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Reconstructing Disturbances and Their Biogeochemical Consequences over Multiple Timescales

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Key words:	disturbance ecology, landscape ecology, nutrient cycling, plant ecology
Abstract:	Ongoing changes in disturbance regimes are predicted to cause acute changes in ecosystem structure and function in coming decades, but many aspects of these predictions are uncertain. A key challenge is to improve the predictability of post-disturbance biogeochemical trajectories at the ecosystem level. Both ecosystem ecologists and paleoecologists have generated complementary datasets about disturbance (type, severity, frequency) and ecosystem response (net primary productivity, nutrient cycling) spanning decadal to multi-millennial timescales. Here, we take the first steps toward a full integration of these datasets by: (1) reviewing how disturbances are reconstructed using dendrochronological and sedimentary archives, and (2) summarizing the conceptual frameworks for carbon, nitrogen, and hydrologic responses to disturbances. Key research priorities include further development of paleoecological techniques that reconstruct both disturbances and terrestrial ecosystem dynamics. Additionally, mechanistic detail from disturbance experiments, long-term observations, and chronosequences can help to increase the temporal understanding of ecosystem resilience.

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Manuscripts

Reconstructing disturbances and their biogeochemical consequences over multiple timescales

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8 49 **Abstract**

9 50
10 51 Ongoing changes in disturbance regimes are predicted to cause acute changes in
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12 52 ecosystem structure and function in coming decades, but many aspects of these
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14 53 predictions are uncertain. A key challenge is to improve the predictability of post-
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16 54 disturbance biogeochemical trajectories at the ecosystem level. Both ecosystem
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18 55 ecologists and paleoecologists have generated complementary datasets about
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20 56 disturbance (type, severity, frequency) and ecosystem response (net primary
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22 57 productivity, nutrient cycling) spanning decadal to multi-millennial timescales. Here,
23
24 58 we take the first steps toward a full integration of these datasets by: (1) reviewing
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26 59 how disturbances are reconstructed using dendrochronological and sedimentary
27
28 60 archives, and (2) summarizing the conceptual frameworks for carbon, nitrogen, and
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30 61 hydrologic responses to disturbances. Key research priorities include further
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32 62 development of paleoecological techniques that reconstruct both disturbances and
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34 63 terrestrial ecosystem dynamics. Additionally, mechanistic detail from disturbance
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36 64 experiments, long-term observations, and chronosequences can help to increase the
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38 65 temporal understanding of ecosystem resilience.
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47 66
48 67 **Keywords**

49 68 ecosystem ecology, disturbance, fire regime, nitrogen cycling, resilience
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3 71
4 72 **Introduction**
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7 73 Ecological disturbances in terrestrial systems are discrete events that reduce
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9 74 primary producer biomass and strongly regulate material and energy flows. Several
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11 75 studies document increasing size and/or severity of a variety of disturbance types,
12
13 76 including beetle outbreaks, storms, and wildfires during recent decades (Seidl et al.
14
15 77 2011). This amplification is raising questions about which disturbances are
16
17 78 unprecedented, and what conditions may lead to threshold behavior in ecosystem
18
19 79 responses. It has been particularly difficult to identify incipient shifts in disturbance
20
21 80 regimes (Turner 2010). The synergistic effects of interactions among disturbance
22
23 81 characteristics, life history parameters, and climate change are of particular concern
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25 82 because disturbances have potentially substantial consequences for carbon and
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27 83 nutrient cycling, and hydrology.
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34 84 We cannot properly evaluate modern disturbance events without the baseline
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36 85 information provided by decadal- to multimillennial-scale records of disturbances
37
38 86 and ecosystem responses. Such long-term records are derived primarily from tree-
39
40 87 ring and sedimentary archives, and they allow the reconstruction of past
41
42 88 disturbance events and regimes. When combined with both detailed observational
43
44 89 data about individual events and with novel proxies of ecosystem processes, the
45
46 90 biogeochemical consequences of changing disturbance regimes can begin to be
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48 91 assessed. An important challenge now is to integrate these two approaches with
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50 92 respect to ecological processes and spatiotemporal patterns.
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3 93 A variety of disturbance agents, such as fire, wind, drought, or insect outbreaks, are
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5 94 familiar to most ecologists as regulating short-term ecosystem processes, such as
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8 95 nutrient cycling and carbon storage (Pickett and White 1985). These same
9
10 96 disturbances can also determine long-term ecosystem trajectories. For example,
11
12 97 without soil-replenishing disturbance, productivity may decrease and ecosystems
13
14 98 may retrograde (Wardle et al. 2004). Chronic disturbance, however, can also lead to
15
16 99 long-term reduction of productivity through nutrient depletion (Boerner 1982).
17
18 100 Although the importance of disturbance to ecosystem functioning is recognized, the
19
20 101 biogeochemical consequences of disturbances have been particularly difficult to
21
22 102 quantify over long timescales. An increasing amount of complementary research
23
24 103 now provides the basis for understanding decadal- to millennial-scale ecosystem
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26 104 change, including high temporal-resolution post-disturbance chronosequences,
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28 105 innovative techniques applied to sediment and tree-ring records to reconstruct past
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30 106 ecosystem change, and long-term ecological measurements of several decades in
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32 107 duration.
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39 108 The integration of ecosystem processes across multiple timescales is crucial for
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41 109 identifying the biogeochemical consequences of disturbances. Ecosystem ecologists
42
43 110 working on the modern landscape (“neo-ecologists”) have had success in
44
45 111 quantifying short-term responses to disturbance through direct measurements
46
47 112 (Mack et al. 2011). In contrast, paleoecologists have been successful at quantifying
48
49 113 past disturbance regimes to contextualize modern disturbances (Marlon et al. 2012).
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52 114 The long-term perspective that paleorecords provide has proven increasingly useful
53
54
55 115 to conservation management (Willis et al. 2007), ecosystem service analyses
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3 116 (Dearing et al. 2012), and restoration ecology. There is now potential to use records
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6 117 of disturbance events and regimes to calculate the biogeochemical consequences of
7
8 118 disturbance across climate types, vegetation types, and soil development stages.
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10
11 119 Integrating the two views— paleo- and neo-ecosystem ecology— is a challenge due
12
13 120 to differences between disciplines in methodology, timescale, and spatial extent. Yet
14
15 121 this integration is crucial for understanding the dynamic nature of disturbance and
16
17 122 potential feedbacks in the Earth system among climate change, disturbance, and
18
19 123 biogeochemistry. Without additional information from paleorecords, extrapolating
20
21 124 modern ecosystem responses across time and space will likely be ineffective, and
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23 125 direct measurements of ecosystem response are needed to interpret the indirect
24
25 126 proxies available in paleorecords. Here, we take the first steps toward this
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27 127 integration by (1) reviewing how disturbances are reconstructed using
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29 128 dendrochronological and sedimentary archives, (2) summarizing the conceptual
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31 129 frameworks for carbon, nitrogen, and hydrologic responses to disturbances, and (3)
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33 130 identifying key research priorities for linking the mechanistic knowledge of
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35 131 disturbance from neo-ecology with the long-term perspective granted by
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37 132 paleoecology.
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134 **Reconstructing past disturbance**

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49 135 A variety of disturbance events can be reconstructed from annual to multi-
50
51 136 millennial timescales (Table 1). Disturbance agents particularly important in mid-
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53 137 latitude terrestrial ecosystems are fire, drought, and insect outbreaks. From
54
55 138 paleorecords, it is possible to reconstruct disturbance agent, severity, spatial
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3 139 location, and spatial extent. We focus here on records available from tree-rings and
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5 140 lacustrine sediment sequences. Long-term instrumental data are helpful for
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8 141 calibration of potential proxy records. The most robust reconstructions use either
9
10 142 multiple approaches or a detailed calibration dataset.

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13 143 Observational, remote-sensing, and historical records of disturbance offer detailed
14
15 144 accounts of event timing, location, and ecosystem impacts, but they typically span
16
17 145 only decades or centuries at most (Raffa et al. 2008). Tree-ring records can extend
18
19 146 this timeframe by providing spatially-precise, sub-annually resolved records of low-
20
21 147 and high-severity fires (Falk et al. 2011), insect outbreaks (Veblen et al. 1994), and
22
23 148 drought (Cook et al. 2004) across a range of forest types. Proxy variables from
24
25 149 sedimentary records can provide disturbance reconstructions on even longer
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28 150 timescales, such as several millennia, but generally have lower spatial and temporal
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31 151 resolution.

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35 152 *Fire.* Wildfires are the most commonly reconstructed disturbance events from
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37 153 sediment records that span centuries to millennia (Gavin et al. 2007, Higuera et al.
38
39 154 2010). Two aspects of past fire occurrence are focal points for these reconstructions,
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41 155 spanning two spatial scales. Stand-level or “local” fire occurrence may be
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43 156 interpreted from distinct charcoal peaks in high-resolution sediment records of
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45 157 macroscopic charcoal. For example, fire events in a coniferous montane forest were
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47 158 reconstructed from macroscopic charcoal influx to a small lake basin (figure 1).
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50 159 Several fire regimes can be distinguished over a 6,000 year time period. Regional-
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53 160 level biomass burning or area burned is inferred from standardized summaries of
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55 161 total charcoal accumulation in a record or network of records, and quantified as an
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3 162 index (Marlon et al. 2012), or area burned (Higuera et al. 2011). Biomass burning
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5 163 can then be linked to independent proxies for vegetation to infer the fuel types and
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8 164 fire severity.
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11 165 Fire history investigations based on tree-ring records employ dating of fire scars,
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13 166 stand establishment, or both depending on the type of wildfire regime being
14
15 167 investigated. Fire scar networks in western North America have been particularly
16
17 168 effective for identifying past landscape and climate drivers of low-intensity fire
18
19 169 events (Falk et al. 2011). Mixed severity fire regimes can be detected with
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21 170 dendrochronological techniques along with reconstructions of fire frequency and
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23 171 area burned (Hessl 2011).
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28 172 *Drought*. Paleorecords can also provide accurate reconstructions of drought
29
30 173 conditions. Drought is defined in multiple ways based on spatial and temporal
31
32 174 changes in precipitation, frequency of precipitation events, and severity of impacts.
33
34 175 Relatively short (< 200 years) instrumental time series provide data on streamflow,
35
36 176 precipitation, soil moisture, Palmer Drought Severity Index, flood events, and lake
37
38 177 levels (Biondi and Strachan 2012). Dendrohydrology uses growth records from
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40 178 long-lived tree species to extend the contemporary instrumental hydrologic records
41
42 179 one to two orders of magnitude (Stine 1994).
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48 180 Lake sediment cores further extend the temporal scale of hydrologic reconstruction
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50 181 to millennial and longer timeframes. From sedimentary records, droughts can be
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52 182 reconstructed from changes in lake level, diatom communities, and geochemical
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54 183 proxies (Laird et al. 2003). These paleolimnological methods have been most
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3 184 commonly applied to closed-basin lakes in arid or semi-arid landscapes where
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6 185 changes in lake-level and salinity are closely related to shifts in hydrologic balance.
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9 186 *Insect outbreaks*. During recent decades, irruptive populations of native bark beetles
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11 187 (*Dendroctonus* spp., *Ips* spp.) have exceeded the spatial scale and intensity of past
12
13 188 outbreaks by an order of magnitude (Raffa et al. 2008). In western North America,
14
15 189 these disturbances have been attributed to the complex interaction of accelerating
16
17 190 insect reproductive cycles related to warming air temperatures, decreased tree
18
19 191 vigor from moisture deficiencies and concomitant pathogen infestation, and high
20
21 192 stand densities of suitably-sized host trees (Mitton and Ferrenberg 2012). Tree-ring
22
23 193 reconstructions provide evidence that native bark beetles are an essential
24
25 194 component of ecosystem function and promote forest regeneration (Sherriff et al.
26
27 195 2011). However, what is believed to be the unprecedented scale and severity of
28
29 196 recent outbreaks suggests that neither the dynamics of beetle disturbances nor their
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31 197 biogeochemical impacts are well understood (Rhoades et al. 2013). Longer-term
32
33 198 ecological records would help to contextualize the recent episodes. For example,
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35 199 tree-ring records indicate that the recent mountain pine beetle epidemic in central
36
37 200 British Columbia was preceded by at least two 19th century outbreaks (Hrinkevich
38
39 201 and Lewis 2011). The need to extend the temporal record of past bark beetle
40
41 202 disturbances is currently being pursued using lake sediments. Proxy and direct
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43 203 evidence, including pollen and plant and insect macrofossils, has been useful to
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45 204 reconstruct past insect disturbances over centennial to millennial timescales
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47 205 (Morris and Brunelle 2012).
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206 **Reconstructing biogeochemistry**

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3 207 In addition to reconstructing disturbance regimes, the paleoecological record has
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6 208 the potential to provide key information on ecosystem pools and fluxes. This
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9 209 information is provided through proxy records that range from standard
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11 210 paleoecological techniques to newly-developed methods in lacustrine sediment
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13 211 cores (Smol 2001). The pollen record indicates plant community composition, and
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15 212 pollen influx can provide an index of terrestrial biomass. Organic matter content of
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17 213 lake sediment is an indicator of combined lake productivity and detrital terrestrial
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19 214 sources; separating these sources is possible through the carbon:nitrogen ratio.
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21 215 Stable isotopes of carbon and nitrogen also relate to the cycling and productivity
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23 216 within the aquatic and terrestrial ecosystems. The toolbox of methods is growing
24
25 217 rapidly. Newer methods include quantifying the elemental composition of
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27 218 sediments to identify sources of mineral matter, measuring compound-specific
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29 219 isotopes that relate to sources and types of organic matter (e.g., terrestrial vs.
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31 220 aquatic), and using additional proxies such as magnetic properties to interpret
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33 221 geochemical changes.
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40 222 **The biogeochemistry of disturbance**

41 223 **Carbon**

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45 226 The cycling of carbon (C) is a fundamental characteristic of ecosystem functioning,
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47 227 and it is highly sensitive to disturbance. In the absence of change in the disturbance
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49 228 regime, C storage in terrestrial ecosystems is constrained by stable rates of
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51 229 accumulation and degradation. Ecosystems therefore have a characteristic C
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53 230 carrying capacity that may fluctuate around this capacity in time and space, i.e. the
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55 231 shifting mosaic steady state (Bormann and Likens 1979). Thus, over decadal
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3 232 timescales the net ecosystem C balance (NECB) trends toward zero (Chapin et al.
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5 233 2006), assuming no directional changes in ecosystem composition and/or climate.
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8 234 However, changes in the disturbance regime can lead to dramatic shifts in C cycling
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10 235 and NECB. Often these changes are realized over long time frames (hundreds to
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12 236 thousands of years) that require a broad and integrated temporal perspective
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14 237 available only from paleorecords or chronosequences. Regional-scale vegetation
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16 238 shifts have the potential to influence the global C balance. For example, 96 Tg of C
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18 239 accumulated in boreal peatlands during the Holocene (Dean and Gorham 1998).
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21 240 Thus, one key challenge is to detect and understand how fundamental shifts in
22
23 241 disturbance regimes can alter C cycling through ecosystem reorganizations.
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26 242 Changes in C status may be triggered by unusually severe events, compound
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28 243 disturbances, or major changes in disturbance regimes that exceed the ecological
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30 244 resilience of the system and trigger changes in the vegetation and soil. Therefore
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32 245 understanding long-term C dynamics in ecosystems requires integration of
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34 246 contemporary and paleoecological approaches. Studies of the modern landscape are
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36 247 constrained by their narrow temporal context, without baseline information and
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38 248 past histories to fully interpret measured recent changes in NECB. Thus, it is difficult
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40 249 to determine if recently-observed changes have precedent or if they represent novel
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42 250 conditions.
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45 251 Paleoecological studies capture the long-term patterns in disturbance regimes, state
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47 252 changes in the vegetation, and coarse measures of C stores. The example
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49 253 sedimentary sequence from an old-growth conifer forest shown in figure 1
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51 254 simultaneously records organic content, changes in fire regimes, and terrestrial
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3 255 vegetation composition, in some cases in response to rapid climate change. But the
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6 256 nature of proxy records is indirect, and the temporal and spatial resolution is often
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9 257 too low to identify precise mechanisms or consequences of change.
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11 258 For many modern ecosystems, C lost during a disturbance event is largely recovered
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13 259 before the next disturbance, and NECB can be considered relatively stable. The
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15
16 260 initial driver of this trajectory is C-fixation in aboveground live biomass such as
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19 261 wood, with relatively linear rates of C accumulation in the first few decades after a
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21 262 disturbance event (figure 2). However, the subsequent redistribution of this C into
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23 263 soil pools with longer turnover times can take centuries, and these slower processes
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25
26 264 lead to a saturation of C accumulation rate on those timescales. When comparing
27
28 265 data from post-disturbance chronosequences with the paleoecological record, it is
29
30 266 clear that both compound disturbances and state changes in the disturbance regime
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32
33 267 alter these trajectories. Additionally, some of these state changes occurred rapidly
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35 268 (i.e. a shift from forest to grassland in less than 300 years) (Williams et al. 2010).
36
37 269 There are also contemporary examples of incipient shifts in the disturbance regime,
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39
40 270 and model results that predict futures with drastically different C balances (Bond et
41
42 271 al. 2005). In short, changes in the disturbance regime can drive state changes in
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44
45 272 ecosystem structure and function with respect to C storage.

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48 273 We present a conceptual model of how NECB and disturbance interval interact over
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50 274 long timescales, and how this would be seen in a sedimentary paleorecord (figure 3).
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53 275 Our current understanding suggests that a single disturbance event, even one of
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55 276 high severity such as a stand-replacing wildfire, will not necessarily alter long-term
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57 277 C balance. Net Primary Productivity (NPP) can keep pace with periodic disturbance,
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3 278 and NECB is stable. We recognize that recovery patterns will vary, but we expect
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6 279 that these two processes will usually balance, as illustrated by the 1:1 line that
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8 280 compares the time to recover the C lost with the time since disturbance as 1:1
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10
11 281 (figure 3b). Observations greater than the 1:1 line imply gains in NECB and slopes
12
13 282 less than 1 imply losses in NECB. Conceptually we expect ecosystems with a
14
15 283 consistent disturbance regime and ecological community to fluctuate around this
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17
18 284 1:1 line.

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21 285 Severe, progressive, and/or continual deviations away from this 1:1 ratio would
22
23 286 indicate a potential biogeochemical state change (figure 3). We set the 1:1 line as
24
25 287 the neutral state where C dynamics and disturbance regime are in balance.

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28 288 Perturbations of either the disturbance regime or time to C recovery would drive
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30 289 the system away from the 1:1 line. When there is a state shift, there is a potential for
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33 290 changes in the C balance compared to the previous system. It is important to note
34
35 291 that there is no assumption that the baseline is constant. Indeed, the value of this
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37 292 conceptual figure is to explicitly acknowledge that the baseline changes with a shift
38
39 293 in the biogeochemical state. Also the interpretation of the ratio of time to
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41
42 294 disturbance: time to C recovery is not interpreted in isolation (Turner et al. 1993).
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45 295 Other information from the paleoecological record (e.g., charcoal-inferred change in
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47 296 fire frequency, pollen-inferred change in plant community) is necessary to infer the
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49 297 processes involved.

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52 298 Small changes in the disturbance return interval can lead to cascading
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55 299 biogeochemical effects, as demonstrated by an example from the tallgrass prairie
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57 300 ecosystems of North America. This grassland disturbance regime is defined by a
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3 301 cycle of annual fires with the C losses replaced each year (i.e. 1:1). If these fires are
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5 302 suppressed by human intervention, the time between disturbances lengthens,
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7 303 leading to net C gain as woody plants increase (C balance > 1). Evidence from
8
9 304 experimental burning suggests that increasing fire return interval from 1 to 4 years
10
11 305 can initiate the first steps of vegetation change (Briggs et al. 2005). The initial C
12
13 306 gains are intensified by a series of feedbacks that culminates in woody plant
14
15 307 encroachment (Knapp et al. 2008). When C ceases to accumulate (apex of curve)
16
17 308 under the new disturbance regime, we define it as a biogeochemical state change.
18
19 309 There is evidence from many other grassland systems that this state change is
20
21 310 currently occurring, leading to increased net aboveground C stocks (Throop and
22
23 311 Archer 2008). In these cases, the biogeochemical state change is driven by a shift in
24
25 312 plant community composition, which is detectable in the pollen record.
26
27 313 The reverse process has also been seen in the paleorecord. Over time, non-
28
29 314 stationary changes in climate (warming and drying) increase the fire frequency
30
31 315 leading to a shorter disturbance interval relative to the C recovery time, eventually
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33 316 resulting in a net loss of C and potentially establishment of a different ecological
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35 317 community. When the C losses stabilize around a new average value, another
36
37 318 biogeochemical state change is indicated. Again this change is driven by a vegetation
38
39 319 shift: this time from woodland to shrubland, such as has been seen in paleorecords
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41 320 from the southwestern U.S. It is an open question at this time if it is possible to get a
42
43 321 large change in carbon balance without a major vegetation shift.
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45 322 Finally, disturbance and atmospheric carbon dioxide concentrations are tightly
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47 323 linked, with C stored in biomass often released to the atmosphere during a single
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3 324 disturbance event. The rate and mechanism of release varies with the type of
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5 325 disturbance, and again NECB may not change in the short-term without significant
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8 326 feedbacks to long-term climate change. Human management of fire regimes has
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10
11 327 been demonstrated to affect C budgets from local to global scales (Bowman et al.
12
13 328 2011).

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16 329
17 330 **Nitrogen**

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19 331 Nitrogen (N) dynamics have many similarities with C dynamics after a single
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21
22 332 disturbance event, and widespread N limitation of NPP and C accumulation often
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24 333 yields close couplings of N and C dynamics after disturbance (Vitousek and Reiners
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26
27 334 1975). However, there are also important differences in the cycling of these two
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29 335 elements that can lead to different long-term trajectories. N₂ is more energetically
30
31 336 costly to fix than CO₂, and yet once fixed into an ecosystem, a single atom of N can be
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34 337 recycled to fix many atoms of C. Despite persistent challenges in calculating
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36 338 ecosystem N budgets on decadal timescales, the general paradigm for temperate
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39 339 ecosystems is that after a single disturbance event, ecosystem N stocks recover to
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41 340 pre-disturbance levels within decades. Such dynamics have been seen in the post-
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43 341 fire chronosequences of *Pinus cortorta* forests in the Greater Yellowstone Ecosystem
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46 342 (figure 4, Smithwick et al. 2009). Rapid N accumulation is also demonstrated in the
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48
49 343 well-studied primary successional chronosequence in Glacier Bay, Alaska, where
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51 344 spruce and hemlock forests accumulated their maximum N contents within 150
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53 345 years of glacial retreat (Milner et al. 2007). Thus, on decadal to century timescales,
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56 346 disturbance alone often does not alter long-term nutrient stocks or limitation.
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3 347 However, the biogeochemical impact of a single disturbance event may also serve to
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5 348 initiate an increase or a decrease in long-term ecosystem N stocks (figure 5a). The
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8 349 disturbance initially causes nutrient loss, but recovery trajectories can vary
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10 350 depending on interactions between the disturbance, ecosystem properties, and
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12 351 ecosystem boundary conditions. One scenario is a net long-term loss of N due to
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14 352 disturbance, as in systems where N fixation is so limited that it cannot replace
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17 353 hydrologic and gaseous N losses. Changes in disturbance intensity or frequency as a
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19 354 result of altered boundary conditions, such as changing climate or vegetation shift,
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21 355 may also prevent reaccumulation of N after disturbance (Yelenik et al. 2013).
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23
24 356 Alternatively, disturbance can in some cases promote long-term increases in N
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26 357 stocks, such as where robust post-fire N fixers like *Alnus* colonize and add more N
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28 358 than is lost in a single disturbance event (Perakis et al. 2011). Similarly, altered
29
30 359 boundary conditions such as a wetter climate might permit higher N accumulation
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33 360 after disturbance.
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36 361 Thus, disturbance-driven increases or losses of nutrients can alter longer-term
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38 362 nutrient limitation when considered over long timescales (figure 5b). Each step of
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40 363 this biogeochemical staircase can therefore serve to either increase (“step up”) or
41
42 364 decrease (“step down”) nutrient availability. In turn, these disturbance-driven
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45 365 changes in nutrient status fall within a broader biogeochemical template derived
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47 366 from very long-term chronosequences that reveal how ecosystems change over
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49 367 long-term soil development. Whereas soil fertility often increases through primary
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51 368 succession, ultimately very old ecosystems undergo “retrogression” in the absence
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53
54 369 of catastrophic disturbance over millennial timescales (Peltzer et al. 2010).
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3 370 Characteristics of retrogressive ecosystems include reduced availability and
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6 371 amounts of soil nutrients (notably phosphorus), reduced NPP, and low standing
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8 372 biomass. In retrogressive ecosystems, disturbance events that expose unweathered
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10 373 soil parent material can replenish available nutrient pools and increase NPP (figure
11
12 374 5c). This seemingly counterintuitive response to disturbance works much the same
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14
15 375 way as geological uplift, glacial retreat, or volcanism on longer timescales (Porder et
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17
18 376 al. 2006).

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21 377 Our understanding of these millennial- and longer-scale processes has been derived
22
23 378 primarily from chronosequences, a series of sites with different histories used to
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25 379 reconstruct both primary (soil-development) and secondary (stand development)
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27
28 380 successional processes. This approach has been very powerful in ecosystem ecology,
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30 381 even if it often only provides information from discrete points in time (Laliberte et
31
32 382 al. 2012). Paleorecords, for example from lake sediment cores, can complement
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35 383 chronosequence studies by providing continuous records that span many millennia
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37 384 (Engstrom et al. 2000). Such records can be used to test theoretical predictions of
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40 385 the conditions when ecosystems should be responsive to disturbance, and the
41
42 386 direction of the response. Thus, ecosystems in very young stages of development
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45 387 where NPP is limited by N would be expected to respond differently to disturbance
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47 388 than very old retrogressive ecosystems limited by phosphorus (Peltzer et al. 2010).

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49
50 389 There are general patterns in how nutrient limitation shifts from primary
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52 390 succession to very old ecosystems that can aid interpretation of paleorecords
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55 391 through time in different landscapes. Nutrient limitation is strong in early stages of
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57 392 soil development, and is most commonly driven by N as it slowly accumulates from
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3 393 primary atmospheric sources. Nutrient limitation decreases in intermediate stages
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5 394 as N accumulates and rock-derived nutrients (e.g., phosphorus, calcium) are
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8 395 released into available forms by weathering. Nutrient limitation intensifies in late
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10 396 stages of soil development due to depletion of rock-derived nutrients on
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12 397 geologically old substrates (Chadwick et al. 1999), shift of rooting zone from mineral
13
14 398 to organic soil horizons (Bormann et al. 1995), or sequestration of accumulated N in
15
16 399 unavailable organic forms (Wardle et al. 2012). The sensitivity of ecosystem N
17
18 400 cycling and loss to disturbance also changes over geological time. At the young and
19
20 401 old ends of the spectrum, N cycling should show weak sensitivity to disturbance,
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22 402 and intermediate-aged ecosystems should exhibit strong sensitivity. In early soil
23
24 403 development, intense N limitation prevents strong changes in N cycling with
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26 404 disturbance. In intermediate stages of soil development, high rates of biotic N
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28 405 cycling elicit strong responses of N biogeochemistry to disturbance, with large
29
30 406 potential for nutrient losses. Late in soil development, nutrient limitation by rock-
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32 407 derived nutrients (phosphorus, calcium) intensifies, attenuating the response of N
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34 408 cycling as N limitation is relaxed, causing weak response of N cycling and loss to
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36 409 disturbance. Grounding novel geochemical proxies from sediment records in
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38 410 biogeochemical theory is likely to improve interpretation of paleorecords in an
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40 411 ecosystem context.
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51 413 **Hydrology**

52 414 Many studies have documented hydrologic responses to disturbance, and these
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54 415 provide an opportunity to consider ecosystem resilience to a range of disturbance
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56 416 types. A foundational paradigm is that removal of plant biomass, such as in clear-
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3 417 cutting experiments, would increase water available for stream discharge (Bormann
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6 418 and Likens 1979). In addition, a variety of ecosystem responses have been observed
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8 419 following disturbances, including changes in water balance (runoff, recharge,
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10 420 evaporation), chemistry (major ions, nutrients, organic C), and sediment yield.
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12 421 Additionally, responses may vary from regime shifts to no detectable change
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14 422 depending on the combination of disturbance magnitude and the resilience of the
15
16 423 ecosystem (figure 6). The total disturbance magnitude could represent the effects of
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18 424 a single event, but may also indicate the combined consequences of repeated
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20 425 disturbances. There are highly-variable observed examples of responses ranging
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22 426 from regime change to no response across the spectrum of total magnitude with the
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24 427 character of the response.
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30 428 Land cover change in different ecosystems, at different times and rates, reveals the
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32 429 influence of ecosystem resilience and resistance on the scale and style of responses
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34 430 to similar magnitude disturbances (figure 6). For example, at Hubbard Brook, New
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36 431 Hampshire, ecosystem resilience of a mesic forest has led to observed recovery of
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38 432 hydrologic behavior in a clear-cut watershed within decades (Bormann and Likens
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40 433 1979), but paleoecological studies demonstrate that land clearance can produce
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42 434 persistent change in less resilient ecosystems such as pine barrens (Lytle 2005) or
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44 435 in combination with other long-term influences (like climate change) that weaken
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46 436 ecosystem resistance and reduce the range of resilience. However, disturbance
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48 437 severity can also be important for shaping the type of hydrologic response. For
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50 438 example, forest thinning may produce a smaller change than total removal of trees.
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3 439 The importance of small disturbances for producing regime changes and severe
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6 440 responses in systems close to thresholds can be demonstrated by permafrost melt
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8 441 events (Smith et al. 2004) and changes in snowpack-drive stream discharge, where
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10 442 small warming can cross the melting threshold and cause a regime change in system
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12 443 state and behavior. Consequently, some ecosystems have the potential to jump
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14 444 permanently from one state to another (e.g., where warming is small, but persistent)
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16 445 even in the face of weak perturbations, while others have the capacity to return to
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18 446 their initial state after some period of recovery (e.g., where only a temporary change
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20 447 in watershed functioning occurred).
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25 448 The frequency of a type of event in a system may predict the magnitude and nature
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27 449 of the hydrologic response. Volcanism provides an example of an infrequent event,
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29 450 and an ecosystem-independent disturbance agent. With no ecosystem feedbacks on
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31 451 disturbance magnitude or frequency, paleo-ecosystem studies reveal otherwise
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33 452 unknown and infrequent events that cannot be studied on short time frames. For
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35 453 example, the eruption of Mount St. Helens, Washington in 1980 A.D. had immediate
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37 454 and well-documented effects on local hydrology, such as the loss of the mountain
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39 455 glaciers (figure 6). Similarly, an ancient eruption in the same region—the eruption
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41 456 of Mount Mazama in 7600 cal yBP—illustrates how volcanism can have far-reaching
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43 457 impacts. The Mazama ash more than doubled the water-holding capacity of soils
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45 458 which resulted in major vegetation changes especially in dry sites (Gavin et al.
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47 459 2001). Post-eruption regime shifts in lake level and diatom communities indicate
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49 460 hydrologic changes well to the east in Montana (Stone and Fritz 2006).
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3 461 Likewise, paleo-ecosystem studies reveal that many other historic observations of
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6 462 apparently large or severe disturbances are not unprecedented when viewed in a
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8 463 long-term context. Recent drought, fire and hurricane events have been placed in
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10 464 the context of many past events, which show that large infrequent disturbances
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12 465 (Foster et al. 1998) are poorly-understood without data representing a long time
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14 466 span (Laird et al. 2003). Disturbances that may initially be considered “catastrophic”
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16 467 such as the 1988 fires in Yellowstone National Park may be viewed as “natural”
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18 468 when compared with the paleorecord (Romme et al. 2011), while others such as the
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20 469 2007 Anaktuvuk River Fire in Alaskan tundra may be considered unprecedented
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22 470 over the late Holocene (Hu et al. 2010).
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28 471 Hydrological responses can also have implications for other biogeochemical cycles
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30 472 in a fashion that depends upon the character of the response. For example, severe
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32 473 droughts in the Great Plains have lowered lake levels and limited surface water
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34 474 supply both historically and during the past several millennia (Laird et al. 2003).
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37 475 One consequence of the most severe droughts during the mid-Holocene was to
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39 476 concentrate migrating birds at the few remaining lakes in the Great Plains, and
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41 477 create localized hot spots of nutrient delivery, but because these hydrologic systems
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43 478 are highly resilient, the end of the droughts led to a recovery of earlier broad
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45 479 distribution of bird stop-over locations and nutrient loading (Donovan and Grimm
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47 480 2007). We suspect that other interactions among systems (e.g., hydrology, fauna, N)
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49 481 and their levels of resilience will produce similarly contingent effects on ecosystem
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51 482 characteristics and function.
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3 483 On the other hand, hydrologic case studies also reveal important absence of
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5 484 responses. This resilience was also seen in long-term ecosystem response to
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7 485 disturbance for C and N cycling. In particular, multiple studies have sought to
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9 486 document significant changes in water chemistry in lakes following fires in their
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11 487 watersheds and have found little to no change. The absence of such a response has
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13 488 been observed both historically and in response to repeated fires in the same
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15 489 watershed using paleo-ecosystem techniques (Marchand et al. 2009). The absence
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17 490 of expected changes illustrates important gaps in our current understanding of
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19 491 ecosystem response to disturbance.
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25 26 492 **Research priorities** 27 493

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29 494 We identify a number of theoretical and practical approaches to strengthen the
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31 495 insights of ecosystem response to disturbance by linking information across
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33 496 multiple temporal scales. First, the paleorecord is replete with examples of
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35 497 unexpected complacency to relatively large magnitude disturbance events
36
37 498 (DiMichele et al. 2004). The disturbances can be reconstructed, but a focus on the
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39 499 drivers that allow ecological resilience or persistence is needed (Swetnam and
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41 500 Betancourt 1998). A general framework for identifying ecosystem characteristics
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43 501 that confer resilience has been difficult to achieve. Disturbance experiments may
44
45 502 provide a way forward. For example, a large-scale girdling experiment in a
46
47 503 secondary temperate forest identified the role of residual vegetation in retaining N
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49 504 and sustaining ecosystem C accumulation during a period when 20-50% of the
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51 505 existing trees were experimentally killed (Nave et al. 2011). Manipulations of fire
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3 506 severity can provide insights regarding the magnitude of C loss under different fire
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6 507 regimes and the mechanisms influencing post-fire C accumulation in both
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8 508 vegetation and soils. Manipulations of fire frequency can provide insights into
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10 509 resilience and resistance of different vegetation types to disturbance regimes
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12 510 (Briggs et al. 2005). The results of these types of experiments will add short-term
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14 511 mechanistic-oriented detail to long-term pattern-oriented records of past
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16 512 disturbance, providing a powerful joint perspective to understand resilience.
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20 513 Second, it is clear that disturbances of varying magnitude and type interact at
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22 514 the landscape scale. These interactions have led to the perceived difficulty of
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24 515 predicting multiple successional pathways after disturbance. Directly studying these
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26 516 interactions can increase our predictive power about post-disturbance
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28 517 biogeochemical trajectories. For example, the interactions of sequential
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30 518 disturbances from logging, fire, and wind disturbance in mixed conifer forests in
31
32 519 Colorado were non-additive (Buma and Wessman 2011). Surprisingly, disturbance
33
34 520 agent, combinations, and resilience mechanisms were more important than number
35
36 521 of disturbances. Because multiple disturbance events and agents can be identified in
37
38 522 paleorecords, this approach may provide further insights about the interactive
39
40 523 nature of disturbance events if explicitly studied. A better understanding of how
41
42 524 different disturbances are represented in the paleorecord is required. While fires
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44 525 are well-studied and well-represented in sedimentary charcoal and tree rings, other
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46 526 important disturbance agents (e.g. insect outbreaks, ice storms, windthrow) are
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48 527 often less clear, despite having implications for biogeochemical cycles. Identifying
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50 528 the mechanism of disturbance instead of the disturbance type or agent, such as in
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3 529 the conceptual framework described by Peters et al. (2011) is likely to improve
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6 530 predictability of biogeochemical response.
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8 531 Several practical avenues can speed the integration of neo and paleo
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10 532 timescales and concepts. A central one is to develop techniques to make the
11
12 533 sediment record more informative of terrestrial ecosystem dynamics. We identify
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14 534 three main technical developments in paleoecology as high research priorities : 1)
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16 535 developing improved quantitative proxies of NPP, 2) developing proxies of
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18 536 terrestrial nutrient cycling, and 3) assessing the age and structure of ecosystem C
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20 537 stocks from sedimentary records. There are significant challenges to each of these
21
22 538 research areas. But there are also recent advances such as spatially-explicit
23
24 539 reconstructions of terrestrial C budgets from pollen records (Boyle et al. 2011),
25
26 540 compound-specific N isotopic analysis in sediments (Enders et al. 2008), and rapid
27
28 541 high-resolution quantification of sedimentary elemental concentrations (Kylander
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30 542 et al. 2011). Another key is to routinely conduct multiproxy analyses, especially
31
32 543 identifying the source of sedimentary organic matter, and the application of whole-
33
34 544 lake flux measurements as derived from multiple sediment cores (Engstrom and
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36 545 Rose 2013). Finally, high-resolution temporal sampling across disturbance events,
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38 546 such that disturbance and biogeochemical responses are measured in same record,
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40 547 would be useful.
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49 548 The primary challenge is to bridge the spatial and temporal scales of neo-and
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51 549 paleo-studies, to produce a more consistent, synthetic understanding of
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53 550 biogeochemical processes than possible from either perspective alone. We identified
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55 551 centennial timescales as an area with the least information on post-disturbance
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3 552 biogeochemical processes, yet critical for understanding ecosystem function during
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6 553 ongoing climate change. Centennial timescales are also where some paleorecords
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8 554 lack sufficient resolution to capture dynamics associated with several disturbance
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10 555 intervals.

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13 556 The conceptual models developed in this manuscript are largely aspatial, but
14
15 557 could be made spatial. There are also particular places and spatial scales that may
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17 558 be most synergistic for linking across timescales. For example, the Arctic tundra is a
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19 559 region already experiencing the effects of global change, including altered fire
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21 560 regimes and permafrost melting (Hu et al. 2010). This region is especially promising
22
23 561 as a test area for this new synthetic approach to paleoecosystem ecology, because it
24
25 562 contains both detailed long-term ecosystem measurements and many sedimentary
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27 563 records of climate, vegetation, and geochemical change.
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33 565

34 566 **Conclusions**

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39 567 1. The unification of long-term ecological monitoring data, chronosequences, and
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41 568 paleorecords provides exciting opportunities for exploring the interactions of
42
43 569 multiple disturbance events against a changing background of climate and
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45 570 disturbance frequency.
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49 571 2. The biogeochemical consequences of a single disturbance are often short-lived,
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51 572 with biogeochemical parameters returning to pre-disturbance conditions on the
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53 573 same timescale as the dominant vegetation type. Thus, biogeochemical cycles in
54
55 574 many ecosystems can be considered resilient to disturbance. The true impact of
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3 575 disturbances may come through changing disturbance regimes, which requires
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6 576 the context of long timescales.
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8 577 3. Centennial and multi-centennial timescales are critical timeframes for improving
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10 578 our understanding of the biogeochemical consequences of disturbances. This
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12 579 scale is when the understanding of a single disturbance event from a
13
14 580 chronosequence begins to break down, while the temporal resolution of
15
16 581 paleorecords is usually insufficient to capture biogeochemical processes through
17
18 582 multiple disturbance events.
19
20 583 4. Developing proxy records of ecosystem characteristics is a priority for the
21
22 584 paleoecological community. Three proxies would be particularly useful: NPP,
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24 585 terrestrial nutrient dynamics, and age of ecosystem C.
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26 586 5. The existence of unexpected post-disturbance trajectories such as multiple
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28 587 successional pathways continues to challenge ecosystem ecologists and poses a
29
30 588 particular problem for the next generation of Earth System Models. The
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32 589 empirical and conceptual perspectives described here help move us closer to the
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34 590 goal of increased predictability of post-disturbance trajectories, including
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36 591 incorporating the role of disturbance agent, severity, and their interactions over
37
38 592 centennial to millennial timescales.
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52 596 (Paleo Reconstructions Of Biogeochemical Environments) workshop where these
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54
55 597 ideas were developed. Any use of trade names is for descriptive purposes only and
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Uncorrected version

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602
603 **Table 1.** Types of disturbances and how they are reconstructed. The key
604 contribution of paleoecology is being able to reconstruct enough disturbance events
605 to characterize disturbance regimes. These are often summarized over a given
606 spatial and temporal extent, and expressed as a distribution or summary statistics
607 [e.g, mean fire return interval (FRI)]. Additional disturbance properties include
608 spatial extent, timing, and interactions (Pickett and White 1985).
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610

disturbance agent	severity	frequency	technique
fire	low	high	fire scars recorded in wood
	high	low	tree establishment dates, sedimentary charcoal
insects	high	low	tree demography, insect fossils in sediment
wind	high	low	downed wood, clastic material in sediments
drought	variable	variable	tree rings, long-term instrumental records, lake level, lake salinity
tree removal (clearance, deforestation)	high	low	pollen preserved in sediment, sedimentation rates, tree demography
mass movement	variable	variable	sedimentation rate

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4 615 **Figure 1.** Paleoecological studies capture the long-term disturbance histories, state
5 616 changes in vegetation, and coarse measures of carbon stores. Multiproxy
6 617 sedimentary data from Yahoo Lake, Washington, USA show how climate affects
7 618 biomass and how fire mediates tree species composition during periods of rapid
8 619 climate change (Gavin et al. in press, 2013). (a) Organic content (loss-on-ignition,
9 620 LOI) and climate as estimated by sea surface temperature (SST) in the NE Pacific
10 621 Ocean. (b) Macroscopic charcoal stratigraphy quantified at 1-cm intervals. Red lines
11 622 indicate interpreted fire events near the lake. (c) Pollen percent (colored areas) and
12 623 needle macrofossil concentration (top axis, needle equivalents per 5 cubic
13 624 centimeters of sediment) for three coniferous tree species.

14 625

15 626 **Figure 2.** Post-fire aboveground carbon (C) accumulation in forests in Alaska. Post-
16 627 fire vegetation type determines the rate of C accumulation in living aboveground
17 628 biomass.

18 629

19 630 **Figure 3.** Conceptual diagram of long-term ecosystem carbon state in the context of
20 631 disturbance regimes. (a) how this would be recorded in the charcoal and pollen
21 632 paleorecords of a sediment core, (b) the carbon storage in the ecosystem, and (c) the
22 633 underlying processes of a shift in NECB, including state changes in vegetation and
23 634 disturbance regime.

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25 636 **Figure 4.** A well-replicated post-disturbance chronosequence of lodgepole pine
26 637 (*Pinus contorta* var. *latifolia*) forests in the western U.S. shows equilibration of
27 638 ecosystem nitrogen (N) pools but redistribution of ecosystem N among pools after
28 639 disturbance (Smithwick et al. 2009).

29 640

30 641 **Figure 5.** Conceptual diagram showing response of available stocks of growth-
31 642 limiting nutrients to disturbance. (a) a single disturbance event may serve to
32 643 increase, decrease, or not affect net stocks of a nutrient depending on post-
33 644 disturbance accumulation and loss rates. (b) over time, a disturbance regime may
34 645 gradually increase or decrease nutrient availability, a phenomenon we term the
35 646 “biogeochemical staircase”, (c) over longer timescales, ecosystems often experience
36 647 the onset of phosphorus limitation and reduced net primary productivity known as
37 648 retrogression (reproduced from Peltzer 2010). In the oldest stages of ecosystems,
38 649 disturbances can serve to replenish nutrients.

39 650

40 651 **Figure 6.** Conceptual diagram of four types of hydrologic responses to disturbances,
41 652 encapsulating the range of responses demonstrated in various paleorecords,
42 653 observations, and modern experiments. Red colors indicate a response of water
43 654 quality, blue colors indicate a response of water quantity.

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4 661 **Box 1.** Background on disturbance ecology.
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The theoretical basis for linking biogeochemistry with **disturbance ecology** was developed in the 1970s and largely focused on impacts from, and recovery to, steady-state or equilibrium conditions (Vitousek and Reiners 1975). Newer conceptualizations of complex adaptive systems and resilience in ecological systems now highlight the important role of interactions, system feedbacks, and landscape contingencies when forecasting how disturbances affect ecosystems. Interactions between biogeochemical cycles and vegetation can either reinforce current states or result in shifts to alternate states, depending on the magnitude and timing of the interactions. Importantly, potential shifts in system states can be caused by the disturbance itself, as well as centuries to millennia of vegetation-soil feedbacks that constrain, mediate, or amplify ecological responses (Turner et al. 1993).

The concept of **biogeochemical resilience** incorporates the idea that biogeochemical characteristics are spatially differentiated prior to disturbance, are sensitive to the disturbance event itself, and are coupled to vegetation response patterns and rates (Smithwick 2011). Classically, ecological resilience is focused on organisms, while biogeochemical resilience is focused on pool sizes and input/output rates. Over long spatial and temporal scales, shifts in biogeochemistry influence vegetation response trajectories and *vice versa*. Understanding long-term ecosystem resilience to disturbance requires a full characterization of disturbance regimes—the agent or mechanism of disturbance, as well as the magnitude, severity, and return interval (Table 1, Peters et al. 2011)— alongside an understanding of elemental pools and flows.

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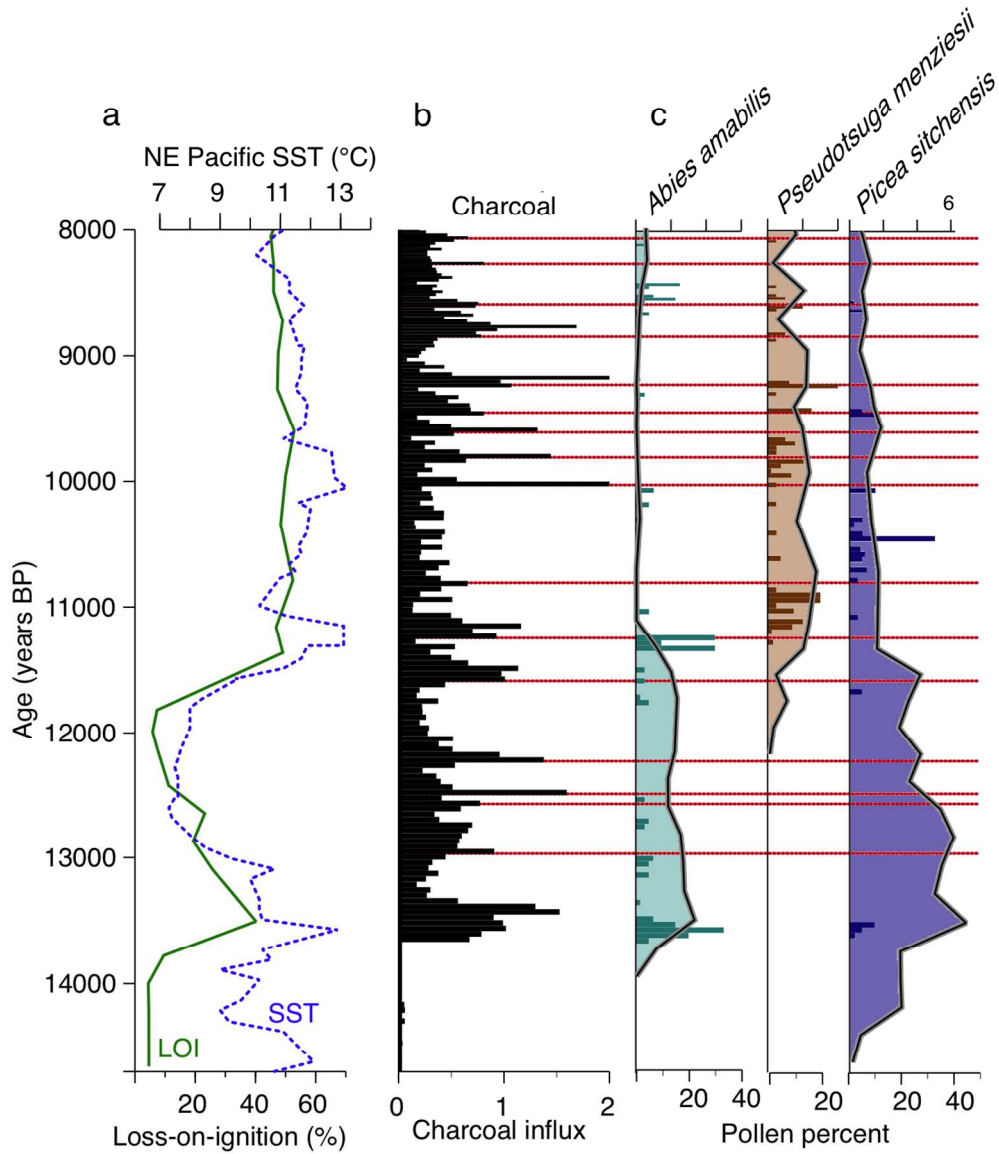


Figure 1. Paleoeological studies capture the long-term disturbance histories, state changes in vegetation, and coarse measures of carbon stores. Multiproxy sedimentary data from Yahoo Lake, Washington, USA show how climate affects biomass and how fire mediates tree species composition during periods of rapid climate change (Gavin et al. in press, 2013). (a) Organic content (loss-on-ignition, LOI) and climate as estimated by sea surface temperature (SST) in the NE Pacific Ocean. (b) Macroscopic charcoal stratigraphy quantified at 1-cm intervals. Red lines indicate interpreted fire events near the lake. (c) Pollen percent (colored areas) and needle macrofossil concentration (top axis, needle equivalents per 5 cubic centimeters of sediment) for three coniferous tree species.

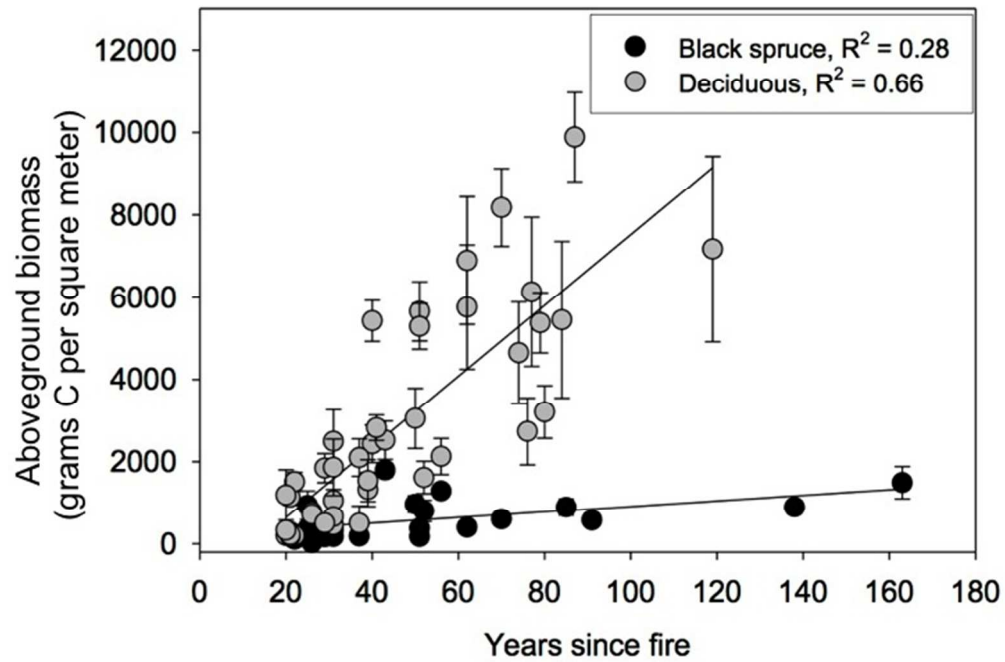
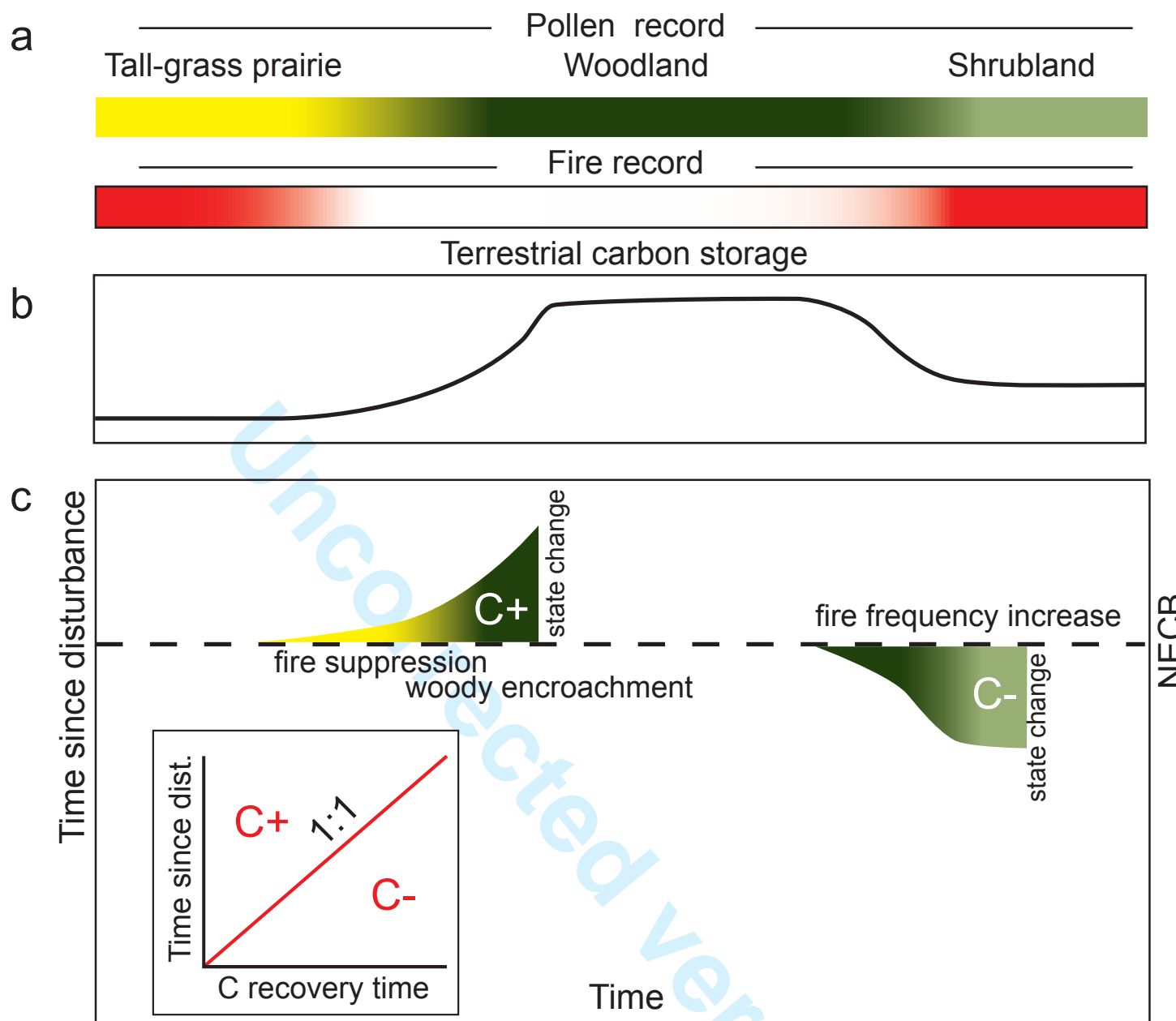


Figure 2. Post-fire aboveground carbon (C) accumulation in forests in Alaska. Post-fire vegetation type determines the rate of C accumulation in living aboveground biomass.
254x190mm (72 x 72 DPI)



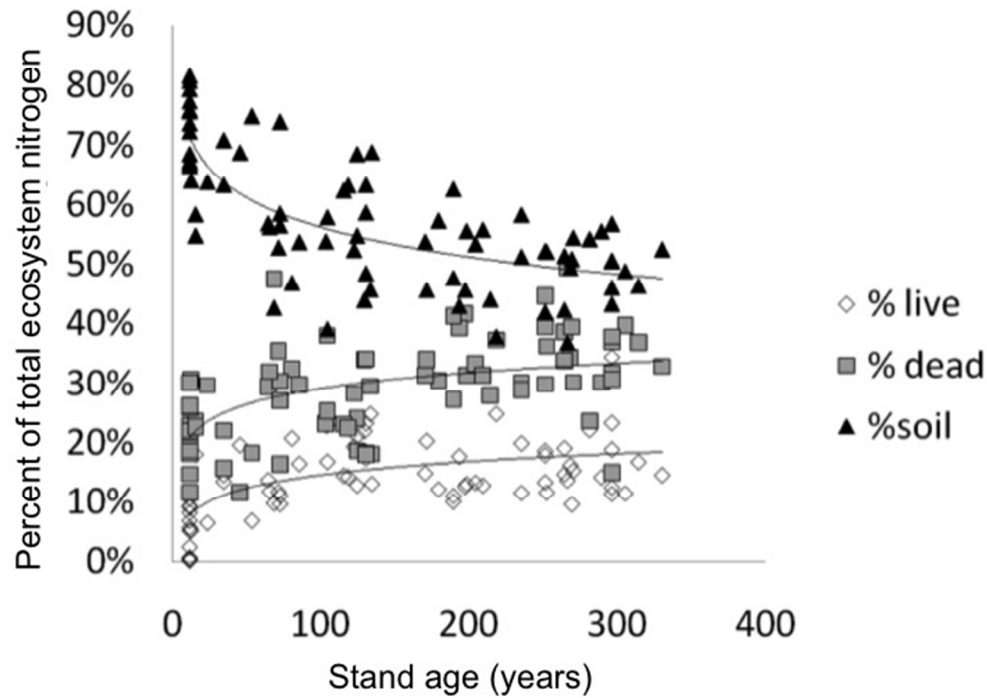


Figure 4. A well-replicated post-disturbance chronosequence of lodgepole pine (*Pinus contorta* var. *latifolia*) forests in the western U.S. shows equilibration of ecosystem nitrogen (N) pools but redistribution of ecosystem N among pools after disturbance (Smithwick et al. 2009).
254x190mm (72 x 72 DPI)

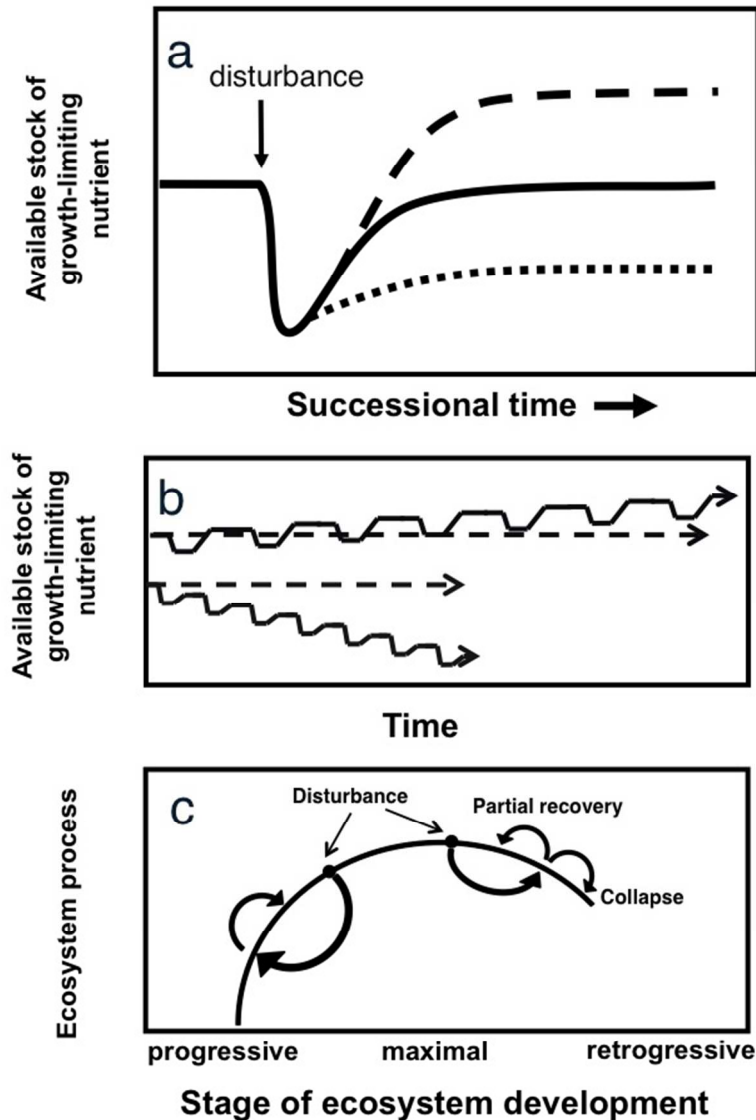


Figure 5. Conceptual diagram showing response of available stocks of growth-limiting nutrients to disturbance. (a) a single disturbance event may serve to increase, decrease, or not affect net stocks of a nutrient depending on post-disturbance accumulation and loss rates. (b) over time, a disturbance regime may gradually increase or decrease nutrient availability, a phenomenon we term the “biogeochemical staircase”, (c) over longer timescales, ecosystems often experience the onset of phosphorus limitation and reduced net primary productivity known as retrogression (reproduced from Peltzer 2010). In the oldest stages of ecosystems, disturbances can serve to replenish nutrients.
254x338mm (72 x 72 DPI)

Style of ecosystem responses

Water quality changes
Water quantity changes

Time series

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