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### *Population genetic structure of the European lobster (*Homarus gammarus*) in the Irish Sea and implications for the effectiveness of the first British marine protected area*

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*Published in:*  
Fisheries Research

*DOI:*  
[10.1016/j.fishres.2016.06.015](https://doi.org/10.1016/j.fishres.2016.06.015)

*Publication date:*  
2016

*Citation for published version (APA):*  
Watson, H., McKeown, N., Coscia, I., Wootton, E., & Ironside, J. (2016). Population genetic structure of the European lobster (*Homarus gammarus*) in the Irish Sea and implications for the effectiveness of the first British marine protected area. *Fisheries Research*, 183, 287-293. <https://doi.org/10.1016/j.fishres.2016.06.015>

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1 **Population genetic structure of the European lobster (*Homarus gammarus*) in the Irish**  
2 **Sea and implications for the effectiveness of the first British marine protected area**

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15

16 **Abstract**

17 Levels of self-recruitment within and connectivity among populations are key factors  
18 influencing marine population persistence and stock sustainability, as well as the  
19 effectiveness of spatially explicit management strategies such as Marine Protected Areas  
20 (MPAs). In the United Kingdom (UK), Lundy Island in the Bristol Channel was designated a  
21 No-Take Zone (NTZ) in 2003 and became the UK's first Marine Conservation Zone (MCZ)  
22 in 2009. This NTZ is expected to represent an additional resource for the sustainable  
23 management of the European lobster (*Homarus gammarus*) fishery. As the first step in a  
24 genetic monitoring program, this study aimed to investigate population genetic structure of  
25 lobster within and between the Irish Sea and Bristol Channel and in doing so to assess the  
26 functioning of the Lundy NTZ in the context of connectivity and other genetic parameters.  
27 Analysis of microsatellite data indicated that lobsters within the study area are genetically  
28 homogeneous and supports the view of a single panmictic population wherein the Lundy  
29 NTZ is highly connected. Levels of genetic variability were universally high with no  
30 evidence of differences for the NTZ. Furthermore, there was no evidence of recent genetic  
31 bottlenecks, and estimates of effective population sizes were infinitely large. The results  
32 suggest that if current management and breeding stock sizes are maintained genetic drift will  
33 not be strong enough to reduce neutral genetic diversity.

34 **Keywords:** Marine protected area; population genetics; sustainability; larval connectivity;  
35 recruitment; crustacean

36

## 37 **1. Introduction**

38           Spatial patterns of self-recruitment and connectivity are recognised as key factors  
39 shaping the dynamics of marine populations and how they respond to natural and/or human  
40 disturbances (Hastings & Botsford 2006). For harvested species, failure to identify  
41 independent (non-connected) population units can lead to local overfishing and ultimately,  
42 severe declines. Self-recruitment and connectivity also determine the efficacy of management  
43 strategies, such as marine protected areas, that are increasingly being implemented as tools to  
44 simultaneously achieve both fisheries management and biodiversity conservation objectives  
45 (McCook et al. 2009; Gaines et al. 2010; McCook et al. 2010). Specifically, the functioning  
46 of marine reserves is dependent on the degree to which MPA individuals contribute to  
47 populations within the MPA (Claudet et al. 2008; Fenberg et al. 2012), and through  
48 spillover/dispersal, to areas beyond reserve borders (Grüss et al. 2011a; Grüss et al. 2011b;  
49 Harrison et al. 2012). Connectivity thus influences the extent to which such MPAs may  
50 contribute recruits to fished areas, as well as other MPAs (Palumbi 2003).

51           In benthic-orientated marine species, which are often relatively sedentary as adults,  
52 connectivity is largely shaped by dispersal during a larval phase (Cowen et al. 2007); the  
53 sampling of which is logistically challenging. Combined with a generally poor understanding  
54 of the interactions between dispersal and oceanic features, the measurement of marine  
55 connectivity remains extremely difficult. In this sense, genetic tools are recognised as an  
56 extremely powerful resource to provide indirect estimates of connectivity, and as such to  
57 describe population dynamics and to predict, validate and quantify the ecological and  
58 economical success of MPAs (Christie et al. 2010b).

59           The European lobster, *Homarus gammarus*, is widely distributed in the northeast  
60 Atlantic, with its range extending from the Arctic Circle to Morocco, although it is not  
61 present in the Baltic Sea due to lowered salinity and temperature extremes (Triantafyllidis et

62 al. 2005). Population genetic studies of *H. gammarus* have been performed using  
63 microsatellite (Ferguson, unpublished data), mitochondrial DNA (mtDNA) (Triantafyllidis et  
64 al. 2005), and allozyme (Jørstad et al. 2005) markers but have to date focused on  
65 macrogeographical scales. While *H. gammarus* has a long pelagic larval duration offering the  
66 potential for extensive dispersal, genetic studies of other marine invertebrates indicate that  
67 local oceanography, and other physical and biological factors can contribute to fine scale  
68 genetic structuring and recruitment variability (Selkoe et al. 2006; Banks et al. 2007; Iacchei  
69 et al. 2013). Hence, finer scale genetic studies of *H. gammarus* are required to understand  
70 population structure at geographical scales more relevant to the species' fishery, especially in  
71 light of signatures of over-exploitation across Europe (Browne et al. 2001).

72 *H. gammarus* has been a socioeconomically important resource for Irish coastal  
73 communities for over a century (Browne et al. 2001). In this time there have been significant  
74 fluctuations in landings (>460%) and the relative importance of the fishery has changed with  
75 the development of ancillary crustacean fisheries (Browne et al. 2001). The species supports  
76 an important fishery in the Irish Sea where it is intensively harvested by both Irish and British  
77 fishers (Table S1). High levels of exploitation, by British fishers only, are also reported for  
78 the adjoining Bristol Channel (Table S1). A prominent feature of the Bristol Channel fishery  
79 is the marine protected area at Lundy Island (off the North Devon coast). This was first  
80 designated as a No-Take Zone (NTZ) in 2003 and became the United Kingdom's (UK) first  
81 Marine Conservation Zone (MCZ) in 2009.

82 The Irish Sea receives Atlantic water and influences from the Celtic Sea and St.  
83 Georges Channel in the south, and via the North Channel and Malin Shelf Sea in the North  
84 (Howarth 2005). The semi diurnal tides are the dominant physical process in the region with  
85 the North and South tidal streams meeting in a region just south of the Isle of Man, where  
86 they form a standing wave (Howarth 2005). The southern emanating tidal streams represent a

87 potential vector for northward larval dispersal, a likely mechanism linking the Bristol  
88 Channel and Irish Sea lobster populations. However, hydrodynamic modelling indicates  
89 patterns of stratification and residual flows that may significantly constrain larval recruitment  
90 while models of larval dispersal incorporating different behaviours (time of release and  
91 swimming) have reported substantially different patterns of predicted dispersal  
92 (<http://www.susfish.com/index.php/workPackagesDetails/2>). As such, empirical data are  
93 needed to assess patterns of connectivity within and between the Irish Sea and Bristol  
94 Channel, and in doing so to assess the connectivity of the Lundy NTZ.

95         Microsatellites have been shown to be capable of resolving fine scale population  
96 structure (McKeown et al. 2015) and were employed in this study to test the general  
97 hypothesis of a panmictic population of *H. gammarus* across the Irish Sea and Bristol  
98 Channel against alternatives of inter- and intra-regional structuring. Encapsulated within this  
99 was an assessment of the connectivity and other genetic parameters of the Lundy NTZ at  
100 local and regional scales. Studies such as this one will provide a baseline for future genetic  
101 monitoring of the lobster stock within the area and contribute to holistic assessments of the  
102 impact of small marine reserves on fisheries enhancement and sustainability.

103

## 104 **2. Materials & Methods**

### 105 *2.1. Sample collection, DNA isolation and microsatellite genotyping*

106         Samples were collected from lobsters at Lundy Island from within the NTZ (Lundy  
107 No-Take Zone, LNTZ) and outside the NTZ (Lundy Island Control Zone, LICZ), and from  
108 seven locations within the Irish Sea and Bristol Channel (Table 1, Figure 1). All samples  
109 consisted of mixed cohorts of adults.

110 Haemolymph was extracted from lobsters using a 2 ml Terumo syringe with a  
111 G23x25 mm needle (VWR International Ltd.) and preserved in absolute ethanol (1:8). DNA  
112 was extracted using the QIAGEN DNeasy Blood & Tissue Kit (QIAGEN, CA-USA) after an  
113 initial centrifugation step of 400  $\mu$ l of the haemolymph/ethanol mixture for 5 minutes at 7000  
114 x g in order to precipitate the cells into a pellet easily separated from the alcoholic fraction.

115 Twelve species-specific microsatellite loci (André & Knutsen 2010) were amplified in  
116 two multiplex PCR reactions. Amplification was carried out using the QIAGEN Multiplex  
117 PCR Kit (QIAGEN, CA-USA) in a final volume of 10  $\mu$ l. This contained 5  $\mu$ l of Multiplex  
118 Kit Buffer, 1  $\mu$ l of genomic DNA (~100ng) and 0.2  $\mu$ l of each forward and reverse primer for  
119 the specific multiplex reaction. The PCR cycle involved an initial denaturation step at 95 °C  
120 for 15 minutes, followed by 34 cycles of 45 seconds at 94 °C, 45 seconds at 59 °C and 45  
121 seconds at 72 °C, and a final extension step at 72 °C for 45 minutes. Products were then run  
122 on an ABI 3730 Genetic Analyzer (Applied Biosystems) alongside a GS500LIZ size standard  
123 and alleles were inferred using GeneMapper 4.0 (Applied Biosystems).

124

## 125 *2.2. Statistical analysis of microsatellite data*

126 Genetic variation within samples was characterised using number of alleles ( $N_A$ ),  
127 effective number of alleles (AEFF), allelic richness (AR; El Mousadik & Petit 1996),  
128 observed heterozygosity ( $H_o$ ), and expected heterozygosity ( $H_e$ ) (Nei 1978), calculated using  
129 GENALEX 6.2 (Peakall & Smouse 2006). Deviations from Hardy-Weinberg equilibrium  
130 (HWE) expectations and genotypic linkage equilibrium (LD) between pairs of loci were  
131 tested using exact tests (10 000 batches, 5000 iterations) in GENEPOP 3.3 (Raymond &  
132 Rousset 1995). Deviations from HWE were measured using  $F_{IS}$ , calculated according to Weir  
133 & Cockerham (1984) and tested for significance by 10 000 permutations in FSTAT 2.9.3

134 (Goudet 1995). Mean pairwise relatedness within samples was calculated using the  
135 relatedness estimator,  $r_{qg}$ , of Queller & Goodnight (1989) in GENALEX with associated 95%  
136 confidence intervals determined by 1000 bootstraps. The 95% confidence intervals for the  
137 expected range of  $r_{qg}$  under a panmictic model were calculated following 999 permutations of  
138 genotypes among samples.

139 Genetic differentiation between and among samples was assessed using pairwise and  
140 global (i)  $F_{ST}$  values, calculated in FSTAT with significance assessed by 10 000  
141 permutations, and (ii) exact tests of allele frequency homogeneity performed in GENEPOP  
142 (10 000 batches 5000 iterations). The simulation method implemented in POWSIM (Ryman  
143 & Palm 2006) was used to estimate the sample size-dependent Type I and Type II error  
144 probabilities of exact tests.  $F_{ST}$  values were also calculated using the null allele correction  
145 method in FreeNA (Chapuis & Estoup 2007). The assumption of neutrality of the  
146 microsatellite loci was assessed using the FDIST outlier test (Beaumont & Nichols 1996)  
147 implemented in LOSITAN (Antao et al. 2008). Outlier tests were performed globally and  
148 between pairs of samples.  $F_{ST}$  matrices were visualised in a Multi-Dimensional Scaling plot  
149 constructed in R (R Core Team 2013) with the pcoa function. Genetic relationships were also  
150 visualised by a Principal Component Analysis (PCA) performed in PCAGEN (Goudet 1999).  
151 Mantel tests, implemented in the IBDWS software (Jensen et al. 2005) were used to test for  
152 correlation between pairwise  $F_{ST}$  and geographical distances between sample sites (i.e.  
153 isolation by distance). Geographical distances were calculated as the shortest sea distances  
154 between approximate centres of sampling locations using NETPAS 2.5. Genetic structure  
155 was also investigated without *a priori* sample information included using the Bayesian  
156 clustering analysis implemented in the program STRUCTURE (Pritchard et al. 2000; Falush  
157 et al. 2003, 2007). Following recommendations by Hubisz et al. (2009) analyses were



158 replicated for both the original 'no locprior' and new 'with locprior' models. Each run  
159 consisted of a burn-in of 1 000 000 steps followed by 5 x 1 000 000 steps.

160 To investigate asymmetric dispersal patterns (source-sink recruitment) recent  
161 migration rates were inferred using the Bayesian inference approach implemented in the  
162 program BAYESASS 3.0.3 (Wilson & Rannala 2003). Simulations with a unique seed value  
163 were run for 10 000 000 iterations with MCMC chains sampled every 1 000 iterations,  
164 following an initial burn-in of 1 000 000 iterations. As suggested in the program  
165 documentation, the five mixing parameters were adjusted to ensure acceptance rates between  
166 20% and 60%, with  $\Delta M$ ,  $\Delta A$  and  $\Delta F$  set to 0.70. In order to examine convergence, the  
167 posterior mean parameter estimates of multiple runs were compared.

168 Effective population sizes ( $N_e$ ) were estimated using LDNE (Waples & Do 2008), in  
169 which a random mating model was assumed. To test for recent effective population size  
170 reductions, BOTTLENECK 1.2 (Piry et al. 1999) was utilised using 10 000 permutations for  
171 the Infinite Allele (IAM) (Crow & Kimura 1970), Stepwise Mutation (SMM) (Ohta &  
172 Kimura 1973) and Two-Phase Mutation (TPM) (Di Rienzo et al. 1994) models of  
173 microsatellite evolution. Approximately 90% of microsatellite mutations are single step  
174 (Garza & Williamson 2001), thus the parameters of the TPM model were set to run at 90% of  
175 single step mutations with a variance of 10 among multiple steps. The Wilcoxon test was  
176 used to determine whether any of the sample sites show an excess of heterozygosity, which is  
177 expected after a severe bottleneck (Cornuet & Luikart 1996). In addition, the graphical mode-  
178 shift test was incorporated to detect shifts from the normal L-shaped distribution of allele  
179 frequencies that are expected at equilibrium (Luikart et al. 1998). A second approach to  
180 assess for evidence of a population bottleneck was the  $M$ -ratio between the total number of  
181 alleles ( $k$ ) and the overall range in allele size ( $r$ ) (Garza & Williamson 2001).  $M$  was  
182 calculated in ARLEQUIN 3.5 (Excoffier & Lischer 2010) by taking the mean population

183 value from the modified Garza-Williamson (GW) index computation. The modification of the  
184 GW index avoids a division by zero when a gene sample is fixed for a single allele (Excoffier  
185 et al. 2005). Values of  $M < 0.68$  were considered as a sign of a recent population bottleneck  
186 (Garza & Williamson 2001), as described in Coscia et al. (2012).

187

### 188 **3. Results**

189 The total number of alleles per locus ranged from seven to 17 (mean = 10.17) and  
190 levels of genetic variability were highly similar among all samples (Table 1; Table S2; Figure  
191 S1). Three locus-pairings were found to be in global linkage disequilibrium: HGD106 &  
192 HGC131b ( $P = 0.025$ ); HGC129 & HGC118 ( $P = 0.019$ ); and HGD111 & HGC103 ( $P =$   
193  $0.042$ ). However, these test results became non-significant after Bonferroni correction.  
194 Furthermore, significant global test results were due to significant results occurring only in  
195 one or two samples in each case (HGD106 & HGC131b locus-pairing: WEX,  $P = 0.003$ ;  
196 HGC129 & HGC118 locus-pairing: LNTZ,  $P = 0.043$ , and; HGD111 & HGC103 locus-  
197 pairing: SW,  $P = 0.002$ ; NW,  $P = 0.016$ ) and became non-significant upon omission of these  
198 samples. All deviations from HWE were due to heterozygote deficits. Eight out of a total 108  
199 individual locus/sample tests exhibited deviations from HWE (at  $P < 0.05$ ), with two  
200 remaining significant after Bonferroni correction: LNTZ and LICZ at locus HGA8 ( $P <$   
201  $0.0005$ ). Multilocus deviations from HWE were reported for the Waterford (WF) and North  
202 Dublin (ND) samples from Ireland and both Lundy samples from the Bristol Channel (LNTZ,  
203 LICZ) with LNTZ exhibiting the largest  $F_{IS}$  value (Table 1). In all cases these were due to  
204 single locus deviations occurring at HGA8, and became non-significant after omission of this  
205 locus. Mean kinship values were within ranges predicted for a panmictic model but there  
206 were significant differences between samples (Figure 2).

207 A global exact test of genic homogeneity was non-significant ( $P = 0.727$ ) as were all  
208 pairwise comparisons except that between LNTZ and SW ( $P < 0.05$ ) (Table 2). Simulation  
209 analysis performed in POWSIM indicated that the locus polymorphism and sample sizes  
210 used in this study would have conferred sufficient power for tests to detect even low levels of  
211 ‘true’ genetic structuring (estimated probability of 0.99 that the data would identify  
212 differentiation at true  $F_{ST} = 0.01$ ) while retaining a low Type I error probability ( $P = 0.04$ ).  
213 Similarly, global  $F_{ST}$  (-0.000; 95% CI: -0.002 – 0.001) and all pairwise  $F_{ST}$ s were non-  
214 significant for both unedited and null allele corrected data sets (Table S3). LOSITAN  
215 confirmed the neutrality of the 12 microsatellite markers that were used, with none being  
216 selected as candidates for balancing or positive selection in global or pairwise tests.  
217 Visualisation of pairwise  $F_{ST}$  using MDS revealed no geographic pattern (Figure 3) and there  
218 was not significant IBD ( $r = 0.057$ ,  $P = 0.35$ ). Bayesian clustering analysis did not provide  
219 evidence of population structuring, with all runs unanimously supporting a model of  $K=1$ .  
220 Inconsistencies between the posterior mean parameter estimates for the BAYEASS  
221 simulations indicated there was insufficient information in the data to accurately infer recent  
222 migration rates among LNTZ, SW, DEV, and WF, WEX. Estimates of  $N_e$  suggest that the  
223 sampled populations were very large, with upper 95% confidence intervals including infinity  
224 (Table 1). BOTTLENECK reported significant results under the IAM for the SW ( $P = 0.032$ )  
225 and DEV ( $P = 0.017$ ) samples but all tests assuming a TPM or SMM were not significant.  
226 Allele frequency distributions revealed no mode shifts for any sample (i.e. standard L shapes  
227 detected). The  $M$ -ratio values for all samples suggested historical population size reductions.

228

#### 229 **4. Discussion**

230 Microsatellite markers represent powerful tools to investigate demographic processes  
231 that may be beyond the resolution of other genetic marker types (Shaw et al. 2004). This

232 study is the first to employ microsatellites to look at fine scale population processes in *H.*  
233 *gammarus*. Simulation analysis indicated that marker polymorphism and sample sizes in this  
234 study conferred considerable power to detect even low levels of genetic differentiation  
235 between samples. Global and pairwise ( $F_{ST}$  and exact) tests provided no evidence of  
236 significant differentiation among samples. This pattern of genetic homogeneity was also  
237 replicated for analysis of sex partitioned samples. For many marine species, estimates of  
238 genetic structure may be compromised by adult dispersal (resulting in mechanical mixing of  
239 differentiated populations) (Nielsen et al. 2004). However, as *H. gammarus* adults are largely  
240 sedentary (Jensen et al. 1993; Bannister et al. 1994; Smith et al. 2001; Moland et al. 2011;  
241 Øresland & Ulmestrand 2013), postlarval dispersal is unlikely to influence results.  
242 Furthermore, Bayesian clustering analysis provided no evidence of genetic structuring, a  
243 pattern which is typical of lobsters (Tolley et al. 2005; García-Rodríguez & Perez-Enriquez  
244 2008; Naro-Maciel et al. 2011) except where gene flow is restricted by conspicuous physical  
245 drivers (e.g. oceanography; geographic barriers) (Gopal et al. 2006; Palero et al. 2008). The  
246 lack of population structure indicates high levels of gene flow and connectivity within and  
247 between the Irish Sea and Bristol Channel.

248         The overall genetic homogeneity is also consistent with connectivity of the Lundy  
249 NTZ. Samples collected outside of the Lundy NTZ reported similar levels of genetic diversity  
250 to the Lundy NTZ sample, which is also compatible with a lack of isolation for the Lundy  
251 NTZ. Calò et al. (2016) reported similar levels of genetic diversity for protected and  
252 unprotected areas in the saddled sea bream (*Oblada melanura*). This pattern differs from the  
253 study including two MPAs in the western Mediterranean by Pérez-Ruzafa et al. (2006) that  
254 reported higher values of total and standardised allelic richness in protected populations of  
255 *Diplodus sargus* than in unprotected ones.

256           The similar levels of genetic diversity between MPA and non-MPA sites observed for  
257 *H. gammarus* may be a combination of (i) the high level of connectivity throughout the  
258 studied area and (ii) a lack of fishery induced genetic erosion. The significant *M*-ratio test  
259 results obtained for all *H. gammarus* samples likely reflect ancestral bottlenecks common to  
260 all populations within the study area (Girod et al. 2011). Similar genetic signatures of  
261 ancestral population size changes associated with historical climate change events such as the  
262 Pleistocene glaciations have been detected in a range of lobster species (Gopal et al. 2006;  
263 García-Rodríguez & Perez-Enriquez 2008; Naro-Maciel et al. 2011). No compelling evidence  
264 of more recent bottlenecks was discovered in any *H. gammarus* sample. Although the Gower  
265 (SW) and Ilfracombe (DEV) samples revealed significant bottleneck signatures using the  
266 Wilcoxon sign rank test, this is not considered as strong evidence for a bottleneck, as test  
267 results were only significant under the IAM, a mutation model deemed unlikely for  
268 microsatellites (Piry et al. 1999), and it is recommended to use the SMM to be statistically  
269 conservative when testing for recent bottlenecks (Luikart & Cornuet 1998). This was further  
270 confirmed by the lack of L-mode shifts, intrasample relatedness values within expected  
271 ranges for a panmictic population, and estimates of  $N_e$  which included infinity. Levels of  
272 intrasample variation ( $H_o$ ,  $H_e$ ,  $N_A$ ) were also comparable to values reported for putative ‘non-  
273 bottlenecked’ populations sampled from other regions (Ferguson, unpublished data)  
274 indicating that *H. gammarus* has not undergone any recent genetic erosion due to fishing  
275 activities within the Irish Sea or Bristol Channel and retains a high level of genetic variation.

276           An important consideration is that the high background levels of genetic diversity  
277 may make any MPA associated increase in diversity difficult to detect. Moreover, any genetic  
278 changes, positive or negative, associated with the NTZ may not be detectable for a number of  
279 generations and thus may not be apparent due to its recent establishment. As such, this study  
280 must form part of an ongoing genetic monitoring program.

281 In this study significant heterozygote deficits were reported for the Irish Waterford  
282 and North Dublin samples and the two Lundy samples, with the Lundy NTZ exhibiting the  
283 largest deficit. In all cases these deficits were driven by a single locus (HGA8), indicating a  
284 null allele effect. However, in light of the low level of genetic structure throughout the  
285 studied area, the restriction of such heterozygote deficits to particular samples, with the  
286 largest deficit reported for the NTZ, suggests there may be some biological component.  
287 Significant heterozygote deficits are a common feature among marine invertebrates with  
288 biological drivers including inbreeding, selection and Wahlund effects (Addison & Hart  
289 2005). Inbreeding can be discounted as values of relatedness for all samples are in line with  
290 expectations of panmixia and provide no evidence that closely related individuals have been  
291 sampled. Selection against heterozygotes has rarely been documented (Teske et al. 2013) and  
292 outlier tests reported no evidence of selection effects at the employed loci. Based on the low  
293 level of genetic structure throughout the studied region spatial Wahlund effects due to the  
294 sampling of mixtures of individuals derived from genetically differentiated populations can  
295 also be discounted. However, for many highly fecund marine species with pelagic larval  
296 stages, large variances in reproductive success occur among individuals (sweepstakes  
297 recruitment) such that in any reproductive event only a small fraction of adults contributes to  
298 the next generation (Hedgecock 1994). This can generate genetic differences within and  
299 between cohorts (cohort Wahlund) (Ruzzante et al. 1996) without restricted gene flow where  
300 recruitment is temporally or spatially variable (Johnson et al. 1993). By a process of  
301 elimination, sweepstakes effects as a component of overall localised recruitment  
302 stochasticity, represents the most plausible biological mechanism for generating the reported  
303 heterozygote deficits. Palero et al. (2011) suggested that for the lobster *Palinurus elephas*, the  
304 biased exploitation of large individuals led to a reduction in the number of large females and  
305 a consequent reduction in variance in reproductive success. In the case of the Lundy NTZ, the

306 predicted persistence of larger females (Wootton et al. 2012) may increase the scope for  
307 variance in reproductive success. Furthermore, a paternity study in the American lobster, *H.*  
308 *americanus*, reported evidence of high levels of multiple paternity at exploited sites but single  
309 paternity at the only unexploited site (Gosselin et al. 2005). Similar fishery associated  
310 reduction in multiple paternity at the Lundy NTZ would serve to increase variance in  
311 reproductive success compared to other multiple paternity areas (McKeown & Shaw 2008).  
312 Confirmation and identification of any such NTZ effects is vital and will require comparative  
313 analysis of other species and ideally comparisons of intraspecific patterns associated with  
314 other NTZs.

315         In conclusion, the data indicate that within the studied area, *H. gammarus*, including  
316 those within the NTZ, belong to a single panmictic population. Comparable levels of genetic  
317 diversity within the Irish Sea and with other regions indicate that the species retains high  
318 levels of neutral genetic diversity despite being heavily exploited. Spatial genetic  
319 homogeneity and large effective population size estimates suggest that if current breeding  
320 stock sizes are maintained, loss of neutral genetic variation within the Irish Sea and Bristol  
321 Channel is unlikely. The high connectivity throughout the studied area emphasises the need  
322 for cooperation between British and Irish fishery managers and also has implications for the  
323 spatial distribution of future MPAs and the establishment of a MPA connected network  
324 (Larson & Julian 1999). Future research aimed at describing even finer elements of  
325 demographic connectivity would benefit from employing genome wide SNP analyses. These  
326 offer the potential to assay a huge number of loci for use in isolation-with-migration models,  
327 and may reveal asymmetric gene flow dynamics (Wang & Hey 2010; Marko & Hart 2011)  
328 that were beyond the resolution of the current data set, as indicated by the simulations  
329 performed for BAYEASS. Additionally, direct estimates of dispersal could be obtained

330 through genetic parentage analysis of new recruits (Jones et al. 2005; Planes et al. 2009;  
331 Christie et al. 2010a; Saenz-Agudelo et al. 2011; Buston et al. 2012; D'Aloia et al. 2013).

332

### 333 **Acknowledgements**

334 This research formed part of SUSFISH, a 3 year project funded by the European Union  
335 Regional Development Fund (ERDF) under the Ireland Wales Programme 2007-2013  
336 (Interreg 4A, Project No. 042).

337

338



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559

560 **Tables**561 **Table 1.** Sample information and summary statistics of *Homarus gammarus* in the Irish Sea:562  $H_e$  expected heterozygosity,  $H_o$  observed heterozygosity,  $N_A$  number of alleles, AEFF563 effective number of alleles,  $A_R$  allelic richness,  $F_{IS}$  inbreeding coefficient, and  $N_e$  estimates of

564 effective population size with 95% confidence intervals (parametric). Significant values of

565  $F_{IS}$  are denoted by \* ( $P < 0.05$ ) or \*\*\* ( $P < 0.001$ ).

Sampling Location	Code	Coordinates	N	$H_e$	$H_o$	$N_A$	AEFF	$A_R$	$F_{IS}$	$N_e$ (95% CI)
Howth,	ND	53.469° N,	36	0.69	0.68	7.25	3.44	6.99	0.016	-272.7
Dublin		6.084° W								(474.0 - $\infty$ )
Carne,	WEX	52.184° N,	29	0.67	0.64	7.00	3.42	7.00	0.037	677.7 (98.3
Wexford		6.302° W								- $\infty$ )
Dunmore	WF	52.085° N,	48	0.68	0.64	7.92	3.59	7.24	0.050*	-768.4
East,		7.033° W								(400.7 - $\infty$ )
Waterford										
North Llŷn	NW	52.806° N,	40	0.67	0.64	7.92	3.47	7.30	0.049*	2407.4
Peninsula,		4.823° W								(187.2 - $\infty$ )
Gwynedd										
Aberystwyth,	CB	52.415° N,	44	0.67	0.66	7.58	3.34	7.07	0.012	-1234.9
Ceredigion		4.236° W								(253.3 - $\infty$ )

Gower, Swansea	SW	51.550° N, 4.144° W	48	0.68	0.69	7.83	3.62	7.15	-0.017	-924.1 (378.0 - ∞)
Lundy Island (outside NTZ)	LICZ	51.205° N, 4.682° W	44	0.68	0.65	8.00	3.66	7.42	0.046*	1252.1 (175.6 - ∞)
Lundy Island NTZ	LNTZ	51.189° N, 4.649° W	48	0.69	0.64	7.75	3.61	7.17	0.072***	-406.8 (642.2 - ∞)
Ilfracombe, Devon	DEV	51.228° N, 4.125° W	48	0.69	0.68	7.50	3.65	6.97	0.014	-916.0 (363.0 - ∞)

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566

567

568 **Table 2.** Genetic differentiation of *Homarus gammarus* samples from the Irish Sea and  
 569 Bristol Channel.  $F_{ST}$  pairwise comparisons (lower diagonal) and exact G tests (upper  
 570 diagonal). \* denotes significant G-tests ( $P < 0.05$ ).

	LNTZ	LICZ	WF	CB	ND	WEX	SW	NW	DEV
LNTZ		21.057	23.444	25.968	24.981	19.784	36.454*	24.002	31.472
LICZ	-0.002		20.833	26.380	23.484	21.954	20.792	17.966	17.061
WF	0.002	-0.002		23.220	27.522	23.410	25.288	27.539	17.068
CB	0.000	0.001	0.002		30.277	22.052	23.771	15.187	23.773
ND	-0.002	-0.001	0.004	-0.001		23.467	24.216	18.863	24.525
WEX	-0.001	0.001	0.004	0.000	0.003		24.576	18.967	23.904
SW	0.001	-0.002	0.000	-0.001	0.000	0.001		14.771	17.262
NW	0.000	-0.002	0.004	-0.004	-0.003	0.001	-0.002		11.897
DEV	0.000	-0.001	0.000	-0.001	-0.002	0.003	-0.003	-0.004	

571

572

573 **Figure legends**

574 **Figure 1.** Sampling locations of the European lobster, *Homarus gammarus*, in the Irish Sea,  
575 highlighting the marine protected (white circle) and non-protected (black circle) areas.

576 **Figure 2.** Mean within-sample pairwise relatedness,  $r_{qg}$ , and associated 95% confidence  
577 intervals derived from bootstrap resampling. Grey bars are the 95% upper and lower expected  
578 values for a null distribution generated from 999 permutations of data from all samples.

579 **Figure 3.** MDS plot of pairwise  $F_{ST}$  values of *Homarus gammarus* in the Irish Sea.

580

Figure 1  
[Click here to download high resolution image](#)

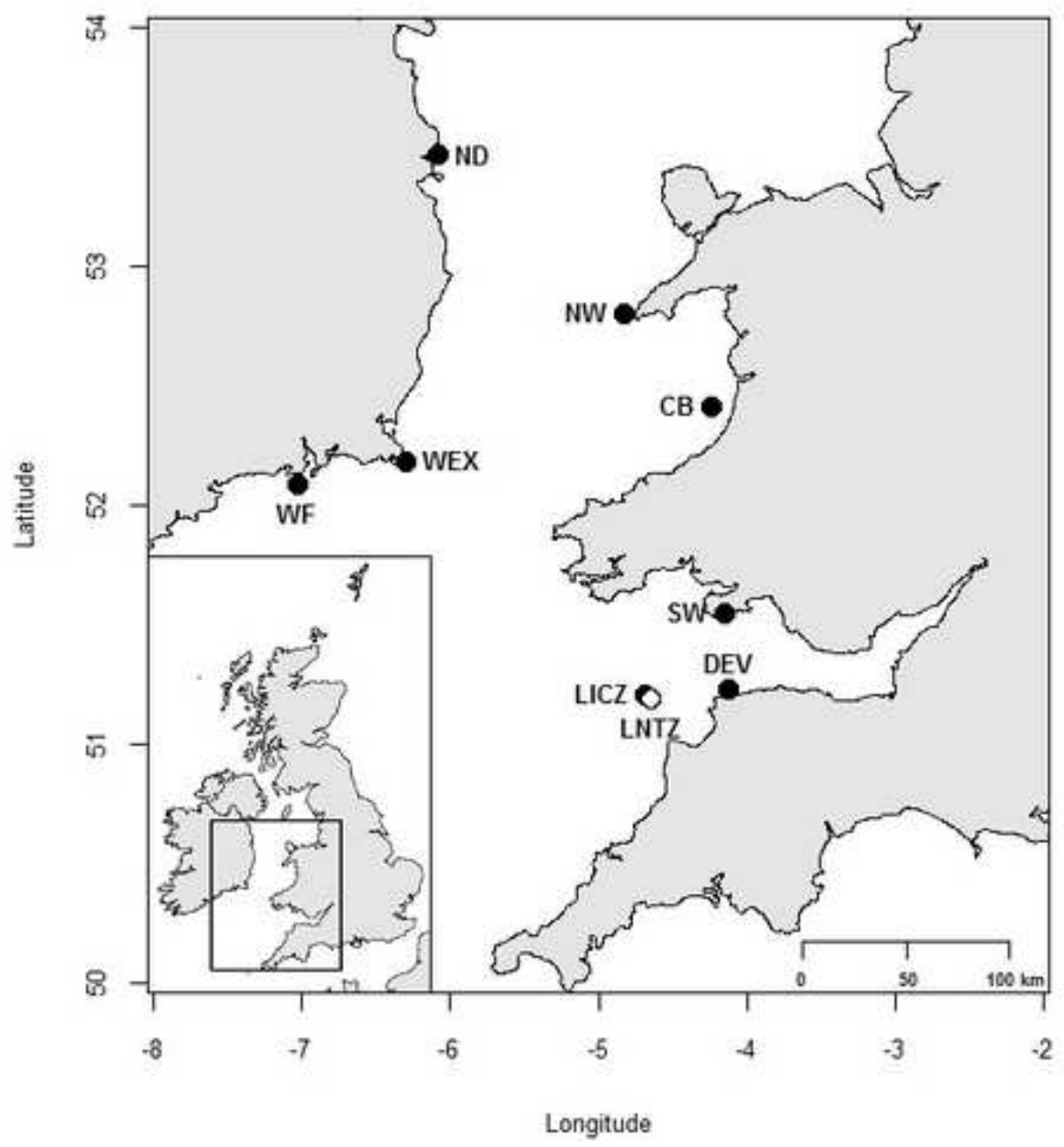


Figure 2

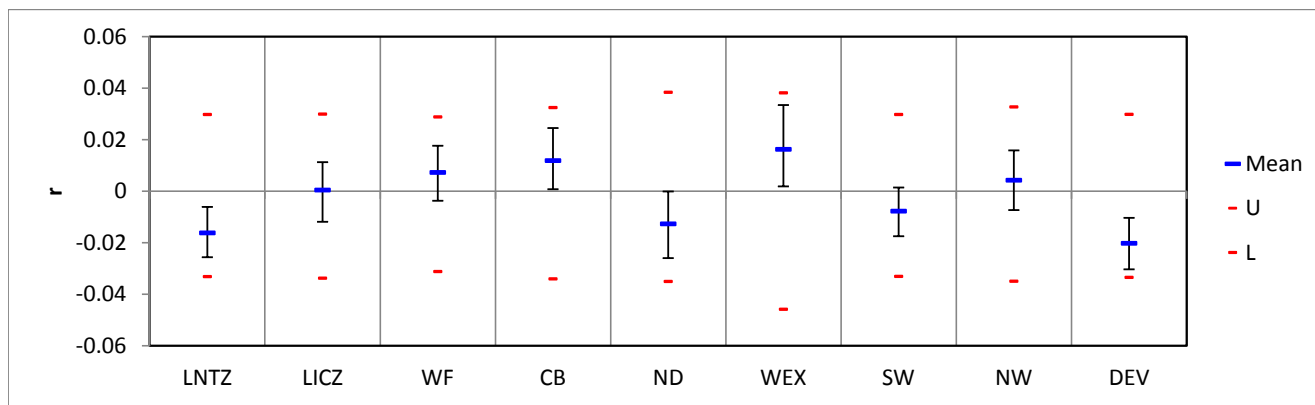
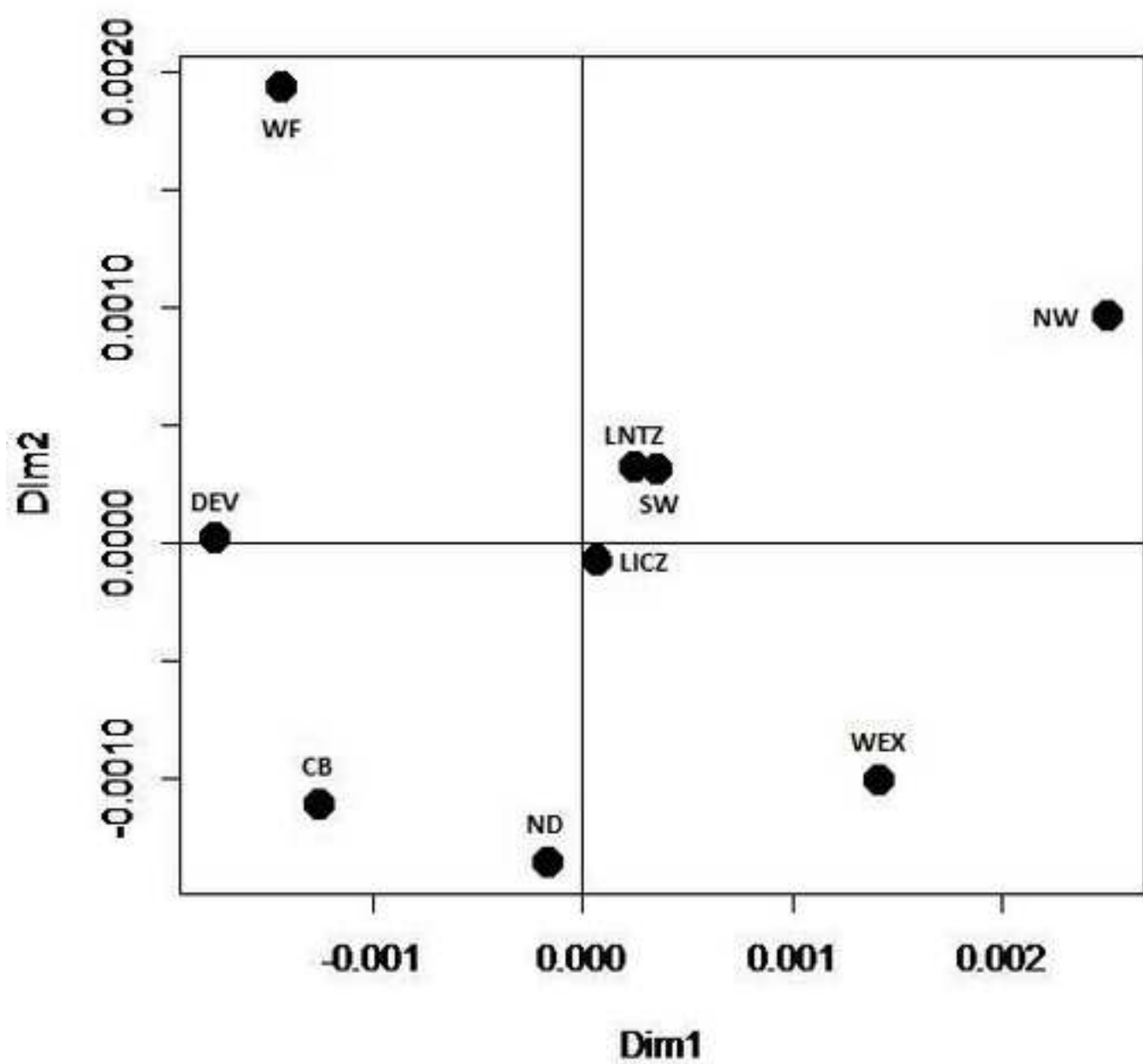




Figure 3  
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