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Genetic analyses reveal high levels of seed and pollen flow in hawthorn (*Crataegus monogyna* Jacq.), a key component of hedgerows

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1 Abstract Hedgerows represent important components of agri-environment landscapes that 2 are increasingly coming under threat from climate change, emergent diseases, invasive 3 species and land-use change. Given that population genetic data can be used to inform best-4 practice management strategies for woodland and hedgerow tree species, we carried out a study on hawthorn (Crataegus monogyna Jacq.), a key component of hedgerows, on a 5 6 regional basis using a combination of nuclear and chloroplast microsatellite markers. We found that levels of genetic diversity were high and comparable to, or slightly higher than, 7 8 other tree species from the same region. Levels of population differentiation for both sets of 9 markers, however, were extremely low, suggesting extensive gene flow via both seed and pollen. These findings suggest that a holistic approach to woodland management, one which 10 11 does not necessarily rely on the concept of "seed zones" previously suggested, but which also 12 takes into account populations with high and/or rare chloroplast (i.e. seed-specific) genetic 13 variation, might be the best approach to restocking and replanting.

14

15 KEYWORDS: *Crataegus monogyna*, gene flow, genetic diversity, hawthorn, microsatellites,
16 woodland management

17 Introduction

18

Hedgerows represent an important component of agri-environment landscapes, not just as 19 20 boundaries for fields, but also in providing a wide range of ecosystem services such as 21 habitats, food sources and wildlife corridors for animals, acting as wind-breaks, and preventing soil erosion (reviewed in Burel and Baudry 1995). The last few decades have 22 23 seen the emergence of an increasing number of threats to tree and shrub species, including those which make up hedgerows. These threats include plant pathogens and pests, habitat 24 25 loss, land use change, invasive species and climate change (Rackham 2008). Consequently, 26 managing hedgerow species, especially trees, is becoming increasingly important, requiring 27 clear knowledge and understanding of their ecological attributes and requirements. Trees lost 28 due to disease or other factors may have to be replaced under management programmes. In 29 Great Britain, the Forestry Commission has drawn up a map of 24 areas (seed zones) with the goal of maintaining provenance by only restocking woodlands with seed from the same zone 30 31 (Herbert et al. 1999). These zones are defined by similar geographic, climatic and ecological 32 characteristics, but it is becoming clear that population genetic structuring can also play an important role in woodland management (Müller-Starck et al. 1992; Ennos et al. 1998). 33 Recent studies on ash (Fraxinus excelsior L.; Sutherland et al. 2010; Beatty et al. 2015a) and 34 35 alder (Alnus glutinosa [L.] Gaertn.; Beatty et al. 2015b) found very little genetic 36 differentiation between woodlands across Great Britain and Ireland, suggesting that the 37 sourcing of local seeds for replanting might not be required. Natural recolonization, regeneration and succession in plants, including trees, relies on 38 39 seed dispersal, which ultimately determines adult vegetation composition (Howe and Smallwood 1982; Nathan and Muller-Landau 2000; Levine and Murrell 2003). Likewise, 40 pollen dispersal also has an important role to play in the same processes, especially across 41

42 deforested landscapes (reviewed in Bacles and Ennos 2008, but see Provan et al. 2008). Measuring seed dispersal is difficult, with both "tagging" and "trapping" methods having 43 drawbacks (reviewed in Beatty et al. 2015a). The use of high-resolution chloroplast 44 microsatellites markers, however, permits the elucidation of maternal gene flow (i.e. via seed 45 in angiosperms; Provan et al. 2001), and in combination with data from nuclear markers can 46 indicate the relative importance of seed and pollen flow in natural populations (Ennos 1994). 47 48 Furthermore, fine-scale studies wherein both parent plants and saplings are sampled allow individual parental assignment and the ability to directly measure seed dispersal distances, as 49 50 demonstrated in a recent study on European ash (Beatty et al. 2015a). Hawthorn (*Crataegus monogyna*, also known as common or one-seeded hawthorn) is a 51 species which grows in both woodland and hedgerows, generally occurring primarily in the 52 53 latter (Gosler 1990). It prefers areas with a high light intensity and exposed soil, and is adept 54 at colonising abandoned or eroded areas (Fineschi et al. 2005). Hawthorn hedges provide habitats for many vertebrate and invertebrate species, and are important nesting sites for 55 56 birds, with their thorns providing protection from predators (Pollard et al 1974; Fineschi et al. 2005). The species is indigenous to Europe, and is also found from North Africa to the 57 Himalayas (Christensen 1992). Like all hawthorns, C. monogyna facilitates more than one 58 group of pollinators, including primarily bumblebees, honey bees and hoverflies (Gyan and 59 60 Woodell 1987; Gosler 1990; Campbell et al. 1991). The fruits it produces persist throughout 61 the winter, acting as an important food source for birds, which are the major vectors for seed dispersal. These small, fleshy fruits each contain a single seed, and are produced from open 62 pollination, with self-pollination generally prevented by gametophytic self-incompatibility 63 64 (Clapham et al. 1990). However, a study on a British population showed the production of fruits in the absence of insect pollination, indicating either selfing or apomixis (reviewed in 65 66 Jacobs et al. 2009).

67 In recent years, hawthorns have increasingly come under threat from several diseases, including fire blight, which is caused by the bacterial pathogen Erwinia amylovora (Burrill) 68 Winslow et al. (Schroth et al. 1974), and leaf spot, which results from a fungal infection by 69 70 Entomosporium maculatum Lev. (Stowell and Backus 1966). Given that very little is known on the genetic structure of natural hawthorn populations, and that population genetic studies 71 72 on hedgerow species are rare compared to those on woodland trees, we analysed the genetic diversity in populations of hawthorn on a regional basis across Northern Ireland to inform 73 management strategies. A combination of nuclear and chloroplast microsatellites was used to 74 75 determine the relative importance of pollen- and seed-mediated gene flow, and to facilitate comparison with woodland species across the same region (Beatty et al. 2015a, 2015b, 2016). 76

77 Materials and methods

78

79 *Sampling and DNA extraction*

80 Samples were collected from 23 sites across Northern Ireland along with one site in the Republic of Ireland (Figure 1; Table 1). Samples were taken from a combination of 81 hedgerows and woodland, depending on the site. Woodlands were selected that had been 82 83 previously designated as ancient or semi-natural based on data collected for the Woodland Trust Inventory of ancient and long-established woodland in Northern Ireland 84 85 (www.backonthemap.org.uk) and the National Survey of Native Woodlands 2003-08 in the Republic of Ireland (www.npws.ie). Woodlands were also selected based on government 86 information from the Department of the Environment such ASSIs (Areas of Special Scientific 87 88 Interest), as well as the landscape character areas listing the woodlands and species present in 89 each region (https://www.doeni.gov.uk). During sampling, 3-4 leaves were taken from each of a maximum of 30 trees and stored in silica gel. The GPS coordinates of each tree were 90 91 recorded. DNA was extracted using the cetyltrimethylammonium bromide (CTAB) method (Doyle and Doyle 1987). Nuclear genotypes were obtained for between 19 and 30 92 individuals per population (Table 1; N = 677; mean = 28.208), and chloroplast haplotypes 93 were obtained for between 20 and 30 individuals per population (Table 1; N = 701; mean = 94 29.208). 95

96

97 *Genotyping*

All samples were genotyped for eight nuclear microsatellites (Table 2) and six chloroplast
microsatellites (Table 3). Seven of the nuclear microsatellite loci (excluding CH02D11) had
previously been used in a population genetic study of two different *Crataegus* species, *C*. *douglasii* and *C. suksdorfii* (Lo et al. 2009). All eight microsatellite markers were originally

developed for studies of apple (*Malus x domestica*; Liebhard et al. 2002). To develop *de novo*chloroplast microsatellite markers, *Crataegus monogyna* chloroplast sequences in the
GenBank database were searched for mononucleotide repeats of nine or more (Provan et al.
2001). Primers were designed using the Primer3 program to amplify the six loci in two
multiplexes, which were combined for a single genotyping run (Table 3).

107 PCR was carried out in a total volume of 10 µl containing 100 ng genomic DNA, 5 pmol of 6-FAM labelled M13 primer, 0.05 pmol of each M13-tailed forward primer, 5 pmol each 108 reverse primer, $1 \times PCR$ reaction buffer, 200 µM each dNTP, 2.5 mM MgCl₂ and 0.25 U 109 110 GoTaq Flexi DNA polymerase (Promega, Sunnyvale, CA, USA). PCR was carried out on a number of machines: the MWG Primus thermal cycler (Ebersberg, Germany), MJ Research 111 PTC-200 and PTC-220 Gradient Peltier thermal cyclers (Quebec, Canada) and Biometra T-112 113 Gradient thermal cycler (Göttingen, Germany) using the following conditions: initial denaturation at 94 °C for 3 min followed by a range of cycles – between 30 and 55 for the 114 nuclear loci (see Table 2) and 30 for the chloroplast loci – of denaturation at 94 °C for 30 s, 115 annealing at 58 °C for 30 s, extension at 72 °C for 30 s and a final extension at 72 °C for 5 116 min. Genotyping was carried out on an AB3730xl capillary genotyping system. (Applied 117 Biosystems, Foster City, CA, USA). Allele sizes were scored using the GENEMARKER 118 software (V1.8, Softgenetics). 119

120

121 Data analysis

GENEPOP (V3.4; Raymond and Rousset, 1995) was used to test for linkage disequilibrium between nuclear microsatellite loci. To estimate genetic diversity within the populations, levels of observed (H_0) and expected (H_E) heterozygosity were calculated using the ARLEQUIN software package (V3.5.1.2; Excoffier and Lischer, 2010), whilst levels of allelic richness (A_R) and fixation indices (F_{IS}) were calculated using the FSTAT software package 127 (V2.9.3.2; Goudet, 2001). Significance of F_{IS} was determined by 10,000 randomisation steps. Chloroplast microsatellite allele sizes were combined into haplotypes, and levels of 128 genetic diversity (H) based on haplotype frequencies were calculated using ARLEQUIN. 129 130 The overall level of genetic differentiation between populations was estimated using Φ_{ST} , which gives an analogue of F_{ST} (Weir and Cockerham, 1984) calculated within the analysis 131 of molecular variance (AMOVA) framework (Excoffier et al. 1992) using ARLEQUIN. To 132 133 further identify possible patterns of genetic structuring, the software package BAPS (V6; Corander et al. 2003) was used to identify clusters of genetically similar populations using a 134 135 Bayesian approach. Ten replicates were run for all possible values of the maximum number of clusters (K) up to K = 24, the number of populations sampled, with a burn-in period of 136 10,000 iterations followed by 100,000 iterations. Multiple independent runs always gave the 137 138 same outcome.

It has been shown previously that the chloroplast genome is maternally inherited in all
Rosaceae tested to date (Fineschi 2005). Thus, the pollen to seed migration ratio (*r*) was
calculated using the formula:

142
$$r = \frac{mp}{ms} = \frac{\left(\frac{1}{\phi STn} - 1\right)(1 + FIS) - 2\left(\frac{1}{\phi STc} - 1\right)}{\left(\frac{1}{\phi STc} - 1\right)}$$
(Ennos 1994)

143 where m_p and m_s are the pollen and seed migration rates respectively, F_{IS} is the mean 144 inbreeding coefficient over all loci, and Φ_{STn} and Φ_{STc} are the levels of genetic differentiation 145 calculated for nuclear and chloroplast markers respectively. 146 **Results**

147

No significant evidence of consistent linkage disequilibrium (i.e. involving the same loci) 148 was detected between any of the eight nuclear microsatellites analysed (41 out of 694 tests). 149 Between twelve (CH05G11) and 36 (CH01F02) alleles were detected per locus, with a total 150 of 206 (mean = 25.75 per locus; Table 2). Levels of observed (H_0) and expected (H_E) 151 152 heterozygosity ranged from 0.600 (CH06G11) to 0.924 (CH01F02; mean = 0.764), and from 0.568 (CH05G11) to 0.931 (CH05D04; mean = 0.803), respectively. Levels of $F_{\rm IS}$ ranged 153 154 from -0.033 (CH05G11) to 0.130 (CH03C02), with a mean value of 0.047. Within populations, levels of allelic richness (A_R) averaged over loci ranged from 8.779 155 (Glenarm) to 12.074 (Portaferry) with a mean value of 10.556 (Table 1). Levels of observed 156 157 (H_0) and expected (H_E) heterozygosity ranged from 0.684 (Belleek) to 0.836 (Keady; mean = (0.763), and from (0.769) (Belleek) to (0.838) (Portaferry; mean = (0.803)) respectively. Levels of 158 F_{IS} ranged from -0.022 (Keady) to 0.122 (Belleek) with 13 out of the 24 values being 159 significantly different from zero. No significant difference was observed for any of the above 160 diversity measures between hedgerow and woodland populations. 161 Five of the six chloroplast microsatellite loci studied were polymorphic in the samples 162 analysed, exhibiting between two and seven alleles (Table 3). Combining allele sizes across 163 164 loci gave 23 haplotypes, twelve of which were found in a single individual (Table 1). The 165 most common (H1) was found in 75% (523 out of 701) of the trees studied, and the four most common haplotypes (H1-H4) accounted for 93% of individuals. Levels of haplotype 166 diversity (H) ranged from 0.193 (Drumshanbo Lough and Belleek) to 0.713 (Eglinton; Table 167 168 1).

169 Levels of population differentiation based on the nuclear and chloroplast markers

170 calculated from the AMOVA were $\Phi_{ST(n)} = 0.0092$ and $\Phi_{ST(c)} = 0.0385$ respectively. The

- 171 BAPS analysis assigned all 24 populations to a single genetic cluster, including the Coolure
- population from the Republic of Ireland. Finally, the ratio of pollen to seed migration (r) was
- 173 2.515.

Discussion

176	The results of the present study, the first to apply high-resolution microsatellite markers to
177	examine patterns of genetic diversity in Crataegus monogyna, indicate high levels of
178	diversity, but low levels of population differentiation, suggesting extensive gene flow via
179	both seed and pollen. Previous studies on hawthorn have either used low-resolution
180	(chloroplast restriction fragment length polymorphism – cpRFLP) markers over large
181	geographic scales to study the phylogeography of the species, thus focusing more on long-
182	term historical factors (Fineschi et al. 2005), or have used nuclear markers exhibiting
183	relatively limited variation (randomly amplified polymorphic DNA – RAPD) at more local
184	scales (Ferrazzini et al. 2008). The combination of nuclear and chloroplast microsatellites
185	allows the identification of any fine-scale structuring of genetic variation, as well as
186	providing insights into the relative contributions of seed and pollen flow in natural
187	populations (Powell et al. 1996; Provan et al. 2001).
188	The lack of comparable studies (i.e. using similar markers over similar geographical
189	scales) means that it is somewhat difficult to put the levels of diversity observed in the
190	present study into context. The sole study examining nuclear genetic variation in C .
191	monogyna used dominant RAPD markers (Ferrazzini et al. 2008), and thus it is not surprising
192	that the reported value of expected heterozygosity (mean $H_E = 0.291$) is much lower than that
193	in this study (mean $H_E = 0.803$). Likewise, the only other application of nuclear
194	microsatellite markers to Crataegus species was carried out in the polyploid species C.
195	suksdorfii and C. douglasii, meaning that comparable diversity statistics (e.g. H_E) could not
196	be calculated (Lo et al. 2009). With regard to other tree species from the same region
197	analysed using microsatellites, levels of diversity in hawthorn were comparable to those
198	reported in ash (mean $H_E = 0.765$; Beatty et al. 2015a), but slightly higher than those reported

for alder (mean $H_E = 0.663$; Beatty et al. 2015b), sessile oak (mean $H_E = 0.720$; Beatty et al. 2016) and pedunculate oak (mean $H_E = 0.714$; Beatty et al. 2016). Levels of inbreeding, 201 measured as F_{IS} , were similar to those in ash (Beatty et al. 2015a), but lower than those 202 reported for alder (Beatty 2015b), probably as a result of the more patchy distribution of the 203 latter in Northern Ireland.

The only previous study using chloroplast markers in *C. monogyna* found four haplotypes 204 across the whole of Europe using cpRFLP (Fineschi et al. 2005). As in the present study, one 205 of these haplotypes was found in the majority of individuals studied, including those from 206 207 Great Britain (no samples from Ireland were analyzed) but it seems likely that this corresponds to the dominant haplotype H1 found in Ireland in this study. Interestingly, the 208 209 same study found no variation at six chloroplast microsatellite loci, in comparison with the 23 210 haplotypes from the six loci used in the present study. This is most likely due to the fact that 211 these six loci are "universal", rather than species-specific, and tend to exhibit limited polymorphism in general (Provan et al. 2001). 212

The level of population differentiation observed for nuclear markers ($\Phi_{ST(n)} = 0.0092$) is

over an order of magnitude lower than those reported previously for *C. monogyna* and other

215 *Crataegus* species. Ferrazzini et al. (2008) reported a value of $\Phi_{ST(n)} = 0.203$ in six

216 populations of *C. monogyna* from northern Italy, despite the study area being approximately

the same as that in the present study. This discrepancy is almost certainly due to the different

218 properties of the markers used, since it has been shown previously that measures of

219 population differentiation based on RAPD phenotypes, as estimated by Ferrazzini et al.

220 (2008), are generally around one order of magnitude greater than those estimated using multi-

allelic codominant markers (Isabel et al. 1999). Microsatellite analyses of natural populations

of *C. douglasii* and *C. suksdorfii* yielded $\Phi_{ST(n)}$ values of 0.22 and 0.37 respectively, although

the latter fell to 0.15 when only diploid samples were included, as much of the differentiation

was postulated to arise from some degree of reproductive isolation between individuals with
differing ploidy levels (Lo et al. 2009). It was also inferred that gene flow was further
reduced between populations as a result of apomixis.

227 The fourfold higher Φ_{ST} value observed for chloroplast markers ($\Phi_{ST(c)} = 0.0385$) compared to that for nuclear markers ($\Phi_{ST(n)} = 0.0092$) is consistent with the fact that 228 chloroplast markers have a lower effective population size, and are maternally inherited in 229 230 angiosperms and thus dispersed via seed. As a result, they generally show higher levels of genetic structuring (Provan et al. 2001). Nevertheless, both values are very low, indicating 231 232 extensive gene flow across the study area via both pollen and seed. The ratio of pollen:seed gene flow (r = 2.515) is among the lowest reported for a range of plant species (Ennos 1994), 233 indicating efficient seed dispersal. In hawthorn, this is primarily via ingestion by mammals 234 235 and birds, and the latter, particularly thrushes (Turdus sp.), are most likely responsible for 236 long-distance dispersal events (García and Chacoff 2007). Extensive pollen flow is also expected due to the wide foraging ranges of the bees and flies that pollinate the species 237 (Steffan-Dewenter et al. 2002). 238

The BAPS analysis indicated that all hawthorn populations studied, including that from 239 Coolure in the Republic of Ireland, belonged to a single genetic cluster. This was also the 240 case in ash (Beatty et al. 2015a), and similar to that in alder, where 25 of the 26 populations 241 242 studied were assigned to the same cluster (Beatty et al. 2105b). Thus, as in the previous 243 studies, there is no obvious genetic structuring that could be used as the basis for management units or seed zones. Nevertheless, examination of the chloroplast haplotypes, 244 which are maternally inherited and thus indicative of seed diversity, can further inform 245 246 management decisions with respect to replanting and restocking. Populations like that at Portaferry, which contains not only seven of the eleven non-unique haplotypes but also four 247 248 unique haplotypes, and that at Eglinton, which has the highest overall chloroplast diversity,

249 would appear to represent good sources of genetic variation. In addition, the Glenarm population is the only one which is not dominated by the most common haplotype (H1), and 250 thus may merit special consideration. Conversely, the Belleek and Drumshanbo Lough 251 252 populations have the least amount of haplotypic diversity, along with significant F_{IS} values, and thus may not represent the best options for the acquisition of seeds for restocking. The 253 lack of differentiation between the Northern Ireland populations and Coolure might indicate 254 255 that, on a broader scale, seeds could be sourced from anywhere in Ireland, but further work on populations from the Republic of Ireland would be needed to confirm this, and to identify 256 257 any potential structuring of chloroplast variation. It should be noted that the use of ostensibly neutral markers, as in the present study, may not reflect local adaptive variation, which would 258 259 only be evident from reciprocal transplant or common garden experiments.

260 It is concluded that extensive pollen- and seed-mediated gene flow occurs in the populations of hawthorn analysed in the present study, giving rise to high levels of genetic 261 diversity but low levels of genetic differentiation between populations. These results mirror 262 those from other woodland tree species from the same region. Thus, there may be an 263 emerging picture that management of these species may be most efficiently carried out at a 264 regional level, although a more holistic approach might also target rare or high chloroplast 265 (seed-specific) variation. Such approaches might not be so suitable for other common tree 266 267 species such as hazel (Corylus avellana), which have very large seeds and thus may have 268 much more limited dispersal potential. Further information on such species is needed to complement the present and earlier findings. 269

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271

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278 Data archiving statement

279

All data will be deposited in DRYAD on acceptance.

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282

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Table 1 Details of hawthorn (*Crataegus monogyna*) populations studied. Type – H hedgerow, W woodland; Lat/Long given in degrees; N – number of individuals analysed; A_R – allelic richness; H_O – observed heterozygosity; H_E – expected heterozygosity; F_{IS} – inbreeding coefficient (* P < 0.05; ** P < 0.01; *** P < 0.001; NS non-significant); H1-H11 – frequency of chloroplast haplotypes (Un – unique haplotypes found in a single individual); H – gene diversity.

No	Name	Type	Lat	Long			Nuclea	ar								Chlor	oplast						
			(N)	(W)	N	A_R	H _o	H_E	F _{IS}	N	H1	H2	H3	H4	H5	H6	H7	H8	H9	H10	H11	Un	Н
1	Portaferry	Н	54.394	5.527	28	12.074	0.783	0.838	0.067**	30	17	2	2	2	-	1	-	1	1	-	-	4	0.681
2	Glenarm	W	54.951	5.965	25	8.779	0.729	0.772	0.058*	25	10	2	12	-	-	-	1	-	-	-	-	-	0.627
3	Kilkeel	Н	54.065	6.087	30	10.231	0.747	0.809	0.077*	30	22	2	1	3	2	-	-	-	-	-	-	-	0.458
4	Rea's Wood	W	54.710	6.234	30	10.231	0.782	0.802	0.026^{NS}	30	21	4	4	1	-	-	-	-	-	-	-	-	0.490
5	Breen Wood	W	55.140	6.240	29	11.056	0.714	0.804	0.114***	30	20	1	1	5	2	-	-	1	-	-	-	-	0.538
6	Loughbrickland	Н	54.292	6.322	30	10.188	0.763	0.798	0.044^{NS}	30	25	2	-	-	1	-	1	-	-	1	-	-	0.308
7	Ballyronan	Н	54.701	6.553	27	10.288	0.778	0.793	0.020^{NS}	30	26	-	2	1	1	-	-	-	-	-	-	-	0.251
8	Kilrea	Н	54.943	6.573	30	10.273	0.795	0.801	0.008^{NS}	30	25	3	2	-	-	-	-	-	-	-	-	-	0.301
9	Portrush	Н	55.195	6.577	30	10.356	0.731	0.795	0.082**	30	21	4	2	3	-	-	-	-	-	-	-	-	0.494
10	Peatlands Park	W	54.486	6.617	29	11.798	0.799	0.812	0.016^{NS}	30	20	2	5	1	-	-	-	-	-	-	-	2	0.538
11	Aghadowey	Н	55.015	6.645	30	10.637	0.789	0.812	0.029 ^{NS}	29	23	1	2	2	-	-	-	-	-	-	1	-	0.372
12	Errigal Glen	W	54.975	6.729	30	11.317	0.775	0.821	0.057*	30	25	3	-	1	1	-	-	-	-	-	-	-	0.303
13	Keady	Н	54.282	6.735	30	10.860	0.836	0.818	-0.022 ^{NS}	30	22	2	2	2	1	-	-	-	-	-	1	-	0.462
14	Killylea	Н	54.373	6.841	22	11.001	0.815	0.824	0.011 ^{NS}	30	24	2	1	1	-	1	-	-	-	1	-	-	0.363
15	Bannagher Glen	W	54.885	6.957	30	9.393	0.739	0.782	0.056*	30	26	-	-	-	-	-	3	-	-	-	-	-	0.246
16	Eglinton	Н	55.047	7.214	28	11.612	0.753	0.834	0.099***	30	16	2	2	2	-	-	-	1	-	-	-	4	0.713
17	Plumbridge	Н	54.743	7.227	27	10.464	0.806	0.808	0.003 ^{NS}	28	18	4	1	5	-	-	-	-	-	-	-	-	0.553
18	Lisnaskea	Н	54.240	7.395	19	9.649	0.768	0.817	0.061 ^{NS}	20	14	2	1	-	3	3	-	-	-	-	-	-	0.500
19	Crom	W	54.167	7.454	30	10.271	0.748	0.787	0.050*	30	24	2	2	1	-	-	-	-	-	-	-	1	0.361
20	Drumshanbo Lough	Н	54.654	7.493	27	10.580	0.709	0.775	0.087**	30	27	1	-	1	-	-	-	-	1	-	-	-	0.193
21	Lough Erne	Н	54.454	7.714	27	10.259	0.723	0.781	0.076**	30	23	3	-	2	1	1	-	-	-	-	-	1	0.409
22	Marble Arch	W	54.266	7.814	30	10.834	0.744	0.808	0.080**	30	26	1	-	1	2	2	-	-	-	-	-	-	0.251
23	Belleek	Н	54.465	8.107	29	10.008	0.684	0.769	0.122***	30	28	1	1	-	-	-	-	-	-	-	-	-	0.193
24	Coolure	W	53.671	7.360	30	11.185	0.789	0.814	0.031 ^{NS}	29	21	3	-	5	-	-	-	-	-	-	-	-	0.451

Locus	Cycles	Ν	Range (bp)	A_R	H_O	H_E	F_{IS}
CH01F02	45	36	213-305	15.600	0.924	0.927	0.004
CH02D11	40	31	159-221	12.067	0.787	0.768	-0.030
CH03C02	40	33	117-193	14.394	0.760	0.871	0.130
CH04G04	40	16	171-213	6.433	0.767	0.768	0.001
CH05D04	40	33	188-258	16.333	0.846	0.931	0.094
CH05D11	45	19	183-231	7.182	0.654	0.709	0.095
CH05G07	40	26	174-224	12.689	0.771	0.883	0.116
CH05G11	30	12	215-247	4.220	0.600	0.568	-0.033

Table 2 Hawthorn (*Crataegus monogyna*) nuclear microsatellite loci analyzed in this study. Loci were originally described in Liebhard et al. (2002). Cycles – number of cycles used in PCR; N – number of alleles; $A_{\rm R}$ – mean allelic richness; $H_{\rm O}$ – mean observed heterozygosity; $H_{\rm E}$ – mean expected heterozygosity; $F_{\rm IS}$ – mean inbreeding coefficient

Table 3 Hawthorn (*Crataegus monogyna*) chloroplast microsatellite loci analyzed in this study.

Multiplex	Accession	Region	Repeat	Primers*	Alleles (bp)
1	EU500411	rpl2-trnH intergenic	(A)9	ATAAAAACAAAAATAGGAGTAATTAATTGTGAC	100,101
				TTCTTAATAAATGATTTGCTACAAAAGG	
	JQ390913	rpl20-rps12 intergenic	(T) ₁₃	TATAACCTTCCCGACCACGA	119,121,122,123
				ATTTACTACTTTTATGTGTTTTTGATACCT	
	JQ392044	trnG-trnS intergenic	(A)9	GATTCGTTGGAACAATAAATGG	135
				GGATTGAAAGAGCCCTTCATAA	
2	JQ391567	rpl16 intron	(T)9	TTGCTTTACAACCCATAATCAGA	159,161,162,163,164,166
				ACCAACTCATCACTTCGTGTT	
	HG764984	trnH-psbA intergenic	$(T)_{13}(A)_{10}$	AGATAAAATACAACCTAAATTGAAAACTT	195,196,219,220,221,222,223
				ATATATGAGTTCTTGAAAGTAAAGGAGTAA	
	FJ395291	atpF-atpH intergenic	(T) ₁₀	CATTTTTCATATGATATCCTCTTTCTT	244,246
				CGGGTACCTAATTCTAATAAGTATCATTC	

* Forward tailed with CACGACGTTGTAAAACGAC; Reverse tailed with GTGTCTT

Table 4 Analysis of molecular variance (AMOVA) for nuclear and chloroplast microsatellitemarkers analysed in hawthorn (*Crataegus monogyna*).

Markers	Source of variation	Sum of squares	Variance	% variation
Nuclear	Among populations	105.747	0.02811	0.92
	Within populations	4006.636	3.01251	99.08
Chloroplast	Among populations	10.378	0.00833	3.85
	Within populations	140.752	0.2079	96.15

Figure Legend

Fig. 1 Locations of sites sampled in this study. Numbers correspond to those in Table 1. The dashed box in the top-right map show the area in Europe which the study was carried out.

