

1 **Permeability is the Critical Factor Governing the Life Cycle Environmental Performance**
2 **of Drinking Water Treatment Using Living Filtration Membranes**

3

4 Daqian Jiang ^{1*}, Dianxun Hou², Carson Bechtel ¹, Katherine R. Zodrow ¹, Rupert J. Myers³,
5 Tianyu Zhang ⁴

6

7 Affiliations:

- 8 1. Environmental Engineering Department, Montana Technological University, Butte MT,
9 59701, USA
- 10 2. WaterNova Group, Lakewood, CO 80227, USA
- 11 3. Department of Civil and Environmental Engineering, Imperial College London, London,
12 SW7 2AZ, UK
- 13 4. Department of Mathematical Sciences, Montana State University, Bozeman, Montana 59717,
14 USA

15

16

17 Correspondence to:

18 djiang@mtech.edu; 1300 W Park St, Butte MT, 59701

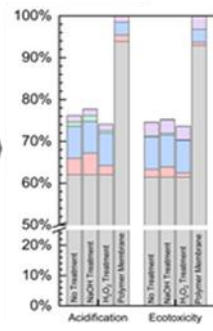
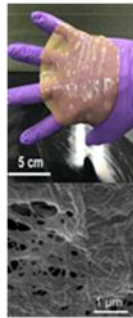
19

20

21 TOC

Design
Treatments

Operation
of fibers
Regeneration
Backwash
Lifespan
TMP



22

23

24

25 **Abstract**

26 Living Filtration Membranes (LFMs) are a water filtration technology that was recently
27 developed in the lab (Technology Readiness Level 4). LFMs have shown filtration performance
28 comparable with that of ultrafiltration, far better fouling resistance than conventional polymer
29 membranes, and good healing capabilities. These properties give LFMs promise to address two
30 significant issues in conventional membrane filtration – fouling and membrane damage. To
31 integrate environmental considerations into future technology development (i.e., Ecodesign), this
32 study assesses the life cycle environmental performance of LFMs treating drinking water under
33 likely design and operation conditions. It also quantitatively ranks the engineering design and
34 operation factors governing the further optimization of LFMs’ environmental performance using
35 a global sensitivity analysis. The results suggest that LFMs’ superior fouling resistance will
36 reduce the life cycle environmental impacts of ultrafiltration by 25% compared to a conventional
37 polymer membrane in most impact categories (e.g., Acidification, Global Warming Potential and
38 Carcinogenics). The only exception is the eutrophication impacts, where the need for growth
39 medium and membrane regeneration offsets the benefits of LFMs’ fouling resistance.
40 Permeability is the most important factor that should be prioritized in future R&D to further
41 improve LFMs’ life cycle environmental performance. A 1% improvement in the permeability
42 will lead to a ~0.7% improvement in LFMs’ environmental performance in all the impact
43 categories, whereas the same change in the other parameters investigated (e.g. LFM lifespan and
44 regeneration frequency) typically only leads to a <0.2% improvement.

45

46

47 **Introduction**

48 Recently, Living Filtration Membranes (LFMs) were successfully developed as a water filtration
49 technology for the first time using native microorganisms of a kombucha symbiotic culture of
50 bacteria and yeast.¹ Composed of a bacterial cellulose (BC) network with embedded
51 microorganisms, LFMs have physical properties highly suitable for water filtration (e.g., high
52 tensile strength and hydrophilicity^{2,3}). In bench-scale dead-end filtration tests, LFMs have
53 accomplished permeability and size cutoff similar to commercial ultrafiltration membranes (i.e.
54 $135 \pm 25 \text{ L m}^{-2}\text{h}^{-1}\text{bar}^{-1}$ and a 90% rejection of 30 nm).⁴

55 Most notably, LFMs have demonstrated high fouling resistance. When treating the influent used
56 by the Basin Creek Drinking Water Treatment Plant (Butte, MT, USA), the flux loss with LFMs
57 was only 50% after 7 hours of operation, considerably lower than the >90% flux loss
58 experienced by a commercial polymer membrane (Millipore VMWP02500, Massachusetts,
59 USA) after the same time (unpublished results attached for review only). This suggests that
60 LFMs can potentially address the greatest challenge in membrane filtration – membrane fouling.
61 Fouling greatly reduces the efficiency of membrane filtration, and can occur even under harsh
62 conditions.^{5,6} Despite the high attention given to membrane fouling, there have been limited
63 breakthroughs in its mitigation.⁷⁻¹⁰

64 In addition to fouling resistance, LFMs have also demonstrated great healing capabilities. Even
65 after severe damages (e.g., 3 mm holes), LFMs achieved 75-80% recovery of flux in a period of
66 4-17 days simply by placing it in growth solutions at 25 °C.^{1,4,11} The healing capabilities of
67 LFMs have the potential of addressing another common operational challenge in membrane
68 filtration – fiber damage.^{12,13} While the frequency of fiber damage may be moderate,¹⁴ it is an

69 operational nuisance, and a number of common operating conditions (e.g., chemical cleaning)
70 can increase the frequency of its occurrence.¹⁵

71 LFMs have created a new category of water filtration technologies, as they are a biodegradable
72 material that may be fabricated into virtually any shape from μm to mm dimensions,¹⁶ and be
73 structurally and chemically modified to impregnate anti-fouling^{17,18} and/or diffusion
74 properties.^{4,19} A multitude of subsequent studies can be developed upon this proof of concept,
75 e.g., permeability modifications,¹¹ functional modifications through incorporation of other
76 molecules,²⁰⁻²² and optimization of the engineering design and operation conditions.²³

77 To integrate environmental considerations into future technology development (i.e.,
78 Ecodesign),^{24,25} herein we conduct a preliminary life cycle assessment of LFMs as a drinking
79 water filtration technology. Since LFMs are currently in the lab development stage, we
80 extrapolate its potential future full-scale environmental performance based on a combination of
81 lab data (e.g. membrane characteristics, such as permeability and regeneration frequency) and
82 real-world operational data (e.g. backwash frequency, and chemical and energy consumption)
83 from the Basin Creek Drinking Water Treatment Plant (Butte, MT, USA). We then identify high-
84 priority design parameters using a global sensitivity analysis over a comprehensive design space
85 informed by the conditions accomplished in the lab and the full-scale plant in Butte, MT.

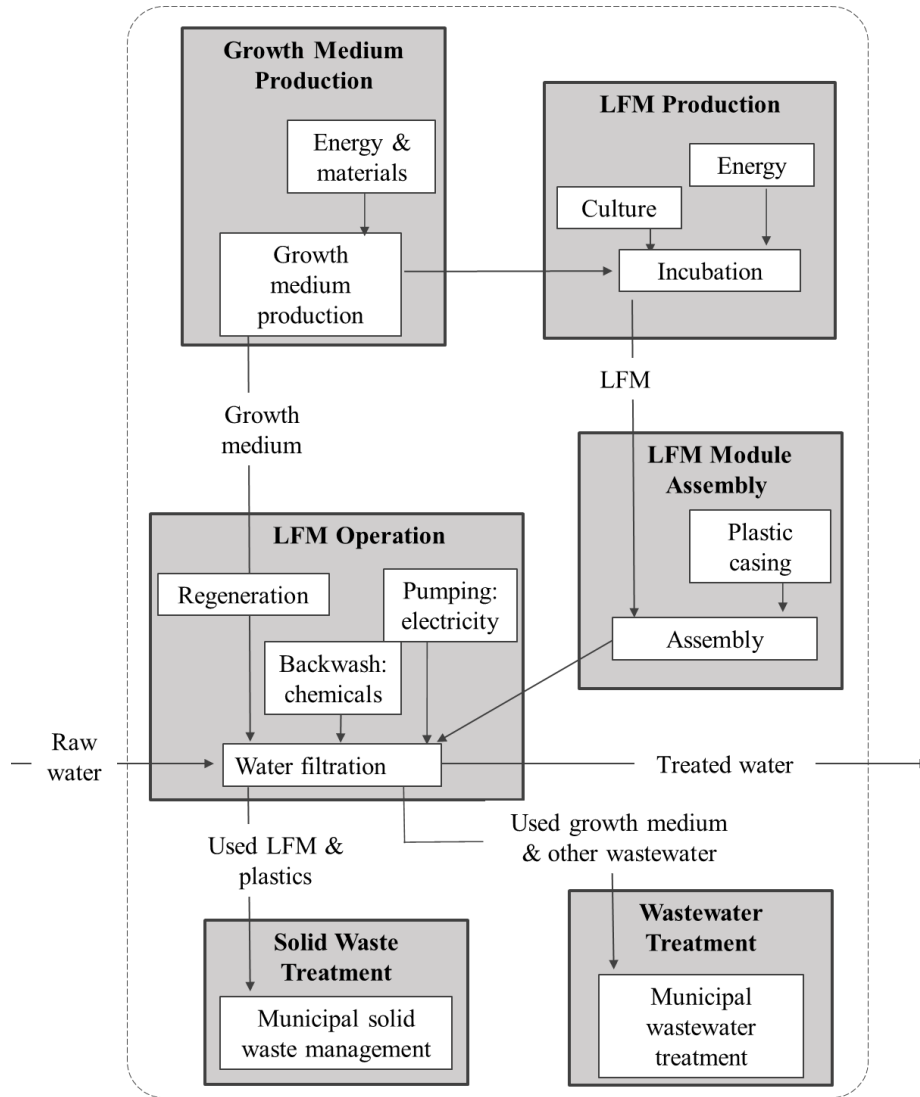
86

87 **Methods and Data**

88 Overview of LFMs

89 A detailed, technical description of LFMs is available elsewhere and in the SI.⁴ Briefly, LFMs
90 currently are a lab-scale water filtration technology (Technology Readiness Level 4). It has

91 successfully treated deionized water at the 100 L/day scale, and achieved filtration performance
92 comparable to that of ultrafiltration with a size cutoff of 30 nm.¹¹



93
94 Figure 1. Overview of LFM-based water filtration. Dotted box is the system boundary. The
95 treated water produced by LFM does not require modifications to subsequent drinking water
96 treatment processes (e.g., disinfection) based on current knowledge.

97
98 Goal of the Life Cycle Assessment

99 The goal of this study is to estimate the life cycle environmental impacts of LFMs in full-scale
100 drinking water operation, and to the extent possible, compare with conventional full-scale
101 membrane filtration technologies to assess whether LFMs can reduce the environmental impacts
102 associated with drinking water treatment in its current state. A functional unit (FU) of 1 MGD
103 (4645m³/day) treated water is thus chosen. Two things should be clarified. First, again, since
104 LFMs are still in the early stages of development, this LCA is indicative of the current
105 technology readiness level (TRL 4) rather than a more mature level that would be expected in
106 commercial production. Second, this LCA assessment has not been certified by an independent
107 LCA analyst and is not intended to be used for any commercial or marketing purposes.

108

109 Scope of the Life Cycle Assessment

110 The scope of this LCA study is from raw materials extraction through to end-of-life waste
111 disposal (Figure 1). Included in the system boundary are the growth medium production, LFM
112 production, LFM module assembly, LFM module operation, and end-of-life disposal (Figure 1).
113 Key upstream and downstream processes, such as electricity production and end-of-life disposal
114 of LFMs are included. Construction of the water treatment facilities (e.g., concrete and pipes)
115 and transportation are excluded. The omission of construction and transportation likely has
116 insignificant impacts, since all previous studies reported that the life cycle impacts of
117 conventional membrane filtration plants are dominated by the operation stage.²⁶⁻²⁸ The LFM
118 modules (plant) are assumed to operate 20 hours/day for 300 days/year, with a lifespan of 20
119 years.

120 This LCA study compares LFMs with a mixed cellulose esters polymer membrane (Millipore
121 VMWP02500, Massachusetts, USA, 0.05 μm pore size), which is picked based on similarities
122 with LFMs in terms of filtration performance (i.e. cutoff and permeability). Identical FU, plant
123 life span, and operation time are used. The same life cycle stages are included. Growth medium
124 production, LFM regeneration and LFM replacement are excluded, as they are not needed for
125 polymer membranes.

126

127 Life Cycle Inventory

128 The inventory data used are a combination of lab data (Table S1), real-world operation data from
129 a full-scale membrane filtration plant (Table S2), and Ecoinvent v3.4 (Table S3). The lab data
130 (foreground data) primarily include the relationships between flux and trans-membrane pressure
131 (TMP) under different treatments, the development of fouling and its impact on flux over time,
132 the material and energy consumption in the growth medium production, and the frequency and
133 material and energy consumption in LFM regeneration. The full-scale operation data (foreground
134 data) are based on the Basin Creek Water Treatment Plant in Butte, MT, which include the
135 frequency of backwash and the material and energy consumption in backwash. Ecoinvent is used
136 as the background database, which include data such as the impacts of wastewater treatment and
137 electricity supply (Table S3).

138

139 Life Cycle Impact Assessment

140 The U.S. EPA's Tool for the Reduction and Assessment of Chemicals and Other Environmental
141 Impacts (TRACI 2.1 version 1.02) was used to assess the impact of LFM filtration in OpenLCA

142 (v1.8). All of its midpoint impact category indicators are reported. The results are grouped into
143 module production, operation – backwash, operation – electricity, operation – regeneration, end-
144 of-life (EoL) – wastewater, and EoL – others. Module production includes the material and
145 energy consumption associated with producing LFM modules and fibers (e.g., plastics, growth
146 medium etc.). “Operation – backwash” includes the energy and chemicals needed for all three
147 types of backwash. “Operation – electricity” is the electricity consumption during operation (i.e.
148 primarily pumping). “Operation – regeneration” includes the materials and energy needed for
149 LFM regeneration (e.g. the growth medium needed). “EoL – wastewater” is the treatment of the
150 wastewater generated during operation, which includes the growth medium used in regeneration
151 and 1% of FU (i.e. 99% recovery assumed). “EoL – others” include the disposal of EoL plastics
152 and LFMs.

153

154 Design Parameters

155 Two types of parameters are selected to understand the potential of improving the life cycle
156 environmental impacts of LFMs through engineering design and operation: LFM property (i.e.
157 permeability and fouling resistance) and design/operation parameters (Table 1). These
158 parameters are directly related to the filtration performance and the material/energy consumption
159 of LFMs, according to lab results and/or theoretical understandings of the mechanisms.

160

161 Table 1. The design parameters and the ranges of values studied in the sensitivity analysis. The
 162 flux-TMP relationship is based on (Song 1998).²⁹ The fouling resistance is based on lab results.

		Flux-TMP/time relationships			
LFM Property	<u>Permeability</u>	Flux ₀ = A×TMP			
		No treatment	A=119.9 L/(m ² h bar)		
		0.3% H ₂ O ₂	A=217.7 L/(m ² h bar)		
		0.5M NaOH	A= 89.4 L/(m ² h bar)		
	<u>Fouling Resistance</u>	Flux _t =Flux ₀ ×[exp (a×t)]			
	LFMs	a= -0.265 hr ⁻¹			
	Polymer membrane	a= -1.233 hr ⁻¹			
		Baseline values	Range simulated in global sensitivity analysis	Accomplished in lab or real-world operation	
Design/Operation Parameters	<u>Module Assembly</u>				
		Number of fibers per module	800	(400, 1800)	NA
	<u>Maintenance</u>				
		LFM regeneration frequency (times/year)	40	(20, 90)	(20, 60)
		Backwash frequency (flux loss)	15%	(10%, 45%)	10-15%
	<u>Durability</u>				
	LFM fiber lifespan (year)	0.8	(0.4,1.8)	NA	
<u>Operation</u>					
	TMP (bar)	4	(0.7, 5.6)	(0.7, 3.1)	

163
 164 A: membrane-specific constant, TMP: trans-membrane pressure, Flux_t: flux at time t, Flux₀: flux at time
 165 0, t: time, a: constant; NA: not available.

166 **Permeability** is an LFM property that determines the flux of LFMs under a certain TMP. It was
 167 shown that LFMs has the capability of undergo chemical treatments to achieve different
 168 permeability and selectivity.¹¹ This capability has direct engineering and environmental
 169 implications. Higher permeability will reduce the energy consumption of LFM filtration (and
 170 thereby likely reducing the environmental impacts of LFM filtration), while higher selectivity
 171 will allow LFMs to adapt changing source water qualities. Herein we explore the environmental

172 impacts of three LFM treatments (i.e. two chemical treatments and a no treatment baseline). The
173 flux-TMP relationship for each treatment is extrapolated from lab results and shown in Table 1.¹¹

174 **Fouling resistance** is an LFM property that characterizes how the flux decreases as fouling
175 develops. In preliminary lab results, LFMs demonstrated a considerable advantage over
176 conventional membranes. With the real influent used by the Basin Creek Drinking Water
177 Treatment Plant in Butte, MT, the flux loss was 50% with LFMs after 7 hours of operation, as
178 opposed to a >90% flux loss with a commercial polymer membrane after the same operation
179 time (unpublished results attached for review only).¹¹ The fouling resistance of LFMs is modeled
180 using a time-dependent equation based on lab results ($a = -0.265$ and -1.233 hour^{-1} for LFMs and
181 polymer membrane respectively, Table 1). The LFMs are assumed to operate in the constant-flux
182 mode, which has advantages over the constant-TMP mode and is popular in industrial
183 applications.^{30,31}

184 **Module assembly** explores the sensitivity of environmental performance to the number of LFM
185 fibers assembled into each module. A typical hollow-fiber design often used in commercial
186 ultrafiltration is followed (length 2m, diameter 0.2m). Each LFM fiber is assumed to be 2m in
187 length and 4mm in diameter, with a wall thickness of 1.5mm (LFM fibers are currently produced
188 as 1.5mm-thick flat sheets in the lab). With this design, the theoretical maximum number of
189 LFM fibers per module is ~2500. The number of fibers per module simulated in this analysis is
190 (400, 1800) (Table 1). A conservative range is chosen given the lack of demonstrated success.
191 The impact of increasing this parameter can be inferred from the global sensitivity analysis.

192 **Maintenance** first includes the periodic regeneration of LFMs to maintain their structural
193 robustness, which can be done by filling each module with growth medium at designated
194 frequencies during the down time. Given the size of the modules, 63L growth medium is needed

195 for each regeneration. The range simulated in this study is (20, 90) times per year (Table 1). The
196 most feasible regeneration frequency needs to be further investigated in future research and
197 development.

198 Also included in maintenance is backwash, which includes three types: daily backwash with
199 reverse water flows, weekly chemical cleaning with NaClO, and semi-annual cleaning with
200 NaClO and citric acid. Daily backwash is initiated once the flux decreases to a certain fraction of
201 the initial flux (e.g., backwash every time the flux decreases to 85% of the initial flux). The
202 values used are specified in Table 1.

203 Weekly chemical cleaning and semi-annual cleaning are modeled using real-world operation
204 data from the Basin Creek Water Treatment Plant in Butte, MT (which uses ceramic
205 membranes), due to the lack of LFM-specific lab data. It is possible that LFMs do not require or
206 could not withstand the same chemical treatments as polymer or ceramic membranes. The real-
207 world operation data are followed in this study as it is the best available data. Further LMF-
208 specific research is needed in the future.

209 **Durability** characterizes the frequency of LFM fiber replacement, which is dependent on the
210 lifespan of LFMs, the conditions, and the water treated. Lab data on the range of LFM lifespan is
211 not yet available. An arbitrary range of 0.4-1.8 years is explored.

212 **Operation** involves the application of a TMP to generate treated water. The range of TMP
213 accomplished in the lab is (0.7, 3.1) bar for LFMs. The range simulated in this study is (0.7, 5.6)
214 bar, which covers the typical pressure applied in the full-scale plant in Butte, MT (4-5 bars). The
215 flux under each TMP and treatment is modeled using the flux-TMP relationships specified in
216 Table 1.

217 The parameters used in the “baseline” case are 800 fibers per module, regeneration frequency
218 40/year, replacement frequency 0.8/year, TMP 4 bar, and backwash 4 times per day (5
219 hours/run), which are based on what is typically accomplished in the lab (e.g., regeneration
220 frequency) and full-scale operation data (e.g., TMP and backwash frequency). The commercial
221 polymer membrane used as a reference is based on the design and operation conditions typically
222 accomplished in commercially available filtration modules and full-scale plants, namely 2000
223 fibers per module, no regeneration needed, replacement frequency 0.1/year, TMP 4 bar, and
224 reverse water flow backwash when there is a 15% loss of flux.

225

226 Sensitivity Analysis

227 A global sensitivity analysis is conducted by combining Latin–Hypercube (LH) sampling³² and
228 One-factor-At-a-Time (OAT) approach³³ as outlined in a previous study.³⁴ The LH-OAT
229 analysis is done for each chemical treatment (Table 1). Within each chemical treatment,
230 $M(e_1, \dots, e_P)$ is one of the mid-point impact categories in TRACI (e.g., acidification) that
231 depends on P parameters (here $P = 5$, i.e., the five design/operation parameters detailed in Table
232 1). Each parameter e_i is assumed to be uniformly distributed on an interval $[a_i, b_i]$ (shown in
233 Table 1), and divided into N (here $N = 7$) strata with a probability of occurrence equal to $1/N$ in
234 the LH sampling method. Thus, the total parameter space is divided into N^P (here $7^5 = 16807$)
235 LH cubes. In each LH cube, one random sample of the parameters (e_1, \dots, e_P) is generated, and
236 the partial effect of parameter e_i on the impact factor M is calculated by OAT approach
237 (Equation 1):

$$S_i = \frac{1}{f_i} \cdot \frac{M(e_1, \dots, e_i * (1 + f_i), \dots, e_P) - M(e_1, \dots, e_i, \dots, e_P)}{[M(e_1, \dots, e_i * (1 + f_i), \dots, e_P) + M(e_1, \dots, e_i, \dots, e_P)]/2}$$

(Equation 1)

238

239 Where S_i is the sensitivity and f_i is the fraction by which the parameter e_i is changed (a
 240 predefined constant with value 10^{-5} for all i). $P + 1$ evaluations of M are then conducted at each
 241 LH point and the total simulation requires $(P+1)*N^P$ evaluations of M .

242

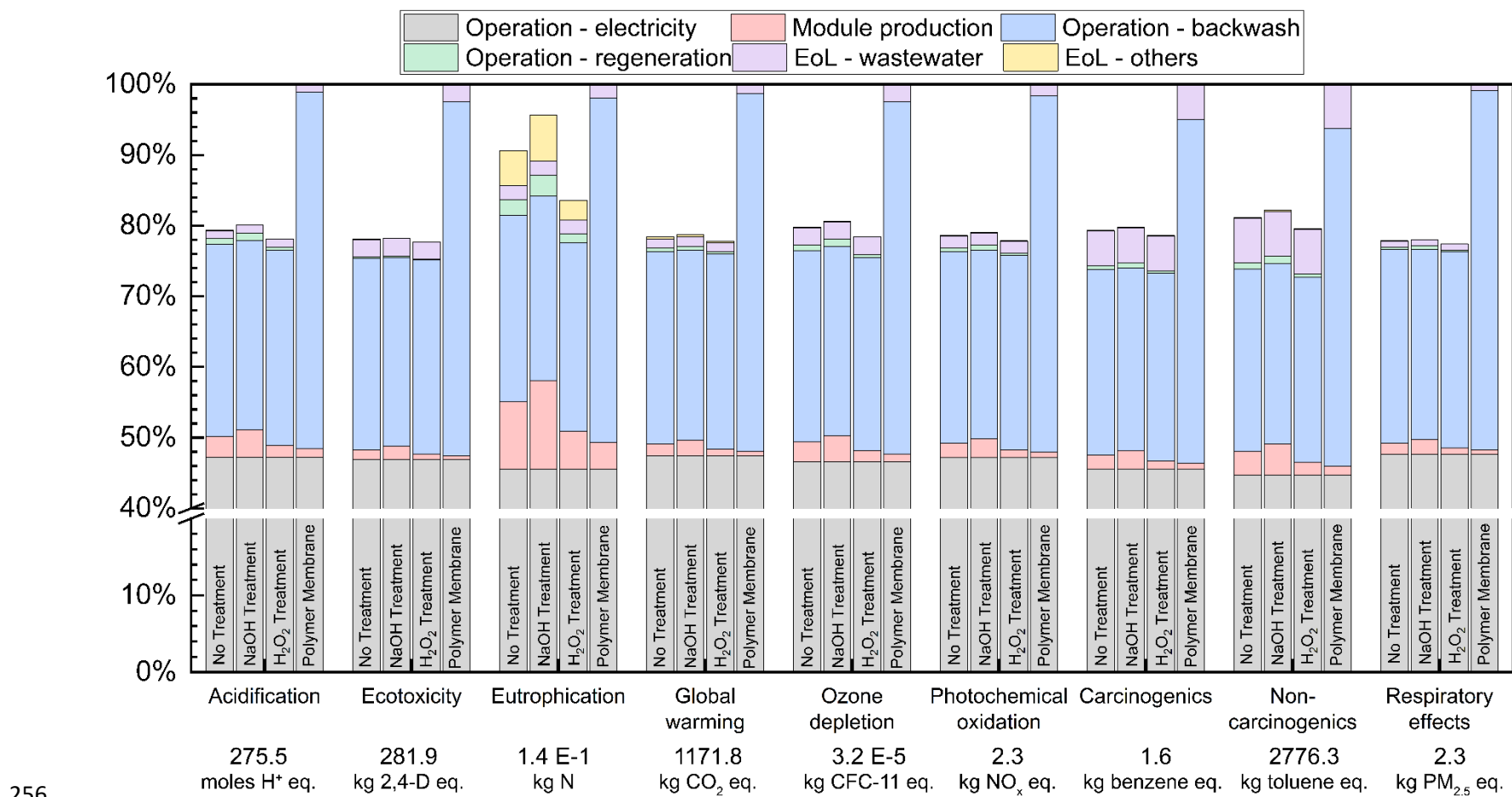
243 **Results and Discussion**

244 Comparison of LFMs and the commercial polymer membrane used in this study

245 In the “baseline” projection, LFMs outperform the polymer membrane by 20-25% in all but one
 246 of the impact categories (Figure 2). The advantage is primarily attributable to LFM’s superior
 247 fouling resistance and the resultant decrease in the electricity consumption during operation.
 248 However, for eutrophication impacts, LFMs only show a 3-18% improvement over the polymer
 249 membrane (Figure 2), mainly because the electricity savings are offset by the increased impacts
 250 from the production of growth medium and LFM fibers during the initial module production and
 251 the LFM regeneration in operation. Different LFM treatments show limited impact on the life
 252 cycle impacts of LFMs in all of the impact categories (i.e., < 3%), except for eutrophication,
 253 where different LFM treatments can lead to a 20% difference (Figure 2).

254

255



256

257

258

259

Figure 2. Relative contributions to each midpoint impact category by process and chemical treatments in LFMs and polymer membrane. The un-normalized 100% impacts are impacts of polymer membrane per FU and the numbers are shown below the x-axis.

260 In terms of relative contributions by process, LFMs and polymer membrane are similar in that
261 the electricity consumption during operation and backwash is the largest contributor to most
262 impact categories (Figure 2). For LFMs, the operation and backwash typically account for >85%
263 of the total impact in all categories; for polymer membranes, the percentage contributions of the
264 two processes are over 95% (Figure 2). The exception again is the eutrophication impacts, in
265 which the other processes (such as regeneration and EoL) can account for up to 25% of the total
266 impact (Figure 2).

267

268 Comparison of LFMs with conventional membrane filtration in previous studies

269 The life cycle impacts of LFM filtration are also compared with previous studies on polymer
270 membranes (Table 2). Despite the differences in the technology (e.g., ultrafiltration vs.
271 nanofiltration), the unit processes included (e.g., with or without construction and pre-treatment),
272 and the LCIA method, a general observation is that the life cycle GWP impacts of LFMs and the
273 reference polymer membrane used in this study do not significantly exceed that of previous
274 studies on the basis of per m³ water treated (Table 2). A comprehensive validation should be
275 attempted when more information becomes available.

276

277

278 Table 2. Comparison with the life cycle global warming potential impacts of LFM (baseline
 279 projection) and membrane filtration technologies reported in previous studies

Reference (technology readiness)	GWP per m ³ treated (kg CO ₂ eq/m ³)	LCIA model	Water treated	Note
This study (lab scale)	0.20	TRACI 2.1	DW	LFMs, excluding construction and pretreatment (coagulation, prefiltration etc.)
This study (lab scale)	0.25	TRACI 2.1	DW	VWMP6 polymer membrane, excluding construction and pretreatment (coagulation, prefiltration etc.)
Ribera et al. 2014 ²⁶ (full scale)	0.13-0.15	Recipe Midpoint (H)	DW	Nanofiltration
Bonton et al. 2012 ²⁸ (full scale)	0.04	Impact 2002+	DW	Nanofiltration
Tangsubkul et al. 2006 ³⁵ (full scale)	0.03-0.30	Unspecified	WW	Microfiltration, including different operating conditions such as TMP and flux
Carre et al. 2017 ²⁷ (full scale)	0.25 0.41 0.42	CML	WW	Ultrafiltration Ultrafiltration + UVB Microfiltration + UVD
Ortiz et al. 2007 ³⁶ (full scale)	0.02-0.5	CML Eco-indicator 99 Ecopoints 97	WW	Ultrafiltration, estimated based on the difference between two processes: activated sludge, and activated sludge + ultrafiltration
Godskesen et al. 2013 ³⁷ (full scale)	1.18	EDIP 1997	DW	Ultrafiltration, including pretreatment, desalination, and UV

280 DW: drinking water. WW: wastewater.

281

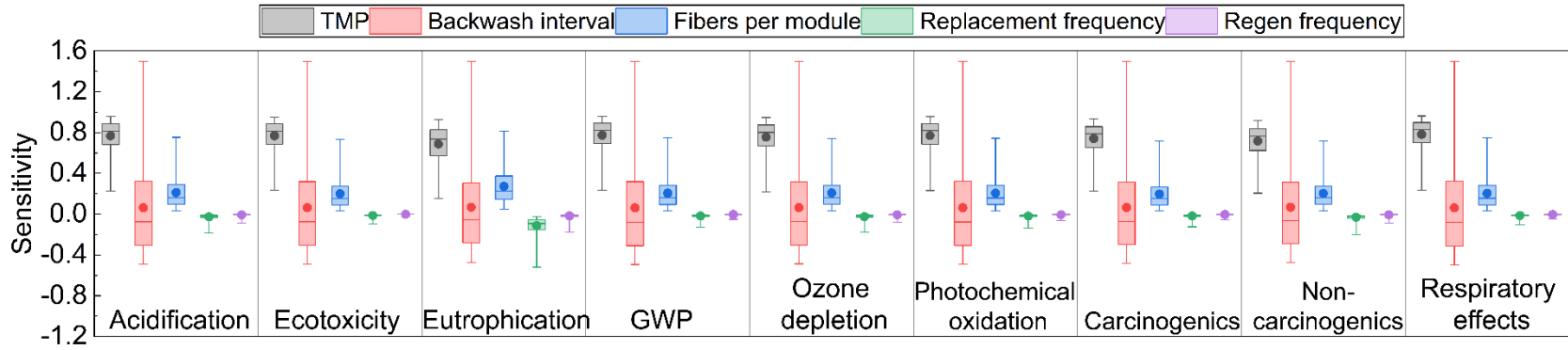
282 Global sensitivity analysis

283 The global sensitivity analysis assesses the importance of each parameter in terms of further
 284 improving LFM's environmental performance by quantitatively ranking the sensitivity of each
 285 parameter over a comprehensive design space. The result suggests that TMP (permeability) is the
 286 most critical design/operation parameter to further reduce the life cycle environmental impacts of

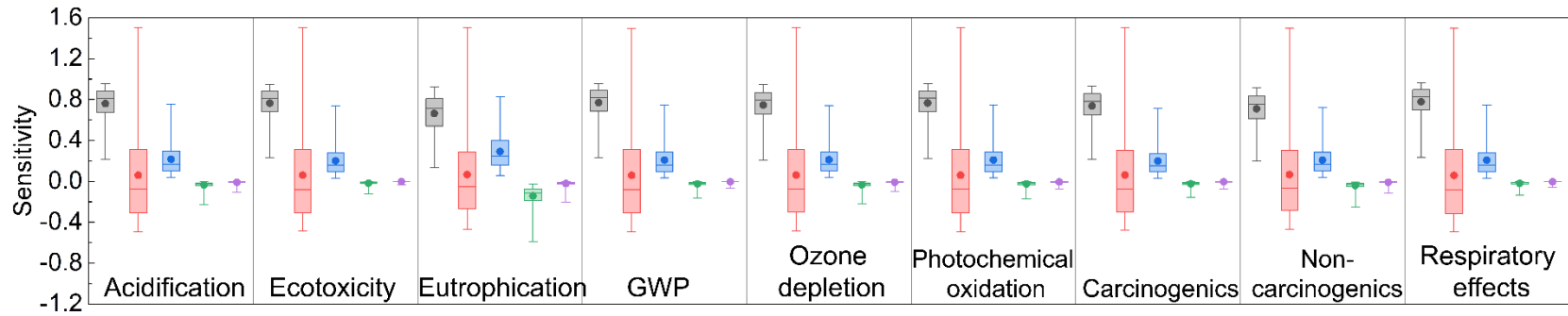
287 LFMs. Changing the TMP by 1% typically results in a 0.6-0.9% change in the life cycle impacts
288 of LFMs. This is statistically significantly higher ($p < 0.05$) than the sensitivities of the other four
289 parameters in most of the impact categories (Figure 3). For example, with untreated LFMs,
290 reducing the TMP by ~ 0.04 bar will reduce the GWP impact by ~ 7.0 kg CO₂ eq./FU, whereas
291 changing the LFM replacement frequency by 1% only changes the GWP impact by ~ 0.5 kg CO₂
292 eq./FU (Figure 3).

293 Comparing the three treatments, H₂O₂ treatment dampens the sensitivity of LFMs' life cycle
294 impacts to TMP as it increases the permeability of LFMs, whereas NaOH treatment enhances the
295 sensitivity as it decreases the permeability of LFMs (Figure 3). The changes in sensitivities due
296 to treatments, however, are typically insignificant ($< 2\%$). For example, the sensitivity of GWP to
297 TMP is 76.0%, 77.4%, and 77.8% for untreated, NaOH-treated, and H₂O₂-treated LFMs
298 respectively (Figure 3).

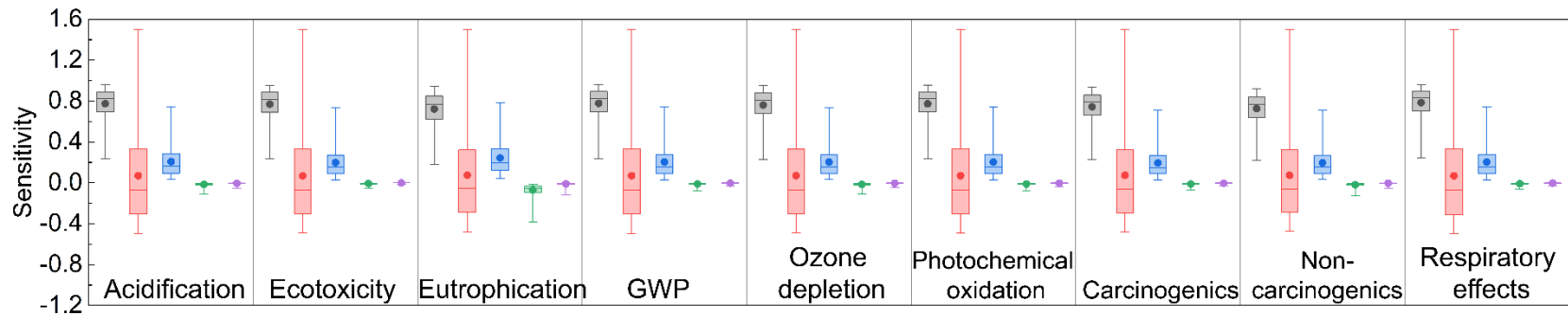
299 To reduce the eutrophication impacts of LFMs, which is LFMs' worst performing category
300 relative to the polymer membrane (Figure 2), increasing the permeability is still the most
301 effective measure (Figure 3). A 1% improvement in permeability can reduce the eutrophication
302 impacts of LFMs by $\sim 0.6\%$ across all treatment types (Figure 3). This is aligned with the results
303 that electricity remains the biggest contributor to LFMs' eutrophication impacts, despite the
304 increased relative importance of LFM fiber production, regeneration and disposal (Figure 2). If it
305 is desirable to reduce the impacts of those processes, new engineering ideas are needed, because
306 simply reducing the frequency of LFM fiber regeneration and replacement have little impacts on
307 the overall eutrophication impacts (sensitivities ~ 0 in Figure 3).



No Treatment



NaOH Treatment



H₂O₂ Treatment

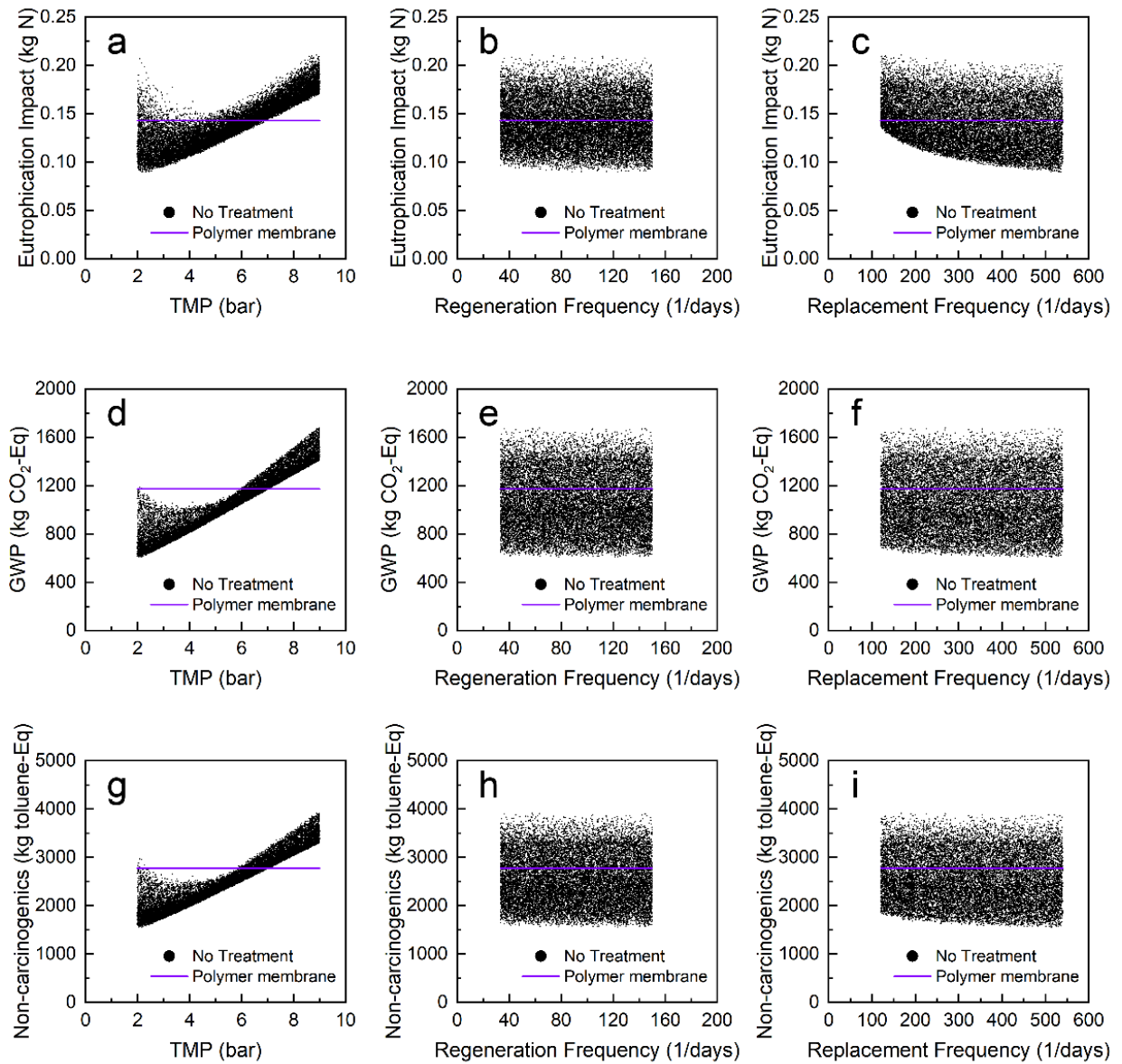
Figure 3. Sensitivity of LFM's environmental performance to each parameter and treatment

310 One-factor-at-a-time sensitivity analysis

311 All 16,807 simulation results are visualized to assess whether the sensitivity of each parameter
312 on final impacts is range-dependent. Figure 4 confirms that the sensitivity of life cycle impacts to
313 the TMP, regeneration frequency, and replacement frequency is rather consistent throughout the
314 entire range of simulated values (Figure 4a, d and g). For example, increasing the TMP from 2 to
315 9 bars results in a narrow band of consistently increasing GWP impacts (Figure 4a), suggesting
316 that TMP dominates the other parameters throughout the range simulated. In contrast, increasing
317 the regeneration frequency from 30 to 150 times/year results in a wide band of randomly
318 changing GWP impacts (Figure 4b), indicating that the impact of regeneration frequency on the
319 overall GWP impacts is overshadowed by other factors.

320 Sensitivity of the eutrophication impacts to the LFM fiber lifespan is range-dependent. A
321 narrower band in the high-impact range is seen when LFM fibers need to be replaced frequently
322 (one replacement per 100-200 days, Figure 4f), indicating a larger impact of LFM fiber lifespan
323 on the final eutrophication impact. In contrast, the impact of LFM fiber lifespan is dominated by
324 the other factors when it reaches the high range (one replacement per 400-500 days), as
325 suggested by a wider band of impact values (Figure 4f).

326 Different LFM treatments follow similar trends as the untreated LFMs (Figure S1-2). The NaOH
327 treatment reduces the sensitivity of life cycle impacts to the design and operation parameters,
328 while the H₂O₂ treatment enhances the respective sensitivities (Figure S2-3).



329

330

331

332

333

334

335

Figure 4. Sensitivity of GWP, eutrophication, and non-carcinogenics to TMP, regeneration frequency, and replacement frequency with untreated LFM. Each black dot is the result of one simulation. The purple line in all the figures is the respective impact of polymer membranes under 2000 fibers per module, no regeneration needed, replacement frequency 0.1 times/year, TMP 4 bar, and backwash at 15% flux loss.

336 In summary, the results reveal that permeability is the most important parameter to target to
337 further improve the life cycle environmental performance of LFMs. It can be accomplished by
338 further improving the fouling resistance or increasing the permeate flux through chemical
339 treatments (e.g. the H₂O₂ treatment compared in this study), as demonstrated in previous studies
340 of LFMs and polymer membranes.³⁸ Between the two options, improving the fouling resistance
341 can improve all nine impact categories, whereas the chemical treatments studied so far are only
342 effective at reducing the eutrophication impacts. In addition, with chemical treatments, the
343 tradeoff between environmental and technical performance needs to be balanced carefully. While
344 chemical treatments can improve the environmental performance of LFMs, they also change the
345 technical performance (i.e. size cutoff).^{4,39}

346

347

348 **ACKNOWLEDGEMENTS**

349 Research was sponsored by the US National Science Foundation (Grant Number: 1706097) and
350 the Combat Capabilities Development Command Army Research Laboratory (Cooperative
351 Agreement Number W911NF-15-2-0020). The views and conclusions contained in this
352 document are those of the authors and should not be interpreted as representing the official
353 policies, either expressed or implied, of the Combat Capabilities Development Command Army
354 Research Laboratory or the U.S. Government. The U.S. Government is authorized to reproduce
355 and distribute reprints for Government purposes notwithstanding any copyright notation herein.

356

357 **Supporting Information Available**

358 Overview of LFM preparation and use; life cycle inventory data; simulation results for NaOH-
359 and H₂O₂-treated LFMs.

360

361 **References**

- 362 (1) Zodrow, K. Living Filtration Membranes. Provisional Patent Application US Application
363 No. 62880397.
- 364 (2) Yamanaka, S.; Watanabe, K.; Kitamura, N.; Iguchi, M.; Mitsuhashi, S.; Nishi, Y.; Uryu, M.
365 *The Structure and Mechanical Properties of Sheets Prepared from Bacterial Cellulose*;
366 1989; Vol. 24.
- 367 (3) Caro, G.; Zuluaga, R.; Mondragon, I.; Gañán, P.; Putaux, J.-L.; Castro, C. Structural
368 Characterization of Bacterial Cellulose Produced by *Gluconacetobacter Swingsii* Sp. from
369 Colombian Agroindustrial Wastes. *Carbohydr. Polym.* **2010**.
370 <https://doi.org/10.1016/j.carbpol.2010.10.072>.
- 371 (4) Eggensperger, C. G.; Giagnorio, M.; Holland, M. C.; Dobosz, K. M.; Schiffman, J. D.;
372 Tiraferri, A.; Zodrow, K. R. Sustainable Living Filtration Membranes. *Environ. Sci.*
373 *Technol. Lett.* **2020**. <https://doi.org/10.1021/acs.estlett.0c00019>.
- 374 (5) Nguyen, T.; Roddick, F. A.; Fan, L. Biofouling of Water Treatment Membranes: A Review
375 of the Underlying Causes, Monitoring Techniques and Control Measures. *Membranes*
376 **2012**, 2 (4), 804–840. <https://doi.org/10.3390/membranes2040804>.
- 377 (6) Flemming, H.-C.; Schaule, G.; Griebe, T.; Schmitt, J.; Tamachkiarowa, A. Biofouling—the
378 Achilles Heel of Membrane Processes. *Desalination* **1997**, 113 (2–3), 215–225.
379 [https://doi.org/10.1016/S0011-9164\(97\)00132-X](https://doi.org/10.1016/S0011-9164(97)00132-X).
- 380 (7) Najjar, A.; Sabri, S.; Al-Gaashani, R.; Atieh, M. A.; Kochkodan, V. Antibiofouling
381 Performance by Polyethersulfone Membranes Cast with Oxidized Multiwalled Carbon
382 Nanotubes and Arabic Gum. *Membranes* **2019**, 9 (2).
383 <https://doi.org/10.3390/membranes9020032>.
- 384 (8) Geng, Z.; Yang, X.; Boo, C.; Zhu, S.; Lu, Y.; Fan, W.; Huo, M.; Elimelech, M.; Yang, X.
385 Self-Cleaning Anti-Fouling Hybrid Ultrafiltration Membranes via Side Chain Grafting of
386 Poly(Aryl Ether Sulfone) and Titanium Dioxide. *J. Membr. Sci.* **2017**, 529, 1–10.
387 <https://doi.org/10.1016/j.memsci.2017.01.043>.
- 388 (9) Sun, J.; Hu, C.; Tong, T.; Zhao, K.; Qu, J.; Liu, H.; Elimelech, M. Performance and
389 Mechanisms of Ultrafiltration Membrane Fouling Mitigation by Coupling Coagulation and
390 Applied Electric Field in a Novel Electrocoagulation Membrane Reactor. *Environ. Sci.*
391 *Technol.* **2017**, 51 (15), 8544–8551. <https://doi.org/10.1021/acs.est.7b01189>.
- 392 (10) Wang, H.; Park, M.; Liang, H.; Wu, S.; Lopez, I. J.; Ji, W.; Li, G.; Snyder, S. A. Reducing
393 Ultrafiltration Membrane Fouling during Potable Water Reuse Using Pre-Ozonation. *Water*
394 *Res.* **2017**, 125, 42–51. <https://doi.org/10.1016/j.watres.2017.08.030>.
- 395 (11) Holland, M.; Eggensperger, C.; Giagnorio, M.; Schiffman, J. D.; Tiraferri, A.; Zodrow, K.
396 R. Facile Post-Processing Alters Permeability and Selectivity of Microbial Cellulose
397 Ultrafiltration Membranes. *Submitted*.
- 398 (12) Childress, A. E.; Le-Clech, P.; Daugherty, J. L.; Chen, C.; Leslie, G. L. Mechanical
399 Analysis of Hollow Fiber Membrane Integrity in Water Reuse Applications. *Desalination*
400 **2005**, 180 (1), 5–14. <https://doi.org/10.1016/j.desal.2004.12.026>.
- 401 (13) Cote, P.; Alam, Z.; Penny, J. Hollow Fiber Membrane Life in Membrane Bioreactors
402 (MBR). *Desalination* **2012**, 288, 145–151. <https://doi.org/10.1016/j.desal.2011.12.026>.
- 403 (14) Gijsbertsen-Abrahamse, A. J.; Cornelissen, E. R.; Hofman, J. A. M. H. Fiber Failure
404 Frequency and Causes of Hollow Fiber Integrity Loss. *Desalination* **2006**, 194 (1), 251–
405 258. <https://doi.org/10.1016/j.desal.2005.11.010>.

- 406 (15) Porcelli, N.; Judd, S. Chemical Cleaning of Potable Water Membranes: A Review. *Sep.*
407 *Purif. Technol.* **2010**, *71* (2), 137–143. <https://doi.org/10.1016/j.seppur.2009.12.007>.
- 408 (16) Petersen, N.; Gatenholm, P. Bacterial Cellulose-Based Materials and Medical Devices:
409 Current State and Perspectives. *Appl. Microbiol. Biotechnol.* **2011**, *91* (5), 1277.
- 410 (17) Kurniawan, H.; Ye, Y.-S.; Kuo, W.-H.; Lai, J.-T.; Wang, M.-J.; Liu, H.-S. Improvement of
411 Biofouling Resistance on Bacterial Cellulose Membranes. *Biochem. Eng. J.* **2013**, *78*, 138–
412 145.
- 413 (18) Yang, G.; Xie, J.; Deng, Y.; Bian, Y.; Hong, F. Hydrothermal Synthesis of Bacterial
414 Cellulose/AgNPs Composite: A “Green” Route for Antibacterial Application. *Carbohydr.*
415 *Polym.* **2012**, *87* (4), 2482–2487.
- 416 (19) Kollarigowda, R. H.; Abraham, S.; Montemagno, C. D. Antifouling Cellulose Hybrid
417 Biomembrane for Effective Oil/Water Separation. *ACS Appl. Mater. Interfaces* **2017**, *9*
418 (35), 29812–29819. <https://doi.org/10.1021/acsami.7b09087>.
- 419 (20) Leitão, A. F.; Silva, J. P.; Dourado, F.; Gama, M. Production and Characterization of a New
420 Bacterial Cellulose/Poly (Vinyl Alcohol) Nanocomposite. *Materials* **2013**, *6* (5), 1956–
421 1966.
- 422 (21) Wang, H.; Zhu, E.; Yang, J.; Zhou, P.; Sun, D.; Tang, W. Bacterial Cellulose Nanofiber-
423 Supported Polyaniline Nanocomposites with Flake-Shaped Morphology as Supercapacitor
424 Electrodes. *J. Phys. Chem. C* **2012**, *116* (24), 13013–13019.
- 425 (22) Shibazaki, H.; Kuga, S.; Onabe, F.; Usuda, M. Bacterial Cellulose Membrane as Separation
426 Medium. *J. Appl. Polym. Sci.* **1993**, *50* (6), 965–969.
- 427 (23) Zeman. *Microfiltration and Ultrafiltration : Principles and Applications*; Routledge, 2017.
428 <https://doi.org/10.1201/9780203747223>.
- 429 (24) Bovea, M. D.; Pérez-Belis, V. A Taxonomy of Ecodesign Tools for Integrating
430 Environmental Requirements into the Product Design Process. *J. Clean. Prod.* **2012**, *20* (1),
431 61–71. <https://doi.org/10.1016/j.jclepro.2011.07.012>.
- 432 (25) Karlsson, R.; Luttrupp, C. EcoDesign: What’s Happening? An Overview of the Subject
433 Area of EcoDesign and of the Papers in This Special Issue. *J. Clean. Prod.* **2006**, *14* (15),
434 1291–1298. <https://doi.org/10.1016/j.jclepro.2005.11.010>.
- 435 (26) Ribera, G.; Clarens, F.; Martínez-Lladó, X.; Jubany, I.; Martí, V.; Rovira, M. Life Cycle
436 and Human Health Risk Assessments as Tools for Decision Making in the Design and
437 Implementation of Nanofiltration in Drinking Water Treatment Plants. *Sci. Total Environ.*
438 **2014**, *466*, 377–386.
- 439 (27) Carré, E.; Beigbeder, J.; Jauzein, V.; Junqua, G.; Lopez-Ferber, M. Life Cycle Assessment
440 Case Study: Tertiary Treatment Process Options for Wastewater Reuse. *Integr. Environ.*
441 *Assess. Manag.* **2017**, *13* (6), 1113–1121.
- 442 (28) Bonton, A.; Bouchard, C.; Barbeau, B.; Jedrzejak, S. Comparative Life Cycle Assessment
443 of Water Treatment Plants. *Desalination* **2012**, *284*, 42–54.
- 444 (29) Song, L. Flux Decline in Crossflow Microfiltration and Ultrafiltration: Mechanisms and
445 Modeling of Membrane Fouling. *J. Membr. Sci.* **1998**, *139* (2), 183–200.
446 [https://doi.org/10.1016/S0376-7388\(97\)00263-9](https://doi.org/10.1016/S0376-7388(97)00263-9).
- 447 (30) Field, R. W.; Wu, D.; Howell, J. A.; Gupta, B. B. Critical Flux Concept for Microfiltration
448 Fouling. *J. Membr. Sci.* **1995**, *100* (3), 259–272. [https://doi.org/10.1016/0376-7388\(94\)00265-Z](https://doi.org/10.1016/0376-7388(94)00265-Z).
- 449

- 450 (31) Miller, D. J.; Paul, D. R.; Freeman, B. D. A Crossflow Filtration System for Constant
451 Permeate Flux Membrane Fouling Characterization. *Rev. Sci. Instrum.* **2013**, *84* (3),
452 035003. <https://doi.org/10.1063/1.4794909>.
- 453 (32) McKay, M. D.; Beckman, R. J.; Conover, W. J. Comparison of Three Methods for
454 Selecting Values of Input Variables in the Analysis of Output from a Computer Code.
455 *Technometrics* **1979**, *21* (2), 239–245. <https://doi.org/10.1080/00401706.1979.10489755>.
- 456 (33) Morris, M. D. Factorial Sampling Plans for Preliminary Computational Experiments.
457 *Technometrics* **1991**, *33* (2), 161–174. <https://doi.org/10.1080/00401706.1991.10484804>.
- 458 (34) van Griensven, A.; Meixner, T.; Grunwald, S.; Bishop, T.; Diluzio, M.; Srinivasan, R. A
459 Global Sensitivity Analysis Tool for the Parameters of Multi-Variable Catchment Models.
460 *J. Hydrol.* **2006**, *324* (1), 10–23. <https://doi.org/10.1016/j.jhydrol.2005.09.008>.
- 461 (35) Tangsubkul, N.; Parameshwaran, K.; Lundie, S.; Fane, A. G.; Waite, T. D. Environmental
462 Life Cycle Assessment of the Microfiltration Process. *J. Membr. Sci.* **2006**, *284* (1–2), 214–
463 226.
- 464 (36) Ortiz, M.; Raluy, R. G.; Serra, L. Life Cycle Assessment of Water Treatment Technologies:
465 Wastewater and Water-Reuse in a Small Town. *Desalination* **2007**, *204* (1–3), 121–131.
- 466 (37) Godskesen, B.; Hauschild, M.; Rygaard, M.; Zambrano, K.; Albrechtsen, H.-J. Life-Cycle
467 and Freshwater Withdrawal Impact Assessment of Water Supply Technologies. *Water Res.*
468 **2013**, *47* (7), 2363–2374. <https://doi.org/10.1016/j.watres.2013.02.005>.
- 469 (38) Goosen, M. F. A.; Sablani, S. S.; Al-Hinai, H.; Al-Obeidani, S.; Al-Belushi, R.; Jackson, D.
470 Fouling of Reverse Osmosis and Ultrafiltration Membranes: A Critical Review. *Sep. Sci.*
471 *Technol.* **2005**, *39* (10), 2261–2297. <https://doi.org/10.1081/SS-120039343>.
- 472 (39) Mehta, A.; Zydney, A. L. Permeability and Selectivity Analysis for Ultrafiltration
473 Membranes. *J. Membr. Sci.* **2005**, *249* (1), 245–249.
474 <https://doi.org/10.1016/j.memsci.2004.09.040>.
- 475