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Integrated evaluation of wine lees valorization to produce value-added products

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8 Abstract

The integrated evaluation of the valorization of wine lees to produce value-added 9 10 products was carried out in this study from a life-cycle perspective. The consumption of steam has been demonstrated as the main hot spot, reaching 85.7% of the impact on 11 Fossil Depletion and 85.3% on Climate Change. Bearing in mind that four different 12 value-added products are produced, a sensitivity analysis was carried out in order to 13 ascertain the influence of the functional unit and the allocation method on the 14 environmental outcomes. The performance of this system was compared to other 15 processes that produce antioxidants from different raw materials. These processes were 16 phycocyanin recovery from Spirulina platensis cyanobacterium, the production of the 17 red antioxidant astaxanthin by microalgae and the valorization of the macroalgae 18 Sargassum muticum. Wine lees valorization showed a better environmental profile 19 throughout the entire life cycle, due to the fact that most of the operations performed are 20 physical (solid/liquid separations, distillations, evaporations, etc.) and do not involve a 21 large consumption of electricity or chemicals. However, there is still room for 22 improvement, and future research should focus on optimizing the extraction of 23 antioxidants from wine lees using two-stages aqueous systems, ultrasonic or microwave 24

- assisted extraction, in the pursuit of better performance and lower environmental
- 26 impact.
- 27 Keywords: Wine lees valorization; Biorefinery; Life Cycle Inventory; Environmental
- 28 assessment; Value-added products.

29 Abbreviations

- LCA Life Cycle Assessment
- LCI Life Cycle Inventory
- FU Functional Unit
- SS Subsystem
- ISO International Organization for Standardization
- CC Climate Change
- OD Ozone Layer Depletion
- TA Terrestrial Acidification
- FE Freshwater Eutrophication
- ME Marine Eutrophication
- HT Human Toxicity
- POF Photochemical Oxidant Formation
- TET Terrestrial Ecotoxicity
- FET Freshwater Ecotoxicity
- MET Marine Ecotoxicity
- FD Fossil Depletion

30

32 **1. Introduction**

The primary sector is one of the largest industrial sectors in terms of resources and energy consumption (Roy et al., 2009) and has therefore been identified as one of the main actors in climate change, with 30% of total greenhouse gas (GHG) emissions. With the aim of developing a competitive low-carbon economy (European Comission, 2011), action plans and measures have been put forward to reduce the current level of GHG emissions throughout the food supply chain.

Beyond environmental indicators related to climate change, the low efficiency in food 39 40 production and processing has been recognized as a major issue. According to a recent FAO report (FAO, 2011), up to 33% of the food produced for human consumption is 41 lost or wasted along the supply chain. Not only is food wasted, but the resources needed 42 43 to produce it, such as water, energy, chemicals or fuels used, are also misused, representing a value of nearly 1.3 billion tons per year. In addition, according to data 44 estimated by United Nations (2017), in relation to population growth, world's 45 population is expected to increase to 8.6 billion by 2030, which will probably aggravate 46 47 the problem of food waste.

Biorefinery is a clear example of the change of paradigm in the framework of 48 sustainable development. Biorefineries represent the transition from oil refineries to 49 50 sustainable systems based on the valorisation of waste flows with the aim of producing value-added compounds such as biogas, electricity, chemical products or biomaterials 51 (Cherubini, 2010). Following this principle, food waste can be valorised through 52 53 different technologies, such as anaerobic digestion to generate bio-hydrogen and biomethane (Algapani et al., 2019), co-composting with other types of organic waste for 54 the production of bioenergy and fertilizers (Vico et al., 2018), conversion into animal 55

feed (Makkar, 2018) or recovery of sugars, organic acids, pigments, fibre, proteins, oils,
antioxidants and vitamins from food waste (García-Herrera et al., 2010).

Within the primary sector, wine production is becoming increasingly important as a 58 59 symbol of a quality product, with a growing influence on exports from producing countries. World wine production is dominated by Italy, France and Spain, which 60 together account for 48% of total production (OIV, 2018). From this perspective, the 61 wine sector can be considered a reference point in the EU strategy within the primary 62 production sector (Christ and Burritt, 2013). The winemaking process comprises a 63 numerous sequence of activities (Escribano-Viana et al., 2018), from the cultivation of 64 the vine, the harvest, the process of fermentation and maturation of the wine in the 65 66 winery to the management of waste generated at each stage of the process. The main 67 effluents from the wine sector are wastewater and organic solid waste (Ruggieri et al., 2009). Solid organic waste includes wine lees and grape pomace, which is composed of 68 stalks, skins and seeds. In general, the volume of waste generated is around 20-30% of 69 70 total wine production, which can be considered a meaningful percentage (Zabaniotou et 71 al., 2018).

In order to improve the overall efficiency and environmental impact of the winemaking process, actions have been identified that allow for the minimization, management and effective recovery of waste streams from a circular economy perspective (Musee et al., 2007). The approach of circular economy grants and consolidates the value of each element of the productive chain and deepens the awareness of action, essential to achieve a real change towards sustainability, with efficient use of resources and valorization of by-products and wastes.

In this framework, Gullón et al. (2018) analysed the environmental impacts of different 79 80 routes for the valorisation of vine shoots. Poveda et al. (2018) proposed revaluing the by-products of winemaking, grape pomace and stems as a source of natural 81 preservatives. Nayak et al. (2018) developed a method for recovering polyphenols from 82 exhausted grape pomace through activated carbon. Zhang et al. (2017) compared two 83 methods for the valorisation of grape pomace, which is the major component of wine 84 production waste, to add value to economic and environmental balance of the overall 85 process. These processes were combustion to produce electricity, and pyrolysis to 86 produce bio-gas, bio-oil and bio-char. The detailed analysis of the wine lees fraction 87 88 presents high concentrations of macronutrients and polyphenols and low concentrations of micronutrients and heavy metals (Devesa-Rey et al., 2011). Moreover, the presence 89 of other compounds of potential interest such as polyphenols and antioxidants identifies 90 91 this stream as an ideal candidate to be valorized (Dimou et al., 2015; Kopsahelis et al., 2018; Martinez et al., 2016). 92

These valorization options can be evaluated according to their environmental
performance. Several environmental assessment methods can be found in literature, as
material flow analysis, energy balance, exergy analysis and life cycle assessment
(Vandermeersch et al., 2014). The latter seems to be the best choice since it can
consider the full life cycle (cradle-to-gate or cradle-to-grave).

The objective of this work seeks to delve into the different strategies for valuing winemaking-derived waste, proposing the identification of critical points in the environmental profile of the process under study, prior to its development and marketing of the products obtained. The function of the system under study is the use of wine lees to produce some bio-based products with marketable added value. It is therefore a question of identifying the most suitable process alternatives in a system

under development, which may suffer from limitations in terms of data availability, but
it may also make it possible to establish a roadmap in the search for viable options from
a techno-economic and environmental point of view.

107 **2. Materials and methods**

108 2.1. Definition of scope and system boundaries

The functional unit is a measure of the function of the system studied and provides a 109 reference to which the inputs and outputs can be related (ISO 14040; 14044). The 110 selection of the functional unit in biorefinery studies is made on the basis of three 111 112 possible options, i.e. total flow of raw materials, quantity of a single target product or the combination of products (Khoshnevisan et al., 2018). The FU considered in this 113 study was 1 tonne of wine lees processed in the winery facility. This feedstock-based 114 FU is consistent with the choice of other similar studies, in which a similar FU was 115 chosen because of their multiple-output nature (Lam et al., 2018; Vaskan et al., 2018). 116 117 The production plan was evaluated considering all the processes from the production of raw materials to the final products obtained from the wine lees, in such a way that the 118 processes in the winery are analyzed, mainly those associated with the production of 119 120 raw materials, electricity, fertilizers, chemical products and water, as well as the consumption of fuel used in the transport of materials. 121

122 The foreground system includes process units that are the direct object of the study.

123 This system is divided into two subsystems represented in Figure 1: SS1. Wine

124 production and SS2. Processing plant. Figure 1 shows the block diagram of the process,

identifying the system boundaries, the subsystems considered, and the main inputs and

126 outputs. No infrastructure process was considered in the study since the environmental

127 impacts of construction, installation, etc. have been considered negligible during the

useful life of the facility. This assumption is a common practice in other life cycle

assessment studies of different biorefineries (Jeswani et al., 2015; Karlsson et al., 2014).

130 As far as storage processes, it was considered unnecessary in this study, since only a

small warehouse within the facilities is needed to store the wine lees at room

temperature, so it is included within the infrastructure processes.

133 Subsystem 1 is divided into two sections: SS1.1. Viticulture comprises different

activities considered in the agricultural phase of wine production, including fertilization,

field operations or soil management and SS1.2. Vinification includes the processing of

136 grapes in the winery: wine production, bottling and packaging (Vázquez-Rowe et al.,

137 2012). Wine lees are a residue generated during this process, which is further processed

in Subsystem 2.



Figure 1. System boundaries of the wine lees biorefinery to produce value-added 140 141 products. Caption: T: Transport; EtOH: Bioethanol; CaT: Calcium tartrate; YC: Yeast cells; AntiOX: Antioxidants; WW: Wastewater; WWTP: Wastewater Treatment Plant 142 143 Subsystem 2 is the industrial process in which wine lees are valorized to produce four end-products, in this case bioethanol, calcium tartrate, an extract rich in antioxidants 144 and solid fraction rich in protein, which can be marketed with an economic return. The 145 146 description of the stages of the valorization process is described below. The first step is the separation of the liquid and solid fractions by centrifugation of the wine lees in a 147 disc centrifuge. Wine lees fed the facilities at room temperature (25°C) and with a 148 149 content of 62.9% (w/w) water, 5.7% (w/w) ethanol and 31.4% (w/w) solids (Dimou et al., 2016). It is possible to recover the residual fraction of ethanol present in the liquid 150 fraction by distillation at 100°C. In this step, the product stream of the bottom contains 151 mainly water and is sent to treatment. The solid fraction is then sent to an extraction 152 tank in which it is mixed with the ethanol recovered in the previous stage. The recovery 153 154 of ethanol implies a significant reduction in the cost associated with this stage since it is available in the original wine lees (Dimou et al., 2016). A new solid-liquid separation is 155 then carried out in a disc centrifuge so that the liquid fraction is fed to another 156 157 distillation column in which the antioxidant-rich extract is separated as a bottom product. The removal of water from this stream takes place in an evaporator and in a 158 spray dryer which works with high pressure steam. The solid stream from the second 159 solid-liquid separation is mixed with HCl and the tartrate salt (insoluble in water) is 160 161 transformed into tartaric acid (water-soluble). This compound is mixed with calcium 162 carbonate (CaCO₃) and calcium chloride (CaCl₂), transforming water-soluble tartaric acid into water-insoluble calcium tartrate. Calcium tartrate is obtained as a solid product 163

after the drying stage. Yeast cells are also obtained from the solid fraction, which can bemarketable as animal feed to obtain an economic return.

166 *2.2. Data acquisition*

The data used in this study come from complementary studies that consider all stages of 167 the life cycle of wine production, which includes the vine cultivation and the winery 168 169 processing stage (Subsystem 1) to the valorization of the different fractions of wine and 170 its by-products within the biorefinery (Subsystem 2) with a processing capacity of 500 kg/h of wine lees developed in Dimou et al. (2016). An environmental assessment was 171 172 carried out with a common perspective defined on the basis of an identical functional unit for both subsystems: 1 tonne of wine lees. The inventory data for the production of 173 all system inputs from background processes were taken from the Ecoinvent® database. 174 175 These inputs include the production of the different chemical products necessary for the extraction of calcium tartrate, the electricity consumed in the different stages, the 176 fertilizers for the vine cultivation and any other type of raw material or resource. An 177 average transport distance of 800 km was considered within continental Europe for 178 chemical products (Pérez-López et al., 2014b) and 100 km of average distance for the 179 transport of the wine lees from the winery to the processing plant were assumed (Hajjaji 180 et al., 2013). 181

182 *2.3. Life cycle inventory*

The life cycle inventory is the compilation of the data set for evaluation and involves the collection of quantitative input/output data for the system. In this study, a cradle-togate approach has been proposed, taking into account all stages, from the cultivation of the vine to the activities of the winery and the valorisation of the wine lees. Considering that the economic allocation reflects the function and objective of the production

process, which is obviously the marketing of a main product: wine and the associated 188 189 economic revenues, mass or energy-based allocations were discarded (Rugani et al., 2013). Since wine lees are a residue derived from the production of wine, it is necessary 190 to make a cost estimate as a requirement for an economic allocation that allows the 191 environmental impacts corresponding to each fraction to be assessed. Thus, the selling 192 price of a bottle of wine produced after SS1 was compared with the potential selling 193 194 price of the various products obtained from the valorisation route (extract rich in antioxidants, ethanol, yeast cells and calcium tartrate). A market prize of €4 for a 750 195 mL bottle of wine with designation of origin was considered. Considering that the 196 197 production of a bottle of this wine generates 11.48 ml of wine lees, the sale of these products generates a profit of €0.022 and when comparing the contributions of both 198 products in terms of their market value, the economic allocation factors are 99.45% for 199 200 wine and 0.55% for wine lees.

201 2.4. Life cycle impact assessment: methodology

SimaPro 8.5.2 (PRé Consultants, 2017) has been the software used for the 202 implementation of the Life Cycle Inventory. To analyze the inputs and outputs of the 203 LCI, the Classification and Characterization guidelines defined by ISO were followed. 204 In this phase, in order to translate the long list of life cycle inventory results into a small 205 number of environmental impact indicators, the ReCiPe Midpoint methodology was 206 207 used. This method provides a common framework in which both mid-point and endpoint indicators can be used (Goedkoop et al., 2009). The characterization factors 208 reported by the ReCiPe Midpoint methodology were applied, and the potential impact 209 categories considered were: Climate Change, Ozone Layer Depletion, Terrestrial 210 Acidification, Freshwater Eutrophication, Marine Eutrophication, Human Toxicity, 211

Photochemical Oxidant Formation, Terrestrial Ecotoxicity, Freshwater Ecotoxicity,
Marine Ecotoxicity and Fossil Fuel Depletion.

214 **3. Results and discussion**

215 3.1. Quantitative analysis of inputs and outputs to the overall system

Life Cycle Assessment is a structured and comprehensive method for quantifying
material and energy flows and associated impacts on the life cycle of products (i.e.
goods and services). Within the LCA methodology, Life Cycle Inventory is a
mandatory stage and the availability of LCI data is often the greatest barrier to
completing an LCA study.

LCI shows that the consumption of different pesticides is very high in the viticulture 221 222 stage, which is in line with other studies on different crops (Caldeira et al., 2018; Liang et al., 2019). In addition, the consumption of organic fertilizers is also remarkable, 223 leading to relevant nitrate emissions, so the impact on the eutrophication categories is 224 expected to be considerable. As for the winemaking phase, the main inputs are 225 226 electricity and chemicals such as NaOH and SO₂. In this system, CO₂ emissions due to 227 wine fermentation were calculated, but excluded from the environmental assessment, as 228 biogenic CO₂ at this stage was not taken into account. LCI has also allowed to quantify 229 the main inputs and outputs of the wine lees valorization system, highlighting the 230 consumption of High-Pressure Steam and the production of four value-added products. The main product obtained in this process is the extract rich in antioxidants. This 231 232 extract, as calculated in Dimou et al. (2016), has a lower total polyphenol content than other studies (26.1 mg of gallic acid equivalent per g of wine lees). Therefore, this 233 extraction is not optimized, and the product could be purer. However, wine lees 234

valorization can be considered as a sustainable and environmentally friendly process 235 236 due to the fact that most of the operations carried out are physical, such as solid-liquid extraction or centrifugation, while other studies involve complex operations, such as 237 238 ultrasound or microwave assisted extraction (Castro-López et al., 2017; Mohammadpour et al., 2019; Pereira et al., 2017). Since most of the treatments carried 239 240 out at this stage are physical, the use of chemical products is low, as only 174 kg per ton 241 of treated wine lees are consumed. Further details of Life Cycle Inventory are provided 242 in electronic Supplementary data.

243 *3.2. Analysis of processing plant energy requirements*

Electricity and steam consumption reported in supplementary material correspond to the operation of the plant and is broken down by equipment in Table 1. Although the total electricity consumption is not too high (125 kWh per FU), more than 95% of this consumption corresponds to disc centrifuges, which are used to separate the solid phase

from the liquid phase. This separation process is essential to obtain value-added

249 products, as most of the treatments carried out are physical, such as stages of

distillation, separation, evaporation and spray drying with compressed hot air.

251 Steam consumption is very high, more than 5 tons of steam per ton of wine lees treated.

In the process, steam is used in distillation columns (units E-102 and E-104) while in

the unit E-105, steam is used to evaporate water from the antioxidant-rich stream.

Finally, in the E-106 and E-107 exchangers, steam is used to heat the compressed air

that will be used in the spray dryers to remove the remaining water from the calcium

- tartrate and the antioxidant-rich extract. This consumption, together with the fact that
- obtaining steam is an activity with high energy requirements (Nieuwlaar et al., 2015),

- makes it possible to anticipate that the environmental impact derived from the use of
- steam will be high.

Table 1. Analysis of the electricity, steam and cooling water consumption of Subsystem

Operation	Electricity	Steam	Cooling water
Dise contrifuçõe	<u>consumption (kwn)</u>	consumption (kg)	(m ³)
Disc centringes	119.19		
CF-101	29.80		
CF-102	29.80		
CF-103	29.80		
CF-104	29.79		
Mixing tanks	4.68		
V-101	2.60		
V-102	1.20		
V-103	0.88		
Blower	6.06		
C-101	4.06		
C-102	2.00		
Heat exchangers		5104.84	224.46
E-102		819.44	
E-104		3994.44	
E-105		238.19	
E-106		39.58	
E-107		13.19	
E-101			33.61
E-103			190.85

262 2. Processing plant per FU (1 tonne of wine lees)

263

264 3.3. Environmental characterization of wine lees valorization process

The environmental impacts expressed through different impact categories are presented 265 in the characterization stage, these results are summarized in Table 2. Figure 2 presents 266 the contributions relative to the environmental burdens of each subsystem considered, 267 identifying the most problematic ones with the greatest environmental impacts in the 268 process of valorisation of wine lees. Specifically, in Subsystem 1, wine production is 269 270 the main contributor to the categories of MET, FET, POF, HT, ME and TA. In MET and TET, SS1 represents almost 90% of the total environmental impact, which is related 271 to the treatment of solid waste produced during winemaking. SS1.1 includes different 272

273	activities of the agricultural phase of wine production such as soil management, field
274	operations or fertilisation. This phase is clearly the hotspot in ME, POF and TA
275	categories with remarkable contributions of 67.9%, 56.8% and 52.7% respectively. It
276	should be noted here that the use of fossil fuels for the operation of machinery such as
277	broadcasters and rotary cultivators and the use of compost for fertilization cause
278	emissions of nitrogen oxides that affect POF category. As for the environmental impacts
279	of ME, the application of fertilizers in the agricultural phase of wine production
280	involved the greatest relative impact of this category. These results are in accordance
281	with the results obtained in another study in which an LCA of red wine production from
282	Catalonia was performed (Meneses et al., 2016). This study showed that the vinification
283	process has a better environmental profile than the viticulture stage. Regarding HT
284	category, there are no major differences between the two considered subsystems
285	because the main contributors to the environmental impact in this category are the
286	emissions of heavy metals into the atmosphere derived from the consumption of fossil
287	fuels for the operation of equipment, transport and steam production.

Table 2. Impact assessment results associated with the valorization of 1 tonne of winelees

Impact category	Unit	SS1.1	SS1.2	SS2	Total
СС	kg CO ₂ eq	$8.79 \cdot 10^2$	$3.26 \cdot 10^2$	$1.33 \cdot 10^{3}$	$2.54 \cdot 10^3$
OD	kg CFC-11 eq	$1.14 \cdot 10^{-4}$	$2.09 \cdot 10^{-5}$	$1.52 \cdot 10^{-4}$	$2.87 \cdot 10^{-4}$
ТА	kg SO ₂ eq	6.52	1.00	4.85	12.37
FE	kg P eq	0.15	0.05	0.22	0.42
ME	kg N eq	2.52	1.05	0.14	3.72
НТ	kg 1,4-DB eq	$1.85 \cdot 10^2$	$1.86 \cdot 10^2$	$2.43 \cdot 10^2$	$6.14 \cdot 10^2$
POF	kg NMVOC	4.25	0.68	2.56	7.48
TET	kg 1,4-DB eq	$3.67 \cdot 10^{-2}$	8.61·10 ⁻³	$1.08 \cdot 10^{-1}$	$1.53 \cdot 10^{-1}$
FET	kg 1,4-DB eq	5.36	57.81	7.73	70.88
MET	kg 1,4-DB eq	5.25	49.77	7.01	62.03
FD	kg oil eq	$1.10 \cdot 10^2$	46.85	$4.35 \cdot 10^2$	$5.92 \cdot 10^2$

291 In relation to subsystem 2, the processing plant is the main contributor in FD and TET impact categories. In the FD category it reaches a maximum contribution of 75% 292 293 associated with the use of natural gas for steam production. In the TET category, it 294 accounts for 70% of the impact due to emissions of heavy metals into the atmosphere linked to the use of fossil fuels, either in transport or in the production of high 295 temperature steam. However, regarding FE, OD and CC categories, the difference 296 between the two subsystems is minimal. Focusing on CC, direct emissions into the 297 298 atmosphere associated with fermentation process are quantified in the winemaking process. However, direct CO₂ emissions from SS1 should not be considered as fossil 299 300 carbon, but as biogenic CO₂. If these emissions were considered as fossil carbon, SS1 301 contribution to the CC category would increase to 48.3%, which would mean a global value of 2570 kg of CO₂ eq per ton of wine lees. 302





Figure 2. Relative contributions (in %) by subsystems in the valorization of wine lees

Taking into account the economic allocation factors calculated in Section 2.3, the 305 306 characterisation results of the wine lees biorefinery were calculated (Figure 3). According to the results obtained, wine lees can be classified as a winery waste that is 307 not assigned an associated environmental impact. On the contrary, the use of high 308 309 temperature steam can be considered as the most burdensome component. In this sense, the other categories include FD (85.7%), CC (85.3%), TA (79.4%) and POF (76.7%), 310 311 all of which are associated with the high consumption of fossil fuels for the production of steam as well as the associated emissions of GHG, SO₂ and NOx. 312



313

Figure 3. Relative contributions (in %) by component in the overall biorefinery
production process

316 Electricity has a low impact in almost all impact categories, which is attributed to a low-

moderate consumption (Table 1). Only FET shows a remarkable value of 24.2%. HCl

becomes the second most contributing element with an average of 16.7%, standing out

- in the categories of toxicity, either Human (HT) or Ecosystem (TET, MET and FET);
- 320 due to emissions of heavy metals associated with the production of HCl, H₂SO₄ and Cl₂
- 321 compounds. In contrast, the impact of CaCl₂ and CaCO₃ are of much lesser importance

with average contributions of 4.7% and 0.03% respectively. Transport presents a 322 323 considerably uniform distribution of environmental impacts in all categories, with contributions almost always lower than 7%, except in the case of the TET category, in 324 325 which a relative contribution of around 27.2% is reached. This is because this category gives more weight to emissions to air and soil (which are abundant in transport 326 processes); in contrast, other toxicity categories such as MET and FET are more 327 328 dependent on water emissions, with negligible impacts. Wastewater treatment contributions are practically negligible except for a percentage of 8.2% in ME category, 329 due to nitrogen compounds directly discharged into the water. 330

331 3.4. Sensitivity analysis. Effect of the selection of the Functional Unit and allocation
332 factors on the environmental profile

The results obtained are based on a FU that focuses on the amount of biomass valorized 333 rather than on the products obtained. In fact, the choice of FU is a critical point in any 334 LCA study as it is a subjective action that must be consistent with the objectives of the 335 study. In this case the function of the system is the treatment of a waste, but this system 336 allows to obtain four different products. The extract rich in antioxidants is the product 337 of greatest interest for its potential applications in the food industry, cosmetics and 338 pharmaceutical industry (Szabo et al., 2018). Therefore, the new FU was selected as 1 339 kg of antioxidant-rich extract obtained. 340

In addition to the choice of the FU, the allocation of impacts is fundamental, especially in multi-production processes such as the one studied here. If a mass or economic allocation is followed, the impacts assigned to each product are different. Therefore, the factors of mass and economic allocation have been calculated by the quantity produced

- of each element and its potential market price obtained from various scientific
- publications. A summary of the considered allocation factors is given in Table 3.
- 347 **Table 3.** Computation of the mass and economic allocation factors for SS2. Processing
- 348 plant

Product	Production ^a	Market price	Mass allocation	Economic allocaion
Bioethanol	28.22 kg	0.67 €kg ^b	8.4%	1.3%
Calcium tartrate	58.50 kg	4.41 €kg ^c	17.5%	17.3%
Yeast cells	241.00 kg	0.88 €kg ^c	72.0%	14.2%
Antioxidants	6.78 kg	147.67 €kg ^d	2.0%	67.6%

349	^a Results per	Functional	unit (1	tonne of wine l	ees)
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350 ^b (Joelsson et al., 2016)

351 ^c (Dimou et al., 2016)
 352 ^d (Vieira et al., 2013)

352 353

These studies show how the market prices of each of the products obtained in the 354 355 biorefinery varies and how the mass and economic allocations factors vary accordingly. Joelsson et al. (2016) conducted a research on the production of biogas and bioethanol 356 from wheat straw impregnated with acetic acid. In addition to the experimental study, 357 they also performed a techno-economic evaluation. In this paper a bioethanol price of 358 359 0.57-0.68 €kg is estimated. In Dimou et al. (2016) a techno-economic evaluation of the 360 biorefinery which is environmentally evaluated in the present study was carried out and 361 a market cost for yeast cells of 1 \$/kg was estimated. This price can be converted to 362 euros and is equivalent to 0.88 €kg. Finally, in Vieira et al (2013) a chemical and 363 economic evaluation of antioxidants extracted from pulp of Euterpe edulis was carried out. The manufacturing costs of the crude extracts obtained in this paper were 165.34 364 \$/kg, which is equivalent to 147.67 €kg. These manufacturing costs were assumed as 365 366 the market cost of this extract in order to obtain an economic return. The effect of the alternative FU and allocation factors on the environmental profiles is 367

The effect of the alternative FO and anocation factors on the environmental promes is

shown in Table 4. When allocation factors (mass or economic) are used, a decrease in

environmental impact is always observed, as the environmental impact is distributed
among the different products. As the amount of the extract rich in antioxidants is low, in
the case of mass allocation its environmental impact is small. However, this is not
realistic, as it is the product of greatest interest, so, as shown in Table 4, the economic
allocation distributes the environmental impacts among the products more accurately.
To the production of 1 kg of extract rich in antioxidants is assigned, among others,
251.3 kg of CO₂ eq in CC category and 58.6 kg of oil eq in FD category.

Table 4. Environmental results for each impact category considering only the

377 production of 1 kg of antioxidants-rich extract with no allocation, mass and economic

378 allocation factors

Impact category	Units	No allocation	Mass allocation	Economic allocation
CC	kg CO ₂ eq	374.1	7.6	251.3
	kg CFC-11	$4.2 \cdot 10^{-5}$	8.6·10 ⁻⁷	$2.8 \cdot 10^{-5}$
OD	eq			
ТА	kg SO ₂ eq	1.8	$3.7 \cdot 10^{-2}$	1.2
FE	kg P eq	$6.2 \cdot 10^{-2}$	$1.3 \cdot 10^{-3}$	$4.2 \cdot 10^{-2}$
ME	kg N eq	$5.5 \cdot 10^{-1}$	$1.1 \cdot 10^{-2}$	$3.7 \cdot 10^{-1}$
HT	kg 1,4-DB eq	90.6	1.8	60.9
POF	kg NMVOC	1.1	$2.2 \cdot 10^{-2}$	$7.4 \cdot 10^{-1}$
TET	kg 1,4-DB eq	$2.3 \cdot 10^{-2}$	$4.6 \cdot 10^{-4}$	$1.5 \cdot 10^{-2}$
FET	kg 1,4-DB eq	10.4	$2.1 \cdot 10^{-1}$	7.0
MET	kg 1,4-DB eq	9.1	$1.9 \cdot 10^{-1}$	6,15
FD	kg oil eq	87.3	1.8	58.6

379

380 *3.5. Comparative environmental performance of different methods to obtain*

381 *antioxidant-rich extracts*

382 A comparison was made with some processes published in the scientific literature on

the basis of an identical FU: 1 kg of extract rich in antioxidants and evaluation

methodology: CLM. Pérez-López et al. (2014a) evaluated the environmental 384 385 performance of several Sargassum muticum macroalgae valorization strategies. In this study it was considered that the combined extraction of antioxidants and alginates 386 stands out as the most sustainable scenario. Pérez-López et al. (2014b) conducted a life 387 cycle assessment of astaxanthin production on a laboratory and pilot scale. In Papadaki 388 et al. (2017) an evaluation of the life cycle of the recovery of phycocyanin from 389 390 Spirulina platensis cyanobacterium is performed. This study compares six different methods based on Ultrasound-Assisted Extraction to recover phycocyanin. In order to 391 simplify the comparative study, the results are scaled to 100 and represented in Figure 392 393 4.



394

Figure 4. Relative environmental profile of the compared valorization scenarios with

the process published in (Pérez-López et al., 2014a) as baseline (index = 100)

397 According to the results depicted in Figure 4, the production of antioxidants through the

valorization of wine lees would be the most appropriate route in all the impact

399 categories studied. Except in the case of EP, where the recovery of phycocyanin from a

400 cyanobacterium is the process with the best environmental profile, since the impacts of

EP are directly related to emissions linked to fertilization in the viticulture phase. The 401 402 production of astaxanthin from microalgae cultivated in photobioreactors on a pilot scale reported the worst environmental profile, with results ranging from 37% worse 403 than the reference (index = 100) in FEP category up to 180% worse in ODP impact 404 405 category. If the present study is compared with the process published by Pérez-López et 406 al. (2014a), the production of antioxidants from the treatment of wine lees presents an 407 environmental profile that is, on average, 75% better in all impact categories. A maximum improvement rate of 92% is reached in the FEP impact category. However, in 408 409 order to obtain 1 kg of extract rich in antioxidants, it is necessary to treat almost 148 kg 410 of wine lees, while in order to obtain this same amount of extract from the valorization of Sargassum muticum, it is only necessary to process 1.5 kg of biomass. In particular, 411 the treatment of wine lees has better environmental results because most of the 412 413 operations performed are physical (solid/liquid separations, distillations, evaporations, etc.) and do not involve the large consumption of electricity or chemicals. There is only 414 415 one determinant consumption in the system, the high temperature steam, while in the 416 rest of the comparative studies, the consumption of electricity and chemicals is relatively high. Comparative analysis has allowed us to compare the valorization of 417 418 wine lees with others related to the production of antioxidants from different raw materials published in the scientific literature. The valorization of wine lees presents the 419 best environmental profile in almost all compared categories. 420

421 4. Conclusions

422 Nowadays special interest is being paid into the valorization of different wastes in order

423 to reduce raw materials consumption. Thus, it has been shown that the integral

424 valorization of wine lees is a very attractive process to produce value-added products.

425 Winemaking is a process with a high demand for chemicals, which leads to a high

environmental impact. Therefore, this system causes 50% of the total impact of wine 426 427 lees valorization. However, on a comparative level with other processes in which an antioxidant-rich extract can be obtained, it has been demonstrated that the valorization 428 429 of wine lees presents the best environmental profile throughout the entire life cycle in almost all the impact categories studied. However, steam consumption has proven to be 430 431 an important hotspot in the process, so it will be necessary to reduce this consumption 432 in the future. In order to achieve this objective, other residues from the winery, such as grape stalks, could be used as raw material to obtain high temperature steam in a boiler. 433

434 Wine waste biorefining is an appropriate way of obtaining products from waste

435 according to the principles of the Circular Economy, where waste is converted into new

436 raw materials. This work shows that the LCA methodology is a useful tool for assessing

the environmental impact of wine lees treatment in order to obtain value-added

438 products. The results of this work should be considered in order to develop a more

439 sustainable way of obtaining an antioxidant-rich extract from agricultural residues.

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445 **References**

Algapani, D.E., Qiao, W., Ricci, M., Bianchi, D., Wandera, S.M., Adani, F., Dong, R.,
2019. Bio-hydrogen and bio-methane production from food waste in a two-stage
anaerobic digestion process with digestate recirculation. Renew. Energy 130,

- 449 1108–1115. https://doi.org/10.1016/j.renene.2018.08.079
- 450 Caldeira, C., Quinteiro, P., Castanheira, E., Boulay, A., Dias, A.C., Arroja, L., Freire,
- 451 F., 2018. Water footprint profile of crop-based vegetable oils and waste cooking
- 452 oil: Comparing two water scarcity footprint methods. J. Clean. Prod. 195, 1190–
- 453 1202. https://doi.org/10.1016/j.jclepro.2018.05.221
- 454 Castro-López, C., Ventura-Sobrevilla, J.M., González-Hernández, M.D., Rojas, R.,
- 455 Ascacio-Valdés, J.A., Aguilar, C.N., Martínez-Ávila, G.C.G., 2017. Impact of
- 456 extraction techniques on antioxidant capacities and phytochemical composition of
- 457 polyphenol-rich extracts. Food Chem. 237, 1139–1148.
- 458 https://doi.org/10.1016/j.foodchem.2017.06.032
- 459 Cherubini, F., 2010. The biorefinery concept: Using biomass instead of oil for
- 460 producing energy and chemicals. Energy Convers. Manag. 51, 1412–1421.
- 461 https://doi.org/10.1016/j.enconman.2010.01.015
- 462 Christ, K.L., Burritt, R.L., 2013. Critical environmental concerns in wine production:
- 463 An integrative review. J. Clean. Prod. 53, 232–242.
- 464 https://doi.org/10.1016/j.jclepro.2013.04.007
- 465 Devesa-Rey, R., Vecino, X., Varela-Alende, J.L., Barral, M.T., Cruz, J.M., Moldes,
- 466 A.B., 2011. Valorization of winery waste vs. the costs of not recycling. Waste
- 467 Manag. 31, 2327–2335. https://doi.org/10.1016/j.wasman.2011.06.001
- 468 Dimou, C., Kopsahelis, N., Papadaki, A., Papanikolaou, S., Kookos, I.K., Mandala, I.,
- 469 Koutinas, A.A., 2015. Wine lees valorization: Biorefinery development including
- 470 production of a generic fermentation feedstock employed for poly(3-
- 471 hydroxybutyrate) synthesis. Food Res. Int. 73, 81–87.
- 472 https://doi.org/10.1016/j.foodres.2015.02.020

473	Dimou, C., Vlysidis, A., Kopsahelis, N., Papanikolaou, S., Koutinas, A.A., Kookos,
474	I.K., 2016. Techno-economic evaluation of wine lees refining for the production of
475	value-added products. Biochem. Eng. J. 116, 157–165.
476	https://doi.org/10.1016/j.bej.2016.09.004
477	Escribano-Viana, R., Portu, J., Garijo, P., Gutiérrez, A.R., Santamaría, P., López-
478	Alfaro, I., López, R., González-Arenzana, L., 2018. Evaluating a preventive
479	biological control agent applied on grapevines against Botrytis cinerea and its
480	influence on winemaking. J. Sci. Food Agric. 98, 4517–4526.
481	https://doi.org/10.1002/jsfa.8977
482	European Comission, 2011. A Roadmap for moving to a competitive low carbon
483	economy in 2050.
484	FAO, 2011. Global food losses and food waste - Extent, causes and prevention. Rome.
485	García-Herrera, P., Sánchez-Mata, M.C., Cámara, M., 2010. Nutritional characterization
486	of tomato fiber as a useful ingredient for food industry. Innov. Food Sci. Emerg.
487	Technol. 11, 707–711. https://doi.org/10.1016/j.ifset.2010.07.005
488	Goedkoop, M., Heijungs, R., Huijbrets, M., de Schryver, A., Struijs, J., Van Zelm, R.,
489	2009. ReCiPe 2008: A life cycle impact assessment method which comprises
490	harmonised category indicators at the midpoint and the endpoint level. Report I:
491	Characterisation.
492	Gullón, P., Gullón, B., Dávila, I., Labidi, J., González-García, S., 2018. Comparative
493	environmental Life Cycle Assessment of integral revalorization of vine shoots
494	from a biorefinery perspective. Sci. Total Environ. 624, 225–240.
495	https://doi.org/10.1016/j.scitotenv.2017.12.036

496	Hajjaji, N.,	Pons, M.N.	, Renaudin,	V., H	louas, A.,	, 2013. (Com	parative	life cycle	
									2	

- 497 assessment of eight alternatives for hydrogen production from renewable and fossil
 498 feedstock. J. Clean. Prod. 44, 177–189.
- 499 https://doi.org/10.1016/j.jclepro.2012.11.043
- 500 ISO, 2006a. ISO 14040 Environmental Management Life Cycle Assessment -
- 501 Principles and Framework. International Organization for Standardization.
- 502 ISO, 2006b. ISO 14044 Environmental Management Life Cycle Assessment -
- 503 Requirements and Guidelines. International Organization for Standardization.
- Jeswani, H.K., Falano, T., Azapagic, A., 2015. Life cycle environmental sustainability

of lignocellulosic ethanol produced in integrated thermo-chemical biorefineries.

506 Biofuels, Bioprod. Biorefining 9, 661–676. https://doi.org/10.1002/bbb

507 Joelsson, E., Dienes, D., Kovacs, K., Galbe, M., Wallberg, O., 2016. Combined

508 production of biogas and ethanol at high solids loading from wheat straw

impregnated with acetic acid: experimental study and techno-economic evaluation.

510 Sustain. Chem. Process. 4, 14. https://doi.org/10.1186/s40508-016-0058-5

511 Karlsson, H., Börjesson, P., Hansson, P.A., Ahlgren, S., 2014. Ethanol production in

512 biorefineries using lignocellulosic feedstock - GHG performance, energy balance

- and implications of life cycle calculation methodology. J. Clean. Prod. 83, 420–
- 514 427. https://doi.org/10.1016/j.jclepro.2014.07.029

515 Khoshnevisan, B., Rafiee, S., Tabatabaei, M., Ghanavati, H., Mohtasebi, S.S., Rahimi,

- 516 V., Shafiei, M., Angelidaki, I., Karimi, K., 2018. Life cycle assessment of castor-
- 517 based biorefinery: a well to wheel LCA. Int. J. Life Cycle Assess. 23, 1788–1805.
- 518 https://doi.org/10.1007/s11367-017-1383-y

519	Kopsahelis, N., Dimou, C., Papadaki, A., Xenopoulos, E., Kyraleou, M., Kallithraka, S.,
520	Kotseridis, Y., Papanikolaou, S., Koutinas, A.A., 2018. Refining of wine lees and
521	cheese whey for the production of microbial oil, polyphenol-rich extracts and
522	value-added co-products. J. Chem. Technol. Biotechnol. 93, 257-268.
523	https://doi.org/10.1002/jctb.5348
524	Lam, C.M., Yu, I.K.M., Hsu, S.C., Tsang, D.C.W., 2018. Life-cycle assessment on food
525	waste valorisation to value-added products. J. Clean. Prod. 199, 840-848.
526	https://doi.org/10.1016/j.jclepro.2018.07.199
527	Liang, L., Wang, Y., Ridoutt, B.G., Lal, R., Wang, D., Wu, W., Wang, L., Zhao, G.,
528	2019. Agricultural subsidies assessment of cropping system from environmental
529	and economic perspectives in North China based on LCA. Ecol. Indic. 96, 351-
530	360. https://doi.org/10.1016/j.ecolind.2018.09.017
531	Makkar, H.P.S., 2018. Review: Feed demand landscape and implications of food-not
532	feed strategy for food security and climate change. Animal 12, 1744–1754.
533	https://doi.org/10.1017/S175173111700324X
534	Martinez, G.A., Rebecchi, S., Decorti, D., Domingos, J.M.B., Natolino, A., Del Rio, D.,
535	Bertin, L., Da Porto, C., Fava, F., 2016. Towards multi-purpose biorefinery
536	platforms for the valorisation of red grape pomace: production of polyphenols,
537	volatile fatty acids, polyhydroxyalkanoates and biogas. Green Chem. 18, 261–270.
538	https://doi.org/10.1039/c5gc01558h
539	Meneses, M., Torres, C.M., Castells, F., 2016. Sensitivity analysis in a life cycle
540	assessment of an aged red wine production from Catalonia, Spain. Sci. Total
541	Environ. 562, 571–579. https://doi.org/10.1016/j.scitotenv.2016.04.083
542	Mohammadpour, H., Sadrameli, S.M., Eslami, F., Asoodeh, A., 2019. Optimization of

- 543 ultrasound-assisted extraction of Moringa peregrina oil with response surface
- 544 methodology and comparison with Soxhlet method. Ind. Crops Prod. 131, 106–

545 116. https://doi.org/10.1016/j.indcrop.2019.01.030

- 546 Musee, N., Lorenzen, L., Aldrich, C., 2007. Cellar waste minimization in the wine
- industry: a systems approach. J. Clean. Prod. 15, 417–431.
- 548 https://doi.org/10.1016/j.jclepro.2005.11.004
- 549 Nayak, A., Bhushan, B., Rodriguez-Turienzo, L., 2018. Recovery of polyphenols onto
- 550 porous carbons developed from exhausted grape pomace: A sustainable approach
- for the treatment of wine wastewaters. Water Res. 145, 741–756.
- 552 https://doi.org/10.1016/j.watres.2018.09.017
- 553 Nieuwlaar, E., Roes, A.L., Patel, M.K., 2015. Final energy requirements of steam for
- use in environmental life cycle assessment. J. Ind. Ecol. 20, 828–836.
- 555 https://doi.org/10.1111/jiec.12300
- 556 OIV, 2018. State of the Vitiviniculture World Market, April 2018, International
- 557 Organisation of Vine and Wine.
- 558 Papadaki, S., Kyriakopoulou, K., Tzovenis, I., Krokida, M., 2017. Environmental
- impact of phycocyanin recovery from Spirulina platensis cyanobacterium. Innov.
- 560 Food Sci. Emerg. Technol. 44, 217–223.
- 561 https://doi.org/10.1016/j.ifset.2017.02.014
- 562 Pereira, P., Cebola, M.J., Oliveira, M.C., Bernardo Gil, M.G., 2017. Antioxidant
- capacity and identification of bioactive compounds of Myrtus communis L. extract
- obtained by ultrasound-assisted extraction. J. Food Sci. Technol. 54, 4362–4369.
- 565 https://doi.org/10.1007/s13197-017-2907-y

- Pérez-López, P., Balboa, E.M., González-García, S., Domínguez, H., Feijoo, G., 566
- 567 Moreira, M.T., 2014a. Comparative environmental assessment of valorization
- strategies of the invasive macroalgae Sargassum muticum. Bioresour. Technol. 568
- 161, 137-148. https://doi.org/10.1016/j.biortech.2014.03.013 569
- Pérez-López, P., González-García, S., Jeffryes, C., Agathos, S.N., McHugh, E., Walsh, 570
- D., Murray, P., Moane, S., Feijoo, G., Moreira, M.T., 2014b. Life cycle assessment 571
- 572 of the production of the red antioxidant carotenoid astaxanthin by microalgae: from
- lab to pilot scale. J. Clean. Prod. 64, 332-344. 573
- 574 https://doi.org/10.1016/j.jclepro.2013.07.011
- Poveda, J.M., Loarce, L., Alarcón, M., Díaz-Maroto, M.C., Alañón, M.E., 2018. 575
- 576 Revalorization of winery by-products as source of natural preservatives obtained
- 577 by means of green extraction techniques. Ind. Crops Prod. 112, 617-625.
- https://doi.org/10.1016/j.indcrop.2017.12.063 578
- PRé Consultants, 2017. SimaPro Database Manual (No. Methods Library). The 579 Netherlands.
- 580
- Roy, P., Nei, D., Orikasa, T., Xu, Q., Okadome, H., Nakamura, N., Shiina, T., 2009. A 581
- 582 review of life cycle assessment (LCA) on some food products. J. Food Eng. 90, 1-10. https://doi.org/10.1016/j.jfoodeng.2008.06.016 583
- 584 Rugani, B., Vázquez-Rowe, I., Benedetto, G., Benetto, E., 2013. A comprehensive
- review of carbon footprint analysis as an extended environmental indicator in the 585
- wine sector. J. Clean. Prod. 54, 61–77. 586
- 587 https://doi.org/10.1016/j.jclepro.2013.04.036
- 588 Ruggieri, L., Cadena, E., Martínez-Blanco, J., Gasol, C.M., Rieradevall, J., Gabarrell,
- X., Gea, T., Sort, X., Sánchez, A., 2009. Recovery of organic wastes in the Spanish 589

590	wine industry. Technical, economic and environmental analyses of the composting
591	process. J. Clean. Prod. 17, 830-838. https://doi.org/10.1016/j.jclepro.2008.12.005
592	Szabo, K., Cătoi, AF., Vodnar, D.C., 2018. Bioactive compounds extracted from
593	tomato processing by-products as a source of valuable nutrients. Plant Foods Hum.
594	Nutr. 73, 268–277. https://doi.org/10.1007/s11130-018-0691-0
595	United Nations, 2017. World population prospects: The 2017 Revision, key findings
596	and advance tables.
597	Vandermeersch, T., Alvarenga, R.A.F., Ragaert, P., Dewulf, J., 2014. Environmental
598	sustainability assessment of food waste valorization options. Resour. Conserv.
599	Recycl. 87, 57-64. https://doi.org/10.1016/j.resconrec.2014.03.008
600	Vaskan, P., Pachón, E.R., Gnansounou, E., 2018. Techno-economic and life-cycle
601	assessments of biorefineries based on palm empty fruit bunches in Brazil. J. Clean.
602	Prod. 172, 3655–3668. https://doi.org/10.1016/j.jclepro.2017.07.218
603	Vázquez-Rowe, I., Villanueva-Rey, P., Moreira, M.T., Feijoo, G., 2012. Environmental
604	analysis of Ribeiro wine from a timeline perspective: Harvest year matters when
605	reporting environmental impacts. J. Environ. Manage. 98, 73-83.
606	https://doi.org/10.1016/j.jenvman.2011.12.009
607	Vico, A., Pérez-Murcia, M.D., Bustamante, M.A., Agulló, E., Marhuenda-Egea, F.C.,
608	Sáez, J.A., Paredes, C., Pérez-Espinosa, A., Moral, R., 2018. Valorization of date
609	palm (Phoenix dactylifera L.) pruning biomass by co-composting with urban and
610	agri-food sludge. J. Environ. Manage. 226, 408–415.
611	https://doi.org/10.1016/j.jenvman.2018.08.035
612	Vieira, G.S., Cavalcanti, R.N., Meireles, M.A.A., Hubinger, M.D., 2013. Chemical and

613	economic evaluation of	natural antioxidant ext	tracts obtained by u	ultrasound-assisted
-----	------------------------	-------------------------	----------------------	---------------------

and agitated bed extraction from jussara pulp (Euterpe edulis). J. Food Eng. 119,

615 196–204. https://doi.org/10.1016/j.jfoodeng.2013.05.030

- 616 Zabaniotou, A., Kamaterou, P., Pavlou, A., Panayiotou, C., 2018. Sustainable
- bioeconomy transitions: Targeting value capture by integrating pyrolysis in a
- 618 winery waste biorefinery. J. Clean. Prod. 172, 3387–3397.
- 619 https://doi.org/10.1016/j.jclepro.2017.11.077
- Zhang, N., Hoadley, A., Patel, J., Lim, S., Li, C., 2017. Sustainable options for the
- 621 utilization of solid residues from wine production. Waste Manag. 60, 173–183.
- 622 https://doi.org/10.1016/j.wasman.2017.01.006