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Clocks in the wild: biological rhythms of great tits and the environment



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BSc.

Submitted in fulfilment of the requirements of the Degree of Doctor of Philosophy
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May 2020

Abstract

Biological clocks play a fundamental role in the physiological and behavioural processes of organisms. Internal timekeepers evolved to anticipate environmental changes, the most important of these being the geophysical light-dark cycle, to coordinate external changes with the timing of internal processes. Research into functions of the biological clock during captive studies has provided valuable insight into mechanisms by which clocks function, and how small environmental changes can affect the clock and its outputs. However, biological clocks have so far been understudied in ecology.

In this thesis, this gap in knowledge was addressed by placing studies of chronobiology into the context of the natural environment. A model species in avian ecology, the great tit (*Parus major*) was used to investigate biological rhythms in the wild at three levels; behaviour, transcripts and life histories. This thesis investigated how features of the natural environment shapes rhythms of behaviour and physiology in a wild animal, using experimental and observational approaches.

Differences in timing of individual rhythms, or chronotype, may provide wild animals with different consequences for fitness. In this thesis, individual behavioural rhythms of incubating great tits were quantified for birds in city and forest environments. There were strong effects of both the number of days to hatching and site on timing of incubation activities, where city birds rose earlier, and stayed out later, than forest birds. Maternal chronotype was then linked to fitness traits.

City birds face a number of new challenges in the urban habitat. The impacts of one feature of the urban habitat, artificial light at night, was tested using a forest nest box system. Nestling great tits were experimentally exposed to low-level artificial light at night, and aspects of condition and clock and immune gene transcripts were compared for nestlings under light at night and dark-night control. Nestlings under light at night treatment weighed less than control birds, and suppressive effects of light at night treatment were found for genes involved in the core pathways of the circadian clock and immune system. Time of day differences were also observed in transcript levels of genes.

Parasitic infections can cause consequences for fitness and reproductive success of wild birds. In this study, effects of infection with avian malaria parasites on nestling condition and immune system were investigated, at city and forest sites. The prevalence of *Leucocytozoon* parasites was higher at forest sites than city sites and increased with the

season. Infection had no suppressive effects on immune genes of nestlings, and no negative effects on condition were found.

In mammals, malaria is otherwise known as the "circadian disease" due to rhythmic development of parasites during their life cycle. In this study, host-parasite interactions with avian malaria parasites were investigated in the context of biological rhythms in wild great tits. Transcript levels of nestlings were determined by field sampling across a temporal profile and linked to infections with *Leucocytozoon* parasites. *Leucocytozoon* infection reduced overall transcript levels for circadian clock and immune gene targets, but did not alter the timing of expression.

This study ultimately demonstrated the importance of biological clocks for the ecology of great tits and provided important advances for studies of clocks in the wild.

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Acknowledgements

I am indebted to many people for their help during this project, whom I must thank. Firstly, I thank my supervisors Barbara Helm, Jane Robinson, Francesco Baldini and Peter O'Shaughnessy for their endless support and advice from the beginning to the end of the project, and for all the great lunches. My advisors, Michelle Bellingham and Dorothy McKeegan for their advice and input. I also thank Simon Babayan, Barbara Mable, Kathryn Elmer, Dan Haydon, Davide Dominoni and Luke Powell for all help during the project.

Secondly, I thank Christian Delles, Karine Pinel, Lesley Graham for their help in the use of the QuantiGene® Plex assay, and the Luminex® MAGPIX machine. Natascha Brinskelle-Schmal from ThermoFisher, and Margaret Mills, for their patience in giving us technical support during optimisation of the QuantiGene® Plex assay!

I thank the amazing technicians in IBAHCM: Kate Griffiths, Elizabeth Kilbride, Winnie Boner and Ana Monteiro, for patiently answering all my questions and for help with processing samples in the lab. I thank Nosrat Mirzai and the MVLS Bioelectronics team for building LED nest boxes used in this project, and Rona Brennan and Colin Adams for welcoming me at SCENE. I thank honours student Gillian O'Flaherty for the molecular sexing of the birds, and Barbara Mable for supervising this work. Pablo Capilla-Lasheras for developing the incR package used in this thesis, and MRes student Ciara McGlade for initial analyses of the iButton data.

I would like to thank Stewart "Stinky" White for the banter, early morning ringing sessions, and for taking me to hospital to have my "stookie" removed! Paul Jerem, and Ben the dog, for carrying out fieldwork while I was on crutches. Crinan Jarrett, Darren Talbot, Rachel Steenson, and all of the invaluable field volunteers over the years, for their hard work in the field tackling the rain, ticks, and biting midges of wild Scotland. The UofG Market Intelligence team: Elizabeth, Bo, James, Allison and Campbell, for helping me right through to the end. Lucy Cotgrove, Georgia Kirby, Matt Pace, Julie Miller, Hans Recknagel, and Megan Wright and Rosie Saxcoburg (my odd combination of friends), for their support and for keeping me grounded. Bedur Albalawi, my PhD sister (we did it!) for keeping me inspired. Allan MacRitchie, for his patience. Finlay Hind, for encouraging me to go for it. Finally, I thank Mum, Dad, Jamie, Shannon and Cody for their endless support and patience.

Author Declaration

I declare that the work presented in this thesis is my own unless otherwise stated, and that			
no part of this thesis has been submitted for any other degree.			
Signature			
Printed name			

Contributions

The data in this thesis span three years and field seasons from 2016-2018. The following section provides a list of chapters in this thesis and the various contributions made by coauthors, students and collaborators.

Chapter Three: Individual timing of incubation activities in city and forest birds, and implications for offspring fitness

Data were collected for this chapter during field seasons 2016, 2017 and 2018. Concept and experimental design by Robyn Womack (RW), Barbara Helm (BH), Francesco Baldini (FB), Jane Robinson (JR) and Peter O'Shaughnessy (PO). Incubation activity and nestling condition data were collected by RW, BH and field volunteers during 2016 and 2017. Incubation activity data were collected by field volunteers and MRes student Ciara McGlade (CM) in 2018. Initial data sorting on incubation data was carried out by CM using incR package developed by Pablo Capilla-Lasheras (PC). Data analyses for this manuscript was carried out by RW. Manuscript was written by RW.

Chapter Four: Effects of experimental exposure of low-level artificial light at night on circadian rhythms and health of wild nestlings

Data and samples for this chapter were obtained during the field season in 2017. Concept and experimental design by RW, BH, FB, JR and PO. Data collection was carried out by BH, Paul Jerem (PJ) and field volunteers. Development of the QuantiGene[®] Plex RNA assay for use on avian blood samples was carried out by RW, and assays to determine transcript levels of immune and clock genes of samples were carried out by RW. Molecular sexing of samples was carried out by Gillian O'Flaherty (GO) under supervision of Barbara Mable (BM). Data analyses was carried out by RW. Manuscript was written by RW.

Chapter Five: Avian malaria infection and condition in wild nestlings at city and forest sites

Nestling data and samples for this chapter were obtained during field seasons in 2016 and 2017. Concept and experimental design by RW, BH, FB, JR and PO. Malaria PCR assays to determine infection prevalence were carried out by RW. Development of the QuantiGene® Plex RNA assay for use on avian blood samples was carried out by RW. Assays to determine transcript levels of immune genes of all samples were carried out by

RW. Molecular sexing of samples was carried out by GO under supervision of BM. Data analyses and manuscript writing was carried out by RW.

Chapter Six: Malaria parasite effects on the circadian physiology of wild birds

Nestling data and samples were obtained during field seasons in 2016 and 2017.

Concept and experimental design by RW, BH, FB, JR and PO. Methodology to measure malaria infection intensity using qPCR was developed by PhD student Bedur Albalawi (BA) and FB. Malaria qPCRs to determine infection intensity were carried out by RW.

Development of the QuantiGene® Plex RNA assay for use on avian blood samples was done by RW. Assays to determine transcript levels of immune and clock genes of all samples were carried out by RW. Molecular sexing of samples was carried out by GO and BM. Data analyses and writing of the manuscript was carried out by RW.

Chapter One: General Introduction

Endogenous circadian rhythms of around 24 hours (from "circa" and "dies", meaning "about a day") are a core feature of animal physiology, vital for the normal function of rhythmic biological processes (Aschoff, 1965; Dunlap, 1999; Reppert and Weaver, 2002). Circadian clocks function to anticipate daily environmental changes and co-ordinate external factors with the internal behavioural and physiological processes of an organism, for example the releasing of enzymes to anticipate food intake (Stokkan et al., 2001).

Decades of research into biological clocks and their functions has elucidated the intricate workings of inner timekeeping mechanisms through *in vitro* and captive studies (Pittendrigh, 1954; Aschoff, 1967; Stephan & Zucker, 1972; Yamazaki et al. 1998; Vitaterna et al., 1999; Yoo et al., 2005), yet clocks remain understudied in evolutionary ecology. This thesis aimed to place existing work of clocks into a wider ecological context, by investigating biological rhythms in a wild model.

Birds are widespread, conspicuous and sensitive to habitat alterations, and therefore make great models for physiological studies in the wild. Much of avian behaviour is exhibited in rhythmic patterns, such as annual reproduction and migration behaviours, and daily foraging, singing and breeding activities. The avian system was therefore an ideal model to investigate ecological and evolutionary questions regarding biological clocks in an environmental context.

This review first outlines functions and mechanisms of the circadian clocks of mammals and birds, and then an appraisal of the literature concerning effects of environmental features on the timing of avian behaviour and physiology.

1.1 The circadian system: mechanisms and functions

Internal timekeeping mechanisms are ubiquitous throughout living organisms and are highly conserved, containing largely the same molecular components from cyanobacteria to mammals (Dunlap, 1999). In mammals, circadian systems are organised in a hierarchy of multiple circadian oscillators (Reppert and Weaver, 2002), with the suprachiasmatic nucleus (SCN) within the hypothalamic region of the brain acting as the master pacemaker, governing rhythms of "slave" oscillators in peripheral organs and tissues (Gaspar and Brown, 2015). Timing mechanisms between the SCN and peripheral clocks are very similar (Yagita et al., 2001), although the resulting temporal patterns of clock gene expression are often tissue specific (Pagani et al., 2010). Many processes of an organism are under circadian influence, such as the sleep/wake cycle (Saper et al., 2005), cycles of body temperature (Refinetti & Menaker, 1992), blood pressure (Pritchett & Reddy, 2015), metabolism (Karatsoreos et al., 2011), and cyclical release of hormones (Nicolaides et al., 2014).

In order to adapt to changing environmental conditions, circadian clock mechanisms must be actively synchronised to external cycles in a process known as entrainment (Pittendrigh, 1981). The light-dark cycle is the most potent environmental cue, otherwise known as zeitgeber, for clock synchronisation. However, it is also possible to entrain clocks to temperature cycles, or any environmental factor that fluctuates with consistent timing over the 24 h day (Pittendrigh, 1981).

1.1.1 Cogs of the circadian clock

Circadian studies have successfully characterised the molecular mechanisms behind the mammalian clock, since the first clock gene was discovered in mice (Vitaterna et al., 1994). In short, the mammalian clock is comprised of interconnecting core and stabilising loops, and overall functions as a transcriptional-translational feedback loop (Reppert and Weaver, 2002), (Figure 1.1.). The core loop is made up of positive elements CLOCK and BMAL1 and negative elements PER and CRY and functions as follows: CLOCK-BMAL1 heterodimers bind to specific E-box elements in target gene promoters, activating the rhythmic transcription of three period genes (*PER1*, *PER2*, *PER3*) and two cryptochrome genes (*CRY1*, *CRY2*), (Buhr and Takahashi, 2013; Gekakis et al., 1998). Negative elements PER and CRY proteins subsequently dimerize, inhibiting further transcriptional activity of

CLOCK and BMAL1 (Buhr and Takahashi, 2013; Reppert and Weaver, 2002). One full cycle of this loop is completed in just over 24 h, with positive element *BMAL1* expressed around 12 h out of phase with *PER* and *CRY* (Reppert and Weaver, 2002). It is this cycle of core clock genes within the SCN that governs biological cycles downstream in peripheral tissues and cells, driving rhythmic expression of proteins. Additional loops buffer and refine this molecular cycle, and link together the circadian and metabolic systems, with gene REVERBA (also known as NR1D1) at the interface, promoting transcription of *BMAL1* and *CLOCK* (Ueda et al., 2005).

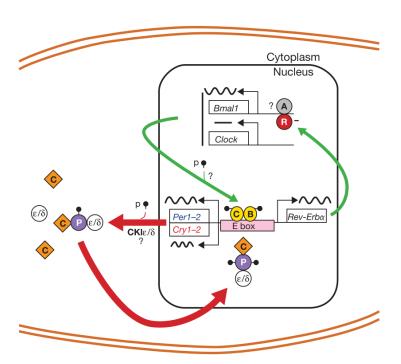


Figure 1.1. The current mammalian circadian clock model. From (Reppert and Weaver, 2002). The mammalian circadian clock consists of an interconnecting positive clock loop (green) and a negative loop (red). Positive elements *CLOCK* (C, yellow oval) and *BMAL1* (B, yellow oval) form heterodimers and facilitate the transcription of negative elements *PER* (purple circles) and *CRY* (C, orange diamond). *PER* and *CRY* are phosphorylated outside the nucleus, and within the nucleus the newly phosphorylated complex shuts down further transcription of *BMAL1* and *CLOCK*. During the positive loop of the clock, *REVERBA* (R, red circle) promotes transcription of *BMAL1* and *CLOCK*.

Circadian oscillators can become desynchronised in the absence of a strong zeitgeber for entrainment (Aschoff et al., 1967). In humans, air travel through multiple time zones, shiftwork and sleep deprivation can all lead to desynchronization of peripheral oscillators to the SCN (Karatsoreos et al., 2011). Many homeostatic and immune functions are under circadian control (Scheiermann et al., 2013), and therefore chronic desynchronization of circadian oscillators to the SCN has been demonstrated to have severe adverse effects on

health. Experimental evidence on human and mouse demonstrates disrupting natural rhythms to have profound effects on the brain, metabolism and behaviour (Karatsoreos et al., 2011), metabolic dysregulation (Huang et al., 2011), mood disorders (McClung, 2013), inflammation (Preuss et al., 2008), and increasing incidences of cancer (Fonken and Nelson, 2014).

Following a phase shift of the clock, a stable relationship between peripheral circadian oscillators and the SCN must be re-established (Karatsoreos et al., 2011). Clocks can be reentrained and normal rhythms restored by exposure to light-dark cycles via light-induced signals from the retina, which in turn modify expression of clock genes in the SCN (Schroeder and Colwell, 2013; Wright et al., 2013).

1.2 Avian clocks – mechanisms, functions and outputs

Although not as well characterised as the mammalian clock, current models of avian clock mechanisms are largely accepted. Three pacemakers make up the core avian clock mechanism, with the pineal gland rather than the SCN acting as "master pacemaker", along with additional core oscillators in the retina and hypothalamic region (Cassone, 2014; Gwinner and Brandstätter, 2001; Kumar et al., 2004), (Figure 1.2). These structures are thought to reinforce each other's rhythmicity in a "neuro-endocrine loop" that influences processes downstream (Cassone and Menaker, 1984), however the exact contribution of these three structures to avian circadian organisation has still yet to be defined (Karaganis et al., 2009).

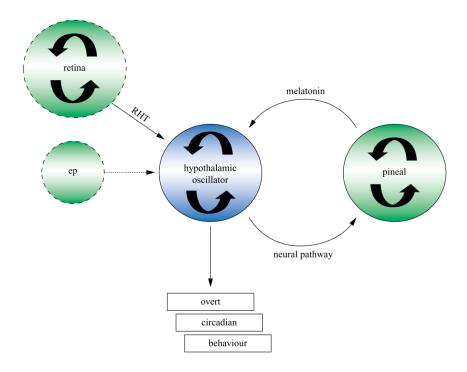


Figure 1.2. Components of the avian circadian clock. From (Gwinner and Brandstätter, 2001). Oscillators in the pineal, hypothalamus and retina contribute to avian circadian outputs in birds. Sites of photoreceptors include the retina and pineal (green), and encephalon (ep). RHT = retinohypothalamic tract.

Avian circadian clock genes are expressed in a similar fashion to those in the mammalian clock except for *PER1*, which is not present in avian genomes (Yasuo et al., 2003; Yoshimura et al., 2000). Molecular analyses of avian clock genes demonstrate *BMAL1* and *CLOCK* to be expressed during the late subjective day (in diurnal birds, twilight) and *PER2* and *PER3* expressed during subjective night to mid-subjective day (Cassone, 2014; Chong et al., 2003; Karaganis et al., 2009; Yoshimura et al., 2000).

1.2.1 The pineal gland and melatonin

Research into avian circadian systems has shown the pineal gland to be the major pacemaker in songbirds (Cassone, 2014; Wang et al., 2014). Pinealectomy studies, or surgical removal of the pineal gland, have revealed the pineal to be essential for sustaining activity rhythms during constant darkness (Gaston and Menaker, 1968), maintaining cycles of body temperature (Binkley et al., 1971), and daily patterns of singing behaviour (Wang et al., 2012). The pineal gland modulates rhythms by secretion of the hormone melatonin (Gwinner et al., 1997; Menaker and Zimmerman, 1976), which has downstream influences on patterns of behaviour. In addition to the control of circadian behaviour, pineal melatonin is possibly indirectly involved in the regulation of some seasonal behaviours (Gwinner, 2008), (Figure 1.3.).

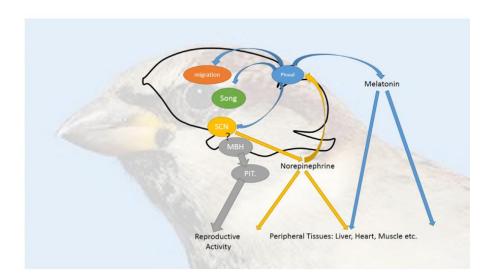


Figure 1.3. The avian pineal gland may have modifying effects on downstream rhythms of seasonal reproduction, migration and song behaviour. From Cassone (2014).

1.2.2 Avian photoreception and entrainment

For the circadian clock to entrain to external light-dark cycles, light information must first be received, transduced and translated. Unlike mammals, where the SCN receives all photic input via the retina, birds and other non-mammalian vertebrates possess extra-ocular photoreceptive structures (Cassone, 2014; Thakur and Kumar, 2015). Functional photoreceptors are present within the avian pineal gland and deep-brain regions, in addition to the retina (Thakur and Kumar, 2015). Rhythms of pineal melatonin can also be entrained by light entering the retina (Barrett and Underwood, 1991).

1.2.3 Development of avian rhythms

In avian embryos, transcripts of *BMAL1* and *CLOCK* are already present in the foregut at 48-52 hours old (Gonçalves et al., 2012). Oscillations of *BMAL1* are later present in the chick retina at eight days (De Lima et al., 2011) and in the liver at four days old (Zeman et al., 2009). Circadian rhythmicity of melatonin biosynthesis occurs in the pineal gland during the first day post-hatching (Zeman and Herichová, 2011) for both precocial and altricial species of birds (Zeman & Gwinner, 1993). The amplitude of this rhythm increases with age within the first two weeks post-hatching (Zeman & Gwinner, 1993; Van't Hof and Gwinner, 1996).

1.3 Environmental influences on avian clocks

Many features within an organisms' environment can have influence on the timing of behaviour and physiology. In the following section, current knowledge of how several important environmental features influence the circadian rhythms of birds is reviewed.

1.3.1 Ambient light

The light-dark cycle is the most important synchronising cue for clock mechanisms (Pittendrigh, 1981). Given that birds possess photoreceptors in neural tissue and the pineal as well as in the retina (Thakur and Kumar, 2015), birds are arguably more light-sensitive than mammals. Previous captive studies on birds have shown rhythmic avian behaviours, such as song and activity, to be dictated by light and subsequent secretion of pineal melatonin (Wang et al., 2012).

Photoperiodic changes are important triggers for the seasonal biology of birds (Dawson et al., 2001; Gwinner, 2008). For example, exposure to long photoperiods can override other environmental cues to trigger physiological changes associated with reproduction (Dominoni et al., 2013b; Lambrechts and Perret, 2000). Light intensity can also influence perception of daylength; male European starlings (*Sturnus vulgaris*) exposed to low intensity light took longer to respond to photoperiodic changes than individuals under higher light intensities (Bentley et al., 1998).

Exposure to light during the subjective night can have particularly disruptive effects on rhythms (Fonken and Nelson, 2011). A captive study exposed great tits (*Parus major*) to five different intensities of nocturnal light and found strongest disruptive effects on activity patterns to occur with strongest light intensity (de Jong et al., 2016). This study also measured plasma melatonin concentrations and found a decrease in concentration with increasing intensity of light treatment. Light at night also disrupted rest and increased nocturnal activity in zebra finches (Alaasam et al., 2018). In great tits, dim light at night treatment shifted the onset of activity, but had no effect on circadian period length (Spoelstra et al., 2018).

1.3.2 Ambient temperatures

Birds and other animals possess a circadian rhythm of internal body temperature, which shows robust rhythmicity independent of environmental temperature oscillations, constant ambient temperatures and constant light (Merrow et al., 2005; Refinetti and Menaker, 1992). However, ambient temperature cycles can also entrain circadian rhythms (Pittendrigh, 1981). One of the fundamental features of the circadian clock is that rhythms are temperature-compensated, where the rate of the clock remains consistent during ambient temperature fluctuations that are physiologically permissive (Hastings, 2013; Pittendrigh, 1954).

Ambient temperature changes may have some influence on rhythms, however. A captive experiment by Lehmann et al., (2012) exposed great tits to temperatures of either 18 °C or 8 °C. Birds exposed to the higher temperature started their daily period of activity later, and terminated activity earlier in the day than those kept under 8 °C. Although observed shifts in daily activity were small, the authors suggest natural conditions where the temperature difference is larger than the 10 °C tested may have a larger effect on activity patterns (Lehmann et al., 2012). In another experiment, ambient temperature fluctuations influenced activity rhythms of Malachite sunbirds (*Nectarinia famosa*) directly by inducing torpor at low temperatures (Downs and Brown, 2002).

1.3.3 Food availability

Both the timing and the quality of food intake can affect biological rhythms in animals. Bird species follow daily patterns of foraging activity, usually peaking in the morning and the afternoon (Bednekoff and Houston, 1994), and can be entrained to feeding cycles under captive conditions (Hau and Gwinner, 1992). Timing of food intake acts as a zeitgeber in a process known as feeding-entrainment, coupling metabolic activity to molecular clocks of the SCN and liver (Stokkan et al., 2001). Mistimed or restricted feeding has been shown to decouple oscillators of metabolic and molecular clocks, causing health problems (Chabot and Menaker, 1992; Damiola et al., 2000). In addition, adequate nutrition is required for the generation of normal metabolic rhythms (Potter et al., 2016). An example of this from mouse models shows high-fat diets have been shown to alter period length and locomotory activity (Kohsaka et al., 2007).

1.3.4 Immunity, disease and parasites

Circadian clocks are closely linked to the immune system (Scheiermann et al., 2013). In birds, pineal melatonin regulates rhythms of innate immunity and seasonal immunity (Markowska et al., 2017). Diurnal changes in mRNA of interleukins *IL6* and *IL18* have been observed in chicken (*Gallus gallus domesticus*) leucocytes (Turkowska et al., 2013). Rhythmic oscillations of proinflammatory cytokines for inflammatory defences have been shown in the avian spleen, also regulated by melatonin (Naidu et al., 2010).

In their natural environment, birds are exposed to parasites and disease (Hamilton and Zuk, 1982). Parasites may directly impact on avian rhythms; for example, infestations with ectoparasites increased nocturnal preening activity and restlessness in chickens (Jacobs et al., 2019). Host-parasite interactions may also be complex, given the temporal fluctuations in host immunity (Martinez-Bakker and Helm, 2015). For example, malaria blood parasites (*Leucocytozoon, Haemoproteus* and *Plasmodium* species) exhibit rhythmicity during their life cycle within a host (Reece et al., 2017). Avian malaria species are also rhythmic (Gore et al., 1982), but interactions with the host circadian system have yet to be defined.

1.4 Measuring rhythms of individuals

1.4.1 Behaviour

In the absence of any environmental cues to entrain to, circadian clock mechanisms will "free-run" at a period length of approximately 24 h (Buhr and Takahashi, 2013). Endogenous circadian rhythms of individuals can be quantified by measurement of this free-running period length (or *tau*) under constant conditions such as constant light (LL) or dark (DD). Period length is highly repeatable in individuals and is usually determined either by periods of sleep and wake or by timing of locomotory activity, such as wheel-running in mice (Brown et al., 2005), or perching activity in birds (Underwood et al., 2001).

In the presence of cues in the natural environment, such as the light-dark cycle, individual timing in animals can be measured by quantifying rhythms of behavioural activity. Differences in behavioural timing can be used as a proxy for differences in endogenous rhythms, as the phase angle between the zeitgeber and activity timing depends on individual free-running period (Roenneberg et al., 2003).

Timing of individual activity patterns of animals can be quantified in their natural environment, given that the behaviour is repeatable and consistent. In birds, this can for example be timing of incubation behaviours (Bulla et al., 2016), sleep patterns (Stuber et al., 2015), and song (Naguib et al., 2019). Individual sleep-wake activity is often classified using the term "chronotypes", with early chronotypes (colloquially known as "larks") rising earlier than later "owl" chronotypes (Roenneberg et al., 2003).

1.4.2 Molecular mechanisms

Despite central oscillators of the circadian system in the SCN being synchronised by external cues, molecular clock mechanisms are cell autonomous, and rhythms can persist in isolated tissues. This allows for *in vitro* studies of circadian rhythms in biopsies from individual animals (Brown et al., 2005; Gaspar and Brown, 2015). In addition, the advance of microarray technologies has facilitated studies of cyclical gene expression, allowing for comparisons of clock mechanisms between species (Reddy, 2013). Since Harmer et al., (2000) mapped the circadian transcriptome in *Arabidopsis*, further studies have elucidated the circadian transcriptomes within avian tissues such as liver and heart (Zeman et al., 2009), brain (Karaganis et al., 2009), spleen (Naidu et al., 2010), ovary (Laine et al.,

2019), retina (Bailey et al., 2004), and pineal (Bailey et al., 2002). As in other species, expression of avian core clock genes appears to be tissue specific (Laine et al., 2019).

Despite advances in molecular methods facilitating studies of mechanisms of the circadian clock in many different species, the potential for these methods to be used in studies of wild animals has yet to be fulfilled.

1.5 Clocks in the wild: studies of daily rhythms of birds in the natural environment

Despite the importance of circadian rhythms for the physiology and behaviour of animals, relatively little is known about biological clocks in ecology (Kronfeld-Schor et al., 2013). In birds, observational differences in rhythms have been recorded at species' level by measurement of timing of behaviours such as arrival at a feeder (Fitzpatrick, 1997), timing of dawn song (Kempenaers et al., 2010; Naguib et al., 2019) or departure from nocturnal roosts (Graham et al., 2017; Hubálek, 2017).

For wild studies in individual birds, behavioural activity patterns are quantified to later infer rhythms. Usually, data loggers are used to measure activity patterns around the clock. One study used remote temperature sensors to record timing of incubation behaviours and infer activity patterns in different wader and seabird species (Bueno-Enciso et al., 2017; Bulla et al., 2016). Time–depth recorders have also been used to infer activity budgets of seabirds (Linnebjerg et al., 2014), or radio telemetry (Jansen et al., 1998; Keitt et al., 2004; Steiger et al., 2013). Another study investigated plasticity in clocks by quantifying sleep behaviours of individual roosting great tits from nest box camera footage (Stuber et al., 2015).

Observational studies of rhythms of individual birds can help to answer ecological questions on the function and modulations of clocks, such as how environmental changes influence timing. For example, overnight ambient temperatures influenced the start of song in male great tits, although the differences in song timing between individuals remained consistent (Naguib et al., 2019). In addition, the presence of artificial light at night influences the start of daily activity in passerines, with an earlier onset of activity dependent on the species (Kempenaers et al., 2010).

To date, few wild studies have explored the adaptive significance of behavioural timing. For example, in a study on male blue tits (*Cyanistes caeruleus*), individuals that rose earlier had higher success at extra-pair paternity (Poesel et al., 2006), suggesting that individual variation in timing may be under selection.

Experimental studies on individual rhythms can help to further understanding of how birds respond to environmental change. A nest box study investigating impacts of light at night on avian physiology and behaviour exposed free-living great tit nestlings to constant illumination at night, finding negative effects on sleep (Raap et al., 2016c). In addition,

another nest box study that experimentally increased nest temperatures caused birds to wake up approximately 30% more frequently during the night (Stuber et al., 2017).

1.5.1 Wild clocks in a model avian species; the great tit

The great tit (*Parus major*) is a model avian species and a widely studied passerine in ecology (Gosler, 1993; Lack, 1964). Great tits are a diurnal species, and show consistent daily rhythms in behaviours such as dawn and dusk song (Da Silva et al., 2015; Naguib et al., 2019), early morning feeding (Fitzpatrick, 1997) and incubation efforts (Bueno-Enciso et al., 2017). Intraspecific variation in rhythms have been observed for sleep behaviours in individuals (Stuber et al., 2016, 2015), individual timing of breeding (Graham et al., 2017), and nocturnal resting behaviours (Caorsi et al., 2019).

Great tits have been studied in the context of biological timing, and particularly effects of the environment on annual and daily cycles of behaviour. For example, the warming global climate has led to advancing vegetation phenology, and consequently, adaptive selection for earlier initiation of breeding activity in great tits (Visser et al., 1998).

Moreover, observations of behaviour of great tits as a response to environmental changes including an increase in vigilance time and reduction of feeding effort during periods of anthropogenic noise (Klett-Mingo et al., 2016). In addition, great tits under light at night have been shown to sing earlier in the morning (Kempenaers et al., 2010). Experimental studies on rhythms of wild great tits have shown sleep to be disrupted by light pollution (Raap et al., 2016c, 2015). In particular, white light exposure has been linked to an increase in nocturnal activity, decrease in oxalic acid levels (a marker for sleep debt), and an increase incidence of infection with avian malaria in adult great tits (Ouyang et al., 2017). Light pollution has also been linked with an increase in feeding rate of great tit parents during nestling provisioning (Titulaer et al., 2012).

1.6 Knowledge gaps

Previous research into biological rhythms of wild birds has provided evidence that individual activity patterns can be measured and quantified using logger technologies (Bulla et al., 2016) or behavioural observations (Stuber et al., 2015). However, classifying circadian phenotype into individual "chronotype" in a wild animal has rarely been done (Graham et al., 2017). In addition, despite the wealth of knowledge obtained from captive studies of avian molecular clock mechanisms, there has been no previous attempt to quantify circadian clocks at the molecular level in a wild bird, and few studies have investigated how individual differences in timing might relate to fitness of wild birds (Hau et al., 2017; Poesel et al., 2006).

Captive studies have demonstrated timing of cues such as light (de Jong et al., 2016) and food (Hau and Gwinner, 1992) to affect daily rhythms of birds. However, there is limited evidence into how such changes affect timing in birds in the natural environment, especially at the mechanistic level. Further research is needed into how rhythms of individuals are shaped by these environmental changes, such as the presence of light at night (Raap et al., 2016c; Stuber et al., 2017).

Birds face many pathogens in their natural environments (Ellis et al., 2014). Although timing of the avian immune system is under circadian control (Markowska et al., 2017), there has been no investigation into time of day differences of immune system activity in wild birds. Research is needed into how fitness challenges for birds, such as wild infections with parasites (van Riper et al., 1986), affect the host circadian and immune systems.

1.7 Aims of Thesis

The overarching aim of this thesis was to put studies of biological rhythms into the context of the organisms' natural environment. Specifically, the aims of this thesis were:

- 1. To further the understanding of how environmental factors influence timing in a wild animal
- To investigate how changes in the environment, such as the city habitat and artificial light at night, affects timing in individuals at behavioural and transcript levels
- 3. To investigate consequences of wild infections with avian malaria parasites for host condition and immune and circadian physiology

In this thesis, the great tit was used as the study species, a common passerine widely used in ecological research (Gosler, 1993; Lack, 1964). Observational and experimental approaches were used to quantify individual rhythms of great tits at behavioural and molecular levels, using knowledge of circadian clock mechanisms and available genetic tools for the species (Laine et al., 2016; Santure et al., 2011). Many abiotic and biotic features can have influence on the clock, and therefore, two aspects of the environment were selected as the focus of chapters in this thesis: ambient light and host-parasite interactions (Figure 1.4.).

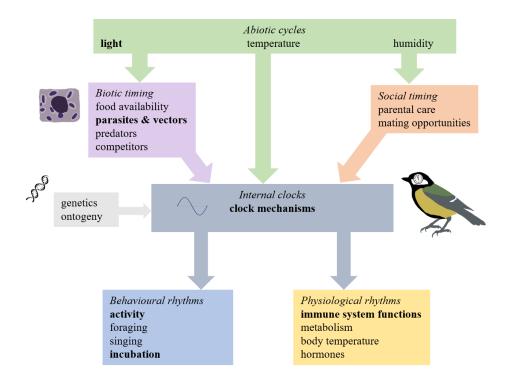


Figure 1.4. Influences of biotic and abiotic factors on internal clocks, and outputs. Factors studied in this thesis are highlighted in bold. Diagram adapted from (Helm et al., 2017).

1.8 Thesis outline

Chapter Two provided an overview of the General Methods used in the laboratory and field for data collection in this thesis, and an overview of the study species and nest box system.

Chapter Three of this thesis investigated differences in individual timing in great tits in city and forest habitats, using thermal sensors to quantify incubation behaviours. Individual "chronotype" was then determined from the timing of onset of activity of the mother inferred from incubation patterns. Chronotype of individual birds was then linked to parameters of condition in offspring, to determine whether differences in individual timing are important for the fitness of wild birds.

Chapter Four of this thesis investigated whether disruptive effects of artificial light at night on rhythms found in captive studies was reflected in wild birds. An experimental approach was used to investigate effects of low-level artificial light at night on nestlings. Molecular tools were used to quantify time of day differences in circadian clock transcripts, and transcripts of genes involved in the avian immune system. This chapter explored effects of exposure to artificial light on nestlings' clock, immune system and parameters of condition such as weight and haematocrit levels.

Chapter Five explored how wild infections with pathogens can affect condition of nestlings in city and forest habitats. This study took advantage of the high prevalence of *Leucocytozoon* avian malaria parasites in the study system (Capilla-Lasheras et al., 2017), to investigate fitness effects of infections on immune gene transcripts and condition of nestlings.

Chapter Six of this thesis investigated host-parasite effects on the circadian system of wild birds. In this study, transcript levels of circadian clock and immune genes were measured across a temporal profile to determine rhythmicity. Avian malaria (*Leucocytozoon* species) infections were also quantified and linked to rhythmicity in nestlings, to explore whether parasites impact on circadian and immune mechanisms in wild birds.

Finally, the **General Discussion** section provided an overview of the findings of this thesis, limitations of the studies undertaken, and contributions to the advancement of the fields of chronobiology and ecology.

Chapter Two: General Methods

Data in this thesis were based on samples collected during three breeding seasons from April to June 2016-2018 in city and forest field sites in Scotland, United Kingdom. The following General Methods section outlines the field and laboratory methods that apply to chapters in this thesis.

2.1 Study species

In this thesis, the great tit (*Parus major*) was used as the study species. The great tit is a passerine species, widely distributed across Europe and Asia, and commonly sighted in oak woodlands and urban parks (Gosler, 1993). A small passerine, adult great tits weigh 13-21 g, and their plumage colouration consists of a black cap, nape and collar, white cheeks, yellow underparts and green and blue upperparts (Gosler, 1993). Male great tits are distinguishable from females by their larger size, and wider black belly-stripe (Norris, 1993). In juveniles, plumage areas which are white in adults are yellowish in colour, before their post juvenile-moult (Gosler, 1993). Brightness of feather colouration in great tits is dependent on diet and overall condition (Hõrak et al., 2001; Slagsvold and Lifjeld, 1985).

Great tits have a predominantly insectivorous diet (Perrins, 1991), and time their breeding activities to match peak caterpillar abundance (Noordwijk et al., 1995). Egg laying dates vary from April to May, and clutch size ranges from 7-12 eggs (Perrins and McCleery, 1989). Incubation of eggs is carried out by the female only, and lasts for approximately two weeks before hatching (Gosler, 1993; Haftorn, 1981).

Both parents are involved in nestling provisioning of prey such as caterpillars and spiders (Hinde, 2006; Naef-daenzer et al., 2016). Nestlings remain in the nest for 18-21 days until their full plumage has developed, finally fledging with a body weight of around 16-22 g (Gosler, 1993). The weight of nestlings at fledgling predicts their prospects for survival (Tinbergen and Boerlijst, 1990) and recruitment into the great tit population post-fledging (Both et al., 1999). Predators of the great tit include great spotted woodpeckers (*Picoides major*) or weasels (*Mustela nivalis*), (Curio and Onnebrink, 1995; Dunn, 1977), which commonly raid passerine nests. Sparrowhawks (*Accipiter nisus*) are also common predators of adult great tits and young fledglings (Götmark and Andersson, 2005).

Great tits are primarily cavity nesters (Gosler, 1993), however, they will readily nest in nest boxes. Nest boxes studies using great tits have facilitated studies of biological rhythms in ecology (Graham et al., 2017; Raap et al., 2015; Stuber et al., 2015). In this thesis, a nest

box system was used for observational and experimental studies on a breeding population of great tits and nestlings. This thesis also made use of the genetic tools available for great tits, including an annotated reference genome (Laine et al., 2016) and transcriptome (Santure et al., 2011; Watson et al., 2017).

2.2 Descriptions of field sites used

This study was conducted using an existing nest box system across two urban field sites and two forest field sites (Table 2.1.), which represent an urban (city) to rural (forest) gradient from Glasgow city to Loch Lomond and the Trossachs National Park, Scotland, United Kingdom.

Given the limited occupancy of great tits at nest boxes at city sites and resulting low numbers of nestlings, for this study, sites were classified as belonging to either city or forest categories to increase power of city and forest comparisons.

Table 2.1. Descriptions of study sites used in this thesis.

Site

	Cashel Native Forest (56° 6' N, 4°34' W)	Scottish Centre for Ecology and the Natural Environment (SCENE) (56°, 7.73N 4° 36.79W)	Garscube Campus (55° 9' N, 4°31' W)	Kelvingrove Park (55°52' N, 4°17' W)
Type	Deciduous forest	Deciduous forest	Suburban parkland	Inner city parkland
Site classification	Forest	Forest	City	City
Composition of tree species	Oak (<i>Quercus</i> sp.), and birch (<i>Betula</i> sp.)	Oak (<i>Quercus</i> sp.)	Open land with abundance of introduced and native species	Parkland, introduced species and sparse birch and oak
Total number of nest boxes at site (2016)	131	289	40	71

2.3 Nest box monitoring protocol

Throughout the breeding season (April-June), all nest boxes (woodcrete material, 260H x 170W x 180D, hole diameter 32 cm, supplier Schwegler, Germany) were checked at least weekly for signs of nest building activity and egg laying. Nests were identified as belonging to a great tit if eggs were >13mm in width (Van Noordwijk et al., 1981). Nest boxes were checked once every two days, two weeks after the estimated start of incubation to determine exact date of hatching (Haftorn, 1981). Except in the case of experimentally manipulated nest boxes in Chapter Four, where samples were taken on day eight, disturbance to nest boxes was avoided from hatch date until day thirteen post-hatching. On day thirteen, all nestlings within a brood were weighed and ringed for individual identification. Nest boxes were checked again after fledging (>21 days post-hatching) to determine the numbers and identities of any dead nestlings. In total, 177 great tit nest boxes were monitored during this study, 137 of these at forest sites, and 40 in the city (see Table 2.2. for breakdown by year and site).

Table 2.2. Number of great tit nests monitored per field site and nestlings processed during this study (2016-2018).

Site	Site Type	Year	Number of great tit broods monitored
		2016	29
SCENE	Forest	2017	32
		2018	8
		2016	31
Cashel Forest	Forest	2017	33
		2018	4
		2016	7
Kelvingrove Park	City	2017	11
		2018	7
		2016	3
Garscube Estate	City	2017	7
		2018	5

2.4 Citronella treatment

During the 2017 breeding season, fifteen forest nest boxes were treated with citronella-based insect repellent as per Krams et al., (2013). According to Krams et al., (2013), inclusion of citronella treatment in their study nest boxes reduced the presence of avian malaria vectors by 80 %. In this study, citronella repellent was added to nest boxes with the aim of reducing abundance of avian malaria vectors (and therefore, avian malaria infections) in nest boxes used in Chapters Four and Five. Repellent consisted of 0.5 ml citronella oil, 200 mg carrageenan kappa, and 0.5 ml water. Four eppendorfs per nest box were first filled with citronella oil mixture and pierced with a 3 mm hole in the lid. One day prior to the expected hatch date entrance (two weeks after the start of incubation), these eppendorfs were fastened to the inner wall of nest box doors, 3 cm away from the nest box entrance.

2.5 Nestling processing and blood sampling

Ringing and measurements of all nestlings took place on day thirteen post-hatching during the 2016 and 2018 breeding seasons. In 2017, ringing and measurements of nestlings took place on day eight post-hatching, as part of the light at night study in Chapter Four. Blood sampling of all nestlings took place on day thirteen, and usually during the daylight hours. For some studies, blood sampling was carried out around the clock (see Chapters Four and Six).

During sampling, all nestlings of a brood were taken from the nest box and placed into a fabric drawstring bag, to avoid escapees. Nestlings were then individually processed as follows. Nestlings were ringed with an aluminium ring detailing individual identification number using rings obtained from the British Trust for Ornithology, and weighed using a portable balance (Spencer, 1976). Where tarsometatarsus (tarsus) measurements were required (Chapter Four and Five), the base of the metatarsals of each nestling were measured using dial callipers (Eck et al., 2011).

Blood sampling was only carried out on nestlings of a healthy weight (17-22 g) to reduce burden on nestlings. Prior to blood sampling, a cotton bud dipped in water was used to dampen feathers and expose the brachial wing vein, to increase accuracy of sampling. The brachial wing vein was then punctured following sampling protocol outlined in Owen (2011). The resulting blood bead was then collected using a heparinised microcapillary (Vitrex) and immediately pumped into an eppendorf containing 250 µl of RNA*later*® stabilization solution (Invitrogen) for RNA sampling, and into an eppendorf containing 500

μl 100% ETOH for DNA sampling. In total, a maximal amount of two 75 μl capillaries were collected per nestling. If this was not possible, samples were obtained for RNA and DNA from different siblings from the same nest. During the field season of 2017, some blood samples were collected for measurement of haematocrit levels (%), (see below section for details). On returning from the field, samples were placed into a refrigerator at 5 °C for three days, before storage at -40 °C.

2.6 Haematocrit measures

During the 2017 breeding season and for Chapter Four and Five, blood samples were taken for measurement of haematocrit, or the percentage volume of red blood cells compared to total blood volume, as a measure of nestling health (Ots et al., 1998). On day thirteen of life, blood samples were obtained from nestlings using the sampling protocol detailed above, taking ½ to ¾ of a 75 µl heparinised haematocrit microcapillary (Vitrex) per nestling. Immediately following sampling, capillaries were sealed directly using a Cristaseal wax plate (Hawksley) and stored on wet ice for transport to the laboratory. Capillaries were placed into a centrifuge and spun for 5 minutes at 5000 RPM. Following this, capillaries were measured against a silver ruler (Figure 2.1.). Three measures were taken, starting at the wax end of the capillary each time: 1) end of the capillary to the seal wax plug; 2) end of the capillary to the packed red blood cell fraction; 3) end of the capillary to the end of the plasma region.

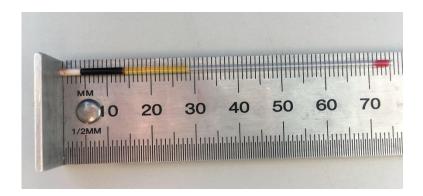


Figure 2.1. Measurement of plasma and packed red blood cell volume from a capillary containing great tit blood after centrifugation, to determine haematocrit levels (%)

Following these measurements, the length of the seal wax was subtracted from the total length of the sample (including the length of the red blood cell fraction), prior to calculation of haematocrit levels. Haematocrit levels (%) were then calculated as the length of the packed red blood cells, divided by the length of the total sample (red blood cells and

plasma). Each capillary was then scored with a glass cutter, and using a Pasteur pipette, pumped blood sample into an eppendorf containing 500 μ l 100% ETOH. Blood samples were then placed into storage at -40 $^{\circ}$ C.

2.7 Sample sizes and breakdown

In total, 297 nestlings were sampled during this study across the two sampling years (2016-2017). Of these samples, 259 were collected from forest sites, and 38 from city sites (see Table 2.3. by a breakdown by year and site).

Table 2.3. Number of nestlings sampled per site

			8 1	
Year	Cashel Forest	SCENE	Garscube	Kelvingrove
				Park
2016	50	56	5	8
2017	97	56	13	12
2018	No	blood samples wer	re obtained for this	year

Site / Number of nestlings processed

This study was conducted across three breeding seasons, and nestling data were used in more than one chapter of this thesis. Nestlings used as controls in Chapter Four were included in Chapters Three, Five and Six as part of the chronotype and malaria studies. However, nestlings exposed to low-level light at night treatment in Chapter Four were not included in any other chapter in this thesis.

Sample sizes for each chapter also differed due to the methods used within this thesis. For example, in Chapter Three, nestlings were only used in the chronotype and fitness analyses if iButton recordings were also obtained for incubating females of that brood. Similarly, nestlings were only used in analyses for Chapters Five and Six if samples had also been obtained for quantification of malaria infection and immune and clock (in the case of Chapter Six) gene transcripts. Samples in Chapter Five were taken during the noon hours, whereas samples in Chapter Six were taken at times throughout the day and night. A breakdown of number of nestlings used in analyses for each chapter is available in Table 2.4.

Table 2.4. Overview of number of nestlings from which data were collected for each chapter of this thesis.

Chapter	Study Years	Number of Nestlings
Chapter Three	2016, 2017, 2018	253
Chapter Four	2017	213
Chapter Five	2016, 2017	174
Chapter Six	2016, 2017	199

2.8 Ethical statement

All blood sampling of nestlings was carried out under the appropriate Home Office project licence held by Barbara Helm, and personal licences for avian sampling held by Robyn Womack and field technicians. Ringing of nestlings was carried out by persons holding appropriate licences obtained from the British Trust for Ornithology.

2.9 DNA extraction

In this thesis, genomic DNA samples were required for molecular sexing (Chapters Four and Five), for avian malaria prevalence testing (Chapter Five), and quantification of avian malaria infection (Chapter Six). Genomic DNA extraction was carried out on previously frozen (-40 °C) whole blood samples stored in 500 µl of 100 % ETOH, using a commercial DNeasy whole blood extraction kit (Qiagen). Following extraction, DNA samples were eluted in 80 µl nuclease-free water. Successful extraction of DNA was confirmed by measuring 1.5 µl of each sample using a Nanodrop ND-1000 Spectrophotometer (Nanodrop Technologies Inc., Wilmington, DE).

2.10 Molecular sexing

For Chapters Four and Five, the sex of each individual nestling was determined by application of a PCR-based protocol as described in Griffiths et al., (1998). This method amplifies target avian genes CHD-W and CHD-Z. Two distinct bands in a well indicates presence of two products of different sizes and is said to be female (ZW). One distinct band in a well indicates presence of only one product and is said to be male (ZZ).

Molecular sexing work for this thesis was carried out by honours student Gillian O'Flaherty under the supervision of Elizabeth Kilbride and Prof. Barbara Mable.

Each reaction was carried out in volumes of $10 \,\mu l$, and contained the following components: $1 \,\mu l$ buffer (Promega), $2 \,\text{mM}$ MgCl₂ (Promega), $0.8 \,\text{mM}$ each of forward and reverse primer of target genes, $0.16 \,\text{mM}$ dNTP, $0.375 \,\text{unit}$ Taq DNA Polymerase (Promega) and $5\text{-}10 \,\text{mg}/\,\mu l$ of extracted DNA template. Primer sequences for CHD-W and CHD-Z were obtained from Griffiths et al., (1998).

PCR runs included a negative control of nuclease-free water and two positive control samples, one from a known female and one from a male. Samples were placed in a PCR machine (MJ Research DyadTM Peltier Thermal Cycler) and amplified at 94 °C for 2 min, 49 °C for 45 s, 72 °C for 45 s, 94 °C for 30 s, step 2×29, 49 °C for 1 min, 72 °C for 5 min.

Samples were run on a 2% agarose gel stained with $10 \,\mu l$ SYBR safe along with a $100 \,bp$ marker, and bands were visualised using a Gel Documentation System (Bio-Rad). Any unclear results were repeated, along with one randomly selected sample from every ten successful samples, to test for potential error.

2.11 Avian malaria prevalence testing

To test for the presence of *Haemoproteus / Plasmodium* and *Leucocytozoon* parasite DNA within samples (Chapter Five), an existing nested polymerase chain reaction (PCR) protocol (Hellgren et al., 2006) was used. This protocol uses two primer sets, the first to amplify the 478 bp fragment of the mitochondrial cytochrome b, and the second to distinguish between *Plasmodium* or *Haemoproteus* and *Leucocytozoon* species (Table 2.5.). Positive controls were used from birds with known avian malaria infections evident from sequencing results, and negative controls where nuclease-free water was added to reactions in place of DNA template. Presence of DNA in negative samples was later checked for presence of an avian housekeeping gene, an 80 bp fragment of reference gene glyceraldehyde-3-phosphate dehydrogenase (GAPDH, Table 2.5.). GAPDH primer sequences were obtained from Atema et al., (2013).

PCR reactions were performed in volumes of 20 μ l, where each reaction contained 1.2 μ M of each of the two primers, 10 μ l GoTaq® universal PCR master mix (Promega), 5.6 μ l nuclease free water and 2 μ l genomic DNA. PCR reaction was amplified for 20 cycles at 94°C for 30 s, 50°C for 30 s, and 72°C for 45 s, and then a final extension at 72°C for 10 min.

Table 2.5. Primer sequences for nested PCR reactions.

Parasite	Primer name	Forward sequence
genera		
Initial PCR	HaemNFI	5'-CATATATTAAGAGAAZTATGGAG-3'
	HaemNR3	5'-ATAGAAAGATAAGAAATACCATTC-3'
Haemoproteus/	HaemF	5'-ATGGTGCTTTCGATATATGCATG-3'
Plasmodium	HaemR2	5'-GCATTATCTGGATGTGATAATGGT-3'
Leucocytozoon	HaemFL	5'-ATGGTGTTTTAGATACTTACATT-3'
	HaemR2L	5'-CATTATCTGGATGAGATAATGGZGC-3'
GAPDH	FWD	5'-TGTGATTTCAATGGTGACAGC-3'
	REV	5'-AGCTTGACAAAATGGTCGTTC-3'

A 2 μl aliquot of the products of the first reaction was further amplified with the same thermal profile, but for 35 cycles instead of 20 cycles. For *Haemoproteus/Plasmodium* screening, primers HaemF and HaemR2 were used. For *Leucocytozoon* screening, primers HaemFL and HaemR2L were used. Finally, 13 μl of each product from the second PCR rounds were loaded on a 1% agarose gel (0.5 x TBE; 54 g Tris Base, 27.5g boric acid and 4.65 g EDTA in 1 L distilled water), stained with SYBR Safe (ThermoFisher Scientific, California, USA) and visualised under ultraviolet light.

2.12 Quantifying avian malaria infection

To quantify malaria infection in young birds (Chapter Six), a sensitive qPCR protocol newly developed by PhD student Bedur Albalawi (2019) was used. This methodology compares quantities (gene copy number) of a reference bird gene and parasite DNA present in one avian DNA sample, to determine infection intensity using a standard curve created from serial dilution of plasmids. To create the serial dilutions for standard curves, the 18S gene for each of the two parasite genera present in the study system (*Leucocytozoon* and *Haemoproteus*) as well as a reference bird gene (GAPDH), were cloned via bacterial transformation (Albalawi, 2019).

Primers for the qPCR were designed based on the criteria that they were distinguishable between parasite genera *Leucocytozoon* and *Haemoproteus*, but identical among parasite species within each genera, based on alignments from existing sequencing among the study bird population (Table 2.6.). As *Plasmodium* species have not been previously detected in

the study system (Capilla-Lasheras et al., 2017), no primers were designed for *Plasmodium*. For the reference bird gene (GAPDH), existing primer sequences for great tit (Atema et al., 2013) were used. Primers were then tested for cross-reactivity, where samples known to be infected with *Leucocytozoon* were run using *Haemoproteus* primers, and vice versa (Albalawi, 2019).

Table 2.6. qPCR primer sequences used to measure malaria infection intensity.

Target Gene	Primer	Primer sequence
Leucocytozoon	qLeucoF3	5'-GCTTCTATCGGTGAACTCTCGA-3'
18S	qLeucoR2	5'-TATTCTTTGCCTGGAGGTAATG-3'
Haemoproteus	qHaemoF2	5'-AGCTCACGCATCGCTTCTAA-3'
18S	qHaemoR1	5'-ATTTTCTTTGCCTGGAGGTTAC-3'
GAPDH	FWD	5'-TGTGATTTCAATGGTGACAGC-3'
	REV	5'-AGCTTGACAAAATGGTCGTTC-3'

Each DNA sample was run in duplicates in a separate plate for each of the three target genes (Haemoproteus, Leucocytozoon and GAPDH). Serial dilutions of plasmid ($10^4 - 10^0$) was included, as well as two reactions containing nuclease-free water in place of DNA template as negative controls.

The total volume of each qPCR reaction was 15 µl, with 5 µl (12ng/µl) of DNA template. Reactions were run in MicroAmp Optical 96-well plates (0.1 ml, Applied Biosystems) on Stratagene Mx3000 (ThermoFisher), and contained 7.5 µl SYBR Select Master Mix (Applied Biosystems), and primer and nuclease-free water volumes outlined in Table 2.7.

Quality checks were performed on the data and any duplicates from individual birds with >0.5 cycle threshold (Ct) value were repeated. Amplification efficiency of each plate was calculated using: $E = -1 + 10^{(-1/\text{slope})}$, where E = 1 indicates 100% efficiency, and values between 0.9-1.1 representing the acceptable range (Kubista et al., 2006). Following qPCR runs, the relative infection intensity of malaria parasites was calculated for each sample as: $x = 2^{-(\text{parasite Ct} - \text{GAPDH Ct})}$.

Table 2.7. Optimal primer concentrations for each target used in the malaria qPCR assay. From (Albalawi, 2019).

Reagent Target

	Haemoproteus	Leucocytozoon	GAPDH (100nM)
SYBR Select Master Mix volume (Applied Biosystems)	7.5 μΙ	7.5 μ1	7.5 μΙ
Primer concentration (volume)	600/600 nM (1.8 µl)	300/300 nM (0.9 μl)	100/100 nM (0.3 μl)
Nuclease-free water volume	0.7 μ1	1.6 μl	2.2 μ1
Total volume	15 μ1	15 μ1	15 μ1
Annealing temperature (°C)	57 °C	58 °C	60 °C

2.13 QuantiGene® Plex RNA assay

For quantification of transcript levels from RNA samples in Chapters Four, Five and Six, a platform for gene expression analysis, the QuantiGene® Plex RNA assay 2.0 (ThermoFisher), was used. QuantiGene® Plex is used as an alternative to RNA-seq and RT-qPCR, and was chosen for this study due to its ability to provide simultaneous readouts of relative gene expression of multiple target genes from a single sample (Gunawardana et al., 2019; Mills and Gallagher, 2016; Xue et al., 2013).

2.13.1 Overview of QuantiGene® Plex

The QuantiGene® Plex assay combined branched DNA amplification and multi-analyte profiling beads (xMAP®) to detect and quantify RNA targets in samples (Affymetrix, 2015). The workflow of the QuantiGene® Plex assay consisted of several stages, including sample preparation, target hybridisation, signal amplification and signal detection (Figure 2.2.), which took place across two days per assay run. On the first day of assay setup, samples were lysed and incubated overnight along with signal amplifiers. On the second day, there were several wash steps which aided signal amplification, followed by quantification of fluorescence signal of each target.

The QuantiGene® Plex assay worked as follows: custom designed gene probes for the target species were incubated along with blood samples containing RNA and xMAP® beads, and during wash steps on the second day of the assay workflow, created an amplification unit from RNA within the sample. Each amplification unit contained

hybridisation sites for Label Probes, which bound fluorescent protein Streptavidin-conjugated R-Phycoerythrin (SAPE). The fluorescence signal associated with individual Capture Beads was then read on a Luminex[®] MAGPIX flow cytometer and was reported as median fluorescence intensity (MFI). MFI was proportional to the number of target RNA molecules present in the sample (Affymetrix, 2015), and therefore was used as a proxy for gene transcript levels within the sample.

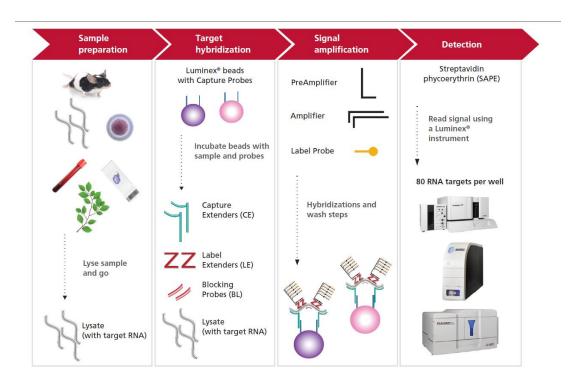


Figure 2.2. QuantiGene® Plex Assay workflow. Diagram obtained from Affymetrix, ThermoFisher (Affymetrix, 2015)

2.13.2 Gene targets

Eighteen probes were commissioned for circadian clock and immune gene targets of interest for great tits (*Parus major*) from ThermoFisher, using sequence information for great tit targets obtained from NCBI database (see Appendix 1). Additionally, probes for three housekeeping gene targets for great tits were commissioned. A list of gene targets, and the chapters in this thesis that they are included in, can be found in Table 2.8. An overview of all gene targets included in this thesis, and their main functions, is shown in Table 2.9.

Table 2.8. A list of gene targets used in this thesis, the gene type (Housekeeping, Clock or Immune targets), and the chapters that they are included in.

Gene target	Type	Ch.4	Ch. 5	Ch. 6
RPL19	Housekeeping	X	X	X
SDHA	Housekeeping	X	X	X
HMBS	Housekeeping	X	X	X
BMAL1	Clock	X	-	X
CLOCK	Clock	X	-	X
CRY1	Clock	X	-	X
CKA	Clock	-	-	X
CK1E	Clock	X	-	X
AANAT	Clock	X	-	X
PER2	Clock	X	-	X
REVERBA	Clock	X	-	X
LY86	Immune	X	X	X
TLR4	Immune	X	X	X
IL1	Immune	X	X	X
IL6	Immune	X	X	X
IGF1	Immune	X	X	X
PRKCA	Immune	X	X	X
GATA3	Immune	X	X	X
IKBA	Immune	X	X	X
<i>NKRF</i>	Immune	X	X	X
NRF2	Immune	X	X	X

Table 2.9 An overview of gene targets used in this thesis and their main functions.

Gene	Gene Type	Function	Citation(s)
RPL19	Housekeeping	Ribosomal Protein L19	(Olias et al., 2014)
		This gene encodes a ribosomal protein of	
CDILA	TT 1	the 60S subunit.	(01' 1 - 2014)
SDHA	Housekeeping	Succinate Dehydrogenase Complex	(Olias et al., 2014)
		Flavoprotein Subunit A	
		This gene is involved in the	
		mitochondrial respiratory chain.	
HMBS	Housekeeping	Hydroxymethylbilane Synthase	(Olias et al., 2014)
		This gene encodes a member of the	
	~	hydroxymethylbilane synthase family	
BMAL1 /	Clock	Brain and Muscle ARNT-Like 1	(Reppert and Weaver,
ARNTL		Transgriptional activator which is a core	2002; Shearman et al., 2000)
		Transcriptional activator which is a core component of the circadian clock.	2000)
CLOCK	Clock	Circadian Locomotor Output Cycles	(Gekakis et al., 1998;
CLOCK	O100H	Protein Kaput	Reppert and Weaver,
		11000m 11mp m	2002; Shearman et al.,
		This gene is involved in the regulation of	2000)
		the circadian clock.	
CRY1	Clock	Cryptochrome Circadian Regulator 1	(Kume et al., 1999;
		m	Reppert and Weaver,
		This gene is involved in the core	2002)
		circadian oscillator. Upregulated by CLOCK/BMAL1 heterodimers.	
CKalpha	Clock	Choline Kinase Alpha	(Reppert et al., 1995)
Cimpin	Clock	Chomic Timase Tilpha	(respect et al., 1995)
		CKa provides a molecular substrate for	
		the actions of melatonin	
CK1E	Clock	Casein Kinase 1 Epsilon	(Steinmeyer et al., 2012)
		This mustain phasphogulates DED	
		This protein phosphorylates PER, involved in the circadian clock.	
AANAT	Clock	Aralkylamine N-Acetyltransferase	(Jin et al., 2011;
			Turkowska et al., 2014)
		Involved in melatonin biosynthesis,	
		controls the day-night rhythm in	
		melatonin production in the vertebrate	
DEVEDDo	Cloals	pineal gland.	(Donnart and Wasser
REVERBa / NR1D1	Clock	Nuclear Receptor Subfamily 1 Group D Member 1	(Reppert and Weaver, 2002; Ueda et al., 2005)
/ INKIDI		Wember 1	2002, Ocda et al., 2003)
		Represses the circadian clock	
		transcription factor BMAL1.	
NKRF	Immune	NF-Kappa-B-Repressing Factor	(Lu et al., 2011)
		This gene mediates transcriptional	
		repression of NFKB responsive genes which control production of cytokines.	
NRF2 /	Immune	Nuclear Factor, Erythroid 2 Like 2	(Kensler et al., 2007; Mills
NFE2L2		1Jour 1 actor, Dry antona 2 Direc 2	and Gallagher, 2016)
· - 		The encoded transcription factor	
		regulates genes which contain	
		antioxidant response elements; many of	
		these genes encode proteins involved in	
		response to injury and inflammation	

		which includes the production of free radicals.	
PER2	Clock	Period Circadian Regulator 2	(Reppert and Weaver, 2002)
		Involved in the core circadian clock loop. Upregulated by CLOCK/BMAL1 heterodimers.	
LY86	Immune	Lymphocyte Antigen 86	(Capilla-Lasheras et al., 2017)
		Involved in the innate immune response to bacterial lipopolysaccharide (LPS) and cytokine production.	
TLR4	Immune	Toll Like Receptor 4	(Akira and Takeda, 2004; Vallance et al., 2017)
		Plays a fundamental role in pathogen recognition and activation of innate immunity.	, ,
IL1	Immune	Interleukin 1 Beta	(Akira and Takeda, 2004; Cuesta et al., 2016)
		An important mediator of the inflammatory response, involved in cell proliferation, differentiation, and apoptosis.	
IL6	Immune	Interleukin 6 Involved in inflammation and the	(Cuesta et al., 2016; Mishra et al., 2019)
IGF1	Immune	maturation of B cells. Insulin Like Growth Factor 1	(Holzenberger et al., 2003;
		Similar to insulin in function, and is involved in mediating growth and development.	Lupu et al., 2001)
PRKCA	Immune	Protein Kinase C Alpha	(Nishikawa et al., 1997)
		Involved in cellular signaling pathways, and serve as receptors for tumour promoters.	
GATA3	Immune	GATA Binding Protein 3	(Capilla-Lasheras et al., 2017; Wang et al., 2011)
		A regulator of T-cell development and plays an important role in endothelial cell biology.	
IKBa /	Immune	NFKB Inhibitor Alpha	(Cabannes et al., 1999)
NFKBIA	minune		
	minuic	Inhibits NFKB, which is involved in inflammatory responses.	

2.13.3 Recommended settings for QuantiGene® Plex RNA assay using avian blood samples

As the QuantiGene® Plex RNA assay had not previously been run for passerine blood samples, assay protocols needed to be optimised for use in this thesis. Initial assay runs carried out using great tit whole blood samples resulted in low florescence signals for all targets, possibly due to the high viscosity of passerine blood preventing probes from binding to target RNA during the overnight incubation step. The next section outlined recommended settings for running of the QuantiGene® Plex RNA assay using passerine blood, after optimisation.

This methodology has been adapted from QuantiGene® Sample Processing Kit and QuantiGene® Plex RNA assay protocols (Affymetrix, 2015). Prior to assay setup, great tit lysates were prepared from blood samples previously frozen at -40°C and thawed on ice. Samples were subject to centrifugation for 3 min at 6000 rpm, and RNA*later*® stabilising solution was carefully removed from each sample using a pouring method and pipetting. Whole Blood Working Lysis Mixture was prepared using 1472 µl lysis mix (pre-warmed at 37°C), 2254 µl nuclease-free water and 92 µl proteinase K. From this, 84 µl of Whole Blood Working Lysis Mixture was added to 12 µl of the blood pellet of each sample in a 1.5 ml eppendorf. Samples were then briefly vortexed and placed in a shaking incubator at 60°C and at 250 RPM for one hour.

QuantiGene® Plex RNA assay setup (Day one)

Assays were prepared in a low-skirted 96 well plate (Invitrogen) and samples were all run in technical duplicates. Working Bead Mix was prepared by combining QuantiGene® Plex kit reagents (Affymetrix, 2015), in the order listed in Table 2.10. The final proportions of this mix were always 33% Lysis Mixture. Capture Beads were taken out of storage right before use and protected from light using foil. Once Capture Beads had been added to the Working Bead Mix, the mix was kept at room temperature to avoid freezing and shattering of beads.

Table 2.10. Reagents contained in Working Bead Mix for QuantiGene® Plex RNA assay.

Order of	Reagent	1 well (μl)	96 wells (μl)
addition			
1	Nuclease-free	2.6	250
	Water		
2	Lysis Mixture	3.3	317
3	Blocking Reagent	1	96
4	Proteinase K	0.1	9.6
5	Capture Beads	0.5	48
	(vortex 30 secs		
	prior to adding)		
6	Probe Set	2.5	240
Total		10 μ1	960 μ1

Each sample was diluted to 1:8 using Diluted Lysis Mixture (one volume of Lysis Mixture, plus two volumes of nuclease-free water, prepared fresh), where 35 μl of Diluted Lysis Mixture was added to each well. Lysed whole blood samples were removed from incubation and subjected to centrifugation at 6000 RPM for one minute. Working Bead Mix was vortexed for 10 seconds, then pipetted 10 μl into each well of the 96 well plate. Finally, 5 μl of the lysed supernatant of each sample was added into wells of the new plate. The total final volume of each well was 50 μl . For background control wells, 40 μl of diluted Lysis Mixture (one volume Lysis Mixture plus two volumes nuclease-free water) were added along with 10 μl Working Bead Mix. Plates were then sealed using a clear Pressure Seal by removal of the backing and placing seal onto the plate. The assay plate was then covered with aluminium foil and placed into a shaking incubator. Plate was incubated overnight at 54°C \pm 1°C at 250 RPM, and any unused lysed samples were placed into storage at -80°C.

QuantiGene® Plex® RNA assay setup (Day two)

On day two of the assay, the plate was removed from incubation and secured onto the Hand-Held Magnetic Plate Washer (Affymetrix) and then carefully unsealed. After one minute of acclimatisation, the assay plate was inverted into the sink to remove any sample not bound to magnetic Capture Beads. Using a multichannel pipette, 50 µl wash buffer (prepared using 300 µl Wash Buffer Component 1, 5 ml Wash Buffer Component 2 and 47.5 ml nuclease-free water) was added to each well. After 30 seconds, the plate was

inverted into the sink again. These wash steps were repeated twice more, for a total of three washes.

Next, 50 μ l Pre-Amplifier solution was added to each well, before plate was resealed, removed from the hand-held magnet and placed into the shaking incubator at 50°C \pm 1°C at 250 RPM for one hour. The three washes and hour of incubation at 50°C \pm 1°C and 250 RPM steps were repeated in the same way, between additions of the following reagents in each well: 1) 50 μ l Amplifier solution; 2) 50 μ l Label Probe solution; 3) SAPE Working Reagent (prepared with 18 μ l SAPE and 6000 μ l SAPE Diluent). The last step was followed by a 30 minute incubation at room temperature and at 250 RPM. Wash steps were then repeated, using SAPE Wash Buffer in place of Wash Buffer. Finally, 130 μ l SAPE Wash Buffer was added to each well and assay plate was shaken on a plate shaker at 750 rpm for 3 min, before immediate reading on the Luminex® MAGPIX instrument.

Extraction and initial analyses of QuantiGene® data

Following reading of the assay plate on the Luminex® MAGPIX, Median Fluorescence Intensity (MFI) data were extracted into Microsoft Excel (2019). Data were visually inspected and assay wells were excluded from further analyses if the total Capture Bead count for that well did not exceed 50, as fluorescence was deemed too low for viable results (Thermofisher Technical Support, personal communication). To normalise fluorescence values, MFI values of each well were divided by the mean of the four background wells for each plate. MFI values were then normalised to housekeeping genes, by dividing the MFI value for each well by the geometric mean of the MFI of the three housekeeping genes (*RPL19*, *SDHA*, *HMBS*). Finally, the coefficient of variation for all technical duplicate wells was calculated. Technical duplicates with a coefficient of variation >35% were excluded from all further analyses, following advice from Thermofisher Technical Support.

Chapter Three: Individual timing of incubation activities in city and forest birds, and implications for offspring fitness

3.1 Introduction

Circadian clocks are internal timekeepers regulating the behavioural and physiological rhythms of organisms with a period length of around 24 hours. Clocks are synchronised by daily changes within an organisms' environment, such as photoperiod and fluctuations in ambient temperature, by a process known as entrainment (Golombek and Rosenstein, 2010).

Individuals often differ in entrainment phase between internal clocks and the synchronising cues within their environment, leading to consistently advanced or delayed rhythms (Roenneberg et al., 2013). These differences in phase can give rise to various circadian phenotypes, where timing of activity differs relative to daylight. This is known as an individual's "chronotype" and is most obviously expressed by individual preference for timing of sleep and activity. "Early" chronotypes usually prefer to sleep earlier than "late" chronotypes (Roenneberg et al., 2003). Chronotype is a highly repeatable personality measure, and is usually classified by the midpoint of sleep (Roenneberg et al., 2003). Alternatively, the daily onset of timing of activity (wake-up time) has been used as a proxy for chronotype (Dominoni et al., 2013).

Individual chronotypes of animals have been documented through captive studies, and include fish, rodent and bird species (Amin et al., 2016; Labyak et al., 1997; Lehmann et al., 2012; Refinetti, 2006). In wild animals, chronotype is harder to study, given that it requires repeatable measurements of individual behaviour in relation to environmental cues, but individual chronotypes have been shown in mammals and birds (Graham et al., 2017; Ocampo-Garcés et al., 2006; Mueller et al., 2012).

For wild animals, many risks and opportunities, such as the presence of predators, food availability, and competition from conspecifics, depend on the time of day. Therefore, differences in individual chronotype may provide wild animals with different consequences for fitness, and may be under selection (Helm et al., 2017). Indeed, emerging evidence suggests that some differences in individual timing may prove advantageous. Experimental studies on passerines have demonstrated that individuals that rise earlier

often have higher reproductive success (Greives et al., 2015; Kempenaers et al., 2010; Poesel et al., 2006), suggesting that for blue tits and great tits, an earlier chronotype is beneficial. Other birds may benefit from extending their activity into the night (Russ et al., 2015). Given that the environmental opportunities as well as risks presented to wild animals are time-dependent, different chronotypes will have varied fitness consequences under different environmental conditions (Helm et al., 2017).

Environmental features that influence individual chronotype are becoming increasingly complex, as human changes to the natural habitat alter the selection pressures incurred on wild animals. Growing evidence suggests that human activity is having substantial effects on wild species' rhythms, including a global shift of activity into the night in response to human disturbance (Gaynor et al., 2018). Moreover, a recent study on the effects of climate change suggested that diurnal mammals may shift their activity into the night, as a buffer for temperature increases (Levy et al., 2019).

One dramatic change to the natural environment by humans is the presence of artificial light at night (ALAN). Exposure to ALAN has significant effects on timing of activity in wild birds, where species living in artificially-lit environments often begin their activity earlier in the morning (Da Silva et al., 2015; Kempenaers et al., 2010). These extensions of activity have also been found in experimental studies of ALAN exposure (de Jong et al., 2016; Raap et al., 2016c). Additionally, the presence of ALAN has been associated with an earlier onset of daily activity in urban birds, such as advanced timing of song in a number of passerine species (Kempenaers et al., 2010). Another study on city blackbirds (*Turdus merula*) noted an advanced onset of activity, but no association between light at night and the end of daily activity (Dominoni et al., 2014).

Differences in individual activity and chronotype arising from human changes to the environment may have fitness consequences for birds (Dominoni et al., 2016). One study on city birds revealed individuals living under ALAN rose earlier, and interestingly, had higher levels of extra-pair paternity than those without light at night in their territory (Kempenaers et al., 2010). It is therefore highly possible that human changes to the natural environment, such as the addition of ALAN, may alter the current selection pressures on activity patterns and individual chronotype in wild animals.

Aside from changes to chronotype, current research suggests that animals living in the city environment may be experiencing restless nights, possibly as a result of increased levels of ALAN (Alaasam et al., 2018; Ouyang et al., 2017). Living in the city environment may lead to disruption of natural circadian rhythms; a captive study revealed city blackbirds to

have less robust rhythms of behavioural activity than their forest conspecifics (Dominoni et al., 2013). Chronodisruption has huge negative implications for health (Dominoni, 2015; Preuss et al., 2008). Wild birds living in city environments therefore may, through chronic chronodisruption, experience a decline in their fitness. Evidently, there is pressing need for studies exploring the effects of the city environment on the activity rhythms, chronotype and fitness measures of wild birds.

One crucial avian life history stage where timing of behavioural activity is important for fitness is incubation, the process by which embryos receive an external supply of thermal energy for development (Deeming, 2002). During incubation, parents must maintain constantly high egg temperatures, as even short bouts of cooling can slow nestling development (Nord and Nilsson, 2011). Longer disruptions to incubation affect nestling prospects (Berntsen and Bech, 2016; Hepp et al., 2015). However, there is large variation in what is considered to be "normal" nest temperatures, due to parents periodically leaving the nest during the day to feed (Webb, 1987; Ar & Sidis, 2002). These nest "off-bouts" are interspersed between longer "on-bouts" of incubation on the nest (Cooper and Voss, 2013). Timing of the parent's first "off-bout" from the nest in the morning marks the onset of daily activity, and the last off-bout marks the end of daily activity (Gwinner et al., 2018). The length of incubation bouts vary, even within a species (Bulla et al., 2016).

During the incubation period, parents may alter behaviours depending on environmental conditions (Gwinner et al., 2018; Hainsworth et al., 1998). In the city environment, conditions are challenging for birds, with the presence of ALAN (Dominoni and Nelson, 2018), ambient noise (Klett-Mingo et al., 2016), pollutants (Isaksson et al., 2009) and reduced quality and availability of food (Pollock et al., 2017). Wild birds may alter their activity rhythms during incubation, to better suit the city environment. However, these changes in incubation patterns may have implications for the fitness of offspring, contributing to the overall decline of bird species associated with urbanisation (Sol et al., 2014).

3.2 Study aims and hypotheses

This study aimed to close the gap in knowledge on the chronotype and urban ecology of wild birds, by comparing daily timing and robustness of incubation activities of wild great tits living in city and forest environments. Furthermore, this study was unique in its aim to investigate the effects of differences in chronotype of wild great tits in city and forest environments on their reproductive success, and subsequent fitness, of offspring.

Increasingly affordable technologies such as thermal sensors and radio-frequency identification (RFID) readers have facilitated studies of incubation rhythms in the wild (Bulla et al., 2016; Graham et al., 2017). Thermal sensors can be placed into nests and behavioural activity can be inferred from the warmth of females while on the nest (Capilla-Lasheras, 2018). Great tits are uniparental intermittent incubators, with the full incubation period lasting for around thirteen days (Gosler, 1993; Haftorn, 1981), and therefore in this study it was possible to measure repeatable behavioural patterns of individual adult females for chronotyping. Individual chronotypes of females were determined by the onset of daily activity as in Dominoni et al., (2013), as morning timing has most relevance for avian ecology (Kacelnik and Krebs, 1983; Kempenaers et al., 2010; Poesel et al., 2006), and is highly repeatable in great tits (Stuber et al., 2015).

Timing of offset of activity was also measured in this study, to determine the duration of daily activity of each bird. Offset of activity was determined as the last daily on-bout to the nest. Nest attendance was measured as the percentage of time a bird is within a nest, assuming the female was in contact with eggs (Skutch, 1962). Lastly, night-time temperature variance was used as a measure for restlessness (Gwinner et al., 2018).

It was predicted that females incubating in the city environment would have less robust rhythms than forest conspecifics, due to the potential effects of ALAN on timing information from natural zeitgebers. This would give rise to differences in daily incubation activities in city birds, such as higher night-time restlessness (from sleep disturbance due to ALAN) and lower percentage of time spent in nest box. Furthermore, city birds were predicted to have an earlier onset of activity (chronotype), later offset of activity, and longer duration of daily activity.

Parameters of offspring fitness used in this study were hatching success, fledging success and nestling weight at day thirteen. The weight of nestlings was used as a fitness proxy given that heavier fledglings have a better chance of survival during the post-fledgling

period compared with lighter fledglings (Perrins, 1965). It was predicted that overall, city nestlings would have lower hatching success, weight and fledging success than forest counterparts as in previous studies in this system (Pollock et al., 2017). An association between chronotype of the mother, and the weight, hatching, and fledging success of her nestlings was expected. Given that in city environments, males with earlier chronotypes have higher reproductive success (Kempenaers et al., 2010), it was expected that nestlings whose mothers have an early chronotype would weigh more than those whose mother had a later chronotype. Likewise, it was expected that this effect was stronger in the city, where pressure to be early appears to be particularly high (Kempenaers et al., 2010; Dominoni et al., 2013). Ultimately, it was predicted that within each habitat, early birds relative to the others within the habitat do well and have better reproductive success.

3.3 Materials and Methods

3.3.1 Study sites

Data for this chapter were obtained during the breeding seasons (May-June) of 2016, 2017 and 2018 in free-living populations of nesting great tits. Field sites were existing forest nest box study systems located at the Scottish Centre for Ecology and the Natural Environment (SCENE; 56° 7.73'N, 4° 36.79'W), Cashel Forest (56°6'N, 4°34'W), and city sites at Kelvingrove Park, Glasgow (55°52' N, 4°17'W) and Garscube estate (55° 9'N, 4°31' W). All sites have been previously used in ecological studies of wild passerines (Pollock et al., 2017; Capilla-Lasheras et al., 2017). See General Methods for more information on study sites.

3.3.2 iButton temperature sensors

iButtons (DS1922L-F5, Thermochron) are small, cylindrical sensors using 1-Wire [®] technology that can be pre-programmed to take temperature readings at equal intervals. In this study, iButtons were pre-programmed at a resolution of 0.0625 ^oC and a sampling interval of three minutes to record within nest temperatures. Pre-programmed iButtons were covered with a small piece of material and wrapped together with a thin wire before insertion into great tit nests (Figure 3.1.). iButtons were inserted into nests containing three or more eggs prior to, or as close as possible to the start of, the incubation period. iButtons were placed carefully next to eggs and the wire thread through to the bottom of the nest and attached to a small 28.35 g fishing weight to secure its position within the box.



Figure 3.1. iButton setup for incubation data collection. 3.1.1. Pre-programmed iButton wrapped in material and attached to wire, 28.35 g fishing weight, label containing serial number, date and target nest box. 3.1.2. iButton positioned in nest, alongside eggs.

Additionally, two iButtons were pre-programmed to record ambient temperature data at each field site with a resolution of 0.5 °C and sampling interval of five minutes, for direct reference to in-nest recordings during iButton data analyses. These iButtons were attached to the outside of nest boxes by taping them to the bottom of the box.

In years 2016 and 2017, replacement of iButtons took place weekly (due to limited data storage capacity on individual iButtons) to ensure that no gaps were left in the recordings. In 2018, iButtons were programmed to roll-over once data capacity was reached, and weekly readings were taken in the field using a TDHC Data Downloader (Thermodata Corporation, Milwaukee, WI). iButton data were downloaded as .csv files and read using OneWire Viewer software (Maxim Integrated, California). Sample sizes for the study are shown in Table 3.1.

Table 3.1. Sample sizes for total number of days of iButton recordings (and total number of nest boxes) per site and study year

	Site		
Year	City	Forest	
2016	14 (4)	218 (30)	
2017	104 (11)	203 (34)	
2018	89 (12)	104 (12)	

C!4 -

3.3.3 MET Office data

Environmental temperature data for Daily Mean Temperature (°C) were obtained for each year from the UK Met Office for an area nearby to forest field sites, Tyndrum (56°25'39.57"N, 4°42'18.20"W), and nearby to city sites, Bishopton (55°54'24.75"N, 4°30'11.20"W), to incorporate into statistical models.

3.3.4 Nestling processing

All nestlings (N = 253) were ringed for individual identification on day thirteen of life. Measurements were taken of nestling weight in all study years. To determine hatch dates, brood sizes and fledging success, all nest boxes were checked weekly during the breeding

season as part of a general nest box monitoring protocol (see General Methods for further details of nest box monitoring).

3.3.5 Ethical statement

See General Methods for details of ringing licences.

3.3.6 Data analyses

Data analyses were carried out using R v. 3.4.4 (R Core Team, 2018). All statistical analyses were performed using packages *lme4* (Bates et al., 2014). Initial global models for response variables were linear-mixed models (LMM) with Gaussian error distribution. All global models included individual nest box as a random factor to account for repeated measures within the same nest box. Residuals were plotted and inspected to check assumptions of normality. Pairs of explanatory variables were also assessed for collinearity by calculating tolerance and variance inflation factors (VIF) and by visual inspection of pairwise plots, before statistical relevance was accepted.

Statistical relevance of fixed factors within all models was determined by dropwise model selection via Likelihood Ratio Testing using package *lmtest* (Zeileis & Hothorn, 2002), until the minimum adequate model (MAM) was found. Factors with a significance level of <5% were kept in the model. Site was always kept within the MAM regardless of significance.

3.3.7 Incubation timing analysis

Prior to statistical modelling, data recordings from iButtons were visualised using ggplot2 (Wickham, 2016). During data collection, some iButtons were pushed to the side of the nest box and did not successfully record within nest temperatures. Therefore, recordings that were identified by visual inspection to closely follow the pattern of environmental temperatures at the field site were removed from the dataset. This was done blind of site where recordings were taken. Analyses were restricted to up to sixteen days prior to the hatch date, and also after all eggs had been laid, as sporadic incubation behaviour may begin prior to clutch completion (Marasco and Spencer, 2015). In total, 732 days of iButton recordings were included in the final analysis, where 207 were taken from nest

boxes at urban sites and 525 from nest boxes at forest sites, using some data as a control group for another experiment in McGlade et al., (2020, *in prep*).

An R package, incR (Capilla-Lasheras, 2018; Gwinner et al., 2018), was used to determine parameters of incubation using iButton data. The incR package uses gradients of temperatures between the nest (around 35 °C in incubating great tits) and the environment (between 5 and 15 °C) to predict if a bird is currently sitting on the nest (Capilla-Lasheras, 2018). Using the incR algorithm, several parameters of incubation behaviour were determined for each nest box: first off-bout, last on-bout, duration of active day, and the percentage of time spent on the nest on a given day. To account for differences in photoperiod across the breeding season, activity onset (relative to sunrise) for each bird was calculated for each day of recordings by subtracting the timing of sunrise from the timing of the first off-bout on that day. Similarly, activity offset was determined for each day of recordings by subtracting the timing of the last on-bout that day. Timing information for sunrise and sunset at SCENE (56° 7.73'N, 4° 36.79'W) and Glasgow (55°52'N, 4°15'W) were obtained from www.timeanddate.com.

Duration of daily activity was calculated for each bird as the daily offset of activity subtracted by daily onset of activity. Percentage of time in nest box over a 24 hour period was calculated as the total time a bird was predicted by incR to be on the nest and in contact with eggs (Gwinner et al., 2018). Night-time nest temperature variance was used as a determinant for restlessness during incubation, as in Gwinner et al., (2018). Night-time restlessness was calculated from 22:00 – 03:00 GMT for each 24-hour period, with the assumption that all birds were within nest boxes during this time period.

Explanatory variables within initial global models for incubation activity include site as a two-level fixed factor (city/forest), the number of days before hatching, clutch size as the maximal number of eggs within a nest, recording date in April day format (where April 1st = day 1), and study year (2016/2017/2018). Mean ambient temperature was also included in global models as low temperatures may affect incubation bout length (Conway & Martin, 2000). The number of days before hatching was included in all models, both as a linear term and as a quadratic, as in (Gwinner et al., 2018). Additionally, interactions between the number of days before hatching and site were included in all models.

3.3.8 Analyses of nestling fitness data

To test whether chronotype predicts fitness of offspring, the mean activity onset of the mother was used as a proxy for chronotype. Then, chronotype of the mother was used as a predictor of nestling hatching success, weight at day thirteen, and fledging success. Nestling weight data were modelled using a Gaussian error distribution, with explanatory variables for study year (2016/2017/2018), hatch date (where April 1st = day 1), brood size, site (city/forest) and mother's chronotype (mean activity onset). Hatching success and fledging success data were modelled using Binomial error distributions, with additional explanatory variables for study year (2016/2017/2018), clutch size (brood size for fledging success), and mother's chronotype (mean activity onset). Nest box was included in all models as a random factor. An interaction was also included between site and chronotype in all models. Statistical relevance of fixed factors within all models was again determined by dropwise model selection via Likelihood Ratio Testing using the package *lmtest* (Zeileis & Hothorn, 2002).

3.4 Results

3.4.1 Timing of incubation activities

Over the duration of the study, onset of daily activity (relative to sunrise) ranged from 03:42 until 07:14. Onset of daily activity was not affected by the clutch size, recording date, mean ambient temperatures or study year (Table 3.3). The onset of daily activity became progressively later as incubation period progressed closer to hatching ($t_{2,732} = -2.15$; P=0.024, LMM), (Figure 3.2.). This trend was stronger in forest birds than in city birds, as indicated by the significant interaction between site and days before hatching (Figure 3.2.; Table 3.3). Forest birds had an overall later onset of activity than city birds, where forest birds were estimated to leave their nest box $10 \pm 3:55$ minutes later than city birds (Table 3.3.).

Table 3.2. Minimum adequate models after likelihood ratio testing using incubation activity data. The global model for all incubation activity data was: Nest box + site + site*days before hatching² + site*days before hatching + days before hatching² + days before hatching + clutch size + mean temperature + recording date + study year. Nest box was included in all models as a random factor. Data were modelled using Gaussian error distribution. An interaction between terms is denoted by an asterisk (*). Addition of terms is denoted by a plus sign (+).

Variable	Minimum Adequate Model (MAM)
Onset of activity (relative to sunrise)	Nest box + site + site*days before hatching² + site*days before hatching + days before hatching² + days before hatching + year
Offset of activity (relative to sunset)	Nest box + site + site*days before hatching² + site*days before hatching + days before hatching² + days before hatching + recording date + clutch size + year
Duration of active day	Nest box + site + site*days before hatching² + site*days before hatching + days before hatching² + days before hatching + year
Percentage time in box	Nest box + site + site*days before hatching² + site*days before hatching + days before hatching² + days before hatching + year
Night time restlessness	Nest box + site + site*days before hatching + days before hatching + year

The offset of daily activity (relative to sunset) over the duration of the study ranged from 17:03 until 22:01. There were no effects of mean ambient temperatures or study year on offset of activity (Table 3.3). Offset of activity became progressively earlier as incubation period came closer to hatch date ($t_{2,708} = 4.07$; P<0.001, LMM), and this trend was stronger in forest birds than in city birds as indicated by the significant interaction between site and days before hatching (Figure 3.2.; Table 3.3). Forest birds had an earlier offset of activity

than city birds, where forest birds were estimated to retire into nest boxes 10.35 (s.e. \pm 06:59) minutes earlier than city birds (Table 3.3).

Offset of activity also was significantly affected by date of recording, where offset of activity was later as the season progressed ($t_{1,708} = -6.14$, P<0.001, LMM), (Table 3.3).

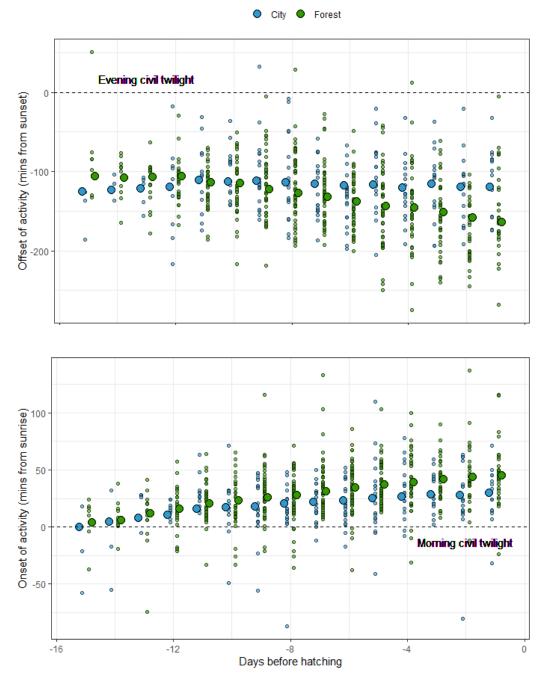


Figure 3.2. Offset (top) and onset (bottom) of activity in city (blue) and forest (green) birds, relative to sunrise and sunset and the number of days before hatching. Large dots represent the mean minimum adequate model predictions of activity onset/offset for that day. Dotted lines represent timing of morning and evening civil twilight, the time of day when the sun first rises in the morning or last sinks in the evening, respectively, below -6 or relative to the horizon.

Duration of activity across the study ranged from 660 minutes (11 hours) to 1011 minutes (16 hours 51 minutes). Duration of activity was not associated with mean ambient temperature, clutch size, recording date or study year (Table 3.3). As with onset and offset of activity, duration of activity was associated with the number of days before hatching ($t_{1,708}$ = -6.03; P<0.001, LMM). The observed effect, a shortening of the active day closer to hatching, was stronger in forest birds than city birds as indicated by the significant interaction between site and days before hatching and steeper drop for forest birds (Figure 3.3.; Table 3.3). During early incubation, the active day was shorter in city birds, but this difference tended to reverse close to hatching (Figure 3.3.).

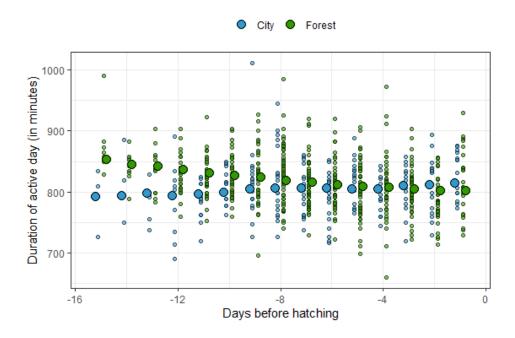


Figure 3.3. Duration of active day (minutes) and number of days before hatching in city (blue) and forest (green) birds. Large dots represent the mean minimum adequate model predictions of duration of activity for that day.

The percentage time within nest box ranged over the duration of the study from 40 % to 98.89 %. There was no effect of mean ambient temperature or recording date on percentage of time spent in nest box (Table 3.2.). Percentage time within nest box increased significantly as incubation progressed closer to hatching (linear: $t_{1,732} = 19.76$; P<0.001, LMM; quadratic: $t_{1,732} = -17.85$; P=0.006, LMM), (Figure 3.4.). During early incubation, city birds spent more time in nest boxes than forest birds, but this difference tended to disappear close to hatching (Figure 3.4.). When compared with 2016, the percentage of time within nest box was estimated to be 0.88 % (\pm 0.88) higher in 2017 and -2.02 % (\pm 1.05) lower in 2018 (Table 3.3). Moreover, females with a larger clutch size spent significantly longer in nest boxes than those with smaller clutches ($t_{1,732} = 2.00$; P=0.041, LMM; Table 3.3).

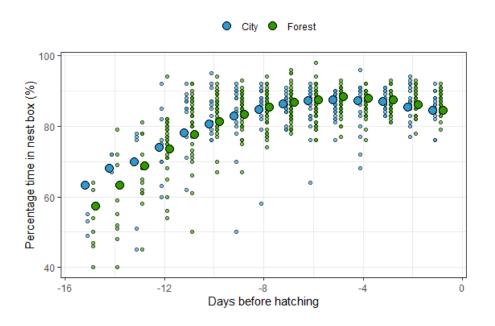


Figure 3.4. The percentage time in nest box (minutes) and number of days before hatching for city (blue) and forest (green) birds. Large dots represent the mean minimum adequate model predictions of percentage time spent in nest box for that day.

Night time nest temperature variance ranged over the duration of the study from 0.028 °C to 7.66 °C. There was no effect of hatch date or clutch size on night time temperature variance (Table 3.3). Forest birds had significantly higher levels of night time temperature variance than city birds ($t_{1,629} = 2.64$; P=0.008, LMM), (Table 3.3). In addition to site, night time temperature variance was also significantly affected by mean ambient temperatures, date of recording and study year. When compared with 2016, night time temperature variance was estimated to be 0.14 °C (\pm 0.16) lower in 2017 and 0.63 °C (\pm

0.18) higher in 2018 (Table 3.3). Night time temperature variance also significantly decreased as the season progressed ($t_{1,629} = -3.73$; P=0.002, LMM), but there was no interaction with site (Table 3.3).

Global model summaries for incubation activity data can be found in Table 3.4.

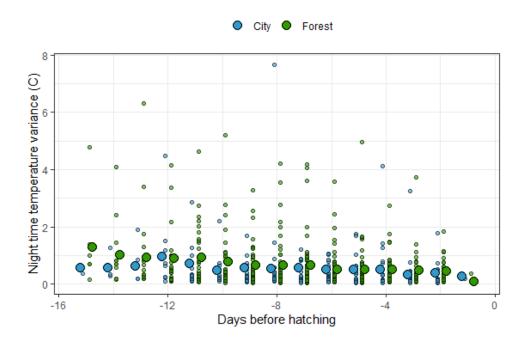


Figure 3.5. Night time temperature variance (restlessness) and the number of days before hatching for city (blue) and forest (green). Large dots represent the mean minimum adequate model predictions of night time temperature variance for that day.

Table 3.3. Minimum adequate model summaries from linear mixed models of incubation activity data. p-values shown here were obtained from likelihood ratio testing. Reference levels for intercept were: Year (2016), Site (City). An interaction between terms is denoted by an asterisk (*).

Variable	N	Estimate	s.e.m	d.f.	<i>t</i> -value	<i>p</i> -value LRT
Onset of activity	732					
(Intercept)		39.61	3.55		11.16	
Site (Forest)		10.35	3.55	1	2.91	
Days before		2.87	0.27	1	10.57	
hatching						
Days before		-41.70	19.41	2	-2.15	0.024
hatching ²						
Site(City)*Days		-91.79	46.67	2	-1.96	0.037
before hatching						
J						
Offset of activity	708					
(Intercept)		-23.39	21.28		-1.10	
Site(Forest)		10.35	6.98	1	1.48	
Days before		-1.54	0.70	1	-2.20	
hatching						
Date		-2.87	0.47	1	-6.14	< 0.001
Site(City)*Days		342.69	84.12	2	4.07	<0.001
before hatching						
D 4' 6	700					
Duration of	708					
active day		5 00 51	< 27		105.00	
(Intercept)		783.51	6.27		125.02	
Site(Forest)		11.58	5.81	1	1.99	
Days before		-3.38	0.56	1	-6.03	
hatching						
Site*Days before		456.2	95.13	2	4.80	< 0.001
hatching						
	=					
Percentage time	732					
in box		00.60	1.72		51.00	
(Intercept)		89.69	1.73		51.82	0.442
Site(Forest)		-0.68	0.90	1	-0.75	0.442
Days before		1.30	0.07	1	19.76	
hatching					. =	
Days before		-101.79	5.70	2	-17.85	
hatching ²						
Clutch size		0.48	0.24	1	2.00	0.041
Year(2017)		0.88	0.88	1	0.10	0.006
Year(2018)		-2.02	1.05	1	-1.92	0.006
Night time	629					
temperature						
variance						
(Intercept)		1.79	0.37		4.86	
Site(Forest)		0.44	0.17	1	2.64	0.008
Mean		-0.03	0.01	1	-2.54	0.012
Temperature						
Date		-0.03	0.01	1	-3.73	0.002
Year(2017)		-0.14	0.16	2	-0.85	<0.001
Year(2018)		0.63	0.18	2	3.41	<0.001

Table 3.4. Global model summaries from linear mixed models of incubation activity data. p-values shown here were obtained from likelihood ratio testing. Reference levels for intercept were: Year (2016), Site (City). An interaction between terms is denoted by an asterisk (*).

Variable	N	Estimate	s.e.m	d.f.	<i>t</i> -value	<i>p</i> -value
Onset of activity	732	Littinate	3.6.111	u.i.	ı varuc	p value
(Intercept)	, 32	-2.50	15.63		-0.16	
Site (Forest)		6.37	4.67	1	1.36	
Days before		1.32	0.51	1	2.60	
hatching						
Days before		-61.39	36.01	2	-1.70	
hatching ²						
Clutch size		0.20	1.01	1	0.20	0.843
Mean temp		0.34	0.35	1	0.97	0.333
Date		0.56	0.30	1	1.83	0.061
Year (2017)		9.72	3.90	1	2.49	0.038
Year (2018)		4.12	4.50	1	0.91	0.038
Site*Days before		105.26	48.01	2	2.19	
hatching						
Site*Days before		28.77	43.16	4	0.67	0.051
hatching ²						
Offset of activity	708					
(Intercept)		3.67	26.30		0.14	
Site(Forest)		6.67	7.71	1	0.86	
Days before		1.77	0.91	1	1.95	
hatching						
Days before		-73.98	63.73	2	-1.16	
hatching ²						
Clutch size		-1.81	1.66	1	-1.10	0.259
Mean temp		-0.46	0.64		-0.71	0.479
Date		-2.51	0.51	1	-4.91	<0.001
Year (2017)		5.80	6.39	1	0.91	0.133
Year (2018)		-7.98	7.41	1	-1.08	0.133
Site*Days before		-350.69	86.70	2	-4.04	
hatching		1.12.02	== 10		4.0.5	0.04
Site*Days before		143.02	77.10	4	1.86	0.067
hatching ²						
D	722					
Percentage time in box	732					
(Intercept)		86.25	3.86		22.32	
Site(Forest)		-1.40	1.11	1	-1.26	
Days before		1.08	0.14	1	7.73	
hatching		1.00	0.14	1	1.13	
Days before		-94.26	10.58	2	-8.91	
hatching ²		-74.20	10.50	2	-0.71	
Clutch Size		0.47	0.23		2.01	0.039
Mean Temp		-0.11	0.10		-1.11	0.273
Date		0.09	0.08		1.13	0.225
Year(2017)		1.42	0.92	1	1.55	0.004
Year(2018)		-1.90	1.05	1	-1.80	0.004
Site*Days before		18.58	13.84	2	1.34	3.00.
hatching						
Site*Days before		-9.85	12.66	4	-0.78	0.423
hatching ²			-		-	

Variable	N	Estimate	s.e.m	d.f.	<i>t</i> -value	<i>p</i> -value
Duration of	708					
active day						
(Intercept)		862.64	27.79		31.04	
Site(Forest)		13.48	8.01	1	1.68	
Days before		1.90	1.01	1	1.88	
hatching						
Days before		-11.53	73.18	2	-0.16	
hatching ²						
Clutch Size		-1.42	1.70		-0.84	0.386
Mean Temp		-0.67	0.73		-0.92	0.366
Date		-0.45	0.54		-0.83	0.393
Year(2017)		-10.34	6.59	1	-1.57	0.170
Year(2018)		-12.28	7.63	1	-1.61	0.170
Site*Days before		-479.22	98.45	2	-4.87	
hatching						
Site*Days before		104.70	88.43	4	1.18	0.240
hatching ²						
Night time	629					
temperature						
variance						
(Intercept)		2.89	0.63	1	4.57	
Site(Forest)		0.62	0.19	1	3.20	
Days before		0.04			3.20	
-		0.04	0.02	1	1.97	
hatching						
•		0.04	0.02 1.21	1 2		
hatching					1.97	
hatching Days before					1.97	0.058
hatching Days before hatching ²		0.04	1.21	2	1.97 0.03	0.058 0.002
hatching Days before hatching ² Clutch Size		0.04	1.21 0.04	2	1.97 0.03 -1.85	
hatching Days before hatching ² Clutch Size Mean Temp		0.04 -0.08 -0.04	1.21 0.04 0.01	2 1 1	1.97 0.03 -1.85 -3.03	0.002
hatching Days before hatching ² Clutch Size Mean Temp Date		0.04 -0.08 -0.04 -0.04	1.21 0.04 0.01 0.01	2 1 1 1	1.97 0.03 -1.85 -3.03 -2.97	0.002 0.003
hatching Days before hatching ² Clutch Size Mean Temp Date Year(2017)		0.04 -0.08 -0.04 -0.04 -0.15	1.21 0.04 0.01 0.01 0.16	2 1 1 1 1	1.97 0.03 -1.85 -3.03 -2.97 -0.90	0.002 0.003 <0.001
hatching Days before hatching ² Clutch Size Mean Temp Date Year(2017) Year(2018)		0.04 -0.08 -0.04 -0.04 -0.15 0.72	1.21 0.04 0.01 0.01 0.16 0.19	2 1 1 1 1 1	1.97 0.03 -1.85 -3.03 -2.97 -0.90 3.83	0.002 0.003 <0.001
hatching Days before hatching² Clutch Size Mean Temp Date Year(2017) Year(2018) Site*Days before		0.04 -0.08 -0.04 -0.04 -0.15 0.72	1.21 0.04 0.01 0.01 0.16 0.19	2 1 1 1 1 1	1.97 0.03 -1.85 -3.03 -2.97 -0.90 3.83	0.002 0.003 <0.001

3.4.2 Nestling fitness

Nestling weight was not significantly affected by brood size, hatch date, or study year (Table 3.5.). There was no significant interaction between site and chronotype of the mother (Figure 3.6), and effect sizes of later chronotype on nestling weight were very small (0.004 g \pm 0.007), (Table 3.6). However, effects of site on nestling weight were evident, where forest nestlings weighed significantly more than nestlings at city sites ($t_{1,231}$ = 4.80; P<0.001, LMM), (Table 3.6). Forest nestlings were estimated to weigh 18.99 g (\pm 0.27) and city nestlings 17.70 g (\pm 0.26), (Table 3.6.).

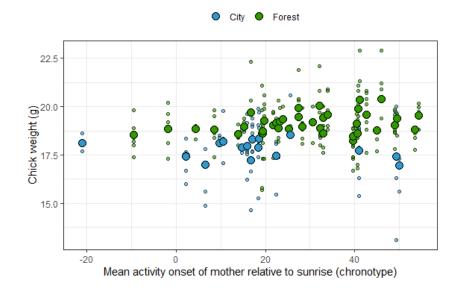


Figure 3.6. Nestling weight (g) at day thirteen and mean activity onset of mother (chronotype), for city (blue) and forest (green) nestlings. Large dots represent the minimum adequate model predictions of maternal chronotype for each nest box.

Hatching success overall was 80.96 % across the study period and all nest boxes. Hatching success was not affected by the study year, clutch size or the hatching date (Table 3.6.). In addition, hatching success was not significantly affected by chronotype of the mother $(z_{1.394} = -1.37; P=0.170, LMM)$, or the site $(z_{1.394} = 0.45; P=0.650, LMM)$.

Fledging success overall was 95.64 % across the study period and all broods. As with hatching success, fledging success was not significantly affected by study year, brood size, or the hatching date (Table 3.6.). Fledging success was also not significantly affected by chronotype of the mother ($z_{1,321} = 9.19$; P=0.755, LMM), or site ($z_{1,321} = 250.16$; P=0.768, LMM).

Global model summaries for nestling fitness traits can be found in Table 3.7.

Table 3.5. Global and minimum adequate models after likelihood ratio testing using nestling data. Nestling weight at day thirteen was modelled using Gaussian error distributions. Fledging success and hatching success data were modelled using Binomial error distributions. Nest box was included in all models as a random factor.

Variable	Global Model	Minimum Adequate Model (MAM)
Nestling weight	Nest box + chronotype*site + chronotype + site + year + hatch date + brood size	Nest box + chronotype*site + chronotype + site + hatch date
Hatching success	Nest box + chronotype*site + site + chronotype + clutch size + hatch date + year	Nest box + chronotype + site
Fledging success	Nest box + chronotype + site + brood size + hatch date + year	Nest box + chronotype + site

Table 3.6. Minimum adequate model summaries of nestling weight, hatching success and fledging success. Nestling weight data were modelled using Gaussian error distribution. Hatching success and Fledging success data were modelled using Binomial error distributions. The p-values shown here were obtained from likelihood ratio testing. Reference levels for intercept were: Site (City). An interaction between terms is denoted by an asterisk (*).

Variable	N	Estimate	s.e.m	d.f.	t-value	<i>p</i> -value LRT
Nestling weight	231					
(Intercept)		17.70	0.26		67.79	
Chronotype		0.004	0.007	1	0.63	0.840
Site (Forest)		1.29	0.27	1	4.80	< 0.001
	N	Estimate	s.e.m	d.f.	z-value	<i>p</i> -value LRT
Hatching success	394					
(Intercept)		2.01	0.42		4.78	
Chronotype		-0.02	0.01	1	-1.37	0.170
Site (Forest)		0.19	0.42	1	0.45	0.650
Fledging success	321					
(Intercept)		6.15	0.01		3416.93	
Chronotype		0.02	0.01	1	9.19	0.755
Site (Forest)		0.45	0.01	1	250.16	0.768

Table 3.7. Global model summaries of nestling weight, hatching success and fledging success. Nestling weight data were modelled using Gaussian error distribution. Hatching success and Fledging success data were modelled using Binomial error distributions. The p-values shown here were obtained from likelihood ratio testing. Reference levels for intercept were: Site (City). An interaction between terms is denoted by an asterisk (*).

Variable	N	Estimate	s.e.m	d.f.	<i>t</i> -value	<i>p</i> -value LRT
Nestling weight	231					
(Intercept)		20.75	1.58		13.09	
Chronotype		-0.01	0.01	1	-0.77	
Site (Forest)		1.33	0.53	1	2.50	
Year(2017)		-0.57	0.34		-1.70	0.120
Year(2018)		-0.11	0.32		-0.34	0.120
Hatch date		-0.05	0.03	1	-1.50	0.115
Brood Size		-0.08	0.07		-1.08	0.247
Chronotype*Site		0.02	0.02	2	1.05	0.253
	N	Estimate	s.e.m	d.f.	z-value	<i>p</i> -value
						LRT
Hatching success	394					
(Intercept)		5.23	2.21		2.37	0.018
Site(Forest)		1.87	0.75	1	2.48	0.013
Chronotype		0.01	0.02	1	0.62	0.534
Clutch Size		-0.07	0.11	1	-0.63	0.530
Hatch Date		-0.08	0.05	1	-1.73	0.083
Year(2017)		0.04	0.49	1	0.09	0.930
Year(2018)		0.29	0.42	1	0.68	0.493
Site*Chronotype		-0.04	0.02	2	-1.54	0.123
Fledging success	321					
(Intercept)		5.23	2.21		2.37	0.018
Site (Forest)		1.87	0.75	1	2.48	0.013
Chronotype		0.01	0.02	1	0.62	0.534
Clutch Size		-0.07	0.11	1	-0.63	0.530
Hatch Date		-0.08	0.05	1	-1.73	0.083
Year(2017)		0.04	0.49	1	0.09	0.930
Year(2018)		0.29	0.42	1	0.68	0.493
Site*Chronotype		-0.04	0.02	2	-1.54	0.123

3.5 Discussion

3.5.1 Incubation patterns

In this Chapter, activity patterns of free-living incubating great tits were compared for city and forest habitats. There were strong effects of the number of days before hatching on the onset, percentage of time in the nest box and night time variance, as mothers spent more time in the nest box the closer to hatching date. Similar changes in parental behaviour during critical points of the embryonic developmental period have been shown in previous studies on incubation, such as a later onset of activity (Graham et al., 2017), and lower variance in night time temperatures (Gwinner et al., 2018) as hatch date becomes closer. This trend for changes in behaviour across the incubation period was stronger in the city site than the forest, as shown by significant interactions of site and the number of days before hatching in statistical models (Table 3.4.). This is evident in Figure 3.2., where city birds appeared to "break off" from forest birds halfway through the incubation period, starting their daily activity earlier and retiring later than forest birds. In addition, there was a shortening of active day in forest birds across the incubation period; a trend which was stronger than in city birds (Figure 3.3.).

As predicted, great tits incubating at city sites had earlier onsets of daily activity, and later offsets of daily activity overall, than forest birds (Figure 3.2.). These differences are consistent with previous studies of activity patterns of birds and urbanisation, with city birds rising earlier (Dominoni et al., 2013; Da Silva, Valcu and Kempenaers, 2015) and staying out later (Russ et al., 2015) than forest conspecifics. Curiously, earlier onset and later offset of activity for city birds did not translate to longer duration of active day for city birds in model predictions (Figure 3.3.). This result may require further analyses.

A possible cause for these observed site differences in timing of incubation activities may have been the high level of ALAN at city sites. Environmental light values extracted from a satellite map (VIIRS, 2019; https://www.lightpollutionmap.info) using global positioning system (GPS) co-ordinates of nest box locations used in the study revealed that mean light levels at city nest boxes were 55 lux, and forest nest boxes 0.44 lux. Previous research investigating the effects of ALAN on birds show that under artificially lit areas, birds are active earlier (Dominoni et al., 2014; Kempenaers et al., 2010; Kumar et al., 2018) and sometimes extend their activity later (Russ et al., 2015) than forest birds; trends which are reflected in the activity patterns of city birds in this study. The mechanisms behind these shifts in activity are currently unknown, but may be due to circadian disruption (de Jong et

al., 2018), or perception of longer daylength as shown for ALAN of as little as 0.3 lux (Dominoni et al., 2013a; Dominoni and Partecke, 2015).

In addition to ALAN, ambient noise in urban areas also has potential to disrupt normal activity patterns of wild birds (Slabbekoorn and Peet, 2003; Fuller, Warren and Gaston, 2007; Gil et al., 2015), although other authors argue that light is more important predictor of activity than noise (Dominoni et al., 2014). For this study, light and noise levels at city and forest sites were not included in statistical models but may have played a role in the observed significant differences between activity patterns of birds at the two habitats. Inclusion of environmental information on light and noise in further studies would help to untangle these effects.

It was predicted that rhythms in city birds would be less robust than in forest birds, and therefore a higher level of night time restlessness would be observed in city birds. This was predicted as per a previous captive study (Dominoni et al., 2013) and from experimental studies of wild birds (Ouyang et al., 2017; Raap et al., 2016b). Contrary to expectation, night time restlessness was higher in forest birds than in city birds (Figure 3.5.), possibly due to the abundance of nocturnal predators in forest sites such as weasels, martens (*Martes martes*) and owls (*Strix aluco*), (Dunn, 1977; Stuber et al., 2014). Although the observed differences in variability of egg temperature were small (0.18 °C), this may be ecologically significant, as even small changes in temperature of embryos during incubation can slow development and affect nestling condition (Ardia et al., 2010; Berntsen and Bech, 2016).

Previous studies have shown ALAN exposure can disrupt sleep in wild birds (Raap et al., 2016c, 2015). Future wild studies measuring activity patterns in city birds ought to measure markers of sleep deprivation (Ouyang et al., 2017), or nocturnal melatonin concentrations (Dominoni et al., 2013) to find out whether city birds are indeed sleep deprived. Moreover, in-nest cameras could be used to quantify sleep behaviours as an addition to incubation patterns (Stuber et al., 2015).

A limitation of this study was the omission of day of year from statistical models of incubation activities. Changes in the environment across the season, such as an increase in day length, warmer temperatures (Avery & Krebs 1984; Kendeigh et al., 1969) and decreased rainfall (Radford et al., 2001) may affect timing of foraging activities. Given that city birds often lay earlier in the year than forest birds (Deviche & Davies 2014), these seasonal changes may have contributed to the observed differences between activity

patterns of city and forest birds. Inclusion of day of year in future analyses would therefore act as a control for differences in phenology between forest and city birds.

3.5.2 Chronotype and condition of offspring

This study was unique in investigating whether individual differences in timing could predict fitness traits of wild birds, by linking chronotype of the mother to hatching success, fledging success and nestling weight. Contrary to expectation, there were no effects of mother's chronotype on the parameters of chick fitness used in this study. Chronotype of the mother also had no significant effect on either hatching success or fledging success of nestlings, although both were higher at forest sites than city sites.

Many limiting resources within a birds' environment fluctuate with time of day (Helm et al., 2017), such as the availability of food (Bednekoff and Houston, 1994). Previous research carried out at the sites used in this study showed that food is limited in the city habitat (Pollock et al., 2017). Therefore, having an earlier chronotype may have provided a foraging advantage for great tits living in the city, both during incubation and nestling provisioning.

In this study, nestlings raised in the city habitat weighed less than those at forest sites, possibly due to this scarcity of food, and in accordance with previous research undertaken at these sites (Capilla-Lasheras et al., 2017; Pollock et al., 2017). Observed differences in nestling weight between sites may have ecological significance, as heavier nestlings have better prospects for survival and recruitment into the population post-fledging (Ringsby et al., 1998). Future studies might consider re-capturing study nestlings as adults, to investigate whether birds whose mother had an earlier chronotype are in better condition than those with later chronotypes.

3.5.3 Conclusions

This study provided no evidence to support the idea that differences in individual timing and chronotype have consequences for condition of offspring in wild birds. However, activity patterns of individual birds differed between city and forest habitats used in this study, which may have been due to the disruptive presence of ALAN on biological rhythms. In city environments where resources are already limited (Pollock et al., 2017), these differences in timing may give early individuals the edge for better survival of chicks or survival prospects post-fledging.

Chapter Four: Effects of experimental exposure of lowlevel artificial light at night on circadian rhythms and health of wild nestlings

4.1 Introduction

Urban wildlife must navigate increasingly noisy, polluted, and fragmented habitats as the human population increases and cities continue to expand (McKinney, 2002). One predominant feature of the urban environment is ecological light pollution, defined as light emissions from sources such as streetlamps and buildings (Longcore and Rich, 2004).

Currently, the world is experiencing a loss of the night (Kyba et al., 2017), and therefore it is increasingly important to consider effects of ecological light pollution on wildlife living in artificially-lit environments. To date, artificial light at night (otherwise known as ALAN) has been shown to have observable effects on many wild animal groups, from a decrease in juvenile growth in amphibians (Dananay and Benard, 2018), to altered species composition of bat communities (Cravens and Boyles, 2019), and attraction of moth species to lamps (van Langevelde et al., 2011). Ecological light pollution, particularly artificial light at night, has therefore rapidly become a subject of ecological interest over the past decade.

A main suspected reason for potential harm of ALAN is through its effects on biological rhythms of wildlife and humans (Fonken and Nelson, 2011; Navara and Nelson, 2007). Ambient light-dark cycles are the most potent zeitgeber for endogenous timekeeping mechanisms (Pittendrigh, 1981). Diel rhythms of physiology and behavioural activity are tied strongly to light and darkness, for example by the timely release of the hormone melatonin, a key regulator of circadian rhythms (Dawson et al., 2001; Gwinner et al., 1997). Light at night has a suppressive effect on melatonin release (Navara and Nelson, 2007), providing one way in which ALAN has the potential to disrupt normal circadian rhythms.

For organisms living in artificially-lit environments, disruption of circadian rhythms by ALAN may have downstream effects on physiology and epidemiology. In mice, ALAN has been shown to increase metabolic disruption (Fonken and Nelson, 2014), and the likelihood of cancer (McFadden et al., 2014). Also in mice, dim light exposure disrupts expression of several clock genes in both the suprachiasmatic nucleus of the brain and

peripheral tissues (Fonken et al., 2013; Shuboni and Yan, 2010). Clocks are strongly tied to the immune system (Scheiermann et al., 2013) and therefore any circadian disruption has the potential to affect inflammatory and immune responses. Indeed, ALAN suppresses immune system function in hamsters (Bedrosian et al., 2013, 2011).

In birds, the potential of ALAN to disrupt rhythms is even greater than in mammals, given that birds possess deep-brain photoreceptors that are sensitive to light (Thakur and Kumar, 2015; Underwood et al., 2001). Captive studies have illuminated the adverse effects that ALAN exposure can have on avian physiology and health (Spoelstra and Visser, 2013). One study found an increase in corticosterone levels in adult great tits as a response to ALAN (Ouyang et al., 2015). Experimental exposure to light at night also results in shifts in the timing of behaviour, such as earlier onsets of daytime activity in Indian weaver birds (*Ploceus philippinus*; Kumar *et al.*, 2018) and also in great tits (de Jong et al., 2016).

Ecological studies on how ALAN affects avian rhythms are rare, but have so far demonstrated its presence to have similar effects to captive studies on behavioural timing, advancing onset of dawn singing in several passerine species in the wild (Da Silva et al., 2015; Kempenaers et al., 2010; Raap et al., 2015). Although it is well known that ALAN has effects on the timing of behaviours in wild birds, the mechanisms behind the observed shifts in activity timing as a result of exposure have yet to be elucidated. Furthermore, despite disruptive effects of ALAN on circadian rhythms in captive studies, and the detrimental effects of disruption on health (Fonken and Nelson, 2011), effects on the condition of wild birds have largely been unexplored. Therefore, there is a need for studies exploring effects of ALAN on both circadian rhythms and fitness of birds in the natural environment.

In ecological studies, the use of nest boxes allows for experimental manipulations such as the implementation of ALAN treatments, and provide valuable insights into effects on nestling birds. A previous nest box study exposing great tit nestlings to constant light for two successive nights found negative effects on nestling growth and increased oxidative stress (Raap et al., 2016c). Another study found nestlings exposed to ALAN for seven consecutive nights to have a decline in overall body condition (Grunst et al., 2019). However, these studies were limited in their approach, as the authors did not expose nestlings to ALAN for the entire growth period.

4.2 Study aims and hypotheses

To address current gaps in knowledge, in this Chapter, wild great tit nestlings were experimentally exposed to low-level artificial light at night from their hatching up until fledging to compare condition and health parameters of ALAN-raised nestlings with those raised under a dark night. Ecological studies of the effects of ALAN in urban environments are often limited by the presence of other environmental effects such as ambient noise, temperature and atmospheric pollutants (Dominoni, 2015). As such, it is difficult to untangle the effects of noise and light on birds, although a few ecological studies have done so (Caorsi et al., 2019; Casasole et al., 2017; de Jong et al., 2018). This study was conducted in a forest habitat in order to isolate the effects of ALAN and minimise confounding environmental factors such as urban noise.

Condition and health effects of being raised under ALAN were assessed using commonly used parameters of nestling condition, namely nestling weight and tarsus length (Eck et al., 2011), and fledging success. For information on health, blood samples were used to measure red blood cell count (haematocrit levels (%)) and to monitor transcript levels of relevant genes.

In this Chapter, the possibility of similar circadian disruption from ALAN exposure was addressed, by comparisons of transcript levels of seven core avian clock genes in ALAN-raised nestlings with those raised under dark nights. The following components of the molecular clock were included in this study: positive clock elements and transcriptional activators *BMAL1* (brain and muscle ARNT-like 1) and *CLOCK* (circadian locomotor output cycles kaput); negative clock elements and inhibitors *CRY1* (cryptochrome circadian regulator 1) and *PER2* (period circadian regulator 2); and accessory loop protein *REVERBA* (Kumar and Sharma, 2018; Yoshimura et al., 2000). Additional gene targets included *CK1E* (casein kinase 1), which catalyses phosphorylation of E4bp4, a light-inducible gene in the nestling pineal gland (Doi et al., 2004), and *AANAT* (aralkylamine N-acetyltransferase), the rate-limiting enzyme regulating melatonin synthesis (Bernard et al., 2002).

This study also addressed the possibility that ALAN affects immune system function in nestlings, by comparing transcript levels of eleven immune targets in ALAN-raised birds to those raised under dark nights. Targets included LY86 (lymphocyte antigen 86) and TLR4 (toll-like receptor 4) which are involved in anti-bacterial and anti-malarial responses (Medzhitov, 2001), and the type 2 transcription factor *GATA3* (GATA binding protein 3) involved in adaptive immunity (Wang et al., 2011). Additional immune targets included: IL1 and IL6 (interleukins 1 and 6) which are inflammatory response mediators (Klasing, 1998); PRKCA (protein kinase C alpha), an enzyme responsible for antiviral effects and cell growth regulation (Clemens and Elia, 1997); and IKBA (NF-kappa-B inhibitor alpha) and NKRF (NF-kappa-B repressing factor), which are two inhibitors of the immune response regulator NF-kappa-B (Cabannes et al., 1999; Nourbakhsh and Hauser, 1999). Finally, immune targets NRF2 (nuclear factor erythroid 2) and IGF1 (insulin like growth factor 1), which are both involved in resistance to oxidative stress and metabolic regulation (Holzenberger et al., 2003; Kensler et al., 2007) were included in this study. A table of the known functions of each gene used in this study is available in the General Methods section.

In order to capture fluctuations of transcript levels of circadian clock and immune system elements across a 24 hr day, the sampling protocol in this study was designed so that samples were obtained from nestlings during the day and at night.

Following previous studies that show a reduction of overall condition as a response to experimental exposure to ALAN (Grunst et al., 2019; Raap et al., 2016a, 2016b), it was expected that birds under ALAN treatment would have reduced condition, such as lower weight and haematocrit levels, shorter tarsi, and lower fledging success than control birds.

Previous studies have shown ALAN to disrupt rhythmic expression of clock genes (Bedrosian et al., 2013, 2011) and immunity (Mishra et al., 2019). In this study, it was expected that ALAN treatment would have significant downregulatory effects on the timing of expression of clock and immune genes, with an overall reduction in time of day differences.

In addition, ALAN has been shown to elevate (Raap et al., 2016b; Saini et al., 2019; Zhang et al., 2019), and also downregulate (Moore and Siopes, 2000) aspects of immune activity. In this study, it was expected that ALAN would have significant effects on the levels of immune genes, with target specific responses.

4.3 Materials and Methods

4.3.1 Study sites

Data for this chapter were obtained between April and June of 2017 at forest field sites on a free-living population of breeding great tits. Field sites were existing forest nest box study systems located at the Scottish Centre for Ecology and the Natural Environment (SCENE; 56°, 7.73'N, 4° 36.79'W), and Cashel Forest (56° 6'N, 4°34'W), and were previously used in ecological studies of wild passerines (Capilla-Lasheras et al., 2017; Pollock et al., 2017). To determine hatch dates, brood sizes and fledging success, all nest boxes were initially checked weekly and, closer to hatching, every second day as part of a general nest box monitoring protocol (see General Methods for further details of nest box monitoring).

4.3.2 Light at night treatment

Fifteen treatment nest boxes (Schwegler, Germany) contained one small light-emitting diode (LED) of broad-spectrum low intensity (1 lux) within the ceiling of the nest box. Lighting intensity was selected as the lowest intensity ALAN treatment to have a significant effect on great tit activity patterns as per de Jong et al., (2016). Treatment (ALAN) and control nest boxes were assigned to nest boxes previously identified as containing great tit clutches. Seventeen nest boxes containing no LED (dark-night) were monitored as controls. Nest boxes were matched by synchronous hatch date. Treatment nest boxes were swapped with existing nest boxes, by careful removal of the whole nest (including eggs) and replacement into the nest box containing an LED. This occurred on the expected hatch date, which was calculated to be fourteen days after laying of the first egg as a conservative estimate of great tit incubation length. Lights were switched on in treatment nest boxes on the expected hatch date and were left on constantly until after nestling fledging. Similar levels of experimental disturbance occurred for all nest boxes within the study, where both treatment and control nest boxes were visited on the expected day of hatching, once on day eight of life (where hatch date = day 0 of life), and twice on the sampling day on day thirteen of life. On this day, nest boxes were visited during the midday hours (11:00-15:00 GMT) and then again during the hours surrounding midnight (23:00-02:00 GMT).

Fifteen of the nest boxes within this study (seven ALAN and eight control nest boxes) were treated with a citronella oil-based insect repellent as per (Krams *et al.*, (2013) as part of an avian malaria study (see General Methods section for more details). Citronella treatment was kept as a variable in all statistical models in this study.

4.3.3 Nestling processing

In total, 97 nestlings from fifteen nest boxes were subjected to ALAN treatment, and 116 nestlings from 17 nest boxes were under "dark-night" controls. All nestlings (N=213) were ringed for individual identification on day eight of life. Measurements of nestling weight and length of the tarsometatarsus bone (tarsus) were taken from each nestling. On day thirteen of life, nestlings were individually weighed again, and their tarsi measured as in Eck *et al.*, (2011). All nestlings were weighed during the day between (08:00 and 17:00 GMT) to reduce diurnal fluctuations in weight. To reduce burden of sampling on nestlings, no nestling was sampled twice on one day. Sampling of broods on day thirteen was split so that up to three of the nestlings within a brood were sampled during the day (10:00-15:00 GMT), and up to three others during the night (23:00-02:00 GMT). Table 4.1. summarises sample sizes obtained during the study.

Two blood samples of approximately 50 µl were obtained from each nestling via the brachial wing vein. During noon sampling, the first sample was taken up by a heparinised capillary, sealed using a wax plug and stored on ice for haematocrit level (%) readings (See General Methods for more details on protocol for haematocrit readings). After haematocrit readings were taken, this sample was removed from the capillary via pipette and stored in 0.5 ml 100% ETOH for molecular analyses. For night sampling, the first sample was taken up by heparinised capillary and immediately stored in 0.5 ml 100% ETOH for molecular analyses. Haematocrit level readings were not obtained for nestlings during night sampling given the difficulties of night sampling in the field. For both noon and night sampling, a second blood sample was obtained for each nestling via heparinised capillary and stored in 250 µl RNAlater® stabilising solution (Invitrogen) for analyses of gene transcript levels. The General Methods section provides further details on nestling blood sampling protocol and ethical licencing.

Table 4.1. Breakdown of number of samples obtained from ALAN treatment and control nest boxes by day. A maximum of six nestlings were sampled per nest on day thirteen.

Treatment	Day Eight	Day Thirteen	Day Thirteen
		Noon 10:00-15:00	Night
		GMT	23:00-02:00 GMT
ALAN	87	30	30
Control	105	43	29
		73	59

4.3.4 Laboratory procedures

Haematocrit levels (%) were measured for noon samples, before pipetting out the remaining blood sample into 100% ETOH. Samples in 100% ETOH were frozen and stored at -40 °C. Genomic DNA was later extracted from these samples via DNAeasy Blood and Tissue Kit (Qiagen) and molecular sexing was carried out via PCR (See General Methods for details).

Measurement of Median Fluorescence Intensity units (as a proxy for transcript levels) for immune, clock and housekeeping gene targets outlined in Table 4.2. was carried out on all samples stored in *RNAlater*® using the QuantiGene® Plex RNA 2.0 assay (ThermoFisher). The QuantiGene® Plex RNA 2.0 assay is a platform that allows for simultaneous readouts of transcript levels of many target genes from a single blood sample using branched DNA signal amplification and multi-analyte profiling beads (Mills and Gallagher, 2016). The General Methods section provides further details on lab procedures carried out in this Chapter.

Table 4.2. Housekeeping, clock and immune gene targets used within this study. Primer information for gene targets can be found in the Appendix.

Transcript targets	Туре
RPL19, SDHA, HMBS	Housekeeping genes
BMAL1, CLOCK, CRY1,CK1E, AANAT, PER2, REVERBA	Circadian clock genes and melatonin synthesis
LY86, TLR4, IL1, IL6, PRKCA, GATA3, IKBA, NKRF, NRF2, IGF1	Immune genes

4.3.5 Ethical Statement

See General Methods for details of ringing licences and blood sampling.

4.3.6 Statistical Analysis and Model Selection

All statistical analyses were performed using R v. 3.4.4 (R Core Team, 2018) and packages *lme4* (Bates et al., 2015). Data were visualised using ggplot2 (Wickham, 2014). Initial global models for response variables were linear-mixed models (LMM), including individual nest box as a random factor to account for repeated measures within the same nest box. Residuals were plotted and inspected to check assumptions of normality whilst modelling using LMMs. Pairs of explanatory variables were also assessed for collinearity by calculating tolerance and variance inflation factors (VIF) and also by visual inspection of pairwise plots, before statistical relevance was accepted.

ALAN treatment and control were included in all models as a two-level fixed factor. Fixed factors in all models included sex of the nestling (M/F) for sex differences, brood size, hatch date in April day format (where April 1st = 1) to account for seasonal changes, and citronella oil repellent (presence/absence). Nestling weight, tarsal length and haematocrit were modelled using a Gaussian distribution, and fledging success was modelled using a binomial generalized linear-mixed model (GLMM).

Median fluorescence intensity data obtained from the QuantiGene® Plex RNA assay 2.0 (Thermofisher) were normalised by division of the mean fluorescence of background wells and then dividing each sample by the geometrical mean of levels of housekeeping genes in Table 4.2, thus, transcript levels are expressed as relative levels to housekeeping genes. Transcript level data for all targets were then modelled using Gamma log-link function for analysis using LMMs. "Sample time" was included as a fixed two-level factor by "noon" for samples obtained during 10:00-15:00 GMT or "midnight" for samples obtained during 23:00-02:00 GMT. An interaction between ALAN treatment and sample time was also included in all gene models.

Statistical relevance of fixed factors within all models was determined by sequential model selection via Likelihood Ratio Testing using package *lmtest* (Zeileis and Hothorn, 2002), until the minimum adequate model (MAM) was found. Factors with a significance level of <5% were kept in the model. ALAN treatment was kept in final models regardless of significance to monitor effect size.

4.4 Results

4.4.1 Artificial light at night treatment and nestling health

No significant effects of ALAN treatment were observed during modelling of nestlings' body weight at day eight, tarsus length or haematocrit (Table 4.4.; Figures 4.1., 4.2., 4.3.). However, nestling weight at day thirteen was significantly affected by presence of ALAN treatment ($t_{1,129}$ =-2.04: P=0.038, LMM), (Figure 4.1.2.). Body weight of control nestlings (20.53 g ± 0.94) was on average higher than nestlings under ALAN (19.68 g ± 0.42).

There was little variation in fledging success between treatment (97%) and control (99%) groups, and possibly because of this, no effect of ALAN treatment on fledging success was found ($z_{1,207}$ =0.74, P=0.452, GLM). Hatch date significantly affected fledging success, however, with fledging success decreasing as the season progressed ($z_{1,207}$ =-2.38: P=0.002, GLM).

There were no significant effects of brood size or hatch date on body weight (at day eight), or tarsus length of nestlings (Table 4.3.). However, brood size significantly affected body weight of nestlings at day thirteen ($t_{1,129} = -1.96$: P = 0.045; LMM), where nestlings in larger clutches had lower body weight. The presence of citronella treatment had no significant effect on nestling body weight, tarsus length, or haematocrit levels on day eight or thirteen (Table 4.4.).

Sex differences in nestling morphology were evident in this study, with male nestlings significantly heavier than females both at day eight ($t_{1,153}$ = 2.66: P = 0.010; LMM) and at day thirteen ($t_{1,129}$ = 2.90: P=0.004; LMM). On day eight, male nestlings weighed 14.7 g (± 0.20) and females 14.17 g (± 0.25), (Table 4.4.). On day thirteen, males weighed 21.13 g (± 0.21) and females 20.53 g (± 0.94), (Table 4.4.). In addition, male nestlings had significantly longer tarsi than females at day thirteen ($t_{1,163}$ = 3.36: P<0.001; LMM), with tarsi of male nestlings 20.64 mm (± 0.10) and females 20.29 mm (± 0.21), (Table 4.4.).

Global model summaries for nestling health data can be found in Table 4.5.

Table 4.3. Final (MAM) linear-mixed models of nestling weight, tarsus length and haematocrit levels at day eight and thirteen. Sex was a two-level fixed factor (M/F). Hatch date was defined as the number of days from the 1st of April 2017, brood size was the number of nestlings within a nest box, ALAN treatment a two-level fixed factor (ALAN/Control), Citronella presence was a two-level fixed factor (Presence/Absence). The global model for fledging success was Nest box (random) + Hatch date + ALAN + Citronella. Global models for all other variables of nestling condition were: Sex + Nest box (random) + Brood size + Hatch date + ALAN + Citronella. Nest box was included in all models as a random factor. Addition of terms is denoted by a plus sign (+).

Variable	ANOVA Minimum Adequate Model (MAM)
Weight day eight	Nest box $+$ Sex $+$ ALAN
Weight day thirteen	Nest box + $Sex + ALAN + Brood size$
Tarsus length day eight	Nest box + ALAN
Tarsus length day thirteen	Nest box + Sex + ALAN
Haematocrit % day eight	Nest box + Hatch date + Brood size + Sex + ALAN
Haematocrit % day thirteen	Nest box + ALAN
Fledging success	Nest box + Hatch date + ALAN

Table 4.4. Minimum adequate model summaries from linear mixed models for nestling weight, tarsus length and haematocrit (%) at days eight (d8) and day thirteen (d13), and fledging success. All variables were modelled using a Gaussian distribution except for fledging success, which was modelled using Binomial distribution. Significance (p<0.05) were shown in bold and values were obtained from likelihood ratio testing. Reference levels for the intercept were: Sex (F), Treatment (Control), Citronella (Citronella present, 1).

Variable	N	Estimate	s.e.m	d.f.	t-value	<i>p</i> -value LRT
Weight d8	153					
(Intercept)		14.17	0.25		56.94	
Sex (M)		0.53	0.20	1	2.66	0.010
ALAN		-0.22	0.34	1	-0.65	0.504
Weight d13	129					
(Intercept)		20.53	0.94		21.89	
Sex (M)		0.60	0.21	1	2.90	0.004
Brood size		-0.25	0.13	1	-1.96	0.045
ALAN		-0.85	0.42	1	-2.04	0.038
Tarsus d8	191					
(Intercept)		17.59	0.31		56.59	
ALAN		-0.50	0.45	1	-1.10	0.251
Tarsus d13	163					
(Intercept)		20.29	0.21		95.57	
Sex (M)		0.35	0.10	1	3.36	< 0.001
ALAN		-0.05	0.31	1	-0.14	0.876
Haematocrit % d8	142					
(Intercept)		65.45	8.48		7.72	
Sex(M)		2.86	1.27	1	2.25	0.023
Brood Size		-1.00	0.40	1	-2.48	0.013
Hatch Date		-0.41	0.17	1	-2.34	0.018
ALAN		-0.74	1.35	1	-0.55	0.575
Haematocrit % d13	45					
(Intercept)		38.92	1.16		33.52	
ALAN		2.33	1.77	1	1.32	0.175
Variable	N	Estimate	s.e.m	d.f.	z-value	<i>p-</i> value LRT
Fledging Success	207					
(Intercept)		34.85	13.20		2.64	
Hatch date		-0.59	0.25	1	-2.38	0.002
ALAN		1.34	1.82	1	0.74	0.452

Table 4.5. Global model summaries from linear mixed models for nestling weight, tarsus length and haematocrit (%) at days eight (d8) and day thirteen (d13), and fledging success. All variables were modelled using a Gaussian distribution except for fledging success, which was modelled using Binomial distribution. Significance (p<0.05) were shown in bold and values were obtained from likelihood ratio testing. Reference levels for the intercept were: Sex (F), Treatment (Control), Citronella (Citronella present, 1).

Variable	N	Estimate	s.e.m	d.f.	<i>t</i> -value	<i>p</i> -value LRT
Weight d8	153					
(Intercept)		16.93	2.36		7.17	
Sex (M)		0.53	0.20	1	2.65	0.012
Brood Size		-0.06	0.12	1	-0.54	0.548
Hatch Date		-0.05	0.05	1	-0.96	0.302
ALAN		-0.13	0.38	1	-0.34	0.695
Citronella(0)		-0.22	0.36	1	-0.60	0.514
Weight d13	129					
(Intercept)		24.49	3.38		7.30	
Sex (M)		0.60	0.21	1	2.92	0.004
Brood Size		-0.22	0.13	1	-1.71	0.066
Hatch Date		-0.09	0.07	1	-1.26	0.177
ALAN		-0.75	0.43	1	-1.75	0.059
Citronella(0)		-0.06	0.43	1	-0.14	0.087
Tarsus d8	191					
(Intercept)		19.93	2.92		6.82	
Sex(M)		0.16	0.16	1	1.04	0.290
Brood Size		-0.11	0.14	1	-0.73	0.423
Hatch Date		-0.03	0.06	1	-0.44	0.626
ALAN		-0.43	0.47	1	-0.92	0.319
Citronella(0)		-0.82	0.45	1	-1.84	0.051
Tarsus d13	163					
(Intercept)		24.24	2.52		9.60	
Sex(M)		0.35	0.10	1	3.34	< 0.001
Brood Size		-0.11	0.10	1	-1.09	0.237
Hatch Date		-0.07	0.05	1	-1.29	0.166
ALAN		-0.04	0.32	1	-0.11	0.888
Citronella(0)		0.17	0.32	1	0.52	0.561
Haematocrit %	142					
d8						
(Intercept)		65.39	8.52		7.68	
Sex(M)		2.96	1.29	1	2.29	0.020
Brood Size		-1.03	0.41	1	-2.51	0.011
Hatch Date		-0.41	0.17	1	-2.35	0.017
ALAN		-0.77	1.36	1	-0.57	0.564
Citronella(0)		0.65	1.27	1	0.51	0.600
Haematocrit %	45	0.05	1.27	1	0.51	0.000
d13	43					
(Intercept)		39.00	14.97		2.60	
Sex(M)		0.53	1.58	1	0.34	0.728
Brood Size		-0.55	0.71	1	-0.77	0.728
Hatch Date		0.07	0.71	1	0.24	0.729
ALAN		2.49	2.07	1	1.20	0.129
		-0.33	2.07	1	-0.15	0.183
Citronella(0)		-0.55	2.14	1	-0.13	0.832

Variable	N	Estimate	s.e.m	d.f.	z-value	<i>p</i> -value LRT
Fledging Success	207					
(Intercept)		34.42	13.29		2.59	< 0.001
Hatch date		-0.58	0.25	1	-2.35	0.019
ALAN		1.34	1.86	1	0.72	0.472
Citronella(0)		0.25	1.12	1	0.22	0.824

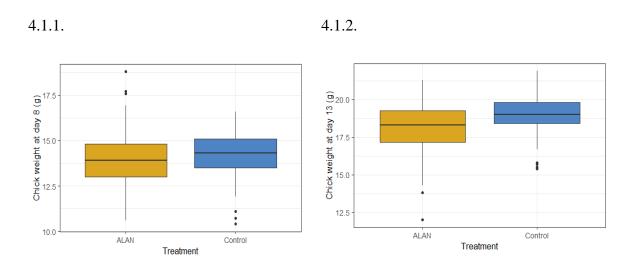


Figure 4.1. Artificial light at night and control treatment and nestling weight (g) at 4.1.1. Day eight and 4.1.2. Day thirteen. Box plots show the interquartile range of the raw data, where boxes above and below the medians (horizontal lines) show the first and third quartiles, respectively. Boxplot whiskers extend to 1.5 times the interquartile range. Dots represent outliers.

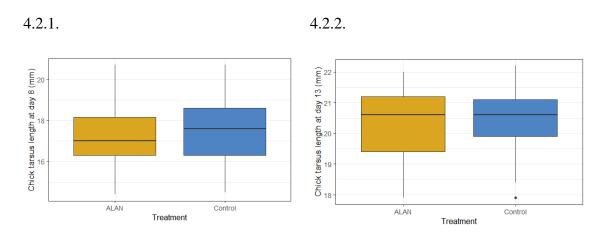
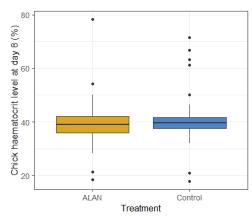


Figure 4.2. Artificial light at night and control treatment and tarsus length (mm) at 4.2.1. Day eight and 4.2.2. Day thirteen. Box plots show the interquartile range of the raw data, where boxes above and below the medians (horizontal lines) show the first and third quartiles, respectively. Boxplot whiskers extend to 1.5 times the interquartile range. Dots represent outliers.





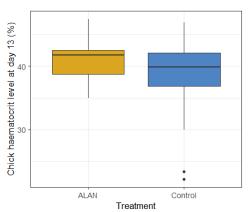


Figure 4.3. Artificial light at night and control treatment and haematocrit levels (%) at 4.3.1. Day eight and 4.3.2. Day thirteen. Box plots show the interquartile range of the raw data, where boxes above and below the medians (horizontal lines) show the first and third quartiles, respectively. Boxplot whiskers extend to 1.5 times the interquartile range. Dots represent outliers.

4.4.2 Artificial light at night treatment and gene transcript levels

Gene transcript levels (measured as median fluorescence intensity) were plotted for ALAN treatment in Figures 4.4. and 4.5. and outcomes of minimum adequate models listed in Table 4.6.

Transcript levels for target genes did not differ significantly between ALAN treatment and control groups, except for the clock gene *REVERBA* ($t_{1,93}$ =-2.35: P=0.026, LMM), (Table 4.7). For this gene, transcript levels were significantly lower in ALAN birds (0.24 ± 0.32) than in control birds (0.51 ± 0.22), with a >2-fold difference overall (Figure 4.4.1.). Overall, no indication was found for interactions between time of day effects and ALAN treatment, and therefore this interaction term was dropped from minimum adequate models for all gene targets (Table 4.6).

Eleven (*IGF1*, *TLR4*, *IL1*, *IL6*, *PRKCA*, *GATA3*, *IKBA*, *BMAL1*, *CRY1*, *CK1E*, *AANAT*) out of seventeen target genes (both clock and immune) differed in transcript levels between day and night, with transcript levels lower at noon compared to midnight. These day-night differences ranged from a 1.13-fold change to 1.54-fold change (Figures 4.4., 4.5.). The remaining six gene targets (*LY86*, *REVERBA*, *CLOCK*, *NKRF*, *NRF2* and *PER2*) did not differ significantly in transcript levels between day and night (Figures 4.4., 4.5.).

Male nestlings had significantly lower transcript levels than females (ranging from a 1.14-fold change to 1.43-fold change) in eleven (*IGF1*, *TLR4*, *IL1*, *IL6*, *PRKCA*, *IKBA*, *BMAL1*,

CRY1, *CK1E*) clock and immune gene targets (Table 4.7). Male nestlings had significantly higher levels of *NKRF* and *NRF2* than females (Table 4.7). There was no evidence for effects of hatch date, brood size or citronella on transcript levels of target genes (Table 4.6.).

Global model summaries for clock and immune gene transcript data can be found in Table 4.8.

Table 4.6. Global and final (MAM) linear-mixed models of circadian clock and immune gene expression. SampleTime was a two-level factor (NOON/MIDNIGHT), ALAN treatment a two-level fixed factor (ALAN/Control) and ALAN*SampleTime a two-way interaction. All gene expression data were modelled using Gamma distribution except for *NRF2*, which was modelled using a Gaussian distribution. Global gene models were: Nest box (random) + Sample Time + Sex + ALAN + Brood size + Hatch date + Citronella + (Sample Time*ALAN). Nest box was always retained as a random factor. Addition of terms is denoted by a plus sign (+). Interaction of terms is denoted by an asterisk (*).

Variable	Minimum Adequate Model (MAM)
LY86	Nest box $+$ ALAN
<i>REVERBA</i>	Nest box $+$ ALAN
IGF1	Nest box + Sample Time + Sex + ALAN
TLR4	Nest box + Sample Time + Sex + ALAN
IL1	Nest box + Sample Time + Sex + ALAN
IL6	Nest box + Sample Time + Sex + ALAN
PRKCA	Nest box + Sample Time + Sex + ALAN
GATA3	Nest box + Sample Time + ALAN
<i>IKBA</i>	Nest box + Sample Time + ALAN
BMAL1	Nest box + Sample Time + Sex + ALAN
CLOCK	Nest box $+$ ALAN
CRY1	Nest box + Sample Time + Sex + ALAN
CK1E	Nest box + Sample Time + Sex + ALAN
AANAT	Nest box + Sample Time + Sex + ALAN
<i>NKRF</i>	Nest box $+$ Sex $+$ ALAN
NRF2	Nest box $+$ Sex $+$ ALAN
PER2	Nest box + ALAN

Table 4.7. Minimum adequate model summaries of transcript levels of clock and immune targets from linear mixed models. Estimates are log-transformed. Significance (p<0.05) is emphasised in bold and values were obtained through likelihood ratio testing. Reference levels for the intercept were: Sex (F), SampleTime (Midnight), Treatment (Control).

Variable	N	Estimate	s.e.m	d.f.	<i>t</i> -value	<i>p</i> -value LRT
LY86	106					
(Intercept)	100	-1.55	0.002		-565.19	
ALAN		-0.03	0.002	1	-11.16	0.853
				_	22120	0.000
REVERBA	93					
(Intercept)		-0.68	0.22		-3.13	
ALAN		-0.76	0.32	1	-2.35	0.026
IGF1	93					
(Intercept)		-1.80	0.16		-11.26	
Sex(M)		-0.25	0.11	1	-2.21	0.030
Sample		-0.30	0.11	1	-2.81	0.006
Time(Noon) ALAN		0.08	0.20	1	-0.44	0.660
ALAN		0.08	0.20	1	-0.44	0.000
TLR4	96					
(Intercept)	90	-1.64	0.17		-9.68	
Sex(M)		-0.26	0.17	1	-2.24	0.028
Sample		-0.26	0.11	1	-2.53	0.013
Time(Noon)		0.20	0.10	1	2.33	0.013
ALAN		-0.15	0.22	1	-0.70	0.486
IL1	92					
(Intercept)		-1.64	0.14		-11.82	
Sex(M)		-0.23	0.10	1	-2.24	0.027
Sample Time(Noon)		-0.35	0.10	1	-3.49	0.001
ALAN		-0.23	0.16	1	-1.43	0.172
IL6	88					
(Intercept)		-1.73	0.16		-10.63	
Sex(M)		-0.28	0.11	1	-2.59	0.011
Sample Time (Noon)		-0.33	0.11	1	-3.13	0.002
ALAN		-0.17	0.20	1	-0.88	0.396
DDVCA	96					
PRKCA	86	1.05	0.10		10.40	
(Intercept)		-1.85 -0.30	0.18 0.12	1	-10.49 -2.51	0.014
Sex(M) Sample		-0.34	0.12	1	-3.03	0.014
Time(Noon)		-0.54	0.11	1	-3.03	0.003
ALAN		-0.07	0.21	1	-0.34	0.734
GATA3	98					
(Intercept)		-1.84	0.13		-14.05	
Sample Time(Noon)		-0.17	0.09	1	-1.86	0.033
ALAN		-0.10	0.17	1	-0.59	0.561

Variable	N	Estimate	s.e.m	d.f.	t-value	<i>p</i> -value LRT
IKBA (Intercept)	102	-1.78	0.16		-10.99	
Sex(M)		-0.20	0.10	1	-2.15	0.033
Sample		-0.23	0.09	1	-2.13	0.006
Time(Noon)		-0.23	0.00	1	-2.77	0.000
ALAN		-0.05	0.22	1	-0.23	0.818
BMAL1	94					
(Intercept)		-1.75	0.14		-12.65	
Sex(M)		-0.30	0.09	1	-3.46	0.001
Sample		-0.20	0.08	1	-2.37	0.019
Time(Noon)		0.02	0.10	1	0.12	0.000
ALAN		-0.02	0.18	1	-0.13	0.896
CLOCK	101					
(Intercept)	101	-1.56	0.07		-22.39	
ALAN		-0.17	0.07	1	-1.64	0.108
ALAI		-0.17	0.10	1	-1.04	0.100
CRY1	109					
(Intercept)	10)	-0.86	0.08		-10.26	
Sex(M)		-0.12	0.06	1	-2.02	0.046
Sample		-0.12	0.05	1	-2.22	0.028
Time(Noon)						
ALAN		-0.14	0.10	1	-1.38	0.180
CK1E	93					
(Intercept)		-1.82	0.16		-11.27	
Sex(M)		-0.25	0.10	1	-2.72	0.016
Sample		-0.27	0.10	1	-2.72	0.007
Time(Noon) ALAN		-0.04	0.21	1	-0.19	0.852
TILI II V		0.04	0.21	1	0.17	0.032
AANAT	92					
(Intercept)	/ -	-1.83	0.16		-11.33	
Sample		-0.23	0.11	1	-2.21	0.013
Time(Noon)						
ALAN		-0.19	0.22	1	-0.85	0.369
NKRF	108					
(Intercept)		0.48	0.10		4.88	
Sex(M)		0.11	0.04	1	3.17	0.002
ALAN		-0.2	0.15	1	-1.38	0.175
ND EA	110					
NRF2	110	5 75	0.25		16.26	
(Intercept)		5.75	0.35	1	16.26	40 001
Sex(M)		1.07	0.21	1	4.99	<0.001
ALAN		-0.50	0.51	1	-0.98	0.227
PER2	105					
(Intercept)	103	-1.44	0.15		-9.35	
ALAN		-0.19	0.13	1	-9.33 -0.82	0.421
ALAN		-0.19	0.23	1	-0.02	0.441

Table 4.8. Global model summaries of transcript levels of clock and immune targets from linear mixed models. Estimates are log-transformed. Significance (p<0.05) is emphasised in bold and values were obtained through likelihood ratio testing. Reference levels for the intercept were: Sex (F), SampleTime (Midnight), Treatment (Control).

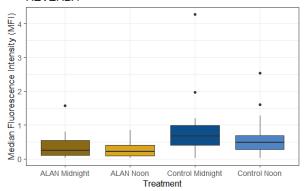
Variable	N	Estimate	s.e.m	d.f.	<i>t</i> -value	<i>p</i> -value
						LRT
LY86	106					
(Intercept)		-2.09	1.26		-1.66	0.096
SampleTime(Noon)		-0.04	0.08	1	-0.46	0.649
Sex(M)		-0.08	0.06	1	-1.26	0.207
ALAN		0.03	0.17	1	0.18	0.857
Brood Size		-0.02	0.05	1	-0.37	0.710
Citronella(0)		0.12	0.15	1	0.81	0.420
Hatch Date		0.01	0.03	1	0.53	0.597
SampleTime*ALAN		-0.15	0.11	2	-1.27	0.203
REVERBA	93					
Intercept)		2.17	2.57		0.85	0.397
SampleTime(Noon)		-0.16	0.21	1	-0.80	0.424
Sex(M)		-0.28	0.16	1	-1.76	0.078
ALAN		-0.60	0.36	1	-1.66	0.097
Brood Size		-0.06	0.09	1	0.59	0.555
Citronella(0)		0.05	0.31	1	0.17	0.863
Hatch Date		-0.06	0.05	1	-1.19	0.236
SampleTime*ALAN		-0.02	0.29	2	-0.71	0.481
GF1	93					
(Intercept)		-4.02	1.58		-2.54	0.011
SampleTime(Noon)		-0.27	0.14	1	-1.88	0.060
Sex(M)		-0.25	0.12	1	-2.15	0.031
ALAN		-0.10	0.23	1	-0.45	0.652
Brood Size		-0.03	0.06	1	-0.45	0.650
Citronella(0)		-0.07	0.19	1	-0.36	0.718
Hatch Date		0.05	0.03	1	1.52	0.127
SampleTime*ALAN		-0.08	0.21	2	-0.36	0.717
TLR4	96					
Intercept)		-3.55	1.78		-2.00	0.046
SampleTime(Noon)		-0.21	0.14	1	-1.48	0.140
Sex(M)		-0.26	0.12	1	-2.26	0.024
ALAN		-0.13	0.25	1	-0.53	0.599
Brood Size		-0.02	0.06	1	-0.25	0.805
Citronella(0)		-0.08	0.21	1	-0.36	0.719
Hatch Date		0.04	0.04	1	1.15	0.251
SampleTime*ALAN		-0.12	0.21	2	-0.58	0.564
IL1	92					
Intercept)		-2.45	1.35		-1.81	0.070
SampleTime(Noon)		-0.25	0.14	1	-1.86	0.063
Sex(M)		-0.23	0.10	1	-2.25	0.025
ALAN		-0.14	0.21	1	-0.69	0.489
Brood Size		-0.01	0.05	1	-0.11	0.911
Citronella(0)		0.07	0.17	1	0.44	0.660
Hatch Date		0.02	0.03	1	0.56	0.573
SampleTime*ALAN		-0.17	0.20	2	-0.87	0.384

Variable	N	Estimate	s.e.m	d.f.	<i>t</i> -value	<i>p</i> -value LRT
IL6	88					
(Intercept)		-3.40	1.55		-2.19	0.029
SampleTime(Noon)		-0.22	0.15	1	-1.51	0.131
Sex(M)		-0.31	0.11	1	-2.76	0.006
ALAN		-0.11	0.23	1	-0.46	0.643
Brood Size		-0.02	0.06	1	-0.42	0.675
Citronella(0)		-0.07	0.19	1	-0.37	0.711
Hatch Date		0.04	0.03	1	1.17	0.240
SampleTime*ALAN		-0.23	0.21	2	-1.09	0.278
PRKCA	86					
(Intercept)		-4.00	1.69		-2.37	0.018
SampleTime(Noon)		-0.30	0.15	1	-1.95	0.051
Sex(M)		-0.29	0.12	1	-2.44	0.015
ALAN		-0.09	0.25	1	-0.35	0.730
Brood Size		-0.05	0.06	1	-0.81	0.416
Citronella(0)		-0.11	0.20	1	-0.53	0.599
Hatch Date		0.05	0.03	1	1.50	0.134
SampleTime*ALAN		-0.09	0.22	2	-0.39	0.695
GATA3	98					
(Intercept)		-3.05	1.45		-2.11	0.035
SampleTime(Noon)		-0.16	0.13	1	-1.24	0.214
Sex(M)		-0.16	0.10	1	-1.61	0.108
ALAN		-0.10	0.21	1	-0.47	0.635
Brood Size		-0.01	0.05	1	-0.23	0.819
Citronella(0)		-0.02	0.18	1	-0.13	0.897
Hatch Date		0.03	0.03	1	0.13	0.347
SampleTime*ALAN		-0.08	0.03	2	-0.43	0.669
IKBA	102	0.00	0.17	2	0.15	0.00)
(Intercept)	102	-4.77	1.69		-2.82	0.005
SampleTime(Noon)		-0.18	0.12	1	-1.54	0.125
Sex(M)		-0.20	0.09	1	-2.07	0.038
ALAN		-0.20	0.03	1	-0.28	0.038
Brood Size		-0.01	0.23	1	-0.20	0.777
Citronella(0)		-0.01	0.20	1	-0.20	0.843
Hatch Date		0.06	0.20	1	1.81	0.977
SampleTime*ALAN		-0.10	0.03	2	-0.61	0.070
BMAL1	94	-0.10	0.10	2	-0.01	0.541
(Intercept)	24	-2.62	1.44		-1.81	0.070
SampleTime(Noon)		-0.14	0.12	1	-1.17	0.070
Sex(M)		-0.14	0.12	1	-3.54	< 0.001
ALAN		0.03	0.09	1	0.16	0.876
Brood Size		-0.02	0.20	1	-0.33	0.870
		-0.02		1		
Citronella(0)			0.18		-0.62	0.537
Hatch Date		0.02	0.03	1	0.69	0.489
SampleTime*ALAN	101	-0.13	0.16	2	-0.81	0.419
CLOCK	101	0.60	0.20		2.27	0.022
(Intercept)		-0.69	0.30	1	-2.27	0.023
SampleTime(Noon)		-0.07	0.08	1	-0.91	0.362
Sex(M)		-0.12	0.06	1	-2.10	0.036
ALAN		-0.17	0.12	1	-1.40	0.161
Brood Size		-0.02	0.03	1	-0.79	0.428
Citronella(0)		0.04	0.10	1	0.36	0.715
Hatch Date		-0.01	0.01	1	-3.10	0.002
SampleTime*ALAN		-0.02	0.11	2	-0.22	0.829

Variable	N	Estimate	s.e.m	d.f.	<i>t</i> -value	<i>p</i> -value LRT
CRY1	109					
(Intercept)		0.36	0.24		1.46	
SampleTime(Noon)		-0.05	0.03	1	-1.54	
Sex(M)		-0.04	0.02	1	-1.62	0.103
ALAN		-0.05	0.04	1	-1.23	
Brood Size		0.01	0.01	1	0.77	0.374
Citronella(0)		-0.01	0.03	1	-0.41	0.638
Hatch Date		<0.01	0.01	1	0.14	0.912
SampleTime*ALAN		< 0.01	0.05	2	-0.04	0.955
CK1E	93					0.000
(Intercept)		-4.65	1.59	1	-2.93	0.003
SampleTime(Noon)		-0.20	0.13	1	-1.49	0.137
Sex(M)		-0.25	0.10	1	-2.49	0.013
ALAN		-0.03	0.23	1	-0.12	0.908
Brood Size		-0.02	0.06	1	-0.40	0.692
Citronella(0)		-0.04	0.19	1	-0.19	0.850
Hatch Date		0.06	0.03	1	1.88	0.061
SampleTime*ALAN	02	-0.14	0.20	2	-0.74	0.461
AANAT	92	4.25	1.75		2.40	0.012
(Intercept)		-4.35	1.75	1	-2.49	0.013
SampleTime(Noon)		-0.21	0.14	1	-1.46	0.145
Sex(M)		-0.21	0.12	1	-1.78	0.075
ALAN		-0.20	0.25	1	-0.82	0.411
Brood Size		-0.02	0.06	1	-0.37	0.715
Citronella(0)		-0.05	0.21	1	-0.23	0.816
Hatch Date		-0.06	0.04	1	1.58	0.113
SampleTime*ALAN	100	-0.11	0.21	2	-0.55	0.585
NKRF	108	4.25	1.75		2.40	0.012
(Intercept)		-4.35 -0.21	1.75 0.14	1	-2.49 -1.46	0.013 0.145
SampleTime(Noon)		-0.21	0.14	1	-1.40	0.143
Sex(M) ALAN		-0.21	0.12	1	-0.82	0.073
Brood Size		-0.02	0.06 0.21	1	-0.37 -0.23	0.715 0.816
Citronella(0) Hatch Date		-0.05	0.21	1	1.58	0.810
		0.06 -0.11	0.04	1 2	-0.55	0.113
SampleTime*ALAN NRF2	110	-0.11	0.21	2	-0.33	0.383
(Intercept)	110	14.67	3.72		3.95	
SampleTime(Noon)		0.03	0.28	1	0.09	
Sex(M)		1.02	0.28	1	4.65	< 0.001
ALAN		0.07	0.22	1	0.13	\0.001
Brood Size		0.07	0.32	1	1.92	0.040
Citronella(0)		0.26	0.14	1	0.58	0.040
Hatch Date		-0.22	0.43	1	-2.93	0.003
SampleTime*ALAN		-0.22	0.08	2	-2.93 -0.97	0.003
PER2	105	-0.57	0.41	2	-0.91	0.300
(Intercept)	103	-4.19	1.82		-2.30	0.021
SampleTime(Noon)		-0.07	0.11	1	-0.63	0.531
Sex(M)		-0.13	0.09	1	-1.47	0.331
ALAN		-0.13	0.09	1	-1.47	0.142
Brood Size		<0.01	0.24	1	-0.07	0.943
Citronella(0)		-0.13	0.07	1	-0.61	0.545
Hatch Date		0.06	0.22	1	1.64	0.101
SampleTime*ALAN		-0.07	0.04	2	-0.40	0.101
bampierinie ALAN		-0.07	0.10	<i>_</i>	-0.40	0.000

4.4.1.



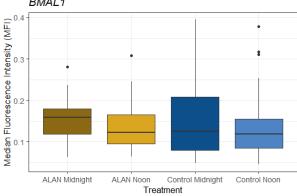


4.4.2.

CRY1 (IW) ALAN Midnight ALAN Noon Control Midnight Control Noon Treatment

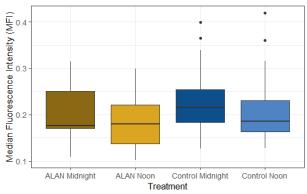
4.4.3.

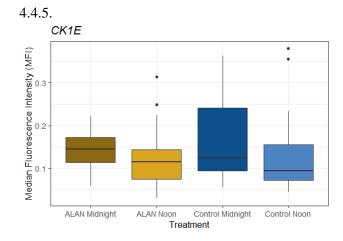
BMAL1



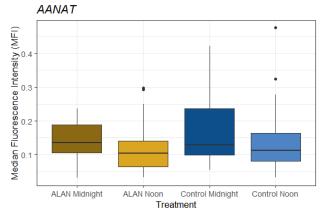
4.4.4.

CLOCK





4.4.6.



4.4.7.

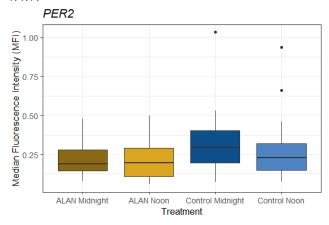
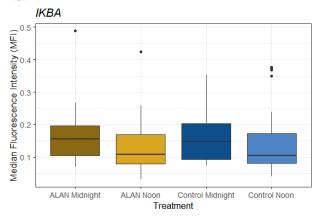


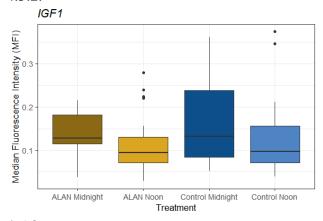
Figure 4.4. Artificial light at night and control treatment and transcript levels of circadian clock gene targets. 1. *REVERBA*, 2. *CRY1*, 3. *BMAL1*, 4. *CLOCK*, 5. *CK1E*, 6. *AANAT*, 7. *PER2*.

Colours indicate treatment and time of day of sampling: ALAN Midnight (Dark yellow), ALAN Noon (Light yellow), Control Midnight (Dark blue), Control Noon (Light blue). Data are shown by sample time at NOON (10:00-15:00 GMT) and NIGHT (23:00-02:00 GMT). Transcript levels are measured in Median Fluorescence Intensity Units (MFI). Box plots show the interquartile range of the raw data, where boxes above and below the medians (horizontal lines) show the first and third quartiles, respectively. Boxplot whiskers extend to 1.5 times the interquartile range, and dots represent outliers.

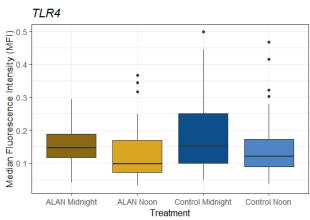
4.5.1.



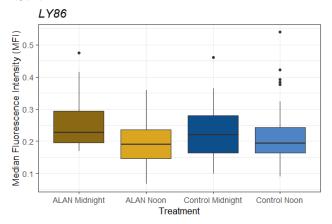
4.5.2.



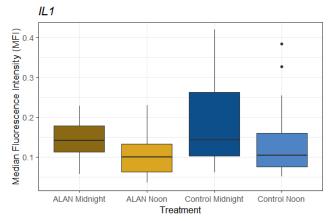
4.5.3.



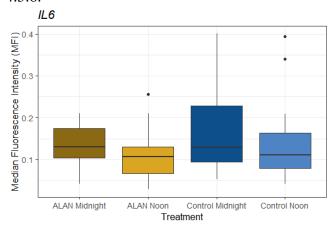
4.5.4.



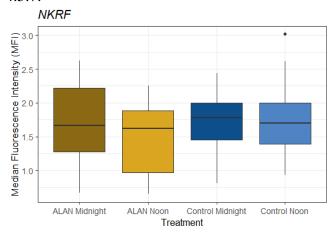
4.5.5.



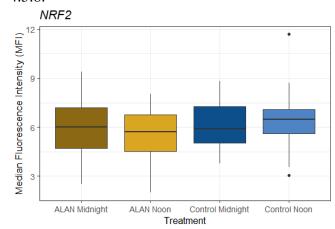
4.5.6.



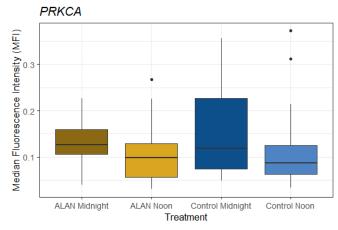
4.5.7.



4.5.8.



4.5.9.



4.5.10

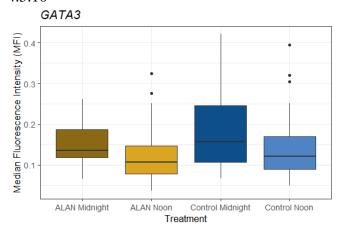


Figure 4.5. Artificial light at night (ALAN) and control treatments and transcript levels of immune gene targets. 1. *IKBA*, 2. *IGF1*, 3. *TLR4*, 4. *LY86*, 5. *IL1*, 6. *IL6*, 7. *NKRF*, 8. *NRF2*, 9. *PRKCA*, 10. *GATA3*.

Colours indicate treatment and time of day of sampling: ALAN Midnight (Dark yellow), ALAN Noon (Light yellow), Control Midnight (Dark blue), Control Noon (Light blue). Data are shown by sample time at NOON (10:00-15:00 GMT) and NIGHT (23:00-02:00 GMT). Transcript levels are measured in Median Fluorescence Intensity Units (MFI). Box plots show the interquartile range of the raw data, where boxes above and below the medians (horizontal lines) show the first and third quartiles, respectively. Boxplot whiskers extend to 1.5 times the interquartile range, and dots represent outliers.

4.5 Discussion

4.5.1 Effects of ALAN on nestling condition

In a previous experimental nest box study on great tits, exposure to ALAN halted nestling growth (Raap et al., 2016a). Here, ALAN treatment did not halt growth, yet nestlings under treatment were significantly lighter than control birds at day thirteen. This difference in result between the two studies may be explained by experimental design; Raap et al., (2016a) exposed nestlings to ALAN treatment for two consecutive nights on days thirteen to fifteen, whereas in this study, nestlings were exposed for a longer period during the nestling growth phase. Nestlings in this study were also weighed on day thirteen, which was the beginning of light treatment in Raap et al., (2016a). It is possible that effects of ALAN depend on stage of growth, where exposure prior to day thirteen of life has more impact on nestling weight than after day thirteen. The light intensity of ALAN treatment used by Raap et al., (2016a) was also three times (3 lux) that used in this study (1 lux), and therefore resulting effects on nestling weight may have been stronger. Given that small reductions in nestling weight have been demonstrated to negatively affect prospects for recruitment into the population post-fledging (Both et al., 1999) and for survival (Ringsby et al., 1998), observed weight differences as a consequence of ALAN exposure may have significant negative impacts on avian life histories.

The tarsus length and fledging success of nestlings was not affected by ALAN treatment as per a previous experimental study on great tit nestlings (Grunst et al., 2019). The lack of observable effect on fledgling was perhaps due to high level of fledging success overall at the study site. Sex effects on both the weight and tarsus length of nestlings were as expected, as male great tits are structurally larger than females (Gosler, 1993).

Nestlings under ALAN treatment did not have significantly different levels of haematocrit to control birds. Haematocrit levels in birds are highly variable, and may have been affected by many features at study sites, such as ambient temperatures (Fair et al., 2007).

4.5.2 Effects of ALAN on transcript levels of clock and immune genes

Previous studies on effects of ALAN on gene expression have shown either downregulations of clock and immune genes (Bedrosian et al., 2013, 2011), or an elevation of aspects of immune activity (Raap et al., 2016b; Saini et al., 2019). Here, it was expected that transcript levels of immune and clock genes would be altered by ALAN. Suppressive effects were found for *REVERBA*, a gene involved in the core pathway of the circadian clock as part of the *REVERBA/ROR* response element, which generates expression of *BMAL1*, and in antiphase to *PER2* (Ueda et al., 2002). It is not clear in this study why *REVERBA* was suppressed by ALAN more than in circadian clock genes of the same pathway, such as *BMAL1*. However, given that *REVERBA* plays an important role in glucose homeostasis (Yin et al., 2007), lipid metabolism (Raspe et al., 2002) and adipogenesis (Fontaine et al., 2003), its observed suppression by ALAN may have negative impacts on nestling physiology.

For most genes, transcripts levels significantly differed between day and night. However, ALAN treatment had no interaction with time of day differences in transcript levels, and therefore in this study there was no evidence of circadian disruption by ALAN. It may be that the low intensity of light used in this study (1 lux) may not have been sufficiently bright to disrupt rhythmic transcript levels of genes. In a study on zebra finches, (*Taeniopygia guttata*), the rhythmic expression of pro and anti-inflammatory cytokines was lost under constant bright light exposure, but not under dim light of 3 lux (Mishra et al., 2019). In another study on Siberian hamsters (*Phodopus sungorus*), dim light at night (5 lux) reduced peaks of daily expression of circadian clock proteins PER2 and BMAL1 in the hypothalamus and SCN (Bedrosian et al., 2013). In both studies, dim light treatment was of higher intensity than the ALAN treatment used here. Given that avian responses to ALAN are light-intensity dependent for behaviour (de Jong et al., 2016) reproductive biology (Dominoni et al., 2018; Zhang et al., 2019) and the endocrine system (Ouyang et al., 2018), it is likely that transcript level responses are also dependent on the intensity of light. Future studies investigating the effects of ALAN on the rhythmicity of avian immune and circadian clock systems should incorporate varying intensities of light treatment to elucidate whether transcript level responses are light intensity dependent.

In this study, there appeared to be no difference in how positive clock elements (e.g. *CLOCK*, *BMAL1*) were affected by ALAN treatment, compared with negative clock elements (*PER2*, *CRY1*). Given that Bedrosian et al., (2013) found ALAN treatment to reduce peaks of proteins found in both the positive circadian clock loop (*BMAL1*) and

negative clock loop (*PER2*), it is possible that ALAN may affect all aspects of the circadian clock in a similar fashion. Indeed, in this study, all circadian clock gene targets (*REVERBA*, *BMAL1*, *CLOCK*, *CRY1*, *CK1E*, *AANAT*, *PER2*) were downregulated by ALAN treatment, although all except *REVERBA* were non-significant effects.

This study makes use of samples taken during two time windows, at noon and at midnight. A potential limitation of this was that "noon" in this case was not exactly "noon" but a time window of a few hours (10:00-15:00). The nature of sampling transcript levels within these two time windows meant that the peak of expression of several genes (e.g. *LY86*, *REVERBA*, *CLOCK*, *NKRF*, *NRF2*) may have occurred for example at 18:00, and therefore the phase would have been missed. As a result, the circadian profile of these genes is unknown. Therefore, it is not possible to say that the lack of day-night differences in levels of these gene targets is due to circadian disruption. This gap in view of the "true" circadian profiles may also explain why circadian clock genes involved in the positive feedback loop (such as *BMAL1* and *CLOCK*) did not appear to be in the opposite phase to negative clock elements (*PER2*, *CRY1*), as both elements may have been in "mid-phase" during sampling. Future studies should incorporate a time profile approach (as in Chapter Six), taking blood samples from nestlings throughout the day and night in order to capture fluctuations in transcript levels of each circadian gene across 24 hr.

For many immune genes in this study, there were significant effects of sex, where female nestlings had higher transcript levels for many genes (*IGF1*, *TLR4*, *IL1*, *IL6*, *PRKCA*, *IKBA*) than males. Previous research of immunity in free-living juvenile birds has shown male nestlings to be less immunocompetent than females (Fargallo et al., 2002; Tschirren et al., 2003). These sex-specific differences in immunocompetence of nestlings are often explained by the modulating effects of sex hormones (Gaillard and Spinedi, 1998). Additionally, levels of several clock gene transcripts (*BMAL1*, *CRY1*, *CK1E*) also appeared to be lower in male nestlings than in female nestlings. In songbird species, there is a sexual dimorphism in the number of melatonin binding sites in telencephalic region of the brain (Aste et al., 2001). Given the role of melatonin in regulation of the avian circadian system (Gwinner et al., 1997), this dimorphism may explain observed differences in levels of circadian transcripts between male and female nestlings in this study.

A key strength of this study was that it was undertaken in the natural environment. However, the wild nature of this study may have made it difficult to capture transcript level responses to ALAN, given that gene expression in wild animals is highly variable in comparison to captive model species (Alvarez et al., 2015). Effects of ALAN were most

conspicuous for *REVERBA*, which interestingly, was one of the few genes in this study without an effect of time of day or sex. For this gene, there may have been less noise from other factors in statistical models, and therefore the effects of ALAN were more evident.

4.5.3 Conclusions

Ultimately, in this study, there was no indication that nestlings under ALAN treatment were in poor condition or had reduced fledging success. However, nestlings under ALAN treatment did weigh less than control birds, which may have fitness consequences post-fledging. This study also provided evidence that even low-levels of light at night may suppress elements of the avian circadian clock mechanism. This was an important new addition to previous studies that have shown shifts in activity (de Jong et al., 2016; Spoelstra et al., 2018) and physiology (Raap et al., 2016b; Saini et al., 2019) as a response to experimental doses of ALAN. Suppression of clock system elements may have consequences for wild birds, particularly during fitness challenges such as parasitic infection.

Finally, a novel aspect of this study was that it showed that it was possible to capture time of day differences in transcript levels of genes involved in circadian and immune pathways in wild animals. This study opened exciting avenues for further research into the effects of light at night on wildlife, and for field chronobiology.

Chapter Five: Avian malaria infection and condition in wild nestlings at city and forest sites

5.1 Introduction

Parasites impose significant costs for the fitness and reproduction of wild birds (Hamilton and Zuk, 1982). Avian malaria is caused by Haemosporidian parasites belonging to genera *Leucocytozoon*, *Haemoproteus* and *Plasmodium*, and is globally distributed except for Antarctica. Given their wide distribution and ability to infect bird species from every order examined to date (Lapointe et al., 2012), avian malaria parasites provide ideal systems for studies of host-parasite interactions and life history trade-offs (Christe et al., 2012; Coon et al., 2016; Marzal et al., 2005). Molecular techniques to determine the presence or absence of parasites in samples of DNA obtained from blood (Hellgren et al., 2006), or direct quantification of parasites from examination of blood smears (Garnham, 1966), have facilitated ecological studies on the impacts of avian malaria infections on wild birds.

All avian malaria parasites have similar life cycles, as shown in Figure 5.1., (Atkinson and Van Riper, 1991). Dipteran vectors (e.g. *Culex, Aedes* and *Culiseta* mosquitoes, *Simuliidae* blackflies and *Culicoides* biting midges) introduce infective-stage sporozoites to the avian host via blood meal. Following successful transmission, parasites migrate to liver cells of the host, and mature into schizonts, which then rupture and release merozoites into the host bloodstream. Merozoites then infect circulating host blood cells and undergo further asexual replication, maturing into schizonts that rupture host cells and release new merozoites. Gametocytes then develop inside host erythrocytes and are taken up by blood feeding insect vectors. The gametocytes then reproduce sexually in the vectors' gut (Beier, 1998), producing infective sporozoites that migrate to salivary glands to transmit to another host.

Hosts experience high levels of parasitaemia during the acute stages of infection, and this is when most pathologies manifest (Williams, 2005). Most early-stage infections result in a launch of host defences via innate, non-specific immunity, with a large number of immune system processes activated during peak parasitaemia, and a decline in activity shortly after the acute phase of infection (Videvall et al., 2015). Following the acute phase, infections may be cleared, or else pass into a chronic stage, which can last for months or years before

the host immune system is able to clear the infection, or persist within the host throughout their lifetime (Bishop et al., 1938).

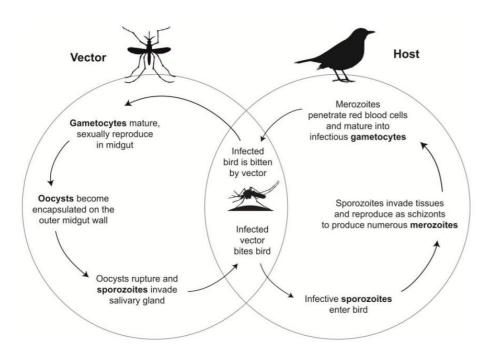


Figure 5.1. Avian malaria parasites life cycle. Diagram adapted from (Atkinson and Van Riper, 1991).

Effects of avian malaria on the host can range from asymptomatic, life-long infections (Bensch et al., 2007) to debilitating effects of acute infection such as anaemia (Palinauskas et al., 2015). In extreme cases, acute infection can result in cerebral ischaemia and mortality (Atkinson et al., 1995; Ilgunas et al., 2019). Responses to infection with avian malaria vary depending on the parasite lineage (Ricklefs and Fallon, 2002), host species (Ilgunas et al., 2019) and degree to which parasites and host are co-evolved (Jenkins et al., 2015). Infections in previously malaria-naïve birds can cause population level impacts, such as the case of parasite *Plasmodium relictum*, which has caused declines and extinctions of endemic bird species on the islands of Hawaii (Atkinson et al., 1995; van Riper et al., 1986), and the deaths of captive non-native birds in zoos (Brossy, 1992). Lowintensity chronic infections may have negative consequences in terms of reproductive success and lifespan (Asghar et al., 2015, 2011; Knowles et al., 2010a). However, other studies have reported no effects of chronic infection on avian fitness (Hahn et al., 2018; Kilpatrick et al., 2006). In wild studies, malaria infection presents a trade-off for host life histories. Studies investigating effects of *Leucocytozoon* infection on life history traits have shown both adverse effects of infection on parental condition and reproductive success (Merino et al., 2000), and an increase in reproductive success (Pigeault et al., 2018).

Many environmental factors affect the abundance and distribution of malaria vectors within a habitat, including seasonal changes and proximity to bodies of water (Chahad-Ehlers et al., 2018; Cosgrove et al., 2008; De Aguilar et al., 2018; Krama et al., 2015). The larvae of different vector families have different developmental needs; for example mosquitoes develop in standing water (Townroe and Callaghan, 2014), while blackfly larvae require clean running water in order to filter feed (Carlsson, 1967; Craig and Chance, 1982). Urban environments typically provide a less ideal habitat for vectors (Bradley and Altizer, 2007; Townroe and Callaghan, 2014), and therefore corresponding lower vector densities in cities may reduce the risk of avian malaria infection for urban bird populations (Bailly et al., 2016; Ferraguti et al., 2016). In an urban study on blackbirds (Turdus merula), there were large reductions in the prevalence of Haemoproteus and *Plasmodium* when compared with forest sites (Evans et al., 2009). However, prevalence of *Plasmodium* in a suburban population of house sparrows was recently shown to be 74%, a higher level than any population of wild birds in Northern Europe (Dadam et al., 2019), likely driven by the increase in vector abundance and environmental conditions favouring mosquito reproduction.

In the city environment, birds must contend with poorer food availability and quality than their forest conspecifics (Pollock et al., 2017), and an increase in environmental pollutants (Isaksson et al., 2009). Previous studies on malaria infections in city birds have indicated infected birds to have reductions in plumage colouration (Hõrak et al., 2001; Jacquin et al., 2011), a lower tolerance for pollutants (Bichet et al., 2013), higher leukocyte counts (Fokidis et al., 2008) and reductions in body condition (Jiménez-Peñuela et al., 2019). Immune responses to infections are energetically and nutritionally costly (Lochmiller and Deerenberg, 2000). City birds may be in poorer condition and have significantly reduced capacity for resisting infections and increased related fitness costs than forest birds. In addition, sleep deprivation and increased stress from exposure to light at night may affect city birds' ability to resist infections (Ouyang et al., 2017).

To date, many ecological studies comparing effects of avian malaria infection on city and forest birds have focused on chronic infections in adult birds (Bichet et al., 2013; Hõrak et al., 2001). However, malaria infections can occur in nestlings (Capilla-Lasheras et al., 2017; Krams et al., 2013a; Pollock et al., 2017). In a study on great tits at a forest site, nestlings infected with malaria parasites had lower survival rates and lower levels of haemoglobin (Krams et al., 2013a). Given that city birds may have reduced capacity to

resist infection than their forest conspecifics, and early infections may have consequences for life histories, it is important to consider the effects of acute malaria infection on urban nestling birds.

5.2 Study aims and hypotheses

The aim of this study was firstly to quantify prevalence of avian malaria infections in great tit nestlings at two city and two forest sites. An existing nest box system was used (Capilla-Lasheras et al., 2017; Pollock et al., 2017; Woodford et al., 2018), and the study was repeated over two years to account for between-year variation in vector and parasite abundance.

The second aim of this study was to investigate impacts of acute infection with avian malaria on nestling condition. To do this, the condition of infected and non-infected individuals was determined using common parameters for nestling condition such as weight and tarsal length (Eck et al., 2011), and fledging success. To determine pathology associated with infection, such as anaemia, red blood cell count (haematocrit levels) were measured as in Marzal et al., (2008).

Previous studies have captured responses of the avian immune system to acute infections under captive conditions (Videvall et al., 2015). However, to date, there have been few investigations of immune systems responses to malaria infections in the wild (Capilla-Lasheras et al., 2017). This study aimed to capture responses to acute infection by comparing transcript levels of genes involved in the inflammatory and immune responses in infected and non-infected individuals. Gene targets included *LY86* (lymphocyte antigen 86) and *TLR4* (toll-like receptor 4) which are involved in anti-bacterial and anti-malarial responses (Medzhitov, 2001), and the type 2 transcription factor *GATA3* (GATA binding protein 3) involved in adaptive immunity (Wang et al., 2011), *IL1* and *IL6* (interleukins 1 and 6) which are inflammatory response mediators (Klasing, 1998).

Additional gene targets involved in growth, oxidative stress and metabolism were included to investigate differences in transcript levels between city and forest birds. *PRKCA* (protein kinase C alpha), an enzyme responsible for antiviral effects and cell growth regulation (Clemens and Elia, 1997); and *IKBA* (NF-kappa-B inhibitor alpha) and *NKRF* (NF-kappa-B repressing factor), which are two inhibitors of the immune response regulator NF-kappa-B (Cabannes et al., 1999; Nourbakhsh and Hauser, 1999). Finally, immune targets *NRF2* (nuclear factor erythroid 2) and *IGF1* (insulin like growth factor 1) which are both involved in resistance to oxidative stress and metabolic regulation (Holzenberger et al., 2003; Kensler et al., 2007) were included in this study.

In this study, it was predicted that there would be lower overall prevalence of avian malaria at city sites than at forest sites, following previous studies (Evans et al., 2009; Jiménez-Peñuela et al., 2019). Nestlings infected with malaria were predicted to have significantly reduced condition, such as lower body weight and shorter tarsi, as resources are allocated to resisting infections instead of growth. This effect was expected to be stronger in city birds, where conditions are generally poorer (Pollock et al., 2017). To test this, an interaction between site and malaria infection was included in statistical models.

Fledging success of infected individuals was predicted to be lower than non-infected individuals, as in a previous study on malaria infection in great tit nestlings (Krams et al., 2013a). Given that anaemia is a common pathology of avian malaria infection (Palinauskas et al., 2015), it was also predicted that infected birds would have significantly reduced haematocrit levels (Krams et al., 2013a).

It was predicted that overall transcript levels of immune and inflammatory targets would be elevated for infected individuals at both city and forest sites, as in a previous transcriptome study (Videvall et al., 2015). Specifically, gene targets such as *GATA3* and *TLR4* were expected to be elevated in nestlings in city sites as in (Capilla-Lasheras et al., 2017). In addition, it was predicted that transcript levels of oxidative stress associated immune targets such as *NRF2* and *IGF1* would be higher for city nestlings than for their conspecifics in the forest (Isaksson et al., 2009; Watson et al., 2017).

5.3 Materials and Methods

5.3.1 Study sites

Data for this chapter were obtained between April and June of 2016 and 2017 at forest and city sites on free-living populations of breeding great tits. Field sites were existing forest nest box study systems located at the Scottish Centre for Ecology and the Natural Environment (SCENE; 56° 7.73'N, 4° 36.79'W), Cashel Forest (56° 6'N, 4°34'W), and city sites at Kelvingrove Park, Glasgow (55°52' N, 4°17'W) and Garscube estate (55° 9'N, 4°31' W). These sites were previously used in ecological studies of wild passerines and malaria vectors (Capilla-Lasheras et al., 2017; Pollock et al., 2017; Woodford et al., 2018). To determine hatch dates, brood sizes and fledging success, all boxes were initially checked weekly and, closer to hatching, every second day as part of a general nest box monitoring protocol (see General Methods for further details of nest box monitoring).

5.3.2 Nestling processing

All nestlings (N = 174) were ringed for individual identification on day thirteen of life. Measurements were taken of nestling weight and length of the tarsometatarsus bone (tarsus) from each nestling. Samples were only taken during the noon hours (11:00-15:00 GMT) to exclude the possibility of potential circadian effects on genes. From each nestling, two blood samples of approximately 50 µl were obtained via the brachial wing vein as in Owen (2011). The first was taken up by a heparinised capillary for haematocrit readings (in the study year 2017), sealed using a wax plug and stored on ice until return to the laboratory. A second blood sample was obtained and stored in 250 µl RNAlater® stabilising solution (Invitrogen) for molecular analyses. A breakdown of number of RNA samples collected for this study can be found in Table 5.1. See General Methods section for further details on the nestling blood sampling protocol and ethical licencing.

Table 5.1. Number of RNA samples collected for 2016 and 2017 at city and forest sites for analyses of transcript levels of immune target genes in this Chapter.

Study year	Site	Number of RNA samples obtained
2016	City	12
2016	Forest	51
2017	City	24
2017	Forest	41

As part of a vector study in 2017, eleven nest boxes were treated with a citronella oil-based insect repellent as per Krams et al., (2013), (see General Methods section). Initial modelling of the dataset via a generalised linear mixed-model revealed citronella treatment to have no significant effect on the prevalence of *Leucocytozoon* parasites in nestlings (GLM; $z_{1,78}$ =-0.68: P=0.488), and therefore citronella was dropped from further analyses to pool 2017 and 2016 datasets. In addition, nestlings from nest boxes that were experimentally exposed to artificial light at night treatment (in Chapter Four) were excluded from this study (See General Methods section for a breakdown of sample sizes for each chapter).

5.3.3 Ethical Statement

See General Methods for details of ringing licences and blood sampling.

5.3.4 Laboratory procedures

Haematocrit readings were taken as described in Chapter Four (See General Methods for more details on protocol for haematocrit readings). After haematocrit readings were taken, this sample was removed from the capillary via pipette and stored in 0.5 ml 100% ETOH for molecular analyses. Samples in 100% ETOH were frozen at -40°C. Genomic DNA was later extracted from these samples via DNAeasy Blood and Tissue Kit (Qiagen) and molecular sexing was carried out via PCR (See General Methods for details).

Measurement of Median Fluorescence Intensity units (as a proxy for transcript levels) for immune and housekeeping gene targets outlined in Table 5.2. was carried out on all samples stored in *RNAlater*® using the QuantiGene® Plex RNA 2.0 assay (ThermoFisher). The QuantiGene® Plex RNA 2.0 assay is a platform that allows for simultaneous readouts of transcript levels of many target genes from a single blood sample using branched DNA

signal amplification and multi-analyte profiling beads (xMAP[®]). The General Methods section provides further details on lab procedures carried out in this Chapter.

Table 5.2. Housekeeping and immune gene targets used within this study. See General Methods section for more details on gene functions, and Appendix 1 for primer information.

Transcript targets	Type
RPL19, SDHA, HMBS	Housekeeping genes
LY86, TLR4, IL1, IL6, PRKCA, GATA3, IKBA, NKRF, NRF2, IGF1	Immune genes

5.3.5 Prevalence testing for *Leucocytozoon*

A nested PCR method was used in this study to test for the presence of *Leucocytozoon*, *Haemoproteus* and *Plasmodium* DNA within avian DNA samples. This PCR method was outlined in the General Methods section.

Four birds were found to be infected with *Haemoproteus* and were excluded from the rest of this study due to small sample size. One bird was infected with both *Haemoproteus* and *Leucocytozoon* and was also excluded from this study. No birds were infected with *Plasmodium*.

5.3.6 Statistical Analysis and Model Selection

All statistical analyses were performed using R v. 3.4.4 (R Core Team, 2018) and packages *lme4* (Bates et al., 2015). Data were visualised using ggplot2 (Wickham, 2014). All global models included individual nest box as a random factor to account for repeated measures within the same nest box. Residuals were plotted and inspected to check assumptions of normality prior to modelling using linear mixed models (LMMs). Pairs of explanatory variables were also assessed for collinearity by calculating tolerance and variance inflation factors (VIF) and also by visual inspection of pairwise plots, before statistical relevance was accepted.

Leucocytozoon prevalence data were modelled using a binomial distribution, (1 = Leucocytozoon detected, 0 = no Leucocytozoon detected). The global generalised linear mixed model for Leucocytozoon prevalence included the following explanatory variables: nestling sex (M/F); study site (Forest/City site); brood size; hatch date (to look for seasonal differences in infections); and study year (2016/2017).

Data for nestling weight, haematocrit level and tarsus length were modelled using linear mixed models and Gaussian distributions. Global models included the following explanatory variables: *Leucocytozoon* infection (two-level fixed factor 0/1), sex (M/F), site (City/Forest), brood size, hatch date. All models included an interaction between *Leucocytozoon* infection and site. This interaction remained in the model regardless of significance.

Additionally, the model for nestling weight included study year (2016/2017). Site (City/Forest) was removed from the global model for haematocrit level, as there was only one infected city bird in the haematocrit dataset, and therefore insufficient power for the model to run. For weight, tarsus length and *Leucocytozoon* prevalence models, an interaction between site and hatch date was included, as city birds tend to nest earlier in the year than forest birds (Deviche and Davies, 2014). Interactions between year and hatch date and year and site were included in global models to account for year to year variation in infections.

Median fluorescence intensity data (as a proxy for transcript levels of genes) were obtained from the QuantiGene[®] Plex RNA assay 2.0 (Thermofisher). These data were normalised by division of the mean fluorescence of background wells and then dividing each sample by the geometrical mean of levels of housekeeping genes in Table 5.2, thus, transcript levels are expressed as relative levels to housekeeping genes. Transcript level data for all targets were then modelled using LMMs and Gamma distribution with log-link function. Global models for transcript level data contained the same explanatory variables as the model for nestling weight.

Statistical relevance of fixed factors within all models was determined by sequential model selection via Likelihood Ratio Testing using package *lmtest* (Zeileis and Hothorn, 2002), until the minimum adequate model (MAM) was found. Factors with a significance level of <5% were kept in the model. *Leucocytozoon* infection and site (City/Forest) was kept in final models regardless of significance.

5.4 Results

5.4.1 Leucocytozoon prevalence

In total, 174 nestlings were sampled and tested for presence of Leucocytozoon parasites. Leucocytozoon prevalence was significantly higher at forest sites than at city sites (GLM, $z_{1,174} = 2.00$; P=0.035). In 2016, Leucocytozoon prevalence at forest sites was 83.9% and city sites 33.33%. In 2017, Leucocytozoon prevalence was much lower than the previous year, with prevalence at forest sites 36.07% and city sites 11.76%, (Figure 5.2.).

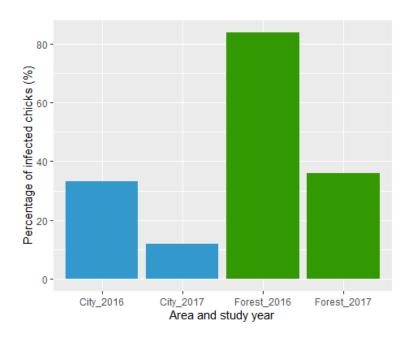


Figure 5.2. *Leucocytozoon* infection prevalence (%) at city (blue) and forest (green) sites for 2016 and 2017.

In addition, *Leucocytozoon* prevalence increased along with hatch date, and this was more evident in 2016 than in 2017 (GLM, $z_{1,174}$ =0.98, P=0.039), (Figure 5.3.). There was no effect of sex or brood size on prevalence of *Leucocytozoon* in nestlings (Table 5.3.).

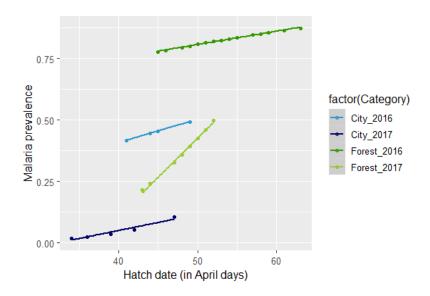


Figure 5.3. *Leucocytozoon* prevalence and hatch date (where April 1st=1) at city (blue) and forest sites (green) for 2016 and 2017. Prevalence data were obtained for each site and year using the minimum adequate model for *Leucocytozoon* prevalence.

The interaction between Leucocytozoon infection and site was non-significant in all models (Table 5.4). In addition, there was no effect of Leucocytozoon infection on the weight of nestlings (LMM; $t_{1,174}$ =0.43: P=0.055), (Figure 5.4.), or their tarsus length (LMM; $t_{1,67}$ =1.13: P=0.080), (Figure 5.5.). Effects of Leucocytozoon infection on haematocrit levels were also not significant (LMM: $t_{1,30}$ =-0.81: P=0.966). Fledging success was very high at both forest (94 %) and city sites (92 %), and due to little variation between sites, the effect of Leucocytozoon infection on fledging was not included in models.

Sex differences in morphology were apparent, where male nestlings were significantly heavier than females (LMM; $t_{1,174}$ =4.84, P<0.001), with mean weight of males 19.45 g (s.e. \pm 0.88) and females 18.6 g (\pm 0.71). In addition, nestlings at forest sites were significantly heavier than city nestlings (LMM; $t_{1,174}$ =4.93, P<0.001), where forest nestlings were 22.26 g (s.e. \pm 0.48) and city nestlings 20.34 g (s.e. \pm 1.27). There were no effects of brood size, hatch date or study year on the weight of nestlings (Table 5.3.).

The tarsus length and haematocrit levels of nestlings were not affected by either the brood size or hatch date (Table 5.3.). Male nestlings had significantly longer tarsi than females

(LMM; $t_{1,67}$ =3.55, P=0.002), where mean male tarsi were 20.01 mm (s.e. \pm 0.14), and females 19.53 mm (s.e. \pm 0.37). Differences in tarsi length of nestlings at city and forest sites were not significant (LMM; $t_{1,174}$ =1.77, P=0.098), (Figure 5.5.).

Global model summaries for nestling data are shown in Table 5.5.

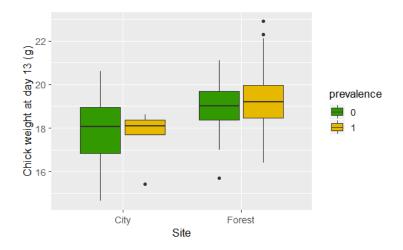


Figure 5.4. Nestling weight (g) at day thirteen and *Leucocytozoon* infection prevalence at city and forest sites. *Leucocytozoon* infection prevalence is indicated by colour: No infection (green), Infection present (yellow). Box plots show the interquartile range of the raw data, where boxes above and below the medians (horizontal lines) show the first and third quartiles, respectively. Boxplot whiskers extend to 1.5 times the interquartile range. Dots represent outliers.

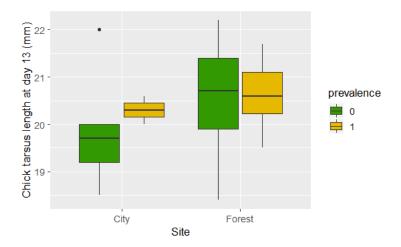


Figure 5.5. Tarsus length (mm) at day thirteen and *Leucocytozoon* infection prevalence at city and forest sites. *Leucocytozoon* infection prevalence is indicated by colour: No infection (green), Infection present (yellow). Box plots show the interquartile range of the raw data, where boxes above and below the medians (horizontal lines) show the first and third quartiles, respectively. Boxplot whiskers extend to 1.5 times the interquartile range. Dots represent outliers.

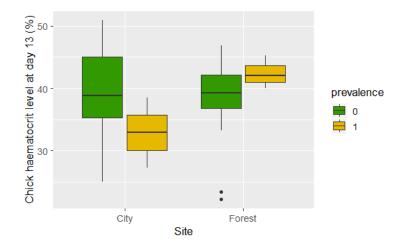


Figure 5.6. Haematocrit levels (%) at day thirteen and *Leucocytozoon* infection prevalence at city and forest sites. *Leucocytozoon* infection prevalence is indicated by colour: No infection (green), Infection present (yellow). Box plots show the interquartile range of the raw data, where boxes above and below the medians (horizontal lines) show the first and third quartiles, respectively. Boxplot whiskers extend to 1.5 times the interquartile range. Dots represent outliers.

Table 5.3. Global and minimum adequate models. *Leucocytozoon* infection and site (City/Forest) were always kept in the model, regardless of significance. *Leucocytozoon* prevalence was modelled using a Binomial distribution (0=no *Leucocytozoon* detected, 1=*Leucocytozoon* detected). Nestling weight, tarsus length and haematocrit data were modelled using Gaussian distributions. Nest box was always included as random factor. Addition of terms is denoted by a plus sign (+). Interaction of terms is denoted by an asterisk (*).

Variable	Global model	Final MA model
Leucocytozoon	Nest box + sex + brood size +	Nest box + site +
prevalence	site*hatch date + year*hatch date +	year*hatch date
	year*site	
Nestling	$Nest\ box + infection + sex + brood$	Nest box $+$ sex $+$
weight	size + hatch date*year + year*site +	site*infection
	site*hatch date + site*infection	
Tarsus length	$Nest\ box + infection + sex + brood$	Nest box $+$ sex $+$
(2017 only)	size + site*hatch date + site*infection	site*infection
Haematocrit	$Nest\ box + infection + sex + brood$	Nest box + infection +
% (2017 only)	size + hatch date + site*infection	hatch date + site*infection

Table 5.4. Minimum adequate model summaries. Estimates and standard error values for immune gene transcript levels are log transformed. Intercept was as follows: Site (City), Year (2016), Sex (F), Infection (0). Interaction of terms is denoted by an asterisk (*).

Variable	N	Estimate	s.e.m	d.f.	z-value	<i>p</i> -value
Leucocytozoon	174					
prevalence						
(Intercept)		-1.90	2.96	1	-0.64	
Site		1.42	0.71	1	2.00	0.035
Hatch date		0.04	0.06	1	0.63	
Year		-6.89	5.22	1	-1.32	
Year*Hatch date		0.10	0.11	2	0.98	0.039
Variable	N	Estimate	s.e.m	d.f.	t-value	<i>p</i> -value
Nestling weight	174					
(Intercept)		20.34	1.27		15.98	
Infection (1)		0.29	0.66	1	0.43	0.055
Sex (M)		0.86	0.17	1	4.93	<0.001
Site (Forest)		1.92	0.48	1	4.02	<0.001
Hatch Date		-0.07	0.03	1	-2.44	0.014
Infection*Site		0.10	0.68	2	0.15	0.884
Tarsus length	67					
(2017 only)						
(Intercept)		19.53	0.37		52.74	
Infection (1)		0.56	0.49	1	1.13	0.080
Sex (M)		0.48	0.14	1	3.55	<0.001
Site (Forest)		0.73	0.41	1	1.77	0.098
Infection*Site		-0.35	0.52	2	-0.68	0.472
Haematocrit %	30					
(2017 only)						
(Intercept)		68.69	15.87		15.87	
Infection (1)		-0.46	4.87	1	4.87	0.696
Hatch Date		-0.79	0.41	1	0.41	0.049
Site		7.20	4.59	1	4.59	0.058
Infection*Site		2.59	6.29	2	6.29	0.601

Table 5.5. Global model summaries. Estimates and standard error values for immune gene transcript levels are log transformed. Intercept was as follows: Site (City), Year (2016), Sex (F), Infection (0). Interaction of terms is denoted by an asterisk (*).

Variable	N	Estimate	s.e.m	d.f.	z-value	<i>p</i> -value
Leucocytozoon	174					
prevalence						
(Intercept)		-10.11	9.43	1	-1.07	
Sex(M)		0.44	0.39		1.12	0.264
Site(Forest)		12.86	9.52	1	1.35	0.177
Brood Size		-0.10	0.13	1	-0.75	0.454
Hatch date		0.21	0.20	1	1.04	0.296
Year(2017)		-9.06	6.07	1	-1.49	0.135
Site*Hatch date		-0.22	0.20	2	-1.10	0.270
Site*Year		0.21	0.13	2	1.60	0.110
Year*Hatch Date		-3.14	1.48	2	-2.12	0.034
Variable	N	Estimate	s.e.m	d.f.	t-value	<i>p</i> -value
Nestling weight	174					
(Intercept)		26.88	5.19		5.18	
Infection (1)		0.32	0.70	1	0.46	
Sex (M)		0.89	0.18	1	4.95	< 0.001
Site (Forest)		-3.17	5.16	1	-0.61	
Brood Size		-0.09	0.10	1	-0.95	0.310
Hatch Date		-0.19	0.11	1	-1.72	
Year(2017)		-5.96	4.23	1	-1.41	
Hatch Date*Year		0.10	0.09	2	1.13	0.214
Site*Year		0.95	1.05	2	0.91	0.318
Infection*Site		0.05	0.74	2	0.07	0.953
Site*Hatch Date		0.10	0.11	2	0.92	0.306
Tarsus length	67					
(2017 only)						
(Intercept)		18.97	3.54		5.35	
Infection (1)		0.48	0.57	1	0.85	
Sex (M)		0.48	0.14	1	3.44	< 0.001
Site (Forest)		5.86	5.35	1	1.10	
Brood Size		-0.05	0.12	1	-0.46	0.601
Hatch Date		0.02	0.09	1	0.23	

Infection*Site		-0.26	0.59	2	-0.44	0.634
Site*Hatch Date		-0.11	0.13	2	-0.85	0.346
Haematocrit %	37					
(2017 only)						
(Intercept)		66.38	18.75		3.54	
Infection (1)		-0.22	4.99	1	-0.04	
Sex(M)		0.16	2.41	1	0.07	0.912
Brood Size		0.39	1.20	1	0.33	0.687
Hatch Date		-0.77	0.44	1	-1.74	0.049
Site(Forest)		6.22	5.77	1	1.08	
Infection*Site		2.11	6.54	2	0.32	0.636

5.4.2 Transcript levels of immune genes

Infection with *Leucocytozoon* did not significantly affect the transcript levels of any of the immune targets tested in this study (Table 5.5.; Figure 5.7.). Moreover, the interaction between *Leucocytozoon* infection and site was found to be non-significant for all immune gene models (Table 5.6). No support was found for effects of brood size on transcript levels of all immune targets (Table 5.6.). For immune targets *IL1*, *PRKCA*, transcript levels increased along with later hatch date (Table 5.7.), and this trend was stronger in forest birds than city birds. For *NKRF*, transcript levels decreased as the season progressed (Table 5.7.).

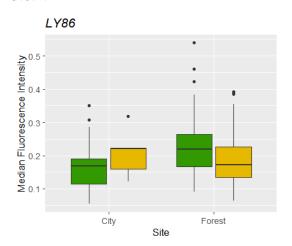
Transcript levels of immune genes at city and forest sites did not differ significantly for many of the targets tested (Table 5.6.). However, transcript levels of immune targets IL1, PRKCA, NKRF and NRF2 were significantly higher at forest sites than at city sites (Table 5.7.). Notably, levels of NKRF were found to be lower by 33 % in city birds than in forest birds (LMM; $t_{1,157}$ =3.66: P<0.001), (Figure 5.7.9.). The same was true for NRF2, (LMM; $t_{1,156}$ =2.75: P=0.006), where transcript levels in city birds were 18 % lower than those of forest birds (Figure 5.7.10.).

No support was found for effects of study year on the transcript levels of immune targets except for *IL1* (Table 5.6.), levels of which were significantly higher in birds in 2017 than in 2016 (LMM; $t_{1,136}$ =2.41: P=0.016). Similarly, no support was found for effect of sex on transcript levels aside from *NRF2*, with males estimated to have higher transcript levels than female birds (LMM; $t_{1,156}$ =2.75: P=0.006).

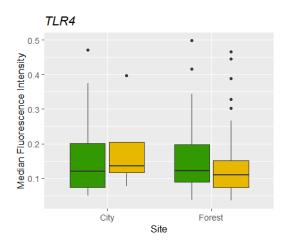
Global model summaries for immune gene data can be found in Table 5.8.

prevalence 🗯 0 븢 1

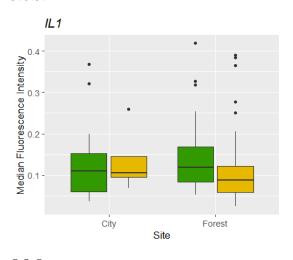
5.8.1.



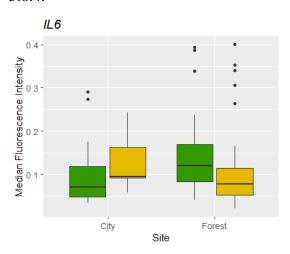
5.8.2.



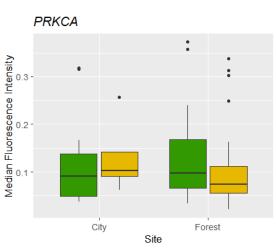
5.8.3.



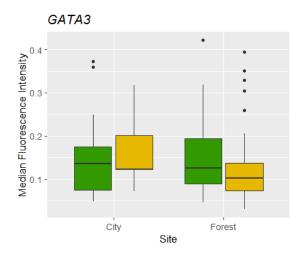
5.8.4.

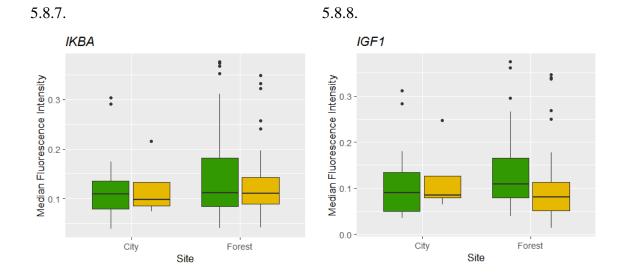


5.8.5.



5.8.6.





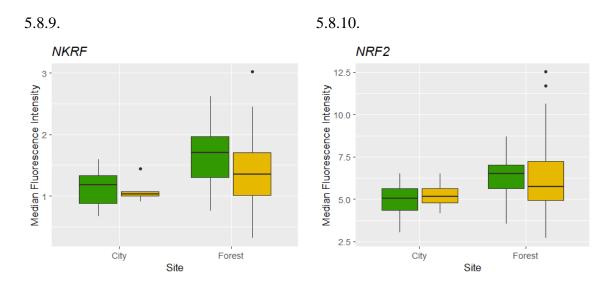


Figure 5.8. *Leucocytozoon* infection prevalence (where green = no infection, yellow = infection) and transcript levels (median fluorescence intensity) for immune gene targets 1) *LY86*, 2) *TLR4*, 3) *IL1*, 4) *IL6*, 5) *PRKCA*, 6) *GATA3*, 7) *IKBA*, 8) *IGF1*, 9) *NKRF*, 10) *NRF2*. Data shown are empirical values obtained from QuantiGene® Plex RNA assay and shown in Median florescence intensity units (MFI). Box plots show the interquartile range of the raw data, where boxes above and below the medians (horizontal lines) show the first and third quartiles, respectively. Boxplot whiskers extend to 1.5 times the interquartile range. Dots represent outliers.

Table 5.6. Global and minimum adequate models. Nest box was always included as random factor. An interaction between *Leucocytozoon* infection and site (City/Forest) was always kept in the model, regardless of significance. Immune transcript data were modelled using Gamma distributions with log link function. Addition of terms is denoted by a plus sign (+). Interaction of terms is denoted by an asterisk (*).

Variable	Global model	Final MA model
LY86	Nest box + infection + sex + brood size + hatch date*year + site*hatch date + infection*site	Nest box + infection*site
TLR4	Nest box + infection + sex + brood size + hatch date*year + site*hatch date + infection*site	Nest box + infection*site
IL1	Nest box + infection + sex + brood size + hatch date*year + site*hatch date + infection*site	Nest box + infection*site
IL6	Nest box + infection + sex + brood size + hatch date*year + site*hatch date + infection*site	Nest box + infection*site
PRKCA	Nest box + infection + sex + brood size + hatch date*year + site*hatch date + infection*site	Nest box + infection*site
GATA3	Nest box + infection + sex + brood size + hatch date*year + site*hatch date + infection*site	Nest box + infection*site
IKBA	Nest box + infection + sex + brood size + hatch date*year + site*hatch date + infection*site	Nest box + infection*site
IGF1	Nest box + infection + sex + brood size + hatch date*year + site*hatch date + infection*site	Nest box + infection*site
NKRF	Nest box + infection + sex + brood size + hatch date*year + site*hatch date + infection*site	Nest box + infection*site + hatch date
NRF2	Nest box + infection + sex + brood size + hatch date*year + site*hatch date + infection*site	Nest box + sex + infection*site

Table 5.7. Model summaries for MAMs. Estimates and standard error values for immune gene transcript levels are log-transformed. Reference level for the intercept: Site (City), Year (2016), Sex (F), Infection (0).

Variable	N	Estimate	s.e.m	d.f.	t-value	<i>p-</i> value LRT
LY86	155					
(Intercept)		-1.86	0.13		-14.06	
Infection (1)		0.20	0.24	1	0.84	0.403
Site (Forest)		0.28	0.15	1	1.83	0.068
Infection*Site		-0.31	0.25	2	-1.23	0.217
TLR4	138					
(Intercept)		-1.94	0.18		-10.68	
Infection (1)		-0.07	0.44	1	-0.16	0.874
Site (Forest)		0.02	0.21	1	0.08	0.940
Infection*Site		-0.13	0.46	2	-0.28	0.778
IL1	136					
(Intercept)		-5.90	1.33		-4.42	
Infection (1)		-0.42	0.43	1	-0.97	0.331
Site (Forest)		3.82	1.49	1	2.57	0.010
Hatch Date		0.08	0.03	1	2.71	0.007
Year(2017)		0.40	0.17	1	2.41	0.016
Infection*Site		0.35	0.44	2	0.79	0.432
Site*Hatch Date		-0.09	0.03	2	-2.67	0.008
IL6	130					
(Intercept)		-2.40	0.19		-12.71	
Infection (1)		0.22	0.35	1	0.64	0.523
Site (Forest)		0.26	0.22	1	1.15	0.251
Infection*Site		-0.47	0.37	2	-1.25	0.210
PRKCA	128					
(Intercept)		-5.12	1.35		-3.78	
Infection (1)		-0.40	0.44	1	-0.90	0.366
Site (Forest)		3.29	1.64	1	2.00	0.045
Hatch Date		0.07	0.03	1	2.10	0.036
Infection*Site		0.24	0.46	2	0.53	0.595
Site*Hatch Date		-0.08	0.04	2	-2.07	0.038
GATA3	142				. • •	
(Intercept)	. 12	-2.05	0.16	1	-12.95	
Infection (1)		0.10	0.31	1	0.34	0.735
Site (Forest)		0.10	0.19	1	0.03	0.733
Infection*Site		-0.25	0.32	2	-0.78	0.434

Variable	N	Estimate	s.e.m	d.f.	t-value	<i>p-</i> value LRT
IKBA	148					
(Intercept)		-2.18	0.15		-14.77	
Infection (1)		-0.13	0.34	1	-0.37	0.710
Site (Forest)		0.10	0.17	1	0.58	0.563
Infection*Site		0.07	0.35	2	0.19	0.851
IGF1	135					
(Intercept)		-2.26	0.19		-11.72	
Infection (1)		-0.05	0.43	1	-0.10	0.917
Site (Forest)		0.07	0.22	1	0.32	0.746
Infection*Site		-0.17	0.45	2	-0.39	0.697
NKRF	157					
(Intercept)		1.04	0.43		2.43	
Infection (1)		0.17	0.18	1	0.95	0.342
Site (Forest)		0.58	0.16	1	3.66	<0.001
Hatch Date		-0.02	0.01	1	-2.38	0.017
Infection*Site		-0.23	0.19	2	-1.25	0.212
Variable	N					
NRF2	156					
(Intercept)		1.52	0.08		18.48	
Infection (1)		0.17	0.16	1	1.02	0.310
Sex (M)		0.10	0.03	1	2.95	0.003
Site (Forest)		0.26	0.09	1	2.75	0.006
Infection*Site		-0.19	0.17	2	-1.12	0.261

Table 5.8. Global model summaries. Estimates and standard error values for immune gene transcript levels are log-transformed. Reference level for the intercept: Site (City), Year (2016), Sex (F), Infection (0).

Variable	N	Estimate	s.e.m	d.f.	t-value	<i>p</i> -value LRT
LY86	155					
(Intercept)		-5.49	1.91		-2.87	
Infection (1)		0.07	0.24	1	0.29	0.774
Sex(M)		0.04	0.06	1	0.71	0.477
Site (Forest)		3.79	1.82	1	2.08	0.037
Brood Size		0.01	0.04	1	0.33	0.738
Hatch Date		0.08	0.04	1	1.90	0.057
Year(2017)		1.98	1.67	1	1.19	0.235
Infection*Site		-0.16	0.25	2	-0.64	0.524
Hatch Date*Year		-0.04	0.03	2	-1.12	0.264
Site*Hatch Date		-0.08	0.04	2	-1.99	0.046
TLR4	138					
(Intercept)		-5.77	2.61		-2.21	
Infection (1)		-0.48	0.46	1	-1.05	0.295
Sex(M)		-0.10	0.10	1	-0.99	0.322
Site (Forest)		2.71	2.45	1	1.11	0.269
Brood Size		0.05	0.05	1	0.98	0.329
Hatch Date		0.09	0.06	1	1.51	0.130
Year(2017)		0.40	2.19	1	0.18	0.855
Infection*Site		0.32	0.47	2	0.68	0.495
Hatch Date*Year		0.01	0.05	2	-0.09	0.931
Site*Hatch Date		-0.07	0.05	2	-1.31	0.191
IL1	136					
(Intercept)		-6.56	2.34		-2.80	
Infection (1)		-0.47	0.43	1	-1.11	0.268
Sex(M)		-0.12	0.09	1	-1.30	0.194
Site (Forest)		4.24	2.18	1	1.95	0.052
Brood Size		0.04	0.05	1	0.78	0.433
Hatch Date		0.10	0.05	1	1.91	0.056
Year(2017)		1.08	1.97	1	0.55	0.585
Infection*Site		0.41	0.44	2	0.93	0.350
Hatch Date*Year		-0.01	0.04	2	-0.36	0.717
Site*Hatch Date		-0.10	0.05	2	-1.20	0.035

Variable	N	Estimate	s.e.m	d.f.	t-value	<i>p-</i> value LRT
IL6	130					LICI
(Intercept)		-4.21	2.43		-1.73	
Infection (1)		0.13	0.35	1	0.36	0.720
Sex(M)		-0.12	0.10	1	-1.19	0.236
Site (Forest)		1.65	2.26	1	0.73	0.465
Brood Size		0.01	0.05	1	-0.08	0.932
Hatch Date		0.04	0.05	1	0.84	0.400
Year(2017)		-0.60	2.09	1	-0.29	0.773
Infection*Site		-0.30	0.37	2	-0.81	0.418
Hatch Date*Year		0.02	0.04	2	0.47	0.635
Site*Hatch Date		-0.04	0.05	2	-0.76	0.449
PRKCA	128					
(Intercept)		-6.40	2.67		-2.40	
Infection (1)		-0.44	0.45	1	-0.98	0.328
Sex(M)		-0.13	0.10	1	-1.28	0.202
Site (Forest)		3.55	2.50	1	1.42	0.156
Brood Size		0.01	0.05	1	0.15	0.882
Hatch Date		0.10	0.06	1	1.65	0.099
Year(2017)		0.85	2.23	1	0.38	0.702
Infection*Site		0.34	0.47	2	0.73	0.468
Hatch Date*Year		-0.01	0.05	2	-0.27	0.785
Site*Hatch Date		-0.09	0.06	2	-1.56	0.118
GATA3	142					
(Intercept)		-5.11	2.13		-2.40	
Infection (1)		-0.04	0.30	1	-0.14	0.893
Sex(M)		0.02	0.08	1	0.25	0.801
Site (Forest)		2.72	1.99	1	1.37	0.170
Brood Size		0.02	0.04	1	0.46	0.643
Hatch Date		0.07	0.05	1	1.46	0.146
Year(2017)		0.54	1.83	1	0.29	0.769
Infection*Site		-0.06	0.32	2	-0.18	0.854
Hatch Date*Year		-0.01	0.04	2	-0.14	0.890
Site*Hatch Date		-0.07	0.04	2	-1.50	0.133
IKBA	148					
(Intercept)		-3.73	2.16		-1.73	
Infection (1)		-0.33	0.36	1	-0.92	0.356
Sex(M)		-0.06	0.07	1	-0.86	0.388
Site (Forest)		1.58	2.05	1	0.77	0.442
Brood Size		0.03	0.04	1	0.58	0.559
Hatch Date		0.04	0.05	1	0.76	0.447

Year(2017)		-0.28	1.84	1	-0.15	0.881
Infection*Site		0.29	0.37	2	0.80	0.426
Hatch Date*Year		0.01	0.04	2	0.21	0.831
Site*Hatch Date		-0.04	0.05	2	-0.84	0.402
IGF1	135					
(Intercept)		-5.40	2.52		-2.15	
Infection (1)		-0.41	0.46	1	-0.89	0.371
Sex(M)		-0.13	0.10	1	-1.30	0.192
Site (Forest)		2.57	2.33	1	1.10	0.270
Brood Size		0.02	0.05	1	0.42	0.675
Hatch Date		0.07	0.05	1	1.32	0.188
Year(2017)		0.18	2.09	1	0.08	0.932
Infection*Site		0.27	0.48	2	0.56	0.579
Hatch Date*Year		0.01	0.04	2	0.09	0.931
Site*Hatch Date		-0.06	0.05	2	-1.24	0.217
NKRF	157					
(Intercept)		-1.40	1.45		-0.97	
Infection (1)		0.13	0.18	1	0.71	0.478
Sex(M)		0.04	0.05	1	0.88	0.379
Site (Forest)		2.50	1.40	1	1.79	0.074
Brood Size		-0.01	0.03	1	-0.16	0.870
Hatch Date		0.03	0.03	1	0.87	0.382
Year(2017)		1.88	1.27	1	1.48	0.138
Infection*Site		-0.18	0.19	2	-0.97	0.332
Hatch Date*Year		-0.04	0.03	2	-1.36	0.173
Site*Hatch Date		-0.04	0.03	2	-1.38	0.168
NRF2	156					
(Intercept)		1.19	1.07		1.11	
Infection (1)		0.19	0.17	1	1.13	0.260
Sex(M)		0.10	0.03	1	2.94	0.003
Site (Forest)		0.74	1.05	1	0.71	0.478
Brood Size		0.01	0.03	1	0.01	0.997
Hatch Date		0.01	0.02	1	0.24	0.813
Year(2017)		0.93	0.96	1	0.97	0.334
Infection*Site		-0.22	0.17	2	-1.24	0.216
Hatch Date*Year		-0.02	0.02	2	-1.01	0.314
Site*Hatch Date		-0.01	0.02	2	-0.36	0.716

5.5 Discussion

5.5.1 *Leucocytozoon* prevalence

Malaria prevalence in this study was substantially higher at forest sites than at city sites, which is consistent with previous studies of avian malaria and urbanisation (Evans et al., 2009; Jiménez-Peñuela et al., 2019). Of the infected birds in this study, most infections were of *Leucocytozoon* species, four were of *Haemoproteus*, and one bird was co-infected with *Haemoproteus* and *Leucocytozoon*. No birds were found to be infected with *Plasmodium*.

In this study, parasites were detected from sampling circulating host blood. For *Leucocytozoon* parasites, the pre-patent stage, where parasites are present in the host liver but not yet in blood, is 5-7 days after introduction of parasites to the host (Eide and Fallis, 1972; Kocan and Clark, 1966). For *Haemoproteus* parasites, the pre-patent period is around 12-14 days (Fallis and Bennett, 1961) and for *Plasmodium* parasites it is 4-12 days (Ilgunas et al., 2019). Nestlings in this study may therefore have been harbouring *Haemoproteus* and *Plasmodium* infections, but levels of parasitaemia were not at a detectable level at day thirteen.

During this study, the overall prevalence of *Leucocytozoon* detected in the population differed between years; prevalence was much higher in 2016 than in 2017. *Leucocytozoon* prevalence also increased during the season, and this effect was strongest in 2016. A study conducted in the forest habitat used in this study observed seasonal variation in blackfly (genus *Simulium*) abundance and species diversity (Woodford et al., 2018). Therefore, it is likely that the increase in *Leucocytozoon* prevalence in great tit nestlings across the season was due to fluctuating abundance of blackfly vectors, or perhaps individuals hatching later were more susceptible to infection. A reduction in food abundance, or a deterioration of food quality, may also have contributed to the increase in infections across the season. In the city environment especially, parents may have stayed away longer to forage, or else have been in a worse state, leaving nestlings susceptible to infections.

In this study, *Leucocytozoon* prevalence was higher at forest sites than at city sites, and this difference in prevalence was conserved across years. Blackflies require lotic habitats for their larval development (Crosskey, 1990). It may be that city sites used in this study were less suitable habitats for black fly vectors, although abundance of vectors was not measured in this study. Indeed, prevalence of *Leucocytozoon* was very low in a study on

house sparrows at a city site (Dadam et al., 2019). Future studies of avian malaria and urbanisation should consider quantifying vector species abundance at study sites.

Previous studies report sex bias in infection, where males are more likely to be infected than females (Calero-Riestra and Garcia, 2016; Jenkins et al., 2015), although other authors have reported no sex bias in infection (Dunn et al., 2011). In this study, no sex bias in infection was found.

5.5.2 Leucocytozoon infection and nestling condition

This study found no evidence of effects of malaria infection on nestling condition, in contrast to previous studies that found a reduction in condition (Calero-Riestra and Garcia, 2016) and fledging success (Krams et al., 2013a). Male nestlings were heavier and had longer tarsi than females, which was expected, due to male great tits being structurally larger than females (Gosler, 1993). City birds also weighed significantly less than forest birds; however, there was no evidence that *Leucocytozoon* infection affected condition more in city birds than forest birds.

Contrary to expectation, haematocrit levels of nestlings were not significantly affected by presence of *Leucocytozoon* infection. Similarly, in a study on four songbird species, *Leucocytozoon* infection did not have an effect on body condition or haematocrit levels (Granthon and Williams, 2017). *Leucocytozoon* parasites infect host thrombocytes instead of red blood cells (Zhao et al., 2015), and therefore may not cause anaemias to the same degree as *Plasmodium* or *Haemoproteus* species. Moreover, many features in the natural environment as well as physiological condition can cause variation in haematocrit levels in wild birds (Fair et al., 2007).

Fledging success was high during both years at both sites in this study, and therefore acute *Leucocytozoon* infection in this system did not seem to affect fledging. The effects of *Leucocytozoon* infection on condition may be more evident later in life, and therefore future studies might consider recapturing nestlings after fledging to measure impacts of chronic infections on health.

5.5.3 *Leucocytozoon* infection and immune gene transcripts

In this study, it was expected that transcript levels of immune gene targets would be elevated in birds infected with *Leucocytozoon*. However, infection with *Leucocytozoon* did

not appear to affect the transcript levels of immune genes studied. In a previous study on siskins (*Carduelis spinus*), transcripts of many immune processes were elevated in response to *Plasmodium* infection (Videvall et al., 2015). Given that adaptive immune responses to infection vary even in closely related bird species (Lee et al., 2006), great tits may respond differently and varying in their susceptibility to infections. Furthermore, some birds may have been tolerant, or else resistant to infection (Palinauskas et al., 2008).

In addition, the nature of *Leucocytozoon* parasites to primarily infect host thrombocytes (Zhao et al., 2015), rather than erythrocytes (like *Plasmodium* and *Haemoproteus* parasites), may have influenced the detection of effects on the host immune system in this study.

5.5.4 Immune transcripts at city and forest sites

Following previous studies, it was predicted that transcript levels of oxidative stress associated immune targets would be higher for city nestlings than for their conspecifics in the forest (Isaksson et al., 2009; Watson et al., 2017). However, contrary to expectation, transcript levels of immune response elements *NRF2* and *NKRF* were significantly higher in forest birds than in city birds. Forest birds may have had other infections not quantified in this study, causing higher transcript levels of these two genes.

Higher levels of *TLR4* and *GATA3* were found in city birds than forest birds in a previous study (Capilla-Lasheras et al., 2017), however, these differences in transcript levels between sites were not reflected here. Ultimately, there was no evidence in this study to support that city birds have higher transcript levels of immune genes than forest conspecifics.

5.5.5 Conclusions and future studies

This study made use of a platform for measuring transcript levels of many genes from a single sample, the QuantiGene® Plex assay. Given that this assay had not been used before for passerine samples, transcript data in this study were noisy overall. Large variation in transcript data may have increased difficulties with detecting signals and responses to *Leucocytozoon* infection (See General Discussion for more details). Additionally, the qPCR method used to detect parasite infections from DNA was sensitive, and therefore nestlings that identified as malaria-positive may have had very low infection intensity. This

may be why there were small differences in immune gene transcript between infected and non-infected individuals. Further studies could quantify the intensity of infection with malaria parasites, rather than prevalence.

A further limitation of this study was that given that this study took place in the natural environment, the time that nestlings were bitten was unknown. This study included samples taken during the noon hours, but not at night, and therefore may have captured just one phase of the parasite life cycle. It is possible that at the time of sampling, erupted parasites during the acute blood stage of parasitaemia had already escaped into host cells, or else had been cleansed by the host spleen. Thus, the levels of parasitaemia in birds in this study may have been underestimated. Given the "circadian nature" of malaria parasites (Mideo et al., 2013), future studies ought to capture the full time profile of parasitaemia and host rhythmicity, instead of solely at the hours surrounding noon.

Chapter Six: Malaria parasite effects on the circadian physiology of wild birds

6.1 Introduction

Parasites are ubiquitous in natural environments and present many challenges for behaviour and physiology of their hosts (Hamilton and Zuk, 1982). During their life cycles, parasites must overcome barriers to transmission, survival and replication. Conversely, the host must actively combat infections, compensate for incurred damage, or tolerate them. Ecological interactions between host and parasite are complex, and selective pressures on fitness drive co-evolutionary relationships (Anderson and May, 1982). For many of these interactions, timing is crucial.

Many parasites show seasonal (circannual) or diel (circadian) traits during their development (Martinez-Bakker and Helm, 2015; Westwood et al., 2019). Endogenous timekeeping is advantageous as organisms can anticipate changes in the natural environment and synchronise internal biological rhythms accordingly. One of the most extensively studied host-parasite interactions in the context of rhythms are those of malaria parasites (*Plasmodium*, *Haemoproteus* and *Leucocytozoon* species) and their vertebrate hosts. Malaria infections typically give rise to symptoms of periodic fevers in their host, with a cycle length of 24 hours or multiples of 24 hours depending on the species of parasite (Kwiatkowski and Greenwood, 1989). This is a consequence of cycles of asexual replication within a host during the blood stages of parasite development. Individual merozoites invade host red blood cells, replicate and then divide to become schizonts (Reece et al., 2017). Subsequently, mature schizonts exhibit bursting behaviour, releasing merozoites into the host bloodstream (Figure 6.1.). Bursting behaviour in parasite species such as *Plasmodium chabaudi* occurs synchronously, and usually beginning at a specific time of the day (Mideo et al., 2013).

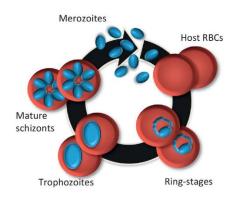


Figure 6.1. Conceptual diagram of development of malaria parasites within host blood cells. From (Mideo et al., 2013)

Rhythmic properties of malaria parasites have so far been documented in humans, mice and in birds (O'Donnell et al., 2013; Pigeault et al., 2015). Studies documenting the periodicity of malaria parasites have used blood smears collected at intervals over a period of 24 hours to estimate abundance of parasites by visual inspection (Gore et al., 1982; Roller and Desser, 1973), or determined abundance through molecular analyses (Hellgren et al., 2006).

6.1.1 Vector-host-parasite clock interactions

Given the rhythmic nature of both parasites and host, evolution has selected different rhythmic strategies for parasites. Selective advantages for malaria parasites in synchronising with rhythms of their hosts have driven evolutionary relationships of parasite clocks at each stage of their life cycle (Reece et al., 2017). For example, parasites exploit diel rhythms of insect vector biting behaviour for optimal transmission to their vertebrate hosts (Pigeault et al., 2018; Rund et al., 2016; Schneider et al., 2018).

Once parasites have successfully invaded a host, they must avoid elimination by host immune system defences (Reece et al., 2017). To avoid attack by the host immune system, it is thought that rhythmic bursting behaviour of mature schizonts is timed to occur when host defences are low (Mideo et al., 2013). In addition, the synchronous bursting of schizonts releases a high number of parasites at once, overwhelming the host immune system with sheer numbers (Reece et al., 2017). Mismatch of synchronicity with the host is costly for parasites; studies on experimental mismatch of *Plasmodium chabaudi* to mouse host results in reduced virulence and a two-fold cost for production and replication (O'Donnell et al., 2013, 2011).

Circadian rhythms in host immune system activity are well documented in humans (Labrecque and Cermakian, 2015; Scheiermann et al., 2013), and in birds (Markowska et al., 2017). Host rhythms can become disrupted in response to infections. For example, infection with *Plasmodium chabaudi* in mice disrupted rhythms in locomotor activity and body temperature, causing lethargy and hypothermia (Prior et al., 2019), although it was unclear whether this disruption was reflected in core circadian oscillators. In birds, malaria infections have been shown to cause significant shifts in the blood transcriptome (Videvall et al., 2015), however it is unknown whether malaria parasites affect circadian rhythms of their avian hosts.

6.1.2 Malaria parasite effects on avian hosts

Avian malaria parasites are globally distributed, occurring in most bird species (Bennett et al., 1994) and are highly prevalent in wild populations (Bensch et al., 2000; Lapointe et al., 2012). Pathologies of avian malaria infections vary from periodic fever, anaemia and mortality (Atkinson et al., 1995; Williams, 2005) to asymptomatic chronic infections (Bensch et al., 2007). Responses to infection with avian malaria vary depending on the parasite lineage (Ricklefs and Fallon, 2002), and bird species (Ilgunas et al., 2019).

Similar to infections with human and mouse malaria, avian malaria parasites show circadian periodicity during their development (Gore et al., 1982; Roller and Desser, 1973; Wolfson, 1936). Time of day effects occur in both acute and chronic infections, with periodic peaks of parasitaemia occurring in the late afternoon in some *Plasmodium* and *Leucocytozoon* species (Pigeault et al., 2018; Roller and Desser, 1973).

Despite the effects that malaria infections can have for fitness of wild birds (Atkinson et al., 1995; Krams et al., 2013b), host-parasite interactions in natural avian populations have not yet been studied in the context of rhythms. Given the complex evolutionary relationships that occur between malaria parasites and their hosts (Mideo et al., 2013; Prior et al., 2019; Reece et al., 2017), it is important to consider timing effects of parasites in ecological studies of avian malaria infections. This study aimed to address this gap in knowledge by investigating whether acute infection with avian malaria influences host circadian physiology, and in particular, the rhythmic expression of biological clock and immune system components in wild birds.

6.2 Study aims and hypotheses

The aim of this study was to test for differences in transcript levels of circadian clock and immune genes in wild passerine nestlings naturally infected with *Leucocytozoon* malaria parasites. Specifically, this study investigated three ways that parasites might affect rhythms of their avian hosts, by asking the following:

- 1) Does acute infection with *Leucocytozoon* result in any changes in host rhythmicity of genes involved in the circadian clock?
- 2) Does acute infection result in changes in the timing of peak levels of circadian clock gene transcripts, indicating that rhythmic effects of *Leucocytozoon* are clock-mediated?
- 3) Are transcripts of immune genes in infected birds upregulated or downregulated by presence of *Leucocytozoon* parasites during acute infection?

In a recent study on *P. chabaudi* infected mice, night peaks of locomotory activity were reduced and body temperature rhythms disrupted in malaria-infected individuals (Prior et al., 2019). It is unknown whether this disruption in activity and body temperature rhythms was reflected in core circadian clock oscillators of mice. However, behaviour is linked to rhythms in the SCN (van Oosterhout et al., 2012). Therefore, it was expected that in this study, rhythmicity of transcript levels of genes involved in the circadian clock would be disrupted in infected birds, compared to non-infected birds.

In (Prior et al., 2019), infected mice were active earlier at night, indicating a phase shift in host rhythms by parasites. Great tits are diurnal (Gosler, 1993), and therefore in contrast to nocturnally active mice, it was expected that infected birds may experience a phase shift, to be earlier in the morning. This shift may be seen as a change in peak timing of transcript levels of circadian clock transcripts.

Acute avian malaria infection has previously been shown to alter the transcriptome of birds by elevating immune transcripts (Videvall et al., 2015). Reductions of immune transcripts in infected birds may occur if birds are immunocompromised, or otherwise in a poor state (Kwon et al., 2008; Meitern et al., 2014). A previous experiment carried out at the field site used in this study showed no association with immune gene transcripts levels and avian malaria infections (Capilla-Lasheras et al., 2017), however, this study had low sample sizes and did not account for host rhythmicity.

In this study, it was expected that there may be changes in transcript levels by either elevation or suppression of immune system components in infected birds compared to non-infected birds. As the immune system of birds is closely tied to the circadian clock (Markowska et al., 2017), changes in the timing of expression of immune transcripts, such as differences in acrophase or amplitude, were also expected.

6.3 Materials and Methods

6.3.1 Study approaches

Rhythmic fluctuations of circadian clock and immune gene transcripts can be determined from blood (Archer et al., 2014). In this study, wild nestlings were blood-sampled throughout the day and night, to capture diel expression of clock and immune gene transcripts.

For analyses of malaria infection on clock gene transcript levels, the following components of the molecular clock were included in this study: positive clock elements and transcriptional activators BMALI (brain and muscle ARNT-like 1) and CLOCK (circadian locomotor output cycles kaput); negative clock elements and inhibitors CRYI (cryptochrome circadian regulator 1) and PER2 (period circadian regulator 2); and accessory loop protein $REVERB\alpha$ (Kumar and Sharma, 2018; Yoshimura et al., 2000). Additional gene targets included $CKI\mathcal{E}$ (casein kinase 1), which catalyses phosphorylation of E4BP4, a light-inducible gene in the nestling pineal gland (Doi et al., 2004) and AANAT (aralkylamine N-acetyltransferase), an enzyme regulating melatonin synthesis (Bernard et al., 2002).

In addition to core avian clock genes, this study aimed to investigate circadian effects of *Leucocytozoon* infection on transcript levels of immune genes across a time profile.

Immune gene targets included *LY86* (lymphocyte antigen 86) and *TLR4* (toll-like receptor 4) which are involved in anti-bacterial and anti-malarial responses (Medzhitov, 2001), and the type 2 transcription factor *GATA3* (GATA binding protein 3) involved in adaptive immunity (Wang et al., 2011). Additional immune targets included: *IL1* and *IL6* (interleukins 1 and 6) which are inflammatory response mediators (Klasing, 1998); *PRKCA* (protein kinase C alpha), an enzyme responsible for antiviral effects and cell growth regulation (Clemens and Elia, 1997); and *IKBA* (NF-kappa-B inhibitor alpha) and *NKRF* (NF-kappa-B repressing factor), which are two inhibitors of the immune response regulator NF-kappa-B ((Cabannes et al., 1999; Nourbakhsh and Hauser, 1999). Finally, immune targets *NRF2* (nuclear factor erythroid 2) and *IGF1* (insulin like growth factor 1) which are both involved in resistance to oxidative stress (Kensler et al., 2007; Holzenberger, 2003 respectively) were included in this study.

Gene transcript data in this study were first fitted to a cosine curve to determine if transcript levels were circadian across the sampling time profile. If transcript levels of

genes were circadian, parameters of the cosine curve, amplitude and acrophase (Figure 6.2.) were determined for both malaria infected and non-infected birds. The amplitude and acrophase of circadian transcripts were then compared for infected and non-infected birds, to determine if *Leucocytozoon* parasites influence rhythmic expression of clock and immune genes in their host.

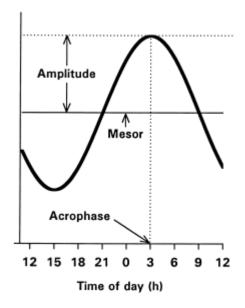


Figure 6.2. Schematic representation of acrophase, mesor and amplitude of a circadian rhythm. Amplitude is one half of the maximal fluctuation between trough and peak. Acrophase is the location of the peak with respect to reference time. Mesor is the midline, or rhythm-adjusted average. Figure obtained from Hayano et al., (1998).

6.3.2 Study sites

Data for this chapter were obtained between April and June of 2016 and 2017 at forest sites on free-living populations of breeding great tits. Field sites were existing forest nest box study systems located at the Scottish Centre for Ecology and the Natural Environment (SCENE; 56°, 7.73'N, 4° 36.79'W) and Cashel Forest (56° 6'N, 4°34'W). These sites were previously used in ecological studies of wild passerines and avian malaria (Capilla-Lasheras et al., 2017; Pollock et al., 2017; Woodford et al., 2018). To determine hatch dates, all boxes were initially checked weekly and, closer to hatching, every second day as part of a general nest box monitoring protocol (see General Methods for further details of nest box monitoring). Nestlings used in ALAN experiment in Chapter Four were excluded from this study.

6.3.3 Nestling processing

All nestlings (N=199) in this study were ringed for individual identification on day thirteen of life. Sampling was carried out at timepoints distributed throughout the day and night on day thirteen (Figure 6.3.). Sampling was undertaken where possible close to the six hour intervals of the day commonly used in circadian biology (06:00, 12:00, 18:00 and 00;00), with the aim of creating a profile of rhythmic expression of transcript levels across a 24 hr day. Due to constraints of sampling nestlings in the field from multiple nest boxes, samples were not taken "on the hour". Therefore, for ease of statistical analysis, sample time was rounded up or down to the nearest hour (e.g. sample taken at 12:30 was rounded to 13:00).

To reduce burden on each nestling, individuals were sampled once, which therefore enabled spread of sampling of a given brood over different times of day. During sampling, two blood samples of ~50 µl were obtained from each nestling via the brachial wing vein. The first was taken by a heparinised capillary and stored in 0.5 ml 100% ETOH for quantification of malaria parasites. A second blood sample was obtained and stored in 250 µl RNAlater® stabilising solution (Invitrogen) for analyses of gene transcript levels. See General Methods section for further details on nestling blood sampling protocol.

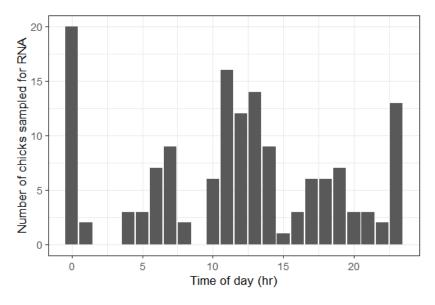


Figure 6.3. Sample time distribution of RNA samples taken for this study. Samples were taken across the breeding seasons in two years (2016 and 2017).

6.3.4 Ethical Statement

See General Methods section for details of ringing licences and blood sampling.

6.3.5 Laboratory procedures

Determining transcript levels of clock and immune genes

Samples in 100% ETOH were frozen and stored at -40 °C. Genomic DNA was later extracted from these samples via DNAeasy Blood and Tissue Kit (Qiagen). Measurement of Median Fluorescence Intensity (as a proxy for transcript levels) for immune and clock gene targets outlined in Table 6.1. was carried out on all samples stored in *RNAlater*® using the QuantiGene® Plex RNA assay 2.0 (ThermoFisher). See General Methods section for further details on the QuantiGene® Plex RNA assay procedure.

Table 6.1. Housekeeping, clock and immune gene targets used within this study.

Transcript targets	Type		
RPL19, SDHA, HMBS	Housekeeping genes		
BMAL1, CLOCK, CRY1 ,CK1E, AANAT, PER2, REVERBA	Circadian clock genes and melatonin synthesis		
LY86, TLR4, IL1, IL6, PRKCA, GATA3, IKBA, NKRF, NRF2, IGF1	Immune genes		

Measuring infection intensity of Leucocytozoon parasites

All avian DNA samples were run using a sensitive qPCR protocol for quantifying malaria infection in young birds developed by another PhD student (Albalawi, 2019). See General Methods section for full details of malaria qPCR protocol.

6.3.6 Statistical Analysis and Model Selection

All statistical analyses were performed using R v. 3.4.4 (R Core Team, 2018), and packages *lme4* (Bates et al., 2015). Data were visualised using ggplot2 (Wickham, 2014). Initial global models for response variables were generalised linear-mixed models (GLM), including individual nest box as a random factor to account for repeated measures within the same nest box. Residuals were plotted and inspected to check assumptions of normality whilst modelling using GLMs. Pairs of explanatory variables were also assessed for collinearity by calculating tolerance and variance inflation factors (VIF) and also by visual inspection of pairwise plots, before statistical relevance was accepted.

Based on the number of *Leucocytozoon* parasite gene copies (as a proxy for number of individual parasites) detected in the sample, infections were quantified into categories.

From visual inspection of the distribution of the data, two main infection groups were identified. These were: low levels of infection (≤1% of blood cells infected) and high levels of infection (>1% of blood cells infected). Additionally, a category was included where no parasites/infection was detected in the sample. Each sample was then assigned to the mean of the infection group and entered into models as a numeric explanatory variable (where no infection=1, low infection=2 and high infection=3), to preserve ordinal categories.

Median fluorescence intensity data (as a proxy for transcript levels of genes) were obtained from the QuantiGene® Plex RNA assay 2.0 (Thermofisher). These data were normalised by division of the mean fluorescence of background wells and then dividing each sample by the geometric mean of levels of housekeeping genes in Table 6.1., thus, transcript levels are expressed as relative levels to housekeeping genes. Transcript data were modelled using a Gamma distribution with log-link function. Time of day information was converted to decimal time, then to radians, and then entered into the model as an explanatory variable in its components of sine(time) and cosine(time). Global models included interactions between infection and sin(time), and also infection and cos(time). Additional explanatory variables included hatch date in April days (where April 1st = 1) and study year (2016/2017). Nest box was included in all models as a random factor.

Statistical relevance of fixed factors within all models was determined by sequential model selection via Likelihood Ratio Testing using package *lmtest* (Zeileis & Hothorn, 2002), until the minimum adequate model (MAM) was found. Factors with a significance level of <5% were kept in the model. Infection and sine and cosine were kept in the model regardless of significance.

Cosinor analyses were carried out using model outputs of genes found to be rhythmic in initial models (where sine or cosine were significant). The amplitude of rhythms of transcript levels was calculated as the square root of the sum of sine and cosine coefficients: sqrt(sine² + cosine². The acrophase was determined as arctan(- sine / cosine).

6.4 Results

In total, infection with *Leucocytozoon* parasites was quantified in 158 nestlings. Out of those sampled, 61 nestlings had no detectable infection, 68 had low-level infection (\leq 1%) and 29 had high levels of infection (>1%).

6.4.1 Clock gene transcripts

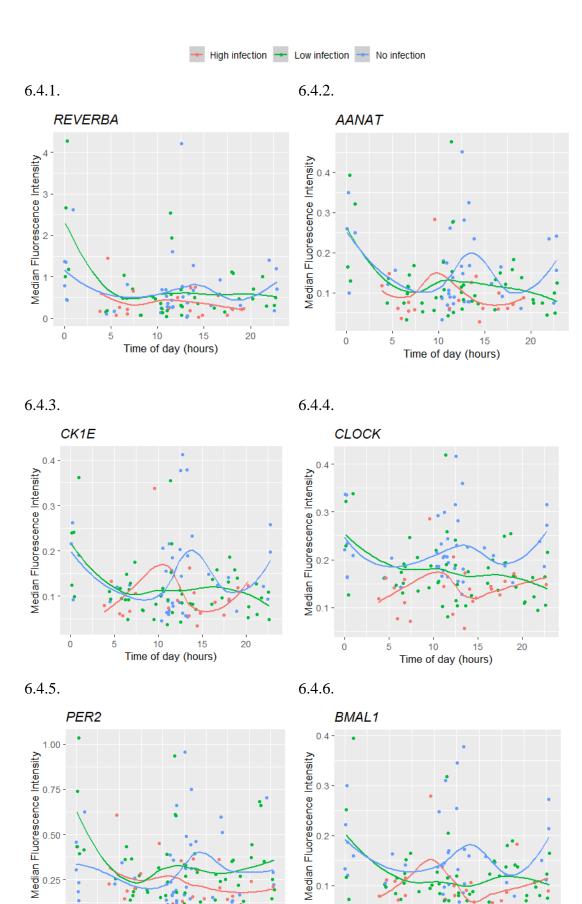
For all clock genes analysed in this study, there was no significant interaction between sine and cosine and *Leucocytozoon* infection (Table 6.2.). Moreover, transcript levels of clock genes did not fit to either sine or cosine curves, and therefore there was no support for rhythmicity (Figure 6.4). Transcript levels of clock genes were also not significantly affected by either hatch date or study year (Table 6.2.).

High levels of infection with Leucocytozoon significantly reduced the levels of transcripts of three clock genes: AANAT ($t_{1,116}$ =-2.62, P=0.009; LMM), CLOCK ($t_{1,131}$ =-4.04, P<0.001; LMM), and CRYI ($t_{1,138}$ =-2.06, P=0.039; LMM), (Table 6.2). For AANAT, transcript levels in infected birds were 24 % lower compared with non-infected birds (Figure 6.4.2.). For CLOCK, transcript levels of infected birds were 16 % lower than non-infected birds (Figure 6.6.4.). Finally, transcript levels for CRYI were 9 % lower in infected birds (Figure 6.4.8.). In addition, reductive effects of high levels of Leucocytozoon infection on transcript levels tended towards significance for PER2 ($t_{1,139}$ =-1.94, P=0.068; LMM), (Figure 6.4.5).

Table 6.2. Model summaries for circadian clock gene targets. Estimates shown here were log-transformed. Global models for all targets included: Nest box + infection * sine(time) + infection * cosine(time) + hatch date + study year. Nest box was included in all models as a random factor. Sine and cosine were tested together during likelihood ratio testing.

Variable	N	Estimate	s.e.m	d.f.	<i>t</i> -value	<i>p</i> -value LRT
REVERBA	112					
(Intercept)		2.10	1.34		1.56	
Infection(Low)		0.01	0.19	1	0.07	0.945
Infection(High)		-0.26	0.23	1	-1.12	0.262
sin(Time)		-0.11	0.28	1	-0.38	0.704
cos(Time)		0.22	0.16	1	1.35	0.177
Hatch date		-0.06	0.03	1	-2.10	0.055
Infection(Low)*sin(Time)		0.06	0.33	2	0.17	0.867
Infection(High)*sin(Time)		0.49	0.37	2	1.33	0.183
Infection(Low)*cos(Time)		-0.04	0.22	2	-0.18	0.856
Infection(High)*cos(Time)		0.01	0.33	2	0.04	0.967
AANAT	116					
(Intercept)		-1.30	0.92		-1.42	
Infection(Low)		-0.18	0.12	1	-1.46	0.145
Infection(High)		-0.43	0.16	1	-2.62	0.009
sin(Time)		-0.05	0.19	1	-0.26	0.792
cos(Time)		0.07	0.11	1	0.66	0.510
Hatch date		-0.01	0.02	1	-0.79	0.431
Infection(Low)*sin(Time)		0.08	0.22	2	0.37	0.709
Infection(High)*sin(Time)		0.18	0.25	2	0.72	0.472
Infection(Low)*cos(Time)		0.02	0.15	2	0.11	0.913
Infection(High)*cos(Time)		-0.07	0.23	2	-0.31	0.760
CK1E	126					
(Intercept)		-1.34	0.77		-1.74	
Infection(Low)		-0.04	0.12	1	-0.35	0.724
Infection(High)		-0.20	0.14	1	-1.41	0.160
sin(Time)		-0.15	0.17	1	-0.90	0.367
cos(Time)		-0.04	0.10	1	-0.39	0.693
Hatch date		-0.02	0.01	1	-1.03	0.303
Infection(Low)*sin(Time)		0.19	0.19	2	0.96	0.338
Infection(High)*sin(Time)		0.25	0.22	2	1.14	0.253
Infection(Low)*cos(Time)		0.18	0.14	2	1.27	0.203
Infection(High)*cos(Time)		0.01	0.19	2	0.05	0.958
CLOCK	131					
(Intercept)		-0.75	0.46		-1.63	0.104
Infection(Low)		-0.12	0.06	1	-1.93	0.054
Infection(High)		-0.34	0.08	1	-4.04	< 0.001
sin(Time)		-0.07	0.09	1	-0.73	0.466
cos(Time)		0.00	0.05	1	0.02	0.987
Hatch date		-0.02	0.01	1	-1.83	0.067
Infection(Low)*sin(Time)		0.19	0.11	2	1.67	0.094
Infection(High)*sin(Time)		0.04	0.12	2	0.35	0.725
Infection(Low)*cos(Time)		0.07	0.07	2	0.98	0.328
Infection(High)*cos(Time)		0.04	0.10	2	0.35	0.728
PER2	139					
(Intercept)		-2.16	1.11		-1.95	
Infection(Low)		-0.04	0.09	1	-0.41	0.683
Infection(High)		-0.23	0.13	1	-1.83	0.068
sin(Time)		0.02	0.16	1	0.10	0.920
cos(Time)		-0.02	0.07	1	-0.30	0.765
Hatch date		0.02	0.02	1	0.75	0.454
Infection(Low)*sin(Time)		-0.03	0.18	2	-0.16	0.874
Infection(High)*sin(Time)		0.15	0.20	2	0.74	0.460
Infection(Low)*cos(Time)		0.13	0.11	2	1.14	0.253

Infection(High)*cos(Time)		0.15	0.16	2	0.96	0.338
BMAL1	125					
(Intercept)		-1.43	0.77		-1.85	
Infection(Low)		-0.15	0.10	1	-1.50	0.135
Infection(High)		-0.20	0.13	1	-1.45	0.146
sin(Time)		-0.05	0.15	1	-0.30	0.761
cos(Time)		-0.03	0.07	1	-0.48	0.633
Hatch date		-0.01	0.01	1	-0.80	0.424
Infection(Low)*sin(Time)		0.08	0.17	2	0.45	0.652
Infection(High)*sin(Time)		0.08	0.20	2	0.39	0.697
Infection(Low)*cos(Time)		0.15	0.11	2	1.32	0.187
Infection(High)*cos(Time)		0.25	0.17	2	1.50	0.134
CKA	142					
(Intercept)		0.82	0.41		2.00	
Infection(Low)		-0.02	0.04	1	-0.52	0.603
Infection(High)		0.00	0.05	1	0.02	0.986
sin(Time)		0.02	0.06	1	0.28	0.779
cos(Time)		0.00	0.03	1	-0.11	0.910
Hatch date		-0.01	0.01	1	-1.56	0.118
Infection(Low)*sin(Time)		-0.04	0.07	2	-0.55	0.579
Infection(High)*sin(Time)		-0.05	0.08	2	-0.65	0.514
Infection(Low)*cos(Time)		0.01	0.05	2	0.27	0.785
Infection(High)*cos(Time)		0.10	0.07	2	1.46	0.143
CRY1	138					
(Intercept)		2.00	0.045		-1.40	
Infection(Low)		-0.52	0.603	1	-0.96	0.340
Infection(High)		0.02	0.986	1	-2.06	0.039
sin(Time)		0.28	0.779	1	-0.76	0.449
cos(Time)		-0.11	0.910	1	-0.11	0.910
Hatch date		-1.56	0.118	1	-0.34	0.733
Infection(Low)*sin(Time)		-0.55	0.579	2	0.25	0.801
Infection(High)*sin(Time)		-0.65	0.514	2	1.45	0.148
Infection(Low)*cos(Time)		0.27	0.785	2	1.32	0.186
Infection(High)*cos(Time)		1.46	0.143	2	0.79	0.428



10 15 Time of day (hours)

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Time of day (hours)

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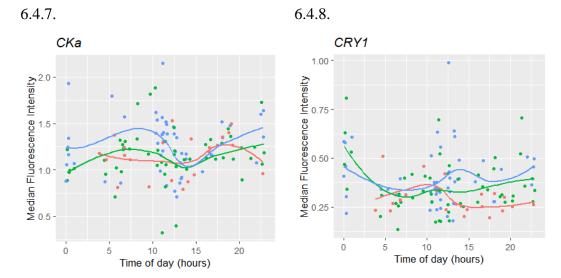


Figure 6.4. Transcript levels of circadian clock gene targets for infection with *Leucocytozoon* parasites, at different times of day: 1) *REVERBA* 2) *AANAT* 3) *CK1E* 4) *CLOCK* 5) *PER2* 6) *BMAL1* 7) *CKA* 8) *CRY1*. Colours indicate intensity of infection with Leucocytozoon parasites: No infection (blue), low-level infection (green) and high-level infection (red). Data shown are empirical values obtained from QuantiGene® Plex RNA assay and are shown in Median florescence intensity units (MFI). Data were fitted with Loess curves.

6.4.2 Immune gene transcripts

For all immune genes studied, there was no significant interaction between sine and cosine and *Leucocytozoon* infection (Table 6.3). Separate to infection, immune genes studied did not fit to either sine or cosine curves, and therefore there was no support for rhythmicity in any of these genes (Figure 6.5). Hatch date or study year did not affect transcript levels of any of the immune genes studied (Table 6.3).

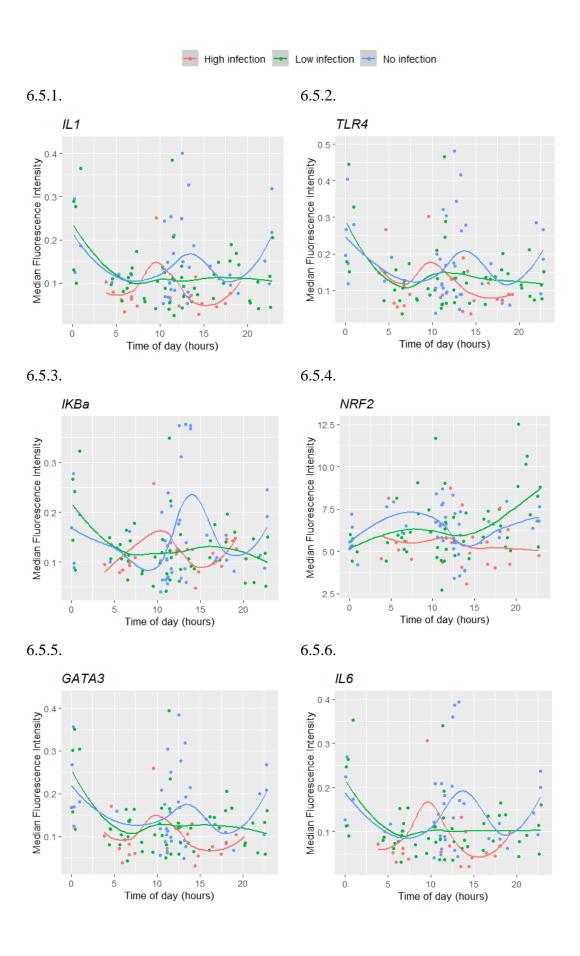
High-level infection with *Leucocytozoon* significantly reduced transcript levels in most of the immune genes studied: *IL1*, *TLR4*, *NRF2*, *GATA3*, *IL6*, *LY86*, *NKRF*, *PRKCA* and *IGF1* (Table 6.3, Figure 6.5.). Fold changes in transcript levels of these genes between infected birds (with high level of infection) and non-infected birds ranged from 0.88 to 0.05.

Table 6.3. Model summaries for immune gene transcripts. Estimates shown here are log-transformed. Global models included: Nest box + infection * sine(time) + infection * cosine(time) + hatch date + study year. Nest box was included in all models as a random factor. Sine and cosine were tested together during likelihood ratio testing.

Variable	N	Estimate	s.e.m	d.f.	<i>t</i> -value	<i>p-</i> value LRT
IL1	120					LKI
(Intercept)	120	-0.59	0.87		-0.68	0.494
Infection(Low)		-0.11	0.11	1	-0.97	0.331
Infection(High)		-0.53	0.16	1	-3.39	0.001
sin(Time)		-0.01	0.17	1	-0.07	0.943
cos(Time)		0.07	0.09	1	0.81	0.417
Hatch date		-0.03	0.02	1	-1.71	0.088
Infection(Low)*sin(Time)		0.08	0.20	2	0.41	0.684
Infection(High)*sin(Time)		0.17	0.23	2	0.75	0.451
Infection(Low)*cos(Time)		0.11	0.13	2	0.82	0.411
Infection(High)*cos(Time)		-0.09	0.21	2	-0.41	0.682
TLR4	115					
(Intercept)		-1.65	0.91		-1.82	0.068
Infection(Low)		-0.24	0.12	1	-2.00	0.045
Infection(High)		-0.41	0.16	1	-2.56	0.011
sin(Time)		-0.09	0.20	1	-0.44	0.657
cos(Time)		0.06	0.08	1	0.71	0.475
Hatch date		0.00	0.02	1	-0.19	0.849
Infection(Low)*sin(Time)		0.11	0.22	2	0.47	0.636
Infection(High)*sin(Time)		0.27	0.25	2	1.09	0.277
Infection(Low)*cos(Time)		0.03	0.13	2	0.21	0.833
Infection(High)*cos(Time)		-0.02	0.21	2	-0.10	0.923
IKBA	132					
(Intercept)		-1.45	0.77		-1.89	0.059
Infection(Low)		-0.04	0.09	1	-0.45	0.650
Infection(High)		-0.11	0.12	1	-0.87	0.383
sin(Time)		-0.13	0.14	1	-0.96	0.339

cos(Time)		-0.03	0.07	1	-0.44	0.661
Hatch date		-0.01	0.01	1	-0.80	0.425
Infection(Low)*sin(Time)		0.16	0.16	2	0.96	0.336
Infection(High)*sin(Time)		0.17	0.18	2	0.98	0.328
Infection(Low)*cos(Time)		0.16	0.11	2	1.49	0.137
Infection(High)*cos(Time)		0.03	0.16	2	0.17	0.864
NRF2	141					
(Intercept)		1.82	0.46		3.92	
Infection(Low)		0.03	0.04	1	0.66	0.508
Infection(High)		-0.09	0.05	1	-1.58	0.114
sin(Time)		0.03	0.06	1	0.43	0.668
cos(Time)		0.00	0.03	1	0.17	0.866
Hatch date		0.00	0.01	1	-0.01	0.989
Infection(Low)*sin(Time)		-0.10	0.07	2	-1.32	0.186
Infection(High)*sin(Time)		-0.03 0.02	0.08 0.05	2 2	-0.36 0.47	0.719 0.636
Infection(Low)*cos(Time) Infection(High)*cos(Time)		0.02	0.03	2	1.58	0.030
GATA3	126	0.11	0.07	2	1.38	0.113
(Intercept)	120	-1.04	0.76		-1.38	
Infection(Low)		-0.14	0.70	1	-1.37	0.171
Infection(High)		-0.42	0.10	1	-3.02	0.171
sin(Time)		0.02	0.14	1	0.10	0.920
cos(Time)		0.08	0.10	1	1.01	0.310
Hatch date		-0.02	0.01	1	-1.19	0.232
Infection(Low)*sin(Time)		0.01	0.19	2	0.04	0.968
Infection(High)*sin(Time)		0.13	0.21	2	0.61	0.541
Infection(Low)*cos(Time)		0.04	0.12	2	0.34	0.734
Infection(High)*cos(Time)		-0.08	0.19	2	-0.42	0.678
IL6	115	0.00	0.17	_	···-	0.070
(Intercept)		-1.16	0.87		-1.34	
Infection(Low)		-0.19	0.12	1	-1.57	0.116
Infection(High)		-0.59	0.17	1	-3.43	0.001
sin(Time)		-0.07	0.20	1	-0.35	0.725
cos(Time)		0.00	0.09	1	0.02	0.986
Hatch date		-0.02	0.02	1	-1.01	0.311
Infection(Low)*sin(Time)		0.10	0.23	2	0.41	0.681
Infection(High)*sin(Time)		0.23	0.26	2	0.86	0.392
Infection(Low)*cos(Time)		0.20	0.14	2	1.41	0.159
Infection(High)*cos(Time)		-0.06	0.22	2	-0.27	0.788
LY86	139					
(Intercept)		-1.16	0.70		-1.65	
Infection(Low)		-0.09	0.07	1	-1.20	0.231
Infection(High)		-0.22	0.10	1	-2.25	0.025
sin(Time)		-0.07	0.11	1	-0.59	0.554
cos(Time)		-0.01	0.05	1	-0.16	0.875
Hatch date		-0.01	0.01	1	-0.56	0.579
Infection(Low)*sin(Time)		0.05	0.13	2	0.39	0.693
Infection(High)*sin(Time)		0.04	0.15	2	0.27	0.787
Infection(Low)*cos(Time)		0.13	0.08	2	1.54	0.125
Infection(High)*cos(Time)		0.01	0.12	2	0.12	0.903
NKRF	141					
(Intercept)	141	1.61	0.62		2.58	
Infection(Low)		0.02	0.02	1	0.34	0.735
Infection(Low) Infection(High)		-0.36	0.03	1	-5.43	0.733 0.001
sin(Time)		-0.30	0.07	1	-0.92	0.360
cos(Time)		0.00	0.07	1	-0.92	0.300
Hatch date		-0.02	0.04	1	-0.02	0.980
Infection(Low)*sin(Time)		-0.02	0.01	2	-0.51	0.607
Infection(High)*sin(Time)		0.03	0.10	2	0.31	0.758
Infection(Low)*cos(Time)		-0.01	0.16	2	-0.16	0.870
Infection(High)*cos(Time)		0.10	0.08	2	1.22	0.222
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Variable	N	Estimate	s.e.m	d.f.	t-value	<i>p</i> -value LRT
PRKCA	113					
(Intercept)		-1.50	0.87		-1.72	
Infection(Low)		-0.16	0.13	1	-1.26	0.207
Infection(High)		-0.47	0.16	1	-2.85	0.004
sin(Time)		-0.12	0.19	1	-0.61	0.540
cos(Time)		0.12	0.09	1	1.31	0.190
Hatch date		-0.01	0.02	1	-0.75	0.454
Infection(Low)*sin(Time)		0.12	0.22	2	0.56	0.578
Infection(High)*sin(Time)		0.27	0.25	2	1.08	0.282
Infection(Low)*cos(Time)		0.02	0.15	2	0.16	0.875
Infection(High)*cos(Time)		-0.11	0.22	2	-0.53	0.595
IGF1	122					
(Intercept)		-1.26	0.83		-1.52	
Infection(Low)		-0.17	0.12	1	-1.44	0.151
Infection(High)		-0.54	0.16	1	-3.28	0.001
sin(Time)		-0.07	0.19	1	-0.38	0.707
cos(Time)		0.04	0.09	1	0.44	0.661
Hatch date		-0.02	0.02	1	-1.00	0.319
Infection(Low)*sin(Time)		0.07	0.22	2	0.31	0.755
Infection(High)*sin(Time)		0.20	0.25	2	0.78	0.434
Infection(Low)*cos(Time)		0.17	0.14	2	1.16	0.244
Infection(High)*cos(Time)		0.02	0.21	2	0.10	0.923



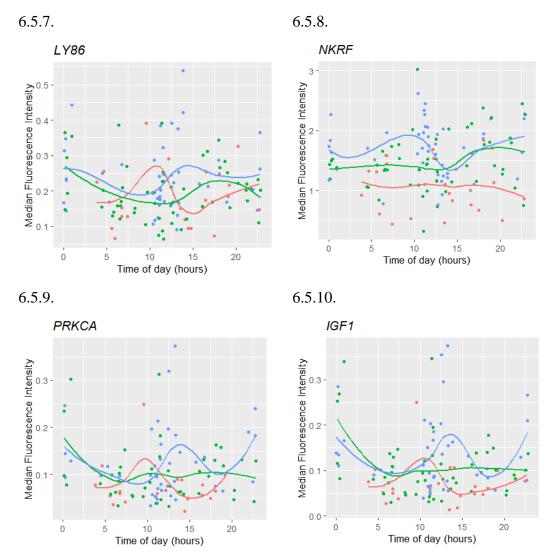


Figure 6.5. Transcript levels of immune gene targets for targets for infection with *Leucocytozoon* parasites, at different times of day. 1) *IL1* 2) *TLR4* 3) *IKBA* 4) *NRF2* 5) *GATA3* 6) *IL6* 7) *LY86* 8) *NKRF* 9) *PRKCA* 10) *IGF1*. Colours indicate intensity of infection with Leucocytozoon parasites: No infection (blue), low-level infection (green) and high-level infection (red). Data shown are empirical values obtained from QuantiGene® Plex RNA assay and are shown in Median florescence intensity units (MFI). Data were fitted with Loess curves.

6.5 Discussion

6.5.1 Rhythmicity in gene transcript levels

This study aimed to detect rhythmic transcripts in nestling great tits for clock and immune gene targets, for the first time in a wild animal. However, there was no support for rhythmicity for the circadian clock and immune genes studied.

Rhythmicity may not have been observed for transcript levels of gene targets in this study for a number of reasons. Firstly, nest disturbances during blood sampling may have influenced circadian gene expression, by way of sleep interruptions (Möller-Levet et al., 2013; Archer et al., 2014), or altered rhythms as a response to stress (Weibel et al., 2002). Furthermore, the nature of outdoor sampling introduces elements such as variations in ambient temperature. Although circadian mechanisms are largely temperature compensated (Merrow et al., 2005; Pittendrigh 1954), changing thermal conditions have nevertheless previously been shown to directly influence timing of activity patterns (Lehmann et al., 2012; Vivanco et al., 2010). Temperature variations therefore may have also played a role in increasing individual variation in circadian gene expression in this study.

In the two previous Chapters, this study made use of a platform for measuring transcript levels, the QuantiGene® Plex assay. Although previously used to measure transcripts in chickens (Gunawardana et al., 2019), this assay had not been used before for passerine samples. Therefore, the assay was not perfectly optimised for use on passerine blood samples, which may have contributed to the high levels of variation in transcript data, and why rhythmicity was detected in only two genes.

Finally, the use of blood samples instead of tissue may have reduced capacity to capture rhythmic expression of genes in this study. Rhythmic expression of genes can be detected from blood samples in humans (Archer et al., 2014). In birds, day-night differences in levels of immune genes have been shown in blood leucocytes of chickens (Turkowska et al., 2013), and previous studies has observed immune gene profiles from passerine blood (Meitern et al., 2014; Watson et al., 2017). However, there can be marked differences in the timing and level of expression of circadian gene transcripts between tissue types (Karaganis et al., 2009). The possibility remains the combined effect of noisy data from the QuantiGene® Plex assay, and low signal from the use of blood samples (in place of other tissues) masked the detection of rhythmicity in most genes in this study.

6.5.2 Effects of infection with *Leucocytozoon* parasites on circadian clock and immune transcript levels

This study investigated the effects of *Leucocytozoon* infection on the rhythmic expression of clock and immune transcript levels in wild nestlings. Transcript levels for many immune genes appeared to be significantly suppressed by *Leucocytozoon* infection. As discussed in Chapter Five, *Leucocytozoon* parasites primarily infect host thrombocytes (Zhao et al., 2015), and therefore may either directly infect cells that are expressing immune transcripts, or cause indirect effects as a consequence of activating T-cells responding to infected thrombocytes. In this study, *Leucocytozoon* parasites may have directly infected and destroyed host thrombocytes, resulting in low levels of immune gene transcripts.

Curiously, many circadian clock gene transcripts also appeared to be suppressed by high infection. Circadian transcripts are produced by thrombocytes (Ando et al., 2009; Turkowska et al., 2013), as well as erythrocytes (Archer et al., 2014). Therefore, circadian transcripts may have been suppressed by *Leucocytozoon* in a similar way to immune gene transcripts, by destruction of host thrombocytes.

6.5.3 Conclusions and future studies

This study provided the first attempt to quantify rhythmic transcripts in birds within the natural environment. However, in this study, no support for rhythmicity was found for the circadian clock and immune gene targets studied.

This study showed infection with *Leucocytozoon* parasites to significantly suppress transcript levels of genes involved in the circadian clock and immune system of wild nestlings. Suppression of the immune system may leave birds susceptible to further infections (Gonzalez et al., 1999; Nordling et al., 1998), and bring negative consequences for life histories in terms of reduced body condition (Marzal et al., 2008b), and reproductive fitness (Hamilton and Zuk, 1982).

The assay used in this study to quantify malaria infections was able to distinguish between infections with *Leucocytozoon*, or other parasite genera such as *Haemoproteus*, but it could not identify parasites at species level (Albalawi, 2019). Given that different lineages and species of malaria parasites have different rhythmic properties (Prior et al., 2019; Rivero and Gandon, 2018), future studies might take the parasite species into consideration when investigating impacts on host rhythms.

In addition, the time and day that nestlings were bitten could not be determined in this study. As such, *Leucocytozoon* parasites may have been in different developmental stages in host blood (e.g. liver stage), and therefore, the diel profile of *Leucocytozoon* parasite bursting activity could not be identified. Future studies should determine the diel stages of *Leucocytozoon* life cycle, by using blood smears to visually identify different developmental stages of parasites. This would help to further elucidate host-parasite interactions between *Leucocytozoon* parasites and avian circadian physiology.

An additional statistical analysis, running models for gene transcripts that were expected to shift in the same direction together (e.g. positive clock elements in the same pathway *CLOCK* and *BMAL1*), may provide further insight into effects of *Leucocytozoon* infection on gene pathways. In these models, gene transcripts should be treated as fixed effects, and cosine and sine curves fitted across all genes.

Chapter Seven: General Discussion

Despite the growing evidence that robust circadian rhythms are important for physiology, studies of biological rhythms of animals in their natural environments remain rare. This thesis aimed to put studies of rhythms into an ecological context. Observational and experimental approaches were used to investigate daily rhythms in a wild bird, the great tit, at three levels: behaviour, gene transcripts, and life histories.

In this thesis, individual differences in behavioural timing were quantified using incubation behaviour. These differences in timing were then linked to fitness traits and put into the context of living in the evolutionarily novel city environment. Furthermore, impacts of a key feature of the city environment, artificial light at night, were investigated in the context of biological rhythms and condition of nestling great tits. Effects of infection with a widespread parasitic disease, avian malaria, on the immune system and condition of nestlings in city and forest habitats were then investigated. Finally, interactions of host clocks and avian malaria, a parasite known to be explicitly linked to biological rhythms, were explored in wild nestlings.

Altogether, this thesis provided new insights into how an animals' environment shapes clocks in the wild, with advances for both the fields of chronobiology and avian ecology research.

7.1 Advancements in avian ecology

This thesis provided further understanding of how biological rhythms of wild great tits are shaped by the environment. Previous literature has shown rhythms in city birds to differ from their forest counterparts in two ways: timing of daily activity (Kempenaers et al., 2010), and period length of endogenous clocks (Dominoni et al., 2013). In Chapter Three of this thesis, individual incubation rhythms were quantified for city and forest birds. There were clear differences in the onset and offset of daily activity between incubating city and forest birds, where city birds arose earlier and retired later than forest conspecifics.

Differences in behavioural timing between city and forest birds may have implications for fitness, given that important aspects of the environment, such as the availability of food, are time-dependent (Helm et al., 2017). As food can be scarce in the city habitat (Pollock et al., 2017), or of poorer quality (Isaksson & Andersson, 2007), the pressure on city birds

to rise earlier may be even greater than in the forest. This may create a trade-off for nestling parents, where the benefits of rising earlier to meet food demand for the energetic costs of incubation (Visser & Lessels, 2001), or nestling provisioning (Naef-Daenzer et al., 2000), will be balanced with the costs of sleep deprivation.

In addition, city sites used in Chapter Three had high levels of ALAN compared to forest sites, where mean light levels at city nest boxes were 55 lux, and forest nest boxes 0.44 lux. In a nest box study on great tits, ALAN increased begging behaviour of nestlings during the night (Raap et al., 2016c). Any reductions in sleep from rising earlier, for foraging opportunities, and through sleep disruption from light at night or begging offspring (Raap et al., 2016c) may mean that city birds have reduced energy (Laposky et al., 2008) to provide for nestlings. Given that city nestlings often weigh less than their forest conspecifics (Pollock et al., 2017), any reductions in provisioning by parents might have negative implications for chick survival (Both et al., 1999).

Site-level differences may have also played a role in the observed differences in activity timing between forest and city birds. For example, ambient temperatures are usually higher in city areas than surrounding areas, due to urban heat-island effects (Oke, 1982; Zhang et al., 2010). Warmer ambient temperatures influence the phenology of vegetation and insect food that breeding birds rely on (Deviche & Davies, 2014), and consequently, warming ambient temperatures influence great tits to breed earlier (Schaper et al., 2017). Great tit individuals that have earlier seasonal breeding have been shown to also have an earlier chronotype (Graham et al., 2017), which may help to explain earlier rising of city birds in this thesis. Intraspecific competition for food resources may also be higher in the city, where there are higher densities of bird populations (Partecke, 2014), which may increase pressure for city birds to rise earlier and be competitive.

Many studies have already shown the detrimental effects of ALAN on the condition of wild birds (Kernbach et al., 2018; Raap et al., 2016c). However, this thesis provided further evidence for negative effects of ALAN on wild bird fitness, and also novel insight into how ALAN may impact on mechanisms of the avian circadian clock in wild birds, by suppression of the core clock gene *REVERBA* (Chapter Four). This thesis also provided further evidence that ALAN exposure affects the condition of birds, given that nestlings exposed to low-level light at night treatment in Chapter Four were lighter than those under dark night. Experimental exposure to ALAN did not affect fledging success, however, small differences in nestling weight can have negative effects on survival and recruitment to the general population post-fledging (Both et al., 1999). From the findings of this thesis,

it is proposed that wild birds may be affected by ALAN in two ways – by directly impacting on circadian clock mechanisms of nestlings (Chapter Four), or indirectly on the behavioural rhythms of parents (Chapter Three).

Biological rhythms play an important role in regulation of the avian immune system (Markowska et al., 2017; Scheiermann et al., 2013). If natural rhythms of birds are disrupted as a result of ALAN, birds living in a city environment under light pollution may be immunocompromised, and therefore more vulnerable to disease (Kernbach et al., 2018). However, in Chapter Five, prevalence of *Leucocytozoon* parasites was significantly lower at city sites than at forest sites. Immune responses of avian hosts to parasites may be mediated by hormones such as corticosterone, the dominant stress steroid in birds (Romero, 2004). In some bird species, levels of corticosterone are elevated under urban conditions (Martin & Boruta, 2014). In a study on red-winged blackbirds (*Agelaius phoeniceus*), a higher endogenous concentration of circulating glucocorticoids was associated with reduced costs of infection with *Plasmodium* and *Haemoproteus* parasites, which indicated higher levels of tolerance (Schoenle et al., 2018). The observed lower prevalence of *Leucocytozoon* at city sites may have been due to increased tolerance in urban great tits. Future studies might explore how stress hormones such as corticosterone affects disease transmission and tolerance of birds in urban environments.

Additionally, the lower rate of infection with *Leucocytozoon* at city sites may be due to a lower abundance of insect vectors. *Leucocytozoon* vectors were not quantified in this study, yet abundance of another vector, the biting midge (*Culicoides*), was higher at forest sites than at city sites in previous years (Heather Ferguson, personal communication). Introductions of avian malaria parasites have proved damaging to naïve ecosystems (Atkinson et al., 1995). If avian malaria vectors continue to move into urban areas through landscape alterations or climate change, city birds may have a reduced ability to cope with acute infections (Dadam et al., 2019; Jiménez-Peñuela et al., 2019).

In this thesis, *Leucocytozoon* infection did not have a significant effect on nestling condition (Chapter Five). However, infections with *Leucocytozoon* significantly suppressed transcript levels of many immune genes in Chapters Five and Six. This may have been due to *Leucocytozoon* parasites infecting thrombocytes that are involved in production of immune responses to pathogens (Zhao et al., 2015). A suppressed immune system as a result of infection with *Leucocytozoon* may leave nestlings susceptible to further infections and also may have consequences for life histories (Hamilton and Zuk, 1982).

7.2 Advances for field chronobiology

This thesis provided new insights for chronobiology for quantifying rhythms of individual animals in their natural environments. In Chapter Three, individual chronotype was quantified for adult female birds from timing of repeatable behavioural activities during incubation. Patterns of incubation activity had been measured before in wild birds (Bulla et al., 2016; Graham et al., 2017), but chronotype had previously not been compared between city and forest habitats. Furthermore, uniquely in this study, individual chronotype was linked to fitness traits, although the effects of chronotype were non-significant. The results of this study add to the current debate in chronobiology that differences in individual timing are under selection (Helm and Womack, 2018; Kempenaers et al., 2010; Poesel et al., 2006).

In addition to behavioural rhythms, another main aim of this thesis was to quantify temporal changes in circadian clock gene transcripts as a proxy for internal clock mechanisms. Previously, this had not been done for a wild animal. In Chapter Four, time of day differences in transcript levels in great tit nestlings were quantified for many of the clock and immune gene targets studied. This thesis provided initial tentative evidence that it is possible to quantify biological rhythms both at the behavioural and transcript levels, which paves the way for future explorations of circadian rhythms of birds and other wildlife in chronobiology studies.

7.3 Challenges of measuring biological rhythms in wild birds

7.3.1 Measuring transcript levels from RNA

A main aim of this thesis was to quantify individual rhythms at the transcript level in a wild animal. To do this, a platform for gene expression analysis, the QuantiGene[®] Plex assay, was used to measure transcript levels of avian clock and immune genes from whole blood samples. The advantage of using the QuantiGene[®] Plex assay for this study was the ability to simultaneously measure transcript levels of all desired gene targets from one small blood sample. This reduced the burden of taking multiple blood samples from nestlings, and the analytic burden of a large series of qPCRs.

This assay has previously been used for samples collected from many species from fish (Mills and Gallagher, 2016), and mice (Song et al., 2015) to insects (Tan et al., 2016). However, avian samples had so far been limited to tissue from domestic chickens

(Gunawardana et al., 2019; Xue et al., 2013). Consequently, the assay had to be optimised for use on passerine blood samples for this thesis. Initial runs of the assay on great tit samples produced data with low fluorescence signal. Therefore, steps were taken to optimise the assay efficiency, such as increasing the time taken for initial digestion of the samples, or diluting samples to increase likelihood of probe binding to target RNAs. Following optimisation, there was marked improvement on the level of fluorescence signal from previous runs, however signal remained low overall for many of the gene targets.

For qPCR, the current standard method of quantifying gene transcript levels, it is often the case that lowly expressed genes show more variable results. In a similar way, the low fluorescence signal resulting from QuantiGene[®] Plex assay may have contributed to the noisiness of gene expression data in this thesis, and the lack of rhythmicity observed in circadian clock gene transcripts in Chapter Six aside from *REVERBA*.

In this thesis, the QuantiGene® Plex assay facilitated quantification of circadian transcript levels in a wild animal, for the first time. However, despite the optimisation steps taken to improve signal, fluorescence remained low. Future studies might further optimise the assay protocols for use of the QuantiGene® Plex on passerine blood samples.

7.3.2 Sampling distributions

This thesis provided support for taking samples across a 24 hr day during ecological studies of clocks. Inclusion of night time samples in Chapter Six showed effects of *Leucocytozoon* infection on several genes studied (*IGF1*, *PRKCA*, *NKRF*, *IL1*, *IL6*, *LY86*, *GATA3*, *TLR4*) not seen during day time sampling in Chapter Five, and therefore differences in results between chapters may be explained by temporal fluctuations of these genes across a day. This thesis highlighted the challenge for measuring temporal differences in gene transcripts, as limiting sampling to just a few timepoints resulted in a snapshot of genes, without knowing when the daily peak occurred.

However, in Chapter Four, transcript levels of many genes (*IGF1*, *TLR4*, *IL1*, *IL6*, *PRKCA*, *GATA3*, *IKBA*, *BMAL1*, *CLOCK*, *CK1E*, *AANAT*) did show time of day differences in transcript levels when comparing noon and midnight. These same genes did not show up as rhythmic during analysis of the full time profile in Chapter Six. As samples for Chapter Four were collected during the 2017 field season, and in Chapter Six were collected across two seasons (2016 and 2017), it may be that differences in years and distribution explains these discrepancies. Furthermore, in Chapter Six, sample size for several timepoints across

the time profile were often limited in number to <5 samples. An increase in samples for each of these timepoints would have improved the power of analysis, and therefore increased likelihood of detection of rhythmicity in all gene targets.

7.3.3 Rhythmicity

In Chapter Six, it was expected that temporal expression of gene transcripts of the core circadian clock (e.g. BMAL1, CLOCK) would be rhythmic. Unexpectedly, no genes out of those studied appeared to be rhythmic. This may be due to the issues of noisy transcript data from QuantiGene® Plex discussed earlier. However, another potential reason behind the lack of observed rhythmicity in circadian clock transcripts is differences in expression between blood and tissues. In this thesis, whole blood samples were used in place of tissue, to minimise impact on nestling welfare. Circadian transcripts have previously been detected in human blood (Archer et al., 2014; O'Neill and Reddy, 2011). In birds, rhythmic expression of circadian transcripts have been found in tissue samples such as ovaries (Laine et al., 2019), brain, liver, heart (Karaganis et al., 2009), spleen (Naidu et al., 2010) and retina (Bailey et al., 2002). Studies using bird blood for analyses of rhythmic gene transcripts are rarer, although one study using blood leucocytes of chickens observed day night differences in levels of immune genes *IL-6* and *IL-18* (Turkowska et al., 2013). Furthermore, great tit whole blood samples have been previously used for transcriptomics (Videvall et al., 2015; Watson et al., 2017). However, the possibility remains that circadian transcripts are not expressed at a detectable level in avian blood, or else are lowly expressed compared to other tissues. Wild studies ought to explore other possibilities of non-lethal sampling for quantifying rhythms, such as the use of fibroblasts (Gaspar and Brown, 2015).

Altogether, this thesis provided new insights into how molecular tools can be used to quantify rhythms in wild birds, and pioneered application of a new method for measuring circadian gene transcripts from avian blood.

7.4 Suggestions for future directions

7.4.1 Further directions

This thesis showed evidence for disruption of natural rhythms as a consequence of living in the city environment in Chapter Three. Also, in Chapter Four, nestlings under ALAN treatment weighed less than those not exposed to ALAN. Previous nest box studies on the effects of ALAN exposure on great tits have shown a reduction in sleep behaviours (Raap et al., 2016c), possibly as a result of circadian disruption. To investigate whether this disruption is contributing to reduced sleep, future studies ought to test for markers of sleep deprivation in city birds or birds experimentally exposed to ALAN.

As discussed in Chapter Four, the intensity of light used in this study (1 lux) may not have been sufficiently bright to disrupt rhythmic transcript levels of genes. Avian responses to ALAN are light-intensity dependent for behaviour (de Jong et al., 2016) reproductive biology (Dominoni et al., 2018; Zhang et al., 2019) and the endocrine system (Ouyang et al., 2018), and therefore it is likely that transcript level responses are also dependent on the intensity of light. Future studies that investigate effects of ALAN on avian immune and circadian clock systems should incorporate varying intensities of light treatment (as in de Jong et al., (2016)), to elucidate whether transcript level responses are light-intensity dependent.

In this thesis, environmental features such as ambient noise and light were not quantified for study sites. Previous ecological studies have shown ambient noise to have impact on rhythms of activity (Fuller et al., 2007; Gil et al., 2015), and others suggest that light has more of an impact than noise (Dominoni et al., 2014). Future studies should quantify environmental features and include these in analyses, to untangle effects of light and noise on rhythms and fitness.

A limiting factor that should also be considered is the number of study sites (two city sites and two forest sites) used in this thesis. Given that the nest box occupancy rate of great tits at all study sites was low, the decision was made to compile data from both urban sites within Glasgow and treat these as "city" and the two forest sites within Loch Lomond National Park as "forest". This increased the sample size for forest and city comparisons throughout the thesis but may have also introduced issues with pseudo replication. For example, it may be that nestlings at Kelvingrove Park react to light at night exposure in a similar way to each other, and differently from birds in the other city site, Garscube Estate. Kelvingrove Park therefore may be unrepresentative of an "urban site". Future studies might consider including more city and forest study sites, and treating these as separate sites within analyses, to reduce possibility of pseudo replication. Additionally, for each site, the degree to which the habitat is urbanised should be quantified, to allow for comparisons between city sites with other avian urban ecology studies.

Furthermore, a focus of this thesis was on nestlings, an important life history stage. From Chapter Four, there was some evidence that ALAN exposure had a cumulative effect on the weight of nestlings. A suggestion for future studies exposing nestlings to ALAN is to re-capture ALAN treatment birds as adults, to see if later life history stages are affected. A similar re-capture study could be carried out on nestlings that were found to be infected with *Leucocytozoon* (Chapters Five and Six), to investigate whether adult birds were able to clear infections, or to explore effects of chronic malaria infection on fitness as in Knowles, Palinauskas and Sheldon, (2010).

In Chapter Six, temporal expression of immune and clock gene transcripts were quantified in nestlings. However, this study did not take into account the temporal activities of malaria parasites themselves. In future studies, the time profile of infection intensity with *Leucocytozoon* parasites ought to be quantified using blood smears, to better understand life cycle of parasites before capturing their rhythmicity. This temporal profile can then be matched to that of clock gene transcripts of infected birds, to further investigate temporal host-parasite interactions.

7.4.2 Further analyses

Statistical analyses within this thesis may be improved by carrying out the following additional steps. Inclusion of coefficients of variance for the random effect of nest box in all statistical models would provide insight into between-nest effects of treatment. For example, analyses of nest box variance in Chapter Four may provide information on whether nest boxes (families) are all equally affected by ALAN. Assessing significance of this random effect of nest box within models would also inform how much within treatment variance is explained by individual differences in e.g. nestling weight.

Additionally, inclusion of forest plots (such as that in Videvall et al., 2015) would be beneficial in Chapters Four and Six. This would help to better visualise the size and direction of the effects of each variable in model (e.g. sex, brood size, hatch date) on clock and immune gene transcripts, and also the direction of their responses to ALAN treatment, or *Leucocytozoon* infection.

7.5 Conclusions

Ultimately, this thesis provided new insights into how biological rhythms of wild birds can be shaped by their environment, at behavioural and transcript levels, and how differences in individual timing may have consequences for life histories. This thesis provided evidence for disruptive effects of artificial light at night on elements of the circadian clock and condition of wild birds, and evidence for differences in timing of activity of birds living in the city habitat. In an increasingly urbanised world, it is important to understand the effects of human activity on wildlife, and to take steps to mitigate the negative impacts of artificial light. Although this thesis focused on one model species in ecology, the great tit, the findings and methods used in this thesis may have wider applications for future studies of wild clocks in other species.

Appendix 1

QuantiGene® Plex RNA Assay Probe information

Table 7.1. GenBank accession numbers and probe size (base pairs) for each gene target used in the QuantiGene® Plex RNA Assay. CE = Capture Extender, LE = Label Extender, BL = Blockers

Bead#	Accession (GenBank)	Symbol	Probeset Region	Full Size	CEs	LEs	BLs	Total
21	XM_015645462	AANAT	180-577	1274	6	10	6	22
61	XM_015630436	BMAL1	674-1080	2608	6	10	3	19
36	XM_015624072	CLOCK	2087-2747	8226	6	12	11	29
53	XM_015628865	CRY1	594-978	4505	6	10	2	18
46	XM_015641820	CSNK1A1	381-746	2994	6	10	1	17
30	XM_015627700	CSNK1E	743-1125	2580	6	10	2	18
33	XM_015629245	GATA3	595-1010	2980	6	10	5	21
12	XM_015649617	HMBS	396-778	1509	6	10	1	17
43	XM_015628060	IGF1	349-792	8910	6	10	4	20
14	XM_015648784	IL-1	1310-1722	2485	6	10	4	20
39	XM_015651893	IL6	310-751	9027	6	10	6	22
44	XM_015618120	LY86	422-829	1398	6	10	1	17
22	XM_015634446	NFE2L2	377-741	2961	6	10	0	16
18	XM_015629571	NFKBIA	343-806	1779	6	12	4	22
47	XM_015626243	NKRF	617-1008	3191	6	10	2	18
13	XM_015651038	NR1D1	293-675	1436	6	10	5	21
55	XM_015637543	PER2	2713-3069	6569	6	10	1	17
51	XM_015645435	PRKCA	1587-2019	4135	6	10	4	20
34	XM_015651012	RPL19	101-513	751	6	10	7	23
29	XM_015619761	SDHA	565-1014	2307	6	10	4	20
15	XM_015644481	TLR4	1157-1608	6908	6	10	3	19

Table 7.2 Great tit target gene sequences for probes used in the QuantiGene® Plex RNA Assay

Gene	Sequence
IGF1	ACACTGTGTCCTACATCCATTTCTTCTACCTTGGCCTGTGTTTGCTTACC
	TT
	AACCAGTTCTGTTGCTGCTGGCCCAGAAACACTCTGTGGTGCTGAGCT
	GGTTGATGCTCTTCAGTTCGTGTGTGGAGACA
	GAGGCTTTTACTTCAGTAAGCCCACGGGGTATGGATCCAGCAGTAGAC
	GTTTACACCACAAGGGAATAGTGGATGAATGC
	TGCTTCCAGAGTTGTGACCTGAGAAGGCTGGAGATGTATTGTGCTCCA
	ATAAAGCCGCCTAAATCCGCACGGTCCGTACG
	TGCCCAGCGCCACACCGACATGCCAAAAGCACAAAAGGAAGTGCATT
	TGAAGAACACAAGTAGAGGGAACACAGGAAACA
	GAAACTACAGAATGTAGGAACATGCCATCCACAAGAATGAAGAATGA
	ATGTGCCATCTGCAGGATACTTTGCT
LY86	ATATTGCAACCCACACCTTTAACATCAGAGCTGCAATGGTCCTAAGAC
	ACAGCATCAAG
	GAACTCTATGTCAAGCTGGACATGATCATAAATGGAAAGACTGTCTTA
	ACCTACTCAGAGACGCTTTGTGGGCCGGGTCA
	TGATAAGCTTATTTTCTGTGGAATGAAGAAAGGAGAACATCTCTATTA
	TGAGGGACCAGTCACACTGGGAATCAAAGAAA
	TCCCTCAGGGAGATTACACACTTTCAGCAAAGCTGACTAATCAAGATC
	ACGTCACTGTCGCCGATTTTACTGTG
	AAAAATTATTTAGACTATTATTAGCAACAAACCAATTTCATGAAAGCT
	ACAGACTACCAAAACTATTCAAGTCAAGTAGG
	CTGCTTGCATGCTACAGTGATTCTGAAGA
RPL19	TGAGTATGCTCCGGCTCCAGAAGAGGCTGGCCTCCAGCGTCCTGCGCT
KI LI)	GCGCCAAGAAGA
	AGGTGTGGCTGGATCCCAACGAGACCAACGAGATCGCCAACGCCAAC
	TCCCGGCAGCAGATCCGGAAGCTGATCAAGGAT
	GGGCTGATCATCCGCAAGCCCGTGACCGTGCACTCCCGCGCCCGCTGC
	AGGAAGAACACGCTGGCCCGGCGGAAGGGCCG
	GCACATGGGCATCGGTAAGAGGAAGGGCACGGCCAACGCCCGCATGC
	CCGAGAAGGTCACCTGGATGCGGCGGATGCGGA
	TCCTGCGGCGGCTGCTCCGGCGCTACCGCGAGTCCAAGAAGATCGACC
	GCCACATGTGAGCCCTGTGCTCGACCCAGAGC
	CTGGAGGATCAGGGACAGGGGACAGG
AANAT	CTGTCCCCTGCACCTGGATGAGATCCGCCACTTTCTGAGCCTGTGCCCG
AANAI	GAGCTGTCCCTC
	GGCTGGTTTGAGGAAGGGCGGCTGGTGGCATTCATCATCGGCTCCCTC
	TGGGACCAGGAGAGGCTCAGCCAGCAGCACT
	GACCTGCACAAGCCACAGGGCTCAGCTGTGCACATCCACGTGCTGGC
	CGTGCACCGCCCCCCCAGCAGGGCAAGG
	GCTCCATCCTGATGTGGCGCTACCTGCAGCACCTGCGCTGCCCTT
	CGCCGCCGCCCTGCTCATGTGCGAGCAC
	TTCCTCGTGCCCTTCTACCAGAAGTGCGGCTTCGAGGCGCTGGGGCCCT
	GCGAGGTGACCGTGAGGGACCTGGCCTTCGT
TT 1	GGAGATGCAGCACCCCG
IL1	TGGCATTCGTCCCTGATTTGGACACACTGGAGAGCGGCAGCTTGAATG
	AGG
	AGACACTGTATGGCCCCAACTGTCTCTGCCTGCAGAAGAAGCCCCGAC
	TGGACTTGGAGGTGACATCACCTGAGGTGGGC
	ATCCAAGTGACAGTGACAAAAGGACACCTTTCCAGGACCTTCCGCCAG
	GCTGCCATCCTGGTGGTTGCTGTGACCAAGCT

	CCTAAAGCAGCCATCGCACAAGGACTTTGCTGATAGTGACCTGGGCAG
	CTTCTTGGATGATATTTTTGAGCCCATCTCCT
	TCCAGTGCATCAAGGGCAGTTATACCAGGCCACCCGTTTTCCGCTACA
	CTCGCTCCCAGTCCTTTGACATCCTGGACATT
	GACCACAAGTGCTTCGTACTGGAGTCACCCACCCAGCTGGTG
IL6	CCCAACGCAGA
	CTCCTCGGGGGACGCCGAGCTGGAGGAGGTGGCGAAGCGGCGGCGG
	CGCTGCCCGGCAGCCTGGAGCAGGCGCGCTGC
	TGCACGCCCGGGCGGCAGAGCTGCGGGACGAGATGTGCGAGAAGTTT
	ACCGTCTGCCAGAACAGCATGGAAATGCTCCTC
	CACAACAACCTCAACCTCCCCAAGGTGACGGAGGAAGATGGGTGTCTG
	CTCGCCGGCTTTAATGAGGATAAATGCTTGAG
	AAAAATCTCCAGTGGGCTTTATACATTTCAGATATACCTCAAATACAT
	ACAAGAAACTTTTATTAGTGAAAACCAAAATG
	TTGAATCGCTATCCCATACTACAGGGCACCTGGCACGTACCCTAAGAG
	AGATGCTGATCAATCCCAAAGAAGTGATCATC
NIETZDIA	CCCGATGCAGCAGCAGCAGCAGCAGCAGCAGCAGCAGCAGCAGCAG
NFKBIA	CTGACGGAGGACGGGGACACTTTTCTGCACTTGGCCATTATTCACGAA
	GAAAAATCCC
	TGAGCCTGGAGGTGATCCAGCAGGCAGCCGGTGACCGGGCTTTCCTGA ACTTCCAGAACAACCTCAGCCAGACTCCTCTT
	CACCTGGCAGTGATAACTGATCAGCCTGAAATTGCTGAGCATCTTCTG
	AAGGCTGGATGTGACCTGGAAATTGCTGAGCATCTTCTG
	CCGAGGAAACACCCCCCTGCACATTGCCTGTCAGCAGGGCTCCCTCAG
	GAGCGTCAGCGTGCTCACACAGTATTGCCTGTCAGCAGGGCTCCCTCAG
	CACACCACCTTCTCACTGTCCTGCAGGCCAACTACAACGGACATA
	CATGTCTTCATTTAGCATCTATTCAAGGATAC
	CTGGCTATTGTCGAGTACCTGTCTTGGGAGCAGATGTAAATGCA
	CAGGAGCCATGCAATGGCAGAACAGCACTGCA
	TTTGGC
CSNK1A1	AGCAGCAGCGCTCCAAGGC
	CGAGTTCATTGTCGGAGGCAAATACAAGCTGGTGCGGAAGATCGGGTC
	GGGCTCCTTCGGGGACATCTATCTGGCGATCA
	ACATCACCAACGGGAGGAAGTTGCTGTGAAGTTGGAGTCTCAGAAG
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HMBS	CAACTTCTCTCCTCGGCTTTACGATTGGCGCTGTCTGCAAGCATTC
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	CAGGGTGCC
CSNK1E	CCACCAGCACATCCCTTATCGGGAAAACAAGAATCTGACTGGCACAGC
COLUMN	CCGTTATGCCTCTATCAACACCCACCTGGGAA
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	CGCAACCGCTGCCAGCAGTGCCGCTTCAAGAA
	GTGCCTGCTGGCATGTCCCGTGATGCCGTGCGCT
CRY1	CTTGAAGAGCTGGGTTTTGACACAGATGGTCTGCCTTCTGCAGTATGG
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NFE2L2	GCTGGAGTTAGACGAAGAGACAGGTGAATTCATTCCTGTGCAGCCAGC
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NKRF	TTTGCAGGCCCAGCAGCAGCC
NKKr	AAGAAGGCTGTGAATAATGTGGACTCTACCACAAGCAGCGCCTTGCAG
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	TOTAL MECCE I GAATATATITATOTI CCTCTOAAAOAOATTOCCCCTOCTO

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TLR4	AGCTGAGAGTGCTCCGTATTACCAAGAACAAAAGACTCAAAAAC
1214	TTCAGTCAGAATTTTGAGGGTCTCACTAACCTGGAGGTTATAGATTTGA
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