



# Linking environmental sustainability and nutritional quality of the Atlantic diet recommendations and real consumption habits in Galicia (NW Spain)

Xavier Esteve-Llorens, María Teresa Moreira, Gumersindo Feijoo, Sara González-García

**Accepted Mansucript** 

# How to cite:

Esteve-Llorens, X., Moreira, M., Feijoo, G., & González-García, S. (2019). Linking environmental sustainability and nutritional quality of the Atlantic diet recommendations and real consumption habits in Galicia (NW Spain). Science Of The Total Environment, 683, 71-79. doi: 10.1016/j.scitotenv.2019.05.20

# **Copyright information:**

© 2019 Elsevier Ltd. This manuscript version is made available under the CC-BY-NC-ND 4.0 license (http://creativecommons.org/licenses/by-nc-nd/4.0/)

1 2

# Linking environmental sustainability and nutritional quality of the Atlantic diet recommendations and real consumption habits in Galicia (NW Spain)

Xavier Esteve-Llorens\*, Maria Teresa Moreira, Gumersindo Feijoo and Sara González-García
 Department of Chemical Engineering, School of Engineering, Universidade de Santiago de
 Compostela. 15782- Santiago de Compostela, Spain.

\* Corresponding author: E-mail address: xavier.esteve.llorens@usc.es

6

7

### 8 Abstract

9 Today's society is increasingly aware of food consumption patterns. Under the perspective that 10 real consumption trends are often not in line with healthy recommendations, this research 11 focuses on the study of the environmental and nutritional sustainability of two types of food 12 consumption habits present in the northern Atlantic area of Spain (Galicia). The main objective 13 is, therefore to detect the existing deviations between the current Galician diet (GD) and the 14 traditional and increasingly relevant Recommended Atlantic Diet (RAD), allowing verifying 15 whether current consumption patterns ensure an optimal and sustainable nutritional profile.. In 16 this sense, the carbon footprint from a Life Cycle Assessment perspective has been estimated 17 as environmental indicator of both dietary patterns and, the nutritional quality has been 18 determined by the Nutrient Rich Diet 9.3 index and the Health gain score. The carbon footprint 19 of both dietary models is moderately high compared to recommended diets such as the 20 Mediterranean one. Comparing the two scenarios, the associated greenhouse gas emissions are about 15% higher for GD than for RAD, mainly due to the higher intake of beef and dairy 21 22 products. On the other hand, nutritional quality is comparatively higher for RAD than for GD, 23 associated with higher consumption of vegetables and fruits. An additional objective of this work 24 has been to consider a sensitivity analysis to determine the effect of replacing beef with 25 alternative sources.

Having in mind this study, it can be concluded that the real consumption pattern in Galicia is far from the recommended one, with worse environmental and nutritional quality. The promotion of social awareness policies to guide consumers in the choice a healthier and more environmentally sustainable dietary pattern should be advisable for regional decision-makers as

well as for those who wish to promote adherence to the Atlantic diet in other regions andcountries.

- 33 Keywords: Atlantic diet; consumption patterns; GHG emissions; LCA; nutritional quality;
- 34 sustainability.
- 35

#### 36 **1. Introduction**

37 There is a crucial need to make a change in the present unsustainable trends in food 38 production and consumption. According to the literature (Garnett, 2011; Tukker and Jansen, 39 2006; Vringer et al., 2010), one third of human impact on climate change is related to the food 40 production chain and consumption patterns. For this reason, consideration of more 41 environmental friendly dietary patterns may be a good option to help climate change mitigation. 42 In order to understand the concept of sustainable diets, it is required to take into account the 43 definition of the Food and Agriculture Organization of the United Nations - FAO (FAO, 2010). 44 According to this organization, sustainable diets are those with low environmental impacts, that 45 contribute both to food and nutritional security as well as to healthy life for present and future 46 generations. In this regard, it is also advised food that comes from agricultural production 47 systems with low environmental impact, such as less limited dependence on food of animal 48 origin, short distance production and consumption networks, and minimal food processing and 49 refining, among others (FAO, 2010).

50 Food consumption trends have a significant influence on environmental impact (Garnett, 51 2011; Irz et al., 2016; Sonesson et al., 2005). Nevertheless, numerous studies have been 52 conducted in recent years to assess the environmental profile of the human diet (Castañé and 53 Antón, 2017; Coelho et al., 2016; Pernollet et al., 2017; van de Kamp et al., 2018) since dietary 54 habits have shifted from traditional to the so-called "Western diet" based on the intake of animal 55 based products in portions greater than recommended. As a result, the amount of GHG 56 associated with a dietary choice varies significantly depending on the different products that 57 makes it up (Carlsson-Kanyama and González, 2009; Committee on Climate Change, 2010; 58 Scarborough et al., 2014), and largely depends on the efficiency of the production chain. On the 59 road to healthier and more environmentally friendly dietary patterns (Van Dooren et al., 2014) 60 governments from countries such as Sweden (Livsmedelsverket, 2009), United Kingdom (Reddy et al., 2009), Germany (Gerlach et al., 2013) and Finland (Steering Group, 2010) have 61 62 strengthened their efforts to set up committees to advise society on more sustainable dietary 63 patterns. Among the recommended diets, the well-known Mediterranean diet, traditionally present in Mediterranean countries (Spain, Italy, Greece, Croatia, Maghreb, Cyprus and 64 65 Portugal), receives special attention. It is considered a healthy diet by global organizations such

as the World Health Organization (WHO, 2012) and FAO (FAO, 2010). The Mediterranean diet
is related to a low incidence of chronic diseases due mainly to the high intake of vegetables,
fruits and whole grains (Cencic and Chingwaru, 2010) and to the low intake of animal fats, with
the moderate use of olive oil as a source of healthy fatty acids (Charro et al., 2006).

70 It is interesting to note how countries outside the traditional area of the Mediterranean diet 71 have begun to promote the Mediterranean diet style (van Dooren and Aiking, 2016; Wilson et 72 al., 2013), as well as to create new dietary choices following that philosophy (e.g. the New 73 Nordic diet) to achieve healthier consumption patterns (De Boer et al., 2014; Donati et al., 2016; 74 Hoek et al., 2017; Saxe et al., 2012; van Dooren and Aiking, 2016). In line with the 75 Mediterranean style, the Atlantic diet (Tojo and Leis, 2009) represents a dietary pattern 76 traditionally associated with the northwest of the Iberian Peninsula including Galicia (Spain) and 77 northern Portugal. The Atlantic diet has been recently considered a world reference for a 78 healthy diet (Vaz Velho et al., 2016), as it maintains the basic characteristics of the 79 Mediterranean diet while promoting the intake of fresh local and seasonal products, avoiding 80 complex cooking methods. This diet also includes plant-based foods, seafood, legumes and 81 nuts, with moderate consumption of animal products (i.e. pork and beef, milk and dairy 82 products, and eggs), a significant intake of potatoes (higher than in Mediterranean countries), 83 preferential use of olive oil for seasoning and cooking, and high consumption of mineral water. 84 Therefore the optimal intake of polyunsaturated fatty acids, complex sugars, vitamins, fiber, 85 minerals and functional components is guaranteed (Tojo and Leis, 2009; Vaz Velho et al., 86 2016). In this sense, the consumption of a balanced diet as in the aforementioned diets is 87 beneficial for the prevention of numerous chronic diseases (Cencic and Chingwaru, 2010). For 88 instance, the consumption of fruits, vegetables, nuts and fish is related with the low incidence of 89 cancers, neurodegenerative and cardiovascular diseases and Type II diabetes (Bach-Faig et 90 al., 2011). Furthermore, consumption of fruits and vegetables is also related to protective effects against cell oxidation (Charanjit Kaur and Harish C. Kapoor, 2001; Wang et al., 1996). 91

92 Numerous studies available in the literature report that there are outstanding differences 93 between the dietary recommendations established by health administrations and actual food 94 consumption patterns (Sáez-Almendros et al., 2013). While the Atlantic diet recommends a high 95 consumption of fresh produce such as vegetables and fruits, data on actual consumption habits

96 indicate that this is not being met as it should be (Sáez-Almendros et al., 2013). Thus, there is a
97 significant gap between recommended dietary patterns and actual food consumption trends,
98 which implies an intensification of resources in the production chain. Current patterns of actual
99 consumption are associated with increased intake of processed food and other resource100 intensive products, such as those of animal origin or processed foods (Cencic and Chingwaru,
2010; Tilman and Clark, 2014; Tukker et al., 2011).

102 Therefore, the main goal of this study is to compare, from a sustainability perspective, the 103 recommendations of the traditional Atlantic diet with the real consumption trends, considering 104 Galicia as case study, as well as to provide an answer to the question whether current 105 consumption patterns ensure an optimal nutritional profile. Finally, the level of concurrence 106 between both dietary patterns was also determined by taking into account both the carbon 107 footprint (CF), from an LCA approach associated with food production, as well as the nutritional 108 quality. Regarding the latter, two different indexes have been proposed for analysis to improve 109 robustness and consistency of results: The Nutrient Rich Diet 9.3 score, which takes into 110 account the intake of certain valuable and harmful nutrients (Van Kernebeek et al., 2014) and 111 the health gain score, which follows a similar approach for food groups (Van Dooren et al., 112 2014). Furthermore, from a practical point of view, the study will allow to identify the weak spots 113 of the GD from both a nutritional and environmental point of view and will serve as a guide for 114 decision-makers to promote a consumption pattern in pursuit of the traditional diet.

115

116 **2. Materials and methods** 

117 The comparative assessment of sustainability between two different dietary patterns related 118 to the recommended Atlantic diet and the actual consumption pattern has been carried out by 119 estimating the CF as a representative environmental indicator, as well as by means of two 120 nutritional quality indexes. A detailed description of both perspectives is presented below.

121

122 2.1. Carbon footprint methodology

123 2.1.1. Description

124

125 In this study, the environmental sustainability of the two different dietary patterns in terms of 126 their CF (i.e., GHG emissions) has been determined from a Life Cycle Assessment (LCA) 127 approach, which systematically assesses the environmental burdens of each type of diet at all 128 stages of its life cycle (ISO 14040, 2006). The carbon footprint is selected as an environmental 129 indicator due to its great relevance and widespread use in related studies of dietary patterns 130 (Aleksandrowicz et al., 2016; Batlle-Bayer et al., 2019; González-García et al., 2018; Ritchie et al., 2018; Springmann et al., 2018). In this case, the CF has been estimated considering the 131 132 stages of production and transport to retailer, as detailed below.

133

#### 134 2.1.2. Functional unit

The selected functional unit to report the results corresponds to the daily amount of food eaten per person, that is, the individual daily diet. This functional unit allows the comparison between the scenarios proposed, as well as with other related studies available in the literature on environmental assessment of different types of daily diets (Castañé and Antón, 2017; Pernollet et al., 2017; Werner et al., 2014) regardless of daily energy intake (i.e. kcal per capita and day).

141

142 2.1.3. Scope of the dietary scenarios

143 The scope of the CF study for both scenarios considered a cradle-to-mouth perspective. Thus, 144 the systems analyzed included the stages of food production (i.e., production of the foodstuffs 145 included in each daily diet) and transport activities (i.e. the distribution of the products from the 146 factory, farm or port to the corresponding retailers and from retailers to households). Therefore, 147 storing at retailers and consumption stage at the households, which should include operations 148 such as food preparation at home, refrigeration and final waste disposal, were disregarded. The 149 rationale behind their exclusion from the scope of the study is that these consumer activities should have a similar impact in both dietary scenarios, considered for the same region (Heller et 150 151 al., 2013). Moreover, other studies (Berlin and Sund, 2010) established that the consumption 152 stage could contribute up to 10% of the total life cycle GHG emissions when estimated 153 considering the food consumed in a typical menu. However, with regard to the estimation of 154 GHG emissions from food cooking in or outside households, it should be necessary to have real

information on the menus and the cooking method (i.e., boiling, frying, baking, ...) considered information that is not available for the GD scenario. Thus and taking in mind the mentioned studies, the exclusion of the consumption stage (i.e., food preparation) from the analysis could be justified. In addition, the exclusion of the stages mentioned also allows the results of this study to be compared with other relevant ones available in the literature (Castañé and Antón, 2017; Risku Norja et al., 2009; Sáez-Almendros et al., 2013; Saxe et al., 2012; Van Dooren et al., 2014).

162 Food losses along the chain considered in both scenarios have been calculated based on 163 García-Herrero et al. (2018), especially regarding the distribution to retailer and consumption 164 stages (Garcia-Herrero et al., 2018). This estimation is based on the losses reported by FAO for 165 European countries (FAO, 2011). Bearing in mind that there is not detailed information on the 166 loss percentage for pre-cooked food - an important item in the current diet, the highest 167 percentage reported by García-Herrero et al. (2018) for processed food (5%) has been 168 assumed for this type of foodstuff as the worst case. With regard to the foodstuffs production 169 stage, information on losses has been included in the corresponding background processes due 170 to the consideration of the cradle-to-gate approach of the references consulted.

171 2.1.4. Description of dietary patterns

172 Galicia (NW Spain) has been historically characterized as the cradle of a wide selection of 173 high quality food products, appellation of origin and organic farming with prestige beyond its 174 borders (Xunta de Galicia, 2005). All these concepts are included within the Atlantic diet model, 175 fulfilling its basic characteristics such as abundance of seasonal, local and fresh products, high 176 intake of plant-based products and seafood, as well as a moderate intake of animal-origin 177 foodstuffs (Vaz Velho et al., 2016). Nevertheless, the current dietary choices of the Galician 178 region may not be at all in line with these recommendations and with traditional patterns, which 179 seem to vary in proportion and quantity of certain categories of foodstuffs. The spread of the 180 occidental culture and the globalization of food consumption and production are behind these 181 alternative choices; however, this trend is also observed in other dietary patterns such as the 182 Mediterranean diet (Da Silva et al., 2009).

183

184

#### 185 Recommended Atlantic Diet – RAD scenario

186 This scenario corresponds to the Atlantic diet recommendations defined by the Health 187 Department of the Xunta de Galicia (2013). The 7 daily-menus reported by Esteve-Llorens et al. 188 (2019) have been taken into consideration (See SM Table 2 to 8) (Esteve-Llorens et al., 2019). 189 This study includes 67 foodstuffs grouped into 11 different categories (i.e., fruits, vegetables, 190 legumes, grains, nuts, dairy products, eggs, meat, fish, sweets and oils/fats), all of which are 191 recommended ingredients in the Atlantic food pyramid (Tojo and Leis, 2009) as well as in the 192 traditional Galician gastronomy (Xunta de Galicia, 2013). In this scenario it has been assumed 193 that the recommended servings according to the Atlantic diet pyramid (Tojo and Leis, 2009) are 194 applied to the entire population even though they are aimed at the adult population. Therefore, 195 the average daily intake of each food group  $(g \cdot day^{-1})$  has been considered for evaluation to 196 facilitate the resulting comparison with the other scenarios proposed for analysis. Table 1 197 summarizes the daily intake of each food category per capita.

#### 198 Galician Diet – GD scenario

199 The second scenario considered for analysis is based on the actual consumption patterns 200 of the Galician diet. The available surveys from the Galician Ministry of Health (SERGAS, 2007) 201 have been analyzed to gather dietary information. The consulted study reports Galician eating 202 habits in 2007 (last year updated) and it is based on data from 3,148 participants, both urban 203 and rural residents. The nutritional analysis included 129 food-items according to the surveys 204 (SERGAS, 2007). As a result, in addition to the food categories indicated in the RAD scenario, 205 an additional group of industrially processed foods has been included in the GD scenario which 206 appears in the current consumption trends but is not included in the Atlantic dietary philosophy 207 due to its low nutritional quality. The surveys were based on a dietary plan of 24 hours, 208 conducted in two different seasons (Spring-Summer and Fall-Winter) to cover seasonal 209 differences in the intake of some foodstuffs (e.g. broccoli, asparagus, peach, fig); in addition, a 210 food consumption questionnaire was also carried out, supported by photographs of food 211 servings to calculate the size of the portions eaten. The reported global food-items intakes (See 212 SM Table 9) have allowed the estimation of the apparent food consumption per capita a whole 213 day (g.person<sup>-1</sup>.day<sup>-1</sup>) as shown in **Table 1**. Nevertheless, the aforementioned SERGAS survey 214 (2007) only provides an average figure of the food consumed per person per day. Thus

- 215 variations between the different individuals surveyed cannot be appreciated. It is therefore
- 216 important to take potential uncertainty into account when discussing the results.

**Table 1.** Daily amount (g) of each food category in the Recommended Atlantic Diet (RAD) and Galician Diet (GD) scenarios.

219

<b>F</b>		GD
Food category	(g·person <sup>-</sup> ' ·day <sup>-</sup> ')	(g·person <sup>-</sup> ·day <sup>-</sup> )
Fruits	1024	439
Vegetables	633	581
Legumes	29.3	20.1
Grains	291	319
Nuts	33	5.0
Dairy	419	472
Eggs	23.7	26.2
Meat	91.9	213
Seafood	195.7	182
Processed food	0.00	88.1
Sweets	11.7	15.5
Oil/fats	29.9	23.6
TOTAL	2753	2387

#### 220

221 2.1.5. Inventory data for carbon footprint estimation

222 After an extensive literature review, a total of 139 food products from 42 LCA studies (see 223 Supplementary material) have been included in the inventory data set to determine the CF 224 scores of both diet scenarios, all of which have been analyzed from a cradle-to-gate 225 perspective. In addition, these foodstuffs have been grouped into 12 representative food 226 categories: fruits, vegetables, legumes, grains, nuts, dairy, eggs, meat, fish, processed food, 227 sweets and oil/fats (Aquilera et al., 2015; Clune et al., 2017; González-García et al., 2013; 228 Gunady et al., 2012; Iribarren et al., 2011; Nielsen et al., 2013; Notarnicola et al., 2017; Noya et 229 al., 2017; Volpe et al., 2015), attending to the Atlantic diet pyramid and other relevant studies in 230 literature (Castañé and Antón, 2017). In order to apply a similar scope for both dietary 231 scenarios, some assumptions have been made. In this sense, most of the foodstuffs LCA 232 studies cover a cradle-to-gate perspective; however, when the CF of a food product includes 233 transportation from retailer to households, food preparation or waste disposal, stages that have 234 not been included in this study, the related GHG emissions have been discarded, by subtracting 235 the corresponding percentage, according to specific LCA study indications. In addition, due to 236 their minor contribution on the daily diets (Castañé and Antón, 2017; Van Kernebeek et al., 237 2014), food condiments, soft drinks, infusions, coffee and alcoholic beverages have been left 238 out of the scope of the study.

On the other hand, certain products have been assimilated to others with similar production process and/or comparable nutritional characteristics due to the lack of data for the estimation of their environmental profiles. This is the case of chard (assimilated as lettuce), curd (as yogurt), semi-cured and cured cheese (as Galician cheese), leek (as onion), nectarine (as peach) and clams, oysters and scallops (as mussels).

244 In terms of distribution, Euro 5 diesel freight lorries (>32 tons) have been chosen for 245 transport activities from the factory/farm gate to retailers for Spanish products. Thus, distribution 246 distances of 60 km and 400 km (on average) have been set for the foodstuffs supply from inside 247 and outside Galicia, respectively, for all products included in the study. In this sense, it is 248 assumed that the vast majority of the products are manufactured in Spanish territory, except 249 certain foodstuffs that are imported such as pineapple, coffee, cod or salmon. In these cases, 250 an average distance by ship and lorry from their country of origin to Galicia has been estimated. 251 Regarding the transport from retailers to households, assumptions from Esteve-Llorens et al., 252 (2019) have been considered. Concerning the estimation of the CF of transport activities, the 253 Intergovernmental Panel on Climate Change (IPCC) characterization factors have been applied 254 to quantify the equivalent  $CO_2$  emissions to be added to those of the food production phase. 255 Inventory data taken from the Ecoinvent ® v3.2 database (Wernet et al., 2016) have been 256 considered for road and sea transport.

257

258 2.2. Nutritional quality estimation

As previously mentioned, the nutritional quality of a diet is as important as its environmental impact, whether or not it is considered a sustainable diet, and it is also an important concept in

our time, when the growing trend towards a healthy lifestyle includes the consumption of nutrient-rich foods instead of high-calorie products (FAO, 2010). In this sense, the nutritional quality of both dietary scenarios (RAD and GD) has been analyzed from an average daily menu perspective rather than from a single meal evaluation, which would not provide sufficient representative information on consumer habits (Van Kernebeek et al., 2014). In this case, the concept of daily menu is based on the average amount of each food-item consumed per person in a day.

268 In this study, two different nutritional indexes have been proposed for analysis, as they 269 could be considered complementary. Firstly, the Nutrient Rich Diet 9.3 (NRD9.3) index 270 proposed by Van Kernebeek et al. (2014) was calculated (see Equation 1). This score considers 271 the daily intake of nine nutrients to encourage (protein, fiber, calcium, iron, magnesium, 272 potassium, vitamin A, vitamin E and vitamin C) and three nutrients to limit (total sugar, saturated 273 fats and sodium) (Castañé and Antón, 2017; Van Kernebeek et al., 2014). Consequently, the 274 greater the amount of nutrients ingested to encourage and the smaller the amount of nutrients 275 to limit, the higher the NRD 9.3 index is. Nevertheless, when the 9 nutrients to encourage 276 exceed the Recommended Daily Value (RDV), they are capped to this former value, in order to 277 avoid overestimation caused by overconsumption.

278 
$$NRD9.3 = \left(\sum_{i=1}^{i=9} \frac{nutrient_{\ i\ capped}}{RDV_i} - \sum_{k=1}^{k=3} \frac{nutrient_k}{RDV_k}\right) * 100 \tag{1}$$

279 On the other hand, a health gain score has also been proposed for estimation, which is 280 based on certain parameters other than the nutrients mentioned above (see Equation 2). This 281 health index has been developed by Van Dooren et al. (2014) taking into account the 282 complexity of determining the health benefits of diets (Van Dooren et al., 2014). To this end, the 283 health parameters established by different health organizations such as WHO<sup>1</sup>, World Cancer 284 Research Fund (WCRF)<sup>2</sup> or Dutch Health Council (DHC)<sup>3</sup> were taken into account.

285 Health gain = 
$$\left(\frac{g \ veg}{200} + \frac{g \ fruit}{200} + \frac{g \ fiber}{30} + \frac{30}{en \ \% \ total \ fat} + \frac{10}{en \ \% \ free \ sugar} + \frac{6}{g \ salt} + \frac{2000}{kcal \ energy}\right) * \frac{100}{7}$$
 (2)

The following parameters have been considered for the estimation: the daily intake of two food groups (i.e., vegetables and fruits), the daily percentage of energy obtained from total fatty

<sup>&</sup>lt;sup>1</sup><u>http://www.who.int/</u> (accessed June, 2018)

<sup>&</sup>lt;sup>2</sup> <u>https://www.wcrf.org/</u> (accessed June, 2018)

<sup>&</sup>lt;sup>3</sup> <u>https://www.gezondheidsraad.nl/en/home</u> (accessed June, 2018)

288 acids and free sugars, the daily intake of sodium and fiber and the total daily energy intake 289 (kcal·day<sup>-1</sup>). Therefore, the amounts of vegetables, fruits and fiber consumed are beneficial 290 elements to encourage. Furthermore, it is not recommended to exceed the reference values for 291 the daily percentage of energy intake obtained from total fatty acids and free sugars and the 292 daily intake of sodium. Moreover, to contextualize the health gain score obtained for the RAD 293 and GD scenarios, the recommended reference values of the mentioned parameters reported 294 by the WHO (WHO, 2003) have been considered. Additionally, the complete nutritional 295 composition of the foodstuffs has been obtained from the Spanish Food Composition Database 296 (AECOSAN, 2018).

Finally, the estimated indexes for both dietary scenarios will be compared with other results (for these scores) available in the literature (van Dooren and Aiking, 2016; Van Kernebeek et al., 2014) to rank their position in terms of nutritional quality.

300

#### 302 **3. Results and discussion**

#### 303 3.1. Carbon footprint assessment

304 The estimated CF for RAD and GD is 4.53 kg CO<sub>2</sub> eq. person<sup>-1</sup> day<sup>-1</sup> and 5.22 kg CO<sub>2</sub> eq. 305 person<sup>-1</sup>-day<sup>-1</sup> respectively. The main factor responsible for the total GHG emissions is the 306 foodstuffs production stage regardless of the scenario (i.e. 4.07 kg CO<sub>2</sub> eq. person<sup>-1</sup> day<sup>-1</sup> and 307 4.80 kg CO<sub>2</sub> eq·person<sup>-1</sup>·day<sup>-1</sup> for RAD and GD, respectively). The food production stage 308 includes all the background processes related to agricultural and farming activities, as well as 309 the corresponding industrial preparation activities if necessary (e.g., slaughterhouse, 310 refrigeration and packaging). Consequently, it can be observed that transport activities are 311 responsible for around 10% of total GHG emissions in both scenarios, specifically, ~0.4 kg CO<sub>2</sub> 312 eq.person<sup>-1</sup>.day<sup>-1</sup>. The rationale behind this fact is that, in the case of food patterns from the 313 same geographical area, the foodstuffs come from the same sources in most cases and are 314 transported over similar distances.

Figure 1 shows the individual CF per scenario for the food production phase, including the distribution by contributing food category. As it can be observed, livestock products (i.e. meat and dairy products) are the main contributor to the CF. Not only because they are some of the most consumed foods (**Table 1**) but also because they are the foods with the worst associated environmental profiles (Aleksandrowicz et al., 2016).



#### 321

Figure 1. Distribution of food category contributions to the global carbon footprint. Acronyms:
 RAD – Recommended Atlantic Diet scenario; GD – Galician Diet scenario.

Focusing on meat products, both scenarios have a similar CF (i.e. 1.9 kg CO<sub>2</sub> eq and 2.2 kg 324 325 CO<sub>2</sub> eq respectively for RAD and GD), despite the fact that the amount of meat ingested is 326 roughly double in the GD compared to RAD as shown in Table 1. The rationale behind this 327 result is associated with beef consumption, which is similar in both scenarios (66.9 g and 56.6 g 328 respectively in RAD and GD), being this type of meat the one with the worst associated 329 environmental profile: 28.60 kg CO<sub>2</sub> eq kg<sup>-1</sup> according to the average value reported by Clune et 330 al. (2017). The CF associated with this amount of beef is 1.91 kg  $CO_2$  eq and 1.62 kg  $CO_2$  eq 331 per person and day, being responsible for 42% and 31% of total GHG emissions in RAD and 332 GD, respectively.

By comparison, the contribution to the total CF from meat consumption, taking into account other types of meat, is much lower than that from beef and veal. In addition, it can be noted that beef alone accounts for about half of the total CF in RAD, and about a third in the GD.

336 For all other food products, dairy products also report a remarkable effect on the CF 337 regardless the scenario under study. In RAD, dairy products are responsible for 16% of the total 338 CF. In the case of GD, their contribution is lightly lower (15%). This is a consequence of the 339 notable intake of dairy products in the Galician region as shown in **Table 1**. The production of 340 the seafood consumed is the third largest in terms of GHG emissions in both scenarios (see 341 Figure 1). However, in this food category it is necessary to distinguish between RAD and GD. 342 Although the amount of seafood products is similar in both dietary patterns (195 g and 182 g 343 respectively for RAD and GD), the derived CF is almost twice as much in GD as in RAD (0.73 344 and 0.41 kg CO<sub>2</sub> eq respectively). The rationale behind this surprising result is mainly explained 345 by the consumption of certain species in the GD, with relatively high GHG emission factors (e.g. 346 salmon, hake, flatfish, prawns and canned tuna), which were not considered within the designed 347 menus of RAD. As a result, seafood products account for about 14% of the total CF for the GD, 348 and about 9% for RAD.

Moreover, it is interesting to note the contribution of the processed food category in the case of GD. As mentioned above, this category only appears in GD, as it includes products not recommended by the health authorities due to their low nutritional quality (Xunta de Galicia, 2013). However, they are actually present in the current consumption trends. In this sense, the consumption of processed foods represents around 6% (0.29 kg CO<sub>2</sub> eq) of the total CF in GD, which is a higher ratio than that associated with other ingredients such as vegetables and fruits ( $\approx$ 4% and  $\approx$ 2% respectively), which are considered basic foods in the diet.

Finally, in terms the contribution to the CF score of food losses along the food supply chain, its relevance to the environmental footprint can be highlighted, since food losses represent around 14% of the total CF for both scenarios (around 0.7 kg CO<sub>2</sub> eq). Therefore, attention should be paid to this hotspot.

360 3.2 Comparing the nutritional quality of RAD and GD scenarios

Regarding the NRD9.3 index, the results estimated for RAD and GD are 474 and 242, respectively, as shown in **Table 2**. The amount of each nutrient ingested to limit and promote is

363 also depicted in the table, as well as its recommended daily intake value (RDV). It is important

364 to note that when the RDV outcomes fall between two values, an average value has been

365 considered in the estimate.

**Table 2.** Nutrient Rich Diet 9.3 (NRD9.3) index for both Recommended Atlantic Diet (RAD) and Galician Diet (GD) and Recommended Daily Value (RDV) for each nutrient considered in the index.

Nutrient	Units	RDV	RAD	GD
Protein	g	50	101	133
Fiber	q	25	39.9	36.8
Vit A	<u>_</u>	700-3000	1448	967
Vit C	ma	60-2000	320	195
	ma	20.4000	02.0	190
VITE	mg	20-1000	23.0	11.7
Са	mg	1000-2500	1083	1182
Fe	mg	18-45	33.0	18.5
к	mg	3500	4938	4151
Mg	mg	400	435	364
Saturated fats	g	20	19.3	35.2
Free sugars	q	50	117	101.1
 Na	ma	1500-2400	1449	2537
NRD9.3	5		474	242

369

370 As can be seen, the intake amount of many nutrients is higher than the RDV in both 371 scenarios. On the other hand, comparing the ingestion values corresponding to the nutrients to 372 encourage, their intake is higher for most nutrients in RAD than in GD mainly due to the large 373 consumption of fruits and vegetables (see Table 1). However, in terms of protein and calcium, 374 the situation is reversed. The higher intake of protein in GD (2.5 times higher than the RDV) is 375 related to the remarkable consumption of meat, considerably higher than the recommended 376 values (see Table 1). The higher intake of dairy products in GD consequently increases the 377 amount of calcium ingested, being 0.2 times higher than the RDV.

378 When comparing the intake of nutrients to limit, saturated fat, free sugars and sodium are 379 ingested in higher amounts in the GD than in the RAD, which is mainly attributed to the consumption of processed foods (AECOSAN, 2018). In this regard, it is important to mention that in RAD, the intake of all limiting nutrients is below the RDV. In contrast, the intake of saturated fat and sodium in the GD is considerably higher than the recommended values. The intake of added free is much higher than the RDV for the GD, mainly due to the intake of processed foods and sweets. Considering that sodium is the leading cause of death due to an inadequate diet, followed by a low intake of fiber and fruits (Afshin et al., 2019), emphasis is placed on avoiding excessive consumption of this element.

The health gain scores for the RAD and GD scenarios are shown in **Table 3**, as well as the required reference values for each parameter (Equation 2). As previously reported, the health gain score is the result of the ratio between the reference intake values considered for vegetables, fruits, total fatty acids, free sugars, fiber, sodium and energy and those for actual intake in both scenarios.

392 Table 3. Health gain score results for both Recommended Atlantic Diet (RAD) and Galician Diet393 (GD) scenarios.

Indicator	Units	Reference value	RAD	GD
Vegetables	q	200	462	424
Fruit	a	200	747	321
Total fatty acids	<u>9</u>	30	31.8	31.9
	0/	10	22.2	22.5
	70	10		23.5
Fiber	g	30.0	39.9	36.8
Sodium	g	6.0	1.4	3.1
Energy	kcal	2000	2100	2381
Health gain score		100	198	115

394

As can be seen, the amount of vegetables and fruits consumed in both scenarios is higher than the reference values set by WHO (WHO, 2003). In this sense, it is important to note that the intake of vegetables is more than double the reference quantity regardless of the scenario analyzed, and almost three times higher for the intake of fruits for RAD. Even the quantity of fruit consumed in the GD is 50% higher than the reference value (321 g versus 200 g). The justification for these differences is associated with the high availability of vegetables and fruits 401 in the Atlantic region throughout the year, as well as the cultural culinary tradition of the region. 402 In terms of fiber intake, it can be seen that the amount consumed in both scenarios is also 403 above the reference value (30 g), an increase of 33% and 23% respectively for RAD and GD. 404 As beneficial parameters, a higher intake of vegetables, fruits and fiber leads to better nutritional 405 quality and, consequently, a higher health gain score. Taking into account the percentage of 406 energy obtained from total fatty acids and free sugars, it should be mentioned that the 407 proportions are above the reference value. While it is only 2% higher for total fatty acids, the 408 percentage of energy from sugars far exceeds the recommended value, which is evidenced by 409 a clear negative effect on the final health gain score. However, it is important to point out that 410 the high intake of free sugars is directly related to high fruit consumption. On the other hand, the 411 daily intake of sodium is lower than the recommended dose (6 g), at values around 75% and 412 50% lower in RAD and DG, respectively.

Energy intake in both scenarios varies slightly from the 2,000 kcal set by WHO (WHO, 2012). An increase in calorie intake is not considered advisable and has a negative impact on the health benefit score. With all these data reported, the health gain score has been estimated for both scenarios according to Equation 2.

417 The values obtained for both scenarios are 198 points and 115 points respectively for RAD 418 and GD (Table 3). Despite this outstanding difference, both scores are above the WHO 419 benchmark (i.e. 100 points). Comparing the RAD and DG scores, it can be seen that the reason 420 for the large difference in health gain scores is directly associated with fruit intake, as no notable 421 differences in the remaining parameters can be identified. Otherwise, this practice also 422 influences the fact that the health gain values in GD are above the reference value, mainly due 423 to higher consumption of vegetables and fruits. Furthermore, it should be borne in mind that due 424 to non-excessive energy intake (Table 3), the Health Score is not affected by this factor which 425 on the contrary, would significantly penalize the nutritional quality.

426 3.3. Benchmarking environmental and health gain scores

Taking into account the results obtained in terms of nutritional and CF indexes, it is necessary to establish a relationship between them and those of the different studies available in the literature (Van Dooren et al., 2014; van Dooren and Aiking, 2016). The health gain score for RAD (198) is above the values found in the literature for other well-positioned dietary options

from environmental and health approaches such as the Mediterranean (MD), vegan (VD), vegetarian (VGD) and semi-vegetarian (SVGD) diets as can be seen in **Figure 2**. On the other hand, the health gain score achieved for GD (114) is above the reference value (100) as mentioned above, and is consistent with those identified for other dietary patterns (**Figure 2**), and even better than VGD or SVGD.





Figure 2. Comparison in terms of Health gain score of RAD and GD with alternative healthy
diets available in literature. Acronyms: MD - Mediterranean diet; VD - Vegan diet; VGD Vegetarian diet; SVGD - Semi-Vegetarian diet.

440

441 In order to establish a relationship between the nutritional quality and the CF, Table 4 442 details the NRD9.3 and CF scores of both scenarios under study and the studies available in 443 the literature; the health gain score has not been included in this table due to lack of information 444 to perform the estimation of this index for these diets. Regarding the NRD9.3 values, RAD and 445 GD obtain a score in line with those of these diets that use the same RDV (Castañé and Antón, 446 2017). As it can be seen in **Table 4**, the RAD score reports a higher nutritional quality than 447 others in the literature. On the other hand, the nutritional score for GD is lower than most of the 448 values cited (e.g., Mediterranean and Healthy diets). Nevertheless, it is necessary to keep in 449 mind that this is a scenario based on real consumption trends and it is not a one based on 450 recommendations such as the other studies mentioned in the literature (Castañé and Antón, 451 2017; Pathak et al., 2010; Risku Norja et al., 2009; Saxe et al., 2012; Van Dooren et al., 2014). 452 It means that certain type of foods of lower nutritional quality, such as processed food and 453 sweets, are included.

	NRD9.3	CF (kg CO₂ eq⋅person <sup>-</sup> ¹⋅day <sup>-1</sup> )
RAD	474	4.53
GD	242	5.22
Mediterranean diet (Castañé and Antón, 2017)	389	2.86
Vegan diet (Castañé and Antón, 2017)	469	1.86
Mediterranean diet (Sáez-Almendros et al., 2013)	_*	2.19
Spanish current diet (Sáez-Almendros et al., 2013)	_*	4.39
Vegan diet (Castañé and Antón, 2017)	469	1.86
Vegetarian diet (Pathak et al., 2010)	424	0.58
Healthy diet (Risku-Norja et al., 2009)	382	3.84
Vegan diet (Risku-Norja et al., 2009)	442	2.47
<b>Vegetarian diet</b> (Van Dooren et al., 2014)	_*	3.2
Danish dietary pattern (Saxe et al., 2012)	112	5.52
*nutritional information not available		

**Table 4.** Summary of NRD9.3 and Carbon Footprint (CF) indexes regarding RAD(Recommended Atlantic diet), GD (Galician diet) and other diets available in literature. 

\*nutritional information not available.

Taking into account the CF for RAD and GD, and comparing these values with others 458 459 reported in the literature, our scenarios involve relatively high CF scores (i.e. 4.53 and 5.22 kg 460 CO<sub>2</sub> eq.pers<sup>-1</sup>.day<sup>-1</sup> respectively for RAD and GD) mainly due to the huge consumption of beef 461 as mentioned above. In view of the results, it is important to refer to the fact that all the dietary 462 patterns mentioned, except the current Spanish dietary pattern of Sáez-Almendros et al. (2013), 463 the Danish dietary pattern of Saxe et al., (2012) and GD, are diets based on recommendations, 464 generally leading to lower CF outcomes. However, attention should be paid to the system 465 boundaries considered since waste production and distribution to households have been 466 included in our study, these stages being responsible for 24% of total CF. In this sense, food 467 distribution and waste are relevant hotspots that significantly increase the CF value and have 468 not been considered in the abovementioned studies. As can be seen, nutritional quality and CF 469 are not always inversely proportional parameters when comparing diets from different studies; 470 however, there is a trend that links both. In this sense, a higher nutritional quality usually 471 translates into a lower CF, as is also observed for the Mediterranean and vegan diets in 472 Castañé and Antón (2017). Considering these interactions, it is important to stand out that the 473 variation on both the nutritional quality and the CF is always related to the products of animal 474 origin (i.e. meat and dairy).

475

476 3.4. Sensitivity analysis of diets sustainability

477 As mentioned above, beef meat is the main source of GHG emissions for both scenarios 478 under study. For this reason, a sensitivity analysis is proposed to determine the effect of the 479 substitution of this type of meat by other types of foods that imply lower GHG emissions and a similar contribution of protein in both diets, as summarized in Table 5, without significantly 480 481 affecting energy intake (kcal per day). In this sense, six protein-rich foods have been selected: 482 Two alternative types of meat (pork and chicken) have been selected as alternatives to beef, 483 taking into account that they are the second (pork) and third (chicken) most consumed meats in 484 Galicia (SERGAS, 2007); two legumes (lentils and peas) have also selected for analysis taking 485 into account the recommendations from Jungbluth at al. (2016), which advise the consumption 486 of vegetable proteins as opposed to animal proteins (Jungbluth et al., 2016); finally a couple of

487 fish products (hake and tuna) have been considered, considering the priority of fish
488 consumption in the Atlantic diet (Álvarez and Peláez, 2018).

	RAD		GD	
Scenario	CF (kg CO₂ eq⋅person <sup>-</sup> ¹⋅day⁻¹)	NRD9.3	CF (kg CO₂ eq∙person⁻¹∙day⁻¹)	NRD9.3
Beef-meat	4.53	474	5.22	242
Meat				
Pork	2.81	512	3.79	245
Chicken	2.84	515	3.77	247
Legumes				
Lentils	2.63	513	3.67	269
Peas	2.62	523	3.62	266
Fish				
Hake	3.08	512	3.96	271
Tuna	2.72	515	3.69	272

489 **Table 5.** Sensitivity analysis of Carbon Footprint (CF) and NRD9.3 results when substituting

490 beef-meat in RAD and GD scenarios by alternative foodstuffs (meat, legumes and fish).

491

492 As regards the simulations carried out for RAD and GD, it is noted that the removal of beef 493 meat in both cases results in a drastic reduction in the CF. In this sense, the highest variation in 494 both the CF and NRD9.3 scores occurs when beef is replaced by legumes, with a reduction in 495 the CF of about 40% and 30% for RAD and GD respectively, and an improvement in the 496 nutritional quality of about 10% in both situations. On the other hand, the consideration of 497 alternative meats reduces the CF by 40% and 30% in RAD and GD respectively, resulting in an 498 improvement of the nutritional quality in both scenarios, around 10% for RAD and 2% for GD. 499 Finally, the alternative of fish products also leads to an improvement in the nutritional quality, in 500 this case the highest one in the GD. Regarding the CF score, it is also reduced in both 501 scenarios although the reduction is lower if hake is considered than tuna, which has a 502 moderately high GHG emission factor. It could therefore be reported that the replacement of 503 beef with alternative food products would be a beneficial measure both environmentally and 504 nutritionally.

#### 505 Analysis of data quality

In terms of data quality, CF is selected as an environmental indicator. In this regard, the variability of LCA data for each product should be taken into account, e.g., for beef meat the carbon footprint ranges from 9.3 kg  $CO_2$  eq·kg<sup>-1</sup> for organic farming (Solid Forest, 2011) to

28.73 kg CO2 eq·kg<sup>-1</sup> for conventional farming (Clune et al., 2017). Thus, most conservative 509 510 figures have been taken into account and the results have been carefully discussed. 511 Furthermore, the beef meat has been identified as a hotspot regarding the results for CF, which 512 could be identified as an opportunity by LCA practitioners to improve their production processes 513 (e.g., technological adaptation at the farm level in order to minimize methane emissions from 514 either enteric fermentation or manure management)(Hyland et al., 2017). Moreover, additional 515 environmental indicators, such as water footprint or land occupation, should be taken into 516 consideration to obtain a more complete environmental profile.

Additionally, with regard to the source of actual food consumption figures for the GD scenario, data from a survey conducted in 2007 has been used, as previously mentioned, due to the lack of more updated real representative data. Thus, this survey is the most recent one for Galicia and the most detailed. However, consumption habits evolve and consequently the CF and nutritional quality. Therefore, efforts should be conducted in the design of a food frequency consumption questionnaire to be supplied to the Galician population for the handling of real parameters.

524

#### 525 4. Conclusions

526 The outcomes of this study prove that there is a deviation between actual consumption 527 patterns and diets based on health recommendations, both from an environmental and 528 nutritional point of view. Thus, in the specific case of Galicia, the current dietary pattern obtains 529 much lower scores in nutritional indexes and a higher CF than the recommendations from the 530 traditional Atlantic diet. Therefore, a change in the current trends of food consumption towards the recommendations of the Atlantic pyramid would be beneficial. In this sense, as weak spots 531 532 in the GD (excessive sodium intake), processed and pre-cooked foods should be left aside, as 533 they are the ones with the worst nutritional quality. However, it has also been proven that both the nutritional and environmental quality of the two studied scenarios can be improved by 534 535 replacing beef with a more sustainable source of protein, taking as reference the methodology 536 used in this study. In this sense, it is advisable to provide more proteins of vegetable origin than 537 those of animal origin, with legumes being the best possible substitute.

The results can be useful for regional policy makers and sanitary authorities to act on the hotspots that cause the greatest loss of nutritional quality and the resulting increased carbon footprint. In the same way, they can be also extended to other regions or countries interested in promoting adherence to the Atlantic diet.

Further research should be based on the design of new variants for the Atlantic diet, focusing on improving environmental quality without affecting its nutritional quality; the changes should be made by replacing foods with a higher environmental impact with more sustainable ones, included in the traditional foods of the Atlantic diet. In addition, taking into account the concept of sustainable diet, future research should include other relevant environmental impacts that are also significant in studies related to food production, such as the water footprint, and socio-economic indicators related to the affordability of diets.

549

#### 550 Acknowledgements

This research has been supported by a project granted by Xunta de Galicia (project ref. ED431F 2016/001). Dr. S. González-Garcia would like to express her gratitude to the Spanish Ministry of Economy and Competitiveness for financial support (Grant reference RYC-2014-14984). The authors belong to the Galician Competitive Research Group GRC 2013-032 as well as to CRETUS (AGRUP2015/02), co-funded by Xunta de Galicia and FEDER (EU).

556

#### 557 References

- 558 AECOSAN, 2018. Spanish Food Composition Database [WWW Document].
- Afshin, A., Sur, P.J., Fay, K.A., Cornaby, L., Ferrara, G., Salama, J.S., Mullany, E.C., Abate,
- 560 K.H., Abbafati, C., Abebe, Z., Afarideh, M., Aggarwal, A., Agrawal, S., Akinyemiju, T.,
- 561 Alahdab, F., Bacha, U., Bachman, V.F., Badali, H., Badawi, A., Bensenor, I.M., Bernabe,
- 562 E., Biadgilign, S.K.K., Biryukov, S.H., Cahill, L.E., Carrero, J.J., Cercy, K.M., Dandona, L.,
- 563 Dandona, R., Dang, A.K., Degefa, M.G., El Sayed Zaki, M., Esteghamati, A., Esteghamati,
- 564 S., Fanzo, J., Farinha, C.S. e S., Farvid, M.S., Farzadfar, F., Feigin, V.L., Fernandes, J.C.,
- 565 Flor, L.S., Foigt, N.A., Forouzanfar, M.H., Ganji, M., Geleijnse, J.M., Gillum, R.F., Goulart,
- 566 A.C., Grosso, G., Guessous, I., Hamidi, S., Hankey, G.J., Harikrishnan, S., Hassen, H.Y.,
- 567 Hay, S.I., Hoang, C.L., Horino, M., Islami, F., Jackson, M.D., James, S.L., Johansson, L.,

568	Jonas, J.B., Kasaeian, A., Khader, Y.S., Khalil, I.A., Khang, YH., Kimokoti, R.W.,
569	Kokubo, Y., Kumar, G.A., Lallukka, T., Lopez, A.D., Lorkowski, S., Lotufo, P.A., Lozano,
570	R., Malekzadeh, R., März, W., Meier, T., Melaku, Y.A., Mendoza, W., Mensink, G.B.M.,
571	Micha, R., Miller, T.R., Mirarefin, M., Mohan, V., Mokdad, A.H., Mozaffarian, D., Nagel, G.,
572	Naghavi, M., Nguyen, C.T., Nixon, M.R., Ong, K.L., Pereira, D.M., Poustchi, H., Qorbani,
573	M., Rai, R.K., Razo-García, C., Rehm, C.D., Rivera, J.A., Rodríguez-Ramírez, S.,
574	Roshandel, G., Roth, G.A., Sanabria, J., Sánchez-Pimienta, T.G., Sartorius, B.,
575	Schmidhuber, J., Schutte, A.E., Sepanlou, S.G., Shin, MJ., Sorensen, R.J.D.,
576	Springmann, M., Szponar, L., Thorne-Lyman, A.L., Thrift, A.G., Touvier, M., Tran, B.X.,
577	Tyrovolas, S., Ukwaja, K.N., Ullah, I., Uthman, O.A., Vaezghasemi, M., Vasankari, T.J.,
578	Vollset, S.E., Vos, T., Vu, G.T., Vu, L.G., Weiderpass, E., Werdecker, A., Wijeratne, T.,
579	Willett, W.C., Wu, J.H., Xu, G., Yonemoto, N., Yu, C., Murray, C.J.L., 2019. Health effects
580	of dietary risks in 195 countries, 1990–2017: a systematic analysis for the Global Burden
581	of Disease Study 2017. Lancet 6736. https://doi.org/10.1016/S0140-6736(19)30041-8
582	Aguilera, E., Guzmán, G., Alonso, A., 2015. Greenhouse gas emissions from conventional and
583	organic cropping systems in Spain. I. Herbaceous crops. Agron. Sustain. Dev. 35, 713-
584	724. https://doi.org/10.1007/s13593-014-0267-9
585	Aleksandrowicz, L., Green, R., Joy, E.J.M., Smith, P., Haines, A., 2016. The impacts of dietary
586	change on greenhouse gas emissions, land use, water use, and health: A systematic
587	review. PLoS One 11, 1–16. https://doi.org/10.1371/journal.pone.0165797
588	Álvarez, J., Peláez, N., 2018. Fundación para la Diabetes [WWW Document].
589	Bach-Faig, A., Fuentes-Bol, C., Ramos, D., Carrasco, J.L., Roman, B., Bertomeu, I.F., Cristià,
590	E., Geleva, D., Serra-Majem, L., 2011. The Mediterranean diet in Spain: Adherence trends
591	during the past two decades using the Mediterranean Adequacy Index. Public Health Nutr.
592	14, 622–628. https://doi.org/10.1017/S1368980010002752
593	Batlle-Bayer, L., Bala, A., García-Herrero, I., Lemaire, E., Song, G., Aldaco, R., Fullana-i-
594	Palmer, P., 2019. National Dietary Guidelines: a potential tool to reduce greenhouse gas
595	emissions of current dietary patterns. The case of Spain. J. Clean. Prod. 213, 588-598.
596	https://doi.org/10.1016/j.jclepro.2018.12.215
597	Berlin, J., Sund, V., 2010. Environmental Life Cycle Assessment (LCA) of ready meals [Online]

- 598 http://environmental-life-cycle-assessment-lca-of-ready-meals-sik-2010804.pdf. Webpage
- 599 accessed: 26th March 2018, SIK-Report No 804 2010.
- 600 Carlsson-Kanyama, A., González, A.D., 2009. Potential contributions of food consumption
- 601 patterns to climate change. Am. J. Clin. Nutr. 89, 1704S–1709S.

602 https://doi.org/10.3945/ajcn.2009.26736AA.1704S

- 603 Castañé, S., Antón, A., 2017. Assessment of the nutritional quality and environmental impact of
- two food diets: A Mediterranean and a vegan diet. J. Clean. Prod. 167, 929–937.
- 605 https://doi.org/10.1016/j.jclepro.2017.04.121
- 606 Cencic, A., Chingwaru, W., 2010. The role of functional foods, nutraceuticals, and food
- 607 supplements in intestinal health. Nutrients 2, 611–625. https://doi.org/10.3390/nu2060611
- 608 Charanjit Kaur, Harish C. Kapoor, 2001. Antioxidants in fruits and vegetables. The millennium's
  609 health. Int. J. Food Sci. Technol. 36, 703–725.
- 610 Charro, A., Tojo, R., Varela, G., Leis, R., 2006. Decálogo de la Dieta Atlántica. Il Congreso
  611 Internacional de la Dieta Atlántica. Baiona 2006, Fundación. ed.
- 612 Clune, S., Crossin, E., Verghese, K., 2017. Systematic review of greenhouse gas emissions for
- 613 different fresh food categories. J. Clean. Prod. 140, 766–783.
- 614 https://doi.org/10.1016/j.jclepro.2016.04.082
- 615 Coelho, C.R. V, Pernollet, F., van der Werf, H.M.G., 2016. Environmental Life Cycle
- Assessment of Diets with Improved Omega-3 Fatty Acid Profiles. PLoS One 11,
- 617 e0160397. https://doi.org/10.1371/journal.pone.0160397
- 618 Committee on Climate Change, 2010. The fourth carbon budget: reducing the emissions
- 619 through 2020.
- 620 Da Silva, R., Bach-Faig, A., Raidó Quintana, B., Buckland, G., Vaz De Almeida, M.D., Serra-
- 621 Majem, L., 2009. Worldwide variation of adherence to the Mediterranean diet, in 1961-
- 622 1965 and 2000-2003. Public Health Nutr. 12, 1676–1684.
- 623 https://doi.org/10.1017/S1368980009990541
- De Boer, J., Schösler, H., Aiking, H., 2014. "Meatless days" or "less but better"? Exploring
- 625 strategies to adapt Western meat consumption to health and sustainability challenges.
- 626 Appetite 76, 120–128. https://doi.org/10.1016/j.appet.2014.02.002
- 627 Donati, M., Menozzi, D., Zighetti, C., Rosi, A., Zinetti, A., Scazzina, F., 2016. Towards a

- 628 sustainable diet combining economic, environmental and nutritional objectives. Appetite
- 629 106, 48–57. https://doi.org/10.1016/j.appet.2016.02.151
- 630 Esteve-Llorens, X., Darriba, C., Moreira, M.T., Feijoo, G., González-García, S., 2019. Towards
- an environmentally sustainable and healthy Atlantic dietary pattern: Life cycle carbon
- 632 footprint and nutritional quality. Sci. Total Environ. 646, 704–715.
- 633 https://doi.org/10.1016/j.scitotenv.2018.07.264
- 634 FAO, 2011. Global food losses and food waste Extent, causes and prevention. Rome.
- 635 FAO, 2010. Biodiversity and sustainable diets united against hunger. Rome.
- 636 Garcia-Herrero, I., Hoehn, D., Margallo, M., Laso, J., Bala, A., Batlle-Bayer, L., Fullana, P.,
- 637 Vazquez-Rowe, I., Gonzalez, M.J., Durá, M.J., Sarabia, C., Abajas, R., Amo-Setien, F.J.,
- 638 Quiñones, A., Irabien, A., Aldaco, R., 2018. On the estimation of potential food waste
- 639 reduction to support sustainable production and consumption policies. Food Policy 80, 24–
- 640 38. https://doi.org/10.1016/j.foodpol.2018.08.007
- 641 Garnett, T., 2011. Where are the best opportunities for reducing greenhouse gas emissions in
- the food system (including the food chain)? Food Policy 37, 463–466.
- 643 https://doi.org/10.1016/j.foodpol.2012.04.006
- 644 Gerlach, A., Hohfeld, L., Sharnhorst, S., Schudak, A., Schoenheit, I., Zwick, Y., 2013. The
- 645 Sustainable Shopping Basket. Berlin, Germany.
- 646 González-García, S., Esteve-Llorens, X., Moreira, M.T., Feijoo, G., 2018. Carbon footprint and
- 647 nutritional quality of different human dietary choices. Sci. Total Environ. 644.
- 648 https://doi.org/10.1016/j.scitotenv.2018.06.339
- 649 González-García, S., Hospido, A., Moreira, M.T., Feijoo, G., Arroja, L., 2013. Environmental life
- 650 cycle assessment of a galician cheese: San Simon da Costa. J. Clean. Prod. 52, 253–262.
  651 https://doi.org/10.1016/j.jclepro.2013.03.006
- 652 Gunady, M.G.A., Biswas, W., Solah, V.A., James, A.P., 2012. Evaluating the global warming
- 653 potential of the fresh produce supply chain for strawberries, romaine/cos lettuces (Lactuca
- 654 sativa), and button mushrooms (Agaricus bisporus) in Western Australia using life cycle
- assessment (LCA). J. Clean. Prod. 28, 81–87.
- 656 https://doi.org/10.1016/j.jclepro.2011.12.031
- Heller, M.C., Keoleian, G.A., Willett, W.C., 2013. Toward a life cycle-based, diet-level framework

- 658 for food environmental impact and nutritional quality assessment: a critical review. Env.
- 659 Sci Technol 47, 12632–12647. https://doi.org/10.1021/es4025113
- Hoek, A.C., Pearson, D., James, S.W., Lawrence, M.A., Friel, S., 2017. Shrinking the food-
- 661 print : A qualitative study into consumer perceptions, experiences and attitudes towards
- healthy and environmentally friendly food behaviours. Appetite 108, 117–131.
- 663 https://doi.org/10.1016/j.appet.2016.09.030
- 664 Hyland, J.J., Henchion, M., McCarthy, M., McCarthy, S.N., 2017. The role of meat in strategies
- to achieve a sustainable diet lower in greenhouse gas emissions: A review. Meat Sci. 132,
  189–195. https://doi.org/10.1016/j.meatsci.2017.04.014
- 667 Iribarren, D., Vázquez-Rowe, I., Hospido, A., Moreira, M.T., Feijoo, G., 2011. Updating the
- 668 carbon footprint of the Galician fishing activity (NW Spain). Sci. Total Environ. 409, 1609–
- 669 1611. https://doi.org/10.1016/j.scitotenv.2011.01.007
- 670 Irz, X., Leroy, P., Réquillart, V., Soler, L.G., 2016. Welfare and sustainability effects of dietary
  671 recommendations. Ecol. Econ. 130, 139–155.
- 672 https://doi.org/10.1016/j.ecolecon.2016.06.025
- 673 ISO 14040, 2006. Environmental management Life cycle assessment Principles and

674 framework, Iso 14040. Switzerland. https://doi.org/10.1136/bmj.332.7550.1107

- Jungbluth, N., Eggenberger, S., Nowack, K., Keller, R., 2016. Life cycle assessment of meals
- based on vegetarian protein sources Impact assessment 1–6.
- 677 Livsmedelsverket, 2009. The Swedish National Food Administration's Environmentally Effective
- 678 Food Choices. Stockholm, Sweden.
- 679 Nielsen, N.I., Jørgensen, M., Rasmussen, I.K., 2013. Greenhouse Gas Emission from Danish
- 680 Organic Egg Production esti mated via LCA Methodology 1–27.
- Notarnicola, B., Tassielli, G., Renzulli, P.A., Castellani, V., Sala, S., 2017. Environmental
- 682 impacts of food consumption in Europe. J. Clean. Prod. 140, 753–765.
- 683 https://doi.org/10.1016/j.jclepro.2016.06.080
- 684 Noya, I., Villanueva-Rey, P., González-García, S., Fernandez, M.D., Rodriguez, M.R., Moreira,
- 685 M.T., 2017. Life Cycle Assessment of pig production: A case study in Galicia. J. Clean.
- 686 Prod. 142, 4327–4338. https://doi.org/10.1016/j.jclepro.2016.11.160
- 687 Pathak, H., Jain, N., Bhatia, A., Patel, J., Aggarwal, P.K., 2010. Carbon footprints of Indian food

- 688 items. Agric. Ecosyst. Environ. 139, 66–73. https://doi.org/10.1016/j.agee.2010.07.002
- 689 Pernollet, F., Coelho, C.R.V., Van der Werf, H.M.G., 2017. Methods to simplify diet and food life
- 690 cycle inventories: Accuracy versus data-collection resources. J. Clean. Prod. 140, 410–
  691 420. https://doi.org/10.1016/j.jclepro.2016.06.111
- Reddy, S., Lang, T., Dibb, S., 2009. Setting the table. Advice to Government on priority
  elements of sustainable diets, Cutting Tool Engineering. London, UK.
- Risku Norja, H., Kurppa, S., Helenius, J., 2009. Dietary choices and greenhouse gas emissions
   assessment of impact of vegetarian and organic options at national scale. Prog. Ind. Ecol.
- 696 An Int. J. 6, 340. https://doi.org/10.1504/PIE.2009.032323
- 697 Ritchie, H., Reay, D.S., Higgins, P., 2018. The impact of global dietary guidelines on climate

698 change. Glob. Environ. Chang. 49, 46–55. https://doi.org/10.1016/j.gloenvcha.2018.02.005

- 699 Sáez-Almendros, S., Obrador, B., Bach-Faig, Serra-Majem, L., 2013. Environmental footprints
- 700 of Mediterranean versus Western dietary patterns: beyond the health benefits of the
- 701 Mediterranean diet. Environ. Heal. 12, 1–8. https://doi.org/10.1186/1476-069X-12-118
- Saxe, H., Larsen, T.M., Mogensen, L., 2012. The global warming potential of two healthy Nordic
- diets compared with the average Danish diet. Clim. Change 116, 249–262.
- 704 https://doi.org/10.1007/s10584-012-0495-4
- Scarborough, P., Appleby, P.N., Mizdrak, A., Briggs, A.D.M., Travis, R.C., Bradbury, K.E., Key,

706 T.J., 2014. Dietary greenhouse gas emissions of meat-eaters, fish-eaters, vegetarians and

- vegans in the UK. Clim. Change 125, 179–192. https://doi.org/10.1007/s10584-014-1169-
- 708

- 709 SERGAS, 2007. Encuesta sobre los hábitos alimentarios de la población adulta gallega, 2007.
  710 Santiago de Compostela.
- 711 Solid Forest, 2011. Informe ISGEA Carne de ternera 1–8.
- Sonesson, U., Anteson, F., Davis, J., Sjödén, P.-O., 2005. Home transport and wastage:
- environmentally relevant household activities in the life cycle of food. Ambio 34, 371–375.
  https://doi.org/10.1579/0044-7447-34.4.371
- 715 Springmann, M., Clark, M., Mason-D'Croz, D., Wiebe, K., Bodirsky, B.L., Lassaletta, L., De
- 716 Vries, W., Vermeulen, S.J., Herrero, M., Carlson, K.M., Jonell, M., Troell, M., DeClerck, F.,
- 717 Gordon, L.J., Zurayk, R., Scarborough, P., Rayner, M., Loken, B., Fanzo, J., Godfray,

718 H.C.J., Tilman, D., Rockström, J., Willett, W., 2018. Options for keeping the food system

719 within environmental limits. Nature. https://doi.org/10.1038/s41586-018-0594-0

- 720 Steering Group, 2010. Food for Tomorrow Proposal for Finland's National Food Strategy Name
- of the publication: Food for Tomorrow Proposal for Finland's National Food Strategy.
- Tilman, D., Clark, M., 2014. Global diets link environmental sustainability and human health.

723 Nature 515, 518–522. https://doi.org/10.1038/nature13959

- Tojo, R., Leis, R., 2009. La Dieta Atlántica, el pescado y las algas Su importancia en el
- neurodesarrollo y la función cerebral. Chapter: El papel de la Dieta Atlántica como
- 726 contrapunto saludable a la Dieta Ocidental actual. Universidad de Santiago de
- 727 Compostela.
- 728 Tukker, A., Goldbohm, R.A., De Koning, A., Verheijden, M., Kleijn, R., Wolf, O., Pérez-
- 729 Domínguez, I., Rueda-Cantuche, J.M., 2011. Environmental impacts of changes to

healthier diets in Europe. Ecol. Econ. 70, 1776–1788.

- 731 https://doi.org/10.1016/j.ecolecon.2011.05.001
- Tukker, A., Jansen, B., 2006. Environmental impacts of products: A detailed review of studies.

733 J. Ind. Ecol. 10, 159–182. https://doi.org/10.1162/jiec.2006.10.3.159

- van de Kamp, M.E., van Dooren, C., Hollander, A., Geurts, M., Brink, E.J., van Rossum, C.,
- 735 Biesbroek, S., de Valk, E., Toxopeus, I.B., Temme, E.H.M., 2018. Healthy diets with
- 736 reduced environmental impact? The greenhouse gas emissions of various diets adhering
- to the Dutch food based dietary guidelines. Food Res. Int. 0–1.
- 738 https://doi.org/10.1016/j.foodres.2017.06.006
- van Dooren, C., Aiking, H., 2016. Defining a nutritionally healthy, environmentally friendly, and

740 culturally acceptable Low Lands Diet. Int. J. Life Cycle Assess. 21, 688–700.

- 741 https://doi.org/10.1007/s11367-015-1007-3
- Van Dooren, C., Marinussen, M., Blonk, H., Aiking, H., Vellinga, P., 2014. Exploring dietary
- 743 guidelines based on ecological and nutritional values: A comparison of six dietary patterns.
- 744 Food Policy 44, 36–46. https://doi.org/10.1016/j.foodpol.2013.11.002
- Van Kernebeek, H.R.J., Oosting, S.J., Feskens, E.J.M., Gerber, P.J., De Boer, I.J.M., 2014. The
- effect of nutritional quality on comparing environmental impacts of human diets. J. Clean.
- 747 Prod. 73, 88–99. https://doi.org/10.1016/j.jclepro.2013.11.028

- Vaz Velho, M., Pinheiro, R., Sofia, A., 2016. The Atlantic Diet Origin and features. Int. J. Food
  Stud. 5, 106–119. https://doi.org/10.7455/ijfs/5.1.2016.a10
- Volpe, R., Messineo, S., Volpe, M., Messineo, A., 2015. Carbon footprint of tree nuts based
  consumer products. Sustain. 7, 14917–14934. https://doi.org/10.3390/su71114917
- 752 Vringer, K., Benders, R., Wilting, H., Brink, C., Drissen, E., Nijdam, D., Hoogervorst, N., 2010. A
- 753 hybrid multi-region method (HMR) for assessing the environmental impact of private
- 754 consumption. Ecol. Econ. 69, 2510–2516. https://doi.org/10.1016/j.ecolecon.2010.07.027
- Wang, H., Cao, G., Prior, R.L., 1996. Total antioxidant capacity of fruits. J. Agric. Food Chem.
- 756 44, 701–705. https://doi.org/10.1021/jf950579y
- 757 Werner, L.B., Flysjö, A., Tholstrup, T., 2014. Greenhouse gas emissions of realistic dietary
- choices in Denmark: The carbon footprint and nutritional value of dairy products. Food
  Nutr. Res. 58. https://doi.org/10.3402/fnr.v58.20687
- Wernet, G., Bauer, C., Steubing, B., Reinhard, J., Moreno-Ruiz, E., Weidema, B., 2016. The
  ecoinvent database version 3 (part I): overview and methodology. Int. J. Life Cycle

762 Assess. 21, 1218–1230. https://doi.org/10.1007/s11367-016-1087-8

- WHO, 2012. Promoting a healthy diet for the WHO Eastern Mediterranean Region: user-friendlyguide.
- 765 WHO, 2003. Diet, nutrition and the prevention of chronic diseases, World Health Organization
- technical report series. https://doi.org/ISBN 92 4 120916 X ISSN 0512-3054 (NLM
  classification: QU 145)
- 768 Wilson, N., Nghiem, N., Ni Mhurchu, C., Eyles, H., Baker, M.G., Blakely, T., 2013. Foods and
- 769 Dietary Patterns That Are Healthy, Low-Cost, and Environmentally Sustainable: A Case
- 770 Study of Optimization Modeling for New Zealand. PLoS One 8.
- 771 https://doi.org/10.1371/journal.pone.0059648
- 772 Xunta de Galicia, 2013. Menús saludables ricos en fibra. Santiago de Compostela.
- 773 Xunta de Galicia, 2005. Ley 2/2005, do 18 de febreiro, de promoción e defensa da calidad
- alimentaria. Santiago de Compostela.