

## OUTDOOR/INDOOR AIR QUALITY IN PRIMARY SCHOOLS IN LISBON: A PRELIMINARY STUDY

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OUTDOOR/INDOOR AIR QUALITY IN PRIMARY SCHOOLS IN LISBON: A PRELIMINARY STUDY. Simultaneous measurements of outdoor and indoor pollution were performed at three schools in Lisbon. Volatile organic compounds (VOCs), formaldehyde and NO<sub>2</sub> were passively monitored over a two-week period. Bacterial and fungal colony-forming units and comfort parameters were also monitored at classrooms and playgrounds. The highest indoor levels of CO<sub>2</sub> (2666 µg/m<sup>3</sup>), NO<sub>2</sub> (40.3 µg/m<sup>3</sup>), VOCs (10.3 µg/m<sup>3</sup>), formaldehyde (1.03 µg/m<sup>3</sup>) and bioaerosols (1634 CFU/m<sup>3</sup>), and some indoor/outdoor ratios greater than unity, suggest that indoor sources and building conditions might have negative effects on air indoors. Increasing ventilation rates and use of low-emission materials would contribute towards improving indoor air quality.

Keywords: indoor air quality; VOCs; formaldehyde.

## INTRODUCTION

Outdoor air quality has become of growing concern during the past 50 years, because of increasing traffic and industrial emissions. However, evidence has been found that citizen spend most of their time in buildings and are far more exposed to pollution indoors than outdoors.<sup>1</sup>

In Lisbon, the number of children with asthma and rhinitis represents, respectively, about 15 and 40% of the school-age population<sup>2</sup> and almost nothing is known about indoor air quality (IAQ) in schools. Mendell *et al.*<sup>3</sup> observed that health problems from poor indoor environments may reduce the performance of occupants in buildings. According to Mendell and Heath,<sup>4</sup> indoor environments in schools need to be studied with the aim of finding connections between IAQ and performance or attendance, due to two primary reasons: (i) schools are seen as particularly likely to have environmental deficiencies because chronic shortages of funding contribute to inadequate operation and maintenance of facilities. (ii) children have greater susceptibility to some environmental pollutants than adults, because they breathe higher volumes of air relative to their body weights and their tissues and organs are actively growing. In addition, a child's immune system is not fully developed. Currently, children also spend more time in school than in any indoor environment other than their home.

Persuasive evidence links higher indoor NO<sub>2</sub> concentrations to reduced school attendance and low ventilation rates to reduced performance. Concerning indirect associations, some studies link indoor dampness and microbiologic pollutants to asthma exacerbations and respiratory infections, which in turn have been related to reduced performance and attendance.<sup>4,5</sup>

The first aim of this preliminary study was to measure comfort parameters, CO<sub>2</sub>, bacterial and fungal contamination and gaseous inorganic and organic pollutants in indoor and outdoor air of three schools in winter. The second aim was to study associations between these factors and possible sources inside or outside the schools. As far as we know, this is the first IAQ monitoring study in schools of Lisbon.

## EXPERIMENTAL

## Schools description

Indoor and outdoor air samples were collected at three schools (183, SJB and SJ) in Lisbon (Portugal), in December 2008. These schools were located in the city centre and were previously considered representative of all the elementary-level educational institutions.<sup>2,6</sup> Two classrooms from each of the three schools were selected for this study. One of the classrooms of both 183 and SJB schools always had the electric heating connected and closed windows. This classroom of the 183 School presented activities of arts with paints and glue in one day during the sampling; the other classrooms had windows and doors opened frequently. In the SJ School, both classrooms were always shut. All the classrooms depend only on the natural ventilation through the doors and windows existent. Details of each sampling site are listed in Table 1.

Table 1. Characteristics of each school

	183	SJB	SJ
Environment	Urban	Urban	Urban
Heating	Yes	Yes	No
Ventilation		Windows/Doors	
Type of board	Blackboard and chalk	White board with pen	Blackboard and chalk
Floor		Ceramic tile	
Material of desks and chairs		Wood, plywood, plastic and metal	
Plants	Outdoor	Outdoor	Indoor/Outdoor
Animals	No	Yes	No

## Sampling and analysis

Pollutants and parameters of interest were carbon dioxide (CO<sub>2</sub>), temperature, relative humidity (RH), total VOCs, bacterial and fungal colony-forming units per cubic metre, NO<sub>2</sub>, speciated VOCs and formaldehyde. Continuous measurements of temperature, relative humidity (RH), CO<sub>2</sub> and total VOCs were

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performed with an automatic portable Indoor Air IQ-610 Quality Probe (GrayWolf® monitor) in one classroom of each school. This IAQ monitor includes a Pt100 probe for measuring temperature, a capacitance probe to sense RH and a CO<sub>2</sub> non-dispersive infrared sensor, all of them with an extremely fast response. The monitor also includes a photo-ionisation detector to track total VOCs over time. It displays measurements in real time allowing logged data to be downloaded to WolfSense® PC software for analysis. The equipment was supplied with a factory calibration certificate, but it is checked prior to next use with appropriate calibration kits. Based on their results, it was expected to evaluate the contribution of ventilation, combustion processes, tobacco smoke and traffic for the IAQ. Bacterial and fungal colony-forming units per cubic metre of air were monitored by liquid impinger sampling in the two classrooms and playgrounds during one day in each school selected.<sup>7</sup> Passive samplers for VOCs, formaldehyde and NO<sub>2</sub> were used for the simultaneous measurements of indoor (in one classroom of each school) and outdoor levels. At each point, samples were collected in duplicate. NO<sub>2</sub> concentrations were passively monitored for a two-week period. The diffusive tubes (with steel grids impregnated with triethanolamine) chemiadsorb NO<sub>2</sub>, as nitrite, which was quantified by visible spectrophotometry.<sup>8</sup> Passive samplers for VOCs and formaldehyde from Radiello®<sup>9,10</sup> were used to obtain a screening of heavy and light molecular weight compounds over a two-week period. Indoor passive samples were collected at a height of about 1.5 m above the floor. They were positioned at a distance that should exceed 1 m from a window or a door. Outdoor passive samples were collected at heights of about 2 m above the ground. VOCs adsorbed in activated charcoal cartridges were recovered by 2 mL of carbon disulfide (CS<sub>2</sub>) with the internal standard, during 30 min. Analyses were performed by gas chromatography (Chrompack CP 9001) coupled to a flame ionisation detection (GC/FID), using nitrogen carrier gas at constant pressure of 20 psi.<sup>9</sup> A 100% dimethylpolysiloxane column (0.2 mm, 50 m, film thickness 0.5 µm) was used under the following temperature program: 50 °C for 5 min, 5 °C/min up to 80 °C, 15 °C/min up to 135 °C, 20 °C/min up to 220 °C, final isotherm for 20 min. Injector and detector temperatures were 240 and 300 °C, respectively. The equipment was calibrated before and during the analyses of samples by injecting standard solutions of all compounds identified in CS<sub>2</sub>, specifically: pentane, *n*-hexane, cyclohexane, *n*-heptane, *n*-butyl acetate, styrene,  $\alpha$ -pinene, sabinene,  $\beta$ -pinene, *n*-decane, (+)-3-carene, limonene (all from Fluka), methyl acetate, ethyl acetate, isooctane, *m,p*-xylene, *o*-xylene (all from Merck), benzene (AnalytiCals), toluene (Lab-Scan), and  $\gamma$ -terpinene (Aldrich). Four standard solutions, each one containing five compounds in CS<sub>2</sub>, have been prepared. The analytes in these four standard solutions were present in concentrations of 40, 20, 10 and 5 ng/µL.

Formaldehyde collected in the 2,4-dinitrophenylhydrazine in sampling cartridges reacted to give the corresponding 2,4-dinitrophenylhydrazones. The analytes were extracted with 2 mL of acetonitrile and analysed by high-performance liquid chromatography (HPLC). The analytical system consisted of a Jasco PU-980 pump, a Rheodyne manual injection valve (sample loop of 20 µL), a Supelcosil LC-18 column (250×4.6 mm; 5 µm; Supelco) and a Jasco MD-1510 diode array detector, all connected in series. Isocratic elution at room temperature was performed using an acetonitrile/water solution (60/40, v/v) as the mobile phase at a flow rate of 1.5 mL/min. The carbonyl concentrations were quantified with external calibrations curves constructed from standard solutions of formaldehyde-DNPH derivative in acetonitrile.<sup>10,11</sup> After compilation of data, the different environments were compared with the aim of finding a relation between indoor and outdoor pollutants and the possible compound sources.

## Evaluation of the chromatographic analysis

Parameters, such as selectivity, linearity, reproducibility and limit of detection, were evaluated by 12 injections of three standard solutions of 10 compounds with 3 concentrations each one, between 5 and 40 ng/µL. The selectivity of an instrumental separation method refers to the ability to discriminate between the analyte and interfering components.<sup>12</sup> As condition for the method selectivity, the absence of peaks in the region of the retention time for the investigated compounds was observed. Linearity is the ability to elicit test results that are directly proportional to the concentration of analytes in samples within a given range.<sup>12</sup> The repeatability measures are the success rate in successive experiments conducted by the same experimenters. It was evaluated from the calculation of the standard deviation of the chromatographic peak areas corresponding to 10-12 injections, each day, in 5 successive days. The limit of quantification (LOQ) represents the lesser concentration of the substance in examination that can be quantitatively analysed with reasonable reliability. Limit of detection (LOD) represents the lesser concentration of the substance in examination that can be detected, but it is not necessarily quantified by a method. The LOQ and LOD have been calculated as described in Ribani *et al.*<sup>12</sup>

## RESULTS AND DISCUSSION

### Results of the evaluation of the chromatographic analysis

The chromatographic analyses of samples exhibited good selectivity and separation capability of analytes. After multiple injections of different concentration standard solutions, it was observed that the plots of peak areas, as a function of analyte mass, produced regression lines that had an intercept not significantly different from 0 and Pearson correlation coefficients ranging from 0.958 to 0.999 (Table 2). Consecutive injections of the same sample under variable conditions showed repeatability among measurements. The maximum standard deviation did not exceed 0.04 (Table 2). Depending on the analyte, LOQ and LOD were in the ranges 1.04-7.64 and 0.34-2.52 ng/µL, respectively (Table 2).

**Table 2.** Average relative response factor, standard deviation (STDEV), linearity, limit of detection and limit of quantification for each compound

Compounds	Average RRF*	STDEV	Pearson correlation coefficients	LOD** (ng/µL)	LOQ*** (ng/µL)
Ethyl acetate	0.27	0.01	0.999	2.52	7.64
Cyclohexane	0.82	0.02	0.999	1.35	4.10
Isooctane	0.89	0.02	0.996	1.06	3.22
<i>n</i> -Heptane	0.97	0.02	0.958	1.03	3.11
Toluene	1.15	0.02	0.999	0.74	2.25
Internal standard	1.00	0.00	0.999	1.06	3.21
<i>o</i> -Xylene	1.28	0.04	0.999	0.43	1.31
$\beta$ -Pinene	1.23	0.04	0.999	0.38	1.15
<i>n</i> -Decane	1.20	0.04	0.999	0.34	1.04
Limonene	1.15	0.04	0.999	0.38	1.14

\*RRF (relative response factor) = (area of compound/mass of compound) × (area of internal standard/mass of internal standard); \*\*LOD (limit of detection) = 3.3(s/S), where s is the STDEV of areas and S is the slope; \*\*\*LOQ (limit of quantification) = 10(s/S), where s is the STDEV of areas and S is the slope.

### Air quality monitoring data

The average room temperature for the three schools was 20 ± 1.4 °C, and the relative humidity presented values between 52 and 61%.

These high indoor relative humidity values are not surprising since, according to the Environment Portuguese Agency, the Lisbon region usually records values between 75 and 85%. It should be stated that higher relative humidity values contribute to the survival and the dispersion of airborne allergens such as mould spores and bacteria, worsening the symptoms of allergy sufferers.

Carbon dioxide concentrations are often used as a surrogate of the rate of outside supply air per occupant. Indoor CO<sub>2</sub> concentrations above approximately 1000 µg/L are generally regarded as indicative of ventilation rates that are unacceptable with respect to body odours.<sup>13</sup> The National System for Energy and Indoor Air Quality Certification of Buildings (*Regulamento dos Sistemas Energéticos de Climatização de Edifícios - RSECE*)<sup>14</sup> establishes an acceptable maximum value (AMV) of CO<sub>2</sub> of 1800 µg/L for buildings in Portugal. The indoor concentrations of CO<sub>2</sub> showed inadequate classroom air exchange rates. Figure 1 depicts the variation of indoor CO<sub>2</sub> concentrations in a typical working day at the three schools. A strong correlation of the CO<sub>2</sub> level with occupancy has been observed. CO<sub>2</sub> spikes were even more pronounced when students started physical activities inside the classrooms, as for example, art classes or entrance and exits to the playgrounds. Room 12 of the 183 School presented the greatest CO<sub>2</sub> concentration (2666 µg/m<sup>3</sup>). This room was the only one that had the electric wall heating constantly connected and windows and doors always closed. Lower outdoor air ventilation rates at homes have been associated with increased prevalence of asthma and allergic symptoms in children.<sup>15</sup>

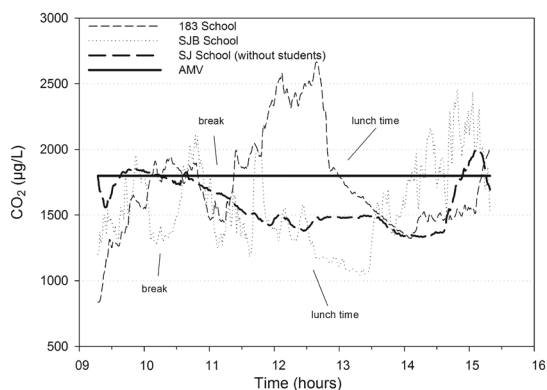


Figure 1. Indoor carbon dioxide levels in the three schools

No standards have been agreed upon for nitrogen oxides in indoor air in Portugal. ASHRAE<sup>13</sup> and the U.S. EPA National Ambient Air Quality Standards list 0.053 ppm as the average 24-h limit for NO<sub>2</sub> in outdoor air. NO<sub>2</sub> concentrations were higher outdoors than indoors (Table 3), probably as a result of vehicular exhaust emissions from nearby traffic. The I/O NO<sub>2</sub> ratio ranged between 0.63 and 0.84. SJ School, which is located near an avenue with intense traffic, presented the smallest level of indoor NO<sub>2</sub>, possibly because the windows and the doors were always closed.

Table 3. Indoor and outdoor NO<sub>2</sub> concentrations (µg/m<sup>3</sup>) in the three schools

	Indoor	Outdoor	Indoor/Outdoor
SJB School	40.3	48.0	0.84
SJ School	36.4	56.9	0.64
183 School	37.1	44.4	0.83

In the SJ and SJB Schools, the total fungal colony-forming units in both indoor and outdoor air (Figure 2) were below the acceptable maximum value (AMV) of 500 CFU/m<sup>3</sup> defined by the Portuguese

Legislation, Decree-Law 79/2006. In the 183 School, fungal colony-forming units higher than this standard were observed in both indoor and outdoor air. Fungal species exceeding 500 CFU/m<sup>3</sup> may be indicative of building-related sources, poor ventilation rates or overcrowding, highlighting the need for remedial action.<sup>16</sup>

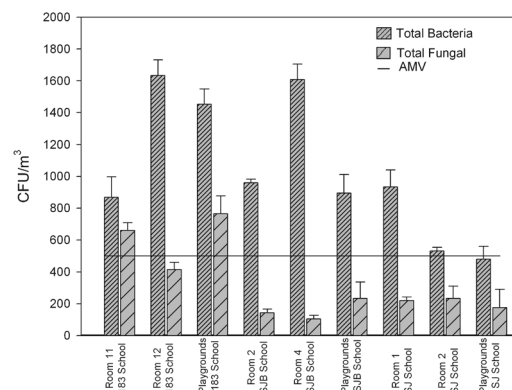


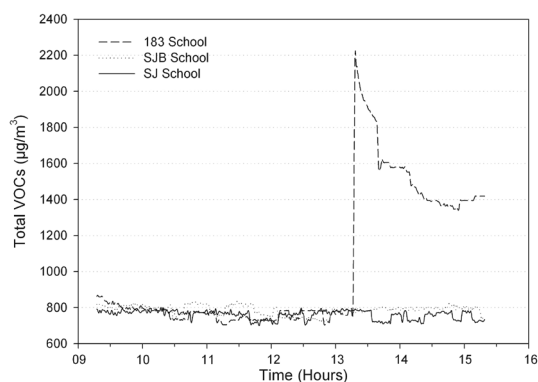
Figure 2. Average of total bacteria and fungal colony-forming units per cubic metre of air and standard deviation

Excepting for the outdoor measurements of SJ School, the total bacteria colony-forming units presented values above 500 CFU/m<sup>3</sup> for all the environments. The main factors affecting atmospheric dispersion and survival of microorganisms are the relative humidity, temperature, oxygen, wind and air turbulence, air pollutants and water and nutrient availability. The high amounts of bacteria in both indoor and outdoor may derive from several factors, including high seasonal level of bioaerosols in outdoor air, from the human self-activities, such as breathing, sweating and movement causing particle resuspension. Cold weather favours children's respiratory infections, which are usually caused by bacteria or virus. Thus, respiratory morbidity among children may also contribute to the airborne spread of bioaerosols.

Two institutions presented indoor/outdoor (I/O) fungal ratios in the range 0.45–0.86, while values higher than 1 have been registered for the SJ School. Depending on classroom, variable I/O bacterial ratios, ranging from 0.62 and 1.95, have been found. Scheff *et al.*<sup>17</sup> reported that, in a middle school of Springfield, the indoor fungal and bacterial counts were significantly higher than the outdoor concentrations. Conversely, Godwin and Batterman<sup>18</sup> found that the outdoor bioaerosol levels exceeded indoor levels in 64 school classrooms in Michigan.

Total VOC concentrations could give information about the influence of aerosol sprays, solvents, cleaning agents, pesticides, paints and repellents. Figure 3 shows a daily profile for the total VOC concentrations. SJ and SJB Schools exhibit very constant levels and similar daily patterns. In the 183 School, there was a huge increase in the VOC concentrations around 13 pm, when pupil's art class was occurring with the use of glue and paints. This makes evident that collage and painting materials increase the VOC levels in indoor air. Zhang *et al.*<sup>19</sup> also identified a visual art classroom with a relatively high level of VOCs. Standards have been agreed upon for total VOCs by Decree 79/2006 of the Portuguese Legislation that establishes the thermal regulations for buildings.<sup>14</sup> The results obtained in schools were above the proposed target guideline value of 600 µg/m<sup>3</sup>.

Table 4 present the results for the VOC concentrations and speciation. In general, concentrations of VOCs were higher indoors than outdoors for all schools. Those compounds that have only been detected in indoor air have a probable indoor source. Ethyl acetate, methyl acetate, styrene, ethanol and limonene were only found in the indoor air. Pollutants identified in both indoor and outdoor samples,



**Figure 3.** Diurnal variation of total VOCs (non-methane hydrocarbons) in the three schools

but with higher concentrations in the indoor environments, may indicate additional indoor sources or inadequate ventilation ratios. I/O ratios higher than 1 were observed for *n*-hexane, *n*-heptane, *n*-butyl-acetate and *o*-xylene at all the schools. I/O ratios exceeding the unity were also determined for pentane, toluene  $\alpha$ -pinene, *n*-decane and terpinene, but not in all institutions. The measured benzene concentrations were below the annual ambient EU limit value of 5  $\mu\text{g}/\text{m}^3$ . Toluene concentrations were higher than those reported by Stranger *et al.*<sup>20</sup> in primary schools of Antwerp, Belgium. The high benzene and toluene concentrations observed in Lisbon are in the same range of those measured in schools of Oporto, Portugal.<sup>21</sup> In this preliminary study, the toluene levels were very similar to those found in schools of Curitiba, Brazil.<sup>22</sup>

**Table 4.** VOC concentrations ( $\mu\text{g}/\text{m}^3$ ) in the three schools

COMPOUNDS	SJB School		183 School		SJ School	
	Indoor	Outdoor	Indoor	Outdoor	Indoor	Outdoor
Pentane	3.61	0.71	0.97	1.08	1.40	1.13
Methyl acetate	52	ni	34	ni	83	ni
Ethyl acetate	2.08	ni	1.30	ni	3.55	ni
<i>n</i> -Hexane	2.98	0.53	1.15	0.62	1.06	0.84
Benzene	2.88	<ld	3.01	3.13	2.54	2.46
Cyclohexane	0.87	0.22	0.17	0.13	1.60	0.16
Isooctane	1.19	0.15	0.16	0.19	0.44	0.21
<i>n</i> -Heptane	3.22	0.37	3.35	0.50	0.95	0.52
Toluene	10.3	2.00	2.51	2.58	4.59	2.93
<i>n</i> -Butyl acetate	4.18	0.62	1.41	0.87	6.74	1.42
<i>m,p</i> -Xylene	8.8	1.22	1.42	1.26	2.82	1.78
Styrene	ni	ni	ni	ni	0.28	ni
<i>o</i> -Xylene	3.09	0.39	1.05	0.46	5.45	0.57
$\alpha$ -Pinene	0.50	ni	0.15	ni	4.27	0.16
Sabinene	0.77	ni	ni	0.14	12.2	0.17
$\beta$ -Pinene	ni	ni	ni	ni	29	ni
<i>n</i> -Decane	1.00	0.40	0.46	0.30	1.71	0.65
(+)-3-Carene	ni	ni	ni	ni	0.24	ni
$\gamma$ -Terpinene	0.65	ni	ni	0.18	0.78	0.23
Limonene	3.17	ni	0.39	ni	86	ni

ld – below limit of detection; ni – not identified.

SJ School, which has the oldest building among all institutions, registered both the highest concentrations and diversity of VOC compounds. Perhaps the inadequate ventilation observed favours accumulation of pollutants with additional indoor sources. The highest levels of limonene,  $\beta$ -pinene, sabinene, *n*-butyl acetate, methyl

acetate and formaldehyde ( $1.03 \mu\text{g}/\text{m}^3$ ) were achieved in this school. *n*-hexane, *n*-heptane and *n*-decane could have indoor sources in some architectural finishes, floor adhesives, PVC flooring, consumer products (e.g. floor waxes and aerosol air fresheners). Limonene could be derived from cleaning products, air fresheners and many other consumer products. Benzene, toluene, xylenes and styrene could be originated from engine vehicle exhaust, gasoline/fuel, tobacco smoke, solvent-based paints, floor adhesives, PVC flooring, carpeting, printed material and solvent-based consumer products.<sup>23</sup> The 183 School registered the lowest concentrations of VOCs, probably because this institution had better ventilation than the other schools, higher classroom volumes and lesser number of pupils.

## CONCLUSIONS

Indoor and outdoor concentrations of  $\text{NO}_2$ , VOCs, formaldehyde and microbiological components were measured for the first time in 3 elementary schools in Lisbon during December 2008. The results suggest that schools with closed windows could have smaller I/O ratios of  $\text{NO}_2$  (0.64), but higher indoor levels of VOCs ( $10.3 \mu\text{g}/\text{m}^3$ ) and formaldehyde ( $1.03 \mu\text{g}/\text{m}^3$ ) with an origin in building materials and consumer products. Total VOC concentrations increase during art classes, reaching about  $2200 \mu\text{g}/\text{m}^3$ . Low ventilation ratios and the children's physical activities have also an impact upon the  $\text{CO}_2$  levels. Fungal and bacterial counts exceeding  $500 \text{CFU}/\text{m}^3$  for one school ( $765 \text{CFU}/\text{m}^3$ ) and for all of them ( $934\text{--}1634 \text{CFU}/\text{m}^3$ ), respectively, may be indicative of building-related sources, poor air exchange rates or overcrowding, highlighting the need for remedial action. Most of the assessed gaseous pollutants can be credited to the traffic emissions and indoor sources (some architectural finishes, floor adhesives, PVC flooring, consumer products and cleaning products). More studies are needed (currently underway), to find additional possible sources of indoor contamination, to calculate air exchange rates on a seasonal basis, to evaluate if there is a causal relationship between pollutant exposure and health symptoms in schools, and to assess if school IAQ can adversely affect academic performance or attendance.

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