

GOPEN ACCESS

Citation: García-Gómez JC, González AR, Maestre MJ, Espinosa F (2020) Detect coastal disturbances and climate change effects in coralligenous community through sentinel stations. PLoS ONE 15(5): e0231641. https://doi.org/10.1371/journal.pone.0231641

Editor: Carlo Nike Bianchi, Universita degli Studi di Genova, ITALY

Received: September 16, 2019

Accepted: March 27, 2020

Published: May 5, 2020

Peer Review History: PLOS recognizes the benefits of transparency in the peer review process; therefore, we enable the publication of all of the content of peer review and author responses alongside final, published articles. The editorial history of this article is available here: https://doi.org/10.1371/journal.pone.0231641

Copyright: © 2020 García-Gómez et al. This is an open access article distributed under the terms of the Creative Commons Attribution License, which permits unrestricted use, distribution, and reproduction in any medium, provided the original author and source are credited.

Data Availability Statement: All relevant data are within the manuscript and its Supporting Information files.

RESEARCH ARTICLE

Detect coastal disturbances and climate change effects in coralligenous community through sentinel stations

José C. García-Gómez*°, Alexandre R. González°, Manuel J. Maestre 6°, Free Espinosa°

Laboratorio de Biología Marina de la Universidad de Sevilla (LBMUS)/Área de Investigación I+D+i del Acuario de Sevilla/Estación de Biología Marina del Estrecho (Ceuta), Seville, Spain

These authors contributed equally to this work.

* jcgarcia@us.es

Abstract

This study was implemented to assess the Sessile Bioindicators in Permanent Quadrats (SBPQ) underwater environmental alert method. The SBPQ is a non-invasive and low-cost protocol; it uses sessile target species (indicators) to detect environmental alterations (natural or anthropic) at either the local or global (*i.e.*, climate change) scale and the intrusion of invasive species. The SBPQ focuses on the monitoring of preselected sessile and sensitive benthic species associated with rocky coralligenous habitats using permanent quadrats in underwater sentinel stations. The selected target species have been well documented as bioindicators that disappear in the absence of environmental stability. However, whether these species are good indicators of stability or, in contrast, suffer variations in long-term coverage has not been verified. The purpose of this study was to assess the part of the method based on the hypothesis that, over a long temporal series in a highly structured and biodiverse coralligenous assemblage, the cover of sensitive sessile species does not change over time if the environmental stability characterising the habitat is not altered. Over a ten-year period (2005–2014), the sublittoral sessile biota in the Straits of Gibraltar Natural Park on the southern Iberian Peninsula was monitored at a 28 m-deep underwater sentinel stations. Analyses of the coverages of target indicator species (i.e., Paramuricea clavata and Astroides calycularis) together with other accompanying sessile organisms based on the periodic superimposition of gridded images from horizontal and vertical rocky surfaces allowed us to assess the effectiveness of the method. We conclude that no alterations occurred during the study period; only minimal fluctuations in cover were detected, and the method is reliable for detecting biological changes in ecosystems found in other geographical areas containing the chosen indicator species at similar dominance levels.

Introduction

The long-term evaluation and quantification of changes in species are crucial for our knowledge of various marine ecosystems [1, 2] and constitute useful tools for the environmental Funding: This research supported by the Regional Activity Centre for Special Protected Areas (RAC/ SPA) and the Mediterranean Protected Areas Network (MedPAN) projects, Consejería de Medio Ambiente de la Junta de Andalucía, and cofinanced by Autoridad Portuaria de Sevilla, Compañía Española de Petróleos S.A.U. (CEPSA) Foundation and Red Eléctrica de España (REE). Also, research was partially supported by the University of Seville providing salaries to the authors JCG-G and, FE, and Research Foundation of University of Seville (different projects) providing salaries to ARG and MJM. The specific roles of these authors are articulated in the 'author contributions' section. The funders had no role in study design, data collection and analysis, decision to publish, or preparation of the manuscript.

Competing interests: This study was partly supported by Autoridad Portuaria de Sevilla, Compañía Española de Petróleos S.A.U. (CEPSA) Foundation and Red Eléctrica de España (REE). There are no patents, products in development or marketed products to declare. This does not alter our adherence to PLOS ONE policies on sharing data and materials. monitoring, management, and conservation of coastal zones and their associated marine protected areas.

Historically, methodological difficulties related to the monitoring and study of subtidal communities have led to their neglect in many environmental monitoring programs [3], particularly in relation to rocky habitats because of their wide heterogeneities, the fixed nature of the resident sessile organisms, the frequent difficulty of achieving access, and the fact that the sampling must be undertaken through indirect methods from boats. Warwick [4] indicated that the study of these communities on wide spatial scales involves a series of technical and implementation difficulties that are not present for work on soft bottoms, the water column, or the intertidal zone. For these reasons, studies related to the environmental monitoring, surveillance, and impacts of various factors on the sublittoral benthic environment have focused on endofauna (macroinvertebrates) associated with soft substrates [5–12]. Many biotic indices based on these animals have been proposed [13–15], and many of these indices have been used for or adapted to the requirements of the European Community Water Framework Directive [16–18].

Indeed, regarding monitoring methods based on indicators in the scope of the European Community Directives [19–20], the research has centred on macroalgae on hard substrates [21–28] and seagrasses [29–32], whereas macroinvertebrates associated with high-diversity rocky subtidal habitats (which have greater abundances of sensitive indicator species) have been less well studied. However, some studies have sought to establish the ecological statuses of coralligenous assemblages [33–39]. The environmental information that can be provided by these assemblages is very powerful and reliable [40–43]; therefore, this faunal approach to rocky habitats could help to fill an important knowledge gap within the European Community Directives [19–20].

Of particular importance is the information that can be provided by pre-coralligenous and coralligenous rocky bottoms because of their high biodiversities and the abundances of sensitive colonial species with long life cycles. Although there are some long-term studies of species that are representative of coralligenous communities [44–49] short- and medium-term studies are more common [50,51].

In contrast, growing concern about the effects of climate change has led to more studies of epibenthic (rocky bottom) species in relation to coral bleaching [52–54], seagrass meadows [55], invasive alien species [56,57], and even the role of the marine reserve biota in climate change monitoring [58–60]. These studies have also highlighted the severe environmental impacts on benthic organisms that have arisen from abnormal temperature increase events in the western Mediterranean [61–63].

The great fragility of highly structured and mature benthic communities associated with hard substrates, such as those of pre-coralligenous and coralligenous areas, has encouraged non-invasive study methods based on video footage or photos of species that are fixed to the substrate in random quadrats [36,51,64–75]. However, the absence of long-term series that have empirically validated these methodologies as systems for detecting changes in communities has been limiting.

Additionally, previous studies have focused on characterising and/or monitoring benthic communities, which implies greater complexity with respect to implementation mainly due to the high diversity of species present in these enclaves, which, in turn, entails greater effort and difficulty associated with the identification of the taxa. However, the Sessile Bioindicators in Permanent Quadrats (SBPQ) methodology, which was proposed by one of the authors of the present paper [43], differs from other methodologies, mainly in that the quadrats are permanently fixed to the rocky bottom and that the coverage of a previously selected target species must be at least 10% within each quadrat. Fixed quadrats allow changes in the coverage of

target species and the rest of the benthic community to be detected without the need for a high number of replicates, and with the certainty that these differences are not due to the characteristic spatial heterogeneity of this kind of habitat. The simplicity and relatively low cost of the method allows the monitoring of the stations to be repeated in the short term, which is essential in the detection of invasive species or other anthropogenic disturbances that require rapid action. The method is not intended to assess the biodiversity or ecological status of these communities, but to establish an early warning system that allows the detection of changes in the coverage of target species. Based on the detection of these changes, specific studies can be designed to assess the degree and origin of the disturbance.

Objectives

This study aims to assess the underwater environmental SBPQ alert method, which focuses on monitoring preselected sessile and sensitive benthic species associated with rocky coralligenous habitats using permanent quadrats in underwater sentinel stations. The selected target species have been well documented as bioindicators that disappear in the absence of environmental stability via acute impacts. However, it has not been verified if these species are good indicators of long-term stability. The purpose of the present study was to assess the portion of the method based on the hypothesis that, over a long temporal series (a ten-year period in this study) in a highly structured and biodiverse coralligenous assemblage, the cover of sensitive sessile species (*i.e.*, the target species) does not change over time if the environmental stability of the habitat is not altered.

Materials and methods

Study zone

This study was performed on the benthic community of two sentry stations on a rocky bottom. The station is situated within the Strait of Gibraltar Natural Park (southern Spain) inside the Grade A protection zone (*i.e.*, the maximum protection area within the Natural Park) characterized by rocky bottoms with a moderate slope that host a high biodiversity and species richness, dominated by species such as *Eunicella sp., Paramunicea clavata, Astroides calycularis, Pentapora sp., Crambe crambe* and *Salmacina sp.* in the Punta Carnero locality (Algeciras), but it is proximal to zones under strong anthropic pressure (Fig 1).

Despite its enormous ecological wealth and high level of protection, the SBPQ sentry station is at notable risk of anthropogenic disturbance. This is due to both to its proximity to the particularly industrialised Bay of Algeciras, where many different pollution sources can be found (thermal power plant, chemical industry, petrochemicals, bunkering activities, etc.) [76,77] and the high level of marine traffic through the area. Due to the second factor, the last decade has seen polluting events related to hydrocarbon spills of varying magnitudes including serious spills from the ships *Sierra Nava* and *New Flame*, both of which occurred in 2007 [78].

Data regarding the bottom water temperature in the study area were obtained from data base of the regional environmental authorities [79]. The implementation of the study was notified to the managers of the the Strait of Gibraltar Natural Park and to the competent environmental authority of Andalusia Government (Regional Ministry of Environment and Territory Planning, CMAYOT). In fact, the CMAYOT has funded the publication of the spanish verison book where the methodology SBPQ is included (43). In adition the first author of the current study is part of the governing board of the Strait of Gibraltar Natural Park. No permits were required for this work.



Fig 1. Location of the sampling point in the Strait of Gibraltar Natural Park.

https://doi.org/10.1371/journal.pone.0231641.g001

Application of the SBPQ methodology for its validation: Target species, permanent grids, and underwater sentinel stations

The SBPQ method was proposed by García-Gómez [43] together with an identification guide for sensitive indicator species vs. tolerant species, and was published by The Regional Activity Centre for Specially Protected Areas (RAC/SPA). The RAC/SPA was established in Tunis in 1985 and is responsible for assessing the status of the natural heritage and helping Mediterranean countries to implement the Protocol concerning Specially Protected Areas and Biological Diversity in the Mediterranean (SPA/BD Protocol), which came into force in 1999.

This methodology has been designed as a simple, non-invasive, underwater environmental alert tool for the potential early detection of environmental impacts of anthropic origin in the sublittoral system: in the short term (local alterations derived from pollutants from industries or emissions of urban origin, coastal dredging or civil engineering works on the coast, intrusion of exotic species with invasive potential, among other sources of alteration of coastal waters), and in the medium or long term (global warming). Even though the method is able to detect the presence of invasive species, further studies are required to test the reliability of the method for detecting other potential impacts of anthropic origin. Recently, the SBPQ methodology has already been used successfully in the early detection of exotic species with invasive potential, in particular the invasive seaweed *Rugulopterix okamurae* [80]. The Fig 2 shows a temporal variation of benthic mean percent species coverage at Tarifa Island monitoring by SBPQ method from years 2013 to 2017.

It is focused on the management of sensitive indicator target species (benthic and sessile) vs. tolerant species in the Western Mediterranean and can be used in other parts of the world once the native species have been selected. The SBPQ method has been proposed for wide-spread use, not only by scientists, but by sports divers linked to diving clubs or centres that could be involved in underwater environmental monitoring of the coastal environment (Citi-zen Science and "Working with Nature" philosophies).

Synthetic adaptation of the SBPQ protocol

- 1. Choice of vertical walls of pristine rocky bottoms, preferably between 20 and 35 metres deep, biologically structured and of high biodiversity, with the presence of adult-sized target species, sensitive, benthic and sessile indicators, that are visible underwater, preferably colonial, with a long life cycle, and abundant compared to the local macrobiota.
- 2. Selective installation of three to five permanent quadrats of $1x1 \text{ m}^2$ (not chosen at random), located on patches of at least one previously selected target species that represents at least 10% of the total coverage per quadrat (Fig 3). The method has recently been updated with the objective of minimizing the degree of intrusion on these fragile habitats. For this purpose, a single hole is drilled on the rock and a small plastic bar marks off four 50 x 50 cm detachable monitoring quadrats that are fitted each time monitoring is carried out. This



Fig 2. Temporal variation of benthic mean percent species coverage at Tarifa Island monitoring station fixed quadrats from years 2013 to 2017 (1, 2, 3, 4 refer to random sampling times within each year). Species with less than 10% coverage (*Alcyonium sp., Aplidium sp., Asparagopsis armata, Crambe sp, Ircinia sp. and Polycitor adriaticum*) were grouped under 'Other species'. Taken from (García-Gómez et al., 2020).

https://doi.org/10.1371/journal.pone.0231641.g002



Fig 3. Sentinel station showing permanent quadrats of 1x1m located on patches of target species.

avoids the use of permanent fixed quadrats with many screws, therefore reducing effort, cost, maintenance and, most importantly, the impact on communities. As explained above, three to five of such monitoring points would be installed per site (Fig 4). A hand drill (Fig 4A), used in climbing, is proposed in the method. This tool is a cheap and relatively simple way to fix the pieces to the rock and can be supplied to diving clubs that are interested in being part of the monitoring network.

The anthozoans *Astroides calycularis* (Pallas, 1766) and *Paramuricea clavata* (Risso, 1826), which are currently listed as 'vulnerable' by the IUCN, were selected for monitoring. These species are vulnerable to increases in temperature and sensitive to deterioration of the environmental quality of the marine environment. Both species are well referenced in ecological studies of coastal benthos [42,43,63, 81–88]. These species should always be moderate to large (preferably colonial) and highly visible while immersed. Moreover, some studies have indicated that typical coralligenous species, such as *Paramuricea clavata*, exhibit extremely low temporal variability [88]. Additionally, these species are highly representative of the community under study. Therefore, the quantitative stability (cover) of these sensitive species over time in the absence of phenotypic signs of stress (*e.g.*, epibiosis, partial necrosis, bleaching) would allow us to infer that the community has remained healthy. Therefore, focusing the monitoring effort on sensitive species that are representative of the community should allow the appropriate inferences to be drawn.

Sampling procedure

Four fixed 1-m² PVC quadrats were installed in two different locations. One setup had a vertical orientation (shady), and the other had a horizontal orientation (sunny) at a depth of 28 m



Fig 4. The updated method with the objective of minimizing the degree of intrusion. A: Hand drill, B: Expansion anchor bolt, C: Drilling and bolt fixing sequence, D: Metal crosshead fixation, E: Sequential allocation of the detachable quadrat, F: Representation of a complete monitoring point.

(coordinates 36° 4,770 N - 05° 25,184 W; GPS, DATUM WGS 84). These quadrats were placed on surfaces containing target species. The covers were monitored over ten consecutive years (2005–2014) in spring and summer seasons.

A four video footage of $50 \ge 50$ cms were filmed sequentially, covering a total of $1 \ge 1$ m quadrat, in order to gain higher resolution of the subsequent photos taken from the video.

A total of 320 photographs were digitally manipulated using the *Adobe Photoshop* 6.0 (2000) program as follows: four 50x50-cm areas were created for each quadrat (Fig 5A); the contrast and colour saturation were adjusted (Fig 5B); a complete digital frame was created (Fig 5C); and a digital grid that was adjusted to the perimeter of the monitoring area was superimposed (Fig 5D). To correct the angular deformation.

To assess the degree of the cover of the target species, a system involving the determination of the presence/absence of cover using the digital grid was applied. This system aided estimation accuracy [89,90] and optimised the working time [64,91]. Nevertheless, it should be noted that, in multistratified communities, the system tends to underestimate the cover [64,69,89–92] because larger species "mask" those that developed underneath them.

Statistical analyses

Repeated measures analysis of variance (RM-ANOVA) was applied to test whether the mean coverage of the target species significantly varied either through time (intra-subject factor 'time'; ten levels: 2005 to 2014) or according to orientation (inter-subject factor 'orientation'; two levels: horizontal and vertical). Mauchly's test of sphericity [93] was used to test the assumption that the variances in the differences between all possible pairs of groups were equal. When the sphericity condition was not verified, the F test value was corrected with the Greenhouse-Geisser epsilon index [94,95]. The factors of time and orientation were considered orthogonal and fixed. The same two-factor design was considered to test for any significant differences in the mean coverages via RM-permutational analysis of variance



Fig 5. Digital manipulation of the captured images prior to analysis.

(RM-PERMANOVA). Univariate analyses were performed using SPSS.13[®] according to the guidelines of Pardo and Ruiz [96].

For the multivariate analyses, square root transformations were applied to the data, and the analyses were performed using the Bray-Curtis similarity. The Bray-Curtis similarity matrix was used to generate a non-metric multidimensional scaling (nMDS) analysis, and the Kruskal stress coefficient was calculated to test the ordination [97].

The multivariate analyses were performed using the PRIMER6 software (complete with the PERMANOVA+ package) [98–101].

Results

The average covers for the five most abundant species (including the two pre-selected sensitive target species *A. calycularis* and *P. clavata*) over the period of 2005 to 2014 at the two locations (horizontal and vertical) are presented in Figs 6 and 7.

Fig.8 illustrates the evolution of the covers of the sampled indicator species *A. calycularis* and *P. clavata* at both locations (horizontal and vertical) together with the time series of the





background water temperatures in the sampling zones. These series reflected no anomalous increases or decreases in temperature and only exhibited oscillations that were attributable to seasonality. The covers of both species were clearly greater in the vertical orientation than in the horizontal orientation. An increase and a decrease were observed in the covers of *A. calycularis* and *P. clavata*, respectively, but there were no significant differences. This decrease in the coverage of P. clavata throughout the period studied coincides with that observed in other studies [47]. Environmental factors such as climate change may be affecting certain species more slowly these trends may be analyzed when they have a longer period.

RM-ANOVA of the cover of each species in each of the two orientations (Table 1) indicated significant differences (p<0.05) in the covers of *Mesophyllum* sp. and *Salmacina incrustans* in the shady (vertical) location over the time period. Regarding the sunny (horizontal) location, the covers of *Mesophyllum* sp. and *Crambe* sp. exhibited significant differences (p<0.05). No significant differences (p>0.05) were detected in the covers of the remaining species.

The RM-PERMANOVA analysis indicated significant differences in the coralligenous assemblages (p < 0.05) for the orientation factor (*i.e.*, vertical vs. horizontal) but not for the time factor. The interaction of these two factors was also not significant (<u>Table 2</u>).

The nMDS analysis (Fig 9) revealed a clear differentiation of the values between the vertical and horizontal orientations (Stress: 0.14).

Discussion

Sampling procedure and analysis of the submarine images

Previous studies have indicated that fixed monitoring stations constitute one of the most robust methods for detecting changes in benthic communities over time [69,102-105]. An area of at least 1 m² is sufficient for integrating various colonies and individuals of diverse fauna in a single sample, and such diversity is a typical characteristic of benthic rocky bottom communities [69,106]. However, the definitive sampling grid size must be defined by the final



Fig 7. Photos of the same replicate taken in 2005 and 2014 for both orientations.

objective of the study [107]. In our case, the objective was not an exhaustive description of the existing community but rather the evaluation of possible changes in the covers of sessile target species within the community. Similarly, the number of quadrats or replicates used must be defined by the target objectives. In other studies of rocky bottoms or surfaces, three [36,108–109], five [110], six [111], and eight replicates [112] have been employed. In all of these studies, the replicates had areas of significantly less than 1 m² with the exceptions of the studies by Fraschetti *et al.* [109] and Piazzi *et al.* [36] who also used areas of 1 m².

Image analysis based on a monitoring system comprises a non-invasive method that does not interfere with the natural development or evolution of the studied community [69, 92, 103, 113,



Fig 8. Graphic representation of the evolution of the cover of the two main indicator species (*A. calycularis* and *P. clavata*), together with the bottom water temperature time series. The bars indicate the standard deviation. H: horizontal; V: vertical.

114]. This method also allows for rapid data collection and permanent data record generation. The images were captured using videos rather than photographs mainly due to video's greater speed and versatility when obtaining data [115–117]. Video was thus time-saving and consequently allowed for the optimisation of diving operations, which was very important given the depth at which the study was performed and the number of replicates that were monitored.

Choices of high biodiversity habitats, target species, and other companion organisms

In general, sublittoral marine habitats are characterised by diversity that increases with depth [66, 118, 119]. Deep communities are dominated by animals, are better structured, and exhibit

| | | A. calycı | ılaris | P. cla | ivata | Cram | be sp. | Eunice | lla sp. | Mesophy | llum sp. | P. fas | cialis | S. incr | ustans |
|------------------------------|----|--------------------|------------|--------------------|------------|--------------------|------------|--------------------|------------|--------------------|------------|--------------------|------------|--------------------|------------|
| Source of variation | df | F | <u>P</u> * | <u>F</u> | <u>P</u> * |
| Vertical | | | | | | | | | | | | | | | |
| Time ⁺ | 9 | 1.159 ² | 0.335 | 1.647 ² | 0.171 | 0.793 ² | 0.557 | 0.632 ¹ | 0.359 | 3.474 ² | 0.018 | 2.735 ² | 0.058 | 2.931 ² | 0.021 |
| Residuals | 10 | | | | | | | | | | | | | | |
| Mauchly's test of sphericity | | p<0.001 | | p<0.001 | | p<0.05 | | p = 0.305 | | p<0.01 | | - | | - | |
| Transformation | | None | | none | | none none | | none | | none | | none | | | |
| Horizontal | | | | | | | | | | | | | | | |
| Time ⁺ | 9 | 0.671^{1} | 0.277 | 0.593 ¹ | 0.446 | 5.720 ² | 0.002 | 1.749 ² | 0.157 | 0.910 ¹ | 0.006 | 1.087^{2} | 0.365 | 0.890 ² | 0.426 |
| Residuals | 10 | | | | | | | | | | | | | | |
| Mauchly's test of sphericity | | p = 0 | .490 | p = 0.052 | | p<0.01 p<0.001 | | p = 0.101 | | - | | - | | | |
| Transformation | | No | ne | none | | no | ne | none | | none | | none | | none | |

Table 1. Results of one-way repeated measures ANOVA on coverage of each species either in vertical and horizontal substrata.

* Greenhouse-Geisser correction was used when the sphericity condition was not verified.

⁺Time correspond with years from 2005 to 2014.

¹ Pillai´s trace.

² Greenhouse-Geisser F.

https://doi.org/10.1371/journal.pone.0231641.t001

| Source | df | <u>SS</u> | MS | Pseudo-F | P(perm) | Unique perms | |
|----------------|------|-----------|--------|----------|---------|--------------|--|
| Time | 9 | 2150.5 | 238.95 | 2.4353 | 0.079 | 998 | |
| Orientation | 1 | 30338 | 30338 | 309.2 | 0.001 | 998 | |
| Ti x Or | 9 | 1207.9 | 134.21 | 1.3678 | 0.359 | 999 | |
| Res | 60 | 5887.1 | 98.118 | | | | |
| Total | 79 | 39583 | | | | | |
| Transformation | None | | | | | | |

Table 2. Results of repeated measures PERMANOVA.

https://doi.org/10.1371/journal.pone.0231641.t002

less abrupt dynamics and smaller temporal changes than shallow communities or those dominated by algae [66]. Additionally, as pointed out by Ballesteros [120], within the scope of marine protected areas (MPAs), the selection of a limited number of representative and/or key species from such communities is a sound strategy for aiding their understanding and management.

Specific fixed benthic organisms have previously been used as indicators for the monitoring of various environmental parameters; such species have been used as indicators of global warming and climate change [63, 121–124], sea level fluctuations [125,126], and the influence of recreational diving on MPAs [62, 83, 127,128].



Fig 9. Graphic representation of the non-metric MDS analysis for the whole monitoring quadrats.

https://doi.org/10.1371/journal.pone.0231641.g009

Moreover, the use of representative taxa as proxies for entire communities has proven to be a reliable alternative for evaluating the state of rocky bottom communities comprising algae because the loss of information associated with the identification of only some specific taxa does not greatly alter the results compared with results obtained with complete datasets based on entire communities [3]. In hard bottom invertebrate communities, it has also been found that sampling efforts can be reduced without significant losses of information via the selection of indicator and representative species [129].

A. calycularis, P. clavata, and *Eunicella* sp. (which were clearly more abundant in the monitored vertical enclaves) play an important structural role in coralligenous assemblages. These species colonise both horizontal bottoms and vertical walls and offer ideal habitats for numerous other organisms [43, 130, 131]. Several previous studies of coralligenous zones in different areas of the Mediterranean, such as the coasts of Italy and France, have determined that the facies of *Paramuricea* and *Eunicella* are distinctive [87, 132].

None of these species exhibited significant cover alterations over the entire monitoring period. Although *P. clavata* and *Eunicella sp.* have similar biological and ecological characteristics, including similar rates of growth and production [133–135], this similarity does not extend to their turnover rates. For *P. clavata*, this rate oscillates between 7 and 9 years [133,135], whereas for *Eunicella* sp., the rate is significantly lower with a range of 3 to 4 years [134]. For this reason, during the long-term temporal monitoring of fixed, limited points, we would expect *Eunicella* sp. to exhibit greater oscillations in cover percentages than those that we observed due to mortality among the colonies within the monitoring quadrats. Nevertheless, this phenomenon was not recorded in our work, possibly because the magnitude of the studied area allowed for the replacement of the dead organisms by new recruits to buffer the potential differences.

Also, it is important to point out that the two main indicator species, *A. calycularis* and *P. clavata* (particularly the latter), are sensitive to changes in normal temperature conditions [122, 136–138], although they both withstand seasonal fluctuations well [120].

The differences found in the covers of certain species can also be explained by natural processes. *Mesophyllum alternans* is among the organisms responsible for the greatest percentage of cover (particularly in the horizontal orientation), but it is considered to be opportunistic [51], which might explain any oscillations in cover in the absence of any significant alterations to the reference species. The other two species that exhibited significant differences, *i.e.*, *Salmacina incrustans* and *Crambe* sp., are organisms that account for very low percentages of the overall cover. Variations in abundances in deep rocky bottom communities are usually much more noticeable among minor species than among the highly abundant or characteristic species [139]. In any case, these two species can be classified as tolerant and relatively resistant to environmental impacts [43].

Recent studies of coralligenous areas have only used vertical enclaves [132], but, despite comprising the same species, the biological communities that occur at the two orientations are clearly different due to the varying percentages of cover formed by those species. Encrusting algae dominate horizontal enclaves, whereas populations of *P. clavata* and *A. calycularis* are more abundant on vertical walls; these differences have previously been mentioned by other authors [120, 140]. Therefore, wherever possible, stations should be set up in both orientations, although vertical coralligenous enclaves provide more information because they have greater biodiversity, are better structured spatially and trophically, and have greater numbers of sensitive colonial species.

Applications of fixed quadrats for the monitoring of rocky bottom habitats

The method tested here would be particularly useful for the WFD/MSFD in relation to the environmental control of littoral zones based on biological indicators. The WFD/MSFD has

very few methodological tools for monitoring sessile benthic invertebrates associated with hard substrates, although the tools that do exist are excellent [141]. The environmental information provided by these sessile benthic species is significant because it covers a wide range of sessile organisms that are long-lived and sensitive to changes in the environment. These species can be used as reliable ecological sentries that keep watch over the quality of littoral environments because they are not able to flee or be displaced (as adults) when conditions deteriorate or significantly change.

The effectiveness of this monitoring method was demonstrated over the 10-year period used to obtain a complete quantitative temporal series. As indicated by Gatti [87,142], such series are very useful for evaluating both reference conditions and environmental changes in ecosystems. Indeed, the main macrobiota species in the sublittoral community in this study were stable and remained in excellent condition from 2005 to 2014.

The use of images is advantageous because, unlike the removal of physical samples by scraping, it is neither destructive nor aggressive to the surroundings. The scraping approach is not effective as a monitoring method in zones in which the fauna is distributed in patches because it entails the destruction of relatively large areas [106]. In protected locations and areas of special interest, this scraping option is particularly undesirable.

The present image-based monitoring method is "low cost" in addition to being "low effort" and requires minimal maintenance (repairs or replacements of the fixed quadrats were necessary only twice over the 10-year monitoring period). Panayotidis [22] indicated that the use of simple monitoring programs (*i.e.*, those involving taxonomic efforts that are limited to representative species and simple statistical treatments) and low budget methods is demonstrably effective for other rocky bottom communities.

According to recent studies, the minimum replication area required for biodiversity studies of communities dominated by *Paramuricea clavata* [143], as well as other hard-substrate benthic invertebrates [92, 144,145], is less than the area used in the replicates in our present work. Additionally, the captured images remain available for subsequent, more detailed descriptions of the community. Higher quality images facilitate the visual identification of more of the macrobenthic species that are present in the community, and the system of photo-quadrats is more efficient than *in situ* visual census methods [92] while simultaneously avoiding the destruction of the monitored areas.

As a final summary, Table 3 presents a comparison of the data from the monitoring area of the present study with the data from several other studies of coralligenous bottoms in terms of time and depth. Accounting for both area and time, the present study represents a long-term monitoring of coralligenous assemblages.

| Table 3. Comparison of the data of me | onitoring area, time and | d depth of several | studies on coralli | genous bottoms |
|---------------------------------------|--------------------------|--------------------|--------------------|----------------|
|---------------------------------------|--------------------------|--------------------|--------------------|----------------|

| Study | | Location | Area (m ²⁾ | Depth (m) | Time (years) |
|---------------|----------------------|---------------------|-----------------------|-----------|--------------|
| [139] | Peckol P (1984) | North Carolina | 0.132 | 20 | 2 |
| [51] | Garrabou J (2000) | Medes Islands | 1.085 | 15-19 | 2 |
| [44] | Garrabou J (2002) | Marseilles, France | 0.4 | 27 | 21 |
| [66] | Garrabou J (2002) | NW Mediterranean | 0.372 | 17-20 | 2 |
| [146] | De Biasi A M (2004) | Aegean Sea | 33.6 | 10-35 | 1 |
| [112] | Bussotti S (2006) | Southern Italy | 13.25 | 6-8 | 1 |
| [147] | Parravicini V (2009) | Ligurian Sea | 25 | 4-5 | 1 |
| [36] | Piazzi L (2017) | Ligurian Sea | 180 | 30-40 | 1 |
| [92] | Sant N (2017) | Cabrera Archipelago | 4.96 | 0-50 | 1 |
| Present study | | | 8 | 27-30 | 10 |

https://doi.org/10.1371/journal.pone.0231641.t003

Finally, the proposed methodology could be extended to the entire geographical area over which the target species of this study are representative of the rocky bottom benthic communities that occur at similar depths and share similar biotic structures. This methodology could also be applied in other areas in which target species can be established and selected by employing the aforementioned implementation criteria. Additionally, the SBPQ method involves minimal implementation difficulty both in terms of cost and installation. The *a priori* selection of the target species simplifies the taxonomic difficulty of visual identification. Therefore, this environmental warning and surveillance system is not only oriented toward the scientific and technical communities of the relevant coastal countries but also toward environmental volunteers associated with diving centres and clubs. This methodology ultimately entails the aim of setting up geographical networks in regions with sufficiently high levels of homogeneity (e.g., the Alboran Sea or, on a larger scale, the western Mediterranean). Citizen science initiatives focusing on the mapping and monitoring of coralligenous assemblages have recently been implemented [148]. In the future, these networks could provide early warnings of changes to structured and unpolluted systems, particularly those in littoral zones close to places subjected to strong anthropogenic pressure, which could be applied to the detection of local or general environmental alterations (*i.e.*, climate change). In contrast, the nature of the method, which is designed to monitor a group of taxa with respect to the total taxa of the community, implies that it is not useful as such in the ecological study of these habitats because the analysed information is biased. However, the generated databases (*i.e.*, stored photographs) will allow for the performance of in-depth studies of the dynamics of the species within the communities over time, assuming that significant changes can be detected, and could aid parallel projects that aim to increase our knowledge of coralligenous assemblages.

Conclusions

The study assessed the Sessile Bioindicators in Permanent Quadrats (SBPQ) underwater environmental alert method in relation to the monitoring of pre-selected sensitive and sessile benthic species (*P. clavata* and *A. calycularis*) associated with rocky coralligenous habitats (which exhibit high stability and biodiversity) via the use of permanent quadrats fixed on rocky shores in underwater SBPQ sentinel stations. The obtained results have allowed for assessing the part of the method based on the initial hypothesis that, over a long temporal duration (a ten-year period in this study) and in a highly structured and biodiverse coralligenous assemblage, the cover of sensitive sessile species does not change over time if the environmental stability characterising the habitat is not altered. This stability of the cover of the sensitive sessile species is a key aspect for confirming the reliability and robustness of the SBPQ method. Given that, as in the present study, the selected species are very sensitive to increases in temperature and deterioration of the environmental quality of the water column, the SBQP method is useful as an underwater environmental alert system because it should be solely sensitive to changes in the coverages of such species that result from physico-chemical changes in the system. Such changes include gradual increases in temperature due to global warming and changes due to the introduction of exotic species.

A future usefulness would be to implement the SBPQ methodology in well-conserved areas so those areas can act as "SBPQ sentinel stations" in the event of possible disturbances. The method has been developed as a simple management tool for use by scientists and specialised technicians in addition to diving clubs that frequent certain areas.

Supporting information

S1 Table. Coverage of each species through the monitoring period. Time: 1 corresponding to 2005 and 10 to 2014. Slope: H, horizontal; V, vertical. (XLS)

Acknowledgments

We want to warmly thank the Campo de Gibraltar, Centro de Investigaciones y Exploraciones Submarinas CIES Algeciras and CIES Tarifa Diving Centers, for their joint participation with the authors in a general "Citizen Science" program related to underwater sentinel stations. Also, we express our gratitude for the collaboration provided to the Seville Aquarium, Town Halls of Tarifa and La Línea as well as to Puerto Deportivo La Alcaidesa, Club Marítimo y Real Club Náutico (RCN) de La Línea. Finally, we would like to mention Academic Editor PLOS ONE and the reviewers of this paper for their valuable recommendations, which improved the contents of this work.

Author Contributions

Conceptualization: José C. García-Gómez.

Data curation: Alexandre R. González.

Formal analysis: José C. García-Gómez, Alexandre R. González.

Funding acquisition: José C. García-Gómez.

Investigation: José C. García-Gómez, Alexandre R. González, Manuel J. Maestre, Free Espinosa.

Methodology: José C. García-Gómez.

Project administration: José C. García-Gómez.

Resources: José C. García-Gómez.

Software: Alexandre R. González.

Supervision: José C. García-Gómez.

Validation: José C. García-Gómez, Alexandre R. González, Manuel J. Maestre, Free Espinosa.

Visualization: José C. García-Gómez, Alexandre R. González.

Writing - original draft: José C. García-Gómez, Alexandre R. González.

Writing – review & editing: José C. García-Gómez, Alexandre R. González, Manuel J. Maestre, Free Espinosa.

References

- Gatti G, Bianchi CN, Morri C, Montefalcone M, Sartoretto S. Coralligenous reefs state along anthropized coasts: Application and validation of the COARSE index, based on a rapid visual assessment (RVA) approach. Ecol Indic. 2015; 52: 567–576.
- Gatti G., Bianchi C. N., Montefalcone M., Venturini S., Diviacco G. Morri C. Observational information on a temperate reef community helps understanding the marine climate and ecosystem shift of the 1980–90s. Mar Pollut Bull. 2017; 114: 528–538. https://doi.org/10.1016/j.marpolbul.2016.10.022 PMID: 27743657
- 3. Puente A, Juanes J A. Testing taxonomic resolution, data transformation and selection of species for monitoring macroalgae communities. Estuar Coast Shelf Sci. 2008; 78: 327–340.
- 4. Warwick R M. Environmental impact studies on marine communities: pragmatical considerations. Australian Journal of Ecology. 1993; 18: 63–80.
- Pearson T H, Rosenberg R. Macrobenthic succession in relation to organic enrichment and pollution of the marine environment. Oceanography and Marine Biology: An Annual Review. 1978; 16: 229– 311.
- 6. Rygg B. Distribution of species along pollution-induced diversity gradients in benthic communities in Norwegian fjords. Mar Pollut Bull. 1985; 16: 469–474.

- 7. Ros J D, Cardell M J. Effect on benthic communities of a major input of organic matter and other pollutants (coast off Barcelona, western Mediterranean). Environ Toxicol Chem. 1991; 31–32: 441–450.
- 8. Dauer D M. Biological criteria, environmental health and estuarine macrobenthic community structure. Mar Pollut Bull. 1993; 26: 249–257.
- Estacio F, García-Adiego E, Fa D, García-Gómez J C, Daza J L, Hortas F, et al. Ecological analysis in a polluted area of Algeciras Bay (Southern Spain): external vs. internal outfalls and environmental implications. Mar Pollut Bull. 1997; 34–10: 780–793.
- Estacio F, García-Adiego E M, Carballo J L, Sánchez-Moyano J E, García-Gómez J C. Interpreting temporal disturbances in an estuarine benthic community under combined anthropogenic and climatic effects. J Coast Res. 1999; 15–1: 155–167.
- 11. Sánchez-Moyano J E, García-Adiego E M, Estacio F. García-Gómez J C. Effects of environmental factors on the spatial variation of the epifaunal polychaetes of the algae Halopteris *scoparia* in Algeciras Bay (Strait of Gibraltar). Hydrobiologia. 2002; 470: 133–148.
- 12. Guerra-García J M, García-Gómez J C. Crustacean assemblages and sediment pollution in an exceptional case study: a harbour with two opposing entrances. Crustaceana. 2004; 77–3: 353–370.
- 13. Majeed S A. Organic matter and biotic indices on the beaches of North Britanny. Mar Pollut Bull. 1987; 18: 490–495.
- 14. Grall J, Glémarec M. Using biotic indices to estimate macrobenthic community perturbations in the Bay of Brest. Estuar Coast Shelf Sci. 1997; 44: 43–53.
- 15. de Juan S, Demestre M. A Trawl Disturbance Indicator to quantify large scale fishing impact on benthic ecosystems. Ecol Indic 2011; 18:183–190.
- Borja A, Franco J, Pérez V. A marine biotic index to establish the ecological quality of soft-bottom benthos within European estuarine and coastal environments. Mar Pollut Bull. 2000; 40: 1100–1114.
- 17. Simboura N, Zenetos A. Benthic indicators to use in ecological quality classification of Mediterranean soft bottom marine ecosystems including a new biotic index. Mediterr Mar Sci. 2002; 3: 77–111.
- Rosenberg R, Blomqvist M, Nilsson H C, Cederwall H, Dimming A. Marine quality assessment by use of benthic species abundance distributions: a proposed new protocol within the European Union Water Framework Directive. Mar Pollut Bull. 2004; 49: 728–739. <u>https://doi.org/10.1016/j.marpolbul.2004</u>. 05.013 PMID: 15530516
- **19.** EC. Directive 2000/60/EC of the European Parliament and of the Council of 23 October 2000 establishing a framework for Community action in the field of water policy. Official Journal of the European Union. 22/12/2000; L 327: 0001–0073.
- EC. Directive 2008/56/EC of the European Parliament and of the Council of 17 June 2008 establishing a framework for community action in the field of marine environmental policy (Marine Strategy Framework Directive). Official Journal of the European Union. 25.6.2008; L 164: 19–40.
- Orfanidis S, Panayotidis P, Stamatis N. Ecological evaluation of transitional and coastal waters: a marine benthic macrophytes-based model. Mediterr Mar Sci. 2001; 2: 45–65.
- 22. Panayotidis P, Montesanto B, Orfanidis S. Use of low-budget monitoring of macroalgae to implement the European Water Framework Directive. Journal of Applied Phycology. 2004; 16: 49–59.
- Ballesteros E, Torras X, Pinedo S, García M, Mangialajo L, Torres M. A new methodology based on littoral community cartography dominated by macroalgae for the implementation of the European Water Framework Directive. Mar Pollut Bull. 2007; 55: 172–180. <u>https://doi.org/10.1016/j.marpolbul.2006</u>. 08.038 PMID: 17045303
- Wells E, Wilkinson M, Wood P, Scanlan C. The use of macroalgal species richness and composition on intertidal rocky seashores in the assessment of ecological quality under the European Water Framework Directive. Mar Pollut Bull. 2007; 55: 151–161. <u>https://doi.org/10.1016/j.marpolbul.2006</u>. 08.031 PMID: 17074370
- Juanes J A, Guinda X, Puente A, Revilla J A. Macroalgae, a suitable indicator of the ecological status of coastal rocky communities in the NE Atlantic. Ecol Indic. 2008; 8: 351–359.
- **26.** Gaspar R, Pereira L, Neto J M. Ecological reference conditions and quality states of marine macroalgae sensu Water Framework Directive: An example from the intertidal rocky shores of the Portuguese coastal waters. Ecol Indic. 2012; 19: 24–38.
- Guinda X, Juanes J A, Puente A, Echavarri-Erasun B. Spatial distribution pattern analysis of subtidal macroalgae assemblages by a non-destructive rapid assessment method. J Sea Res. 2012; 67: 34–43.
- Neto J M, Gaspar R, Pereira L, Marques J C. Marine Macroalgae Assessment Tool (MarMAT) for intertidal rocky shores. Quality assessment under the scope of the European Water Framework Directive. Ecol Indic. 2012; 19: 39–47.

- Krause-Jensen D, Greve T M, Nielsen K. Eelgrass as a bioindicator under the European Water Framework Directive. Water Resources Management. 2005; 19: 63–75.
- Foden J. Assessment metrics for littoral seagrass under the European Water Framework Directive; outcomes of UK intercalibration with the Netherlands. Hydrobiologia. 2007; 579: 187–197.
- **31.** Foden J, Brazier D P. Angiosperms (seagrass) within the EU Water Framework Directive: A UK perspective. Mar Pollut Bull. 2007; 55: 181–195. https://doi.org/10.1016/j.marpolbul.2006.08.021 PMID: 17027036
- 32. Gobert S, Sartoretto S, Rico-Raimondino V, Andral B, Chery A, Lejeune P, Boissery P. Assessment of the ecological status of Mediterranean French coastal waters as required by the Water Framework Directive using the Posidonia oceanica Rapid Easy Index: PREI. Mar Pollut Bull 2009; 58(11): 1727– 1733. https://doi.org/10.1016/j.marpolbul.2009.06.012 PMID: 19700176
- Deter J, Descamp P, Ballesta L, Boissery P, Holon F. A preliminary study toward an index based on coralligenous assemblages for the ecological status assessment of Mediterranean French coastal waters. Ecol Indic. 2012; 20: 345–352.
- 34. Canovas-Molina A, Montefalcone M, Bavestrello G, Cau A, Bianchi C N, Morri C, et al. A new ecological index for the status of mesophotic megabenthic assemblages in the Mediterranean based on ROV photography and video footage. Cont Shelf Res. 2016; 121: 13–20.
- Ferrigno F, Russo G F, Sandulli R. Coralligenous Bioconstructions Quality Index (CBQI): a synthetic indicator to assess the status of different types of coralligenous hábitats. Ecol Indic. 2017; 82: 271–279.
- **36.** Piazzi L, Gennaro P, Cecchi E, Serena F, Bianchi C N, Morri C, et al. Integration of ESCA index through the use of sessile invertebrates. Sci Mar. 2017; 81–2: 283–290.
- Sartoretto S, Schohn T, Nike Bianchi C, Morri C, Garrabou J, Ballesteros E, et al. An integrated method to evaluate and monitor the conservation state of coralligenous habitats: The INDEX-COR approach. Mar Pollut Bull. 2017; 120: 222–231. <u>https://doi.org/10.1016/j.marpolbul.2017.05.020</u> PMID: 28521933
- Enrichetti F, Bo M, Morri C, Montefalcone M, Toma M, Bavestrello G, et al. Assessing the environmental status of temperate mesophotic reefs: A new, integrated methodological approach. Ecol Indic. 2019; 102: 218–229.
- Piazzi L, Cecchi E, Gennaro P, Morri C, Montefalcone M, Serena F, et al. STAR: An integrated and standardized procedure to evaluate the ecological status of coralligenous reefs. Aquat Conserv. 2019; 29: 189–201.
- Carballo J L, Naranjo S A, García-Gómez J C. Use of marine sponges as stress indicators in marine ecosystems at Algeciras Bay (southern Iberian Peninsula). Mar Ecol Prog Ser. 1996; 135: 109–122.
- Naranjo S A, Carballo J L, Garcia-Gómez J C. Effects of environmental stress on ascidian populations in Algeciras Bay (southern Spain). Possible marine bioindicators? Mar Ecol Prog Ser. 1996; 144: 119–131.
- **42.** García-Gómez J C. Biota Litoral y Vigilancia Ambiental en las Áreas Marinas Protegidas. Junta de Andalucía, Consejería de Medio Ambiente. 2007.
- 43. García-Gómez J C. A guide on environmental monitoring of rocky seabeds in Mediterranean Marine Protected Areas and surrounding zones. Marine Biology Laboratory, Department of Zoology, Faculty of Biology, University of Seville. R+D+I Biological Research Area, Seville Aquarium. Ed. RAC/SPA– MedMPAnet Project, Tunis. 2015. http://www.rac-spa.org/sites/default/files/doc_medmpanet/final_ docs_regional/78.guide_suivi_amp_med_en.pdf (Available for free download).
- 44. Garrabou J, Harmelin J G. A 20-year study on life-history traits of a harvested long-lived temperate coral in the NW Mediterranean: insights into conservation and management needs. Journal of Animal Ecology. 2002; 71: 966–978.
- 45. Coma R, Serrano E, Linares C, Zabala M, Ribes M. Effect of a severe storm event on the mortality rate of the gorgonian *Paramuricea* clavata on the Medes Islands Marine Reserve and the nearby Montgrí coast. In: Mateo M. A. and García-Rubies T. (Eds.), Assessment of the ecological impact of the extreme storm of Sant Esteve s Day (26 December 2008) on the littoral ecosystems of the north Mediterranean Spanish coasts. Final Report (PIEC 200430E599). Centro de Estudios Avanzados de Blanes, Consejo Superior de Investigaciones Científicas, Blanes, Spain, 2012; 67–78.
- Bertolino M, Betti F, Bo M, Cattaneo-Vietti R, Pansini M, Romero J, et al. (2016). Changes and stability of a Mediterranean hard bottom benthic community over 25 years. J Mar Biol Assoc U K. 2016; 96(2): 341–350.
- Betti F, Bavestrello G, Bo M, Asnaghi V, Chiantore M, Bava S et al. Over 10 years of variation in Mediterranean reef benthic communities. Marine Ecology. 2017; 38(3): e12439.
- **48.** Montefalcone M, Morri C, Bianchi C N, Bavestrello G, Piazzi L. The two facets of species sensitivity: Stress and disturbance on coralligenous assemblages in space and time. Mar Pollut Bull. 2017; 117: 229–238. https://doi.org/10.1016/j.marpolbul.2017.01.072 PMID: 28185652

- 49. Bianchi CN, Cocito S, Diviacco G, Dondi N, Fratangeli F, Montefalcone M, et al. The park never born: Outcome of a quarter of a century of inaction on the sea-floor integrity of a proposed but not established Marine Protected Area. Aquat Conserv. 2018; 28(5): 1209–1228.
- Garrabou J. Life-history traits of Alcyonium acaule and Parazoanthus axinellae (Cnidaria, Anthozoa), with emphasis on growth. Mar Ecol Prog Ser. 1999; 178: 193–204.
- Garrabou J, Ballesteros E. Growth of Mesophyllum alternans and *Lithophyllum frondosum* (Corallinales, Rhodophyta) in the northwestern Mediterranean. Eur J Phycol. 2000; 35: 1–10.
- Glynn P W. Coral Reef Bleaching in the 1980s and Possible Connections with Global Warming. TREE. 1991; 6–6: 175–178.
- Walther G R, Post E, Convey P, Menzel A, Parmesan C, Trevor J C, et al. Ecological responses to recent climate change. Nature. 2002; 412: 389–395.
- Baker A C, Glynn P W, Riegl B. Climate change and coral reef bleaching: An ecological assessment of long-term impacts, recovery trends and future outlook. Estuar Coast Shelf Sci. 2008; 80: 435–471.
- 55. Jordà G, Marbà N, Duarte C M. Mediterranean seagrass vulnerable to regional climate warming. Nat Clim Chang. 2012; 2: 821–824.
- Dukes J S, Mooney H. Does global change increase the success of biological invaders? Trends Ecol Evol. 1999; 14: 135–139. https://doi.org/10.1016/s0169-5347(98)01554-7 PMID: 10322518
- Occhipinti-Ambrogi A. Global change and marine communities: Alien species and climate change. Mar Pollut Bull. 2007; 55: 342–352. https://doi.org/10.1016/j.marpolbul.2006.11.014 PMID: 17239404
- **58.** Soto C. The potential impacts of global climate change on marine protected areas. Rev Fish Biol Fish. 2002; 11: 181–195.
- Carvalho S B, Brito J C, Crespo E G, Watts M E, Possingham H P. Conservation planning under climate change: Toward accounting for uncertainty in predicted species distributions to increase confidence in conservation investments in space and time. Biol Conserv. 2011; 144: 2020–2030.
- 60. Hawkins S J. Marine conservation in a rapidly changing world. Aquat Conserv. 2012; 22: 281–287.
- 61. Romano J C, Bensoussan N, Younes W A N, Arlhac D. Anomalie thermique dans les eaux du golfe de Marseille durant l'été 1999. Une explication partielle de la mortalité d'invertébrés fixés? Comptes Rendus de l'Académie des Sciences de Paris, Sciences de la vie / Life Sciences. 2000; 323: 415– 427.
- Linares C, Coma R, Diaz D, Zabala M, Hereu B, Dantart L. Immediate and delayed effects of a mass mortality event on gorgonian population dynamics and benthic community structure in the NW Mediterranean Sea. Mar Ecol Prog Ser. 2005; 305:127–137.
- Bensoussan N, Romano J C, Harmelin J G, Garrabou J. High resolution characterization of northwest Mediterranean coastal waters thermal regimes: To better understand responses of benthic communities to climate change. Estuar Coast Shelf Sci. 2010; 87: 431–441.
- Foster M S, Harrold C, Hardin D D. Point vs. photo quadrat estimates of the cover of sessile marine organisms. J Exp Mar Bio Ecol. 1991; 146: 193–203.
- Roberts D E, Fitzhenry S R, Kennelly S J. Quantifying subtidal macrobenthic assemblages on hard substrata using a jump camera method. J Exp Mar Bio Ecol. 1994; 177: 157–170.
- Garrabou J, Ballestero E, Zabala M. Structure and Dynamics of North-western Mediterranean Rocky Benthic Communities along a Depth Gradient. Estuar Coast Shelf Sci. 2002; 55: 493–508.
- Leujak W, Ordmon R F G. Comparative accuracy and efficiency of six coral community survey methods. J Exp Mar Bio Ecol. 2007; 351-1-2: 168–187.
- Van Rein H, Brown C J, Schoeman D S, Quinn R, Breen J. Fixed-station monitoring of a harbour wall community: the utility of low-cost photomosaics and scuba on hard-substrata. Aquat Conserv. 2011; 21: 690–703.
- **69.** Van Rein H, Schoeman D S, Brown C J, Quinn R, Breen J. Development of benthic monitoring methods using photoquadrats and scuba on heterogeneous hard-substrata: a boulder-slope community case study. Aquat Conserv. 2011; 21–7: 676–689.
- 70. Van Rein H, Schoeman D S, Brown C J, Quinn R, Breen J. Development of low-cost image mosaics of hard-bottom sessile communities using SCUBA: comparisons of optical media and of proxy measures of community structure. J Mar Biol Assoc U.K. 2012; 92–1: 49–62.
- Parravicini V, Micheli F, Montefalcone M, Morri C, Villa E, Castellano M, et al. Conserving Biodiversity in a Human-Dominated World: Degradation of Marine Sessile Communities within a Protected Area with Conflicting Human Uses. PLoS One. 2013; 8–10: e75767.
- 72. Kipson S, Fourt M, Teixido N, Cebrian E, Casas E, Ballesteros E, et al. Rapid Biodiversity Assessment and Monitoring Method for Highly Diverse Benthic Communities: A Case Study of Mediterranean

Coralligenous Outcrops. PLoS One. 2011; 6(11): e27103. <u>https://doi.org/10.1371/journal.pone.</u> 0027103 PMID: 22073264

- Deter J, Descamp P, Boissery P, Ballesta L, Holon F. A rapid photographic method detects depth gradient in coralligenous assemblages. J Exp Mar Bio Ecol. 2012; 418–419: 75–82.
- 74. Gattia G, Montefalconea M, Rovere A, Parravicini C, Morria C, Albertelli G, et al. Seafloor integrity down the harbor waterfront: the coralligenous shoals off Vado Ligure (NW Mediterranean). Advances in Oceanography and Limnology. 2012; 3(1): 51–67.
- 75. Piazzi L, Bianchi CN, Cecchi E, Gatti G, Guala I, Morri C, et al. What's in an index? Comparing the ecological information provided by two indices to assess the status of coralligenous reefs in the NW Mediterranean. Sea Aquatic Conserv: Mar Freshw Ecosyst. 2017; 27:1091–1100.
- Alemany J. Puerto Bahía Algeciras. Entre dos mares y dos continentes. 100 años de historia. Lunwerg Eds. 2005.
- 77. Morales C. Caracterización de la calidad de sedimentos afectados por vertidos de petróleo: comparación entre casos de vertidos accidentales (impacto agudo) frente a derrames continuos (impacto crónico). Tesis Doctoral. Universidad de Cádiz, Facultad de Ciencias del Mar y Ambientales. 2007.
- 78. Consejería de Medio Ambiente de la Junta de Andalucía. Acuerdo específico entre la Consejería de Medio Ambiente y la Universidad de Sevilla para el seguimiento y la vigilancia ambiental de las comunidades rocosas intermareales y submareales del parque natural del Estrecho. Tercer informe parcial. 2009.
- 79. Consejería de Medio Ambiente y ordenación del territorio. Red de información ambiental de Andalucía. (http://www.juntadeandalucia.es/medioambiente/portal_web/rediam/documentos/docs/litoral/ Informe_resultados_sst.pdf).
- García-Gómez J, Sempere-Valverde J, González A, Martínez-Chacón M, Ponzone L, Sánchez-Moyano E, et al. From exotic to invasive in record time: the extreme impact of Rugulopteryx okamurae (dictyotales, ochrophyta) in the strait of Gibraltar. Science of The Total Environment. 704 (2020) 135408. https://doi.org/10.1016/j.scitotenv.2019.135408 PMID: 31836226
- Zibrowius H. The "southern" Astroides calycularis in the Pleistocene of the northern Mediterranean an indicator of climatic changes (Cnidaria, Scleractinia). Geobios. 1995; 28: 9–1.
- Benedetti-Cecchi L, Airoldi L, Abbiati M, Cinelli F. Exploring the causes of spatial variation in an assemblage of benthic invertebrates from a submarine cave with sulphur springs. J Exp Mar Bio Ecol. 1996; 208: 153–168.
- Linares C, Coma R, Garrabou J, Díaz D, Zabala M. Size distribution, density and disturbance in two Mediterranean gorgonians: *Paramuricea clavata* and *Eunicella singularis*. Journal of Applied Ecology. 2008; 45: 688–699.
- Cupido R, Cocito S, Barsanti M, Sgorbini S, Peirano A, Santangelo G. Unexpected long-term population dynamics in a canopy-forming gorgonian coral following mass mortality. Mar Ecol Prog Ser. 2009; 394: 95–200.
- Linares C, Doak D F. Forecasting the combined effects of disparate disturbances on the persistence of long-lived gorgonians: a case study of *Paramuricea clavata*. Mar Ecol Prog Ser, 2010; 402: 59–68.
- Terrón-Sigler A, Peñalver-Duque P, León-Muez D, Espinosa-Torre F. Spatio-temporal macrofaunal assemblages associated with the endangered orange coral Astroides calycularis (Scleractinia: Dendrophylliidae). Aquatic biology. 2014; 21: 143–154.
- Gatti G, Bianchi C N, Morri C, Montefalcone M, Sartoretto S. Coralligenous reefs state along anthropized coasts: Application and validation of the COARSE index, based on a rapid visual assessment (RVA) approach. Ecol Indic. 2015; 52: 567–576.
- Casas-Güell E, Teixidó N, Garrabou J, Cebrian E. Structure and biodiversity of coralligenous assemblages over broad spatial and temporal scales. Mar Biol. 2015; 162: 901–912.
- Dethier M N, Graham E S, Cohen S, Tear L M. Visual versus random-point percent cover estimations: 'objective' is not always better. Mar Ecol Prog Ser. 1993; 96: 93–100.
- Dietz H, Steinlein T. Determination of plant species cover by means of image analysis. J Veg Sci. 1996; 7: 131–136.
- Leonard G H, Clark R P. Point quadrat versus video transect estimates of the cover of benthic red algae. Mar Ecol Prog Ser. 1993; 101: 203–208.
- Sant N, Chappuis E, Rodríguez-Prieto C, Real M, Ballesteros E. Cost-benefit of three different methods for studying Mediterranean rocky benthic assemblages. Sci. Mar. 2017; 81–1: 129–138.
- J W. Significance test for sphericity of a normal n-variate distribution. Annals of Mathematical Statistics. 1940; 11; 204–209.
- Geisser S, Greenhouse S. An extension of Box's results on the use of the F distribution in multivariate analysis. Annals of Mathematical Statistics. 1958; 29: 885–891.

- Greenhouse S, Geisser S. On methods in the analysis of profile data. Psychometrika. 1959; 24: 95– 112.
- 96. Pardo A, Ruiz M A. Análisis de varianza (III). El procedimiento Modelo Linear General: Medidas repetidas. In: Sánchez-González C. (Ed.). Análisis de datos con SPSS 13 Base. McGraw-Hill / Interamericana de España S. A. U., Madrid. 2005; 395–427.
- Kruskal J B, Wish M. Multidimensional Scaling. Sage University Paper Series on Quantitative Applications in the Social Sciences, Sage Publications, Newbury Park. 1978; No. 07–011.
- 98. Clarke K R, Gorley R N. PRIMER v6: User manual/tutorial. PRIMER-E Ltd. Plymouth, UK. 2006.
- Clarke K R, Somerfield P J, Airoldi L, Warwick R M. Exploring interactions by second-stage community analyses. J Exp Mar Bio Ecol. 2006; 338: 179–192.
- Anderson M J, Gorley R N, Clarke K R. PERMANOVA + for PRIMER. Guide to software and statistical methods. Plymouth, UK: PRIMER-E Ltd. 2008.
- 101. Anderson M. J. and Walsh D. C. I. PERMANOVA, ANOSIM and the Mantel test in the face of heterogeneous dispersions: What null hypothesis are you testing? Ecol Monogr. 2013; 83: 557–574.
- Lundalv T, Larsson C S, Axelsson L. Long-term trends in algal-dominated rocky subtidal communities on the Swedish west coast—a transitional system? Hydrobiologica. 1986; 142: 81–95.
- Lanyon J M, Marsh H. Temporal changes in the abundance of some tropical intertidal seagrasses in North Queensland. Aquat Bot. 1995; 49: 217–237.
- Hill J, Wilkinson C. Methods for Ecological Monitoring of Coral Reefs. Australian Institute of Marine Science: Townsville. 2004.
- 105. Molloy P P, Evanson M, Nellas A C, Rist J L, Marcus J E, Koldewey H J et al. How much sampling does it take to detect trends in coral-reef habitat using photoquadrat surveys? Aquat Conserv. 2013; 23: 820–837.
- Cattaneo-Vietti R, Albertelli G, Bavestrello G, Bianchi C N, Cerrano C, Chiantore, et al. Can rock composition affect sublittoral epibenthic communities? PSZN: Marine Ecology. 2002; 23: 65–77.
- 107. Bianchi C N, Pronzato R, Cattaneo-Vietti R, Benedetti-Cecchi L, Morri C, Pansini M, et al. Mediterranean marine benthos: a manual of methods for its sampling and study. Hard bottoms. Biol Mar Mediterr. 2004; 11: 85–215.
- Benedetti-Cecchi L, Airoldi L, Abbiati M, Cinelli F. Estimating the abundance of benthic invertebrates: a comparison of procedures and variability between observers. Mar Ecol Prog Ser. 1996; 138: 93– 101.
- 109. Fraschetti S, Bianchi C N, Terlizzi A, Fanelli G, Morri C, Boero F. Spatial variability and human disturbance in shallow subtidal hard substrate assemblages: a regional approach. Mar Ecol Prog Ser. 2001; 212: 1–12.
- Bevilacqua S, Terlizzi A, Fraschetti S, Russo G F, Boero F. Mitigating human disturbance: can protection influence trajectories of recovery in benthic assemblages? Journal of Animal Ecology. 2006; 75: 908–920. https://doi.org/10.1111/j.1365-2656.2006.01108.x PMID: 17009754
- 111. Guidetti P, Bianchi C N, Chiantore M, Schiaparelli S, Morri C, Cattaneo-Vietti R. Living on the rocks: substrate mineralogy and the structure of subtidal rocky substrate communities in the Mediterranean Sea. Mar Ecol Prog Ser. 2004; 274: 57–68.
- 112. Bussotti S, Terlizzi A, Fraschetti S, Belmonte G, Boero F. Spatial and temporal variability of sessile benthos in shallow Mediterranean marine caves. Mar Ecol Prog Ser. 2006; 325: 109–119.
- 113. Bohnsack J A. Photographic quantitative sampling of hard-bottom benthic communities. Bulletin of Marine Science. 1979; 29: 242–252.
- **114.** Martí R, Uriz M J, Ballesteros E, Turon X. Seasonal variation in the structure of three Mediterranean algal communities in various light conditions. Estuar Coast Shelf Sci. 2005; 64: 613–622.
- Whorff J S, Grifting L. A video recording and analysis system used to sample intertidal communities. J Exp Mar Bio Ecol. 1992; 160: 1–12.
- Magorrian B H, Service M. Analysis of Underwater Visual Data to Identify the Impact of Physical Disturbance on Horse Mussel (*Modiolus modiolus*) Beds. Mar Pollut Bull. 1998; 36: 354–359.
- Cabaitan P C, Licuanan W Y, Gomez E D. Comparison between videographic and photographic methods in assessing coral reef benthic communities. Sci Diliman. 2007; 19: 7–13.
- 118. Jackson J B C. Adaptation and diversity of reef corals. Bioscience. 1991; 41: 472–482.
- **119.** Witman J D, Grange K R. Links between rain, salinity, and predation in a rocky subtidal community. Ecology. 1998; 79: 2429–2447.
- Ballesteros E. Mediterranean coralligenous assemblages: a synthesis of present knowledge. Oceanography and Marine Biology: An Annual Review. 2006 44: 123–195.

- 121. Coma R, Linares C, Ribes M, Diaz D, Garrabou J, Ballesteros E. Consequences of a mass mortality in populations of Eunicella singularis (Cnidaria: Octocorallia) in Menorca (NW Mediterranean). Marine Ecology Progress Series. 2006; 327: 51–60.
- 122. Garrabou J, Coma R, Bensoussan N, Bally M, Chevaldonné P, Cigliano M, et al. Mass mortality in Northwestern Mediterranean rocky benthic communities: effects of the 2003 heat wave. Glob Chang Biol. 2009; 15: 1090–1103.
- 123. Bianchi CN, Azzola A, Bertolino M, Betti F, Bo M, CattaneoVietti R et al. Consequences of the marine climate and ecosystem shift of the 1980-90s on the Ligurian Sea biodiversity (NW Mediterranean). Eur Zool J. 2019; 86:1, 458–487. https://doi.org/10.1080/24750263.2019.1687765
- 124. Bianchi CN, Azzola A, Parravicini V, Peirano A, Morri C Montefalcone M. Abrupt Change in a Subtidal Rocky Reef Community Coincided with a Rapid Acceleration of Sea Water Warming. Diversity. 2019; 11, 215. https://doi.org/10.3390/d11110215
- Laborel J, Laborel-Deguen F. Sea-level indicators, biologic. In: Encyclopedia of Coastal Science (Ed. Schwartz M.), Wiley, New York. 2005; 833–834.
- 126. Rovere A, Antonioli F, Bianchi C N. Fixed biological indicators. In: Handbook of Sea-Level Research (Eds. Shennan I., Long A. J. and Horton B. P.). John Wiley & Sons, Ltd, Chichester, UK. 2015; 268–280.
- **127.** Di Franco A, Milazzo M, Baiata P, Tomasello A, Chemello R. Scuba diver behaviour and its effects on the biota of a Mediterranean marine protected area. Environ Conserv. 2009; 36: 32–40.
- **128.** Terrón-Sigler A, León-Muez D, Peñalver-Duque P, Espinosa-Torre F. The effects of SCUBA diving on the endemic Mediterranean coral *Astroides calycularis*. Ocean Coast Manag. 2016; 122: 1–8.
- 129. Urkiaga-Alberdi J, Pagola-Carte S, Saiz-Salinas J I. Reducing effort in the use of benthic bioindicators. Acta Oecologica. 1999; 20: 489–497.
- Sebens K P. Habitat structure and community dynamics in marine benthic systems. In: Bell S. S., McCoy E. D., Mushinsky H. R. (Eds) Habitat structure. Chapman & Hall, London. 1991; 211–234.
- 131. Orejas C, López-Gónzalez P J. Gorgonias y formaciones coralinas. Arquitectos de nuestros mares. In: Martínez R. and Cornejo J. M. (Eds.). Mares de España, Secretaría General del Mar, Dirección General de Sostenibilidad de la Costa y el Mar, Ministerio de Medio Ambiente y Medio Rural y Marino, Madrid. 2008.
- 132. Piazzi L, Balata D, Cecchi E, Gennaro P, Serena F. Effectiveness of different investigation procedures in detecting anthropogenic impacts on coralligenous assemblages. Sci. Mar. 2014; 78–3: 319–328.
- 133. Mistri M, Ceccherelli V U. Growth and secondary production of the Mediterranean gorgonian *Paramuricea clavata*. Mar Ecol Prog Ser. 1994; 103: 291–296.
- 134. Weinbauer M G, Velimirov B. Biomass and secondary production of the temperate gorgonian coral *Eunicella cavolini* (Coelenterata: Octocorallia). Mar Ecol Prog Ser. 1995; 121: 211–216.
- 135. Coma R, Ribes M, Zabala M, Gili J. M. Growth in a Modular Colonial Marine Invertebrate. Estuar Coast Shelf Sci. 1998; 47: 459–470.
- 136. Perez T, Garrabou J, Sartoretto S, Harmelin J G, Francour P, Vacelet J. Mortalité massive d'invertébrés marins: un événement sans précédent en Méditerranée nord-occidentale. Comptes Rendus de l Académie des Sciences de Paris, Sciences de la Vie. 2000; 323: 853–865.
- Bianchi C N. Biodiversity issues for the forthcoming tropical Mediterranean Sea. Hydrobiologia. 2007; 580: 7–21.
- Lejeusne C, Chevaldonne P, Pergent-Martini C, Boudouresque C F, Pérez T. Climate change effects on a miniature ocean: the highly diverse, highly impacted Mediterranean Sea. Trends Ecol Evol. 2010; 25–4: 250–260.
- **139.** Peckol P, Searles R B. Temporal and Spatial Patterns of Growth and Survival of Invertebrate and Algal Populations of a North Carolina Continental Shelf Community. Estuar Coast Shelf Sci. 1984; 18: 133–143.
- 140. True M. A. Étude quantitative de quatre peuplements sciaphiles sur substrat rocheux dans la région marseillaise. Bulletin Institute Océanographique de Monaco. 1970; 1410: 1–48.
- 141. Montefalcone M, Morri C, Bianchi C N, Bavestrello G, Piazzi L. The two facets of species sensitivity: Stress and disturbance on coralligenous assemblages in space and time. Mar Pollut Bull. 2017; 117: 229–238. https://doi.org/10.1016/j.marpolbul.2017.01.072 PMID: 28185652
- 142. Gatti G, Bianchi C N, Montefalcone M, Venturini S, Diviacco G, Morri C. Observational information on a temperate reef community helps understanding the marine climate and ecosystem shift of the 1980–90s. Mar Pollut Bull. 2017; 114: 528–538. https://doi.org/10.1016/j.marpolbul.2016.10.022 PMID: 27743657
- 143. Kipson S, Fourt M, Teixidó N, Cebrian E, Casas E, Ballesteros E, et al. Rapid Biodiversity Assessment and Monitoring Method for Highly Diverse Benthic Communities: A Case Study of Mediterranean Coralligenous Outcrops. PLoS One. 2011; 6–11: e27103.

- 144. Weinberg S. The minimal area problem in invertebrate communities of Mediterranean rocky substrata. Mar Biol. 1978; 49: 33–40.
- 145. Gili J M, Ballesteros E. Structure of cnidarian populations in Mediterranean sublittoral benthic communities as a result of adaptation to different environmental conditions. Oecologia Aquatica. 1991; 10: 243–254.
- 146. De Biasi A M, Bianchi C N, Aliani S, Cocito S, Peirano A, Dando P, et al. Epibenthic communities in a marine shallow area with hydrothermal vents. Chemistry and Ecology. 2004; 20: 89–105.
- 147. Parravicini V, Morri C, Ciribilli G, Montefalcone M, Albertelli G, Bianchi C N. Size matters more than method: Visual quadrats vs photography in measuring human impact on Mediterranean rocky reef communities. Estuar Coast Shelf Sci. 2009; 81: 359–367.
- 148. Gerovasileiou V, Dailianis T, Panteri E, Michalakis M, Gatti G, Sini M, et al. CIGESMED for divers: Establishing a citizen science initiative for the mapping and monitoring of coralligenous assemblages in the Mediterranean Sea. Biodivers Data J. 2016; 4: e8692.