntary carbon-offset projects. These standards were designed to ensure the environmental integrity of carbon offsets by requiring projects to use approved carbonaccounting methodologies for establishing deforestation baselines, monitoring, and reporting, all subject to third-party audits. Among those, the *Verified Carbon Standard* (VCS) (16) has certified the greatest number of voluntary REDD+ projects worldwide (5).

Despite the growing literature on local REDD+ interventions, there have only been a 83 84 few evaluations of their impacts on carbon emissions using rigorous, counterfactual-based methods (17, 18). To our knowledge, this is the first study that systematically compares 85 deforestation baselines established ex-ante with counterfactual estimates of deforestation 86 constructed ex-post. We employ the synthetic control method to construct deforestation 87 88 counterfactuals and assess the reductions in forest loss that can be attributed to voluntary REDD+ projects (19-21). We apply this method to all VCS-certified REDD+ projects for 89 unplanned deforestation implemented in the Brazilian Amazon in the last decade (2008-90 2017; Fig. 2 & SI Appendix, Table S1). We focus on this region for several reasons: its 91 global relevance for conservation and REDD+; the on-going discussions in Brazil about 92 "nesting" voluntary projects into a national REDD+ program (6-8); and the recent 93 availability of a cadastral database (22) that allows us to define a pool of rural properties 94 similar to the REDD+ project areas. We construct synthetic controls from donor pools of 95 properties based on weighted combinations of accessibility and biophysical characteristics 96 that result in the best matches of historical deforestation trends. Unlike the typical approach 97 to crediting baselines, we then construct counterfactual deforestation scenarios based on the 98 actual deforestation observed in those synthetic controls during the period when the REDD+ 99 projects were operational. We evaluate whether the REDD+ projects caused additional 100

reductions in deforestation compared to the counterfactual deforestation as represented by the synthetic controls (i.e., REDD+ *additionality*) and assess the robustness of our results with placebo tests (21). We also examine trends in forest loss in buffer zones around the REDD+ project areas after project implementation to assess the plausibility that any apparent reductions in deforestation may have been displaced instead (23). Finally, we contrast our counterfactuals to the crediting baselines adopted by the voluntary projects.

107 Results

Before assessing the impacts of the REDD+ projects, we explored whether the synthetic controls can accurately replicate deforestation trends in the project areas without REDD+. This "proof of concept" was implemented by dividing the pretreatment period (i.e., before project implementation) into "training" and "testing" periods. We found that the synthetic control method was able to replicate pretreatment deforestation trends reasonably well in 10 of the 12 synthetic controls (SI Appendix, Fig. S2). Our findings for the other two projects (i.e., Jari/Amapá and Suruí) must be interpreted with particular caution.

115 **Deforestation in the REDD+ areas.** Overall, we find no significant evidence that voluntary REDD+ projects in the Brazilian Amazon have mitigated forest loss. Deforestation is 116 consistently lower in the REDD+ project site than in the synthetic control in only four of the 117 projects (Fig. 3 & SI Appendix, Fig. S3), and this difference is only outside the confidence 118 interval around zero established by the placebo tests in one project (Maísa; Fig. 4 & SI 119 Appendix, Fig. S4). The only two REDD+ projects from our sample that were implemented 120 in protected areas, i.e., Suruí and Rio Preto-Jacundá, experienced among the largest 121 cumulative losses of forest cover after REDD+ implementation, along with Jari/Amapá (Fig. 122 3). This is partly a function of their large project areas and the widespread forest fires that 123 124 occurred in those protected areas in 2010–2011 and 2015, respectively (see SI for details). For Rio Preto-Jacundá we find much higher deforestation than in its synthetic control (which 125 126 is the same order of magnitude in size); specifically, the differences between deforestation (both cumulative and annual) in the Rio Preto-Jacundá area and its synthetic controls were 127 substantially greater than the differences between deforestation in the placebos and their 128 synthetic controls (Fig. 4 & SI Appendix, Fig. S4). 129

Across all projects, we find substantial differences between the deforestation baseline scenarios adopted ex-ante by the REDD+ projects and the observed forest loss (ex-post) in the synthetic controls (Fig. 5 & SI Appendix, Fig. S5). The Suruí project, implemented in an indigenous territory, is the only case where the synthetic control deforestation exceeded the baseline deforestation adopted by the project proponents. This may reflect the fact that the
baseline for Suruí was developed based on a participatory, system dynamics model (24), as
opposed to the assumptions based on historical deforestation trends adopted by all other
projects (see SI for details).

138 Carbon offset implications. Credits from the voluntary REDD+ projects are generally 139 issued after a third-party audit (i.e., *verification*) every 1–5 years. These credits are based on 140 the estimated carbon-emission reductions from the avoided deforestation brought about by 141 the projects, calculated as the difference between the carbon emissions under the baseline 142 scenario minus the observed emissions from the project area and leakage.

According to the projects' ex-ante estimates, up to 24.8 million carbon offsets could 143 potentially have been generated by the REDD+ interventions by 2017 (Fig. 5 & SI Appendix, 144 Table S1). According to the VCS database, only 5.4 million tradable credits from these 145 146 projects have been certified and made available to offset GHG emissions from private and 147 public sources by that year (SI Appendix, Table S1) (25). Using the synthetic control method to estimate REDD+ counterfactuals, we find no systematic evidence that the certified carbon 148 offsets claimed by the voluntary projects in our sample (with the exception of Maísa) are 149 associated with additional reductions in deforestation in the REDD+ areas above and beyond 150 the background reduction in deforestation achieved in the Brazilian Amazon over the same 151 152 period (11). Even for the Maísa case, our results suggest that nearly 40% of the 50 thousand tradable carbon offsets issued by the project by 2017 (SI Appendix, Table S1) may not be 153 genuinely additional (Fig. 5). 154

Leakage. If REDD+ implementation mitigates forest loss in project areas by effectively 155 excluding deforestation agents, it could displace, and hence increase, deforestation next to the 156 project areas. Shifts in deforestation after project start in 10-km buffer zones surrounding the 157 REDD+ projects suggest that such leakage effects could have occurred in three cases (i.e., 158 Maísa, Florestal Santa Maria, and Manoa; SI Appendix, Fig. S6). Further, leakage 159 presupposes a direct conservation impact, and all three of the projects exhibited lower 160 deforestation than their synthetic controls, although this estimated effect of REDD+ is only 161 162 larger than the placebo tests in the Maísa project (Fig. 4 & SI Appendix, Fig. S4). It is also worth noting that while deforestation in the buffer zones of these three projects rose between 163 the project start dates and 2017, post-intervention rates were still lower on average than in the 164 pre-REDD+ period. 165

166 **Discussion**

Our findings partially support early skepticism about the contribution of voluntary 167 REDD+ projects to climate change mitigation (15, 26). In particular, they raise questions 168 about the environmental integrity of offsets calculated using deforestation counterfactuals 169 based on the continuation of historical trends (e.g., Fig. 1). In all projects that established 170 crediting baselines using historical trends, we find that the crediting baselines significantly 171 overstate deforestation in comparison to the counterfactual estimates based on synthetic 172 controls. This pattern reflects the confounding effect created by Brazil's post-2004 efforts to 173 control Amazonian deforestation that were uniquely successful (11, 27, 28). If carbon credits 174 175 are expected to reflect changes in emissions caused by REDD+, then using historical baselines leads to excess carbon credits for projects when deforestation at the regional level 176 drops below the historical baseline. The opposite happens when unanticipated forest threats, 177 such as fires, emerge at the regional scale. 178

In contrast, the synthetic control methodology uses historical trends to identify appropriate weighted combinations of comparison areas but then constructs the counterfactual based on the observed deforestation in those areas. These counterfactuals thus incorporate the effects of contemporaneous drivers of deforestation, including agricultural commodity prices, currency exchange rates, and environmental regulations (27–29). As such, the synthetic control method is less prone to incorrectly attribute changes in deforestation to REDD+.

We note some caveats on our analysis. First, we base our evaluation on the project 186 boundaries defined by the polygons available from the VCS project database, which are 187 somewhat larger than the areas officially reported by project proponents (SI Appendix, Table 188 S2). Most of those polygons correspond to Amazonian rural properties registered in the 189 Brazilian Rural Environmental Registry (CAR), whose owners are legally entitled to clear up 190 to 20% of their forest area. Second, our synthetic controls do not perfectly match the REDD+ 191 project areas in terms of size, accessibility, and biophysical characteristics. In particular, the 192 synthetic control for Agrocortex is only 61% the size of their project area (SI Appendix, 193 Table A1-2). While historical deforestation is similar in the synthetic controls and project 194 areas, clearly there is future potential for more deforestation in the larger project areas than in 195 their smaller synthetic controls. Third, the construction of our synthetic controls may not 196 have included all relevant structural determinants of deforestation. Lastly, the period of 197 analysis may not have been long enough to observe significant REDD+ impacts in some 198 199 cases.

Despite these caveats, the weight of the evidence suggests that these projects caused 200 less reduction in deforestation than claimed (Fig. 5 & SI Appendix, Fig. S5) and that few 201 projects actually achieved emission reductions (30). Suspicion about the environmental 202 integrity of carbon offsets is not restricted to REDD+ or voluntary interventions. A series of 203 reports on other market-based initiatives for climate change mitigation, i.e., the Joint 204 205 Implementation (JI) and the Clean Development Mechanism (CDM) of the Kyoto Protocol, also raised concerns about the true climatic contributions from certified carbon offsets. These 206 reports suggest that about three-quarters of JI credits are unlikely to represent additional 207 208 emission reductions (31) and that 73% of the potential 2013–2020 CDM credits have a low likelihood of environmental integrity (in contrast to 7% with high likelihood) (32). 209

The projects that we evaluated may have had little additional impact because they did 210 not adopt the most effective actions to achieve their REDD+ objectives, perhaps because of 211 uncertainties about the future availability of funds and concerns about unfairly raising local 212 213 expectations of carbon payments. Hence, our results do not imply that voluntary REDD+ projects cannot achieve their objectives if designed and implemented effectively. There is 214 both quasi-experimental and experimental evidence that conditional payments for 215 environmental services (PES) can effectively reduce deforestation (3, 33), and recent 216 217 literature suggests that REDD+ implemented through well-designed conditional PES can deliver positive conservation outcomes (34–36). 218

Another possible explanation for the lack of impact is difficulty with the on-the-ground implementation and execution of activities envisioned by project proponents (37, 38). One example is the Suruí project, which attracted international attention as one of the first voluntary REDD+ interventions implemented in an indigenous territory (4). The project aimed to use the financial revenues from carbon sales to promote sustainable land-use practices in the Suruí territory but was not able to prevent the illegal invasion of loggers and miners.

A third possible explanation for under-performance relates to challenges with the 226 commercialization of carbon offsets and correspondingly limited revenues available to 227 implement project activities (39). One way that voluntary REDD+ projects overcome that 228 challenge is by claiming "retroactive credits" (40). Often, projects that are certified in a given 229 year claim to have started much earlier (SI Appendix, Table S1). As a result, those projects 230 231 are eligible to issue large amounts of carbon offsets at the time of certification, retroactively corresponding to the period between the certification and the project start date. This can help 232 233 to fund project start-ups, but it also implies that projects have not actually had access to

carbon revenues during their early years of operation. Carbon crediting rules may thuspartially explain why we find limited evidence for avoided deforestation.

Our results emphasize the need to reassess approaches to measuring project 236 additionality. While ex-post counterfactual methods such as illustrated here would ensure a 237 high level of environmental integrity, they would introduce substantial uncertainty about the 238 credits that can be obtained from a given reduction in deforestation in project areas. An 239 alternative approach often suggested in the literature is to require projects to adopt national or 240 subnational (jurisdictional) baselines that are predefined, and periodically updated, by the 241 242 government (6, 7, 41), as well as default carbon-stock values or a common carbon-density map (42). Imposing one common baseline would have the benefits of facilitating the 243 inclusion of carbon emission reductions claimed by decentralized initiatives into national 244 GHG emission inventories, ensuring consistency in the treatment of leakages, and avoiding 245 double-counting reductions (6, 8, 43), while still offering relative certainty about carbon 246 credits conditional on project performance. However, national and subnational baselines are 247 typically based on historical data and thus are not any more likely to capture 248 contemporaneous deforestation drivers and their dynamism (although it is also possible to 249 apply the synthetic control method to nations [30]). Thus, they do not address the main 250 251 problem identified by our analysis: the limitations of historical data for baseline development.

Periodic baseline updates based on recent deforestation trends could help mitigate the 252 influence of factors external to voluntary REDD+ projects on the carbon credits that they 253 claim. In fact, current VCS rules already require projects to revise their baselines every 10 254 years (16). Our results suggest that this interval should be shorter. Baseline updates could be 255 based on control areas that share similar characteristics as the REDD+ projects, as 256 demonstrated in this study with the construction of the synthetic controls. In addition, 257 coupled human-natural system models, such as was used in the Suruí case, can be used to 258 explore alternative baseline scenarios and quantify the potential downside risks involved in 259 conservation investments under dynamic patterns of land-use change, though at increased 260 261 project development costs (24). These models could also shed light on the potential impacts of REDD+ on local livelihoods and biodiversity (45, 46), which we do not consider here but 262 recognize as fundamentally important. 263

We do provide empirical evidence for a phenomenon that was anticipated in the early policy debate over REDD+ (47), i.e., *de facto* additionality of REDD+ projects depends on both project implementation and national circumstances. Carbon finance and crediting systems must safeguard against both "hot air" from overstated claims of carbon additionality and excessive risks to private conservation investments associated with desirable government
action to combat deforestation, as observed in Brazil from 2005 to 2012.

270 Materials and Methods

We examined the impacts of 12 voluntary REDD+ projects implemented in the 271 Brazilian Amazon since 2008 and certified under the Verified Carbon Standard (VCS) before 272 May 2019 to curb local unplanned deforestation (Fig. 2; SI Appendix, Tables S1 & S2). 273 Project areas were defined by the geospatial polygons reported by the project proponents and 274 available from the VCS project database. Ten of the 12 projects were implemented in 275 privately owned properties, whereas the other two, the Suruí and the Rio Preto-Jacundá 276 projects, were implemented in an indigenous territory and a sustainable-use reserve, 277 respectively. Following VCS-approved carbon-accounting methodologies, historical 278 deforestation rates were the basis of all project deforestation baselines with the exception of 279 the Suruí project (e.g., Fig. 1). In the latter, baseline deforestation rates were informed by a 280 281 participatory, and community-specific, system dynamics model (24).

Rigorous impact evaluations rely on the establishment of credible counterfactuals for 282 what would have happened in the absence of an intervention (48, 49), which are 283 unobservable. We construct "synthetic controls" to serve as counterfactuals for the REDD+ 284 project areas (19, 50). We adopted the synthetic control approach, as opposed to more 285 286 traditional methods from the impact evaluation literature (e.g., difference-in-differences estimator), because of our small number of treated units and likely heterogeneity of the 287 treatment across them (49, 51, 52). Synthetic controls were constructed as a weighted average 288 of selected donor units through a nested optimization procedure that minimizes the 289 290 differences in pretreatment characteristics between the project and the control, with characteristics weighted such that the resulting weighted average outcome of the selected 291 donor units most closely matches the pretreatment outcome in the treated unit (20, 21). 292 Specifically, the iterative procedure minimizes the mean squared prediction error (MSPE) of 293 294 the outcome, or the sum of squared residuals between the treated unit and the synthetic control, over the pretreatment period (50). 295

Two sets of synthetic controls were constructed as a weighted combination of areas selected from "donor pools" (19, 50) composed of Amazonian properties registered in the CAR database (22) that do not overlap with project areas and that had \geq 90% forest cover in the first year of the analysis. In the first set, we used cumulative deforestation as the optimization outcome, whereas the second set was based on annual deforestation. We note

that the optimization algorithm selected different groups of donors for the synthetic controls 301 for each outcome, which allows us to use the second set as a robustness check. Donor pools 302 were preferably based on properties from the same state as the REDD+ project and within 303 $\pm 25\%$ the size of the project area. Whenever the resulting synthetic controls had substantially 304 different land areas or pretreatment annual and cumulative deforestation (i.e., before project 305 306 implementation), the donor pools were expanded to all properties in the Amazon biome (see SI for details). Lastly, for the cases of persistent unbalanced synthetic controls, donor pools 307 were expanded to properties with $\pm 50\%$ the size of the project area. Synthetic controls for the 308 309 REDD+ projects implemented in a sustainable-use reserve (i.e., Rio Preto-Jacundá) and an indigenous territory (i.e., Suruí) were constructed based on donor pools composed of other 310 sustainable-use reserves and indigenous territories, respectively. 311

The spatial covariates structurally related to deforestation (29) used for the construction 312 of the synthetic controls were obtained from official maps produced by government agencies 313 314 in Brazil (SI Appendix, Fig. S7 & Table S4). The covariates represent (i) property size, (ii) initial forest cover, (iii) slope, (iv) soil quality, and distances from (v) state capitals, (vi) 315 316 towns, (vii) federal highways, and (viii) local roads, as well as the proportion of (ix) primary and (x) secondary forest, (xi) pastureland, (xii) agriculture, and (viii) urban areas in 2000, 317 318 2004, 2008, and 2012 (for projects implemented after 2012) within 10-km buffer zones of the project and potential donor areas. In accordance with the previous literature (20, 50), we also 319 320 used the pretreatment annual and cumulative deforestation rates to inform the construction of the two sets of synthetic controls. Temporal land-use information in the buffer zones was 321 322 obtained from the TerraClass dataset produced by Brazil's National Institute for Space Research (INPE). Annual deforestation data for the 2001–2017 period were processed from 323 the MapBiomas land-use/cover dataset v.3.1 for the Brazilian Amazon biome (Fig. 2 & SI 324 Appendix, Fig. S1). 325

While the construction of our synthetic controls was based on all information available 326 from 2001 to the project start year (i.e., pretreatment period), we conducted a separate 327 analysis in which a different set of synthetic controls were constructed based on data 328 constrained to the first-half of the pretreatment period (i.e., "training" period), so they could 329 be tested against the second-half (i.e., "testing" period; SI Appendix, Fig. S2). We evaluated 330 the outcome of this analysis both visually and by comparing training and testing MSPE (SI 331 Appendix, Table S3). This "proof of concept" differs from standard model-validation 332 practices because the donors selected as synthetic controls based on the first-half of the 333

pretreatment periods do not necessarily match the final set of donors when the fullpretreatment period is used.

We examined the robustness of our findings with a series of placebo tests, in which we 336 create synthetic controls for all CAR polygons in the donor pool (i.e., not subject to REDD+ 337 activities) and compute the difference in both annual and cumulative deforestation between 338 each placebo and its synthetic control (Fig. 4 & SI Appendix, Fig. S4). Because placebo areas 339 are not exposed to REDD+, any differences in forest loss between placebos and their 340 synthetic controls are statistical "noise." In order to increase the number of placebo tests, we 341 342 use the expanded placebo donor pools of all Amazonian properties with $\pm 50\%$ the project size. In accordance with the previous literature (21), we discarded placebo tests with 343 pretreatment MSPE five times higher than the pretreatment MSPE of the REDD+ polygon. 344 We used the gaps in deforestation between the placebos and their respective synthetic 345 controls to create 99% confidence intervals around the mean placebo effect estimate, which is 346 347 approximately zero in all cases. Analyses were conducted with the Synth package (v.1.1) available for R software (v.3.6.0) (50). Lastly, we computed the annual deforestation in 10-348 349 km *buffer zones* surrounding the project areas as an indicator of possible leakage effects (23), i.e., because increasing deforestation could reflect the displacement of deforestation due to 350 351 the REDD+ activities.

352 Acknowledgments

353 We thank two anonymous reviewers for valuable comments and suggestions. JB 354 acknowledges support from the German Federal Ministry of Education and Research.

355 **References**

- Angelsen A (2017) REDD+ as result-based aid: General lessons and bilateral agreements of Norway. *Rev Dev Econ* 21(2):237–264.
- 358 2. UN-REDD (2011) *The UN-REDD Programme Strategy 2011–2015* (Geneva).
- 359 3. Börner J, et al. (2018) National and subnational forest conservation policies—What works,
 360 what doesn't. *Transforming REDD+: Lessons and New Directions*, eds Angelsen A, et al.
 361 (Center for International Forestry Research (CIFOR), Bogor), pp 105–116.
- 362 4. West TAP (2016) Indigenous community benefits from a de-centralized approach to REDD+
 363 in Brazil. *Clim Policy* 16(7):924–939.
- 364 5. Donofrio S, Maguire P, Merry W, Zwick S (2019) *Financing Emissions Reductions for the* 365 *Future: State of the Voluntary Carbon Markets 2019* (Washington, D.C.).
- 366 6. Lee D, Llopis P, Waterworth R, Roberts G, Pearson T (2018) *Approaches to REDD+ nesting:* 367 *Lessons learned from country experiences* (World Bank, Washington, D.C.).
- 368 7. Verified Carbon Standard (2017) *Jurisdictional and Nested REDD*+ (*JNR*) *Requirements*369 (Verified Carbon Standard, Washington, D.C.).

370 371 372	8.	FAO (2019) From reference levels to results reporting: REDD+ under the United Nations Framework Convention on Climate Change. 2019 update (Food and Agriculture Organization, Rome).
373 374 375	9.	West TAP, et al. (2018) A hybrid optimization-agent-based model of REDD+ payments to households on an old deforestation frontier in the Brazilian Amazon. <i>Environ Model Softw</i> 100:159–174.
376 377 378	10.	Dezécache C, Salles J-M, Hérault B (2018) Questioning emissions-based approaches for the definition of REDD+ deforestation baselines in high forest cover/low deforestation countries. <i>Carbon Balance Manag</i> 13(1):21.
379 380	11.	West TAP, Börner J, Fearnside PM (2019) Climatic Benefits From the 2006–2017 Avoided Deforestation in Amazonian Brazil. <i>Front For Glob Chang</i> 2:52.
381 382	12.	Ferrante L, Fearnside PM (2019) Brazil's new president and 'ruralists' threaten Amazonia's environment, traditional peoples and the global climate. <i>Environ Conserv</i> 46(4):261–263.
383 384	13.	Mertz O, et al. (2018) Uncertainty in establishing forest reference levels and predicting future forest-based carbon stocks for REDD+. <i>J Land Use Sci</i> 13(1–2):1–15.
385 386	14.	Rifai SW, West TAP, Putz FE (2015) "Carbon Cowboys" could inflate REDD+ payments through positive measurement bias. <i>Carbon Manag</i> 6(3–4):151–158.
387 388	15.	Seyller C, et al. (2016) The "virtual economy" of REDD+ projects: does private certification of REDD+ projects ensure their environmental integrity? <i>Int For Rev</i> 18(2):231–246.
389	16.	Verra (2019) VCS Standard v4.0 (Washington, D.C.).
390 391 392	17.	Sills EO, et al. (2017) Building the evidence base for REDD+: Study design and methods for evaluating the impacts of conservation interventions on local well-being. <i>Glob Environ Chang</i> 43:148–160.
393 394	18.	Duchelle AE, Simonet G, Sunderlin WD, Wunder S (2018) What is REDD+ achieving on the ground? <i>Curr Opin Environ Sustain</i> 32:134–140.
395 396	19.	Sills EO, et al. (2015) Estimating the impacts of local policy innovation: The synthetic control method applied to tropical deforestation. <i>PLoS One</i> 10(7):1–15.
397 398	20.	Abadie A, Gardeazabal J (2003) The economic costs of conflict: A case study of the Basque country. <i>Am Econ Rev</i> 93(1):113–132.
399 400 401	21.	Abadie A, Diamond A, Hainmueller J (2010) Synthetic control methods for comparative case studies: Estimating the effect of California's tobacco control program. <i>J Am Stat Assoc</i> 105(490):493–505.
402 403	22.	Azevedo AA, et al. (2017) Limits of Brazil's Forest Code as a means to end illegal deforestation. <i>Proc Natl Acad Sci</i> 114(29):7653–7658.
404 405	23.	Aukland L, Costa PM, Brown S (2003) A conceptual framework and its application for addressing leakage: The case of avoided deforestation. <i>Clim Policy</i> 3(2):123–136.
406 407	24.	Vitel CSMN, et al. (2013) Land-use change modeling in a Brazilian indigenous reserve: Construction of a reference scenario for the Suruí REDD Project. <i>Hum Ecol</i> 41(6):807–826.
408 409	25.	Hamrick K, Gallant M (2017) <i>Fertile Ground: State of Forest Carbon Finance 2017</i> (Forest Trends' Ecosystem Marketplace, Washington, D.C.).
410 411	26.	Riksrevisjonen (2018) The Office of the Auditor General of Norway's investigation of Norway's International Climate and Forest Initiative (Bergen).
412 413	27.	Börner J, Kis-Katos K, Hargrave J, König K (2015) Post-crackdown effectiveness of field- based forest law enforcement in the Brazilian Amazon. <i>PLoS One</i> 10(4):1–19.
414	28.	Assunção J, Gandour C, Rocha R (2015) Deforestation slowdown in the Brazilian Amazon:

- 415 prices or policies? *Environ Dev Econ* 20(06):697–722.
- 416 29. Busch J, Ferretti-Gallon K (2017) What drives deforestation and what stops it? A meta417 analysis. *Rev Environ Econ Policy* 11(1):3–23.
- 418 30. Bos AB, et al. (2017) Comparing methods for assessing the effectiveness of subnational
 419 REDD plus initiatives. *Environ Res Lett* 12(7). doi:10.1088/1748-9326/aa7032.
- 420 31. Kollmuss A, Schneider L, Zhezherin V (2015) *Has Joint Implementation reduced GHG*421 *emissions? Lessons learned for the design of carbon market mechanisms* (Stockholm).
- 422 32. Cames M, et al. (2016) *How additional is the Clean Development Mechanism?* (Berlin).
- 33. Sills EO, Jones K (2018) Causal inference in environmental conservation: The role of
 institutions. *Handbook of Environmental Economics*, eds P. D, S.K. P, V. KS (Elsevier B.V.,
 Amsterdam), pp 395–437.
- 426 34. Jayachandran S, et al. (2017) Cash for carbon: A randomized trial of payments for ecosystem
 427 services to reduce deforestation. *Science* 357(6348):267–273.
- 35. Simonet G, Subervie J, Ezzine-de-Blas D, Cromberg M, Duchelle AE (2019) Effectiveness of
 a REDD+ Project in Reducing Deforestation in the Brazilian Amazon. Am J Agric Econ
 101(1):211–229.
- 431 36. Cuenca P, Robalino J, Arriagada R, Echeverría C (2018) Are government incentives effective for avoided deforestation in the tropical Andean forest? *PLoS One* 13(9):1–14.
- 433 37. Simonet G, et al. (2018) Forests and carbon: The impacts of local REDD+ initiatives.
 434 *Transforming REDD+: Lessons and New Directions*, eds Angelsen A, et al. (Center for
 435 International Forestry Research, Bogor), pp 117–130.
- 436 38. Duchelle AE, Sassi C De, Sills EO, Wunder S (2018) People and communities. *Transforming*437 *REDD+: Lessons and New Directions*, eds Angelsen A, et al. (Center for International
 438 Forestry Research, Bogor), pp 131–141.
- 439 39. Laing T, Taschini L, Palmer C (2016) Understanding the demand for REDD+ credits. *Environ* 440 *Conserv* 43(4):389–396.
- 441 40. Linacre N, R. O, Ross D, Durschinger L (2015) *REDD+ Supply and Demand 2015-2025* (Washington, D.C.).
- 443 41. Pedroni L, Dutschke M, Streck C, Porrúa ME (2009) Creating incentives for avoiding further
 444 deforestation: the nested approach. *Clim Policy* 9(2):207–220.
- 445 42. Asner GP, et al. (2010) High-resolution forest carbon stocks and emissions in the Amazon.
 446 *Proc Natl Acad Sci* 107(38):16738–16742.
- 447 43. Schneider L, et al. (2019) Double counting and the Paris Agreement rulebook. *Science* 366(6462):180–183.
- 449 44. Roopsind A, Sohngen B, Brandt J (2019) Evidence that a national REDD+ program reduces
 450 tree cover loss and carbon emissions in a high forest cover, low deforestation country. *Proc*451 *Natl Acad Sci* 116(49):24492–24499.
- 45. West TAP, et al. (2018) Impacts of REDD+ payments on a coupled human-natural system in
 453 Amazonia. *Ecosyst Serv* 33:68–76.
- 454 46. Iwamura T, Lambin EF, Silvius KM, Luzar JB, Fragoso JMV (2016) Socio-environmental
 455 sustainability of indigenous lands: Simulating coupled human-natural systems in the Amazon.
 456 Front Ecol Environ 14(2):77–83.
- 457 47. Angelsen A (2008) *Moving ahead with REDD: Issues, Options and Implications* (Center for International Forestry Research, Bogor).
- 459 48. Holland PW (1986) Statistics and Causal Inference. J Am Stat Assoc 81(396):945–960.

- 460 49. Rubin DB (1974) Estimating causal effects of treatments in randomized and nonrandomized
 461 studies. *J Educ Psychol* 66(5):688–701.
- 462 50. Abadie A, Diamond A, Hainmueller J (2011) Synth: An R package for synthetic control methods in comparative case studies. *J Stat Softw* 42(13):1–17.
- 464 51. Rosenbaum PR, Rubin DB (1983) The central role of the propensity score in observational studies for causal effects. *Biometrika* 70(1):41–55.
- 466 52. Abadie A, Diamond A, Hainmueller J (2015) Comparative politics and the synthetic control method. *Am J Pol Sci* 59(2):495–510.

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Figure legends

Fig. 1. Annual deforestation in the Brazilian Amazon from PRODES data (bars). Blue bars indicate voluntary REDD+ project start dates. Red lines represent 10-year deforestation averages prior to project implementation and commonly adopted as projects' deforestation baselines.

Fig. 2. VCS-certified REDD+ projects established during 2008–2017 in the Brazilian Amazon forest biome.

Fig. 3. Cumulative post-2000 deforestation in Amazonian areas with REDD+ projects (red) versus synthetic controls (blue). Dashed black lines are the project start dates.

Fig. 4. Placebo tests: cumulative deforestation in REDD+ project areas minus deforestation in their respective synthetic controls (red), and placebos minus their respective synthetic controls (blue dots). Dashed black lines are the project start dates (assumed the same for placebos). Shaded blue areas represent 99% confidence intervals around the mean of the placebos. The number of placebos varies by project based on whether synthetic controls with low mean squared prediction error could be constructed for the placebo tests.

Fig. 5. Cumulative deforestation from the baseline scenarios adopted by the REDD+ projects (orange) versus observed cumulative deforestation in the synthetic controls (blue). Dashed black lines are the project start dates.