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Long-term nitrogen loading alleviates phosphorus limitation in terrestrial ecosystems

# **Running title: Prolonged N loading eases P limitation**

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Abstract Increased human-derived nitrogen (N) deposition to terrestrial ecosystems has resulted in widespread phosphorus (P) limitation of net primary productivity. However, it remains unclear if and how N-induced P limitation varies over time. Soil extracellular phosphatases catalyze the hydrolysis of P from soil organic matter, an important adaptive mechanism for ecosystems to cope with N-induced P limitation. Here we show, using a meta-analysis of 140 studies and 668 observations worldwide, that N stimulation of soil phosphatase activity diminishes over time. Whereas short-term N loading ( $\leq$ 5 yr) significantly increased soil phosphatase activity by 28%, long-term N loading had no significant effect. Nitrogen loading did not affect soil available P and total P content in either short- or long-term studies. Together, these results suggest that N-induced P limitation in ecosystems is prevented in the long-term through the initial stimulation of soil phosphatase activity, thereby securing P supply to support plant growth. Our results suggest that increases in terrestrial carbon uptake due to ongoing anthropogenic N loading may be greater than previously thought.

**Key words** Phosphorus limitation, Nitrogen addition, Soil phosphatase activity, Soil nitrogen content, Soil phosphorus content, Soil pH, Microbial biomass, Nutrient stoichiometry balance.

### 1. Introduction

Humans have doubled nitrogen (N) inputs into terrestrial ecosystems over the past century by burning fossil fuels and using artificial fertilizers (Galloway *et al.*, 2008; Davidson 2009). Nitrogen can enter natural ecosystems through multiple routes, whereas phosphorus (P) derives mostly from deposition of mineral aerosols and weathering (Thingstad *et al.*, 2005; Peñuelas *et al.*, 2013). Phosphorus inputs are typically small compared with N inputs, leading to a strong stoichiometric imbalance between N and P in unmanaged terrestrial ecosystems (Crowley *et al.*, 2012; Peñuelas *et al.*, 2013; Reinhard *et al.*, 2017).

Enhanced N loading generally increases plant growth, thereby potentially stimulating ecosystem C storage and mitigating climate change (LeBauer, & Treseder 2008; Schulte-Uebbing, & de Vries 2018). However, the imbalance between N and P inputs suggests that plant growth may gradually shift from N to P limitation over time (Elser *et al.*, 2007; Peñuelas *et al.*, 2013; Li *et al.*, 2016). Indeed, several reports show that P fertilization stimulates plant growth more strongly in ecosystems with elevated versus ambient N inputs (Elser *et al.*, 2007; Li *et al.*, 2016), suggesting a N-induced P-limitation of primary productivity. On the other hand, N-fertilization can stimulate plant growth in P-limited ecosystems (LeBauer, & Treseder 2008), and P-limitation of plant growth frequently fails to develop, even after decades of N-fertilization (Hyvönen *et al.*, 2008; Lambers *et al.*, 2008; Finzi 2009; Weand, Arthur, Lovett, Sikora, & Weathers, 2010; Crowley *et al.*, 2012; Tatariw *et al.*, 2018). These contrasting results suggest the existence of unknown mechanisms that can alleviate N-induced P limitation of plant growth over time.

Plants and soil microorganisms preferentially invest metabolic resources to acquire nutrients that limit their growth (Johnson, Wilson, Bowker, Wilson, & Miller, 2010; Bragg 2012; Marklein, & Houlton 2012). Soil phosphatases are enzymes produced by both plants and soil microorganisms to catalyze the hydrolysis of ester-phosphate bonds and phosphoric acid anhydrides, releasing orthophosphate that can be taken up across living cell membranes (Vance, Uhde Stone, & Allan, 2003; Margalef *et al.*, 2017; Liu, Chen, Chen, Guo, Li, 2020). The production of extracellular phosphatases is generally assumed to indicate P limitation of both plant and microbial growth

(Vitousek, Porder, Houlton, & Chadwick, 2010; Marklein, & Houlton 2012; Jian *et al.*, 2016). Shortterm experiments across a wide range of terrestrial ecosystems show that N loading stimulates soil phosphatase activity and accelerates P cycling (Marklein, & Houlton 2012; Jian *et al.*, 2016; Xiao, Chen, Jing, & Zhu, 2018). Yet it remains unclear whether N stimulation of soil phosphatase activity is persistent or temporally dynamic. To address this knowledge gap, we compiled a database of the effects of N loading on soil phosphatase activity consisting of 668 observations from 140 studies (Fig. S1 and Table S1). We synthesized these data using meta-analysis, specifically focusing on how the response of soil phosphatase activity to N loading changes over time.

#### 2. Materials and methods

#### 2.1. Data collection

We searched for peer-reviewed articles published before December 2019 (i.e., ~1900 to 2019) using Web of Science (http://apps.webofknowledge.com/), Google Scholar (http://scholar.google.com/) and China National Knowledge Infrastructure (http://www.cnki.net/). The keywords used for the article selection were: (i) "nitrogen loading" or "nitrogen fertilization" or "nitrogen enrichment" or "nitrogen elevated" or "nitrogen deposition" and (ii) "phosphatase" or "acid phosphatase" or "alkaline phosphatase" and (iii) "terrestrial" or "soil" or "land".

Articles selected for this meta-analysis had to meet the following criteria: (1) vegetation, soil, and climatic parameters were similar for ambient and N loading treatments; (2) results came from field or common garden experiments; (3) N loading methods (rate, duration, form and frequency) were clearly reported; (4) the duration of the experiment was longer than one year; and (5) standard deviation (SD) and sample size were reported or could be calculated from the data presented in the publication. For factorial experiments, we only considered comparisons between control and treatments that differed solely in N loading (e.g., precipitation versus precipitation plus N loading). Measurements from different years from the same study site were included separately. We excluded articles with: (i) incomplete information on the study site (e.g., when the coordinates, climatic variables or ecosystem type could not be determined); (ii) ambiguous N loading methods (e.g., compost or slurry additions); (iii) missing information on study duration. When multiple measurements were taken within the same

year or growing season, we considered the measurement with the most accompanied ancillary variables or that from the peak of the growing season (Tatariw *et al.*, 2018). Based on these criteria, 140 articles and 668 data points were selected (Fig. S1); the original dataset is available from Figshare (https://figshare.com/s/8ebfabf329c09de0277e). When results from published sources were presented graphically, we used Grapher<sup>™</sup> to digitize the data (http://www.goldensoftware.com/products/grapher).

#### 2.2 Climatic and environmental variables

For each study in our dataset, we recorded a wide range of environmental variables, including latitude (with a range from 4.12°S to 68.63°N), longitude (157.78°W to 129.18°E), elevation (4 to 3559 m), background N deposition (0.17 to 13.80 g N m<sup>-2</sup> yr<sup>-1</sup>), mean annual temperature (-8.8 to 27.7°C), mean annual precipitation (110 to 4500 mm), and vegetation type (cropland, grassland, forest, shrubland and wetland). If the required data were not reported in the selected articles or other articles published by the same research group, we contacted the corresponding author. Otherwise, we obtained mean annual temperature and precipitation from the WorldClim Database (Fick, & Hijmans 2017), background N deposition from the Global N deposition database (ORNL DAAC 2017), and vegetation types from the Whittaker Biome Diagram (Whittaker 1962).

#### 2.3 Nitrogen loading methods

Information on N loading rate (g N m<sup>-2</sup> yr<sup>-1</sup>), duration (year), frequency (times per year), and chemical forms (urea, NH<sub>4</sub>NO<sub>3</sub>, and mixed N [combined inorganic and organic N loading]) were tabulated for each study. To make our results comparable to other meta-analyses on enhanced N deposition, continuous variables were divided into different subgroups as in earlier studies (Liu, & Greaver 2010; Li *et al.*, 2016; Chen *et al.*, 2017). For example, N loading rate was grouped by <5, 5–15 and >15 g N m<sup>-2</sup> yr<sup>-1</sup> and N loading frequency by <4, 4–12 and >12 times per year. To assess temporal variation in treatment effects, we made a distinction between short-term (<5 years) and long-term ( $\geq$ 5 years) studies (Kuebbing *et al.*, 2018; Chen *et al.*, 2020). The cutoff of 5 year aligned with the large survey of long-term research in ecology and evolution by Kuebbing *et al.*, (2018).

### 2.4 Meta-analysis

We quantified the effect of N loadings on each variable by calculating the natural log of the response ratio ( $\ln R$ ), a metric commonly used in meta-analysis (Chen *et al.*, 2015; Hedges, Gurevitch, & Curtis, 1999):

$$\overline{\ln R} = \ln\left(\frac{\overline{X_N}}{\overline{X_C}}\right) = \ln\left(\overline{X_N}\right) - \ln\left(\overline{X_C}\right)$$
(1),

where  $\overline{X_N}$  and  $\overline{X_C}$  are the arithmetic mean value of the variables in the N loading and ambient treatments, respectively.

The effects of N loading on soil phosphatase activity were evaluated by mixed-effects models using the rma.mv function from the R package "*metafor*" (Viechtbauer 2010). "Study site" and "observation" were considered as random effects in the meta-analysis, because several study sites contributed more than one effect size (Chen *et al.*, 2018a). Effect sizes were weighted by the inverse of the pooled variance ( $V_i$ ):

$$V_{\rm i} = \frac{SD_N^2}{n_N X_N^2} + \frac{SD_C^2}{n_C X_C^2}$$
(2)

where  $SD_N$  and  $SD_C$  are the standard deviations,  $n_N$  and  $n_C$  are the replicate numbers, and  $X_N$  and  $X_C$  are arithmetric values for the variables for N loading and ambient treatments, respectively. The results for the analyses on  $\ln R$  were back-transformed and reported as percentage change with N loading to ease interpretation. The effects of N loading were considered significant if the 95% confidence interval (CI) did not overlap with zero.

#### 2.5 Model selection and correlation analysis

Mixed-effects meta-regression model selection was adopted to identify the most important predictors of the effects of N loading on soil phosphatase activity using the "*glmulti*" package in R (Calcagno, & de Mazancourt 2010). The model selection was based on maximum likelihood estimation. The

importance of each predictor was computed as the sum of Akaike weights for models that included this predictor. A cut-off of 0.8 was set to differentiate between essential and non-essential predictor variables (Calcagno, & de Mazancourt 2010; Terrer, Vicca, Hungate, Phillips, & Prentice, 2016; Jiang *et al.*, 2019). To avoid possible artifacts associated with arbitrary category definitions, we included study duration, N loading rate and N loading frequency as continuous variables in the model selection analysis. We ran model-selection with all available pairwise predictors (i.e., latitude, longitude, elevation, mean annual temperature, mean annual precipitation, background N deposition, N loading methods [frequency, rate and duration, all as continuous variables], and vegetation type) because missing values were not allowed in the model selection analysis.

To explore the potential mechanisms underlying P limitation, we tabulated the following information from both ambient and N loading treatments from the original studies: aboveground biomass, soil pH, soil total N and P content, soil available P content, and soil microbial biomass. We checked whether these variables were correlated with treatment effects on phosphatase activity. Since most studies reported only a few of those variables, correlation analyses were separately conducted for each variable. When a variable correlated significantly with treatment effects on phosphatase activity, we repeated model selection with the subset of studies that included information on this variable.

#### 3. Results

Averaged across all studies, N loading enhanced soil phosphatase activity by 13% (95% CI, 8 to 18%, p < 0.001) (Fig. 1). This effect of N loading was consistent regardless of vegetation type (farmland, forest, grassland, shrubland and wetland) or the rate, duration and frequency of N loading (Figs. 2 and 3).

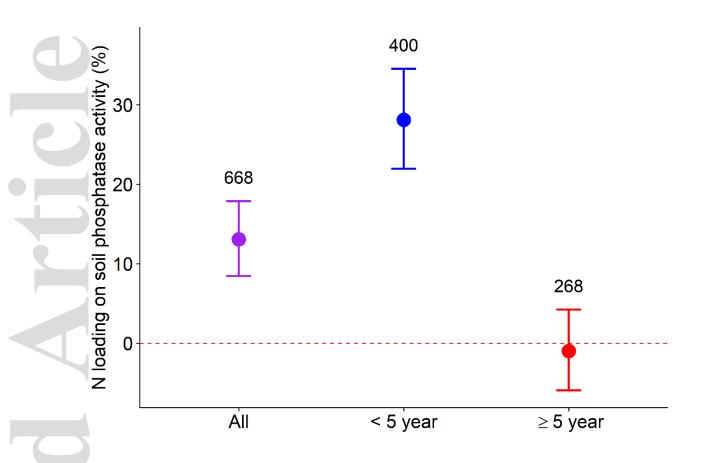


Figure 1. Effect of nitrogen (N) loading on soil phosphatase activity. Error bars represent bootstrap 95% confidence intervals. The sample sizes are shown above the error bars. Results are grouped by study duration for short- and long-term studies (<5 year and  $\geq$ 5 year).

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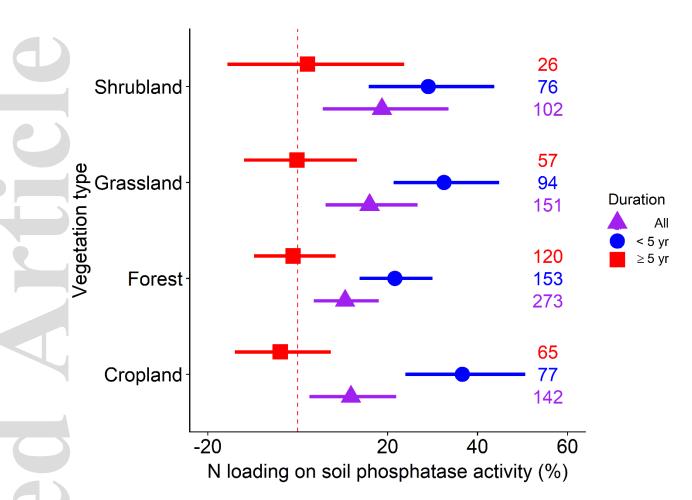


Figure 2. Effect of N loading on soil phosphatase activity for different vegetation types when studies are grouped by study duration. Error bars represent bootstrap 95% confidence intervals. The sample sizes are shown on the right hand of the error bars. Studies are grouped by study duration (<5 year and  $\geq$ 5 year).

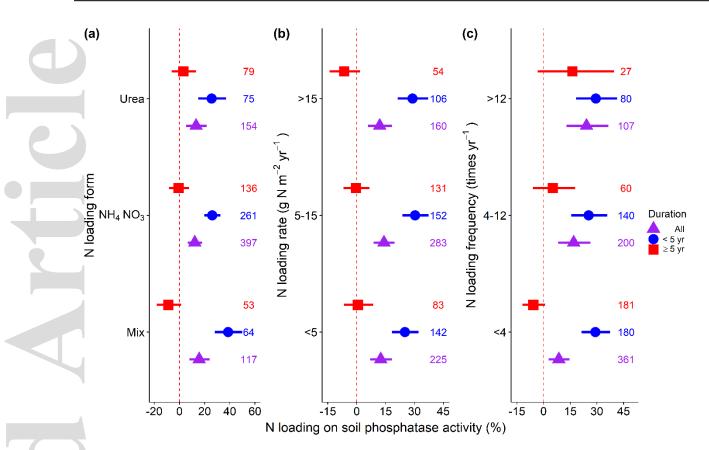
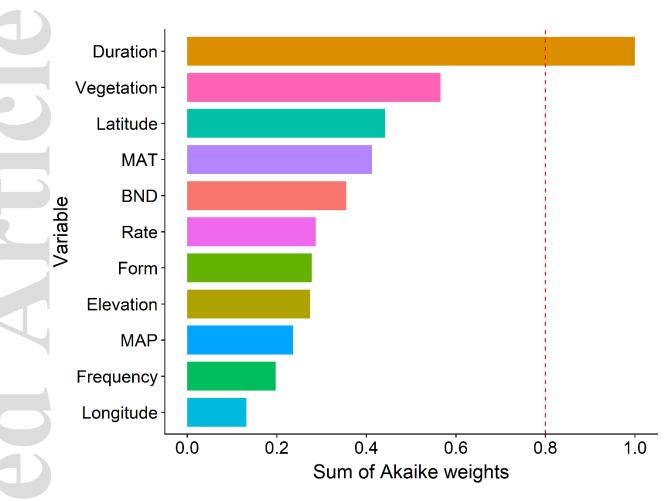


Figure 3. Effect of N loading on soil phosphatase activity for N loading (a) form, (b) rate and (c) frequency when studies are grouped by study duration. Error bars represent bootstrap 95% confidence intervals. The sample sizes are shown on the right hand of the error bars. Studies are grouped by study duration (<5 year and  $\geq$ 5 year).

Model selection analysis of soil phosphatase activity to N loading were best explained by study duration (Fig. 4). The high importance of study duration supported the removal of predictors related to climate and environmental variables (latitude, elevation, mean annual temperature, mean annual precipitation and background N deposition), vegetation types and other N loading methods (Fig. 4). Specifically, short-term (<5 year) N loading significantly increased soil phosphatase activity by 28% (p < 0.001), whereas long-term ( $\geq$ 5 year) N loading had no effect on soil phosphatase activity (p =0.707) (Fig. 1). This differential response of soil phosphatase activity to short- and long-term N loading was also found within study categories based on vegetation type and N loading method (Figs. 2 and 3).



**Figure 4. Model-averaged importance of the predictors of the effects of nitrogen (N) loading on soil phosphatase activity.** Importance is estimated from the sum of Akaike weights based on model selection analysis using corrected Akaike's Information Criteria. Cutoff is set at 0.8 to explore the most essential variables. MAT, mean annual temperature; BND, background N deposition; MAP, mean annual precipitation. Form, duration, rate, and frequency denote different N loading methods.

Increases in soil phosphatase activity positively correlated with soil N content (p < 0.001, Fig. S2A). When we repeated our model selection procedure for the subset of studies reporting soil N content, study duration remained the most important predictor of treatment effects on soil phosphatase activity (Fig. S2B). Nitrogen loading did not affect soil total P or available P content in either short-or long-term studies (Fig. S3A and B). The response of soil phosphatase activity was not correlated to

soil total P or available P content (Fig. S3C and D).

Nitrogen loading significantly increased aboveground biomass by 21% (95% CI, 13 to 29%, p < 0.001), with no difference for short- and long-term N loading studies (Fig. S4A). Nitrogen-stimulation of soil phosphatase activity was positively correlated with the responses of aboveground biomass (Fig. S4B). When we repeated the model selection analysis by including the response of aboveground biomass, we found that study duration and N loading rate were the most important predictors of treatment effects on soil phosphatase activity (Fig. S4C).

Nitrogen loading significantly decreased soil pH by 0.24 units (95% CI, 0.17 to 0.32, p < 0.001, Fig. S5A). The response of soil phosphatase activity to N loading did not directly relate to soil pH, but N-induced reductions in soil pH were positively correlated with N-induced changes in soil phosphatase activity (Fig. S5B, p < 0.001). When we repeated the model selection analysis by including soil pH in ambient treatment and N-induced changes in soil pH, study duration remained the most important predictor of treatment effects on soil phosphatase activity (Fig. S5C).

Across the dataset, N-induced changes in soil microbial biomass were positively correlated with treatment effects on soil phosphatase activity (Fig. S6A). When we limited the model selection analysis to studies that reported soil microbial biomass, study duration remained the most important predictor of the effects of N loading on soil phosphatase activity (Fig. S6B). In addition, short-term N loading significantly increased microbial specific phosphatase activity by 21% (p < 0.001), whereas long-term N loading had no effect (Fig. S7).

### 4. Discussion

#### 4.1 Stimulation of soil phosphatase activity with N loading

Our results indicate that N loading significantly increases soil phosphatase activity across a wide range of ecosystems (Fig. 2). Enhanced plant growth with N loading could drive increased demand for P, causing widespread plant and microbial P limitation (Elser *et al.*, 2007; Vitousek, Porder, Houlton, & Chadwick, 2010; Li *et al.*, 2016). This is supported by the positive relationship between N-stimulation of soil phosphatase activity and aboveground biomass (Fig. S4B). Plants and microbes are therefore investing energy and resources in phosphatase production, which increases the turnover rate of P-containing compounds (Plaxton, & Tran 2011; Richardson, & Simpson 2011; Bragg 2012). Phosphatases are N-rich molecules (Sinsabaugh *et al.*, 2008; Midgley, & Phillips 2016; Pinsonneault, Moore, & Roulet, 2016), and the additional N provides the building blocks to support plant and microbial production of these enzymes (Delgado-Baquerizo *et al.*, 2013).

### 4.2 Acclimation of soil phosphatase activity to prolonged N loading

Total and available soil P content were unaffected by N loading, even in long-term studies (Fig. S3A and B). Further, N-stimulation of soil phosphatase activity decreased over time (Fig. 1). These results suggest that ecosystems may prevent N-induced P-limitation by mechanisms that maintain P availability (discussed below), thereby weakening the initial N-stimulation of soil phosphatase activity over time.

First, enhanced soil phosphatase activity during initial stages of N loading can accelerate organic P mineralization (Vitousek, Porder, Houlton, & Chadwick, 2010; Marklein, & Houlton 2012; Jian *et al.*, 2016), making a portion of P bioavailable and conserved in plant biomass rather than occluded by clays and minerals (Crowley *et al.*, 2012). The P assimilated by plants can then be re-translocated and recycled with extended N loading, e.g., by redistributing P from senescing leaves to developing tissues (Bragg 2012).

Second, shifts in plant community composition could increase P use efficiency. Nitrogen-induced P limitation of plant growth has been observed for individual species, but is less likely for whole communities (Weand, Arthur, Lovett, Sikora, & Weathers, 2010; Turner, Brenes-Arguedas, & Condit, 2018), suggesting that individual species differ in their response to P-limitation. Species able to tolerate P-limitation or those with high P-recycling capabilities will outcompete other species, alleviating the N-induced P limitation of plant growth (Sundqvist, Liu, Giesler, & Wardle, 2014; Farrer, & Suding 2016). For example, arbuscular mycorrhizal-associated herbaceous species gained a competitive advantage over ectomycorrhizal-associated larch in response to enhanced N loading in a

temperate forest (Deng *et al.*, 2016). In addition, plants can mitigate P limitations for growth through adaptation strategies related to root morphology (Gilroy, & Jones 2000; Castrillo *et al.*, 2017; Li *et al.*, 2020), exudation of organic acids (Lambers, Raven, Shaver, & Smith, 2008), and reduced N demands (Maathuis 2009).

Third, N-induced microbial P limitation can increase microbial mobilization of soil P through enhanced expression of P-cycling genes or through changes in microbial community composition (Vance, Uhde Stone, & Allan, 2003; Carrara et al., 2018). For example, Chen et al., (2019) reported that long-term mineral N addition increased crop production, but decreased soil phosphatase activity. These contrasting responses were accompanied by considerable shifts in bacterial communities harboring *phoD* genes coding for PhoD alkaline phosphatases (e.g., *Stenotrophomonas* and Brevundimonas), showing that N addition favored bacteria with the capacity to mineralize recalcitrant organic P rather than stimulating soil phosphatase production (Fox, Kwapinski, Griffiths, & Schmalenberger, 2014; Chen et al., 2019). Similarly, N-induced changes in microbial community composition and physiology can help ecosystems adapt to P limitation (Jakobsen, Abbott, & Robson, 1992; Johnson, Wilson, Bowker, Wilson, & Miller, 2010; Tedersoo, & Bahram 2019; Wei et al., 2013). For example, arbuscular mycorrhizal fungi symbionts enhanced soil available P content, stimulated plant P absorption, and decreased the plant N:P ratio with N loading, which could help alleviate N-induced P limitation over time (Mei, Yang, Zhang, Zhang, & Guo, 2019; Wang et al., 2018). It should be noted though those responses of arbuscular mycorrhizae are ecosystem specific (Cusack et al., 2016; Treseder 2008; Wang et al., 2018; Sekaran, McCoy, Kumar, & Subramanian, 2019), and that N loading typically decreases the abundance of arbuscular mycorrhizae (Treseder 2004). As such, the potential of mycorrhizae to alleviate P limitation with N loading is still unclear.

Fourth, N-induced changes in soil pH could partly attenuate N-induced P limitation. Soil acidification after N loading could induce a selective pressure that drives the evolution of the soil microbial community towards preventing the N-induced P limitation of plant and microbial growth (Chen *et al.*, 2019; Vitousek, Porder, Houlton, & Chadwick, 2010). In addition, N-induced soil acidification can reduce plant growth, soil C input rates, and microbial activity (Carrara *et al.*, 2018;

Chen *et al.*, 2018b), which could gradually reduce plant and microbial P requirements over time. Reductions in soil pH with long-term N loading could also help mobilize P from secondary minerals of aluminum, calcium and iron, alleviating N-induced P limitation over time (Sherman, Fernandez, Norton, Ohno, & Rustad, 2006; SanClements, Fernandez, & Norton, 2010). However, because this latter mechanism does not affect the P requirements of plants or microbes, it will not alleviate P limitation indefinitely.

Finally, prolonged N loading generally reduces belowground C allocation by plants, including decreased production of fine root biomass and root exudates (Carrara *et al.*, 2018; Song *et al.*, 2019). These responses can be ascribed to either N saturation over time or the accumulated deleterious effects on plant and microbial growth with long-term N loading (Aber *et al.*, 1998; Treseder 2008). Thus, it is likely that there is an ecosystem-specific threshold, above which prolonged N inputs do not exacerbate the P limitation, but instead reduce belowground C allocation (Tian, Wang, Sun, & Niu, 2016). Reduced belowground C allocation by plants may suppress microbial enzyme production due to C and energy limitations of microbial metabolism and growth over time (Mooshammer, Wanek, Zechmeister-Boltenstern, & Richter, 2014; Soong *et al.*, 2019). This explanation is supported by our finding that N loading increases microbial specific phosphatase expression in the short-term, but not in the long-term. Similarly, limitation of other nutrients induced by long-term N loading (e.g., potassium, calcium and magnesium) could also constrain plant and microbial metabolism and gradually alleviate N-induced plant and microbial P limitation (Wright *et al.*, 2011; Crowley *et al.*, 2012).

### 4.3 Soil N regulation of soil phosphatase activity with N loading

N loading increased soil phosphatase activity more strongly in ecosystems with greater soil N content (Fig. S2), suggesting that plants and soil microorganisms are more prone to P-limitation in N-rich ecosystems. Soil phosphatase activity can be enhanced by N loading when plant growth and microbial metabolism have shifted from N limitation to P limitation, particularly once N limitation of plant growth has been lifted (Finzi 2009; Li *et al.*, 2016). Since phosphatase production imposes high N costs, N loading could stimulate soil phosphatase activity more strongly in ecosystems already with

relatively high soil N content (Mineau, Fatemi, Fernandez, & Simon, 2014; Ratliff, & Fisk 2016). In ecosystems with relatively low soil N content, plant growth and microbial metabolism may remain N limited, even with enhanced N inputs (Jones, Clode, Kilburn, Stockdale, & Murphy, 2013; Fatemi, Fernandez, Simon, & Dail, 2016; Yokoyama, Imai, & Kitayama, 2017). Under these conditions, N loading would primarily support plant and microbial growth rather than the production of phosphatases (Kuzyakov, & Xu 2013; Pii *et al.*, 2015).

### 4.4 Implications

Our analysis confirms numerous studies showing that N loading leads to P limitation of both plant growth and microbial activity across a wide range of ecosystems (Elser *et al.*, 2007; Marklein, & Houlton 2012; Li *et al.*, 2016). However, our findings underline the importance of ecosystem processes leading to long-term acclimation of P limitation. These processes have not yet been adequately explored, and ignoring them may lead to overestimating the effect of P limitation on ecosystem function. For instance, some model simulations suggest that future P limitation of plant growth will turn ecosystems into net CO<sub>2</sub> sources by the end of this century (Wieder, Cleveland, Smith, & Todd-Brown, 2015; Sun *et al.*, 2017). However, these models do not consider plant and microbial P acquisition strategies, potentially leading to substantial uncertainties in model projections (Reed, Yang, & Thornton, 2015; Jiang, Caldararu, Zaehle, Ellsworth, & Medlyn, 2019). When plant and microbial P acquisition strategies are considered, ecosystems that experience long-term enhanced N deposition can still act as net CO<sub>2</sub> sinks (Fleischer *et al.*, 2019). Our results point in that direction and suggest that omitting plant and microbial P acquisition strategies may underestimate the potential of vegetation to mitigate climate change.

Phosphorus fertilization constitutes a challenge in modern agriculture, as continued P over application has led to enhanced P runoff from agricultural fields, causing eutrophication and hypoxia of lakes and marine ecosystems (Vance, Uhde Stone, & Allan, 2003; Conley *et al.*, 2009). Our results imply that plants and microorganisms will likely strengthen their P acquisition strategies due to N-induced P limitation, eventually leading to ecosystems being acclimated to N-induced P limitation. Indeed, soils store considerable amounts of organic P (Vance, Uhde Stone, & Allan, 2003; Vitousek,

Porder, Houlton, & Chadwick, 2010), which can potentially be taken up by plants and microorganisms through, for example, phosphatase catalyzed P mineralization (Marklein, & Houlton 2012; Margalef *et al.*, 2017). Thus, by managing plant and microbial P acquisition strategies, the need for P fertilizers could potentially be reduced (Vance, Uhde Stone, & Allan, 2003; Lambers, Raven, Shaver, & Smith, 2008; Menezes-Blackburn *et al.*, 2018). Exploiting and applying plant and microbial P acquisition pathways with N loading is a top research priority for developing nutrient-smart and sustainable agricultural systems.

In summary, our synthesis indicates that N-induced increases in soil phosphatase activity diminish over time, even when considering a wide range of climatic, edaphic, and experimental determinants of phosphatase activity. Our results indicate progressive attenuation of P limitation with N loading, highlighting the plant- and microbial-mediated ecosystem acclimation to N-induced P limitation. Thus, our results suggest that P limitation of plant growth with chronic N loading is smaller than previously thought. Our findings reconcile current conflicting results on N-induced P limitation, and underline the importance of ecosystem acclimation strategies to nutrient imbalances. By exploring the plant and microbial mechanisms associated with ecosystem acclimation to N-induced P limitation, we may enhance plant nutrient use efficiency and improve model predictions of net primary production in a warmer and N-enriched world.

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**Data accessibility** The data associated with this paper is available from the online supplementary file or from the figshare (https://figshare.com/s/8ebfabf329c09de0277e).

**Supporting Information** Additional Supporting Information may be found online in the supporting information tab for this article.

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