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1 Abstract

2 The biotransformation and bioconcentration of natural and synthetic steroid estrogens by
3 *Chlorella vulgaris* were investigated using batch shaking experiments with incubation for
4 48 hours in the light or dark. Estradiol and estrone were inter-convertible in both light
5 and dark conditions, however this biotransformation showed a preference to estrone. In
6 the light, 50% of estradiol was further metabolized to an unknown product. Apart from
7 biotransformation, estrone as well as hydroxyestrone, estriol and ethinylestradiol were
8 relatively stable in the algal culture, while estradiol valerate was hydrolyzed to estradiol
9 and then estrone within 3 hours of incubation. All the tested estrogens exhibited a degree
10 of partitioning to *C. vulgaris*, however, the concentrations of estriol, hydroxyestrone,
11 ethinylestradiol and estradiol valerate were always below the quantification limits. For
12 estradiol and estrone, the partitioning of these estrogens in the algal extracts to the
13 filtrates was below 6% of the total present. The average concentration factor for estrone
14 was around 27, however the concentration factor for estradiol is not reported since no
15 equilibrium was reached between aqueous solution and that within the cells due to
16 continuing biotransformation.

1 Introduction

2 Natural phenolic steroid estrogens (estradiol and estrone) and the synthetic
3 estrogen, ethinylestradiol, have been detected in aquatic systems and their occurrence is
4 associated with vitellogenesis and reproductive abnormalities in wild fish (3,5,7,16,21). A
5 more complete understanding of their behavior in the aquatic environment is required to
6 enable accurate risk assessments to be undertaken (9). Algae, given their substantial
7 biomass, extensive range of habitat and diversity play an important role in the fate of
8 organic compounds in the aquatic ecosystem (14,15). They may degrade or take up
9 contaminants thereby acting as a medium for bioconcentration and subsequent
10 biomagnification in higher trophic levels (14,15,20). Biosorption of heavy metals and
11 organic pollutants by algae has been frequently reported (12,14) and some macroalgae
12 have been shown to have the same detoxification enzymes as those found in the
13 mammalian liver, the so called “green liver” (15). It has been demonstrated that
14 *Ochromonas danica* has the ability to break down phenolic compounds to pyruvate and
15 carbon dioxide (19).

16 Ethinylestradiol and estradiol valerate are synthetic estrogens produced by
17 modifying the structure of estradiol. Ethinylestradiol can be deethinylated to estradiol in
18 the human liver, though a majority (up to 80%) of ethinylestradiol is excreted from
19 humans unchanged in conjugated forms (4), while estradiol valerate is readily hydrolysed
20 to estradiol in the liver to exhibit its estrogenic functions (18). In the environment, the
21 persistence of ethinylestradiol has been reported in sewage treatment works and rivers
22 (7,13,22). With respect to the natural estrogens, estradiol and estrone, they are readily
23 transformed via oxidative and reductive pathways by 17 β -hydroxysteroid dehydrogenase

1 (17 β -HSDH) in a range of organisms (4). Microorganisms that do not produce estrogens,
2 such as fungi and bacteria, also exhibit the ability to facilitate this biotransformation
3 through such pathways when exposed to estradiol and estrone (4,11,13,23,27). These
4 steroid estrogens may also be further metabolized to estriol and hydroxyestrone, which
5 play an important role in a range of organisms (4).

6 The aim of this study was to investigate the interaction of a common and widely
7 studied fresh water alga, *Chlorella vulgaris*, with different natural (estradiol, estrone,
8 hydroxyestrone and estriol) and synthetic (ethinylestradiol and estradiol valerate)
9 estrogens and establish whether they are biotransformed or bioconcentrated.

10

11 Materials and Methods

12 Chemicals

13 Individual working solutions of estradiol, estrone, estriol, hydroxyestrone,
14 estradiol valerate, estradiol acetate and ethinylestradiol (Sigma, Poole, UK) were
15 prepared in HPLC grade acetone (Rathburn, Walkerburn, UK) at 10 mg l⁻¹. *Chlorella*
16 *vulgaris* was purchased from the Culture Collection of Algae and Protozoa (CCAP),
17 Windermere, UK. The alga was grown in Jaworski's medium (JM), as recommended by
18 CCAP, with continuous aeration and illumination (cool white fluorescent light at 300 lux
19 was provided by Osram L58 W23 tubes) in a temperature controlled room set at 18 \pm 2°C.
20 The algal cultures were allowed to grow into the stationary phase (20 days) achieving
21 approximately 2 g l⁻¹ (dry weight). Stationary phase cultures maximised the amount of
22 biomass available to allow for determination of trace levels of steroid estrogens sorbed to
23 the solid phase.

1

2 Biotransformation and bioconcentration experiments

3 All the batch experiments were undertaken in triplicate. Individual estrogens
4 (500ng) (estradiol, estrone, estriol, hydroxyestrone, estradiol valerate and
5 ethinylestradiol) were spiked separately into 250 ml Teflon bottles (dark) or 500 ml
6 conical flasks (light) and blown to dryness with a gentle nitrogen stream. Algal culture
7 (200 ml) was transferred into the bottles or flasks so that the initial concentration was of
8 $2.5 \mu\text{g l}^{-1}$ and placed on a rotary shaker in the dark or aerated in the light with triplicate
9 samples taken at time intervals of 3, 6, 24 and 48 hours. The optical density of the algal
10 culture was measured at 680 nm and converted into mass of algae (dry weight) for each
11 sampling time. Further kinetic experiments on the biotransformation of estradiol were
12 undertaken using different algal cell densities (0.8, 1.5 and 2.3 g l^{-1} , dry weight) shaking
13 for 1, 18 and 48 hours and a range of estradiol concentrations per gram of algae (9.7, 97,
14 388 and $969 \mu\text{l l}^{-1} \text{ g}^{-1}$) in the dark for 6 hours.

15

16 Estrogen determination in filtrate and algal cells

17 Algal cells were filtered from the growth culture using a glass fibre filter (GF6)
18 (Scheicher & Schuell, London, UK). Before extraction, 500 ng of estradiol acetate was
19 spiked into the separated algal cells and filtrate for the determination of the final recovery
20 in each sample. Estrogens in the algal cells were extracted overnight by soxhlet extraction
21 using 150 ml dichloromethane (DCM) (Rathburn, Walkerburn, UK). Estrogens in the
22 liquid phase were extracted by liquid-liquid extraction using 200 ml DCM. Extracted
23 estrogens were evaporated to 1 ml using a rotary evaporator, transferred into a reaction
24 vial and dried with a gentle nitrogen stream. The derivatization mixture (*N*-methyl-*N*-

1 (trimethylsilyl)-tri-fluoroacetamide: trimethylsilylimidazole: dithioerythritol; 1000:2:2;
2 v/v/w) (Sigma, Poole, UK), 50 µl, was added, the reaction vial sealed, and placed in a
3 heating block at 60 °C for 30 min. The solution was again evaporated to dryness and 0.2
4 ml of 2 mg l⁻¹ mirex (internal standard) (Promochem, Welwyn, UK) in hexane was then
5 added before gas chromatography-mass spectrometry (GCMS) analysis. The GCMS
6 conditions, calibration, identification and quantification procedures have been described
7 previously (8). The GCMS limit of quantification was 10 µg l⁻¹ in final extracts, however,
8 some samples exhibited responses of target ions at correct retention times below this level
9 and such results have been reported as being less than the quantification level, but above
10 the detection limit of 5 µg l⁻¹.

11

12 Quality Control

13 Before the experiments, quality controls were undertaken to ensure that the
14 estrogens did not sorb to the laboratory equipment. The recovery efficiency of the soxhlet
15 and liquid-liquid extraction for the algae and the filtrate spiked with 500 ng each of
16 estrogen mixture was evaluated with respect to the recovery of estradiol acetate. The
17 recovery for soxhlet extraction was 65% and for liquid-liquid extraction was above 85%
18 for all the estrogens, except for estriol where recoveries were 50% in both extraction
19 methods. Moreover, there was no estrogen peak detected from the blanks of the algae and
20 filtrate extracts.

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1 Results

2 Fate of estrogens in algal culture

3 The concentration of estradiol in the culture medium fell rapidly over the first 3
4 hours of incubation, and this decline was more pronounced in the light than the dark
5 (Figure 1). However, subsequent change in the estradiol concentration was more rapid in
6 the dark, and was accompanied by an increase in the concentration of estrone in the
7 media solution. The stability of all estrogens was tested using crude and autoclaved algal
8 filtrate (without algal cells) instead of the algal culture as controls. Experimental
9 conditions were the same for all sample types. The estrogen concentrations in the crude
10 and autoclaved (sterile) filtrates remained constant, which indicated that no biological,
11 physical or chemical removal occurred in the algal-cell free or the sterile filtrates. Lack of
12 change in the crude filtrates which may have contained bacteria not removed by the GF6
13 filter, would indicate that effects observed were due to the algae rather than other
14 microorganisms. The most probable explanation for the source of estrone is that it is a
15 result of biotransformation of estradiol by the algae, *C. vulgaris*. This is also supported by
16 considering the mass balance of estradiol and estrone in the algal culture, which in the
17 dark decreased by less than 20% in total (Table 1). However, in the light there was an
18 overall decrease in concentration of both compounds, probably as a result of further
19 metabolism (Table 1). These differences in final concentrations of estradiol and estrone in
20 light and dark conditions suggest that biochemical processes in the algae have a different
21 effect on the estrogens in light and dark conditions.

22 Although further transformation products, hydroxyestrone and estriol were
23 observed in media extracts, the concentrations were below quantification levels and these

1 are therefore insignificant products or metabolic intermediates with short half-lives. It
2 was also possible that the estrogens were removed from the media solution through
3 bioconcentration within the algae. The amount of estrogens which partitioned to the algae
4 did not exceed more than 9% of the total estrogens by mass (Table 1). Although the
5 ultimate fate of estradiol is not clear, it is apparent that in the light, *C. vulgaris* reduces
6 the concentration from $2.50 \mu\text{g l}^{-1}$ to $0.37 \mu\text{g l}^{-1}$, although some $0.60 \mu\text{g l}^{-1}$ of the estradiol
7 is converted to estrone, which is a potent steroid estrogen.

8 Media solutions were also spiked with estrone and again incubated in light and
9 dark conditions. In this case, estradiol was observed in the solutions, which would
10 indicate that the alga is able to interconvert estrone and estradiol. However, the amount of
11 estrone transformed was less than 10% in comparison to the reverse reaction where over
12 70% and 22% of estradiol was transformed to estrone in the dark and light, respectively
13 (Figure 2). Both estrone and hydroxyestrone were again observed in solution, yet still
14 below the quantification limit. It was however, clear that mass balance equations (Table
15 2) indicate that overall reduction in concentration of estrone and estradiol was less than
16 when solutions were spiked with estradiol. With respect to the partitioning of estrogens to
17 the algae, the maximum percentage partitioning was 7%, with the majority of compounds
18 remaining in solution (Table 2).

19 The algae also exhibited an ability to hydrolyze estradiol valerate to estradiol and
20 estrone in both light and dark (Figure 3). The concentration of estradiol valerate had
21 decreased after the first 3 hours of incubation to below the quantification level and this
22 drop was associated with the occurrence of both estradiol and estrone. The concentration
23 of estradiol remained constant throughout the experimental period in both dark and light,

1 however, the level of estrone increased continuously in the dark, but peaked at 24 hours
2 in the light. Traces of both estriol and hydroxyestrone were also detected in the filtrate
3 but their concentrations were below the quantification limits. This data may suggest that
4 estradiol valerate was rapidly transformed to an intermediate before a slower hydrolysis
5 to estradiol. Some uptake of estradiol valerate by the algae was observed, however,
6 concentrations were below the limit of quantification.

7 For the samples spiked with estriol and ethinylestradiol, a slight drop in
8 concentrations of both compounds were observed however the data points were not
9 significantly different and thus evidence of removal of these compounds was not
10 conclusive. Similarly, it was not possible to quantify any uptake of these estrogens by the
11 algae since concentrations were generally below the quantification limits and in some
12 cases below the detection levels.

13 The fate of hydroxyestrone in the algal culture was also tested and the result was
14 compared to that in a culture filtrate (no algae present) as a control. A reduction in
15 concentration of hydroxyestrone (<10% of initial spiked) was recorded in the culture with
16 a detectable amount of estriol, while the concentration of hydroxyestrone was unchanged
17 in the control. The relative stability of hydroxyestrone would indicate that it is not a
18 rapidly metabolized intermediate in any transformation pathways.

19 Using the preceding observations, a possible pathway for transformation of the
20 steroid estrogens by *C. vulgaris* is shown in Figure 4. Ethinylestradiol was persistent in
21 the algal culture, while estradiol valerate was hydrolyzed to estradiol and estrone. Due to
22 the interconversion preference for estrone, it is evident that estradiol valerate was
23 hydrolyzed to estradiol and then transformed to estrone. In light conditions,

1 approximately 50% of estradiol was metabolized into an unknown product and some of
2 the estradiol was transformed to estrone, while in dark conditions, biotransformation to
3 estrone was the major pathway. Although estrone, hydroxyestrone and estriol were
4 detected in the algal culture under light conditions, the low concentration and the stability
5 of these estrogens in the culture indicated that as a major metabolic pathway it is
6 insignificant compared to the unknown pathway observed to occur in the light.

7

8 Bioconcentration of estrogens in the algae

9 On exposure of *C. vulgaris* to estradiol and estrone, a degree of partitioning of
10 both estrogens was observed to occur between the algae and aqueous phase. However, a
11 true “bioconcentration factor” cannot be reported as equilibrium was not been reached,
12 and transformation continued. For the samples spiked with estradiol, the concentration of
13 estradiol decreased significantly in the filtrate with time, and the occurrence of estrone in
14 the algal extract was most probably a result of intracellular biotransformation (rather than
15 uptake from the filtrate), thus no bioconcentration factor can be reported (Table 1). For
16 the samples spiked with estrone, although biotransformation did occur, the level of
17 estrone in the filtrate remained relatively constant and the concentration of estradiol was
18 also consistently low. It was therefore likely that estrone extracted from the algae
19 originated from the filtrate solution and thus a concentration factor for estrone between
20 the algae and filtrate was calculated to give an insight into the probable bioconcentration
21 factor for estrone. The average concentration factors for estrone in both light and dark
22 conditions was approximately 27 (Table 2).

23

1 Effect of cell density on transformation of estradiol

2 The kinetics of the biotransformation process from estradiol to estrone were
3 determined over a range of cell densities in the dark only as in light the end products of
4 transformation were unknown. The rate of biotransformation of estradiol to estrone
5 declined with an increase in cell density and shaking time (Figure 5), which indicated that
6 substrate levels were limiting under these experimental conditions.

7

8 Effect of substrate level on the rate of transformation

9 In order to investigate the effect of substrate level on biotransformation, different
10 concentrations of estradiol were spiked into the culture. The final concentration of estrone
11 produced increased with initial estradiol concentration up to $400 \mu\text{g l}^{-1} \text{g}^{-1}$ of algae,
12 however at $969 \mu\text{g l}^{-1} \text{g}^{-1}$ there was no significant increase in the final estrone
13 concentration. The regression line produced using the first 3 data points indicated that
14 within experimental error, the concentration of estradiol became limiting below $400 \mu\text{g l}^{-1}$
15 g^{-1} (Figure 6) which is equivalent to $2.58 \mu\text{g}$ of estrone produced per gram of algae per
16 hour and is likely to indicate the maximum capacity of *C. vulgaris* to transform estradiol
17 to estrone under experimental conditions.

18

19 Discussion

20 In the present study, the hydrolysis of estradiol valerate to estradiol by *C. vulgaris*
21 has been demonstrated, however, under the conditions used in this study, transformation
22 of ethinylestradiol, through deethinylation or other pathways, was not observed. With
23 respect to estradiol and estrone, *C. vulgaris* exhibited a preference toward estrone in both

1 light and dark. This biotransformation may be significant to the alga by reducing the
2 estrogenic potency and increasing the polarity of estrogens for excretion or further
3 metabolism (4,25). The occurrence of estriol and hydroxyestrone is an indicator of further
4 hydroxylation of estrone to more polar metabolites (4,25). However, such a pathway
5 accounts for a small proportion, < 4% of total loss of estrone/estradiol in the light and
6 another, unidentified pathway must therefore be responsible for 50% loss of the
7 estrogens. It is possible that conjugation of the estradiol may occur in the light.
8 Conjugation increases polarity and is the method by which the steroid estrogens are
9 excreted in urine in a range of animal species (4). However, this process also occurs in
10 plants and partially or completely deactivates plant hormones and phytoestrogens (2).
11 This process has also been observed to occur in plants to detoxify xenobiotics such as
12 DDT and chlorinated phenols (17,24). It is evident that conjugation is a rapid
13 detoxification mechanism and an absence of a lag phase for conjugation has been
14 demonstrated for plant hormones (2). Such transformation would explain the rapid
15 decline in estradiol concentration after the first 3 hours of shaking. Moreover, in the same
16 review (2) the effect of light on the stimulation and regulation of conjugation has been
17 exemplified in different plant hormones.

18 Estradiol/estrone biotransformation by the algae was observed to be dependent on
19 cell density and substrate concentration. In this study a relatively high concentration of
20 estrogens in comparison to environmental levels, of around $0.1 - 2 \text{ ng l}^{-1}$ (9), and algal
21 cells, were used to allow for monitoring of changes and uptake. The results of the kinetic
22 study demonstrated that the biotransformation process operated under conditions of
23 substrate limitation, however if the same pattern of substrate limitation occurs in the

1 natural environment, it would be dependent on the ratio of estrogen concentrations
2 present, the amount of algae and a range of other factors not included in this study.
3 However, our results suggest that, it is likely that biotransformation would reach a
4 saturation when the concentration of estradiol in aqueous solution per gram of algae is
5 greater than $400 \mu\text{g l}^{-1} \text{g}^{-1}$. This kinetic study may be used in estimating the probable rate
6 of biotransformation of estrogens by algae in different environmental conditions.

7 The log K_{ow} values for the steroid estrogens indicate that ethinylestradiol should
8 partition to the organic cellular material to a greater extent than other estrogens studied,
9 assuming that simple diffusion controls the partitioning process (10,23). However, this
10 was not the case and the apparently variable uptake of the compounds suggests that an
11 active transport (or exclusion) or a selective binding mechanism may be operating. Since
12 it is generally agreed that estrogens pass across membranes via a simple diffusion
13 pathway (25), it is likely that bioconcentration of estrogens in the algae is due to an active
14 binding mechanism such as binding to receptors or enzymes for biotransformation.

15 It has been suggested that accumulation of persistent organic compounds in
16 phytoplankton is the first step of biomagnification in the food chain (12,20). However,
17 the concentration factors of estrogens in the algae are significantly lower than reported
18 for other endocrine disrupting substances such as DDT (3×10^4) (10). It is likely that
19 biomagnification of estrone is of less significant than for other endocrine disrupting
20 substances in ecosystems.

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1 References

- 2 1. **Axelman, J., D. Broman, and C. Naf.** 1997. Field measurements of PCB
3 partitioning between water and plankton organisms: Influence of growth, particle
4 size, and solute-solvent interactions. *Environ. Sci. Technol.* **31**:665-669.
- 5 2. **Bearder, J. R.** 1980. Biosynthesis and metabolism of plant hormones. *In* J.
6 MacMillan (ed.), *Hormonal regulation and development: molecular aspects of plants*
7 *hormones*, Springer, Berlin.
- 8 3. **Belfroid, A. C., A. Van der Horst, A. D. Vethaak, A. J. Schafer, G. B. J. Rijs, J.**
9 **Wegener, and W. P. Cofino.** 1999. Analysis and occurrence of estrogenic hormones
10 and their glucuronides in surface water and waste water in The Netherlands. *Sci.*
11 *Total Environ.* **225**:101-108.
- 12 4. **Bolt, H. M.** 1979. Metabolism of estrogens-Natural and synthetic. *Pharmac. Ther.* **4**:
13 155-181.
- 14 5. **Desbrow, C., E. J. Routledge, G. C. Brighty, J. P. Sumpter, and M. Waldock.**
15 1998. Identification of estrogenic chemicals in STW effluent. 1. Chemical
16 fractionation and in vitro biological screening. *Environ. Sci. Technol.* **32**:1549-1558.
- 17 6. **Itagaki, E., and T. Iwaya.** 1988. Purification and characterization of 17 β -
18 Hydroxysteroid dehydrogenase from *Cylindrocarpon radicolola*. *J. Biochem.*
19 **103**:1039-1044.
- 20 7. **Jurgens, M. D., A. C. Johnson, and R. J. Williams.** 1999. Fate and behavior of
21 steroid oestrogens in rivers: A scoping study. R&D Technical Report No P161,
22 Environment Agency, UK.

- 1 8. **Lai, K. M., K. L. Johnson, M. D. Scrimshaw, and J. N. Lester.** 2000. Binding of
2 waterborne steroid estrogens to solid phases in river and estuarine systems. *Environ.*
3 *Sci. Technol.* **34**:3890-3894.
- 4 9. **Lai, K. M., M. D. Scrimshaw, and J. N. Lester.** The effects of natural and synthetic
5 steroid estrogens in relation to their environmental occurrence. *Crit. Rev. Toxicol.* (In
6 press)
- 7 10. **Lai, K. M., M. D. Scrimshaw, and J. N. Lester.** Prediction of the bioaccumulation
8 factors and body burden of natural and synthetic estrogens in aquatic organisms in the
9 river systems. *Sci. Total Environ.* (In press).
- 10 11. **Lanisnik, T., M. Zakelj-Mavric, and I. Belic.** 1992. Fungal 17 β -hydroxysteroid
11 dehydrogenase. *FEMS Microbiol. Letts.* **99**:49-52.
- 12 12. **Larsson, P., A. Andersson, D. Broman, J. Nordback, and E. Lundberg.** 2000.
13 Persistent organic pollutants (POPs) in pelagic systems. *Ambio* **29**:202-209.
- 14 13. **Layton, A. C., B. W. Gregory, J. R. Seward, T. W. Schultz, and G. S. Sayler.**
15 2000. Mineralization of steroidal hormones by biosolids in wastewater treatment
16 systems in Tennessee U.S.A. *Environ. Sci. Technol.* **34**:3925-3931.
- 17 14. **Mason, R. P., J. R. Reinfelder, and F. M. M. Morel.** 1996. Uptake, toxicity, and
18 trophic transfer of mercury in a coastal diatom. *Environ. Sci. Technol.* **30**:1835-1845.
- 19 15. **Pflugmacher, S., C. Wiencke, and H. Sandermann.** 1999. Activity of phase I and
20 phase II detoxication enzymes in Antarctic and Arctic macroalgae. *Marine Environ.*
21 *Res.* **48**:23-36.

- 1 16. **Routledge, E. J., D. Sheahan, C. Desbrow, G. C. Brighty, M. Waldock, and J. P.**
2 **Sumpter.** 1998. Identification of estrogenic chemicals in STW effluent. 2. In vivo
3 responses in trout and roach. *Environ. Sci. Technol.* **32**:1559-1565.
- 4 17. **Sandermann, H., R. Schmitt, H. Eckey, and T. Bauknecht.** 1991. Plant
5 biochemistry of xenobiotics: isolation and properties of soybean O- and N-glucosyl
6 and O- and N-malonyltransferases for chlorinated phenols and anilines. *Arch.*
7 *Biochem. Biophys.* **287**:341-350.
- 8 18. **Seibert, B.** 1996. Data from animal experiments and epidemiological data on
9 tumorigenicity of estradiol valerate and ethinylestradiol. UBA TEXTE 3/96:63-68,
10 Berlin.
- 11 19. **Semple, K. T., R. B. Cain, and S. Schmidt.** 1999. Biodegradation of aromatic
12 compounds by microalgae. *FEMS Microbiol. Letts.* **170**:291-300.
- 13 20. **Sijm, D. T. H. M., K. W. Broersen, D. F. de Roode, and P. Mayer.** 1998.
14 Bioconcentration kinetics of hydrophobic chemicals in different densities of *Chlorella*
15 *pyrenodosa*. *Environ. Toxicol. Chem.* **17**:1695-1704.
- 16 21. **Ternes, T. A., J. Mueller, M. Stumpf, K. Haberer, R. D. Wilken, and M. Servos.**
17 1999. Behaviour and occurrence of estrogens in municipal sewage treatment plants-I.
18 Investigations in Germany, Canada and Brazil. *Sci. Total Environ.* **225**:81-90.
- 19 22. **Ternes, T. A., P. Keckel, and J. Mueller.** 1999. Behaviour and occurrence of
20 estrogens in municipal sewage treatment plants-II. Aerobic batch experiments with
21 activated sludge. *Sci. Total Environ.* **225**:91-99.
- 22 23. **Wang, X., Y. Ma, W. Yu, and H. J. Geyer.** 1997. Two-compartment
23 thermodynamic model for bioconcentration of hydrophobic organic chemicals by

- 1 alga: Quantitative relationship between bioconcentration factor and surface area of
2 marine algae or octanol/water partition coefficient. *Chemosphere* **35**:1781-1797.
- 3 24. **Wetzel, A., and H. Sandermann.** 1994. Plant biochemistry of xenobiotics: isolation
4 and characterization of a soybean O-glucosyltransferase of DDT metabolism. *Arch.*
5 *Biochem. Biophys.* **314**:323-328.
- 6 25. **Williams, C., and G. M. Stancel.** 1996. Estrogens and progestins. *In* J. G. Hardman,
7 L. E. Limbird, P. B. Molinoff, and R. W. Ruddon (ed.), *The pharmacological basis of*
8 *therapeutics*, 9th ed. McGraw-Hill, New York.
- 9 26. **Zakelj-Mavric, M., T. Kastelic-Suhadolc, A. Plemenitas, T. L. Rizner, and I.**
10 **Belic.** 1995. Steroid hormone signalling system and fungi. *Comp. Biochem. Physiol.*
11 **112B**:637-642.
- 12 27. **Zeillinger, R., and J. Spona.** 1986. *Pseudomonas testosteroni*: new data about
13 growth and steroid metabolism. *FEMS Microbiol. Letts.* **37**:231-235.
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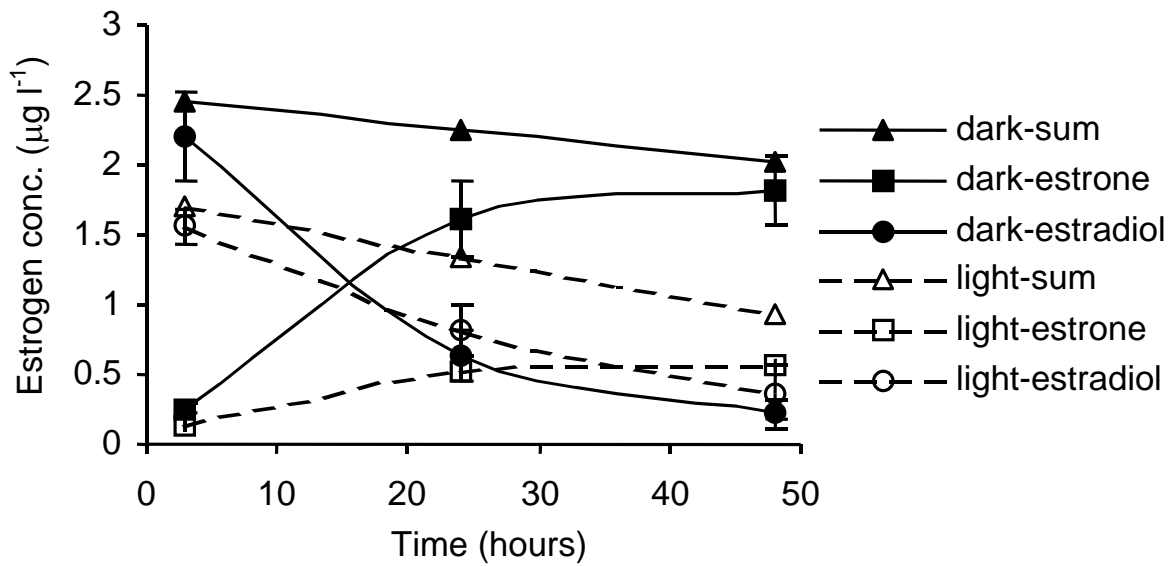


Figure 1. Changes with time (in light and dark conditions) observed after spiking the culture with estradiol

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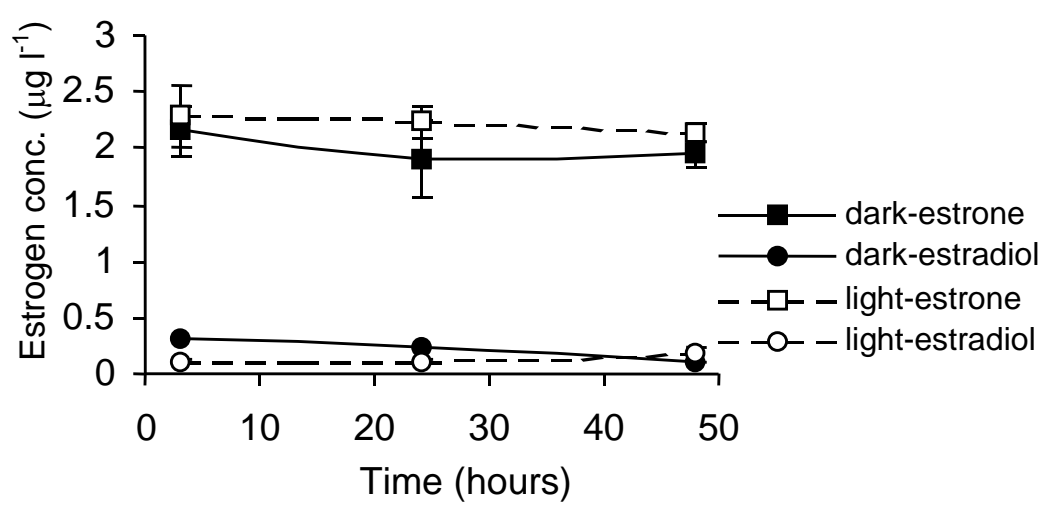


Figure 2. Changes with time (in light and dark conditions) observed after spiking the culture with estrone

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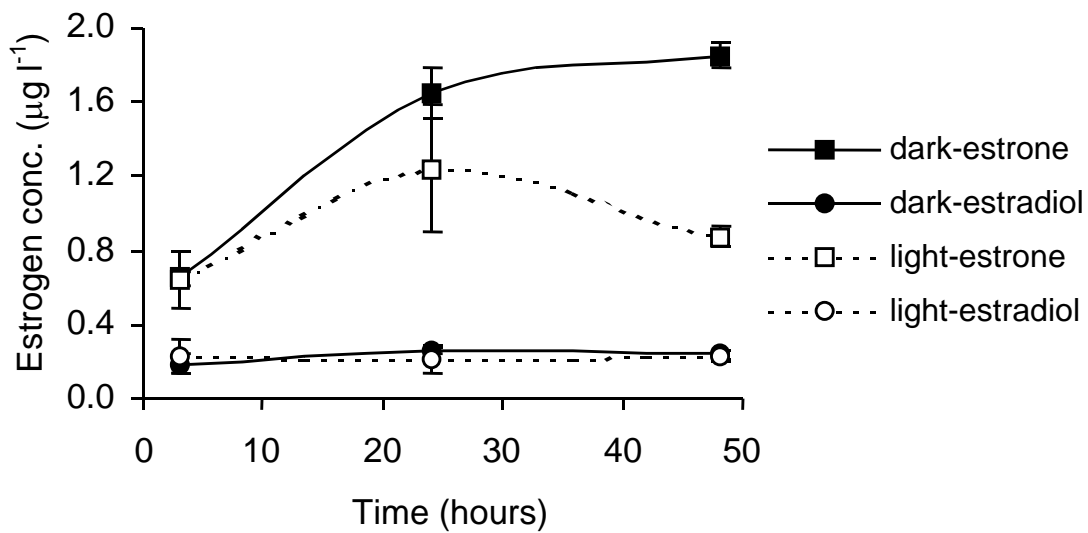


Figure 3. Changes with time (in light and dark conditions) observed after spiking the culture with estradiol valerate

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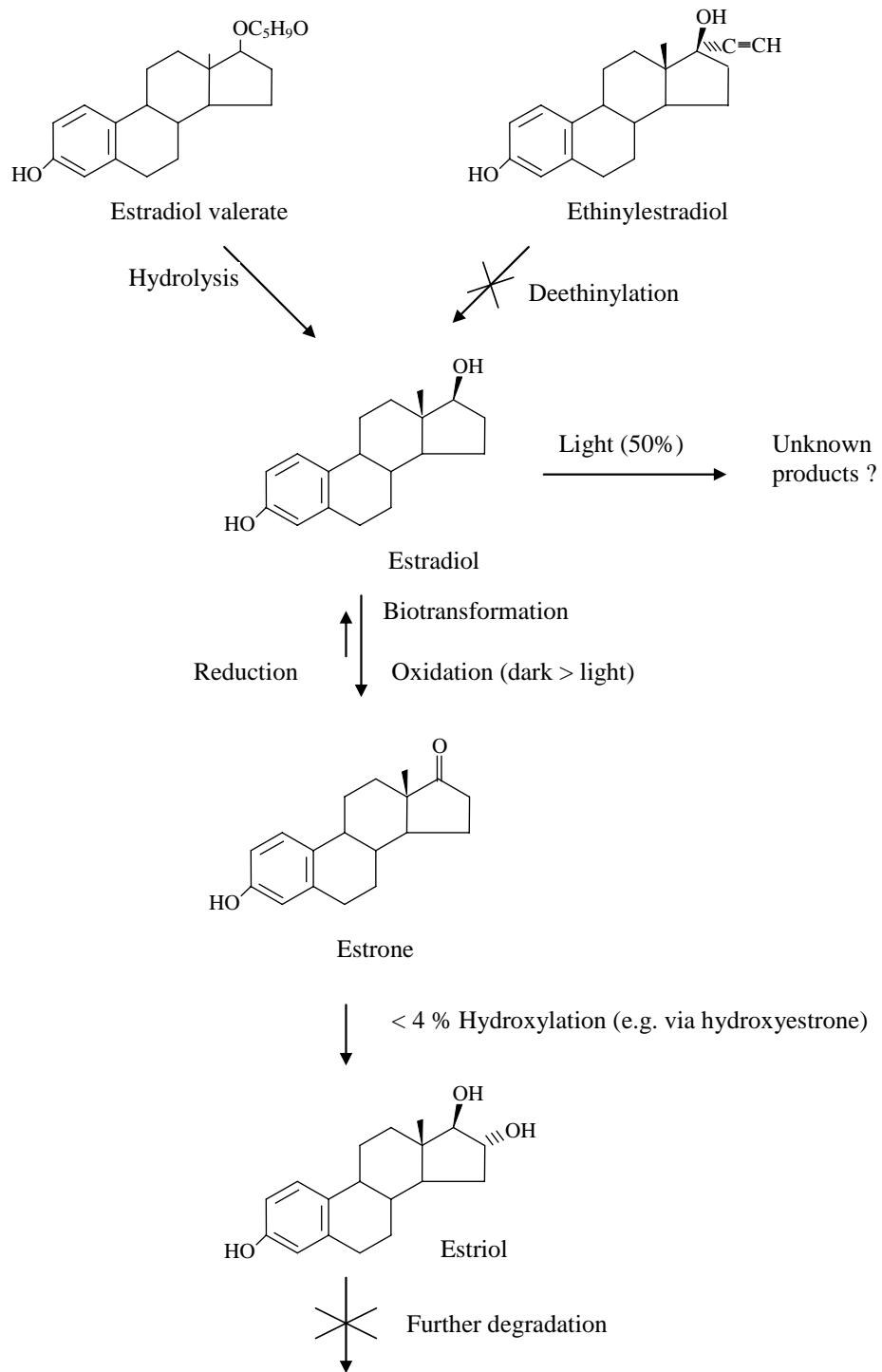


Figure 4. Proposed biotransformation pathway of steroid estrogens by *C. vulgaris*

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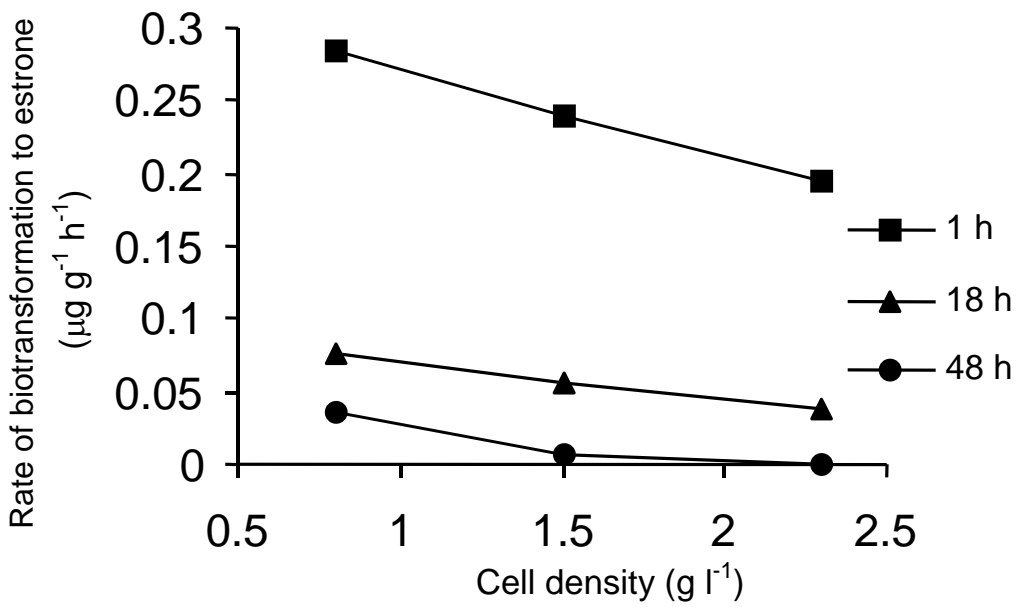


Figure 5. Effect of cell density on the formation of estrone from estradiol

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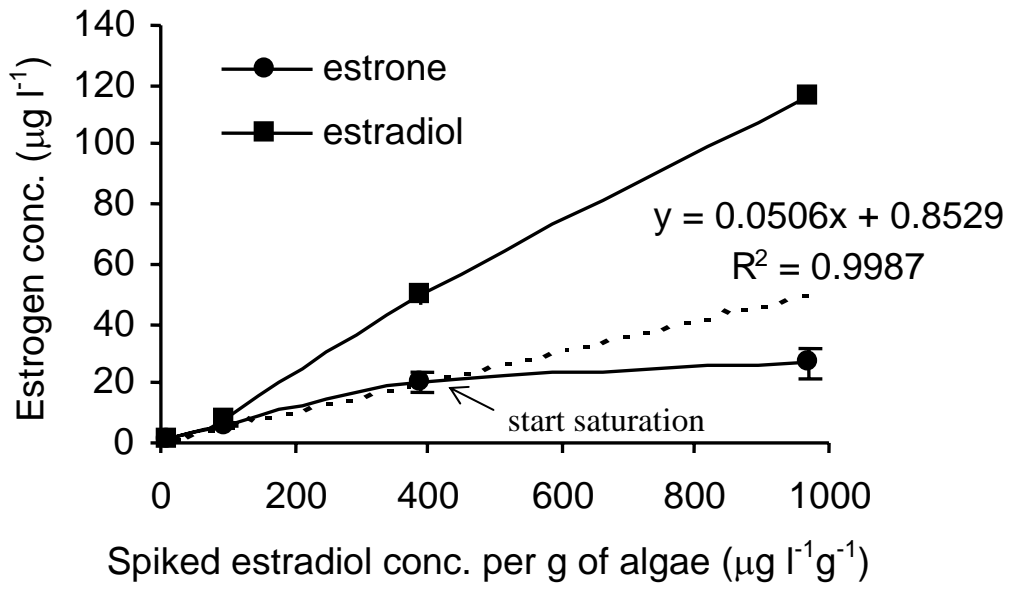


Figure 6. Effect of estradiol concentration on the biontransformation

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Table 1. Mass balance and distribution of estrogens in algae and filtrate in samples spiked with estradiol

DARK Hours	Mass balance (ng)	Algae (ng g ⁻¹)		Filtrate (µg l ⁻¹)		Partitioning to algae* (%)
		Estradiol	Estrone	Estradiol	Estrone	
Initial	500	0	0	2.5	0	0
3	506	98	66	2.2	0.26	3
24	485	201	175	0.63	1.62	8
48	426	137	72	0.22	1.81	5

LIGHT						
Initial	3	24	48	Mass balance (ng)	Algae (ng g ⁻¹) Estradiol	Algae (ng g ⁻¹) Estrone
Initial	3	24	48	500	0	0
				362	206	17
				298	59	269
				252	30	97

14 * Sum of estrogens in algae/ in filtrates

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Table 2. Mass balance, distribution and concentration factor of estrogens in algae and filtrate in samples spiked with estrone

DARK Hours	Mass balance (ng)	Algae (ng g ⁻¹)		Filtrate (µg l ⁻¹)		Partitioning to algae* (%)	Conc. factor [#]
		Estradiol	Estrone	Estradiol	Estrone		
Initial	500	0	0	2.5	0	0	0
3	509	103	47	0.31	2.16	3	22
24	445	83	98	0.24	1.90	4	52
48	424	109	41	0.10	1.95	4	21

LIGHT							
Hours	Mass balance (ng)	Algae (ng g ⁻¹)		Filtrate (µg l ⁻¹)		Partitioning to algae* (%)	Conc. factor [#]
		Estradiol	Estrone	Estradiol	Estrone		
Initial	500	0	0	2.5	0	0	0
3	510	240	70	0.11	2.29	6	31
24	481	72	26	0.12	2.24	2	12
48	487	157	47	0.17	2.14	7	22

14 * Sum of estrogens in algae/ in filtrates; [#] Concentration factor of estrone in algae/ in filtrate

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