

1 **The role of soil carbon in natural climate solutions**

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20 **Summary**

21 Mitigating climate change requires clean energy and removing atmospheric carbon. Building
22 soil carbon is an appealing way to increase carbon sinks and reduce emissions due to the
23 associated benefits to agriculture. However, practical implementation of soil carbon climate
24 strategies lag behind the potential, partly because we lack clarity around the magnitude of
25 opportunity and how to capitalize on it. Here we quantify the role of soil carbon in natural (land-
26 based) climate solutions (NCS), and review some of the project design mechanisms available to
27 tap into the potential. We show that soil carbon represents 25% of the 23.8 GtCO₂yr⁻¹ NCS
28 potential of which 40% is protection of existing soil carbon and 60% is rebuilding depleted
29 stocks. Soil carbon comprises 9% of the mitigation potential of forests, 72% for wetlands, and
30 47% for agriculture and grasslands. Soil carbon is important to land-based efforts to prevent
31 carbon emissions, remove atmospheric carbon dioxide and deliver ecosystem services in addition
32 to climate mitigation.

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34 Protecting and restoring soil organic matter delivers many benefits to people and
35 nature^{1,2}. Globally, soils hold three times more carbon than the atmosphere³, and the role of soil
36 organic matter as a regulator of climate has been recognized by scientists for decades⁴. Recent
37 work has highlighted the historical loss of carbon from this pool³, and the threat of future
38 accelerated loss under warming scenarios^{4,5}. Soil organic carbon as a natural climate solution
39 (NCS) thus has a role both through restoring a carbon sink and protecting against further CO₂
40 emissions in response to predicted land use change and climate change.

41 This dual role for soil in the global carbon budget suggests climate benefits can be
42 achieved through strategies that both conserve existing soil organic carbon stocks (avoid loss),
43 and restore stocks in carbon-depleted soils⁶. There are important additional benefits. Protecting
44 and increasing soil carbon storage can (i) protect or increase soil fertility, (ii) maintain or
45 increase resilience to climate change, (iii) reduce soil erosion, and where implemented through
46 conservation of natural ecosystems iv) reduce habitat conversion, all in line with the United
47 Nations Sustainable Development Goals (SDG's)⁷, the goals of the United Nations Framework
48 Convention on Climate Change (UNFCCC) and the United Nations Convention on Combating
49 Desertification (UNCCD). As such, soil carbon is promoted as a common denominator amongst
50 a variety of global and national initiatives⁷. Although recent academic comment and perspective
51 pieces point the way towards accelerated action on soils^{8,9}, there remains much uncertainty
52 around actionable pathways for achieving the global opportunity. Here we examine the scientific
53 and policy context surrounding soil carbon projects, to aid prioritization and decision making.

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57 **Status of soil carbon as a climate solution**

58 Despite scientific consensus around its potential and multiple benefits, deployment of soil
59 carbon storage and sequestration for climate mitigation remains limited in practice. There is
60 growing interest in soil in international climate mitigation conversations, with the recognition of
61 ‘wetland drainage and rewetting’ (WDR) as an accounting option under the Kyoto Protocol
62 (formalized in 2011), the launch of the ‘4 per 1000’ Initiative in Paris in 2015 and formal
63 recognition of soil carbon sequestration in the UNFCCC process in 2017 (COP23 decision
64 4/CP.23). To date there are only a few dozen projects that address soil organic carbon in
65 registered compliance or voluntary carbon markets. Fewer than 60 projects (half of them in
66 Australia) provided under 50 thousand tonnes of CO₂-equivalent removals by soil in agriculture
67 and grassland projects per year¹⁰. This is less than 0.0001% of the estimated mitigation
68 potential¹¹. As a comparison there are 1,500 carbon projects covering 12 Mha of land in the
69 forest sector¹². The small soil carbon numbers are due in part to the sector’s near exclusion from
70 early carbon market mechanisms, notably the Kyoto Protocol’s Clean Development Mechanism
71 (CDM) which limited potential soil carbon mitigation to afforestation and reforestation projects.
72 Nevertheless, the past two decades have witnessed the emergence of a variety of robust
73 methodological approaches for the calculation of mitigation benefits and the issuance of ‘carbon
74 credits’ in a wide range of project categories covering croplands, grasslands, savannahs, as well
75 as peatlands and coastal wetlands. While still occupying no more than a niche in the toolbox for
76 international climate action, there is experience on soil carbon projects to provide confidence and
77 to support development of mitigation plans at larger scales¹⁰.

78 Experience with implementation has not yet caught up with aspirations in the political
79 arena. While soil targets for mitigation are included in only eight Nationally Determined

80 Contributions (NDCs) to the UNFCCC⁹, the UNFCCC is now exploring agriculture and soils –
81 including with respect to “[improved] soil carbon, soil health and soil fertility under grassland
82 and cropland as well as integrated systems, including water management” as a more explicit part
83 of their agenda¹³. At the same time, nations are moving forward to invest in solutions and set
84 targets that address the food security and land use commitments of the United Nations
85 Sustainable Development Goals (SDGs). Beyond governments, a growing number of companies
86 are including soil organic carbon within their set of options to build the resilience and long-term
87 profitability of agricultural value chains⁹. The enthusiasm arises because, in general, soil carbon
88 enhancement practices are considered to have positive co-benefits, do not require additional land
89 area, have minimal water footprint, and are readily deployable considering that they do not
90 require changes in land use^{11,14}.

91 The science supporting the global technical potential of soil carbon mitigation is
92 relatively well established, even though measuring changes in soil carbon is more difficult than
93 for plant biomass. Recent estimates of global soil carbon sequestration technical potential, i.e.
94 the level of mitigation that could be achieved when accounting only for biophysical constraints,
95 if there were no economic, social, institutional or other barriers to implementation, align around
96 2-5 Gt CO₂ per year^{11, 14-18} albeit many of these estimates rely on the same underlying data.
97 Counter to this relative certainty, recent scholarly debates focused primarily on debunking claims
98 that soil carbon sequestration could fully offset current increases in atmospheric CO₂¹⁹⁻²¹ have
99 created confusion for practitioners. Yet, even these debates do not call into question the
100 significance of the global potential, and the multiple benefits of increasing global soil carbon
101 stocks.

102 Caveats surrounding soil carbon sequestration such as sink saturation and non-
103 permanence risk (reversibility) have also been well explored in the soil science literature. Soil
104 carbon saturation refers to a maximum capacity of the soil to retain organic carbon¹⁵, meaning
105 that soil organic carbon does not increase indefinitely with the exception of some wetland
106 systems¹⁶. For most improved carbon management practices the rate at which soils will store
107 additional carbon therefore begins to decline after some decades, and eventually will reach a new
108 steady state when a higher carbon stock is achieved. The time period before a new steady state is
109 reached will vary greatly depending on soil type, management intervention, climate regime, and
110 pre-existing SOC depletion¹⁵, but is generally on the order of decades²². This aligns with the
111 need to reduce peak atmospheric CO₂ levels and mitigate peak warming. With respect to non-
112 permanence, maintaining high SOC stocks, such as with cover cropping and manuring in
113 croplands, requires some form of maintenance (continuation of improved soil carbon
114 management practices), even after a new steady-state is reached and no further mitigation
115 benefits accrue¹⁴. In other cases, i.e. when there is protection of existing soil carbon stocks, such
116 as avoided grassland conversion, it is likely that SOC levels are at steady-state, and the
117 management activity (in this case protection) also needs to be maintained to maintain those SOC
118 stocks²³. Nevertheless, SOC may be more resilient to fire, pests, and wind than carbon in
119 aboveground biomass in many environments¹⁷, and some forms of soil carbon, such as biochar,
120 can persist for millennia¹⁸.

121 Meanwhile, outside of soil science, carbon project design approaches have moved
122 forward to deal with heterogeneity, uncertainty, additionality, and non-permanence in particular
123 which are challenges for the entire Agriculture, Forestry and Land Use (AFOLU) sector. Soil
124 does not differ substantially from forestry in this regard, and because this has been a topic for

125 decades, significant experience exists in managing these risks as part of project and policy
126 design²⁴. Some methods to account for and resolve these issues in soil carbon project design are
127 reviewed by²⁵. The CDM issues temporary credits that are continuously renewed as long as the
128 removal benefit persists. If a reversal event occurs, renewal of the temporary credit concerned is
129 no longer possible (Decision 5/CMP.1; Decision 14/CMP.1).

130 An alternative approach to non-permanence of soil carbon sequestration is based on the
131 installation of portfolio-wide “buffer” reserves – each project contributes with a share of the
132 credits achieved – that works as an insurance scheme. For any event of intentional (subsequent
133 land degradation, land conversion) or unintentional (usually force majeure events such as
134 extreme weather events, storms, flooding, fire, and so on) that causes sink reversals or carbon
135 stock losses reversal risks, credits held in the buffer account will be released (in an amount
136 equivalent to the reversal or stock loss amount) and permanently canceled²⁶. Most voluntary
137 carbon market standards operate with a buffer reserve²⁷ based on some standardized or project-
138 specific risk assessment. In Australian carbon farming associated with the government’s land
139 based strategies for climate mitigation follow a mixed approach that combines buffer reserves
140 with discount elements: farmers that would receive a certain amount of credits in a 100-year
141 permanence scenario (with maintenance obligations being transferred to subsequent landowners
142 within the 100-year window) will receive 20% less credits if they commit to 25-year stable
143 conditions only¹⁰; the discount comes on top of the general 5% buffer amount. No case is known
144 in which a buffer reserve was ever depleted, which suggests that, while important, permanence is
145 a manageable issue. As a caveat, this experience arises primarily from the forest sector, and
146 given that most soil carbon projects in the agriculture sector are relatively new, there has been
147 little time for permanence issues to appear. Soil carbon sequestration ambitions can benefit from

148 this experience in the markets and the accepted protocols that now exist for most types of soil
149 carbon sequestration project types including for grasslands, peatlands and croplands¹⁰.

150 Practical solutions aside, the relevance of the non-permanence issue is also fading²⁸.
151 While of great importance in the context of project-level offsetting, the non-permanence risk of
152 mitigation action within wider jurisdictional or national schemes is less a concern of
153 environmental integrity but of legal responsibility (liability). Within the Paris Agreement, in
154 particular, nations are expected “to include all categories of anthropogenic emissions or removals
155 in their nationally determined contributions and, once a source, sink or activity is included,
156 continue to include it” (Decision 1/CP.21, paragraph 31.c). Once soil carbon emissions are thus
157 covered under a target, the non-permanence issue in specific measures is solved at the higher-
158 level accounting framework: Any reversal events will translate into a fresh obligation (a priori
159 for the government) to reduce or avoid emissions. As with permanence, issues of additionality
160 and leakage require strong safeguards and binding agreements. Australia’s direct action subsidy
161 approach may fund non-additional projects and therefore deliver less abatement than expected²⁹.

162 There are several other challenges to implementation of soil as a climate mitigation
163 strategy. Historically, there have been limited finance and policy options. The Kyoto
164 mechanisms failed to address soil carbon interventions. Then, carbon prices (the price paid per
165 tCO₂e) collapsed following the 2008 global economic recession and the Copenhagen summit in
166 2009¹⁰ failed to generate a new agreement. Further, carbon pricing currently covers only about
167 20% of global emissions. However, there are some signs that viability of climate financing for
168 soil is improving. There is increased action on agriculture under the Paris agreement. The Green
169 Climate Fund has established a funding window targeting land-use and agriculture. There are a
170 range of fresh private-sector initiatives on soil carbon that promise sufficient funding and

171 transformational change^{30,31}, and impact investors focusing on landscape, soil resources, and
172 payments for ecosystem services schemes¹⁰.

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174 **Soil contribution to Natural Climate Solution Pathways**

175 Experience and trends in the AFOLU market sector, emerging finance opportunities for
176 climate positive agriculture, and earlier global potential analyses provide the framework for
177 actions on soil carbon. Here we extend the analysis of Griscom et al. ³² to offer improved
178 guidance on the set of actions available for realizing the soil carbon climate mitigation
179 opportunity. The recent study by Griscom et al. ³² provides a framework for an integrated
180 assessment of the overall global mitigation potential of “natural climate solutions” (NCS). In the
181 Griscom et al. ³² study the potential of 20 conservation, restoration and improved land
182 management actions, including reforestation, planting trees in croplands, grazing land
183 management, peatland protection and others, to increase carbon storage and/or avoid greenhouse
184 gas emissions across global forests, wetlands, grasslands, and agricultural lands was determined
185 to be 23.8 Gt CO₂e yr⁻¹. This analysis estimated mitigation potentials constrained by a
186 requirement for additionality and by food security and biodiversity safeguards. A benefit of this
187 analysis is that researchers, policy makers and practitioners can prioritize across various sectors
188 of potential activity. An additional benefit is that by using a common framework, the analysis
189 avoids double counting across the various mitigation options, referred to as “pathways” - an
190 important consideration for national accounting with NDC commitments. While soil-related
191 ecosystem services are identified as a co-benefit in 16 of the 20 pathways, the specific
192 contribution of soil carbon storage (avoided losses and enhanced sinks) to each of these
193 pathways, and overall, was accounted for but not reported as a component distinct from biomass

194 carbon. Here we elaborate on Griscom et al.³² by incorporating findings from a few key papers
195 published since 2017 and by separating out the contribution of soils to each pathway (see
196 methods). Table 1 describes the soil carbon protection and sequestration pathways, the annual
197 mitigation potential and benefits for sustainability.

198 Our results (Figure 1) show the global additional mitigation potential of protecting and
199 rebuilding soil carbon to be 5.5 Gt CO₂e yr⁻¹, representing 25% of the total mitigation potential
200 of the 20 NCS pathways. Of this, 4.3 Gt CO₂e yr⁻¹ comes from non-forest pathways, thus soil
201 carbon represents more than half of the 7.6 Gt CO₂e yr⁻¹ NCS potential of non-forested lands,
202 with safeguards for food security, fiber security and biodiversity conservation. Avoidable losses
203 represent 2.2 Gt CO₂e yr⁻¹, or 40%, of the total soil carbon mitigation potential of all NCS
204 pathways. Protection is important not only because the potential is large, but also because soil
205 carbon is lost more quickly than it can be gained³³, and in many cases it is not possible to restore
206 to soil organic carbon to original levels on climate-relevant timescales^{3,34}. These estimates do not
207 include land or agricultural management practices that reduce non-CO₂ GHG emissions (i.e. N₂O
208 and CH₄) without protecting or enhancing soil carbon sinks, for example improved rice, nutrient
209 and livestock management strategies, which together constitute an additional 1.85 Gt CO₂e yr⁻¹
210 ³².

211 The predominance of SOC protection and sequestration within the overall contribution of
212 NCS differs among different biomes (Figure 2). Across forest pathways, the SOC mitigation
213 potential of 1.2 Gt CO₂e yr⁻¹ is a small portion (9%) of the total and is split almost equally
214 between increased sequestration from reforestation and avoidable emissions through prevented
215 conversion. In grasslands and agriculture, 47% of the total potential mitigation (2.3 Gt CO₂e yr⁻¹)
216 arises from soil carbon protection and sequestration, while 20% involves other greenhouse gases

217 involved with improved soil management practices. In wetland pathways soil carbon is estimated
218 to comprise 2.0 Gt CO₂e yr⁻¹, 72% of the total mitigation potential of wetland pathways. For
219 forest pathways, soil carbon can bring an additional component to mitigation accounting which is
220 largely dominated by the above ground tree biomass, while in wetland pathways soil carbon is
221 the main vehicle through which climate mitigation can be achieved (Table 1). In agriculture and
222 grassland pathways overall, soil carbon is approximately half of the abatement potential, and
223 importantly accounting for soil carbon can bring large areas of grasslands and croplands under
224 the Paris Agreement.

225 About half of the soil carbon mitigation potential, 2.8 Gt CO₂e yr⁻¹, is considered cost-effective
226 at \$100/tCO₂ (based on Griscom et al. ³²methodology) which is one estimate of the amount that
227 society is expected to have to pay to mitigate climate change³⁵. About one quarter, 1.2 Gt CO₂e
228 yr⁻¹, is considered to be low cost at \$10/tCO₂. Low-cost and cost-effective removal is, therefore,
229 equivalent to about 3% and 7%, respectively, of recent annual anthropogenic emissions of
230 carbon dioxide to the atmosphere. In other studies, negative costs have been estimated for soil
231 carbon sequestration, based on the co-benefits such as increased productivity and resilience of
232 soils³⁶, and have suggested that many soil-based NCS are cost-effective even without supportive
233 climate policy. The IPCC recently concluded that the cost for soil carbon mitigation is below
234 \$100/tCO₂³⁷. Despite the relatively low or negative costs, soil carbon actions are not yet
235 implemented due to other economic, social, institutional or other barriers as noted and
236 highlighted above.

237

238 **Soil science knowledge gaps**

239 Given the availability of project design mechanisms to realize the potential for soil
240 carbon mitigation actions (see Table 2 for example actions for each pathway), soil management
241 planning and prioritization at various scales would benefit from increasingly more accurate
242 system and practice-specific estimates of climate impacts. For agriculture and grassland
243 pathways, future work should disaggregate mitigation accounting to specific activities each with
244 their own mitigation estimates, trade-offs, and co-benefits. Tillage, cover cropping, enhanced
245 crop rotations and grazing management are in fact broad sets of activities, each with potentially
246 very different impacts on soil organic carbon⁸, different N₂O emissions, and different
247 feasibilities. An activity that builds organic carbon on one soil type might be ineffective on a
248 different soil³⁸. In wetland pathways, more research should focus on accurately predicting the
249 magnitude of increasing CH₄ emissions when soil organic carbon is restored in wetland
250 environments, and improving estimates of the potential and existing carbon storage in peatland
251 soils.

252 Our estimates are lower overall than Fuss et al.¹¹ for the sequestration pathways, and
253 lower for agriculture than Zomer et al.³⁹, which used unconstrained cropland area availability.
254 We provide conservative estimates because we exclude interventions for which there is less
255 consensus on the impact, such as no-till⁴⁰ and we use conservative estimates for pathways with a
256 large range in published numbers, such as biochar^{41,42} and optimal grazing⁴³. Thus, agricultural
257 pathways in our analysis encompass only the best understood options for incremental change to
258 existing farming practices. Opportunities for greater innovation may result in higher per hectare
259 mitigation rates than those reflected here, but data are lacking to make robust global estimates of
260 their potential. Regenerative agriculture, organic farming, agroecology, silvo-pasture, climate
261 smart agriculture, agroforestry, and permaculture are all complex and not mutually exclusive

262 agricultural systems that can have significant positive impacts on soil organic carbon in specific
263 geographies, according to a recent literature review by Toensmeier⁴⁴. Other, less well-
264 established, opportunities for SOC management take advantage of the potential to build organic
265 matter into deeper soil layers through deep rooted grasses and new crop varieties⁴⁵, and deep
266 inversion techniques⁴⁶. Organic biosolids from cities are a large pool of organic material that are
267 often a pollution and waste disposal problem⁴⁷, that could provide substrate to build soil health
268 and sequester carbon in soils. Exogenous organic matter additions can stimulate rangeland
269 productivity and sequester endogenous organic matter beyond the actual tonnage of
270 compost/biosolids applied⁴⁸, but may pose a risk to native plant biodiversity⁴⁹ and more research
271 is needed (and is underway) to understand how universal these findings are. Early research from
272 row-crop systems suggests endogenous vs. exogenous organic matter have similar effects⁵⁰.

273 Soil carbon fluxes associated with forest pathways are often ignored, given the more
274 obvious changes observed in woody biomass, even though the contribution from forest pathways
275 to soil carbon sequestration is substantial (Fig 1). Conversion of forests to permanent croplands
276 and pastures often generates soil carbon emissions and forest restoration is expected to increase
277 soil carbon³⁴. Recent estimates for the extent of potential reforestation vary widely^{51,52}, and our
278 estimate is based on an intermediate spatial extent of potential reforestation (6.8 M km²), and
279 includes food security and biodiversity safeguards³². However, the potential for additional soil
280 carbon storage from improved management practices on natural and plantation forests are much
281 more complex and more research is needed to include the potential soil carbon benefits in this
282 NCS framework.

283

284 **Looking forward**

285 As the urgency to harness all available opportunities to mitigate catastrophic climate
286 change grows^{53,54}, we emphasize that if we are to limit warming well below 2°C as called for by
287 the Paris Agreement, soil carbon can be an important way to increase carbon sinks and reduce
288 emissions. Soil carbon sequestration is not an alternative to emission reductions in other sectors,
289 but rather an additional opportunity for increasing currently insufficient ambition in existing
290 nationally determined contributions (NDCs) to the Paris Agreement. This opportunity should
291 neither be dismissed nor exaggerated. Our analysis disaggregates this opportunity across all land
292 sectors in a way that is relevant to target setting and prioritization efforts at scales from NDCs to
293 sub-national programs.

294 A strong benefit of soil carbon mitigation action is that it can positively engage rural
295 landowners and the agricultural sector as beneficiaries of mitigation incentives that are likely to
296 be produced by successful climate negotiations. Further, the majority of soil carbon pathways are
297 “no regrets” opportunities for climate mitigation, by delivering improved soil fertility, climate
298 resilience and other ecosystem services in addition to climate mitigation. As such, soil carbon
299 aligns targets across different international conventions (SDG, UNFCCC, UNCCD) and agendas
300 by providing measurable benefits towards diverse goals with a common metric. Prospects for soil
301 carbon sequestration action are promising because project design tools are sufficient to address
302 accounting challenges, and climate financing seems to be growing for the sector; notably,
303 because enhancing soil carbon brings multiple benefits, there are opportunities to incentivize
304 action beyond formal carbon markets. Policies in both the climate sector with a focus on
305 mitigation and the agriculture sector with a focus on soil health are needed to achieve significant,
306 cost effective soil carbon protection and enhancement to meet climate targets and improve
307 resilience.

308

309 **Methods**

310 **Estimating soil carbon mitigation potential in NCS pathways**

311 Griscom et al. ³² identified 20 pathways by which natural systems could contribute to
312 mitigation of greenhouse gases. For these pathways, an analysis of over 300 publications was
313 conducted in concert with expert elicitation to define the maximum areal extent, the amount of
314 avoided emissions or sequestration rate (“flux”) and time until a new steady-state, and the
315 amount of total mitigation attainable at different costs informed by marginal abatement curves.
316 For complete sources see Griscom et al. ³² supplementary information. Pathways were
317 constructed carefully to estimate additional annual mitigation potential above a business-as-usual
318 baseline, to avoid double counting and to safeguard biodiversity and human needs for food, fiber
319 and fuel. The analysis also included estimates of uncertainty around extent, flux and mitigation
320 for each pathway, and propagated across all pathways. In this current work, we have separated
321 out the soil contribution of each pathway as briefly described below; full details of pathway
322 methods are found in Griscom et al. ³²:

- 323 • Avoided conversion of forested ecosystems (>25% tree cover) where they are threatened
324 by agriculture preventing the loss of soil carbon. Don et al⁵⁵ estimated that 17.4 Mg C
325 ha⁻¹ are lost when forests are converted to various commercial agricultural uses. Powers
326 et al. ⁵⁶ further found that conversion of forests for shifting cultivation results in a slightly
327 lower impact to soil carbon stocks (14.5 Mg C ha⁻¹). These avoided emission values were
328 then applied to the 5.93 Mha of tropical forest that are lost annually with the assumption
329 that 54% of the loss goes to commercial agriculture and the remainder to shifting
330 cultivation. Most temperate and all boreal regions excluded due to lack of spatial data

331 and/or albedo considerations. Forested wetlands excluded to avoid double-counting with
332 wetland pathways.

333 • Soil carbon sequestration arising through reforestation, including silvopastoral practices.

334 The reforestation pathway quantifies potential conversion from non-forest (< 25% tree
335 cover) to forest (> 25% tree cover) in areas that historically supported forests. This
336 pathway excludes afforestation of grass-dominated ecosystems to avoid negative

337 biodiversity impacts on grassland ecosystem^{57,58} and croplands for food security reasons.

338 The pathway does allow for reforestation of potentially forested grazing lands based on

339 recent analyses that show the potential to shrink the footprint of livestock production

340 through improved efficiencies in production and/or shifts towards a more plant-based

341 diet^{59,60}, but to avoid double counting, the mitigation potential from grazing pathways

342 was deducted from the mitigation potential for reforestation. To further avoid double

343 counting, the area of reforestation opportunity excluded wetland areas. Finally, the

344 reforestation pathway did not include opportunity assessments in boreal zones, since

345 changes in albedo can offset the climate benefits of carbon capture⁶¹, and excluded

346 opportunity within denser human settlements where widespread tree cover expansion is

347 constrained. The original NCS assessment included an average soil carbon accumulation

348 rate of 0.4 Mg C ha⁻¹ yr⁻¹ for tropical and subtropical reforestation from Powers et al⁵⁶,

349 which we disaggregated here. We then further quantified the soil carbon accumulation for

350 temperate forests using a more recent study by Nave et al³⁴. This analysis estimated that

351 reforesting stands accumulated between 0.11 to 0.34 Mg C ha⁻¹ yr⁻¹ in the topsoil. We

352 therefore used the midpoint of this range (0.23 Mg C ha⁻¹ yr⁻¹) to estimate potential soil

353 accumulation in temperate biomes.

- 354 • Biochar amendment to increase the soil carbon pool of agricultural soils is a soil-only
355 pathway in Griscom et al. ³² and remains unchanged in this analysis. Increased soil
356 carbon pool results from conversion of non-recalcitrant carbon (crop residue biomass) to
357 recalcitrant carbon (charcoal) through pyrolysis. Biochar carbon mitigation was estimated
358 using a mid-range estimate of available crop residues and multiplying this value by the
359 amount of persistent biochar assuming 79% is recalcitrant, a 50% conversion efficiency
360 during pyrolysis and a carbon content of crop residues of 45% of available crop residues.
- 361 • Cover cropping is a soil-only pathway in Griscom et al. ³², and remains unchanged in this
362 analysis. We assumed that 50% of the 800 Mha of cropped land were amenable to cover
363 cropping. To this area we applied a mean sequestration rate of 0.32 Mg C ha⁻¹ yr⁻¹ ⁶².
364 Effects of no-till and other potential conservation agriculture practices were not included
365 to avoid double-counting with cover crops, and unresolved questions about long-term
366 efficacy.
- 367 • The trees in annual croplands pathway entails the expansion of three agroforestry
368 practices into annual croplands that currently have low (<10% tree cover). These include
369 expansion of farmer-managed natural regeneration across dry croplands in Africa (150
370 Mha), windbreaks over 50% of non-African croplands (318 Mha), and alley cropping
371 across 22% of non-African croplands (140 Mha). Note that windbreaks and alley
372 cropping were applied to non-African croplands to avoid double counting with farmer-
373 managed natural regeneration. Estimates of soil carbon accumulation derive from a
374 literature review around the soil benefits of windbreaks, or shelterbelts based⁶³⁻⁶⁵ and
375 alley cropping⁶⁶⁻⁶⁸. We estimate that windbreaks capture an additional 0.69 Mg C ha⁻¹ yr⁻¹
376 ¹, whereas alley cropping capture an additional 0.59 Mg C ha⁻¹ yr⁻¹. Because we could not

377 find independent estimates of soil carbon accumulation for farmer managed natural
378 regeneration, we assumed 25% of the mitigation potential was attributable to soil
379 accumulation, averaging together the proportion of mitigation potential for alley cropping
380 and windbreaks. Silvopastoral systems were not included here to avoid double counting
381 with the reforestation pathway.

382 • Avoided grassland conversion refers to avoided soil carbon loss by protecting grasslands
383 from conversion to croplands in areas where grasslands are threatened. For this pathway,
384 we updated the initial NCS analysis of Griscom et al.³² by allowing 28% of soil carbon
385 to be lost down to 1 m in the soil based upon the findings of Sanderman et al³; and the
386 new soil carbon modeling for temperate and tropical grasslands based on ISRIC
387 database³. Thus, we applied this soil carbon loss to the estimated 155 tC ha⁻¹ in temperate
388 grasslands, and 122 tC ha⁻¹ in tropical grasslands over 0.7 Mha and 1.0 Mha respectively
389 for temperate and tropical grasslands converted annually⁶⁹.

390 • Grazing – optimal intensity is a soil only pathway in Griscom et al.³² and remains
391 unchanged in this analysis representing changes in grazing intensity that optimize forage
392 removal and increase soil carbon on both rangeland and planted pasture. We assumed
393 additional sequestration potential of 0.06 MgC ha⁻¹ yr⁻¹ over 712 Mha of land. This
394 includes global rangelands and planted pastures. There is some spatial overlap with
395 Reforestation and Grazing – Legumes, therefore the mitigation potential of this pathway
396 was subtracted from Reforestation mitigation potential to avoid double-counting.
397 Accounting with Grazing – Legumes is additive, so no double-counting concerns.

398 • Grazing – legumes, sowing leguminous crops on planted pastures to increase soil carbon,
399 is a soil only pathway in Griscom et al.³², and remains unchanged in this analysis. The

400 pathway quantifies the net increase in soil carbon (after accounting for increases in N₂O
401 emissions) in planted pastures due to the fertilizing effect of increased nitrogen fixation.
402 We estimate an additional sequestration potential of 0.56 MgC ha⁻¹ yr⁻¹ over 72Mha of
403 land. This was restricted to global planted pastures. Spatial overlap with Reforestation
404 and Grazing – Optimal Intensity. Mitigation potential of this pathway was subtracted
405 from Reforestation mitigation potential to avoid double-counting. Accounting with
406 Grazing – Optimal Intensity is additive, so no double-counting concerns.

- 407 • Peatland restoration includes restoration of global non-tidal freshwater forested and non-
408 forested wetlands. The restoration opportunity across tropical, temperate and boreal
409 peatlands estimated at 46 Mha was not changed³². Avoidable soil carbon losses of 5.44
410 tC ha⁻¹ yr⁻¹ for tropical peatlands, 3.55 tC ha⁻¹ yr⁻¹ for temperate peatlands, and 1.42 tC ha⁻¹
411 yr⁻¹ for boreal peatlands were estimated by assuming an avoided loss of 50% of the
412 original soil carbon⁷⁰⁻⁷² occurring over a 20-year period. Due to the strong likelihood of
413 near-term increased CH₄ emissions arising from increased soil organic carbon in
414 peatlands⁷³, we do not include increased soil carbon sinks in freshwater peatlands upon
415 rewetting for restoration. In other words, we assumed that any possible enhanced carbon
416 sink was at risk of being offset by increased CH₄ emissions³². Recent work shows that
417 this problem may be greater than expected also in coastal wetlands⁷⁴.
- 418 • Avoided peat impacts refers to avoided soil carbon loss by protecting threatened tropical,
419 temperate and boreal peatlands. It includes all threatened non-tidal freshwater forested
420 and non-forested wetlands estimated to cover 0.78 M ha yr⁻¹⁷². Avoidable soil carbon
421 fluxes were estimated to be 217 tC ha⁻¹ for tropical peatlands^{70,72}, 142 tC ha⁻¹ for

422 temperate peatlands^{71,72} and 57 tC ha⁻¹ for boreal peatlands^{71,72}. Forested wetlands were
423 excluded from Avoided Forest Conversion pathway to avoid double-counting.

424 • Restoration of coastal blue carbon ecosystems (mangroves, salt marshes and seagrass
425 meadows) typically leads to significant soil carbon accumulation. Mean literature
426 estimates of carbon sequestration rates during ecosystem restoration were applied to the
427 historic area lost of each of these ecosystems, 11 Mha, 2 Mha and 17 Mha respectively
428 for mangrove, salt marsh and seagrass and was not changed from Griscom et al.³². Here
429 both avoided losses of soil carbon and enhanced sequestration are included, and were
430 estimated based on addition sequestration at an average rate of 1.7 tC ha⁻¹ yr⁻¹^{75,76}, and
431 avoided fluxes averaging 3.4 tC ha⁻¹ yr⁻¹ estimated by assuming a potential 50% loss of
432 the original soil carbon^{77,78} occurring over a 20-year period.

433 • Avoided coastal impacts refer to the avoided soil carbon emissions by protecting
434 threatened blue carbon ecosystems (mangroves, salt marshes and seagrass meadows).
435 This pathway was updated from Griscom et al.³² by using more recent lower estimates of
436 ongoing mangrove loss rates^{79,80}. The soil portion was calculated based on estimates of
437 soil carbon stocks to 1m and expected losses resulting in avoidable fluxes of 197.47 tC
438 ha⁻¹, 133.78 tC ha⁻¹ and 77.43 tC ha⁻¹ respectively over 0.05 Mha yr⁻¹ of mangroves, 0.08
439 Mha yr⁻¹ of salt marshes and 0.45 Mha yr⁻¹ of seagrass meadows^{77,78,80}. Mangroves were
440 excluded from Avoided Forest Conversion pathway to avoid double-counting.

441

442 **Uncertainty estimates**

443 Uncertainty for maximum mitigation estimates of each pathway can be found in Griscom
444 et al.³² In brief, methods consistent with IPCC good practice guidance were used when empirical

445 uncertainty estimation was possible. For other pathways, the Delphi method of expert elicitation
446 involving two rounds of explicit questions about expert opinion on the potential extent and
447 intensity of flux were combined.

448

449 **Data Deposition**

450 A global spatial dataset of reforestation opportunities is available on Zenodo
451 (<https://zenodo.org/record/883444>). Figures 1 and 2 have associated raw data that can be made
452 available upon request.

453

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666

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670 analysis; DB, BG, SC, PW, JF, JS, PS, SW, RZ, MvU, and IE interpreted the data and wrote the
671 paper.

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681 **Figure Legends**

682 Figure 1. Additional soil carbon storage potential for 12 natural pathways to climate mitigation.
683 We estimate annual maximum climate mitigation potential with safeguards for the reference year
684 2030. Light gray portions of bars represent cost-effective mitigation levels assuming a global
685 ambition to hold warming below 2°C (<100 USD MgCO₂e⁻¹ y⁻¹). Dark grey portions of bars
686 indicate low cost (<10 USD MgCO₂e⁻¹ y⁻¹) portions. Ecosystem service benefits linked with each
687 pathway are indicated by colored bars for biodiversity, water (filtration and flood control), food
688 and air filtration. Most pathways also contribute biomass carbon, (see Figure 2), with the
689 exception of pathways that are entirely soil carbon: biochar, cover cropping, both grazing
690 options, and avoided grassland conversion. More than half of the pathways (reforestation, cover
691 cropping, biochar, trees in croplands, grazing, improved pasture options and coastal wetland
692 restoration), represent enhanced soil carbon sinks, while the others are avoided soil carbon
693 losses. The remaining 8 of the 20 pathways from Griscom et al. ³² are not expected to have an
694 impact on soil organic carbon, and therefore have not been included in this figure.

695

696 Figure 2. Maximum climate mitigation potential of soil in 2030 across forest, agriculture and
697 grassland, and wetland biome pathways with safeguards. Bars to the left indicate the magnitude
698 of potential sinks, whereas the bars to the right indicate magnitude of avoided emissions. Dark
699 portions of bars represent soil carbon; white portions of bars represent vegetative biomass; and
700 dotted portion of bar is avoided CH₄ and N₂O through improved nutrient, rice, and animal
701 management. Note that due to the strong likelihood of near-term increased CH₄ emissions arising
702 from increased soil organic carbon in peatlands⁷³, we do not included increased soil carbon sinks
703 in freshwater peatlands upon rewetting for restoration.

704 **Tables**

705 Table 1. Summary of soil carbon elements of natural climate solutions (NCS): the role of soil
 706 and co-benefits for sustainable development. Cells in green indicate Forest pathways, in yellow
 707 indicate grassland/agricultural pathways, and in blue indicate wetland pathways. Adapted from
 708 Table S2 and S5 in Griscom et al. ³².

NCS Pathway	Contribution of soil carbon	Co-benefits for Sustainable development
Avoided Forest Conversion	1.2 Gt CO ₂ e yr ⁻¹ for soil protection and carbon sequestration is about 9% of the mitigation benefit from these two forest pathways.	Water retention and flow regulation. Biodiversity benefits. Maintains soil biological and physical properties ensuring health and productivity of forests.
Reforestation		Measured increase in soil fauna in reforested sites. Drought resilience. Water retention and flow regulation.
Biochar	1.1 Gt CO ₂ e yr ⁻¹ biochar direct mitigation potential.	Soil quality and fertility enhancement in temperate regions.
Cover cropping	0.41 Gt CO ₂ e yr ⁻¹ is entirely soil carbon.	Soil quality and fertility enhancement. Reduced agricultural water demands with appropriate cover crops. Reduced soil erosion and redistribution maintaining soil depth and water retention.
Trees in Croplands	0.28 Gt CO ₂ e yr ⁻¹ in soil carbon is 40% of the total mitigation potential.	Biodiversity, habitat connectivity, erosion control, water recharge, and reduced soil erosion. Tree planting helps capture airborne particles and pollutant gasses.
Avoided Grassland Conversion	0.23 Gt CO ₂ e yr ⁻¹ is entirely soil carbon.	Permanent grasslands provide "biological flood control" and maintain ecosystem water balance assuring adequate water resources. Important habitat for nesting and foraging birds.
Grazing - Optimal Intensity	0.15 Gt CO ₂ e yr ⁻¹ is entirely soil carbon.	Reduces disturbance to plant-insect interactions. Reduce water use on managed pastures, increase the soils ability to trap contaminants.
Grazing - Legumes in Pastures	0.15 Gt CO ₂ e yr ⁻¹ is entirely soil carbon.	Higher insect diversity, biological nitrogen fixation, improved soil structure, erosion protection and greater biological diversity.
Peatland Restoration	0.65 Gt CO ₂ e yr ⁻¹ in soil carbon is 80% of the total mitigation potential.	Restoring reestablishes diverse communities and increases faunal species that help develop soil structure and fertility. Waste water treatment and storm water remediation. Flood attenuation. Reduced fire risk lessening exposure to pollutants associated with lung and pulmonary disorders.
Avoided Peatland Impacts	0.54 Gt CO ₂ e yr ⁻¹ in soil carbon is 72% of the total mitigation potential.	
Coastal Wetland Restoration	0.52 Gt CO ₂ e yr ⁻¹ in soil carbon is 62% of the total mitigation potential.	Maintains the provision of structure, nutrients and primary productivity and nurseries for commercially important fish and shrimp. High economic value for water treatment. Benefits of cross-system nutrient transfer to coral reefs, coastal protection, and water quality regulation.
Avoided Coastal Wetland Impacts	0.24 Gt CO ₂ e yr ⁻¹ in soil carbon is 79% of the total mitigation potential.	

710 Table 2. Example activities to achieve mitigation potentials of soil carbon sequestration
 711 pathways. Cells in green indicate Forest pathways, in yellow indicate grassland/agricultural
 712 pathways, and in blue indicate wetland pathways. Adapted from Table S7 in Griscom et al.³² and
 713 Griscom et al.⁸¹.

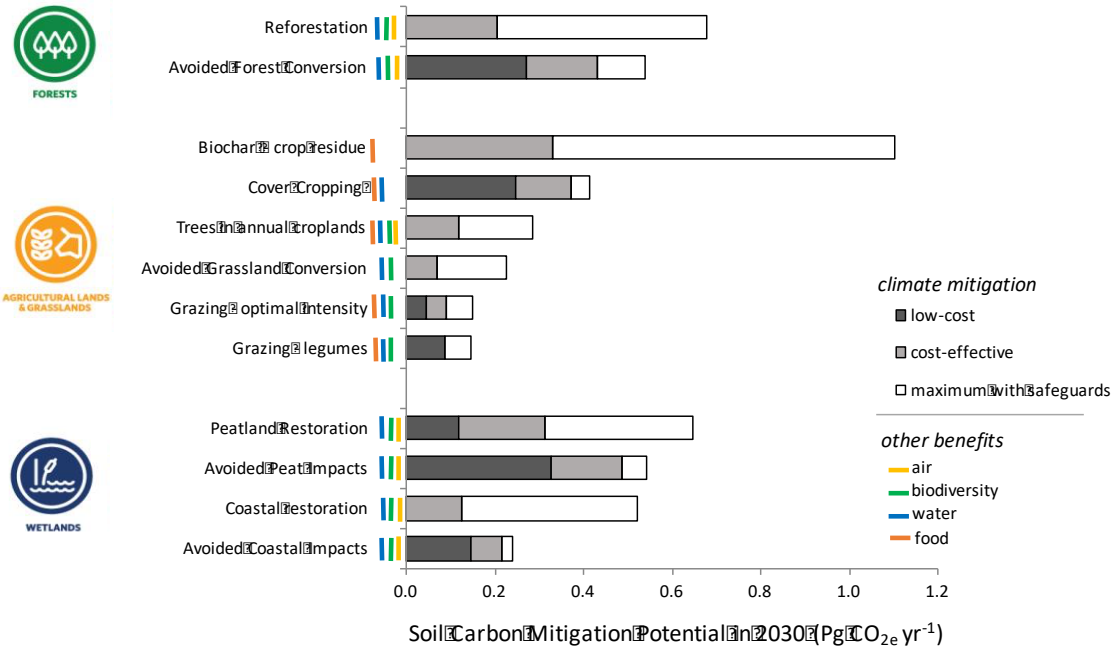
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NCS Pathway	General Activities	Specific Activities
Avoided Forest Conversion	PROTECTION Protected areas establishment and improved enforcement, improved land tenure, indigenous community management	Improved citing of non-forest land use; forest certification; zero deforestation commitments; sustainable intensification of agriculture; diet shifts; avoided loss of high carbon forests.
Avoided Grassland Conversion		Prevent conversion of grasslands to tilled croplands; intensification of existing croplands.
Avoided Peatland Impacts		No-net-loss mitigation regulations; re-siting of oil palm plantation permits to non-peat locations.
Avoided Coastal Wetland Impacts		No-net-loss mitigation regulations; avoided harvest of mangroves for charcoal; avoided consumption of food products with acute impacts on coastal wetlands (e.g. mangrove replacing shrimp farms).
Biochar	MANAGEMENT Realignment of agriculture support programs, ecosystem services payments, certification schemes, improved land tenure, mitigation programs and markets	Extension programs to build capacity on biochar management.
Cover cropping		Cultivation of additional cover crops in fallow periods; shift to reduced-tillage or zero-tillage systems and other conservation agriculture practices may enhance soil carbon benefits of cover crops.
Trees in Croplands		Regulations and certification programs that promote windbreaks (shelterbelts), alley cropping, agroforestry systems and farmer managed natural regeneration (FMNR).
Grazing - Optimal Intensity		Maintaining forage consumption rates that enable maximum forage production.
Grazing - Legumes in Pastures		Sowing legumes in existing planted pastures.
Reforestation	RESTORATION Certification and mitigation programs, indigenous community management	Regulations that advance minimum forest cover requirements; integration of trees into grazing lands (i.e. silvopastoral systems); diet shifts.
Peatland Restoration		Re-wetting and re-planting with native freshwater wetlands species.
Coastal Wetland Restoration		Re-wetting and re-planting with native salt-water wetlands.

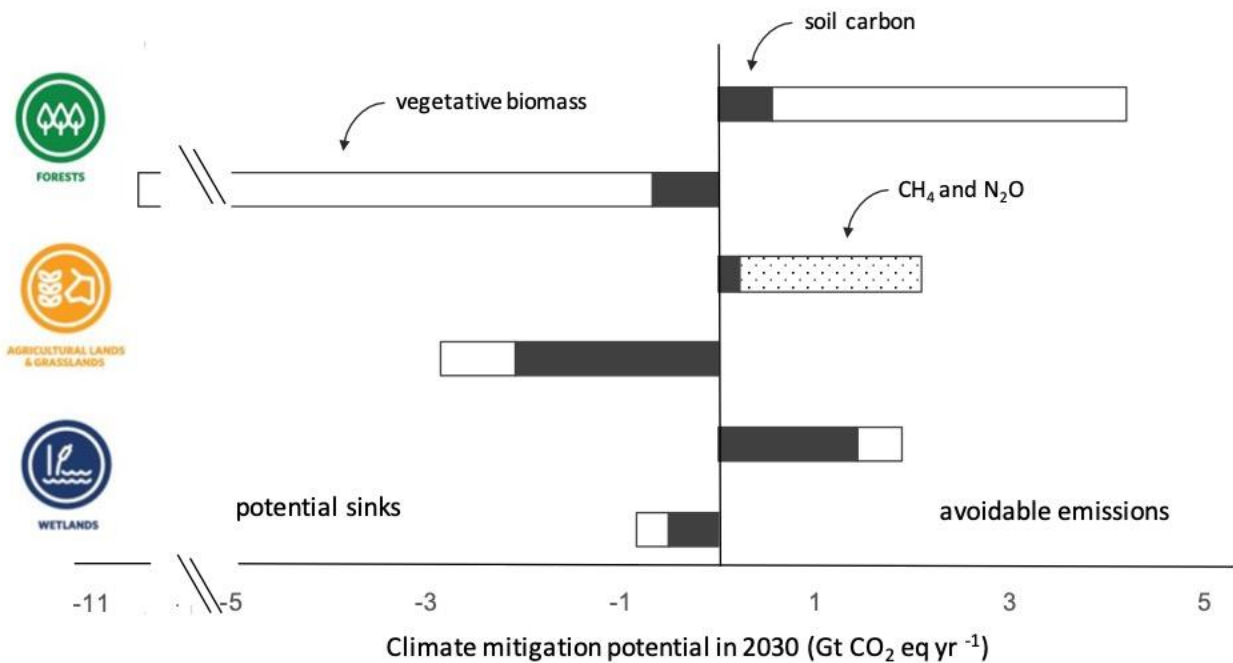
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717 Figure 1
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719 Figure 2
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