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Portable dehumidifiers condensed water: a novel matrix for the screening of semivolatile compounds in indoor air

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Highlights:

- Dehumidifier condensates contain qualitative information of indoor pollutants
- Above 140 compounds identified following a non-target data mining approach
- Variable condensate composition depending on the investigated environment
- Acceptable repeatability between samples obtained with different dehumidifiers
- Extraction efficiency correlated to Henry's law constant

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Abstract

The comprehensive identification of organic species existing in indoor environments is a key

issue to understand their impact in human health. This study proposes the analysis of

condensed water samples, collected with portable dehumidifiers, to characterize semi-

volatile compounds in the gas phase of confined areas. Water samples are concentrated by

solid-phase extraction (SPE). The obtained extracts are analysed by gas chromatography

(GC) time-of-flight mass spectrometry (TOF-MS), following a non-target screening data

mining approach. In first term, spectra of deconvoluted compounds are compared with those

in NIST low resolution library; thereafter, tentative identifications are verified using an in-

house database of accurate electron ionization (EI) MS spectra. Chromatographic (retention

index) and spectral data are combined for unambiguous species identification. The potential

of condensed water samples to reflect changes in the composition of indoor atmospheres,

the match between data obtained using different dehumidifiers, and the relative

concentration efficiency of condensed water compared to that attained by active sampling

of moderate air volumes are discussed. A total of 141 semi-volatile compounds were

identified (98 confirmed against authentic standards) in a set of 21 samples obtained from

different homes and working places. This list contains more than 40 fragrances (including

several potential allergens), solvents and intermediates in the production of polymeric

materials, plasticizers and flame retardants.

Keywords: Indoor air; semi-volatile compounds; dehumidifiers; non-target screening; gas

chromatography accurate mass spectrometry.

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1. Introduction

Indoor environments contain rich, complex mixtures of organic chemicals which have been correlated with allergies, respiratory and dermal diseases (Araki et al. 2018; Steinemann and Goodman, 2019). Some of these compounds have a natural (human) origin (Liu et al., 2017); however, most of them are related to building materials, upholstery, personal care and household products (Lamas et al., 2010; Lucattini et al., 2018), and even with life habits (i.e. smoking and consumption of illict drugs) (Cecinato et al., 2017; Pandey and Kim, 2010). In addition to parent compounds released from above sources, their degradation products might also impair the quality of air in indoor environments (Wakayama et al., 2019).

Compounds released in indoor areas are distributed between dust particles and the gas phase. Analysis of dust is useful to evaluate human exposure through accidental ingestion, whilst the identification of compounds existing in air is required to understand inhalation exposure (Geens et al., 2009). Dust samples are available from domestic vacuum cleaners; however, air sampling requires the use of dedicated gadgets. Active sampling devices employ a vacuum pump forcing a stream of air through a trapping unit, usually consisting of a filter followed by a sorbent (Laboire et al., 2016). Their concentration efficiency depends on the sampling time, the flow of air through the trapping unit (measured using a flowmeter or volumetric counter), and the retention efficiency of the sorbent. Low density sorbents, offering a low pressure drop, such as polyurethane foam (PUF), permit high flow rates of air for effective concentration of semi-volatile compounds (Melymuk et al., 2016). Carbonaceous materials, Amberlite polymers and Tenax display a high affinity for volatile compounds (Hoang et al., 2017). In this case, the flow rate of air through the sorbent and the sampled volume are reduced in comparison with protocols dealing with semi-volatile compounds. To compensate for a lower enrichment factor, trapped compounds are totally transferred to a gas chromatography (GC) instrument through thermal desorption of the sorbent (Ramírez et al., 2010). Passive sampling devices, such as semipermeable membranes filled with liquid or solid sorbents (Gale et al., 2009), microextraction techniques (Esteve-Turillas et al., 2009) and PUF disks (Soheib and Harner, 2002) represent another option to investigate the presence of volatile compounds in confined atmospheres. In this approach, the uptake mass of a given compound depends on thermodynamic factors (affinity for the sorbent) and kinetic variables (diffusion rate from air to the active sorbent).

Portable dehumidifiers are appliances designed to reduce the concentration of water in air from interior environments. They incorporate a fan to create a stream of air, which passes first through a filter (to remove coarse particles), and then gets in contact with a cold region (usually the evaporator of a previously compressed cryogenic fluid), where water vapor condenses to be recovered in a draining tank. The high flow rate of air circulating through the dehumidifier, controlled by the turning speed of the fan, is expected to lead to a significant concentration of semi-volatile compounds in the condensed water. A previous study has described the presence of more than 40 compounds in condensed water obtained from the heater exchanger units of central air conditioning systems (Roll et al., 2015). Authors demonstrated the repeatability of chromatographic profiles for samples obtained from the same building in different dates; moreover, they showed how changes in daily activities affected the composition of water condensates (Roll et al., 2015). Unfortunately, central air conditioning facilities are not available everywhere, they do not provide information regarding the spatial distribution of organic compounds in different rooms from the same house, and the potential effect of each air conditioning facility in the composition of water condensates is not easy to evaluate.

The aim of this manuscript is to investigate comprehensively the range of organic compounds existing in condensed water samples obtained from portable dehumidifiers. To this end, samples were concentrated by solid-phase extraction (SPE). Thereafter, the SPE extracts were characterized by GC combined with accurate mass spectrometry (MS), based on a time-of-flight (TOF) instrument. Compounds identification was carried out by comparison of accurate electron ionization (EI) MS spectra of deconvoluted compounds with those included in the low resolution NIST database; followed by additional confirmation with

an in-house database (personal compound database library, PCDL) of suspected targets (Castro et al. 2019; Gómez-Ramos et al., 2019). Concordance of results obtained using different dehumidifiers, placed in the same room, and major changes in the composition of drained water depending on the sampling place are reported. In addition, the concentration efficiencies attained using portable dehumidifiers and active sampling, of moderate air volumes, are compared and correlated with the properties of identified compounds.

2. Material and methods

2.1. Solvents, standards and sorbents

Ethyl acetate (AcOEt) (trace analysis grade) and methanol (MeOH) (LC gradient quality) were obtained from Merck (Darmstadt, Germany). Standards of compounds employed in the current study were provided by Sigma Aldrich (St. Louis, MO, USA), Riedel de Häen (Seelze, Germany) and Restek (Bellefonte, PA, USA). Compounds were acquired either as individual species, or as technical mixtures (case of some fragrances and other ingredients of personal care products). Stock solutions of each substance were prepared in MeOH. Further dilutions and mixtures were made in AcOEt. The concentrations of standards employed to build the PCDL of accurate EI-MS spectra were in the range from 0.5 to 2 μg mL⁻¹. A standard mixture of n-alkanes (C₈-C₄₀) in dichloromethane was provided by Supelco (Bellefonte, PA, USA). This solution was used to calculate the linear retention index (LRI) of compounds in chromatograms of sample extracts and those for standards included in the PCDL. LRI and accurate EI-MS spectra were employed for confirmation of tentative identifications derived from search of deconvoluted experimental spectra in the NIST database. PCB-30 (2,4,6-trichloro biphenyl) was used as injection standard, added to SPE extracts before GC-MS analysis.

OASIS HLB (200 mg) cartridges were purchased from Waters. This polymer was employed for the concentration of condensed water, and also as trapping sorbent for active air sampling. Dehumidifiers employed to obtain water samples from different indoor areas were purchased in local markets. They were furnished with a cryogenic fluid evaporator where

water vapor condensates. A grid in the back side of these appliances served as a filter to limit the entrance of coarse particles into the evaporator unit. In some experiments, the grid was covered with glass fibre filters (GFF), pore size of 0.7 µm, provided by Sigma-Aldrich.

The membrane pump employed for active sampling of indoor air was purchased from Vaccubrand (model MD 4NT). It was connected to an HLB cartridge, followed by a rotameter. Air was sampled at a flow rate of 16.5 L min⁻¹, similar to values considered in previous studies dealing with determination of fragrances (10 L min⁻¹) (Lamas et al., 2010) and other semi-volatile compounds from indoor air (12 L min⁻¹) (Laboire et al., 2016). Unless otherwise stated, the pump was operated 5 h to concentrate 5 m³ of air.

2.2. Samples and sample preparation

Dehumidifiers were placed in several indoor environments, including private houses and working places. After a given sampling time (5-12 h) condensed water samples were transferred to glass vessels, transported to the laboratory and stored at 4 °C, for a maximum of 24 h, before SPE extraction. Table 1 summarizes some data regarding the sampling places, and the codes assigned to the obtained samples. In brief, samples were taken in 10 different homes, two administrative buildings, a hair dressing in a shopping centre, and the research laboratory where SPE extractions were carried out. Additional information, including temperature and relative air humidity at the beginning of the sampling step, and the collected water volume are provided as supplementary material, Table S1.

Field blanks were prepared filling the drain vessel of dehumidifiers with ultrapure water. After a contact time of 12 hours, water was concentrated by SPE. Another set of blanks corresponded to SPE of ultrapure water without contact with dehumidifiers.

Water samples (300 mL) were concentrated using HLB cartridges previously conditioned with AcOEt, MeOH and ultrapure water (2 mL each). Following the concentration step, the sorbent was rinsed with 5 mL of ultrapure water, dried using a gentle stream of nitrogen and eluted with 2 mL of AcOEt. This extract was fortified with PCB-30 as injection standard, at

1000 ng mL⁻¹, to check the stability in the responses of the GC-MS system. Potential breakthrough problems were investigated by passing the water sample (300 mL) through two cartridges connected in series. After the concentration step, they were disconnected and eluted separately.

HLB cartridges employed in active sampling experiments were conditioned with AcOEt (2 mL). Thereafter, they were dried and wrapped in aluminium foil. During active air sampling experiments, the open edge of the cartridge was faced to floor (floor distance was 0.5 m, similar to that existing in the air entrance of dehumidifiers) to prevent direct deposition of particulate matter on the frit above the sorbent. Breakthrough studies were carried out by spiking a standard mixture of fragrances, with different volatilities, in a plug of glass wool placed inside the barrel of the SPE cartridge, above the sorbent. In these experiments, a 2nd cartridge was connected in series. After sampling a given volume of air, cartridges were eluted separately with 2 mL of ethyl acetate. The plug of glass wool was sonicated with the same volume of solvent for 5 min. In order to compare the range of compounds concentrated using active sampling and dehumidifiers, both systems were simultaneously operated in the same room.

2.3. GC-MS system and data analysis

Identification of compounds in indoor air related extracts was performed with a GC-QTOF-MS instrument, obtained from Agilent (Wilmington, DE, USA). The system was comprised of a 7890A gas chromatograph connected to a 7200 QTOF MS spectrometer, which was furnished with an EI source. The TOF was operated in the 2 GHz mode, offering typical mass resolution values (FWHM) of 6500 at m/z 131. The GC-QTOF-MS system was employed in the single MS mode, that is, as a GC-TOF-MS instrument. MS spectra were obtained at 2.5 Hz (5430 scans accumulated per spectrum) in the range of m/z values from 40 to 600. The m/z axis in the mass analyser was automatically recalibrated, every 3 injections, by infusion of perfluorotributyl amine (PFTBA) in the EI source.

Two different columns were tested for the separation of compounds exiting in SPE extracts. They were a BP-5 MS (30 m x 0.25 mm i.d., 0.25 µm film thickness) and a DB-WAXETR one (30 m x 0.25 mm i.d., 0.5 µm film thickness). Both were acquired from Agilent. The carrier gas (Helium) flow rate was 1.2 mL min⁻¹. Standards and sample extracts (2 µL) were injected in the splitless mode, with the injector maintained at 260 °C. The splitless time and the split flow were 1 min and 50 mL min⁻¹, respectively. The quadrupole and the EI source temperatures were 150 °C and 230 °C, respectively. Two different oven temperature programs were employed. For the BP-5 column the gradient was 50 °C (1 min), rated at 10 °C min⁻¹ to 290 °C (15 min). When using the DB-WAXETR one, temperature varied as follows: 60 °C (1 min), 1st rate at 5 °C min⁻¹ to 220 °C, 2nd rate at 20 °C min⁻¹ to 240 °C (15 min). The temperature of the transfer line between the GC and the MS spectrometer was the same as the upper value employed with each column.

The *MassHunter* software was used to control the acquisition parameters in the GC-QTOF-MS system and to process the obtained data. *MassHunter Qualitative* software (version B.08.00) was employed during inspection of raw GC-TOF-MS chromatograms, to extract the accurate EI-MS spectra of standards and to transfer these spectra to a customized PCDL. In addition to accurate spectra, the PCDL database included LRI values.

Deconvolution of chromatograms obtained for indoor air related samples was carried out using the *Unknowns Analysis* (UA) function (based on the *SureMass* algorithm), integrated in the *MassHunter Quantitative* software (version B.08.00). Responses of deconvoluted species (peak areas corrected with that obtained for PCB30) were employed to evaluate the severity of contamination problems in procedural blanks, to compare the relative efficiencies of active air sampling versus concentrated dehumidifier water samples, to investigate the responses obtained for a given compound depending on the sampling place, and to evaluate the agreement between data obtained with different dehumidifiers.

The *MS Search* (v. 2.3) software was used to manage the spectra compiled in the NIST17 low resolution EI-MS library, and to calculate the theoretical *m/z* ratios of fragment ions with known structures, using the NIST-MS interpreter tool, in this database.

The workflow for the non-target identification of pollutants in sample extracts has been described in detail in a previous publication (Castro et al., 2019). In brief, it involves deconvolution of raw GC-TOF-MS chromatograms, followed by spectral comparison with the low resolution NIST17 and the accurate PCDL libraries. The 2nd database was updated continuously, taking into account tentative identifications derived from comparison with the NIST database. A scheme of the employed workflow is provided as supplementary information, Fig. S1. Currently, the created PCDL contains 265 inputs of volatile and semi-volatile species.

3. Results and discussion

3.1. Preliminary experiments

A typical chromatogram obtained with the DB-WAXETR column for dehumidifier condensed water is provided as supplementary information, Fig. S2A. One of the latter, intense eluting peaks corresponded to dibutyl phthalate (DBP). Less volatile species, such as di-(ethylhexyl) phthalate, tri-(butoxyethyl) phosphate, octocrylene and certain dyes (such as indigo), which are recognized as ubiquitous in dust from indoor areas (Castro et al., 2019; Fromme et al., 2004) were absent in this chromatogram. In order to verify that these less volatile compounds did not remain inside the employed column, the same extract was injected in the BP5-MS column, using a faster gradient and also a higher final temperature (290 °C versus 240 °C in the DB-WAXETR). DBP was again the latter intense peak in the BP-5 chromatogram, Fig. S2B. Thus, either less volatile species are mostly associated to indoor particles instead of to the gas phase (Weschler and Nazaroff, 2014), or they are not concentrated in condensed water samples. So, the more retentive and higher selectivity Carbowax-type column was selected for the rest of the study.

As regards SPE of dehumidifier water, considering a sample volume of 300 mL, rich chromatographic records (Fig. S2) were obtained without noticing breakthrough problems during the concentration step, just traces of some carboxylic acids were found in the security SPE cartridge (their peak areas represented between 1 and 2% of those observed in the extract from the first cartridge, figure not shown). Condensed water samples were concentrated without previous filtration to minimize the risk of losing any compound recovered from the indoor atmospheres. AcOEt was used for the elution of SPE cartridges considering its medium polarity and compatibility with GC-MS analysis. A volume of 2 mL was enough to recover species retained in the SPE sorbent.

The coarse grid of dehumidifiers might lead to the presence of dust particles in the drained water. Consequently, compounds in SPE extracts might correspond to species existing in the vapor phase, but also to those solubilised from dust particles. To investigate this possibility some experiments were carried out fitting a GFF filter behind the coarse grid of the dehumidifier. After operating the system for 48 h, filters were sonicated with ethyl acetate and the extract concentrated to 2 mL. The corresponding chromatograms were similar to those obtained for blank filters. Moreover, no significant peaks were noticed in the filter extract compared to those in the chromatogram of condensed water (Fig. S3). So, the stream of air created by the fan of the dehumidifier does not introduce particulate matter with a size larger than the cut-off of the filter (0.7 µm) in the cooling unit of the system.

Efficiency of the HLB polymer to trap volatile compounds during active sampling (16.5 L min⁻¹) of indoor air was investigated with a mixture of 16 terpenes with different volatilities (their chromatographic retention times varied from 7.81 min for limonene to 30.54 min for carvacrol, see Table S2). Experiments were performed as described in section 2.2. After sampling 5 m³ of air, the extracted ion chromatograms (EIC), mass window 2 mDa, for their more representative ions (Table S2) corresponding to the glass wool plug, the retention and the security HLB cartridges were compared. Compounds were noticed only in the extract from the retention cartridge.

3.2. Database of accurate EI-MS spectra

Mass accuracy of the GC-EI-TOF-MS system employed in this study (first generation of Agilent TOF combined with GC) was in the range of 0.002 Da, equivalent to relative mass errors from 20 to 6.7 ppm for ions with m/z ratios in the range from 100 to 300. Although these errors have been lowered by 2^{nd} generation TOF instruments (Stupak et al., 2018), and particularly by Orbitrap mass analyser (Gómez-Ramos et al., 2019), they sufficed to discriminate between ions with different empirical formula (considering C,H,O and N as their elements) and same nominal m/z values.

During non-target analysis of complex samples, ions from two or more compounds are often merged under the same chromatographic peak; thus, spectral deconvolution is mandatory previously to compounds identification. Even after deconvolution, more often than desired, spectral data alone are not enough for the unambiguous identification of a given compound. Fig. 1A depicts a region of a deconvoluted compounds chromatogram. Spectra assigned to peaks in blue are shown in Fig. 1B and 1C, whilst the accurate spectrum for the best candidate (benzaldehyde) appears in Fig. 1D. The normalized scores (0-100) corresponding to the match of deconvoluted spectra with that recorded for benzaldehyde remained above 90 in both cases (Fig. 1B and 1C); moreover, the differences with m/z values of ions in the PCDL spectra of benzaldehyde were lower than 1.3 mDa (absolute value, Fig. 1D). Thus, without using chromatographic data (retention times or LRI) it would be practically impossible to discriminate which of the two peaks corresponds to benzaldehyde. In this case, the component at 15.97 min matched the LRI of benzaldehyde. Another example of compounds with practically identical EI-MS spectra are the fragrances linalool (C₁₀H₁₈O, CAS number 78-70-6) and linally acetate (C₁₂H₂₀O₂, CAS number 115-95-7), Fig. S4A and S4B. In the latter case, the molecular ion (M⁺) is absent and most fragment ions (Fig. S4A) match those in the spectrum of linalool. Additional problematic cases are positional isomers. Very often, they show practically identical EI-MS spectra, as they are the cases of

ethylbenzene and xylenes, di-butyl and di-isopropyl phthalates, thymol and carvacrol (Fig. S4C and D).

In order to overcome limitations reported in the previous paragraph, the PCDL was implemented with LRI values. Thus, positive identifications were based on spectral and LRI matches between deconvoluted compounds in extracts from indoor air-related samples and the PCDL database. The minimum normalized score (0-100) for the spectral match was set at 70, and the maximum difference with database retention time 10 s (equivalent to a variation in LRI values of 8-10 units). A 3rd requirement for a positive identification was a maximum relative error of 20 ppm for any intense ion in the deconvoluted spectra compared to that in the PCDL spectrum of the candidate. These three conditions were verified by the *Unknowns Analysis* software during library search of deconvoluted components. Despite the restriction in the retention time window, in case of some isomeric species with very close retention times, such as the diastereomers of galaxolide, or *m*-xylene and *p*-xylene, revision of identities assigned by *Unknowns Analysis* was required.

3.3. Dehumidifiers as samplers of semi-volatile compounds from indoor areas

The chromatograms of dehumidifier water samples contained a high number of peaks related to species existing in indoor air (Fig. S2 and Fig. 1); however, in order to demonstrate the usefulness of this matrix to characterize the composition of indoor air, several features need to be considered. Firstly, it is necessary to check that compounds in condensed water are not contaminants introduced by dehumidifiers. Changes in the composition of the indoor environment must be reflected in the presence of new chromatographic peaks, and/or in variations in their response intensities in dehumidifier water extracts. In addition, similar chromatographic profiles should be obtained for the same environment when using different dehumidifiers.

In general, the chromatograms of procedural blanks were very similar to those obtained for SPE of ultrapure water, not in contact with dehumidifier draining tanks. However, for two of

the tested dehumidifiers we found traces of tris (1,3-dichloro-2-propyl) phosphate (TCPP) and tributyl acetylcitrate in procedural blanks, so these devices were not used in further experiments. In addition to substances released from the draining vessel, the stream of air and the condensed water drops might release organic compounds from other elements of dehumidifiers before reaching the draining vessel. Although this problem is difficult to detect, we have compared the normalized responses (peak area divided by the peak area of PCB30) for a series of potentially problematic species (plasticizers, flame retardants, curing agents, fragrances...) in the extracts of water samples taken, with the same dehumidifier, in the living room from 6 different homes. Responses of selected compounds showed significant variations (in some cases up to 10-fold) depending on the sampling site, Fig. S5. Thus, if occurs, the release of above compounds from the dehumidifier unit would not mask variations among samples obtained from different environments.

Fig. 2 shows the ratio of responses obtained for selected compounds in the extracts from two water samples obtained from the same room (46 m³ volume). Samples were taken with the same dehumidifier, before and after installing an air freshener in the room (5 m distance from the dehumidifier). In addition to water extracts, the content of the freshener was diluted (10⁵-fold dilution with AcOEt) and injected in the GC-TOF-MS system. In case of species not detected in the diluted freshener solution, response ratios from 0.8 to 1.2 were obtained, Fig. 2A. On the contrary, for compounds found in the freshener liquid, their responses increased between 2 and 62 times in the sample obtained after activation of the air aromatization device, Fig. 2B.

The effect of the dehumidifier device in the composition of condensed water samples was investigated operating simultaneously two different appliances in the same room. Chromatograms obtained at sampling point S21 (Table 1) are shown in Fig. 3A. The volumes of water recovered by each instrument differed around 20%; however, similar chromatographic profiles were observed for both extracts. Fig. 3B shows the correlation plot between responses (logarithms of peak areas) for 50 identified compounds (after

deconvolution and comparison with the PCDL) in the chromatograms for both samples. An acceptable correlation (determination coefficient, R², 0.959) was observed; moreover, the slope of the plot (0.996) was close to the unit. Fig. 3C shows the equivalent plot for a second pair of dehumidifiers operating at sampling point S3. In this case, the graph compares the responses obtained for 70 positive identifications.

3.4. Comprehensive identification of semi-volatile compounds in dehumidifier water samples

The list of substances identified in dehumidifier water from the 21 sampling points considered in the current study (Table 1) is compiled in Tables 2A and 2B. Compounds are classified attending to their applications (in most cases same species present different uses), and sorted accordingly to their increasing LRI values. Compounds detected in procedural blanks were included in the above tables only when their responses stayed above 10-times that observed in the procedural blank of the corresponding dehumidifier. Species such as toluene, styrene and dimethyl sulfoxide were ubiquitous in condensed water samples, however, they did not satisfy the above requirement; thus, they are not included in the list of positive identifications. Table 2A gathers compounds with applications as fragrances and flavors. They are usually incorporated in personal care, cleaning and other household products. Some of them are natural products whose empirical formulae derive from that of terpene (C₁₀H₁₆), with one or two atoms of oxygen and variations in the number of unsaturated bonds. In other cases, they show a phenolic structure, as it is the case of salicylates, vanillin, eugenol and thymol. The list includes also synthetic musks, such as galaxolide and tonalide. Many compounds in Table 2A have been included in the list of potential contact allergens (Shibuta et al., 2016).

Regarding substances compiled in Table 2B, most of them belong to the category of solvents and intermediates in the production of high volume chemicals, followed by those used as preservatives and plasticizers. Within the first group, 2-ethyl-1-hexanol was ubiquitous in the processed samples. This compound can be released from polymeric materials; moreover, it can arise from hydrolysis of di-ethylhexyl phthalate (Wakayama et al., 2019), which has been

quantified in indoor dust at high concentrations (Castro et al, 2019). Another intermediate often noticed in samples was cyclohexanone, which is involved in the polyurethane foams (Glowacz-Crerwonka, 2017). Likely, phenol and aniline are released from adhesives, paints and cleaning fluids (Palmiotto et al., 2001). The most probable sources of compounds integrated in the category of preservatives are again personal and household products. The list of plasticizers whose presence was confirmed in dehumidifier water samples includes four phthalates, two citrates, a diisobutyrate (2,2,4-trimethyl-1,3-pentanediol diisobutyrate; TXIB) and diacetin. All of them are high production volume chemicals widely employed in building materials, and they have been previously detected in indoor air (Lucattini et al., 2018; Ueta et al., 2019).

Table 3 summarizes a list of 43 compounds tentatively identified (not confirmed by injection of authentic standards) in the processed samples. Compounds in this table showed spectral match scores higher than 90% when their experimental accurate MS spectra are compared with those in the NIST database. Moreover, the difference between m/z ratios in the deconvoluted spectra and the calculated values for ions with known structures in the NIST database remained below 2 mDa, at least for two intense fragment ions. When available, experimental LRI values of tentative identifications were compared with those obtained from Chemspider database (http://www.chemspider.com/), or directly for Carbowax-type columns in previous studies (Moniruzzaman et al., 2014). Most of them show similar applications to positive identifications compiled in Tables 2A and 2B. The larger families of compounds in Table 3 are benzene derivatives, carboxylic acids of natural origin, and several of their esters with industrial applications as plasticizers (particularly those of phthalic, hexanoic and isobutyric acid), glycols and terpene derivatives. Some compounds listed in Table 3 are closely related to those positively identified in Table 2. An example is 2-(methylthio) benzothiazole, which has been reported together with benzothiazole in vehicle tires (Zhang et al., 2018). Another case corresponds to the isomers of 1,3-pentanediol, 2,2,4-trimethylisobutyrate (Texanol) with similar applications to TXIB (Ueta et al., 2019).

Table 1 shows the identities assigned to the three most intense peaks in the chromatograms for condensed water samples from the 21 sampling sites involved in the current study. At least one plasticizer was within the above ranking for 14 out of 21 samples. Within the group of personal care and cleaning products, 2-phenoxyethanol was among the three most intense peaks in 5 sites. Nonanoic acid and diethylene glycol butyl ether were also often identified in the ranking of most intense peaks. Comparison of data for samples obtained at home 1 (sample codes S1, S2, S3, S4 and S5) revealed changes in the identities of the most intense chromatographic peaks between kitchen and garage (S4 and S17, respectively) with the other investigated rooms (S1, S2 and S3). Sample code S10 was obtained from the home of a smoker. In this case, the most intense chromatographic peak was identified as nicotine. In summary, even the major compounds in dehumidifier water extracts varied depending on the sampling area.

3.5. Dehumidifier condensed water versus active air sampling

Data in previous sections show that condensed water samples contain a cocktail of organic compounds potentially related with on-going activities, building materials and life style in indoor areas. The question is whether the qualitative information provided by dehumidifiers and active sampling are somehow correlated, or not. The raw chromatograms corresponding to active sampling (16.5 L min⁻¹) for 5 h through a HLB cartridge, and after concentration of condensed water sample obtained from a portable dehumidifier, both operating simultaneously in the same room, are shown in Fig. S6. The first part of the chromatogram for active sampling contains more intense peaks than those observed for condensed water (Fig. S6A). On the other hand, the latter chromatogram is richer in number and intensity of peaks with retention times above 17 min (Fig. S6B). Even when the sampling time was increased to 15 h, the 2nd half of the resulting chromatograms was relatively poor in peaks when compared to those obtained for dehumidifier condensed water (Fig. S6C).

Fig. 4 shows the logarithmic values of average response ratios for common compounds (sorted following their increasing retention times in the DB-WAXETR column), identified

using both sampling approaches, at four different sampling points (S2, S3, S7 and S20, Table 1). In general, the response ratios plot follows a similar trend that the negative logarithms of the Henry's law constants (K_H, atm m³ mol¹). That is, the relative extraction yield provided by dehumidifier condensed water samples increases with the decrease of the K_H values. Not surprisingly, K_H is defined as the ratio between the pressure of a given compound in the gas phase and its concentration in the equilibrium water phase. In general, for species with K_H values lower than 10⁻⁴ atm m³ mol¹ higher responses are observed after SPE of 300 mL of condensed water than for direct sampling of 5 m³ of air. In addition to compounds in Fig. 4, other species were only noticed using one of the two sampling techniques. For example, camphene and pinene (alpha and beta isomers), and the terpene-related species *p*-cymene were found in chromatograms corresponding to active air sampling, but not in the extracts from dehumidifier water. Their K_H values (data at 25 °C) stayed above 10⁻² atm m³ mol¹¹. A list of compounds present in dehumidifier water, but not detected using active sampling, is provided as supplementary information, Table S3. Their K_H values varied between 10⁻⁶ and 10⁻¹¹ atm m³ mol¹¹.

4. Conclusions

The GC-EI-TOF-MS chromatograms obtained for the extracts of dehumidifier water samples contain a large number of semi-volatile compounds. Variations among sampling places, and the reasonable agreement of data obtained in the same room, with different dehumidifiers, confirm that condensed water samples are useful for qualitative analysis of the chemical composition of indoor areas. Dehumidifier appliances are portable, relatively non-expensive and non-disrupting equipment compared to pumps involved in active sampling of indoor air. The concentration efficiency of indoor air compounds in dehumidifier water is correlated with the values of their Henry constants, being effective for species with values below 10⁻⁴ atm m³ mol⁻¹ at 25 °C. Compounds identified in this study (including tentative identifications) represent around 25% of deconvoluted peaks in the chromatograms of dehumidifier water

samples; therefore, construction of larger EI-MS spectral databases is required to understand the complexity of the mixture of semi-volatile species existing in indoor areas.

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Captions to figures:

- Fig. 1. Deconvoluted compound chromatogram (A), accurate EI-MS spectra of selected peaks (B and C) and PCDL spectrum for a standard of benzaldehyde (D).
- Fig. 2. Response ratios for compounds in condensed water samples obtained from the same room after and before installing an air freshener. A and B correspond to a selection of compounds not detected and detected in the freshener liquid, respectively.
- Fig. 3. A, Chromatograms for condensed water samples obtained in the same room with two different dehumidifiers operating simultaneously. B and C, correlation plots of logarithmic responses for identified compounds in sampling places S21 and S3 using different appliances.
- Fig. 4. Black, logarithmic values of response ratios for compounds identified in condensed water samples and in active sampling extracts (concentrated air volume 5 m³) simultaneously obtained in the same room. Average data obtained in four different places. Red, negative logarithmic values of Henry's law constant (K_H, atm m³ mol⁻¹) estimated using the bond method (values at 25 °C).

References

Araki, A., Bastiaensen, M., Bamai, Y.A., Van den Eede, N., Kawai, T., Tsuboi, R., Ketema, R.M., Covaci, A., Kishi, R., 2018. Associations between allergic symptoms and phosphate flame retardants in dust and their urinary metabolites among school children. Environ. Int. 119, 438-446. https://doi.org/10.1016/j.envint.2018.07.018

Castro, G., Rodríguez, I., Ramil, M., Cela, R., 2019. Assessment of gas chromatography time-of-flight mass spectrometry for the screening of semi-volatile compounds in indoor dust. Sci. Total Environ. 688, 162-173. https://doi.org/10.1016/j.jhazmat.2015.01.043

Cecinato, A., Romagnoli, P., Perilli, M., Balducci, C., 2017. Psychotropric substances in house dusts: a preliminary assessment. Environ. Sci. Pollut. Res. 24, 21256-21261. https://doi.org/10.1007/s11356-017-9549-5

Esteve-Turrillas, F.A., Ly-Verdú, S., Pastor, A., de la Guardia, M., 2009. Development of a versatile, easy and rapid atmospheric monitor for benzene, toluene, ethylbenzene and xylenes determination in air. J. Chromatogr. A 1216, 8549-8559. https://doi.org/10.1016/j.chroma.2009.10.001

Fromme, H., Lahrz, T., Piloty, M., Gebhart, H., Oddoy, A., Rueden, H., 2004. Occurrence of phthalates and musk fragrances in indoor air and dust from apartments and kindergartens in Berlin (Germany). Indoor Air 14, 188-195. https://doi.org/10.1111/j.1600-0668.2004.00223.x

Gale, R.W., Cranor, W.L., Alvarez, D.A., Huckins, J.N., Petty, J.D., Robertson, G.L., 2009. Semi-volatile organic compounds in residential air along the Arizona-Mexico Border, Environ. Sci. Technol. 43, 3054-3060. https://doi.org/10.1021/es803482u

Geens, T., Roosens, L., Neels, H., Covaci, A., 2009. Assessment of human exposure to Bisphenol-A, Triclosan and Tetrabromobisphenol-A through indoor dust intake in Belgium. Chemosphere 76, 755-760. https://doi.org/10.1016/j.chemosphere.2009.05.024

Glowacz-Czerwonka, D., 2017. The polyurethane foams based on melamine-formaldehyde-cyclohexanone. Cell. Polym. 36, 35-50. https://doi.org/10.1177/026248931703600103 Gómez-Ramos, M.M., Ucles, S., Ferrer, C., Fernández-Alba, A.R., Hernando, M.D., 2019. Exploration of environmental contaminants in honeybees using GC-TOF-MS and GC-Orbitrap-MS. Sci. Total Environ. 647, 232-244.

https://doi.org/10.1016/j.scitotenv.2018.08.009

Hoang, T., Castorina, R., Gaspar, F., Maddalena, R., Jenkins, P.L., Zhang, Q., McKone, T.E., Benfenati, E., Shi, A.Y., Bradman, A., 2017. VOC exposures in California early childhood education environments. Indoor Air 27, 609-621. https://doi.org/10.1111/ina.12340

Laboire, S., Moreau-Guigon, E., Alliot, F., Desportes, A., Oziol, L., Chevreuil, M., 2016. A new analytical protocol for the determination of 62 endocrine-disrupting compounds in indoor air. Talanta 147, 132-141. https://doi.org/10.1016/j.talanta.2015.09.028

Lamas, J.P., Sánchez-Prado, L., García-Jares, C., Llompart, M., 2010. Determination of fragrance allergens in indoor air by active sampling followed by ultrasound-assisted solvent extraction and gas chromatography-mass spectrometry. J. Chromatogr. A, 1217, 1882-1890. https://doi.org/10.1016/j.chroma.2010.01.055

Liu, S., Thompson, S.L., Stark, H., Ziemann, P.J., Jimenez, J.L., 2017. Gas-phase carboxylic acids in a university classroom: abundance, variability and sources. Environ. Sci. Technol. 51, 5454-5463. https://doi.org/10.1021/acs.est.7b01358

Lucattini, L., Poma, G., Covaci, A., de Boer, J., Lamoree, J.H., Leonards, P.E.G., 2018. A review of semi-volatile organic compounds (SVOCs) in the indoor environment: occurrence in consumer products. Chemosphere 201, 466-482. https://doi.org/10.1016/j.chemosphere.2018.02.161

Melymuk, L., Bohlin-Nizzetto, P., Vojta, Š., Krátká M., Kukučka, P., Audy, O., Přibylová, P., Klánová, J., 2016. Distribution of legacy and emerging semi-volatile organic compounds in five indoor matrices in a residential environment. Chemosphere 153, 179-186. https://doi.org/10.1016/j.chemosphere.2016.03.012

Moniruzzaman, M., Rodríguez, I., Ramil, M., Cela, R., Sulaiman, S.A., Gan, S.H., 2014. Assessment of gas Chromatography time-of-flight accurate mass spectrometry for identification of volatile and semi-volatile compounds in honey. Talanta 129, 505-515. https://doi.org/10.1016/j.talanta.2014.06.019

Palmiotto, G., Pieraccini, G., Moneti, G., Dolara, P., 2001. Determination of the levels of aromatic amines in indoor and outdoor air in Italy. Chemosphere 43, 355-361. https://doi.org/10.1016/S0045-6535(00)00109-0

Pandey, S.K., Kim, K.H., 2010. A review of environmental tobacco smoke and its determination in air. Trends Anal. Chem. 29, 804-819. https://doi.org/10.1016/j.trac.2010.04.014

Ramírez, N., Marcé, R.M., Borrull, F., 2010. Development of a thermal desorption-gas Chromatography-mass spectrometry method for determining personal care products in air. J. Chromatogr. A 1217, 4430-4438. https://doi.org/10.1016/j.chroma.2010.04.049

Roll, I.B., Pycke, B.F.G., Halden, R.U., 2015. Indoor air condensate as a novel matrix for monitoring inhalable organic contaminants. J. Hazard. Mat. 288, 89-96. https://doi.org/10.1016/j.jhazmat.2015.01.043

Shibuta, S., Imasaka, T., Imasaka, T., 2016. Determination of fragrance allergens by ultraviolet femtosecond laser ionization mass spectrometry. Anal. Chem. 88, 10693-10700. https://doi.org/10.1021/acs.analchem.6b03229

Soheib, M., Harner, T., 2002. Characterization and comparison of three passive air samplers for persistent organic compounds. Environ. Sci. Technol. 36, 4142-4151. https://doi.org/10.1021/es020635t

Steinemann, A., Goodman, N., 2019. Fragranced consumer products and effects on asthmatics: an international population-based study. Air Qual. Atmos. Health 12, 643-649. https://doi.org/10.1007/s11869-019-00693-w

Stupak, M., Goodall, I., Tomaniova, M., Pulkrabova, J., Hajslova, J., 2018. A novel approach to assess the quality and authenticity of Scotch whisky based on gas chromatography coupled to high resolution mass spectrometry. Anal. Chim. Acta, 1042, 60-70. https://doi.org/10.1016/j.aca.2018.09.017

Ueta, I., Takenaka, R., Fujimura, K., Narukami, S., Sasaki, T., Maeda, T., 2019. Quantitative determination of 2-ethyl-1-hexanol, Texanol and TXIB in in-door air using a solid-phase extraction-type collection device followed by gas chromatography-mass spectrometry. Anal. Sci. 35, 855-859. https://doi.org/10.2116/analsci.19P033

Wakayama, T., Ito, Y., Sakai, K., Miyake, M., Shibata, E., Ohno, H., Kamijima, M., 2019. Comprehensive review of 2-ethyl-1-hexanol as an indoor air pollutant. J. Occup. Health 61, 19-35. https://doi.org/10.1002/1348-9585.12017

Weschler, C. J., Nazaroff, W.W., 2014. Dermal uptake of organic vapors commonly found in indoor air. Environ. Sci. Technol. 48, 1230-1237. https://doi.org/10.1021/es405490a

Zhang, J., Zhang, X., Wu, L., Wang, T., Zhao, J., Zhang, Y., Men, Z., Mao, H., 2018. Occurrence of benzothiazole and its derivates in tire wear, road dust, and roadside soil, Chemosphere 201, 310-317. https://doi.org/10.1016/j.chemosphere.2018.03.007

Table 1. Codes of dehumidifier water samples, indicating the sampling place and the identities of the three most intense peaks in the GC-EI-TOF-MS chromatograms for their SPE extracts.

Sample code	Sampling Place	Room	1 st most intense peak	2 nd most intense peak	3 rd most intense peak
S1	Home 1	Laundry	Diethylene glycol butyl ether	Texanol	TXIB
S2	Home 1	Bedroom	Diethylene glycol butyl ether	Texanol	TXIB
S3	Home 1	Living room	Diethylene glycol butyl ether	Texanol	TXIB
S4	Home 1	Kitchen	Alfa-Terpineol	Diethylene glycol butyl ether	Texanol
S5	Home 2	Entrance	Nonanoic acid	Diisobutyl phthalate	2-Phenoxyethanol
S6	Home 3	Living room	Diisobutyl phthalate	Diethyl phthalate	Nonanoic acid Diethylene glycol butyl
S7	Home 4	Living room	TCPP	Nonanoic acid	ether
S8	Home 4	Office	Diethylene glycol butyl ether	TCPP	Hexanoic acid Diethylene glycol butyl
S9	Home 5	Living room	2-Phenoxyethanol	Diethyl phthalate	ether
S10	Home 6	Living room	Nicotine	Methyl Dihydrojasmonate 4,7-Methano-1H-indenol,	Dodecanoic acid
S11	Home 7	Laundry	Linalool	hexahydro-	2-Phenoxyethanol
S12	Shopping centre	Hair dressing	2-Phenoxyethanol	Methyl Dihydrojasmonate	Dodecanoic acid
S13	Home 8	Living room	^a Glycol	^a Glycol	Nonanoic acid
S14	Administrative building 1	Reception area	Texanol	TXIB	Diethyl phthalate
S15	Administrative building 2	Office	Diethyl phthalate	Nonanoic acid	Diisobutyl phthtalate
S16	Administrative building 1	Office	Benzyl alcohol	2-Phenoxyethanol	Texanol
S17	Home 1	Garage	Benzothiazole	TXIB	Texanol
S18	Home 9	Garage	Benzothiazole	2-phenoxyl ethanol	^a Glycol
S19	Home 10	Living room Sample peparation	2-Phenoxyethanol	Methyl Dihydrojasmonate	Diethyl phthalate
S20	Research building	laboratory	Nonanoic acid	Diisobutyl phthalate	Diethyl phthalate
S21	Home 11	Living room	Texanol	Diethyl phthalate	Nonanoic acid

^aThe exact congener of the glycol series was not identified.

Table 2 A. Fragrances and flavoring compounds identified in dehumidifier water samples.

0	Positive	0.4.0		0 I	Positive	0.4.0	
Compound	samples	CAS number	LRI	Compound	samples	CAS number	LRI
Limonene	4	5989-27-5	1219	Beta-Ionone	5	14901-07-6	1972
Eucalyptol	13	470-82-6	1228	1-Dodecanol	5	112-53-8	1981
Octanal	8	124-13-0	1308	Methyleugenol	7	93-15-2	2044
3-Octanol	5	589-98-0	1406	Ethyl maltol	2	4940-11-8	2047
Tetrahydrolinalool	16	78-69-3	1442	Benzaldehyde, 4-methoxy	18	123-11-5	2065
2-Octanol, 2,6-dimethyl-	4	18479-57-7	1456	4-Ethylguaiacol	5	2785-89-9	2073
1-Heptanol	19	111-70-6	1471	Lilial	10	80-54-6	2080
Dihydromyrcenol	19	18479-58-8	1479	Cinnamal	2	104-55-2	2083
Camphor	16	76-22-2	1549	Eugenol	13	97-53-0	220
Linalool	17	78-70-6	1564	Thymol	13	89-83-8	2213
Linalyl acetate	2	115-95-7	1573	Naphthalene, 2-methoxy	13	93-04-9	223
1-Octanol	17	111-87-5	1574	Carvacrol	3	499-75-2	224
Bornyl acetate	6	76-49-3	1604	Hexyl salicylate	7	6259-76-3	224
Menthol	19	15356-70-4	1658	Naphthalene, 2-ethoxy	3	93-18-5	226
Alfa-terpineol	20	98-55-5	1719	Piperonal	15	120-57-0	227
Borneol	19	507-70-0	1726	Methyl anthranilate	17	134-20-3	228
Verbenone	21	1196-01-6	1739	Methyl Dihydrojasmonate	18	24851-98-7	231
				Galaxolide (two			2358
Citronellol	11	106-22-9	1782	diastereomers)	9	1222-05-5	236
Methylsalicylate	8	119-36-8	1815	Alfa-Hexylcinnamaldehyde	3	101-86-0	239
Geraniol	9	106-24-1	1869	Tonalid	5	21145-77-7	244
Alpha-isomethylionone	4	127-51-5	1877	Coumarin	21	91-64-5	2519
Benzyl alcohol	20	100-51-6	1912	Ethyl Vanillin	18	121-32-4	2570
2-Phenylethanol	21	60-12-8	1941	2-Acetonaphthone	20	93-08-3	260°
7-Dimethyl-7-hydroxyoctanal	12	107-75-5	1969				260°

Table 2 B. Summary of semi-volatile compounds identified in dehumidifier water samples, including their most relevant application.

Compound	Positive samples	CAS number	LRI	Application
Nicotine	8	54-11-5	1896	Alkaloids
Quinoline	19	91-22-5	1976	
Nonanoic acid*	18	112-05-0	2188	Carboxylic acid
Benzothiazole	20	95-16-9	1996	Curing agent
Tributyl phosphate	12	126-73-8	2141	
TCPP	17	13674-84-5	2735	Flame retardants
Tris(2-chloroethyl) phosphate	6	115-96-8	3026	
Irgacure 184	16	947-19-3	2684	Paint additives
Irgacure 651	18	24650-42-8	2794	
Diazinone	2	333-41-5	2463	Pesticides
Terbutryn	3	886-50-0	2992	
Ethyl benzene	10	100-41-4	1147	
p-Xylene	16	106-42-3	1155	Petrol components
m-Xylene	8	108-38-3	1163	
o-Xylene	10	95-47-6	1210	
Ibuprofen	4	15687-27-1	2863	Pharmaceutical
2,2,4-Trimethyl-1,3-pentanediol diisobutyrate (TXIB)	11	6846-50-0	1892	
1,2-Diacetin*	21	102-62-5	2096	
Dimethyl phthalate	19	131-11-3	2341	
Diethyl phthalate*	20	84-66-2	2409	Plasticizers
Triethyl citrate	20	77-93-0	2496	
Diisobutyl phthalate*	21	84-69-5	2577	
Dibutyl phthalate*	19	84-74-2	2735	
Tributyl acetylcitrate	13	77-90-7	2933	
Benzaldehyde* 5-Chloro-2-methyl-3(2H)-	21	100-52-7	1555	
isothiazolone	1	26172-55-4	2004	
2-Phenoxyethanol	19	122-99-6	2178	
Benzoic acid	6	65-85-0	2475	Preservatives
o-Hydroxybiphenyl	17	90-43-7	2658	
Benzyl Benzoate	9	120-51-4	2693	
3(2H)-Isothiazolone, 2-octyl-	8	26530-20-1	2784	
Methylparaben	10	99-76-3	3000	

^{*}Compounds also noticed in procedural blanks

Table 2 B, cont. Summary of semi-volatile compounds identified in dehumidifier water samples, including their most significant application.

Compound	Positive samples	CAS number	LRI	Application
Cyclohexanone	17	108-94-1	1321	Application
-	20	111-76-2	1419	
Ethylene glycol butyl ether				
Furfural	6	98-01-1	1499	
2-Ethyl-1-hexanol*	21	104-76-7	1502	
Isophorone	18	78-59-1	1623	
Diethylene Glycol ethyl ether	7	111-90-0	1638	
1-Decanol	3	112-30-1	1775	
Aniline	9	62-53-3	1787	
Diethylene glycol monobutyl ether	14	112-34-5	1814	Solvents/ intermediates
2-Nitrophenol	3	88-75-5	1850	Solvents/ Intermediates
Phenol*	10	108-95-2	2039	
2,3-Dimethylphenol	3	526-75-0	2112	
p-Cresol	15	106-44-5	2119	
4-Ethylphenol	18	123-07-9	2208	
Epsilon-Caprolactam	1	105-60-2	2225	
4-chloro-3-methylphenol	19	59-50-7	2548	
Benzoic acid, p-tert-butyl-	20	98-73-7	2794	
Phthalimide Phthalimide	13	85-41-6	2985	
Dibromochloromethane	4	124-48-1	1322	Trihalomethane

^{*}Compounds also noticed in procedural blanks

Table 3. Compounds tentatively identified in dehumidifier water samples, including the accurate m/z values of two intense ions with mass errors.

Compound	Formula	CAS number	Experimental LRI	Literature LRI	Chemical class	Positive samples	1st ion	2nd ion	Mass error (mDa) 1 st ion	Mass error (mDa) 2 nd ion
Pyridine, 3-ethenyl	C ₇ H ₇ N	1121-55-7	1501		Alkaloids	5	105.0575	78.0459	0.3	-0.4
Nicotyrine	C ₁₀ H ₁₀ N ₂	487-19-4	2368			5	158.0844	130.0657	0.4	0.2
2-Butene, 2-methyl-	C ₅ H ₁₀	513-35-9	1259			6	70.078	55.0542	0.3	-0.1
Cyclobutane, ethyl-	C ₆ H ₁₂	4806-61-5	1363		Alkanes	10	69.0701	56.0625	0.3	0.4
Isopropylcyclobutane	C7H14	872-56-0	1467		Alkalies	17	83.0853	70.0775	-0.3	-0.3
Cyclopentane, 1,1-dimethyl-	C7H14	1638-26-2	1570			8	83.0852	69.0699	-0.4	0.1
Benzene, (1-methoxypropyl)-	C ₁₀ H ₁₄ O	59588-12-4	1783			4	121.0646	91.0543	-0.2	0.1
1-phenyl ethanol	$C_8H_{10}O$	98-85-1	1838	1820a		5	122.0724	107.0488	-0.3	-0.4
1-(4-tert-Butylphenyl)propan-2-one	$C_{13}H_{18}O$	81561-77-5	2047			5	190.1373	175.1137	1.5	1.4
1-Phenoxypropan-2-ol	$C_9H_{12}O_2$	770-35-4	2073		Dannana dani sati sa	8	152.0836	94.0409	0.4	-0.4
N-methyl phthalimide	$C_9H_7NO_2$	550-44-7	2328		Benzene derivatives	4	161.0495	132.0474	1.8	2.5
Benzene, (3-methylcyclopentyl)-	$C_{12}H_{16}$	5078-75-1	2345			6	160.1257	131.0861	1.0	0.6
Dibenzyl ether	$C_{14}H_{14}O$	103-50-4	2423			13	92.0617	91.0546	-0.3	0.3
1,2-Ethanediol, monobenzoate	$C_9H_{10}O_3$	94-33-7	2554			6	123.0449	105.0344	0.8	0.9
Diphenyl methanol	C ₁₃ H ₁₂ O	91-01-0	2803			4	184.0904	105.0355	1.6	1.5
2-(Methylmercapto)benzothiazole	C ₈ H ₇ NS ₂	615-22-5	2484	2422ª	Benzothiazole derivative	4	181.0012	148.0217	-0.3	0.2
Hexanoic acid	C ₆ H ₁₂ O ₂	142-62-1	1865	1851a	donitativo	 7	87.0437	73.0277	-0.4	-0.6
Heptanoic acid	C ₇ H ₁₄ O ₂	111-14-8	1972	1971 ^a		8	101.0596	87.0438	-0.2	-0.3
Octanoic acid	C ₈ H ₁₆ O ₂	124-07-2	2079	2086ª	Carboxylic acids	17	101.0592	87.0438	-0.6	-0.3
n-Decanoic acid	C ₁₀ H ₂₀ O ₂	334-48-5	2292	2275a		4	129.0911	115.0753	0.1	-0.1
Tetradecanoic acid	C ₁₄ H ₂₈ O ₂	544-63-8	2716	2716 ^a		6	185.154	129.0914	0.4	0.4

Table 3 cont. Compounds tentatively identified in dehumidifier water samples, including the accurate m/z values of two intense ions with mass errors.

Compound	Formula	CAS number	Experimental LRI	Literature LRI	Chemical class	Positive samples	1 st ion	2 nd ion	Mass error (mDa) 1st ion	Mass error (mDa) 2nd ion
Butanoic acid, ethyl ester	C ₆ H ₁₂ O ₂	105-54-4	1071	1041a		3	101.0615	88.0536	1.2	1.2
Butanedioic acid, dimethyl ester	C ₆ H ₁₀ O ₄	106-65-0	1616	1595a		10	115.0384	87.0441	-0.6	0.1
Hexanedioic acid, dimethyl ester	C ₈ H ₁₄ O ₄	627-93-0	1842	1819 ^a		10	143.0708	114.0675	0.5	-0.1
Hexanoic acid, 2-ethyl- 3-hydroxy-2,2,4-trimethylpentyl	C ₈ H ₁₆ O ₂	149-57-5	1968	1963ª	Esters	8	101.0594	88.0515	-0.4	-0.4
isobutyrate (Texanol)	C ₁₂ H ₂₄ O ₃	77-68-9 1000373-	1889, 1909			9	173.1168	89.0593	-0.5	-0.5
Isobutyl methyl phthalate	C ₁₃ H ₁₆ O ₄	89-3	2476			3	163.0391	149.0235	0.2	0.2
2(3H)-Furanone, 5-ethenyldihydro- 5-methyl-	C ₇ H ₁₀ O ₂	1073-11-6	1703	1689ª	Furane	4	111.0431	98.0363	-0.9	0.1
2-Propanol, 1-butoxy- 1-Propanol, 2-(2-	C7H16O2	5131-66-8	1354	1364ª		14	87.0805	57.0696	0.1	-0.3
methoxypropoxy)- 1,3-Pentanediol, 2,2,4-trimethyl-	C7H16O3	13588-28-8	1584		Glycols	3	117.0916	73.0644	0.6	-0.4
isobutyrate Ethanol, 2-[2-(2-	C ₈ H ₁₈ O ₂	144-19-4	1947		•	6	85.0646	73.0644	-0.2	-0.4
butoxyethoxy)ethoxy]-	C ₁₀ H ₂₂ O ₄	143-22-6	2217			4	85.0649	57.0697	0.2	-0.2
1H-Inden-1-one, 2,3-dihydro-	C ₉ H ₈ O	83-33-0	2055		Indene	17	132.0583	104.0626	1.3	0.5
Indole	C ₈ H ₇ N	120-72-9	2496	2445a	Indole	2	117.0585	90.0471	1.2	0.7
Tri-ethyl phosphate	C ₆ H ₁₅ O ₄ P	78-40-0	1675	1672a	Organophosphate	21	155.0469	127.0158	0.1	0.4
1H-Pyrrole-2-carboxaldehyde	C ₅ H ₅ NO	1003-29-8	2062	2044a	Pyrrole derivative	3	95.0374	94.029	0.8	0.3
cis-Furanoid linalool oxide	C ₁₀ H ₁₈ O ₂	5989-33-3	1460	1462 ^b		7	111.0799	94.0774	-0.5	-0.4
trans-Furanoid linalool oxide 3-Cyclohexen-1-ol, 1-methyl-4-(1-	C ₁₀ H ₁₈ O ₂	34995-77-2	1489	1483ª	Terpenoids	5	111.0803	94.0778	-0.2	0.1
methylethyl)-	C ₁₀ H ₁₈ O	586-82-3	1593	1589a	·	9	136.1244	121.1012	-0.3	0.1
LalphaTerpineol	C ₁₀ H ₁₈ O	10482-56-1	1718	1722 ^b		14	136.1247	121.1014	0.1	0.3

Cyclohexanol, 4-(1,1-									
dimethylethyl)-, trans-	$C_{10}H_{20}O$	21862-63-5	1766		6	123.1187	81.071	1.3	0.6
4,7-Methano-1H-indenol,									
hexahydro-	$C_{10}H_{14}O$	37275-49-3	2009		11	132.0939	117.0704	0.5	0.5
2,6-Di-tert-butyl-4-hydroxy-4-									
methylcyclohexa-2,5-dien-1-one	$C_{15}H_{24}O_2$	10396-80-2	2118	2049 ^a	8	221.1538	180.114	0.2	-0.5

^aLRI values obtained from ChemSpider database (http://www.chemspider.com/) for Carbowax-type GC columns.

^bLRI obtained from Moniruzzaman et al., 2014.

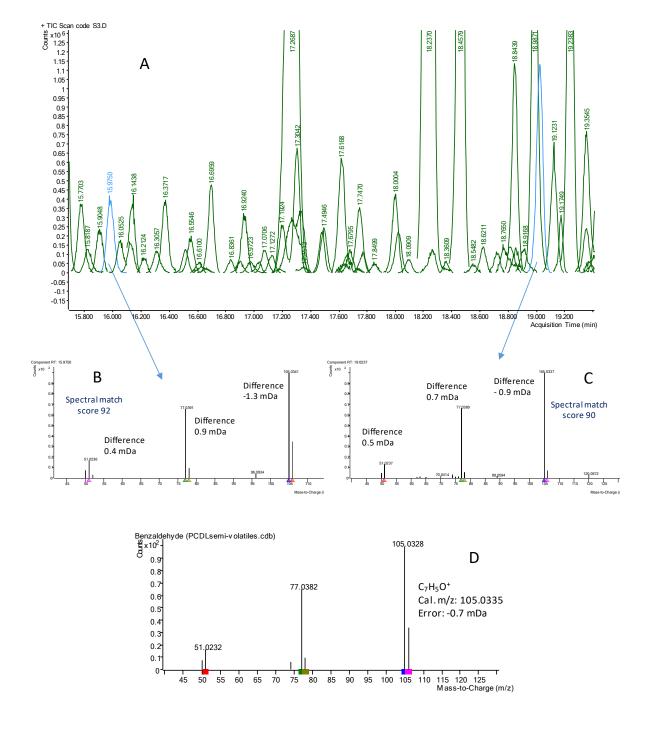
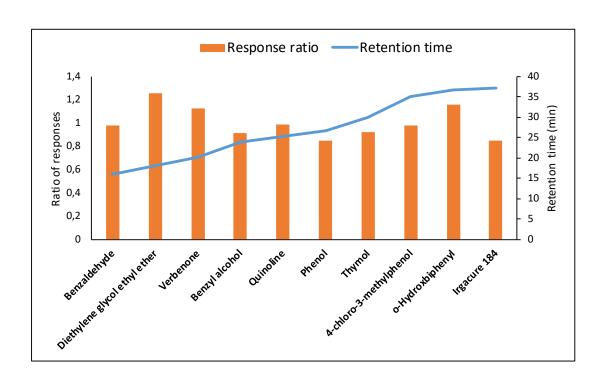


Fig. 1



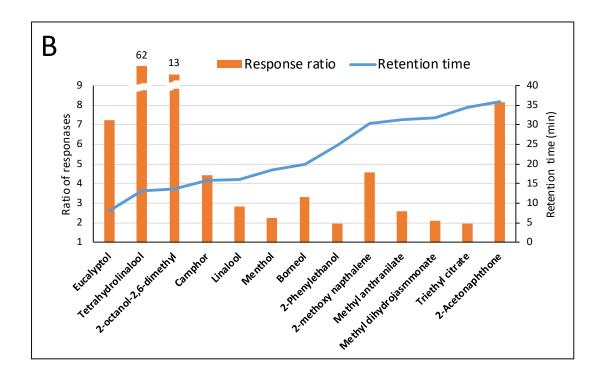
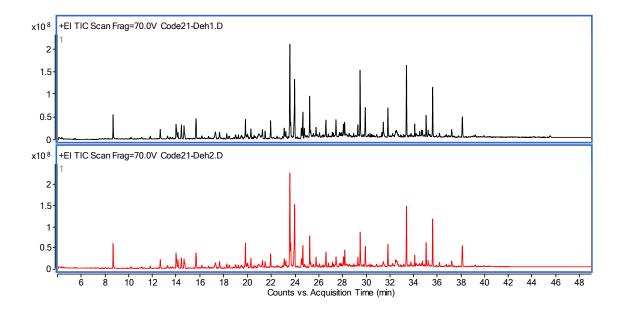
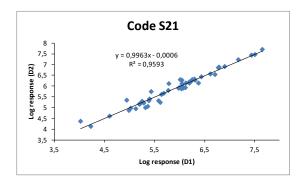


Fig. 2





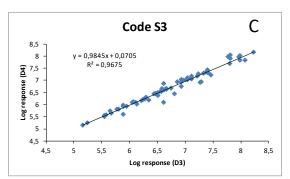


Fig. 3

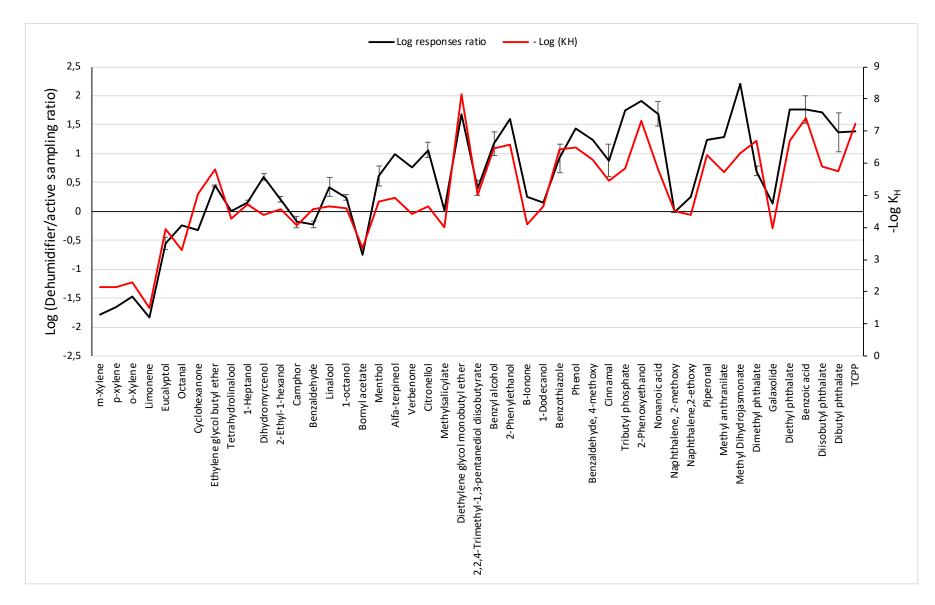


Fig. 4

Supplementary information to manuscript:

Portable dehumidifiers condensed water: a novel matrix for the screening of semi-volatile compounds in indoor air

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Table S1. Data corresponding to sampling places, temperature and relative humidity before dehumidifier operation.

Sample code	Dehumidifier system	Room volume (m³)	Temperature (ºC)	Relative air humidity (%)	Volume of the condensed water sample (mL)
S1	1	46	21	72	650
S2	1	55	19	68	680
S3	1	90	19	68	700
S4	1	30	21	68	520
S 5	1	50	18	65	800
S6	1	33	18	72	700
S 7	1	53	21	62	1000
S8	1	25	22	65	850
S9	1	55	19	81	1000
S10	1	65	20	84	1000
S11	2	25	20	80	900
S12	1	105	22	56	400
S13	1	75	22	70	600
S14	1	35	20	75	500
S15	1	60	23	75	1900
S16	1	70	19	65	600
S17	1	150	16	75	1200
S18	3	150	17	75	1500
S19	4	35	19	80	600
S20	1	400	21	62	2000
S21	5	35	22	85	800

Table S2. Summary of compounds, including their retention times and a representative ion, used to investigate potential breakthrough problems during active sampling of indoor air in the HLB cartridge.

Compound	³ Potentian time (min)	Representative ion
Compound	^a Retention time (min)	(m/z)
Limonene	7.81	93.0701
Eucalyptol	8.01	154.1354
Gamma-Terpinene	8.85	136.1252
Cymene	9.51	119.0847
Camphor	15.77	152.1195
Linalool	16.14	121.1017
Borneol acetate	17.17	136.1247
Neral	19.57	137.096
Alfa-Terpineol	19.84	121.1007
Verbenone	20.3	135.0801
Citral	20.71	137.096
Citronellol	21.26	95.0854
Geraniol	23.07	123.1176
Eugenol	29.78	164.0835
Thymol	29.77	135.0799
Carvacrol	30.54	135.0797

^aValues corresponding to the DB-WAXETR column.

Table S3. Summary of compounds identified in condensed water extracts, and not detected after active sampling of 5 m³ of indoor air. Henry's law constant values (K_H) predicted using the bond method (http://www.chemspider.com/).

Compound	K _H (atm m ³ mol ⁻¹ , 25 °C)
1,2-Diacetin	2.60e-11
2-Acetonaphthone	9.58e-7
4-chloro-3-methylphenol	4.58e-7
4-ethylphenol	8.21e-7
Coumarin	6.95e-6
Ethyl Vanillin	1.1e-10
Irgacure 184	2.78e-6
Isophorone	^a 6.64e-6
Phthalimide	1.02e-8
Quinoline	^a 1.67e-6
Tributyl acetylcitrate	3.78e-10
Triethyl citrate	^a 3.84e-9
Vanillin	°2.15e-9

^aExperimental values.

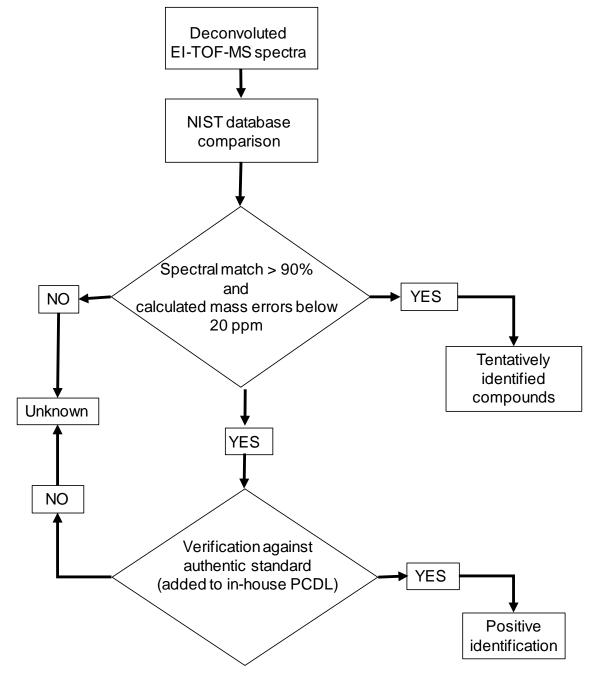


Fig. S1. Scheme of the workflow employed for non-target data mining of compounds existing in indoor air related samples.

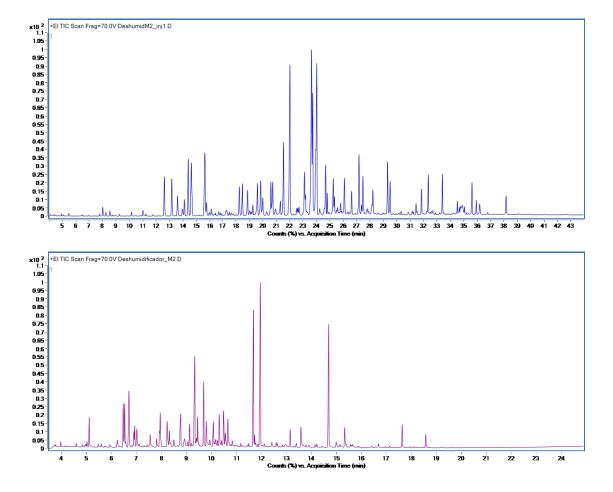
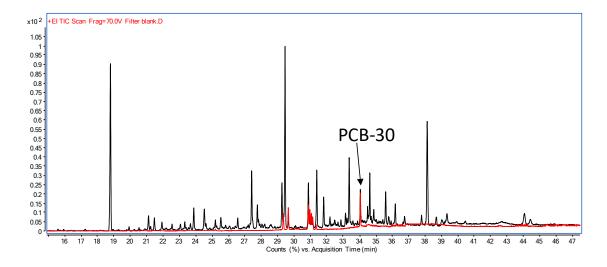


Fig. S2. GC-EI-TOF-MS chromatograms for the same SPE extract from dehumidifier water injected in the DB-WAXETR (A) and the BP-5 (B) column. DBP, dibutyl phthalate.



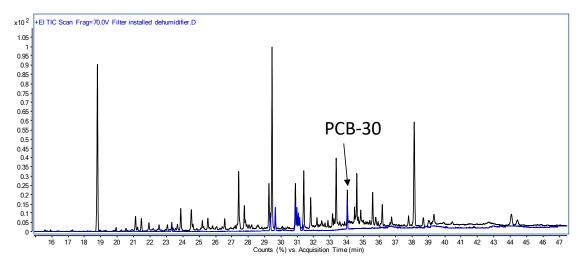
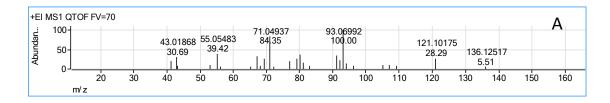
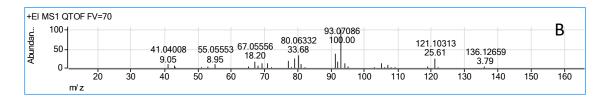
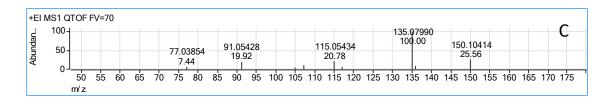


Fig. S3. Chromatograms obtained for condensed dehumidifier water (black), a blank of glass fiber filter (red), and a glass fiber filter placed before the coarse grid of the dehumidifier (blue) while collected the condensed water sample. Peaks in glass fiber filter extracts were tentatively identified as diisopropyl naphthalene isomers.







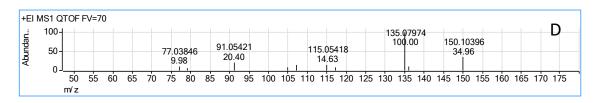


Fig. S4. Accurate EI-MS spectra for pairs of related species. A and B, Linalool ($C_{10}H_{18}O$) and Linalyl acetate ($C_{12}H_{20}O_2$). C and D, Thymol ($C_{10}H_{14}O$) and Carvacrol ($C_{10}H_{14}O$). Figures under m/z values correspond to the relative intensity of each ion.

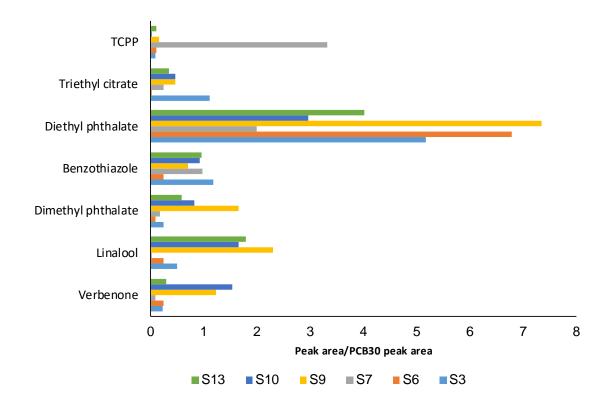


Fig. S5. Corrected responses (deconvoluted compound peak area/PCB 30 peak area) for selected compounds in the living room of different homes. Data corresponding to water samples obtained with the same dehumidifier.

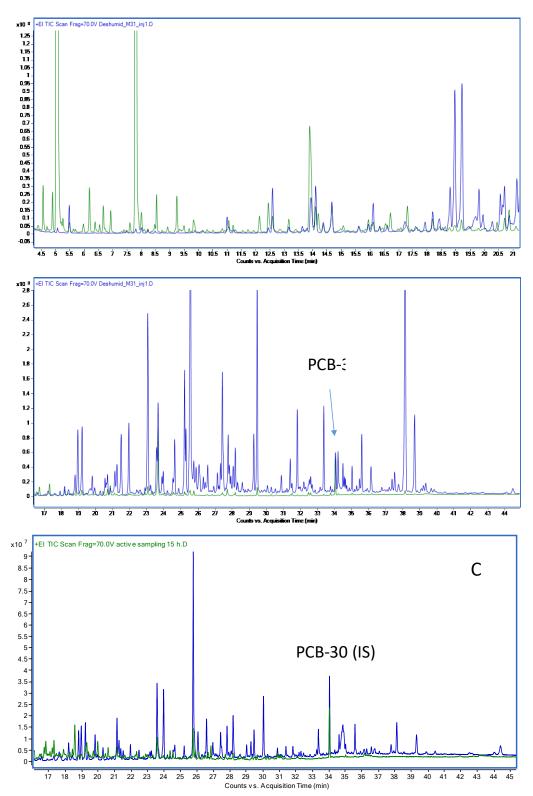


Fig. S6. GC-EI-TOF-MS chromatograms corresponding to active sampling (green) for 5 h at 16.5 L min⁻¹ and the SPE extract from the dehumidifier water sample (blue) simultaneously obtained from the same room. A, retention time interval: 4-21 min. B, retention time range: 16-45 min. C, chromatograms after 15 h of active sampling (green) versus analysis of dehumidifier water (blue) obtained in a different room. Pairs of chromatograms are shown in the same scale.