

## Aberystwyth University

### *Incipient genetic isolation of a temperate migratory coastal Sciaenid fish (Argyrosomus inodorus) within the Benguela Cold Current system*

Henriques, Romina; Potts, Warren; Sauer, Warwick; Shaw, Paul

*Published in:*  
Marine Biology Research

*DOI:*  
[10.1080/17451000.2014.952309](https://doi.org/10.1080/17451000.2014.952309)

*Publication date:*  
2015

*Citation for published version (APA):*

Henriques, R., Potts, W., Sauer, W., & Shaw, P. (2015). Incipient genetic isolation of a temperate migratory coastal Sciaenid fish (*Argyrosomus inodorus*) within the Benguela Cold Current system. *Marine Biology Research*, 11(4), 423-429. <https://doi.org/10.1080/17451000.2014.952309>

#### **General rights**

Copyright and moral rights for the publications made accessible in the Aberystwyth Research Portal (the Institutional Repository) are retained by the authors and/or other copyright owners and it is a condition of accessing publications that users recognise and abide by the legal requirements associated with these rights.

- Users may download and print one copy of any publication from the Aberystwyth Research Portal for the purpose of private study or research.
- You may not further distribute the material or use it for any profit-making activity or commercial gain
- You may freely distribute the URL identifying the publication in the Aberystwyth Research Portal

#### **Take down policy**

If you believe that this document breaches copyright please contact us providing details, and we will remove access to the work immediately and investigate your claim.

tel: +44 1970 62 2400  
email: [is@aber.ac.uk](mailto:is@aber.ac.uk)

1 **Title:** Incipient genetic isolation of a temperate migratory coastal Sciaenid fish (*A.*  
2 *inodorus*) within the Benguela Cold Current system

3 **Authors:** Romina Henriques<sup>1,2</sup>, Warren M. Potts<sup>2</sup>, Warwick H.H. Sauer<sup>2</sup>, Paul W.  
4 Shaw<sup>1,3</sup>

5 **Affiliations:** <sup>1</sup>Centre for Ecology, Evolution and Behavior, School of Biological  
6 Sciences, Royal Holloway University of London, Egham, U.K.; <sup>2</sup>Department of  
7 Ichthyology and Fisheries Science, Rhodes University, Grahamstown, South Africa;  
8 <sup>3</sup>Institute of Biological, Environmental and Rural Sciences (IBERS), Aberystwyth  
9 University, Aberystwyth, U.K.

10

11 **Corresponding author and present address:**

12 Romina Henriques

13 Evolutionary Genomics Group

14 Department of Botany and Zoology

15 Stellenbosch University - Private Bag X1

16 Matieland

17 7602

18 South Africa

19 Email: rhenriques@sun.ac.za

20

21 This work was supported by the Fundação para a Ciência e Tecnologia (FCT –  
22 Portugal), under Grant number SFRH/BD/36176/2007.

23 **Abstract**

24 The Benguela Current is considered to be a major biogeographical barrier for tropical  
25 and warm-temperate marine fish, but there is limited knowledge regarding its influence  
26 on population sub-structuring of cold-tolerant species. Employing genetic variation  
27 within the mitochondrial DNA Control Region and six cross-specific nuclear  
28 microsatellite markers, a preliminary study was conducted to investigate population  
29 substructuring in *Argyrosomus inodorus*, a highly exploited, cool-temperate migratory  
30 species, across the Benguela Current region. Results revealed evidence of incipient  
31 genetic differentiation (mtDNA  $\phi_{ST} = 0.092$ ; nuclear  $F_{ST} = 0.036$  and  $D_{ST} = 0.104$ ,  
32  $P < 0.005$ ) between the two sampling sites, suggesting the presence of two regional  
33 populations. Estimates of contemporary migration rates between populations were low,  
34 and similar in range to those reported in tagging surveys. Although preliminary, these  
35 results suggest that the oceanographic features of the Benguela Current may have  
36 influenced the evolutionary history of *A. inodorus*, and that the species is likely to be  
37 composed of two populations in the Benguela region. As the species is considered  
38 overexploited both in Namibia and South Africa, information on the distribution,  
39 population dynamics and long-term dispersal patterns across the Benguela Current  
40 region would provide a comprehensive evaluation of genetic structure, which should be  
41 incorporated into fishery management arrangements. .

42

43 **Key-words:** *Argyrosomus inodorus*, population structure, Benguela Current, isolation

44

45 **Introduction**

46 The Benguela Cold Current system, located in the southern Atlantic, features cold sea  
47 surface temperatures bounded to the north and south by tropical currents (the Angola  
48 and Agulhas currents, respectively), and a perennial upwelling cell off central Namibia  
49 that divides the region into two sub-systems with different characteristics (Shannon  
50 1985; Hutchings et al. 2009). The colder sea surface temperatures of the Benguela  
51 Current have been considered an important bio-geographical barrier, isolating tropical  
52 and warm-temperate fauna of the Atlantic and Indo-Pacific Oceans (Awise 2000; Floeter  
53 et al. 2008). However, recent studies revealed that other oceanographic features such as  
54 the perennial upwelling cell may also play an important role in shaping the population  
55 structure of warm-temperate fish populations within the Benguela system, as complete  
56 disruption of gene flow was documented both in *Lichia amia* and *Atractoscion*  
57 *aequidens* (Henriques et al. 2012; 2014). Little is known, however, regarding the  
58 influence of the Benguela system on genetic population connectivity of cold water  
59 tolerant species.

60 *Argyrosomus inodorus* is a migratory, benthopelagic sciaenid fish, endemic to the  
61 southeastern Atlantic (Griffiths & Heemstra 1995). Distribution range is restricted to  
62 cold-temperate waters (13°C-16°C), from the nearshore environment to depths of 100m,  
63 between Cape Frio and Meob Bay in Namibia, and between Cape Point and East  
64 London in South Africa (Griffiths & Heemstra 1995; Griffiths 1997; Kirchner &  
65 Holtzhausen 2001). The species distribution overlaps with those of the congeneric *A.*  
66 *coronus* in northern Namibia, and with *A. japonicus* along the southern and Eastern  
67 Cape coasts of South Africa (Griffiths & Heemstra 1995). Contrary to *A. inodorus*, *A.*  
68 *coronus* and *A. japonicus* are considered warm-temperate species, occurring  
69 preferentially in sea surface temperatures of 16°-19°C (Potts et al 2010) and 21°-25°C  
70 (Heemstra & Heemstra 2004), respectively. As *A. inodorus* is absent from the west  
71 coast of South Africa and there is no evidence for significant migration between the two  
72 areas of occurrence (Kirchner & Holtzhausen 2001), the species has been managed as  
73 two independent stocks. Life history characteristics appear to corroborate the hypothesis  
74 of two isolated and locally adapted populations, as features such as maximum size and  
75 size at maturity of Namibian and South African *A. inodorus* are significantly different,  
76 although no evidence of differentiation was observed within either region (Griffiths  
77 1997; Holtzhausen et al. 2001). *A. inodorus* is a critical component of multiple coastal

78 fishery sectors, and exploitation pressure throughout its distribution has led to the  
79 species becoming severely depleted, with spawning stocks estimated to be 69% of  
80 unexploited values (FAO 2012). In Namibia, *A. inodorus* is harvested by the  
81 commercial and recreational fishery sectors and although approximately the same  
82 numbers of fish are captured in each sector, the average size captured in the commercial  
83 sector is larger (Kirchner 1998). While *A. inodorus* is only regulated through an input  
84 control (number of permits) the recreational fishery catch, which comprises 70% of the  
85 total recreational catch, is regulated through bag- and size-limits (Kirchner & Beyer,  
86 1999; Holtzhausen et al. 2001). In South Africa, a 2012 survey reported that total  
87 landings of *A. inodorus* exceeded 400t per year (DAFF 2012). To establish sustainable  
88 management measures, it is necessary to understand how *A. inodorus* populations are  
89 structured across the Benguela Current region and whether migration between the two  
90 centers is absent. To date, no comprehensive genetic survey has been carried out in *A.*  
91 *inodorus*, with the exception of a genetic identification study in 1997 to differentiate  
92 between *A. inodorus* and *A. coronus*, based on allozymes (van der Bank & Kirchner  
93 1997), and a more recent study on shifts of abundance of these two species in central  
94 Namibia (Potts et al. 2014).

95 The distribution range and life history features of this species suggest that, as observed  
96 for warm-temperate species, the oceanographic features of the Benguela Current may  
97 influence the population structure and gene flow across the region. The aim of this study  
98 was to conduct a preliminary assessment of genetic diversity, population substructuring  
99 and connectivity between the two putative populations of *A. inodorus* across the  
100 Benguela Current, using both mitochondrial DNA (mtDNA) and nuclear microsatellite  
101 DNA markers, in order to test whether the regional oceanographic features influence  
102 population connectivity in this cold-temperate fish species.

103

## 104 **Methods**

### 105 **Sampling**

106 A total of 80 fish were captured by rod-and-line fishing from the shore, by local  
107 collaborators in two areas: the West Coast Recreational Area in Namibia ( $n = 40$ ) and  
108 the Eastern Cape Province in South Africa ( $n = 40$ ), representing the two centres of

109 abundance of the species (Figure 1). A clip of the pectoral fin was removed immediately  
110 after capture and stored in 95% ethanol.

111

## 112 Genetic screening

113 DNA extraction was performed using a standard phenol:chlorophorm method  
114 (Sambrook et al. 1989). Genetic variation was assessed as DNA sequence  
115 polymorphism in a fragment of the mtDNA Control Region (CR) and allele frequencies  
116 at six microsatellite loci isolated from *Argyrosomus japonicus* (Archangi et al. 2009). A  
117 total of 36 *A. inodorus* individuals were amplified by polymerase chain reaction (PCR)  
118 for CR, using primers and protocols of Appleyard et al. (2002). PCR products were  
119 purified with an enzymatic digestion, consisting of 0.5U of EXO1 (NewEngland  
120 biolabs) and 1u of shrimp alkaline phosphatase (SAP) in 1x supplied buffer  
121 (Fermentas), and sequenced in the forward direction using the same amplification  
122 primers, by Macrogen Inc. (South Korea). Sequences were visually inspected and a  
123 multiple alignment was performed in CLUSTAL X (Thompson et al. 1997), as  
124 implemented in BioEdit 7.0.1 (accession numbers: JX191998-2033).

125 Forty individuals per sampling site were screened at six microsatellite loci (UBA5,  
126 UBA40, UBA50, UBA91, UBA853 and UBA854). Optimized PCR mixes included 1x  
127 NH<sub>4</sub>Cl buffer, 2mM of MgCl<sub>2</sub>, 0.2mM of dNTPs, 0.5pmol of each primer, 0.2U of Taq  
128 polymerase (Bioline UK) and 50-100ng of extracted DNA, in a final volume of 10 µl.  
129 The Archangi et al. (2009) protocols were modified to ensure accurate amplification:  
130 annealing temperatures and number of cycles (UBA91 T<sub>a</sub> = 52°C, remaining loci T<sub>a</sub> =  
131 48°C, with 35 cycles), and removal of the final extension step of 72°C for 10min. PCR  
132 fragments from multiple loci were combined and genotyped on an AB3500 Genetic  
133 Analyzer (Applied Biosystems). Alleles were scored as PCR product size in base pairs,  
134 and scores were determined against an internal size marker (LIZ 600), using  
135 GeneMapper 4.0 (ABIPrism). In order to ensure accurate allele size scoring between  
136 runs, individuals with known allele sizes were used in each run as positive controls.

137

## 138 Data analyses

139 The CR dataset was assessed for levels of haplotype ( $h$ ) and nucleotide ( $\pi$ ) diversity,  
140 and fits to neutrality tests: Ewens-Watson's  $F$ , Tajima's  $D$  and Fu's  $FS$ , as  
141 implemented in ARLEQUIN (Excoffier et al. 2005). Determination of the most suitable  
142 nucleotide substitution model was performed in jModelTest (Posada, 2008). Preliminary  
143 inference of population connectivity of *A. inodorus* across the Benguela Current region  
144 was estimated as  $\phi_{ST}$  in ARLEQUIN (Excoffier et al. 2005), with a significance level of  
145  $P < 0.05$  determined by 10,000 permutations. Haplotype networks were reconstructed to  
146 evaluate intraspecific relationships among haplotypes, using the Median-Joining (MJ)  
147 algorithm implemented in NETWORK (Bandelt et al. 1999).

148 Microsatellite genotypic frequencies were tested for deviation from Hardy-Weinberg  
149 expectations of random mating and from linkage equilibrium, as implemented in  
150 GENEPOP (Raymond & Rousset 1995). The occurrence of amplification errors such as  
151 large allele drop out and stuttering, and estimation of null allele frequencies were  
152 assessed in MICROCHECKER (van Oosterhout et al. 2006). Levels of genetic diversity  
153 were estimated as number of alleles ( $Na$ ), allelic richness ( $AR$ ), observed and expected  
154 heterozygosity ( $H_O$  and  $H_E$ ), and Wright's inbreeding coefficient ( $F_{IS}$ ), in ARLEQUIN  
155 (Excoffier et al. 2005). A preliminary analysis to investigate the statistical power of the  
156 dataset to infer population substructuring was conducted in POWSIM (Ryman & Palm  
157 2006). Simulations were conducted for six loci and two populations ( $n = 40$ ,  $n = 40$ ),  
158 using the estimated allelic frequencies as the baseline for the ancestral population. Runs  
159 were performed using multiple combinations of effective population size ( $N_e$ ) and  
160 number of generations ( $t$ ) to generate a population differentiation of  $F_{ST} = 0.05$ ,  $F_{ST} =$   
161  $0.02$  and  $F_{ST} = 0.01$ . Each simulation was run for 1,000 replicates, and power was  
162 estimated as the proportion of tests that indicated significant genetic divergence (Ryman  
163 & Palm, 2006). Genetic differentiation was measured as Weir & Cockerham (1984)  $F_{ST}$   
164 estimator, as implemented in FreeNA (Chapuis & Estoup 2007), with significance and  
165 95% confidence intervals estimated after jackknifing. For comparison purposes, genetic  
166 differentiation was also measure using Jost's  $D_{est}$  estimator, which is independent of the  
167 levels of genetic diversity, in SMOGD (Crawford 2010). Contemporary estimates of  
168 long-term average migration rates between the two sampling sites were performed for  
169 the microsatellite dataset using two complementary approaches: the classical method  
170 based on  $F_{ST}$  values ( $F_{ST} = 1/(4N_{em} + 1)$ ) (Excoffier et al. 2005), and by employing the  
171 coalescent-based approach of MIGRATE (Beerli 2009). In MIGRATE, the Bayesian  
172 approach was implemented, enforcing a full migration model, with three replicates run

173 for each dataset (Beerli 2009). Each analysis was performed with four connected chains,  
174 using static heating (1,000,000, 3, 1.5, 1), a burn-in period of 10,000 steps, followed by  
175 90,000 steps, and parameters were recorded every 100 steps. Estimates of migration  
176 rates ( $m$ ) were obtained from  $M$  ( $M = m \cdot \mu$ ) and  $\Theta$  ( $\Theta = 4N_e\mu$ ) (Beerli, 2009). In order to  
177 obtain estimates of migration rates per generation (and not scaled by mutation) three  
178 general mutation rates were used: 0.1%, 0.5% and 1% per generation.

179

## 180 **Results**

### 181 Population structure and phylogeography

182 Sequencing of mtDNA CR yielded a fragment of 704 base pairs (bp). The 36  
183 individuals screened displayed 32 haplotypes defined by 34 variable nucleotide sites, of  
184 which 16 sites were parsimony informative. The Tamura-Nei nucleotide substitution  
185 model was identified as the most suitable for the mtDNA dataset. Haplotype diversity  
186 was high ( $h = 0.991$ ), whilst nucleotide diversity was low ( $\pi = 0.006$ ), with Namibian  
187 samples exhibiting higher values than the South African samples (Table 1). Deviations  
188 from the assumptions of selection neutrality were observed in Fu's  $FS$  for both  
189 populations, but not with either Ewens-Watsonson  $F$  or the Tajima's  $D$  tests (Table 1). As  
190 Fu's  $FS$  is known to be sensitive to abrupt demographic changes it is likely that the  
191 observed deviation to neutrality resulted from past population size changes, rather than  
192 reflecting selection effects. Genetic differentiation ( $\phi_{ST}$ ) between samples was low but  
193 statistically significant ( $\phi_{ST} = 0.092$ ,  $P < 0.05$ ), although haplotype relationships did not  
194 show an obvious geographical pattern (Figure 2): most individuals were represented by  
195 unique haplotypes with no obvious clustering of related haplotypes into Namibian or  
196 South African groups (Figure 2). The majority of haplotypes were closely related,  
197 differing from one another by one to two mutation steps, with the exception of two  
198 HEN individuals that were divergent by 10 mutation steps (Figure 2).

199 None of the six microsatellite loci exhibited evidence of amplification errors, and all  
200 displayed genotype frequencies that confirmed with Hardy-Weinberg and linkage  
201 equilibrium expectations (Table 2). Levels of genetic diversity in terms of  
202 heterozygosity and allelic richness were high (overall values of  $H_E = 0.774$  and  $AR =$   
203 13.7), with both samples displaying very similar values at individual loci and overall



204 (Table 2). Number of private alleles varied between 1 and 7, per locus and region (Table  
205 2). Analyses of statistical power of the dataset revealed that the loci and sample sizes  
206 used in this study could statistically detect genetic differentiation as low as  $F_{ST} = 0.001$   
207 in 99% of tests. As for the mtDNA data, nuclear genetic differentiation between the  
208 Namibian and South African samples was significantly greater than zero ( $F_{ST} = 0.036$ ,  
209  $P < 0.05$ ), with Jost's  $D_{est}$  indicating a slightly higher level of differentiation ( $D_{est} =$   
210  $0.104$ ,  $P < 0.05$ ). Estimates of contemporary migration rates per generation between the  
211 two geographical populations were low, independently of the method used or mutation  
212 rate considered ( $F_{ST}$ -based:  $N_{em} = 6$ ; MIGRATE:  $m_{2 \rightarrow 1} = 0.0014$ ;  $m_{1 \rightarrow 2} = 0.0011$  for  $\mu =$   
213  $0.1\%$  per generation).

214

## 215 Discussion

216 Despite the preliminary nature of the present study, due to the limited number of  
217 sampling sites available, similarly high levels of genetic diversity and evidence for  
218 shallow but significant genetic differentiation between the two regional populations  
219 (Namibia and South Africa) of *Argyrosomus inodorus* was found. The observed  
220 mitochondrial and nuclear genetic diversity ( $h = 0.991$ ,  $\pi = 0.006$ ;  $H_O = 0.771$ ,  $H_E =$   
221  $0.764$ ) were comparable with other commercially exploited fish species occurring in the  
222 Benguela Current region, such as *Argyrosomus japonicus* ( $h = 0.96$ ,  $\pi = 0.009$  –  
223 Klopper 2005), *Lichia amia* ( $h = 0.991$ ,  $\pi = 0.006$  – Henriques et al. 2012), *Atractoscion*  
224 *aequidens* ( $h = 0.853$ ,  $\pi = 0.005$ ;  $H_E = 0.889$  – Henriques et al. 2014) and  
225 *Rhabdosargus holubii* ( $h = 0.91$ ,  $\pi = 0.006$  – Oosthuizen 2007). High genetic diversity  
226 and shallow population structure are common features of marine teleosts, even in  
227 abundant, commercially exploited species. These are thought to result from historically  
228 high effective population sizes and/or high levels of gene flow between adjacent  
229 populations (Waples 1998). Interestingly, the observed genetic divergence between the  
230 Namibian and South African *A. inodorus* populations (mtDNA  $\phi_{ST} = 0.092$ ; nuclear  $F_{ST}$   
231  $= 0.036$  and  $D_{est} = 0.104$ ,  $P < 0.05$ ) was higher than that reported for other migratory  
232 sciaenids such as *Micropogonias undulatus* ( $\phi_{ST} = 0.046$  – Lankford et al. 1999) and  
233 *Sciaenops ocellatus* ( $\phi_{ST} = 0.057$  – Gold & Richardson 1998), but substantially lower  
234 than observed for other fish species with similarly disjunct distributions across the  
235 Benguela Current region (*L. amia*,  $\phi_{ST} = 0.9$  – Henriques et al. 2012; *A. aequidens*,  $\phi_{ST} =$

236 0.902,  $F_{ST} = 0.055$  – Henriques et al. 2014). These results, combined with estimates of  
237 the number of contemporary migrants ( $N_{em} = 0.0014 - 6$  per generation, depending on  
238 the method used), suggest a limited level of gene flow between Namibian and South  
239 African *A. inodorus* populations, and support the presence of incipient population  
240 differentiation. The present findings concur with tagging studies conducted for the  
241 species, where only two of 17,353 *A. inodorus* tagged in Namibia were recaptured in  
242 South Africa, suggesting that connectivity between populations may be limited  
243 (Kirchner & Holzhausen 2001). Therefore, the low but significant genetic  
244 differentiation displayed by *A. inodorus* is likely to result from a present-day disjunct  
245 population distribution, with occasional migrants, and historically high levels of  
246 effective population size, rather than substantial gene flow between Namibia and South  
247 Africa.

248 As with other fish species distributed around southwestern Africa (e.g. *L. amia* –  
249 Henriques et al. 2012; *A. aequidens* – Henriques 2012; *Albula* spp. – Colborn et al.  
250 2001), the distribution break in *A. inodorus* appears to correspond with the areas of cold  
251 water upwelling off southern Namibia and the west coast of South Africa (Griffiths &  
252 Heemstra, 1995; Griffiths 1997; Kirchner & Holtzhausen 2001). Although the limited  
253 sampling precludes the drawing of definitive conclusions, the reported genetic  
254 divergence and breakdown of gene flow across the Benguela Current suggests that the  
255 oceanographic features of the system, namely the cold water region, may be  
256 contributing to disrupt both adult and larval dispersal of *A. inodorus*, and supports the  
257 hypothesis of two isolated populations with limited migration between them. As the  
258 species is considered overexploited both in Namibia and South Africa, information on  
259 the distribution, population dynamics and long-term dispersal patterns across the  
260 Benguela Current region would provide a comprehensive evaluation of genetic  
261 structure, which should be incorporated into fishery management arrangements.

262

## 263 **Acknowledgments**

264 The present work and R. Henriques were supported by a Fundação da Ciência e  
265 Tecnologia – FCT grant (ref. SFRH/BD/36176/2007). The authors would like to thank  
266 Steve “Spyker” Krugger for assisting in sample collection in Namibia, and the

267 University of Namibia, and the Ministry of Fisheries in Namibia, for their invaluable  
268 support.

269

## 270 **References**

271 Appleyard SA, Ward RD, Williams R. 2002. Population structure of the Patagonian  
272 toothfish around Heard, McDonald and Macquarie Islands. *Antarctic Science*  
273 14:364-373.

274 Archangi B, Chand V, Mather PB. 2009. Isolation and characterization of 15  
275 polymorphic microsatellite DNA loci from *Argyrosomus japonicus* (mulloway),  
276 a new aquaculture species in Australia. *Molecular Ecology Resources* 9:412-  
277 414.

278 Avise JC. 2000. *Phylogeography: The History and Formation of Species*. Cambridge:  
279 Harvard University Press. 447 pages.

280 Bandelt HJ, Forster P, Rohl A. 1999. Median-joining networks for inferring  
281 intraspecific phylogenies. *Molecular Biology and Evolution* 16:37-48.

282 Beerli P. 2009. How to use MIGRATE or why are Markov chain Monte Carlo programs  
283 difficult to use? In: Bertorelle G, Bruford MW, Hauffe HC, Rizzoli A, Vernesi  
284 C, editors. *Population Genetics for Animal Conservation*. Cambridge:  
285 Cambridge University Press, p 42-79.

286 Chapuis MP, Estoup A. 2007. Microsatellite null alleles and estimation of population  
287 differentiation. *Molecular Biology and Evolution* 24:621-631.

288 Colborn J, Crabtree RE, Shaklee JB, Pfeiler E, Bowen BW. 2001. The evolutionary  
289 enigma of bonefishes (*Albula* spp.): cryptic species and ancient separations in a  
290 globally distributed shorefish. *Evolution* 55:807-820.

291 Crawford NG. 2010. SMOGD: software for the measurement of genetic diversity.  
292 *Molecular Ecology Resources* 10:556-557.

293 DAFF. 2012. *Status of the South African marine fishery resources 2012*. Cape Town:  
294 DAFF. 71 pages.

295 Excoffier L, Laval G, Schneider S. 2005. ARLEQUIN (version 3.0): An integrated  
296 software package for population genetics data analysis. *Evolutionary*  
297 *Bioinformatics* 1:47-50.

298 FAO. 2012. *The State of World Fisheries and Aquaculture*. Rome: FAO. 209 pages.

299 Floeter SR, Rocha LA, Robertson DR, Joyeux JC, Smith-Vaniz WF, Edwards AJ et al.  
300 2008. Atlantic reef fish biogeography and evolution. *Journal of Biogeography*  
301 35:22-47.

302 Gold JR, Richardson LR. 1998. Mitochondrial DNA diversification and population  
303 structure in fishes from the Gulf of Mexico and western Atlantic. *Journal of*  
304 *Heredity*, **89**, 404-414.

305 Griffiths MH. 1997. The life history and stock separation of silver kob, *Argyrosomus*  
306 *inodorus*, in South African waters. *Fishery Bulletin* 95:47-67.

307 Griffiths MH, Heemstra PC. 1995. A contribution to the taxonomy of the marine fish  
308 genus *Argyrosomus* (Perciformes: Sciaenidae), with description of two new  
309 species from southern Africa. *Ichthyological Bulletin of the J.L.B. Smith*  
310 *Institute of Ichthyology* 1-40.

311 Heemstra P, Heemstra E. 2004. *Coastal Fishes of Southern Africa*. Grahamstown:  
312 NISC, SAIAB. 488 pages.

313 Henriques R, Potts WM, Sauer WHH, Shaw PW. 2012. Evidence of deep genetic  
314 divergence between populations of an important recreational fishery species,  
315 *Lichia amia* L. 1758, around southern Africa. *African Journal of Marine Science*  
316 34:585-591.

317 Henriques R, Potts WM, Sauer WHH, Santos CV, Shaw PW. 2014. Population  
318 connectivity and phylogeography of a coastal fish, *Atractoscion aequidens*  
319 (Sciaenidae), across the Benguela Current region: evidence of an ancient  
320 vicariant event. *PLOS One* 9(2): e87907. doi:10.1371/journal.pone.0087907

321 Holtzhausen JA, Kirchner CH, Voges SF. 2001. Observations on the linefish resources  
322 of Namibia, 1990-2000, with special reference to West Coast steenbras and  
323 silver kob. *South African Journal of Marine Science-Suid-Afrikaanse Tydskrif*  
324 *Vir Seewetenskap* 23:135-144.

325 Hutchings L, Van der Lingen CD, Shannon LJ, Crawford RJM, Verheye HMS,  
326 Bartholomae CH, et al. 2009. The Benguela Current: An ecosystem of four  
327 components. *Progress in Oceanography* 83:15-32.

328 Kirchner CH. 1998. Population dynamics of the exploited silver kob (*Argyrosomus*  
329 *inodorus*) in Namibian waters. PhD thesis, University of Port Elizabeth, South  
330 Africa. 276 pages.

331

- 332 Kirchner CH, Beyer JE. 1999. Estimation of total catch of silver kob *Argyrosomus*  
333 *inodorus* by recreational shore-anglers in Namibia using a roving-roving creel  
334 survey. South African Journal of Marine Science-Suid-Afrikaanse Tydskrif Vir  
335 Seewetenskap 21:191-199.
- 336 Kirchner CH, Holtzhausen JA. 2001. Seasonal movements of silver kob, *Argyrosomus*  
337 *inodorus*, (Griffiths and Heemstra) in Namibian waters. Fisheries Management  
338 and Ecology 8:239-251.
- 339 Klopper AW. 2005. Intraspecific genetic variation in the percoid teleosts, *Argyrosomus*  
340 *japonicus* (Temminck & Schlegel, 1843) and *Pomadasys commersonii*  
341 (Lacepede, 1801) as inferred from the mitochondrial control region. Master  
342 Thesis. University of Pretoria, South Africa. 144 pages.
- 343 Lankford TE, Targett TE, Gaffney PM. 1999. Mitochondrial DNA analysis of  
344 population structure in the Atlantic croaker, *Micropogonias undulatus*  
345 (Perciformes : Sciaenidae). Fishery Bulletin 97:884-890.
- 346 Oosthuizen CJ. 2007. Genetic variation within the Cape Stumpnose *Rhabdosargus*  
347 *holubii* Steindachner (Teleostei: Sparidae). Master Thesis. University of  
348 Pretoria, South Africa. 182 pages.
- 349 Posada D. 2008. jModelTest: Phylogenetic model averaging. Molecular Biology and  
350 Evolution 25:1253-1256.
- 351 Potts WM, Henriques R, Santos CV, Munnik K, Ansorge I, Dufois F, Booth AJ,  
352 Kirchner C, Sauer WHH, Shaw PW. 2014. Ocean warming, a rapid  
353 distributional shift and the hybridization of a coastal fish species. Global Change  
354 Biology, DOI: 10.1111/gcb.12612.
- 355 Raymond M, Rousset F. 1995. GENEPOP version-1.2: Population genetics software for  
356 exact tests and ecumenicism. Journal of Heredity 86:248-249.
- 357 Ryman N, Palm S. 2006. POWSIM: a computer program for assessing statistical power  
358 when testing for genetic differentiation. Molecular Ecology Notes 6:600-602.
- 359 Sambrook J, Fritscher EF, Maniatis T. 1989. Molecular cloning: a laboratory manual.  
360 New York: Cold Spring Harbor Laboratory Press. 2028 pages.
- 361 Shannon LV. 1985. The Benguela ecosystem. 1. Evolution of the Benguela, physical  
362 features and processes. Oceanography and Marine Biology 23:105-182.
- 363 Thompson JD, Gibson TJ, Plewniak F, Jeanmougin F, Higgins DG. 1997. The  
364 CLUSTAL\_X windows interface: flexible strategies for multiple sequence

365 alignment aided by quality analysis tools. *Nucleic Acids Research* 25:4876-  
366 4882.

367 Van der Bank H, Kirchner CH. 1997. Biochemical genetic markers to distinguish two  
368 sympatric and morphologically similar Namibian marine fish species,  
369 *Argyrosomus coronus* and *Argyrosomus inodorus* (Perciformes: Sciaenidae).  
370 *Journal of African Zoology* 111: 441-448.

371 Van Oosterhout C, Weetman D, Hutchinson WF. 2006. Estimation and adjustment of  
372 microsatellite null alleles in nonequilibrium populations. *Molecular Ecology*  
373 *Notes* 6:255-256.

374 Waples RS. 1998. Separating the wheat from the chaff: Patterns of genetic  
375 differentiation in high gene flow species. *Journal of Heredity* 89:438-450.  
376

377 **Tables**

378

379 **Table 1:** Estimates of mitochondrial genetic diversity levels and neutrality tests for *A.*  
380 *inodorus* CR: *n* – number of individuals; *H* – number of haplotypes; *h* – haplotype  
381 diversity;  $\pi$  - nucleotide diversity; *F* – Ewens-Waterson neutrality test; *D* – Tajima  
382 neutrality test; *FS* – Fu neutrality test. Significant departures from expectations (*P* <  
383 0.05) in bold.

	HEN	EastC	Overall
<i>n</i>	18	18	36
<i>H</i>	18	14	32
<i>h</i>	1.000	0.968	0.991
$\pi$	0.008	0.004	0.006
<i>F</i>	-	0.862	0.966
<i>D</i>	-1.486	0.324	-1.554
<i>F<sub>s</sub></i>	<b>-14.762</b>	<b>-10.099</b>	<b>-25.652</b>

384

385

387 **Table 2:** Genetic diversity in *A. inodorus* samples at six cross-specific microsatellite  
 388 loci:  $n$  – number of individuals genotyped;  $NA$  – number of alleles;  $AR$  – allelic  
 389 richness;  $PA$  – number of private alleles  $H_E$  – expected heterozygosity;  $H_O$  – observed  
 390 heterozygosity;  $F_{IS}$  – inbreeding coefficient. Significant deviations to Hardy-Weinberg  
 391 expectations in bold.

	<b>HEN</b>	<b>EastC</b>	<b>Overall</b>	
	$n$	40	40	80
	$NA$	11	11	13
	$AR$	10.803	10.925	10.52
<b>UBA5</b>	$PA$	1	2	3
	$H_E$	0.819	0.825	0.839
	$H_O$	0.875	0.825	0.850
	$F_{IS}$	-0.047	0.003	-0.007
	$n$	40	39	79
	$NA$	8	7	8
	$AR$	7.951	7.000	7.452
<b>UBA40</b>	$PA$	1	0	1
	$H_E$	0.765	0.807	0.790
	$H_O$	0.750	0.795	0.772
	$F_{IS}$	0.038	0.037	0.028
	$N$	39	40	79
	$NA$	14	15	16
	$AR$	13.974	14.899	14.846
<b>UBA50</b>	$PA$	1	2	3
	$H_E$	0.887	0.896	0.914
	$H_O$	0.821	0.800	0.810
	$F_{IS}$	0.066	0.038	0.120

	<i>n</i>	40	40	80
	<i>NA</i>	5	3	5
	<i>AR</i>	4.902	3.000	3.962
<b>UBA91</b>	<i>PA</i>	1	1	2
	<i>H<sub>E</sub></i>	0.361	0.387	0.375
	<i>H<sub>O</sub></i>	0.275	0.475	0.375
	<i>F<sub>IS</sub></i>	0.182	-0.207	0.006
	<i>n</i>	40	40	80
	<i>NA</i>	13	14	17
	<i>AR</i>	12.799	12.924	14.530
<b>UBA853</b>	<i>PA</i>	3	4	7
	<i>H<sub>E</sub></i>	0.831	0.872	0.876
	<i>H<sub>O</sub></i>	0.925	0.900	0.913
	<i>F<sub>IS</sub></i>	-0.100	-0.036	-0.035
	<i>n</i>	40	40	80
	<i>NA</i>	9	7	19
	<i>AR</i>	7.604	11.899	15.152
<b>UBA854</b>	<i>PA</i>	7	1	8
	<i>H<sub>E</sub></i>	0.881	0.776	0.860
	<i>H<sub>O</sub></i>	0.975	0.675	0.825
	<i>F<sub>IS</sub></i>	-0.038	0.130	0.047
	<i>n</i>	40	40	80
	<i>NA</i>	10	9.500	11.333
	<i>AR</i>	9.004	9.833	13.667
<b>Average all loci</b>	<i>PA</i>	14	10	24
	<i>H<sub>E</sub></i>	0.757	0.760	0.776
	<i>H<sub>O</sub></i>	0.770	0.745	0.758
	<i>F<sub>IS</sub></i>	0.000	0.012	0.027

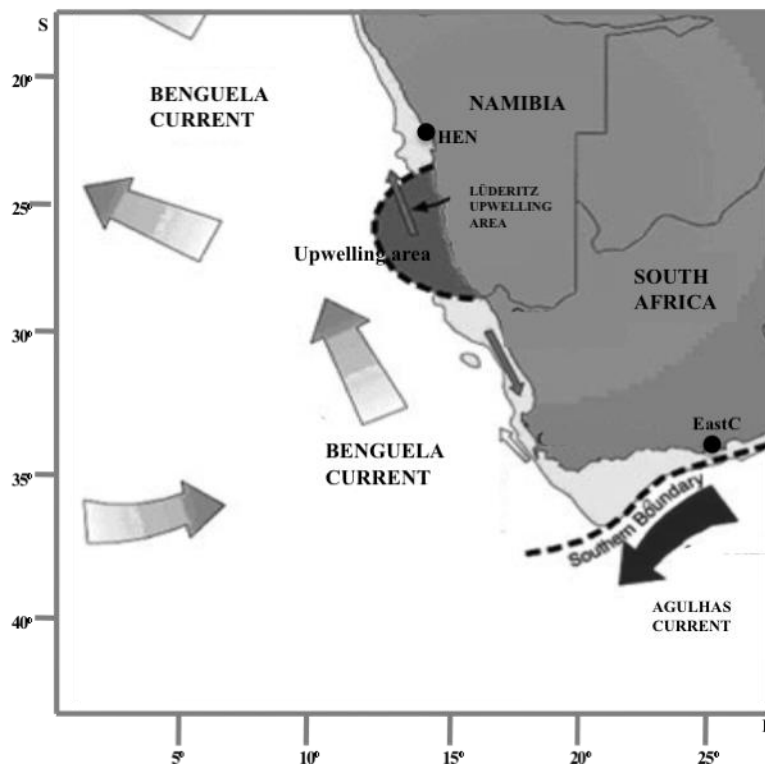




393 **Figure Legends**

394

395 **Figure 1:** Sampling strategy for *A. inodorus* across the Benguela Current region,  
396 highlighting sampling sites, and their position relative to the major oceanographic  
397 features of the system: position of the Benguela and Agulhas Currents, central Namibia  
398 upwelling cell, and continental platform width.



399

400

401

402

403

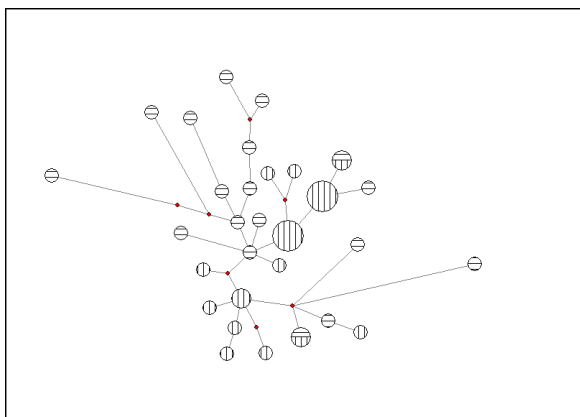
404

405

406

407

408 **Figure 2:** Haplotype network for *A. inodorus* across the Benguela Current region, based  
409 on 704bp of mtDNA CR sequences:  $\ominus$  = HEN;  $\oplus$  = EastC. Branch lengths are  
410 proportional to number of nucleotide differences, and node sizes are proportional to the  
411 number of individuals. Red dots represent unsampled inferred haplotypes.



412