

1 **Energy Recovery by Fast Pyrolysis of Pre-treated Trommel Fines Derived from a UK-**
2 **based MSW Material Recycling Facility**

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10

11 **Abstract**

12 In this experimental study, a physically pre-treated trommel fines feedstock, containing 44 wt%
13 non-volatiles (ash and fixed carbon) and 56 wt% volatile matter (dry basis), was subjected to
14 fast pyrolysis to recover energy from its organic load, using a 300 g h⁻¹ bubbling fluidized bed
15 (BFB) fast pyrolysis rig. A physical pre-treatment method (including crushing, grinding and
16 sieving) was used to prepare a 0.5 – 2 mm sized trommel fines feedstock to make it suitable
17 for fast pyrolysis in the BFB reactor. Experimental results from the fast pyrolysis process
18 showed that the highest yield of organic liquid was obtained at around a temperature of 500
19 °C. However, both char and gas yields increased dramatically at temperatures above 500 °C,
20 as a result of enhanced cracking of liquid components, which reduced the yield of liquid
21 products. Overall, energy recovery from the pyrolysis products (liquid and gas products as
22 well as char pot residues) ranged from 63 – 70 %, generally increasing with temperature. A
23 large proportion of the high ash content (36 wt%) of the feedstock was found in the char pot
24 (>62%), while smaller proportions were found in the reactor bed and some liquid products.
25 The char pot ash residues composed mostly of non-hazardous earth materials and may be
26 applied in bulk construction materials e.g. cement manufacture. Although, there was no
27 problem with the pyrolysis rig during 1 hour of operation, longer periods of operation would
28 require periodic removal of accumulated solid residues and/or char pot modification to ensure
29 continuous rig operation and process safety.

30

31 *Keywords: pre-treated trommel fines, fast pyrolysis, energy recovery, pyrolysis liquid, solid*
32 *residues*

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37 **1 Introduction**

38 Trommel fines are a by-product obtained from the mechanical recycling process of municipal
39 solid waste (MSW). Landfill sites have been the traditional home for trommel fines, however
40 with the increase in landfill taxes and the amendment of the Landfill Tax Regulation,
41 processors are being forced to find alternative methods of handling trommel fines. The
42 introduction of the Loss on Ignition (LOI) test in April 2015 now requires UK-based waste
43 processors to ensure that LOI of trommel fines be less than 10% in order to be considered
44 eligible for the lower rate of tax (HMRC, 2014). In other words, lower LOI values indicate lower
45 calorific values. The solution is inevitable; waste processors must introduce methods for
46 trommel fines clean up to remove the calorific materials e.g. through energy recovery, in order
47 to reduce landfill tax bills. Energy recovery from trommel fines is a new frontier and big
48 questions arise as to what technologies would be suitable for this type of waste.

49

50 A detailed analysis of the organic composition of trommel fines (Eke, Onwudili and Bridgwater,
51 2017), showed that they contained different biodegradable and non-biodegradable materials
52 such as wood, stones, glass, ash, textiles, plastics, paper, cardboard, food waste and fibre.
53 The HMRC estimates that the UK produces about 4.5 million tonnes of trommel fines each
54 year (HMRC, 2014), representing just over 2 wt% of total annual waste generation. This mixed
55 composition means that trommel fines could be likened to a mixed stream of MSW, for which
56 biological technologies such as anaerobic digestion would not be suitable to eliminate the
57 organic load. In contrast, thermochemical processes are usually faster and more robust than
58 biological processes for mixed waste streams. For example, thermochemical processing can
59 handle both biodegradable and non-biodegradable waste, whereas biological processing can
60 only handle readily biodegradable waste (McKendry, 2002).

61

62 MSW pyrolysis can be considered as an innovative process for the pre-treatment MSW to
63 obtain cleaner energy carriers and chemicals (Schaefer, 1975; Malkow, 2004), than the
64 original feedstock. Indeed, pyrolysis gives intermediate products including gases and liquids,
65 which have better handling properties, suitable for downstream processing to more valuable
66 products than the original solid feedstock. Such downstream processing routes may include a
67 more controllable gasification process to convert the pyrolysis vapour into syngas or isolation
68 of high value chemicals from the condensed pyrolysis oil. In comparison, conventional MSW
69 mass-burn incineration process gives heat as the main product, thereby foreclosing the
70 potential for chemical added value. Additionally, the inert atmosphere used during pyrolysis
71 leads to lower emissions of nitrogen oxides (NO_x) and sulphur oxides (SO₂).

72

73 Numerous studies have been performed in laboratories to determine pyrolysis behaviours and
74 product properties of MSW and individual MSW fractions such as paper, sewage sludge,
75 plastics, fabric, lignocellulosic materials, refuse derived fuel (RDF) and RDF pellets
76 (Grammelis et al., 2009; Bosmans et al., 2013; Zhao et al., 2011; Ates et al., 2013; Miskolczi
77 et al., 2013). The interactions between the different individual fractions and the pyrolysis
78 products have also been explored (Sørum et al., 2001; Grieco and Baldi, 2012). Among the
79 listed feedstocks investigated for pyrolysis, RDF, sewage sludge and manures are known to
80 have significant ash contents. High ash contents pose operational problems during
81 thermochemical processing of wastes and can have significant effects on both the design and
82 operability of reactors. For example, accumulation of ash, stones and pieces of concrete in
83 beds of fast pyrolysis reactors can lead to hardware issues as well as rapid attrition of bed
84 material, leading to entrainment of small particles in gas flows, with consequent dust
85 deposition on downstream systems. In some cases, certain ash compounds, may be beneficial
86 in terms the chemistry of pyrolysis. For instance, some ash compounds may serve as catalyst
87 for the cracking of pyrolysis vapours, leading to the production of more valuable liquid and
88 gaseous products (Pradhan, 2015; Renkonen, 2016).

89

90 Therefore, this research has been carried out in an attempt to provide baseline data on the
91 technical feasibility of applying a 300 g h^{-1} bubbling fluidized bed fast pyrolysis system on
92 trommel fines from a UK-based waste recycling facility, which contained a very high ash
93 content. In this present work, fast pyrolysis of a physically pre-treated trommel fines feedstock
94 was investigated in relation to reaction temperature, as a means of recovering energy from
95 this type of problem waste. The aim was to reduce organic load of the feedstock, while leaving
96 an inert solid for final disposal at a reduced landfill cost due to lower LOI or for potential useful
97 application. Detailed analyses and characterisation of the pyrolysis products were carried out
98 to determine the yields and compositions of gaseous and liquid products, as well as the fate
99 of the ash during the fast pyrolysis process in order to draw up possible conclusions.

100

101 **2 Experimental methods**

102 2.1. Trommel fines feedstock

103 The sample used in this study was supplied by a UK commercial waste management company
104 – Biffa Ltd from Leicester. The sample constituted the <10 mm-sized mixed particles left after
105 mechanical sorting and trommel screening of household, thereby qualifying as trommel fines
106 (Fitzgerald, 2013; Pitchell, 2014). The estimated percentage composition of materials in the
107 trommel fines, carried out by visual inspection and hand separation, are given in Table 1. The
108 sample was initially prepared to meet the feeding requirements of an existing bubbling
109 fluidized bed fast pyrolysis reactor. Details of the pre-treatment procedure have been recently

110 published (Eke, Onwudili and Bridgwater, 2017). In brief, the preparation method involved
111 manual removal of visible large stones, glasses and concrete, followed by grinding with a
112 Retsch Ltd., Germany, Heavy-Duty Cutting Mill, Knife Mill Type SM2000 and sieving to obtain
113 a 0.5 - 2 mm size fraction. This fraction accounted for over 70 wt% of the original trommel
114 fines feedstock, which comprised of mostly paper pellets, plastics, glass, stones, bones,
115 miscellaneous wood, textiles and metals (Eke, Onwudili and Bridgwater, 2017). Table 2
116 summarises their main characteristics of this size fraction, which shows that it contained
117 approximately 56 wt% of volatile matter, from which liquid and gaseous products could be
118 obtained.

119

120

121 **2.2 Fast pyrolysis experiments**

122 Fast pyrolysis experiments were conducted in a bench-scale (300 g h⁻¹ throughput) fluidized
123 bed reactor unit shown in Figure 1 (Onwudili and Eke, 2019). Briefly, it consisted of a feeder,
124 a reactor and a product collection system. The feeder consisted of a tubular storage hopper
125 with a dual screw gravimetric feeding system with variable speed motor for feeding, attached
126 to a fast screw. The pyrolysis occurred in a steel tubular reactor with an internal diameter of
127 41 mm and a height of 320 mm. It was filled with approximately 150 g of sand with particle
128 size range of 500 – 600 µm as bed material for fluidization and heat transfer. The product
129 collection system was composed of a water condenser, two dry ice/acetone condensers and
130 a cotton wool filter.

131

132 For each experiment, fluidisation of the bed material was achieved by using inert nitrogen at
133 a flow rate of approximately 6 L min⁻¹. At the start, the nitrogen gas, preheated to 250 °C, was
134 used to remove air within the system for 10 min, followed by preheating the reactor using an
135 electrical furnace. Preliminary tests showed that to achieve the desired reactor temperature of
136 between 400 – 700 °C, the heater needed to be set 50 °C higher than the reactor temperature,
137 due to the endothermic nature of the pyrolysis reactions. Throughout the experiments,
138 temperatures were controlled within 6.5% of the set value. Once the temperatures of the
139 fluidising medium reached a steady state, trommel fines in the feeder were continuously
140 delivered into the reactor at the middle of the fluidised bed by nitrogen entrained flow via the
141 air-cooled feeding tube. In this work, actual feeding rate ranged from 120 to 220 g h⁻¹ and each
142 experiment lasted for 60 min. The pyrolysis temperature was the average value of temperature
143 data recorded by two K-type thermocouples placed within the fluidized-bed during the run.

144

145 The pyrolysis vapours, including aerosols, water and permanent gases produced from the
146 thermal degradation of the organic fractions of the trommel fines, as well as fine solids (char

147 and ash), were carried away from the reactor by the large flow of nitrogen. This is a main
148 feature of fluidized bed reactor systems. On reaching the cyclone, most of the entrained solids
149 were removed and transferred into the char pot. The vapours together with some solid
150 residues fines that could not be removed by the cyclone, then passed through a liquid product
151 collection system consisting of a water condenser, and a series of two dry ice/acetone
152 condensers, where most of the condensable liquids were collected. However, some more
153 volatile products could not be condensed and collected in the condensers and were trapped
154 in the cotton wool filter. This could be attributed to the low partial pressure of the volatiles in
155 the condensers due to the dilution effect of the large flow of nitrogen carrier gas. This problem
156 is widely known in large-scale pyrolysis plants and can be solved by process optimization,
157 which is a subject for further research. After the cotton wool filter, the nitrogen carrier gas
158 together with the non-condensable pyrolysis gases passed through the gas meter where the
159 volume of the exit gas was recorded. Each experiment was conducted three times and
160 average yields reported.

161

162 2.3 Analyses of fast pyrolysis products

163 2.3.1 Liquid product

164 In total, there were six fractions of the liquid products based on the collection point within the
165 pyrolysis rig. These include; (1) water-cooled condenser, (2) transition pipe 1, (3) dry ice
166 acetone condenser 1, (4) transition pipe 2, (5) dry ice acetone condenser 2, and (6) cotton
167 wool filter. For ease of handling all the liquid products obtained from fast pyrolysis process
168 were then classified into primary condensates (water-cooled condenser product) and
169 secondary condensate (the others). Both products consisted of mixtures of oil, water and fine
170 solids.

171

172 Volumetric Karl-Fischer (KF) titration was used to determine the water content of all the fast
173 pyrolysis primary and secondary condensate. The primary and secondary condensates were
174 dissolved in a known volume of acetone (1 part of condensate to 6 parts of acetone) prior to
175 analyses. Solids content in the primary condensate was determined using the vacuum filtration
176 technique suggested by Oasmaa and Peacocke (2001). The primary condensate was filtered
177 through a pre-dried and pre-weighed Whatman No. 2 qualitative filter paper with mean pore
178 size of 8 μm . The product was then washed with excess amount of acetone until the filtrate
179 was clear to ensure that there was no organic liquid left on the filter paper. The filter paper
180 with the residue was air-dried for approximately 15 min and in an oven at 105 °C for around 2
181 h, cooled in a desiccator and weighed. Solvent-free primary and secondary condensates were
182 analysed to determine their elemental (CHNS) analysis using CE-440 and Carlo Erba
183 elemental analysers with $\pm 0.3\%$ absolute accuracy. Oxygen was obtained by difference.

184

185 2.3.2 Pyrolysis liquid product analysis by GC-MS

186 Compositional analysis of liquid product samples was performed using a PerkinElmer Clarus
187 680 GC-MS system. GC samples were prepared by mixing liquid product with GC grade
188 acetone (1:6 vol/vol). 1 μL of the sample was filtered using a 0.2 μm pore size Sartorius filter,
189 and was injected into the GC column via an injection port maintained at 300 $^{\circ}\text{C}$, with 1:50 split
190 ratio. The transfer line was held at 275 $^{\circ}\text{C}$, while the detector temperature was at 250 $^{\circ}\text{C}$. The
191 GC oven programme was as follows: held constant at 50 $^{\circ}\text{C}$ for 2 mins, then ramped at 5 $^{\circ}\text{C}$
192 min^{-1} to 275 $^{\circ}\text{C}$ and held at 275 $^{\circ}\text{C}$ for 3 min. The programme lasted 50 min. Helium was used
193 as the carrier gas with a constant flow of 15 ml min^{-1} . A column splitter was used to enable
194 simultaneous detection of compounds separated on the columns by MS and FID detectors.
195 Mass spectra were obtained using 70 eV ionisation energy in the molecular mass range of
196 $m/z = 35 - 300$, with a scan time of 0.35 seconds. Assignments of the main peaks were made
197 from mass spectral detection (NIST05 MS library).

198

199 2.3.3 Solid product analysis

200 Total solids from the fast pyrolysis was a combination of the solid residues collected in the
201 char pot, reactor bed and metal transition pipe, and the solid fines or solids present in the
202 liquid products. However, the char pot residue accounted for about 90 wt% of the total solids
203 measured and was used for ash determination. The ash contents of char pot solid residues
204 were determined according to the ASTM D1762-84 method (ASTM, 2013). In the procedure,
205 between 4 and 5 g of solid residues was placed a pre-cleaned and pre-weighed crucible and
206 heated in a furnace and heated to 750 $^{\circ}\text{C}$ for 6 h. After cooling in a desiccator for one hour,
207 the crucible was weighed to the nearest 0.1 g. The ash content was obtained by difference
208 and expressed as wt% of the solid residues with standard deviation from five replicate
209 analyses. For each sample, the average of five samples was taken to reduce the deviation.
210 CHNS analyses of the solid residues were also carried out using the same CE-440 and Carlo
211 Erba elemental analysers.

212

213 2.3.4 Gas product analysis

214 During a pyrolysis run, the non-condensable gases were sampled every 3 min into a micro
215 gas chromatograph with a thermal conductivity detector (TCD) from Varian Chromatography
216 System Inc. (Kalgo, 2011; Banks, Nowakowski and Bridgwater, 2014). The gas components
217 were separated on two columns; namely Varian CP-5A Molsieve held isothermally at 80 $^{\circ}\text{C}$
218 (for H_2 , CO , N_2 and O_2) and CP-PortaPLOT held isothermally at 90 $^{\circ}\text{C}$ (for CO_2 and $\text{C}_1\text{-C}_4$
219 hydrocarbon gases prior to detection. Quantitation was achieved by external standard method
220 by calibrating the detector response using a standard gas mix containing these gases at 3

221 vol% concentrations in nitrogen. The mass yields of the gas components (nitrogen-free) were
222 calculated using the general gas equation, based on the gas volume composition obtained
223 from GC analysis, total gas volume and the exit gas temperature and recorded pressure. The
224 calorific values (as higher heating values, HHV) of gas product were estimated from the
225 volume percentage of each gas component and their higher heating values, according to Eq.1.

226
227 $HHV (MJ kg^{-1}) = \sum_{i=1}^n (x_i \cdot HHV_i) \dots\dots\dots Eq. 1.$

228 Where;

229 $i \dots n$ = each combustible component in the gas product

230 x = mass fraction of combustible components in gas product

231 HHV_i = higher heating value of each component in $MJ kg^{-1}$

232

233 2.3.5 Heating value of liquid and solid products

234 The standard bomb calorimeter method was used to determine the higher heating value for
235 the primary condensate and solid products (ASTM, 2000). Results obtained from the bomb
236 calorimeter experiment indicate the higher heating value (HHV) of the samples. In the
237 procedure, approximately 1 g of solvent-free primary condensate was burnt completely at
238 constant volume, in an excess oxygen environment in a steel vessel, called a bomb using a
239 Parr 6100 calorimeter. Each experiment was repeated three times. The same procedure was
240 repeated for the solid residues. In addition, the HHV of the solids were calculated using the
241 generalised formula reported by Channiwala and Parikh (2002), according to equation 1;

242

243 $HHV = 0.3491C + 1.1783H + 0.1005S + 0.1034O - 0.0151N - 0.0211A \dots\dots\dots (1)$

244 Where; C = carbon wt%, H = hydrogen wt%, S = sulphur wt%, O = oxygen wt%, N = nitrogen
245 wt% and A = wt% ash

246

247 **2.4 Mass balance calculation**

248 Mass balances (wt% on dry feed basis) were calculated based on mass of dry trommel fines
249 processed and final fast pyrolysis products of liquid, solid residues and non-condensable
250 gases. The solid residues were distributed into the bed material, char pot and liquid products.
251 The bed material was weighed before and after each experiment to determine the weight of
252 solid residues that remained. The weight of ash in the char pot was determined after char
253 burn-off, while ash contents in the liquid products was determined via filtration and weighing.
254 All metal, glassware items and transition pipes used in the bench-scale pyrolysis unit were
255 weighed before and after each run, to calculate the yields of pyrolysis products. The amount
256 of trommel fines feedstock fed into the pyrolysis reactor was calculated from the difference

257 between the weight of trommel fines added to hopper before and the trommel fines left in
258 hopper after each experiment.

259

260 **2.5 Fast pyrolysis energy recovery efficiency**

261 To determine whether trommel fines samples can be used for energy recovery via fast
262 pyrolysis, with significant reduction in mass and volume of the original waste, the fast pyrolysis
263 energy recovery efficiency (η) was calculated based on the ratio of the energy content in the
264 conversion products to that in the feedstock. In this case, the products of trommel fines fast
265 pyrolysis conversion were the liquid and gas products as well as the char product obtained in
266 the char pot. Therefore, energy recovery efficiency was obtained by using the higher heating
267 values (HHV) and the mass of each component (m) as follows:

268

$$269 \eta (\%) = \frac{(m \text{ HHV})_{\text{liquids}} + (m \text{ HHV})_{\text{gas products}} + (m \text{ HHV})_{\text{solid products}}}{(m \text{ HHV})_{\text{feedstock}}} \times 100 \dots \dots \dots (2)$$

270

271 **3 Results and Discussions**

272 **3.1 Effect of pyrolysis temperature on overall mass balance and product yields**

273 Table 3 details an overview of the process parameters, product yields, product distributions
274 and mass balance closure results obtained from the experiments. The resulting mass
275 balances were compared to establish similarities between results. Reproducibility was more
276 than 95%, so the highest mass balance closure obtained were reported and the rest are
277 presented as standard deviations (Kalgo, 2011; Bank, Nowakowski and Bridgwater, 2014).

278 As can be seen from Table 3, the mass balance closures were above 97% for TIR-400 and
279 TIR-500, which showed acceptable accounting for the mass flows. However, the mass
280 balance closure decreased with increasing temperature, with TIR-700 having the lowest mass
281 balance closure. The relatively poor balances at higher temperatures was be largely due to
282 the presence of unidentified gas components, formed from severe thermal degradation, whose
283 peaks were observed in the GC outlines but not quantified.

284

285 Two sets of data for product gases have been included in Figure 2. One set of data represent
286 the actual yields obtained from GC analysis, while the other set of data was obtained by
287 difference after considering the yields of liquids and solids. Figure 2 therefore shows that there
288 was a good agreement between two sets of data at the lower temperatures, with significant
289 differences at higher temperatures. In fast pyrolysis, the liquid product is generally regarded
290 as the main product, while solid products and permanent gases may be valuable by-products.
291 Typically, liquid products contain some amount of water (Czernik and Bridgwater, 2004). This
292 water is derived from the moisture in the feedstock and the water produced during fast

293 pyrolysis reaction, which is called “reaction water”. In Table 3, only the reaction water is
294 reported, and the non-aqueous part of bio-oil is then referred to as “organics”, which is the
295 desirable product. This fraction is reported without the solids (ash-free). Results showed that
296 pyrolysis temperatures have a dominant effect on the liquid product yield.

297

298 In general, the yields of total liquid increased with increasing pyrolysis temperature from 400
299 °C to 500 °C and then decreased at higher temperatures (Figure 2). The highest liquid yield
300 was observed at TIR-500 (32.6 wt% dry basis) with a primary condensate organic yield of 12.5
301 wt% (daf) and secondary condensate organic yield of 7.00 wt% (daf). When compared to
302 literature on pyrolysis of RDF, due to the similarity in composition with the trommel fines
303 feedstock, the organic yield obtained here was smaller (Mohan et. al 2006; Chen et al., 2015;
304 Ding et al., 2016). In addition, the reaction water yields were found to decrease with increasing
305 temperature, indicating the possible consumption of the produced water in the chemical
306 reactions occurring during fast pyrolysis at high temperatures (Chen et al., 2015; Ding et al.,
307 2016).

308

309 The solid residues obtained from the fast pyrolysis of trommel fines ranged of 48 – 65 wt% on
310 dry basis (Figure 2). This result was consistent with the high ash content (36.2 wt%, dry basis)
311 of the feedstock. In general, it can be observed from Figure 2, that increasing temperature led
312 to a slight decrease in solid residues, while gas yields increased. Hence, it was considered to
313 use 700 °C as the maximum pyrolysis temperature, as higher temperatures would lead to
314 more gasification. Such effect has been reported in previous studies, showing increased gas
315 formation at the expense of liquid and solid products during high-temperature pyrolysis of
316 different organic feedstock, using different reactor configurations (Luo et al., 2004; Lee et al.,
317 2005; Tsai et al., 2007). The increase of gas yields with increasing temperature is possibly
318 due to a combination of secondary thermal cracking of the evolved pyrolysis vapours and the
319 catalytic effects of the ash via vapour-solid reactions at elevated temperatures (Lee et al.,
320 2005; Pradhan, 2015; Renkonen, 2016). Some researchers have shown that the presence of
321 certain types of ash (e.g. alkali metals and alkaline-earth metals) can promote the formation
322 of solids, reaction water and gas products at the expense of pyrolysis liquid yield (Mohan et
323 al., 2006; Buah et al., 2007; Zhou et al., 2013; Chen et al., 2014; Pradhan, 2015; Xue, Braden
324 and Bai, 2017)

325

326 As seen in Table 3, the gaseous products contain mainly of carbon dioxide and propylene with
327 small quantities of C₁ – C₄ hydrocarbon gases. Most of the gases identified, increased in their
328 mass yields with increasing temperature. Moreover, it can be seen from Figure 3 that
329 increasing pyrolysis temperature led to a decrease in CO₂ proportion in the gas product (Figure

330 3A) as other gases became prominent. This could be mainly because most of the CO₂ was
331 already generated by carboxyl release at a relatively low temperature (Luo et al., 2004). In
332 addition, the secondary thermal cracking of pyrolysis vapours produced CO and hydrocarbon
333 gases rather than CO₂.

334

335

336 **3.2 Effect of pyrolysis temperature on liquid product characteristics**

337 The water content of the liquid products from pyrolysis has remained one of the factors
338 affecting their quality and subsequent use. Certainly, the higher the reaction water yields, the
339 higher the water content in the liquid products. The presence of water in the liquid products
340 can be disadvantageous and advantageous. It reduces the heating value, especially the LHV
341 and flame temperature and on the other hand, it improves the flow characteristics of pyrolysis
342 liquid by reducing the viscosity (Czernik and Bridgwater, 2004).

343

344 The water contents of liquid products produced from trommel fines at different pyrolysis
345 temperature are summarised in Table 4. It can be seen from Table 4 that operating
346 temperatures has significant influence on the water contents of liquid products; they seemed
347 to be higher at lower temperature especially in the secondary condensates, as more water
348 condensed in the condensers operated at low temperatures. This agrees with a study by Lee
349 et al., (2005), where fast pyrolysis of rice straw at different temperatures was investigated.
350 The total water contents of trommel fines liquid products (primary and secondary condensate)
351 were in the range of 50 – 78 wt%, which is higher than those of typical wood bio-oils and may
352 need to be reduced prior to use. A promising route to lowering the amount of water in bio-oils
353 has been proposed by Oasmaa et al. (2005), which involves the removal of water together
354 with light reactive volatiles by increasing the condenser temperature to 50 ± 4 °C followed by
355 the addition of alcohol (usually isopropanol) to improve the viscosity and stability of the bio-
356 oils.

357

358 The solids contents entrained in the liquid products produced at different pyrolysis
359 temperatures are also presented in Table 4. The solid content in the trommel fines liquid
360 products increased with increasing temperature. They fluctuated in the range of 0 – 16.2 wt%,
361 which was a higher range than those reported in literature (Miskolczi et al., 2013). Solids in
362 bio-oils are usually composed of char fines entrained out of the cyclone and the secondary
363 char formed by secondary reactions of pyrolysis vapours such as re-polymerization and
364 condensation. In addition to these, the oils from trommel fines also contained inorganic
365 materials (sand and glass fines) from the ash. The solids could escape the cyclone due to
366 their very small particles (less than about 10 µm in diameter) and the high gas stream velocity

367 in the cyclone. The main challenge was that once the solids got into the liquid products, it was
368 difficult to remove by liquid filtration because of the highly viscous nature of the primary
369 condensate.

370

371 Literature has shown that, by introducing a hot vapour filter prior to condensation units, the
372 amounts of solids in liquid products can substantially be reduced (Chen et al., 2011; Pattiya
373 and Suttibak, 2017). The solids present in in the liquid product can cause erosion and
374 blockages to equipment such as nozzles, valves and pumps, (Czernik and Bridgwater, 2004).
375 Additionally, the inorganic compounds present in the solids are important to bio-oils ageing
376 characteristics as they appear to catalyse polymerisation reactions during storage, leading to
377 increased viscosity and growth in the apparent diameter of the suspended char (Diebold,
378 2000). Accordingly, liquids with lower solids content are generally preferred. However, the
379 solids content of liquid product may not be an important issue, if the product is a mixture of
380 bio-oil and solid char in the form of slurry. This slurry which can contains up to 90% of energy
381 from the feedstock could be used as a feedstock for gasification process to produce syngas
382 or producer gas. In fact, the slurry is commercially known as Bio-Oil Plus™ from Dynamotive
383 Energy Systems Corporation, which can be used as fuel in boilers.

384

385 The elemental compositions of both the primary and secondary condensate liquids are also
386 listed in Table 4. The percentages of carbon for both primary and secondary condensate
387 increased with increasing temperature until reaching a maximum. With regards to nitrogen
388 contents, increase in temperature led to an increase in nitrogen content for both the primary
389 and secondary condensates and are higher than typical wood bio-oil, which resulted from the
390 high nitrogen content in the trommel fines feedstock. The presence of textile materials such
391 as those made from polyamides may have significantly increased the nitrogen contents of the
392 trommel fines; which upon thermal degradation would form nitrogen-containing volatile
393 compounds, which condensed into liquids after cooling in the condensers.

394

395 Moreover, there was more nitrogen in the secondary condensates than in the primary
396 condensates, possibly due to the high volatility of the nitrogen-containing degradation
397 products, which were trapped in the colder condenser. The presence of nitrogen compounds
398 can be a drawback when burning the liquids because of the high potential for NO_x emissions.
399 This problem may be prevented by pre-treatment such as the washing of the feedstock with
400 distilled water prior to pyrolysis (Lee et al., 2005). The pre-treatment has shown to also remove
401 some alkali metals such as sodium and potassium that are known to have adverse catalytic
402 effect during fast pyrolysis reactions (Lee et al., 2005).

403

404 The heating values of the primary and secondary condensate liquids are also shown in Table
405 4. It was found that the heating values increased with increasing temperature until reaching a
406 maximum then further increase led to slight decrease for both the primary and secondary
407 condensate with TIR-500 liquids having the highest values for both the primary and secondary
408 condensate 32 MJ kg^{-1} and 17.45 MJ kg^{-1} (dry basis) respectively. The HHV of the primary
409 condensate was found to be higher when compared to literature on pyrolysis of biomass
410 (Diebold, 2000; Czernik and Bridgwater, 2004; Lee et al. 2005) and has potential for energy
411 recovery. The fractional plastic content of the trommel fines feedstock may be responsible to
412 the high heating value.

413

414 **3.3 Effect of pyrolysis temperature on composition of organic liquid products**

415 The organic liquid products produced from pyrolysis temperature investigation are composed
416 of different classes of organic compounds, which are derived primarily from the de-
417 polymerization and fragmentation reactions of the organic components of the original
418 feedstock, namely the biogenic and plastic fractions in the trommel fines (Mohan et al., 2006).
419 The chromatograms of the primary and secondary condensate liquids from TIR-500
420 experimental run are shown in Figure 4, indicating the different profiles of compounds present
421 in both fractions. These chromatograms show that majority of the compounds in the primary
422 condensate had longer retention times in the GC columns compared to compounds in the
423 secondary condensates. Therefore, the secondary condensates contained lighter, more
424 volatile organic compounds than the primary condensates as shown later in Figure 5.

425

426 The major organic compounds have been classified into groups including, oxygenates,
427 carboxylic acids, amines, phenols, aromatic hydrocarbons and others. Oxygenates included
428 compounds such as ketones and aldehydes e.g. cyclopentanones, cyclopentenones, furfural
429 and their alkylated derivatives, from pyrolysis of the sugar fractions (cellulose and
430 hemicellulose) of biomass. Aromatic hydrocarbons included styrene, azulene, naphthalene
431 and bibenzyls; these compounds have been reported as products of polystyrene pyrolysis
432 (Muhammad, Onwudili and Williams, 2015). Phenols included phenol, alkyl phenols and
433 methoxyphenols such as methylphenols and dimethylphenols, which are common compounds
434 from lignin pyrolysis (Zhou et al., 2015). Several amines N-containing compounds were
435 detected in the liquid products including 1-propenamine, pyridine and methyl pyridine, which
436 could be from the pyrolysis of textile materials made from polyamides.

437

438 Figure 5 presents the relative compositions, in terms of peak area percent, of the primary and
439 secondary condensates in relation to pyrolysis temperatures. The compounds with the largest
440 peak areas, up to about 60% of compounds in each sample, were selected for detailed

441 evaluation. The primary condensates consisted of heavier compounds including phenols and
442 2-ringed aromatic hydrocarbons, while the secondary condensates contained a number of
443 lighter compounds such as the ketones and aldehydes (oxygenates), amines and lighter
444 aromatic hydrocarbons (e.g. propynyl benzene and azulene). Increase in temperature
445 increased the presence of phenolic and aromatic hydrocarbon compounds in the primary
446 condensate (and Figure 5A). The increased formation of these compounds is known to be
447 promoted during high temperature pyrolysis of biomass through secondary reactions such as
448 radical combinations and cyclization (Diels-Alder reaction), aromatisation, deoxygenation,
449 demethoxylation and demethylation (Ashmadi et al., 2011; Ledesma et al., 2013; Zhou et al.,
450 2015).

451
452 On the contrary, in the secondary condensate (Figure 5B), the peak area percent of lighter
453 compounds decreased with increasing temperature. These trends agree with trends seen in
454 literature as increase in temperature would have led to more cracking and increases in gas
455 yields as seen on Table 3. The secondary condensate liquid fraction collected from the dry ice
456 condensers showed the presence of nitrogen-containing organic compounds (Figure 5B).
457 These compounds were in the form of amines, pyridines and their derivatives. This finding is
458 in line with the result of the elemental analysis of the secondary condensate liquid which shows
459 an increase in nitrogen content with increasing temperature (Table 4).

460

461 **3.4 Effect of pyrolysis temperature on solid residues characteristics**

462 The solid residues obtained from this work, expectedly contained a mixture of ash (inorganic
463 material) and char. Ash content, elemental composition, and calorific value of the solid
464 products obtained from fast pyrolysis of trommel fines are tabulated in Table 5. The
465 degradation of the organic fraction of the feedstock via fast pyrolysis process would leave the
466 ash in the solid product (Silva et al., 2015; Rotheut and Quicker, 2017; Zaini et al., 2019). The
467 ash content in the solid residues ranged from 62 - 87 wt% (dry basis), with TIR-500 having the
468 highest ash content of 86.3 wt%. Increase in temperature above 500 °C lead to a slight
469 decrease in the ash content (Table 5) and this is possibly due to the secondary decomposition
470 of solid residues (such as carbonates) at elevated temperatures (Scott et al., 1988), which
471 lead to an increase in gas yield (Figure 2). At lower temperatures, there was more char in the
472 solid residues (Table 5).

473

474 The elemental compositions and heating value of solid products are also listed in Table 5. The
475 percentage of carbon did not show a clear trend but mirrored the proportion of char in the solid
476 residues at different pyrolysis temperatures. The heating values of the solid products
477 corresponded to the carbon contents. The HHV were in the range of 4 - 8 MJ kg⁻¹ dry basis,

478 giving a maximum value at the lower temperature. The low calorific value of the solid product
479 would not make it a viable source of process heat but could be combusted in cement kilns and
480 the ash incorporated in building materials (if non-hazardous).

481

482 A scan of the ash in the feedstock was carried out using a Bruker S8 Tiger X-ray Fluorescence
483 (XRF) spectrometer, capable of quantifying elements from sodium to uranium. The major
484 elements in the component with concentrations > 0.1 wt% included calcium, silicon, lead,
485 chlorine, iron, potassium, aluminium, copper, sulphur, manganese, niobium, phosphorus, zinc,
486 magnesium and titanium (Onwudili and Eke, 2019). Hence apart from chlorine, these
487 elements (mainly as their oxides) would remain in the solid residues, which may be burnt off
488 at cement kilns, milled and in combination with clay, lime and cement mortar can be used as
489 an additive for plaster or for bricks and concrete elements at a ratio of up to 80%. Such
490 blending has been reported to create inside walls with excellent insulation and breathing
491 properties, able to maintain humidity levels in a room at 45 – 70% (Schmidt, 2014).

492

493 Alternatively, the solid residues may be used in suitable biochar applications such as in soil
494 amendments as fertilizer and for carbon sequestration. When used as a soil amendment,
495 biochar has been reported to boost soil fertility and improve soil quality by raising soil pH,
496 increasing moisture holding capacity, attracting more beneficial fungi and microbes, improving
497 cation exchange capacity (CEC), and retaining nutrients in soil (Lehmann, Gaunt and Rondon,
498 2006; Lehmann, 2007; Winsley, 2007; Gaunt and Lehmann, 2007; Laird, 2008). Biochar may
499 persist in soil for millennia because it is very resistant to microbial decomposition and
500 mineralisation.

501

502 **3.5 Energy recovery from pre-treated trommel fines**

503 Figure 6 shows the effect of temperature on the fast pyrolysis process conversion efficiency
504 of dry physically pre-treated trommel fines samples calculated using the calorific values of the
505 liquid, solid (char pot) and gas products. The char content left in the bed material was not
506 considered as it would not be readily available for immediate use. Therefore, Figure 6 reflect
507 the energy yields based on the liquid, char pot residues and gas products from the fast
508 pyrolysis of trommel fines. The energy recovery efficiency ranged between 63 – 70%,
509 generally increasing with increasing pyrolysis temperature from 400 °C to 700 °C, except for
510 a dip at 600 °C. Figure 6 shows that the liquid product accounted for over 56% of the recovered
511 energy from the trommel fines feedstock at 500 °C. Whereas, at higher temperatures, the
512 contributions of gas products to the recovered energy became more dominant, such that at
513 700 °C, the gas products contributed 50% to the energy yields.

514

515 From these results, it is clear that running the fast pyrolysis tests at temperatures around 500
516 °C, gave the highest liquid products. Literature shows that at around 500 °C, fast pyrolysis is
517 used to obtain the highest liquid yields from various organic feedstocks including biomass and
518 plastics. Therefore, trommel fines may be co-processed with other feedstock without major
519 changes to fast pyrolysis process conditions. Indeed, co-processing of trommel fines with
520 biomass would reduce the operational impact of its high ash content and possibly give a liquid
521 product with better quality, especially in terms of calorific values.

522

523 **4. Conclusion**

524 Energy recovery from a physically pre-treated trommel fines feedstock has been investigated
525 in a bubbling fluidised bed reactor in relation to pyrolysis temperature. The optimum reaction
526 temperature to obtain the highest pyrolysis liquid was 500 °C (33 wt % dry basis). It was found
527 that trommel fines produced less pyrolysis liquids (approximately 13 - 33 wt%, dry basis)
528 compared to woody biomass (50 - 75 wt%), mainly due to the high ash content (36 wt%) of
529 the former. Similar to other pyrolysis studies, reaction temperature had a dominant effect on
530 the fast pyrolysis product yields. However, the dramatic increase in gas yields above 500 °C,
531 could attributed to synergistic catalytic cracking effect of the high ash content. The trend in the
532 yields of pyrolysis products mirrored the calculated energy recovery efficiency profile; showing
533 that non-volatilised solids, liquid products and gas products dominated the recovered energy
534 at 400 °C, 500 °C and 700 °C, respectively. The high ash content of the feedstock (36 wt %, dry
535 basis) resulted in a large proportion of solid residues in the char pot, which was found to
536 contain mostly non-hazardous earth materials and could be used by the construction industry.
537 Subsequent accumulation of large quantities of solid residues in the char pot may cause safety
538 and operability problems during long periods of rig operation. Possible solutions would involve
539 redesigning the ring to allow for intermittent/periodic removal of accumulated solids and/or the
540 use of a char pot with larger volume. Overall, the results of this work showed that the BFB fast
541 pyrolysis reactor could handle the high ash content after size reduction of the trommel fines
542 feedstock, while producing liquid and gas products as valuable recovered energy carriers.

543

544 **Acknowledgements**

545 The authors are grateful to the Energy and Bioproducts Research Institute at Aston University
546 for their continuous support.

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552 **References**

553 Asmadi, M.; Kawamoto, H.; Saka, S., 2011. Gas- and solid/liquid-phase reactions during
554 pyrolysis of softwood and hardwood lignins. *J. Anal. Appl. Pyrolysis*, 92, 417–425.

555

556 ASTM, 2000. ASTM D2015 Standard Test Method for Gross Calorific Value of Coal and Coke
557 by the Adiabatic Bomb Calorimeter. American Society for Testing and Materials International,
558 United States

559

560 ASTM D1762 – 84, 2013. Standard Test Method for Chemical Analysis of Wood Charcoal.
561 American Society for Testing and Materials International, United States

562

563 Ates, F., Miskolczi, N., Borsodi, N., 2013. Comparison of real waste (MSW and MPW)
564 pyrolysis in batch reactor over different catalysts. Part I: product yields, gas and pyrolysis oil
565 properties. *Biores. Technol.* 133, 443–454.

566

567 Banks, S. W., Nowakowski, D.J., Bridgwater, A.V., 2014. Fast pyrolysis processing of
568 surfactant washed Miscanthus. *Fuel Proc. Technol.* 128, 94-103.

569

570 Bosmans, A., Vanderreydt, I., Geysen, D., Helsen, L., 2013. The crucial role of waste-to-
571 energy technologies in enhanced landfill mining: a technology review. *J. Cleaner Prod.* 55, 10-
572 23.

573

574 Channiwala, S.A., Parikh, J., 2002. A unified correlation for estimating HHV of solid, liquid and
575 gaseous fuels. *Fuel* 81, 1051-1063.

576

577 Chen, S., Meng, A., Long, Y., Zhou, H., Li, Q. and Zhang, Y., 2015. TGA pyrolysis and
578 gasification of combustible municipal solid waste. *J. Energy Institute* 88(3), 332-343.

579

580 Chen, T., Wu, C., Liu, R., Fei, W. and Liu, S., 2011. Effect of hot vapour filtration on the
581 characterization of bio-oil from rice husks with fast pyrolysis in a fluidized-bed reactor. *Biores.*
582 *Technol.* 102(10), 6178-6185.

583

584 Czernik, S., Bridgwater, A.V., 2004. Overview of applications of biomass fast pyrolysis oil.
585 *Energy Fuels* 18, 590–598.

586

587 Diebold, J.P., 2000. A Review of the Chemical and Physical Mechanisms of the Storage
588 Stability of Fast Pyrolysis Bio-Oils. National Renewable Energy Laboratory (NREL).
589 NREL/SR-570-27613.
590

591 Ding, K., Zhong, Z., Zhong, D., Zhang, B., Qian, X., 2016. Pyrolysis of municipal solid waste
592 in a fluidized bed for producing valuable pyrolytic oils, *Clean Technol. Environ. Policy* 18,
593 1111-1121.
594

595 Eke, J., Onwudili, J. A. and Bridgwater, A.V., 2017. Physical Pretreatment of biogenic-rich
596 trommel fines for fast pyrolysis. *Waste Management*. 70, 81-90.
597

598 Fitzgerald, G.C., 2013. Pre-processing and treatment of municipal solid waste (MSW) prior to
599 incineration. In. Klinghoffer, N.B. and Castaldi, M.J. (eds). *Waste to Energy Conversion*
600 *Technology*, 55-71. Woodhead Publishing Limited, Cambridge
601

602 Gaunt, J., Lehmann, J., 2008. Energy balance and emissions associated with biochar
603 sequestration and pyrolysis bioenergy production. *Environ. Sci. Technol.* 42, 4152- 4158.
604

605 Grammelis, P., Basinas, P., Malliopoulou, A., Sakellariopoulos, G., 2009. Pyrolysis kinetics
606 and combustion characteristics of waste recovered fuels. *Fuel* 88, 195-205.
607

608 Grieco, E.M., Baldi, G., 2012. Pyrolysis of polyethylene mixed with paper and wood: interaction
609 effects on tar, char and gas yields, *Waste Management* 32, 833– 839.
610

611 HMRC, 2014. Landfill Tax: Liability of waste ‘fines’. Her Majesty Revenue & Customs.
612 Webpage:Online:[https://www.gov.uk/government/uploads/system/uploads/attachment_data/f](https://www.gov.uk/government/uploads/system/uploads/attachment_data/file/325100/Landfill_tax_condoc.pdf)
613 [ile/325100/Landfill_tax_condoc.pdf](https://www.gov.uk/government/uploads/system/uploads/attachment_data/file/325100/Landfill_tax_condoc.pdf). [Accessed 21 May 2019].
614

615 Kalgo, A. S., 2011. The Development and Optimisation of a Fast Pyrolysis Process for Bio-oil
616 Production, Aston University, PhD Thesis.
617

618 Laird, D. A., 2008. The charcoal vision: A win-win-win scenario for simultaneously producing
619 bioenergy, permanently sequestering carbon, while improving soil and water quality. *Agron.*
620 *J.* 100, 178-181.
621

622 Ledesma, E.B., Campos, C., Cranmer, D.J., Foytik, B.L., Ton, M.N., Dixon, E.A., Chirino, C.,
623 Batamo, S., Roy, P., 2013. Vapor-phase cracking of eugenol: distribution of tar products as
624 function of temperature and residence time. *Energy and Fuels* 27: 868–878
625

626 Lee, K.H., Kang, B.S., Park, Y.K., Kim, J.S., 2005. Influence of reaction temperature, pre-
627 treatment, and a char removal system on the production of bio-oil from rice straw by fast
628 pyrolysis, using a fluidized bed. *Energy Fuels* 19, 2179-2184.
629

630 Lehmann, J., 2007. Bio-energy in the black. *Front. Ecol. Environ.* 5, 381-387.
631

632 Lehmann, J., Gaunt, J., Rondon, M., 2006. Bio-char sequestration in terrestrial ecosystems:
633 a review. *Mit. Adapt. Strat. Global Chang* 11, 403-427.
634

635 Luo, Z., Wang, S., Liao, Y., Zhou, J., Gu, Y., Cen, K., 2004. Research on biomass fast
636 pyrolysis for liquid fuel. *Biomass and Bioenergy* 26, 455-462.
637

638 Malkow, T., 2004. Novel and innovative pyrolysis and gasification technologies for energy
639 efficient and environmentally sound MSW disposal. *Waste Management.* 24, 53-79.
640

641 McKendry P., 2002. Energy production from biomass (Part 2): conversion technologies.
642 *Biores. Technol.* 8, 37-54
643

644 Miskolczi, N., Ates, F., Borsodi, N., 2013. Comparison of real waste (MSW and MPW)
645 pyrolysis in batch reactor over different catalysts. Part II: contaminants, char and pyrolysis oil
646 properties. *Biores. Technol.* 144, 370-379.
647

648 Mohan, D., Pittman, C.U., Steele, P.H., 2006. Pyrolysis of wood/biomass for bio-oil: a critical
649 review. *Energy and Fuels* 20, 848-889.
650

651 Muhammad, C., Onwudili, J.A., Williams, P.T., 2015. Thermal degradation of real-world waste
652 plastics and simulated mixed plastics in a two-stage pyrolysis-catalysis reactor for fuel
653 production. *Energy and Fuels* 29, 2601-2609,
654

655 Oasmaa, A., Peacocke, C., 2001. A guide to physical property characterisation of biomass-
656 derived fast pyrolysis liquids. Technical Research Centre of Finland, Espoo.
657

658 Oasmaa, A., Sipila, K., Solantausta, Y., Kuoppala, E., 2005. Quality improvement of pyrolysis
659 liquid: effect of light volatiles on the stability of pyrolysis liquids. *Energy and Fuels* 19, 2556-
660 2561.

661

662 Onwudili, J. A. and Eke, J., 2019. Effects of ash removal by agitated aqueous washing and
663 sedimentation on the physico-chemical characteristics and fast pyrolysis of trommel fines. *J.*
664 *Energy Institute* 93(1), 312 – 323.

665

666 Pattiya, A. and Suttibak, S. (2017). Fast pyrolysis of sugarcane residues in a fluidised bed
667 reactor with a hot vapour filter. *J. Energy Institute* 90(1), 110-119.

668

669 Pitchell, J., 2014. Municipal solid waste processing: material recovery facilities. *Waste*
670 *Management Practices: Municipal, Hazardous and Industrial*. 2nd Edition, 165-196. CRC
671 Press, London.

672

673 Pradhan, U., 2015. Physical treatments for reducing biomass ash and effect of ash content
674 on pyrolysis products. MSc Dissertation, Auburn University, Alabama, USA. Available online:
675 [https://etd.auburn.edu/bitstream/handle/10415/4726/Master%20thesis_Ujjain.pdf?sequence](https://etd.auburn.edu/bitstream/handle/10415/4726/Master%20thesis_Ujjain.pdf?sequence=2&isAllowed=y)
676 [=2&isAllowed=y](https://etd.auburn.edu/bitstream/handle/10415/4726/Master%20thesis_Ujjain.pdf?sequence=2&isAllowed=y) . [Accessed May 1, 2019].

677

678 Renkonen, S., 2016. Removal of alkali metals for the enhancement of fast pyrolysis process.
679 Master's Thesis. Aalto University, Espoo, Finland. Available online:
680 [https://cris.vtt.fi/en/publications/removal-of-alkali-metals-for-the-enhancement-of-fast-](https://cris.vtt.fi/en/publications/removal-of-alkali-metals-for-the-enhancement-of-fast-pyrolysis-pr)
681 [pyrolysis-pr](https://cris.vtt.fi/en/publications/removal-of-alkali-metals-for-the-enhancement-of-fast-pyrolysis-pr). [Accessed: May 16, 2019].

682

683 Rotheut, M., Quicker, P., 2017. Energetic utilisation of refuse derived fuels from landfill mining.
684 *Waste Management*, 62, 101–117

685

686 Schaefer, W.D., 1975. Disposing of solid wastes by pyrolysis. *Environ. Sci. Technol.* 9, 98-98.

687

688 Schmidt, H. P., 2014. The use of biochar as building material, *Biochar Journal*, Arbaz,
689 Switzerland. ISSN 2297-1114 www.biochar-journal.org/en/ct/3. Accessed: 04.03.2019

690

691 Scott, D.S., Piskorz, J., Bergougnou, M.A., Graham, R., Overend, R.P., 1988. The role of
692 temperature in the fast pyrolysis of cellulose and wood. *Ind. Eng. Chem. Res.* 27, 8-15.

693

694 Silva, R.B., Martins-Dias, S., Arnal, C., Alzueta, M.U., Costa, M., 2015. Pyrolysis and char
695 characterization of refuse-derived fuel components. *Energy and Fuels*, 29,1997–2005.
696

697 Sørum, L., Gronli, M.G., Hustad, J.E., 2001. Pyrolysis characteristics and kinetics of municipal
698 solid wastes. *Fuel* 80, 1217-1227.
699

700 Tsai, W.T., Lee, M.K., Chang, Y.M., 2007. Fast pyrolysis of rice husk: product yields and
701 compositions. *Biores. Technol.* 98, 22-28
702

703 Winsley, P., 2007. Biochar and bioenergy production for climate change mitigation. *New*
704 *Zealand Sci. Review* 64, 5-10.
705

706 Zaini, I. N., López, C.G., Pretz, T., Yang, W., Jönsson, P.G., 2019. Characterization of
707 pyrolysis products of high-ash excavated-waste and its char gasification reactivity and kinetics
708 under a steam atmosphere. *Waste Management*, 97, 149-163.
709

710 Zhao, L., Chen, D.Z., Wang, Z.H., Ma, X.B., Zhou, G., 2011. Pyrolysis of waste plastics and
711 whole combustible components separated from municipal solid wastes: Comparison of
712 products and emissions. In: *Proceedings of the Thirteen International Waste Management and*
713 *Landfill Symposium*, 3–6 October 2011, Sardinia, Italy.
714

715 Zhou, H., Wu, C., Onwudili, J.A., Meng, A., Zhang, Y., Williams, P.T., 2015. Polycyclic
716 aromatic hydrocarbons (PAH) formation from the pyrolysis of different municipal solid waste
717 fractions. *Waste Management* 36, 136-146.
718