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**Multi-residue determination of psychoactive pharmaceuticals,
illicit drugs and related metabolites in wastewater by Ultra-High
Performance Liquid Chromatography-Tandem Mass Spectrometry**

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Abstract

This study presents a new multi-residue analytical method for the simultaneous determination of 38 psychoactive drugs (including benzodiazepines, antidepressants and drugs of abuse) and related metabolites in raw wastewater. Potential analyte losses during sample filtration and stability in wastewater were evaluated. Analyte losses, especially for 12 compounds, were observed during filtration, indicating a strong sorption onto the filter material. In order to overcome this effect, filtered water samples were combined with methanolic washes of the corresponding filters and the resulting

26 solutions were solid-phase extracted on mixed-mode (reverse-phase plus cation-
27 exchange) sorbents. Extracts were analyzed by ultra-high performance liquid
28 chromatography-tandem mass spectrometry. Quantification was performed by the
29 internal standard method with isotopic labeled analogs. Recovery percentages varied
30 between 65% and 137%; method quantification limits ranged between 0.2 and 22 ng/L
31 in ultrapure water and between 0.3 and 30 ng/L in wastewater for all the analytes but
32 three (for which they were ~60-80 ng/L). The analysis of 24 h-composite samples
33 collected during one week in the city of Santiago de Compostela demonstrated the
34 ubiquity of 31 analytes, which were positively quantified in all samples. The highest
35 concentrations were found for some of the antidepressants, with mean and maximum
36 levels exceeding, in some cases, the levels previously reported in literature. This fact
37 could be related to the additional washing step of the filters using methanol, which
38 allowed to desorb retained analytes highlighting the importance of this step during the
39 sample preparation protocol.

40

41 **Keywords:** Benzodiazepines; Antidepressants; Substances of abuse; Sewage analysis;
42 Solid-phase extraction; Ultra-High Performance Liquid Chromatography

43

44 1. Introduction

45 Psychoactive pharmaceuticals, drugs of abuse and their metabolites are widely
46 known to be present in urban wastewaters due to their high rates of production and
47 consumption [1-9]. According to the Health Indicators of the Organization for
48 Economic Co-operation and Development (OECD) [10], the use of antidepressants has
49 increased considerably in most OECD countries since the year 2000, with selective
50 serotonin reuptake inhibitors (SSRIs, e.g. fluoxetine, paroxetine, sertraline, citalopram)
51 being the most popular in Spain [11]. Anxiolytics and hypnotics, particularly
52 benzodiazepines (lorazepam, diazepam, alprazolam), are another group of medicines
53 with remarkable rates of prescription [12]. And, among drugs of abuse, cannabis
54 accounts for the largest estimate of abuse in the European Union, followed by cocaine,
55 ecstasy, other amphetamine-derived compounds and opioids [13]. Following this
56 widespread consumption, residues of licit and illicit psychoactive substances enter
57 sewage systems continuously [1-9], a fact that may imply environmental consequences
58 (if they end in surface waters [9, 14, 15]) and act as a measurable indicator of their use
59 in different communities [2, 5, 16, 17].

60 Most of the analytical methodologies developed for the determination of
61 psychoactive substances in wastewater remove suspended particles by filtration or
62 centrifugation prior to solid-phase extracting the aqueous phase. However, organic
63 molecules may get adsorbed onto solids following a process that depends on the
64 properties of both the substance (pK_a , K_{ow} , etc.), the suspended particle matter (SPM)
65 and the water itself (pH, Total Organic Carbon - TOC) and the filtering materials.
66 Therefore, sorption is very hard to predict and, if a proper evaluation of the portion of
67 substance adsorbed is not performed, it is possible to underestimate its levels in real
68 samples [18-24]. In the case of drugs and pharmaceuticals, Baker et al. [18] assessed the

69 sorption onto wastewater SPM for 16 out of the 38 analytes included in our study and
70 concluded that it was >10% for methadone and 2-ethylidene-1,5-dimethyl-3,3-
71 diphenylpyrrolidine (EDDP), and >30% (up to 89% in one of the samples) for some
72 antidepressants like fluoxetine and norfluoxetine. This result highlighted the need to
73 take sorption processes into consideration for the development of future sample
74 preparation protocols for these analytes.

75 In this line, this study presents a novel sample preparation strategy for the solid-
76 phase extraction (SPE) of 38 psychoactive drugs and metabolites in wastewater.
77 Particular attention was paid to pretreatment steps, including water filtration, washing of
78 the filters and in-sample stability, in order to avoid the under-reporting of
79 concentrations in real wastewater. The analytes were carefully selected to be the most
80 frequently prescribed psychoactive pharmaceuticals, the most frequently abused drugs
81 in Spain and their most relevant metabolites: i) seven benzodiazepines and two of their
82 metabolites; ii) methylphenidate - a psycho-stimulant drug used in the treatment of
83 attention deficit hyperactivity disorder - and its main metabolite ritalinic acid; iii) eight
84 antidepressants and five of their metabolites; and iv) eight illicit drugs, five metabolites
85 and levamisole, the most common adulterant of cocaine. Analytes were separated and
86 detected by ultra-high performance liquid chromatography (UHPLC) coupled to tandem
87 mass spectrometry (MS/MS). Parameters affecting UHPLC separation and MS/MS
88 detection were carefully optimized and the final method was validated in terms of
89 trueness, precision and quantification limits. Finally, it was applied to the analysis of 24
90 h-composite raw wastewater samples collected during one week in the city of Santiago
91 de Compostela (NW of Spain).

92

93 **2. Experimental**

94 **2.1. Reagents and materials**

100 Analyte standards were supplied by Cerilliant (Round Rock, TX, USA) as
101 individual solutions of 100 µg/mL of norsertaline and O-desmethylvenlafaxine, or 1000
102 µg/mL in methanol (MeOH) of alprazolam, α-hydroxyalprazolam, diazepam,
103 nordiazepam, oxazepam, temazepam, lorazepam, lormetazepam, chlordiazepoxide,
104 methylphenidate, ritalinic acid, citalopram, N-desmethylcitalopram, fluoxetine,
105 norfluoxetine, sertraline, venlafaxine, mirtazapine, N-desmethyilmirtazapine, duloxetine,
106 paroxetine, trazodone, amphetamine, methamphetamine, 3,4-
107 methylendioxyamphetamine (MDMA), cocaine, benzoylecgonine, cocaethylene,
108 levamisole, 11-hydroxy-Δ⁹-THC (THC-OH), 11-nor-9-carboxy-Δ⁹-THC (THC-
109 COOH), meta-chlorophenylpiperazine (mCPP), mephedrone, ketamine, methadone and
110 EDDP. Isotopic labeled analogs (α-hydroxyalprazolam-D₅, alprazolam-D₅, diazepam-
111 D₅, nordiazepam-D₅, oxazepam-D₅, temazepam-D₅, lorazepam-D₄, methylphenidate-
112 D₉, ritalinic acid-D₁₀, citalopram-D₆, N-desmethylcitalopram-D₃, fluoxetine-D₆,
113 norfluoxetine-D₆, sertraline-D₃, norsertaline-¹³C₆, venlafaxine-D₆, O-
114 desmethylvenlafaxine-D₆, duloxetine-D₃, paroxetine-D₆, trazodone-D₆, amphetamine-
115 D₆, methamphetamine-D₅, MDMA-D₅, cocaine-D₃, benzoylecgonine-D₃,
116 cocaethylene-D₃, THC-OH-D₃, THC-COOH-D₃, mCPP-D₈, mephedrone-D₃,
117 ketamine-D₄, methadone-D₃ and EDDP-D₃) were also supplied by Cerilliant as 100
118 µg/mL solutions in MeOH and used as internal standards (IS). Mixed stock solutions
119 containing all the analytes (10 µg/mL) or all the IS (2 µg/mL) were prepared in MeOH
120 and stored in the dark at -20°C until use.

121 HPLC-grade MeOH, acetonitrile (ACN), acetic acid (100%) and ammonia
122 solution in ultrapure water (25%) were supplied by Merck (Darmstadt, Germany).
123 Formic acid (95-97%) and NH₃ solution in MeOH (7M) were supplied by Sigma-
124 Aldrich (San Luis, Mi, USA). Ultrapure water was obtained in the laboratory by

125 purifying demineralized water in a Milli-Q Gradient A-10 system (Merck-Millipore,
126 Bedford, MA, USA).

127

128 **2.2. Filtration tests**

129 Potential sorption of analytes onto different filter materials was assessed by
130 vacuum filtering 100 mL aliquots of ultrapure water, spiked with 5 ng/mL of all the
131 analytes, through different types of filters: 0.7 μm glass microfiber filters GF/A
132 (Whatman, Kent, U.K.), 0.45 μm mixed cellulose membranes (Millipore, Bedford, MA,
133 USA), 0.45 μm hydrophilic nylon membranes (Millipore) and 0.45 μm hydrophilic
134 PVDF membranes (Millipore). IS were added after filtration and samples (n=3 in every
135 case) solid-phase extracted as detailed in section 2.4. Losses were calculated as:

$$136 \quad \text{Filtration loss (\%)} = \left(1 - \left(\frac{\text{Response}}{\text{Average}(\text{Response}_{\text{No filtr}})} \right) \right) \times 100$$

137 Where *Response* is the IS-corrected response in a filtered sample and
138 *Average(Response_{No filtr})* is the average of the IS-corrected responses in non-filtered
139 samples. One-way ANOVA ($\alpha=0.05$) were performed to compare the mean losses of
140 every analyte with the four types of filters.

141 For the filters providing the best performance (GF), experiments were repeated
142 with raw wastewater in order to assess the combined sorption onto the filter and the
143 SPM. Aliquots (100 mL, n=3) were spiked with 5 ng/mL of all the analytes before and
144 after being filtered, IS added after filtration and losses calculated as:

$$145 \quad \text{Filtration loss (\%)} = \left(1 - \left(\frac{\text{Average}(\text{Response}_{\text{Before}})}{\text{Average}(\text{Response}_{\text{After}})} \right) \right) \times 100$$

146 Where *Average(Response_{Before})* is the average of the IS-corrected responses in
147 samples spiked before filtration; and *Average(Response_{After})* is the average of the IS-

148 corrected responses in samples spiked after filtration. Individual Student's t-tests
149 ($\alpha=0.05$) were run for all the analytes to assess whether there were statistically
150 significant differences between their mean filtration losses in ultrapure water *versus* in
151 raw wastewater, i.e., whether there were differences between their sorption onto filters
152 (exclusively) or their combined sorption onto filters and SPM.

153 Finally, potential recovery of the analytes adsorbed by means of a methanolic
154 wash of the filter was assessed with raw wastewater samples spiked with 2 ng/mL of all
155 the analytes before and after filtration (n=3). Filters were washed with 2×5 mL of
156 MeOH. Washes were collected, spiked with IS and made to a final volume of 1 mL for
157 instrumental analysis.

158

159 **2.3. Antidepressant biodegradation tests**

160 Biodegradation of benzodiazepines and drugs of abuse was not assessed since it
161 had been already reported in literature [25, 26].

162 Potential biodegradation of antidepressants was evaluated by spiking 10 mL of
163 raw wastewater (n=3) with 500 ng/mL of these analytes and collecting 0.7 mL aliquots
164 at the beginning of the experiment and at different times up to 48 h. Each aliquot was
165 passed through a 0.22 μm GHP membrane syringe filter (Pall laboratory, NY, USA).
166 Subsequently, 0.7 mL of MeOH were used to wash the filter and collected over the
167 water fraction. The resulting solutions were spiked with 100 ng/mL of IS and kept at -
168 20°C until analysis (by direct injection into the UHPLC-MS/MS system). Signals were
169 compared to the response of a standard in ultrapure water:MeOH 1:1 containing 250
170 ng/mL and 100 ng/mL of analytes and IS, respectively.

171

172 **2.4. Sampling and sample treatment**

173 Raw wastewater samples were collected at the wastewater treatment plant
174 (WWTP) of Santiago de Compostela (NW of Spain), which treats mostly domestic
175 wastewater and serves a population of ~136,500 inhabitants. Composite samples of 24 h
176 were taken in April 2016 for seven consecutive days, from 10.00 a.m. to 10.00 a.m. of
177 the following day. A Sigma SD900 portable sampler from Hach (Loveland, CO, USA)
178 worked in time proportional mode collecting 120 mL of water every 10 min. Composite
179 samples were transferred to the laboratory and extracted within 8 h after the end of the
180 sampling.

181 The sample preparation protocol was adapted from two previously published
182 works [26, 27]. Under final working conditions, 100.0 mL aliquots were spiked with 20
183 ng of IS and vacuum-filtered through 0.7 μm glass microfiber filters GF/A. Filters were
184 washed with 2 \times 5.0 mL of MeOH, which were collected together with the filtered
185 aqueous sample. Resulting solutions were solid-phase extracted onto mixed reverse
186 phase-cation exchange cartridges (Oasis MCX-150 mg, Waters Corp., Milford, MA,
187 USA), previously conditioned with 5.0 mL of MeOH containing 5% of NH_3 followed
188 by 5.0 mL of ultrapure water. Sorbents were washed with 10.0 mL of ultrapure water
189 and dried under nitrogen for 30 min. Analytes were recovered with 10.0 mL of 5% NH_3
190 in MeOH. Eluates were evaporated to dryness under nitrogen (99.999%) using both a
191 Turbo-Vap II (Zymark, Hopkinton, MA USA) and a Mini-Vap concentrator (Sigma-
192 Aldrich). They were redissolved in 100 μL of MeOH for instrumental analysis. Every
193 sample was processed in triplicate.

194

195 **2.5. UHPLC-MS/MS analysis**

196 Samples (2 μL) were injected into a Waters Acquity UPLC[®] H class system
197 (Milford, MA, USA) equipped with a sample manager, a quaternary solvent pump and a

198 column oven. Chromatographic separation was carried out at 50°C on a Kinetex® EVO
199 C18 100 Å column (50 × 2.1 mm I.D., particle size 1.7 µm) from Phenomenex
200 (Torrance, CA, USA), protected with a C18 pre-column (4 × 2 mm I.D), also from
201 Phenomenex. A dual eluent system consisting of (A) 5 mM of NH₃ in ultrapure water
202 and (B) 5 mM of NH₃ in MeOH was employed at a flow rate of 0.5 mL/min. The
203 gradient elution started with 30% B, increasing to 60% B in 4 min and then to 100% B
204 in 0.01 min. 100% B was held for 2 min. Return to initial conditions (30% B) was
205 performed in 0.01 min and held for 2 min for reconditioning.

206 The UPLC® system was coupled to a triple quadrupole mass spectrometer Xevo
207 TQD (Waters Corp., Milford, MA, USA) equipped with an electrospray ionization
208 (ESI) source. Nitrogen, used as desolvation and cone gas, was provided by a nitrogen
209 generator (Peak Scientific Spain, Barcelona, Spain). Argon, for the collision induced
210 dissociation, was purchased from Praxair (Madrid, Spain). Ionization was performed in
211 positive mode using the following parameters: 4 kV (capillary voltage), 150°C (source
212 temperature), 500°C (desolvation temperature), 1000 L/h (desolvation gas flow, N₂) and
213 50 L/h (cone gas flow, N₂). Collision energy (CE) and cone voltage (CV) values were
214 adjusted individually for every compound. MS analyses were done in Selected Reaction
215 Monitoring (SRM) mode recording one (IS) or two (analytes) precursor/product ion
216 transitions per compound. Selected transitions, together with their corresponding CE
217 and CV values, retention times (RT) and labeled compounds used as IS are listed in the
218 Supplementary Material, Table S1.

219

220 **2.6. Method validation**

221 The method was validated in terms of linearity, instrumental repeatability,
222 instrumental and method quantification limits (IQLs and MQLs), trueness and

223 precision. Analytes were quantified using the corresponding isotopic labeled analog as
224 IS. In those (five) cases where no labeled analog was available, the labeled compound
225 providing the best results in terms of trueness was selected (Table S1).

226 Calibration was performed using a 13-point calibration curve ranging from
227 individual IQLs to 1500 ng/mL. For sertraline, fluoxetine and lormetazepam, it ranged
228 from IQL to 500 ng/mL; IS level in all cases: 200 ng/mL. IQLs were calculated as the
229 concentration of a standard providing a signal-to-noise ratio (S/N) of 10. MQLs were
230 assessed from measured concentrations in ultrapure water and wastewater samples
231 containing (or spiked with) low concentrations of all the analytes, downscaling the
232 levels for which the signal-to-noise ratio is 10. Instrumental repeatability was assessed
233 as the relative standard deviation (%RSD) of six consecutive injections of two different
234 standards (containing 5 and 50 ng/mL of all the analytes and 200 ng/mL of IS).

235 Trueness and precision of the whole method were estimated from recovery experiments
236 performed in ultrapure water spiked at two concentration levels (20 and 100 ng/L of all
237 the analytes, 200 ng/L of all IS) and in raw wastewater spiked with 500 ng/L of all the
238 analytes and 200 ng/L of all IS. In the latter case, IS only-spiked aliquots were analyzed
239 simultaneously in order to correct for the levels of analytes in sewage. Responses
240 (analyte area/IS area) in ultrapure water or differences between responses of analyte-
241 spiked and non-spiked aliquots of wastewater were compared with calibration curves in
242 MeOH.

243

244 **3. Results and discussion**

245 **3.1. UHPLC-MS/MS optimization**

246 MS/MS conditions (transitions, CE and CV values) were optimized by direct
247 infusion of individual standard solutions (10 µg/mL) in MeOH. Ionization was

248 performed in positive mode. Two SRM transitions (one quantifier, one qualifier) were
249 acquired per analyte and one transition per IS. CV and CE values providing the highest
250 intensities were selected individually (Table S1).

251 Chromatographic separation was carried out on a Kinetex[®] EVO C18 column stable
252 throughout the whole pH range (1-12). Several modifiers giving different pH values
253 were considered for the mobile phase (consisting of ultrapure water – mobile phase A,
254 and MeOH – mobile phase B): formic acid 0.1% (pH 2.7); ammonium acetate 5 mM
255 (pH 7.0); and NH₃ 5 mM (pH 10.5). Figure 1 displays the chromatograms of nine
256 analytes representative of the different behaviours observed in the three different
257 scenarios. Since most substances have basic groups, a basic pH ensures their
258 neutralization, increasing their retention on the C18 phase and improving peaks shape.
259 As an example, amphetamine derivatives split in two in acidic medium, so formic acid
260 was discarded for them. This splitting was also observed for other basic compounds
261 such as mephedrone, ketamine or methylphenidate, whose height was between 1.5 and 3
262 times higher with NH₃ than with ammonium acetate. Higher peaks were also obtained
263 in basic medium for other basic, less polar species (e.g. duloxetine, sertraline) and for
264 amphoteric compounds (e.g. benzoylecgonine, ritalinic acid), what demonstrates the
265 higher sensitivity of the proposed method when a basic eluent system was used. On the
266 contrary, THC-COOH and THC-OH peaks were higher with formic acid but, since this
267 was a minor behaviour, NH₃ was added at a concentration of 5 mM to both the aqueous
268 and the organic phase. The use of MeOH or ACN was also considered, but no
269 significant differences were observed neither on peak shapes nor on analyte intensities
270 (data not shown), so MeOH was selected due to its lower price.

271 The consecutive injection of 1, 2, 3, 4 and 5 µL of a standard showed that the
272 greater the volume injected, the higher the intensity. As it is shown in Figure S1 of the

273 Supplementary Material, a reasonable peak width was maintained in all cases (e.g.
274 trazodone) excepting ritalinic acid and benzoylecgonine, for which peaks were split.
275 Although a certain percentage of water in the solvent used to prepare standards and
276 reconstitute extracts could have avoided the split, it was discarded due to the already
277 known poor stability of some of the investigated species in water (e.g. cocaine [28]).
278 Alternatively, a compromise injection volume of 2 μ L, which provided high signal
279 intensity avoiding significant peak widening for benzoylecgonine and ritalinic acid, was
280 adopted. As an example of the chromatographic performance, Figure 2 shows the
281 extracted ion chromatogram (EIC) for the first transition of all the analytes in a 100
282 ng/mL standard.

283

284 **3.2. Assessment of losses during sample filtration**

285 In first instance, sorption of analytes onto filters (exclusively) was assessed by
286 comparing different filter materials: GF, cellulose membranes, hydrophilic nylon and
287 hydrophilic PVDF (see section 2.2). Losses were lower than 30%, independently of the
288 material used, for all analytes but methadone, its metabolite EDDP, the two
289 cannabinoids and most of the antidepressants (Figure 3). An ANOVA statistical test
290 allowed to conclude that differences between mean losses observed with the four types
291 of filters were statistically significant, at the 95% of confidence level, for thirteen
292 compounds: citalopram, N-desmethylcitalopram, fluoxetine, norfluoxetine, sertraline,
293 mirtazapine, duloxetine, paroxetine, trazodone, THC-OH, THC-COOH, methadone and
294 EDDP. For ten of them, higher losses were observed with cellulose membranes
295 followed by PVDF filters. For mirtazapine and EDDP, there was barely no difference
296 between these two materials. Conversely, lower adsorption occurred on GF and
297 hydrophilic nylon, which were regarded as the best filtering materials. However, THC-

298 OH and THC-COOH disappeared completely after being filtered through nylon, what
299 prevented the selection of this material in favour of the use of GF.

300 Combined sorption onto both SPM and GF filters was further assessed with raw
301 wastewater and compared to the (exclusive) sorption on GF filters occurring with
302 ultrapure water (Figure 4). At the 95% of confidence level, there were no statistically
303 significant differences between the mean losses observed in these two matrices,
304 indicating a strong sorption onto the filter material barely affected by the content of
305 SPM. Only O-desmethylvenlafaxine, mirtazapine and EDDP underwent a significantly
306 higher loss in ultrapure water. This may be attributed to the fact that dissolved organic
307 matter partially prevents sorption to the filter unit.

308 Desorption of analytes from filters by washing them with 2×10 mL of MeOH was
309 evaluated with raw wastewater samples as explained in section 2.2. For most of the
310 compounds, recoveries in the water extract were above 80% (data not shown),
311 demonstrating again that sorption has not a great impact on them. For citalopram, N-
312 desmethylcitalopram, methadone and EDDP, recoveries varied between 60% and 80%,
313 and for some compounds with high K_{ow} (i.e. fluoxetine, norfluoxetine, sertraline,
314 norsesertraline, duloxetine, paroxetine, THC-OH and THC-COOH) they were below 50%
315 (Supplementary Material, Figure S2). For these analytes, recoveries in filter washes
316 reached values above 20% of the total addition in the water sample. For citalopram, N-
317 desmethylcitalopram, methadone and EDDP they varied between 9 and 24% (Figure
318 S2). For the remaining compounds, filter washes recoveries were below 10% in all
319 cases.

320 Therefore, the combination of the filtered water sample and the methanolic filter
321 washes was further extracted and analysed, as explained in sections 2.4 and 2.5, in order
322 to improve the accuracy and the sensitivity of the method. Moreover, the addition of IS

323 before filtration allowed us to compensate for the uncertainty associated to the sorption
324 occurring during filtration, and also to avoid a potential underestimation of the
325 concentrations found in wastewater.

326

327 **3.3. Stability of antidepressants in wastewater**

328 Stability experiments were performed as explained in section 2.3 for
329 antidepressants and their metabolites. After sample filtration, recovery experiments
330 showed that 0.7 mL MeOH were necessary to sweep the analytes from the filter,
331 especially fluoxetine, norfluoxetine, sertraline, nortriptyline, duloxetine and paroxetine.
332 A second MeOH wash was not necessary, since less than 5% of the analytes were eluted
333 in this fraction (data not shown). Figure S3 of the Supplementary Material compiles the
334 biodegradation profiles for all the antidepressants along 48 h at room temperature. Since
335 no significant degradation was observed in any case (relative responses >80%), no
336 degradation is expected during the 24 h sampling.

337

338 **3.4. Method performance**

339 UHPLC-MS/MS performance parameters (linearity, instrumental repeatability
340 and IQLs) are displayed in Table 1. The representation of the ratio analyte area/IS area
341 versus analyte concentration fitted a linear model with determination coefficients (R^2)
342 between 0.9928 and 0.9987. The linear range was IQL-1500 ng/mL for all the analytes
343 but sertraline, fluoxetine and lormetazepam, for which it was IQL-500 ng/mL. %RSD
344 values for six repeated injections of a standard varied between 0.6 and 6.9% at 5 ng/mL
345 and between 1.4 and 13% at 50 ng/mL. IQLs were between 0.1 and 13 ng/mL.

346 The combined SPE-UHPLC-MS/MS method was validated in terms of trueness,
347 precision and MQLs. Percentages of recovery (%R) for quadruplicate analyses of

348 ultrapure water samples spiked with 20 ng/L of all the analytes and 200 ng/L of IS
349 varied between 71% for lormetazepam and 132% for norsertaline (Table 1). THC-
350 COOH could not be recovered in this case since its MQL in ultrapure water was higher
351 than the spiked level (22 ng/L). %RSD were between 1% and 10%. At 100 ng/L in
352 ultrapure water, %R varied from 72% for N-desmethyilmirtazapine to 137% for mCPP
353 .%RSD were between 3% and 27%. In raw wastewater experiments (spiking level: 500
354 ng/L of all the analytes, 200 ng/L of IS), %R varied from 65% for duloxetine to 134%
355 for N-desmethyilmirtazapine, with %RSD between 1% and 14%. Finally, MQLs ranged
356 from 0.2 ng/L to 22 ng/L in ultrapure water and from 0.3 ng/L to 30 ng/L in wastewater
357 for all the analytes but mCPP (82 ng/L), amphetamine (64 ng/L) and mephedrone (83
358 ng/L).

359 Table S2 offers a comparative of the performance of the proposed method *versus*
360 other multi-residue analytical methods for the determination of psychoactive substances
361 in wastewater. MQLs were in the same order of magnitude than MQLs reached by other
362 methodologies [4, 6, 7, 26, 29-31]. Trueness was similar or even better, with IS
363 corrected %R in the 65-134% range *versus* 39-226% [29], 32-125% [30] or 51-130%
364 [4]. It must be noticed, however, that both MQL and %R values can be estimated in
365 different ways and, therefore, performance figures offered by different researchers
366 might not be readily comparable. In terms of analysis time, the optimized SPE protocol
367 is as long as other off-line SPE protocols, being, of course, slower than the in-line SPE-
368 LC-MS/MS method developed by Fedorova et al. [31], or the Auto-SPE on HLB discs
369 optimized by Baz-Lomba et al. [30]. The chromatographic separation is, conversely, the
370 fastest (8 min in total).

371

372 **3.5. Occurrence in 24 h-composite wastewater samples**

373 The optimized and validated method was used to analyze 24 h-composite raw
374 wastewater samples collected during one week in April 2016 at the WWTP of Santiago
375 de Compostela (Spain).

376 Out of the 38 analytes, 31 could be positively quantified in all samples and two
377 (α -hydroxyalprazolam and EDDP) in five and four samples, respectively (Table 2).
378 Amphetamine, methamphetamine, mCPP, mephedrone and ketamine were always
379 <MDL. This may be related to their relatively high MDL values in the method proposed
380 (10-83 ng/L). Moreover, the consumption figures of these particular substances in Spain
381 are usually very low [13], resulting in very low concentrations in wastewater.

382 Among drugs of abuse, benzoylecgonine (196-489 ng/L) was the metabolite
383 quantified at the highest levels, followed by its precursor cocaine (82-259 ng/L), the
384 cocaine adulterant levamisole (41-129 ng/L) and the THC metabolites THC-COOH (26-
385 70 ng/L) and THC-OH (19-62 ng/L). These values reflect the pattern of consumption of
386 illicit drugs in Spain [13], with cocaine and cannabis being the most abused substances
387 and, therefore, the ones found in higher amounts in wastewater [32]. Other compounds
388 such as MDMA (2-13 ng/L), methadone (3-13 ng/L) and its metabolite EDDP (<MDL-
389 14 ng/L) were quantified at lower levels.

390 Benzodiazepine-related compounds were quantified at mean levels <100 ng/L, in
391 the same order of magnitude than in other European studies [6, 9, 26, 29, 33, 34].
392 Lorazepam (44-182 ng/L) and oxazepam (9-100 ng/L) were the most abundant species,
393 but they were exceeded, in any case, by ritalinic acid (metabolite of methylphenidate,
394 63-195 ng/L).

395 Among all the analytes, the highest concentrations were found for the
396 antidepressant metabolite O-desmethylvenlafaxine (1066-1231 ng/L), followed by its
397 precursor venlafaxine (459-1063 ng/L). These values are in good agreement with the

398 values reported for venlafaxine in the Slovakian city of Trenčín (391-947 ng/L) [7], but
399 they are considerably higher than the levels observed in other Slovakian [7], Greek [6],
400 German [9] and British [29, 33] cities. Sertraline (176-455 ng/L) and its metabolite
401 norsesertraline (209-531 ng/L) were also quantified at very high concentrations when
402 compared to other European studies [6, 9, 34, 35], a fact that could be associated to their
403 desorption from filtered particles in the present work. These substances are highly
404 retained onto SPM and common sample preparation methodologies (separating aqueous
405 and solid phases and adding IS after filtration) may underestimate their real levels in
406 wastewater. However, several factors may affect the occurrence of antidepressants in
407 sewage from different countries (i.e. different rates of prescription) and any association
408 with sample preparation/methodological issues may be considered cautiously.
409 Fluoxetine, norfluoxetine and paroxetine concentrations were in line with or lower than
410 the concentrations reported in other countries [6, 9, 26, 29, 34-36]. Citalopram (110-183
411 ng/L), N-desmethylcitalopram (79-136 ng/L) and mirtazapine (27-81 ng/L) had already
412 been found at similar/higher levels in other cities [6, 9, 34, 36], whereas N-
413 desmethyilmirtazapine (14-22 ng/L), duloxetine (17-60 ng/L) and trazodone (35-545
414 ng/L) are usually not detected or detected at lower concentrations.

415 Regarding weekly concentrations profiles, and in accordance with the expected
416 pattern of constant consumption, no clear trend could be observed for any of the
417 investigated pharmaceuticals, neither for antidepressants nor for benzodiazepines.
418 Conversely, higher levels were detected for all drugs/related compounds in the weekend
419 samples (Saturday, Sunday, Monday), reflecting the recreational use of these substances
420 in the city of Santiago de Compostela.

421

422 **4. Conclusions**

423 A new SPE-UHPLC-MS/MS method has been developed for the multi-residue
424 determination of 38 psychoactive drugs (covering the most consumed psychoactive
425 pharmaceuticals and drugs of abuse in Spain) in wastewater. Sample filtration proved to
426 be a critical step in the loss of the most hydrophobic analytes (methadone, EDDP, THC-
427 COOH, THC-OH and eight antidepressants). This led us to include a washing of the
428 filters (and its subsequent extraction together with the filtered water) in the final
429 protocol. This step allowed to reach higher absolute recoveries and lower limits of
430 detection for these compounds, sometimes not detected, or detected at low –potentially
431 underestimated– concentrations in real wastewater by other methodologies that simply
432 extract the filtered aqueous phase (in the best scenario, adding IS before filtration). The
433 analysis of seven 24 h-composite raw wastewater samples demonstrated the ubiquity of
434 most of the analytes, with some of the antidepressants quantified at very high levels
435 when compared to other European studies.

436

437 **Appendix A. Supplementary material**

438

439 **Declarations of interest: none**

440

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446

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Figure Captions

Figure 1. Chromatograms of some representative analytes using a water/MeOH mobile phase at pH 2.7 (purple line), pH 7.0 (green line) and pH 10.5 (red line).

Figure 2. Extracted Ion Chromatogram (EIC) for the first transition (Q1) of all the analytes in a 100 ng/mL standard mixture.

Figure 3. Analyte losses (%) during ultrapure water filtration through glass microfiber filters (GF), cellulose membranes, hydrophilic nylon and hydrophilic PVDF filters. Error bars represent the standard deviation. Numbers above the bars are p-values for ANOVA statistical tests assessing the presence/absence of statistically significant differences between the mean losses observed with the four types of filters ($\alpha = 0.05$).

Figure 4. Analyte losses (%) observed with glass microfiber filters (GF) in ultrapure water and in raw wastewater. Error bars represent the standard deviation. Numbers above the bars are p-values for Student's t-tests assessing the presence/absence of statistically significant differences between the mean losses observed in the two matrices ($\alpha = 0.05$).

Table 1. Method performance parameters: determination coefficients (R^2), relative standard deviations (%RSD), instrumental quantification limits (IQL), recovery values (%R) and limits of quantification of the whole method (MQL).

	Analyte	Linearity (R^2) ^a	Repeatability (% RSD, n=6)		IQL (ng/mL)	Trueness and precisión (%R (%RSD), n=4)			MQL (ng/L)	
			5 ng/mL ^b	50 ng/mL		Ultrapure (20 ng/L)	Ultrapure (100 ng/L)	Wastewater (500 ng/L)	Ultrapure	Wastewater
Benzodiazepines and methylphenidate	Alprazolam	0.9982	6.9	2.2	0.4	106 (8)	104 (3)	107 (3)	0.9	1.4
	α -hydroxyalprazolam	0.9961	4.3	1.4	1.1	102 (3)	119 (27)	100 (3)	2.8	4.5
	Diazepam	0.9975	1.5	2.4	0.2	98 (6)	114 (12)	106 (1)	0.5	1.2
	Nordiazepam	0.9973	1.6	1.8	1.7	89 (4)	108 (10)	91 (2)	2.8	3.0
	Oxazepam	0.9984	4.8	1.8	1.1	101 (3)	100 (4)	100 (6)	1.7	8.0
	Temazepam	0.9960	1.5	2.6	0.7	106 (6)	100 (5)	91 (3)	1.7	3.3
	Lorazepam	0.9948	6.1	3.5	0.5	113 (10)	101 (12)	76 (11)	0.9	21
	Lormetazepam	0.9968	1.5	1.8	1.4	71 (10)	99 (3)	99 (12)	3.4	3.7
	Chlordiazepoxide	0.9978	1.4	2.5	1.1	72 (6)	98 (6)	71 (3)	2.3	4.5
	Methylphenidate	0.9985	2.8	2.5	1.0	100 (5)	110 (8)	75 (1)	1.1	1.6
Ritalinic acid	0.9982	1.5	3.1	0.4	104 (5)	114 (8)	83 (4)	0.4	0.7	
Antidepressants	Citalopram	0.9943	1.5	2.5	0.9	104 (4)	105 (9)	99 (2)	0.9	5.5
	N-desmethylcitalopram	0.9969	3.0	1.4	0.5	104 (1)	103 (9)	107 (3)	1.2	6.3
	Fluoxetine	0.9955	1.8	1.6	0.1	92 (7)	106 (8)	82 (2)	1.0	1.8
	Norfluoxetine	0.9986	2.2	2.7	0.1	106 (5)	110 (9)	94 (12)	0.5	3.3
	Sertraline	0.9981	2.9	3.0	0.3	107 (5)	107 (5)	93 (2)	0.8	4.4
	Norsertaline	0.9948	NA	13	10	132 (9)	92 (16)	114 (2)	10	30
	Venlafaxine	0.9966	3.1	2.9	0.1	104 (4)	104 (7)	98 (3)	0.7	0.9
	O-desmethylvenlafaxine	0.9985	2.4	1.7	0.4	103 (6)	108 (11)	91 (11)	1.1	7.7
	Mirtazapine	0.9976	0.8	2.3	0.1	99 (4)	103 (13)	102 (4)	0.2	1.1
	N-desmethyilmirtazapine	0.9928	3.6	3.1	0.1	105 (6)	72 (8)	134 (6)	0.4	6.8
	Duloxetine	0.9987	4.0	3.8	0.1	106 (8)	119 (11)	65 (2)	0.3	8.6
	Paroxetine	0.9953	2.3	1.4	0.4	106 (8)	102 (9)	108 (11)	1.1	6.9
Trazodone	0.9979	0.6	2.5	0.1	101 (6)	105 (7)	101 (2)	0.4	1.8	

Illicit drugs	Amphetamine	0.9932	NA	3.1	6.6	90 (2)	103 (10)	106 (14)	9.0	64
	Methamphetamine	0.9964	2.0	2.9	1.1	89 (5)	103 (9)	102 (2)	1.7	27
	MDMA	0.9971	1.4	2.6	1.1	101 (4)	123 (16)	100 (6)	1.5	1.2
	Cocaine	0.9975	2.0	2.1	0.1	106 (4)	110 (5)	100 (3)	1.1	3.2
	Benzoyllecgonine	0.9968	2.0	3.6	0.2	105 (5)	105 (4)	95 (7)	0.3	3.3
	Cocaethylene	0.9977	1.9	2.3	0.1	105 (2)	107 (7)	96 (2)	0.2	0.3
	Levamisole	0.9930	0.8	3.8	0.9	97(6)	92 (9)	79 (2)	1.2	5.8
	THC-OH	0.9985	NA	2.3	7.0	102 (2)	106 (4)	88 (3)	8.2	9.3
	THC-COOH	0.9946	NA	4.0	13	NA ^c	117 (7)	121 (8)	22	25
	mCPP	0.9967	6.9	2.4	5.0	101 (7)	137 (18)	102 (4)	5.5	82
	Mephedrone	0.9978	1.1	1.8	1.0	106 (4)	101 (10)	114 (3)	1.3	83
	Ketamine	0.9965	0.6	2.3	0.1	105 (7)	103 (9)	116 (3)	0.7	10
	Methadone	0.9971	1.4	2.0	0.2	104 (7)	97 (6)	95 (2)	1.1	1.7
EDDP	0.9972	3.7	2.6	2.7	87 (6)	98 (11)	79 (3)	3.2	4.8	

^a Linear range: IQL - 1500 ng/mL for all the analytes excepting sertraline, fluoxetine and lormetazepam (IQL - 500 ng/mL)

^b Repeatability at 5 ng/mL was not calculated for norsertraline, amphetamine, THC-OH and THC-COOH since their IQL > 5 ng/mL

^c Recovery at 20 ng/L was not calculated for THC-COOH since its MQL > 20 ng/L

Table 2. Analyte concentration (mean in ng/L and %RSD in brackets) in 24 h-composite raw wastewater samples collected during one week in Santiago de Compostela. <MDL: not detected. <MQL: detected, but below the MQL.

	Analyte	Mo	Tu	W	Th	Fr	Sa	Su
Benzodiazepines and methylphenidate	Alprazolam	5 (24)	2 (28)	4 (11)	6 (29)	3 (12)	3 (5)	4 (29)
	α -hydroxyalprazolam	9 (22)	<MQL	<MQL	14 (15)	7 (19)	6 (19)	9 (21)
	Diazepam	3 (35)	2 (1)	2 (22)	4 (7)	2 (26)	2 (12)	3 (1)
	Nordiazepam	13 (6)	8 (14)	10 (9)	17 (2)	11 (33)	8 (9)	13 (1)
	Oxazepam	25 (8)	10 (25)	27 (16)	100 (20)	32 (3)	9 (25)	57 (19)
	Temazepam	12 (2)	8 (25)	12 (20)	15 (20)	11 (7)	12 (16)	11 (6)
	Lorazepam	71 (8)	68 (4)	60 (9)	182 (24)	44 (17)	78 (38)	72 (15)
	Lormetazepam	11 (5)	12 (15)	13 (23)	10 (29)	6 (8)	7 (20)	8 (18)
	Chlordiazepoxide	12 (9)	8 (21)	12 (18)	16 (11)	9 (27)	7 (3)	10 (23)
	Methylphenidate	8 (25)	8 (16)	10 (6)	13 (19)	11 (8)	6 (11)	8 (12)
	Ritalinic acid	99 (10)	75 (4)	97 (8)	195 (1)	145 (8)	63 (7)	115 (11)
Antidepressants	Citalopram	140 (4)	119 (11)	161 (2)	110 (5)	123 (5)	129 (7)	183 (4)
	N-desmethylcitalopram	136 (3)	79 (16)	113 (3)	92 (10)	79 (12)	88 (9)	110 (4)
	Fluoxetine	38 (13)	29 (4)	49 (9)	23 (8)	36 (1)	36 (6)	64 (7)
	Norfluoxetine	32 (29)	16 (12)	32 (3)	19 (6)	23 (8)	26 (29)	43 (5)
	Sertraline	258 (7)	227 (3)	396 (2)	176 (5)	254 (9)	232 (8)	455 (4)
	Norsertaline	233 (4)	230 (17)	471 (13)	209 (9)	309 (24)	256 (9)	531 (31)
	Venlafaxine	551 (6)	459 (2)	1063 (6)	661 (1)	466 (3)	478 (1)	549 (1)
	O-desmethylvenlafaxine	1231 (5)	1102 (15)	1066 (13)	1124 (6)	1152 (5)	1175 (2)	1213 (6)
	Mirtazapine	81 (8)	31 (16)	33 (7)	27 (3)	30 (2)	33 (13)	36 (19)
	N-desmethylmirtazapine	20 (11)	17 (7)	21 (8)	14 (14)	16 (13)	16 (26)	22 (19)
	Duloxetine	60 (22)	21 (8)	45 (18)	17 (19)	23 (6)	24 (16)	46 (21)
	Paroxetine	24 (10)	21 (10)	26 (8)	13 (8)	18 (10)	17 (1)	35 (8)
	Trazodone	62 (12)	45 (9)	45 (4)	35 (9)	545 (8)	92 (7)	99 (2)

Illicit drugs	Amphetamine	<MDL	<MDL	<MDL	<MDL	<MDL	<MDL	<MDL
	Methamphetamine	<MDL	<MDL	<MDL	<MDL	<MDL	<MDL	<MDL
	MDMA	6 (6)	2 (26)	2 (6)	6 (13)	8 (11)	11 (12)	13 (10)
	Cocaine	169 (25)	82 (24)	165 (1)	230 (5)	215 (1)	259 (4)	240 (9)
	Benzoyllecgonine	398 (6)	196 (31)	289 (11)	331 (7)	371 (9)	464 (3)	489 (24)
	Cocaethylene	4 (11)	1 (36)	4 (9)	7 (3)	7 (11)	11 (11)	13 (5)
	Levamisole	69 (15)	41 (16)	47 (4)	63 (2)	76 (14)	129 (23)	129 (23)
	THC-OH	58 (14)	31 (23)	19 (8)	54 (12)	34 (18)	28 (6)	62 (11)
	THC-COOH	53 (18)	26 (4)	33 (20)	70 (9)	29 (4)	47 (18)	56 (8)
	mCPP	<MDL	<MDL	<MDL	<MDL	<MDL	<MDL	<MDL
	Mephedrone	<MDL	<MDL	<MDL	<MDL	<MDL	<MDL	<MDL
	Ketamine	<MDL	<MDL	<MDL	<MDL	<MDL	<MDL	<MDL
	Methadone	8 (30)	3 (19)	6 (34)	13 (10)	4 (13)	4 (24)	7 (1)
	EDDP	12 (16)	<MDL	<MQL	14 (12)	<MDL	5 (9)	10 (16)

Figure 1

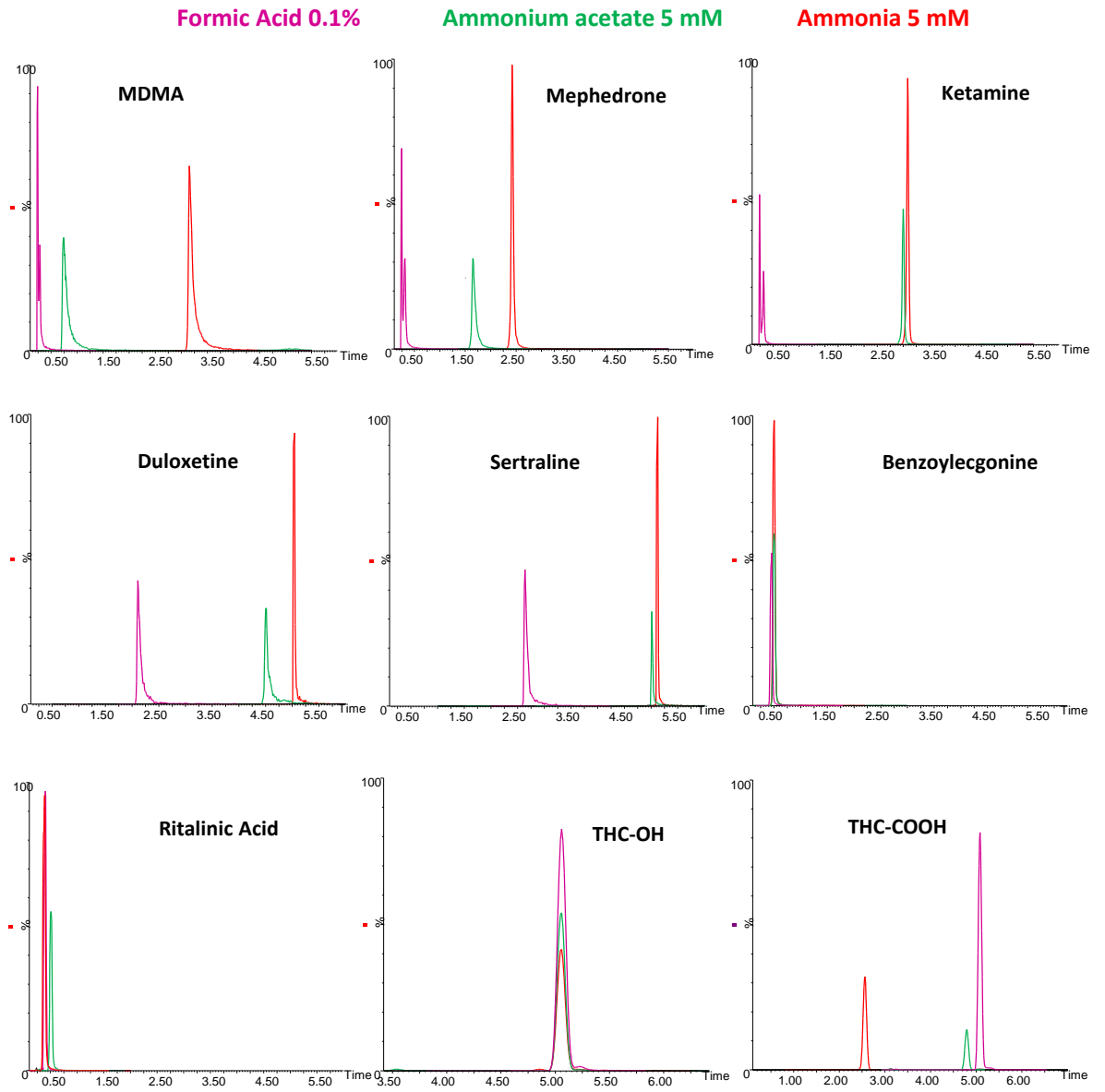


Figure 2

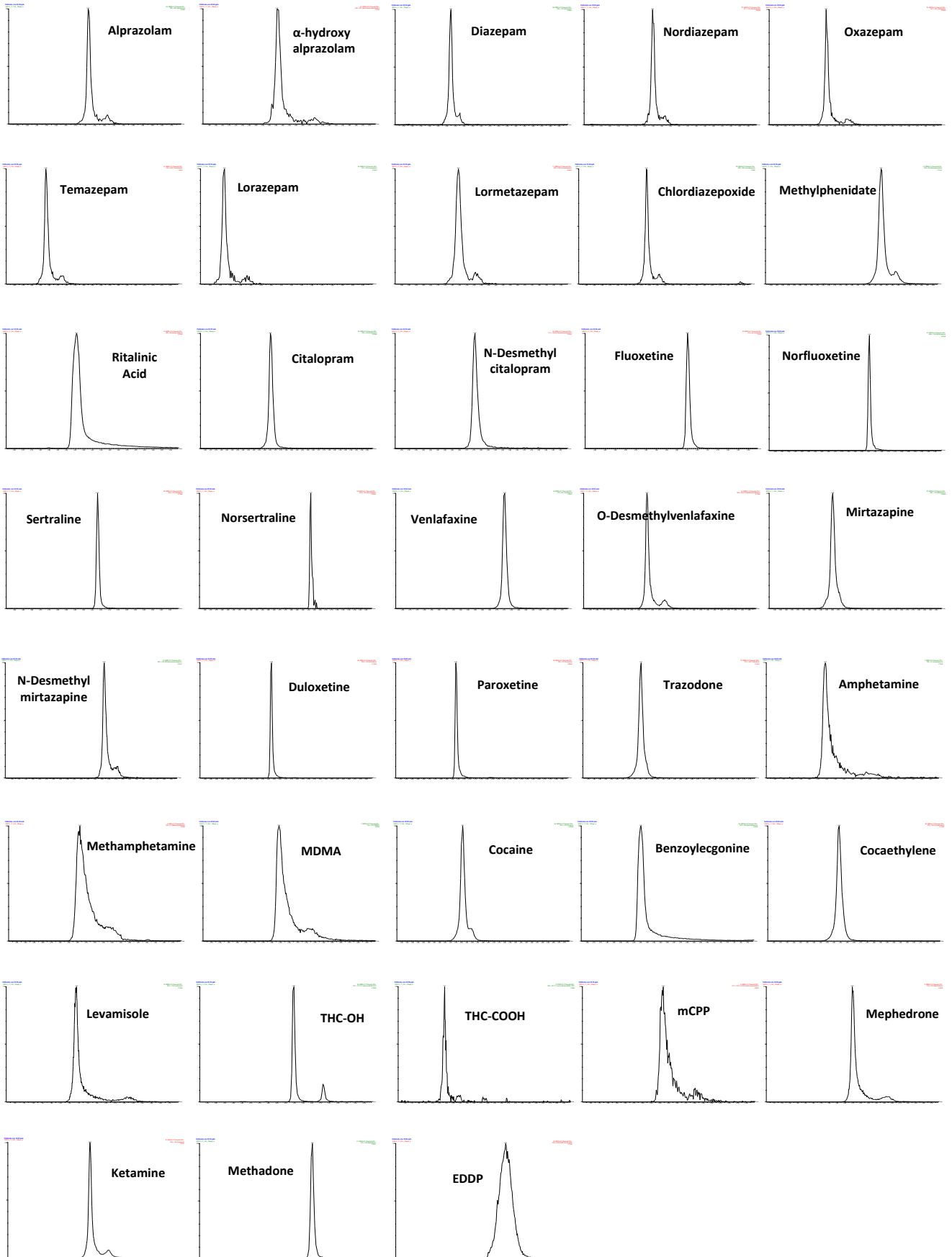


Figure 3

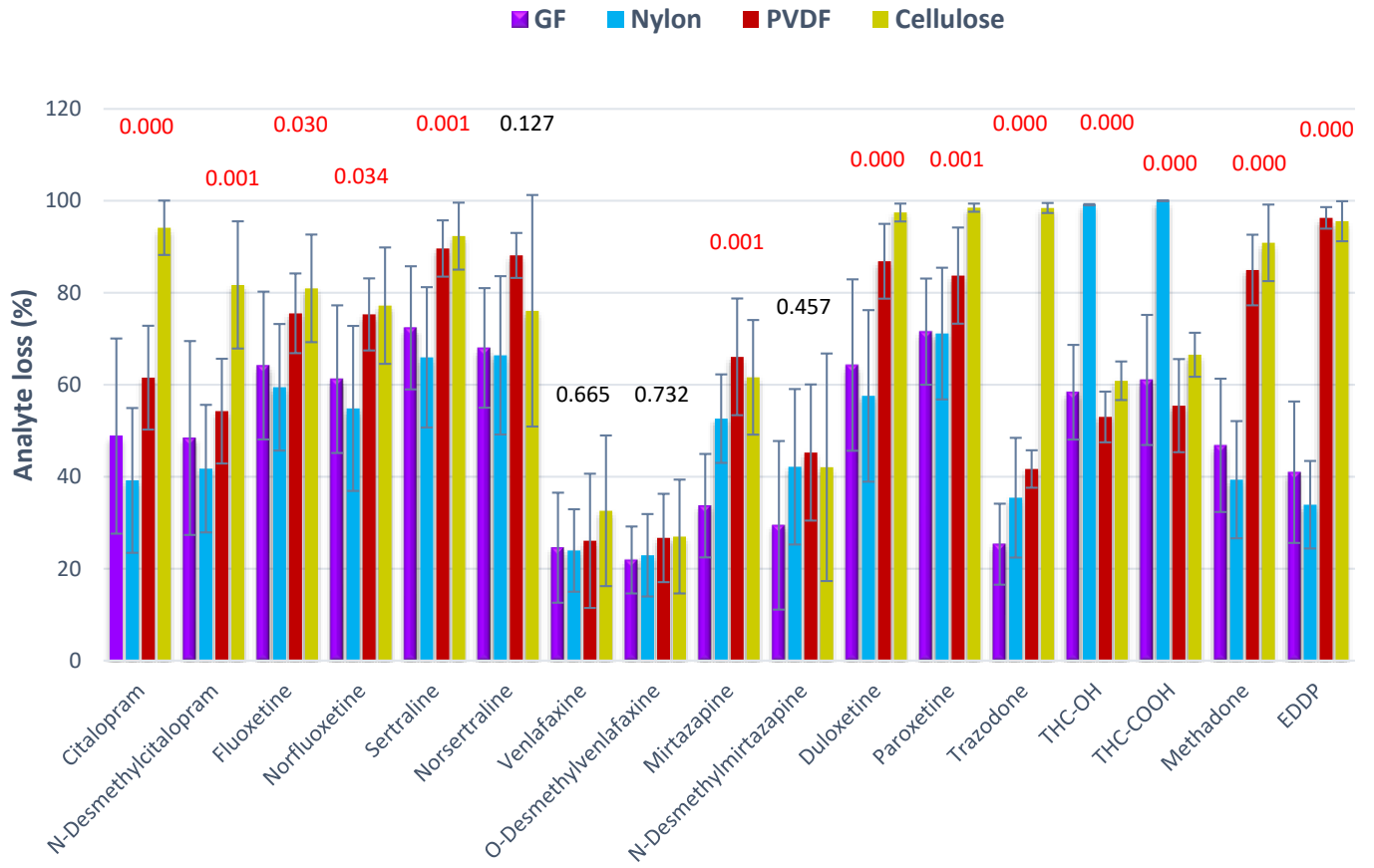
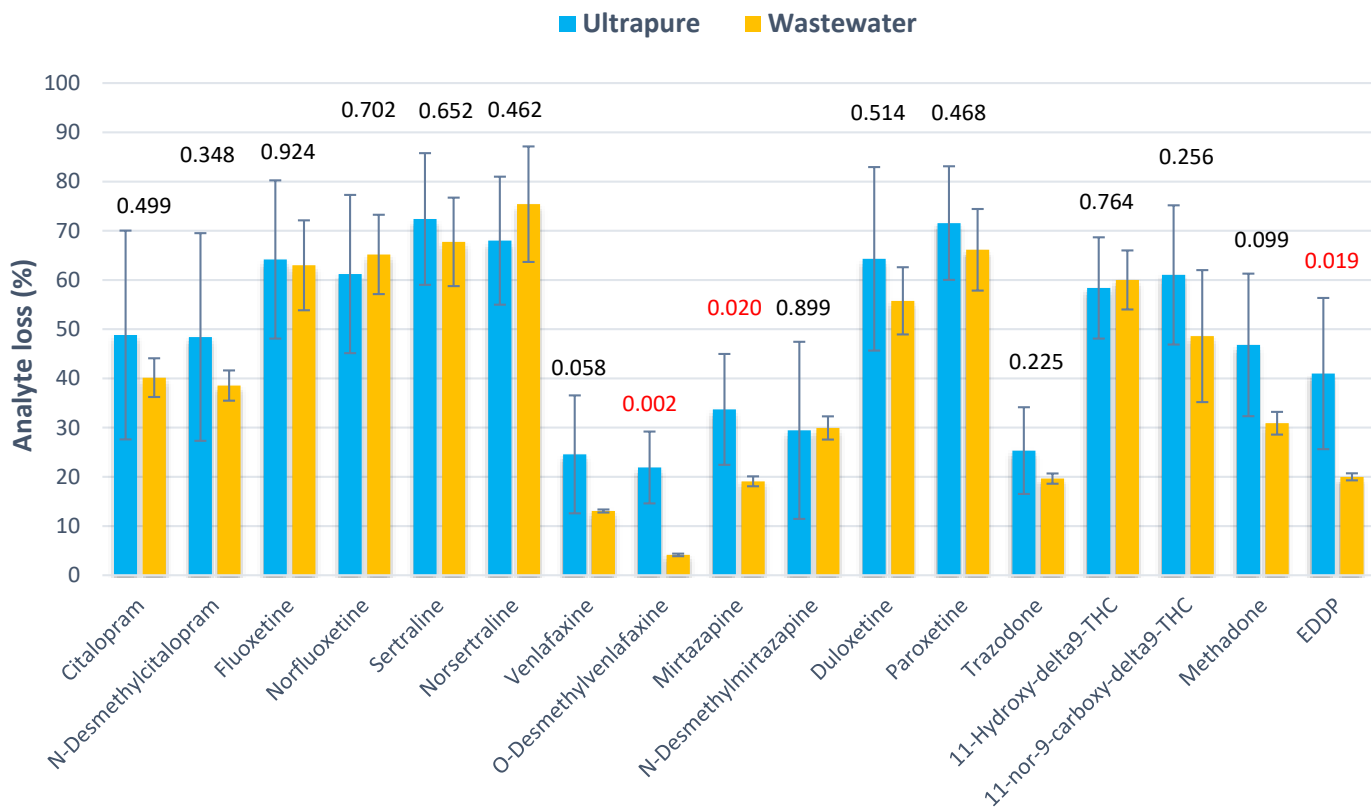


Figure 4



Appendix A. Supplementary material to:

Multi-residue determination of psychoactive pharmaceuticals, illicit drugs and related metabolites in wastewater by Ultra-High Performance Liquid Chromatography-Tandem Mass Spectrometry

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Table S1. Analytes, labeled compounds used as IS, retention time (RT), cone voltage (CV), quantifier (Q1) and qualifier (Q2) *m/z* transition (precursor > product ion), and collision energy (CE) selected in every case.

	Analyte	IS	RT (min)	CV (V)	Q1	CE (V)	Q2	CE (V)
Benzodiazepines and methylphenidate	Alprazolam	Alprazolam-D5	2.7	60	309>205	46	309>281	34
	α -hydroxyalprazolam	α -hydroxyalprazolam-D5	2.5	64	325>297	24	325>216	38
	Diazepam	Diazepam-D5	3.7	60	285>193	32	285>89	62
	Nordiazepam	Nordiazepam-D5	3.3	60	271>140	30	271>165	30
	Oxazepam	Oxazepam-D5	2.6	38	287>241	22	287>269	14
	Temazepam	Temazepam-D5	3.0	42	301>255	20	301>283	14
	Lorazepam	Lorazepam-D4	2.7	45	321>275	20	321>303	14
	Lormetazepam	Alprazolam-D5	3.2	40	335>289	22	335>317	14
	Chlordiazepoxide	Alprazolam-D5	3.2	34	300>283	14	300>227	24
	Methylphenidate	Methylphenidate-D9	3.5	6	234>84	20	234>91	40
Ritalinic acid	Ritalinic Acid-D10	0.4	22	220>84	20	220>56	38	
Antidepressants	Citalopram	Citalopram-D6	4.7	50	325>109	30	325>262	22
	N-desmethylcitalopram	N-desmethylcitalopram-D3	4.8	44	311>109	26	311>262	16
	Fluoxetine	Fluoxetine-D6	5.0	26	310>44	14	310>148	8
	Norfluoxetine	Norfluoxetine-D6	5.0	18	296>134	8	296>30	10
	Sertraline	Sertraline-D3	5.0	24	306>159	24	306>275	12
	Norsertraline	Norsertraline-13C6	5.0	14	292>159	20	292>275	8
	Venlafaxine	Venlafaxine-D6	4.8	38	278>58	20	278>260	12
	O-desmethylvenlafaxine	O-desmethylvenlafaxine-D6	2.9	32	264>58	18	264>246	12
	Mirtazapine	Trazodone-D6	4.1	46	266>195	28	266>72	24
	N-desmethylmirtazapine	Cocaine-D3	2.9	46	252>195	24	252>209	24
	Duloxetine	Duloxetine-D3	5.0	16	298>154	6	298>44	46
Paroxetine	Paroxetine-D6	4.9	48	330>70	30	330>44	22	

	Analyte	IS	RT (min)	CV (V)	Q1	CE (V)	Q2	CE (V)
Illicit drugs	Trazodone	Trazodone-D6	4.2	52	372>148	38	372>176	26
	Amphetamine	Amphetamine-D6	2.2	24	136>91	16	136>119	10
	Methamphetamine	Methamphetamine-D5	3.1	28	150>91	16	150>119	10
	MDMA	MDMA-D5	2.9	28	194>163	12	194>105	26
	Cocaine	Cocaine-D3	3.8	30	304>182	22	304>82	32
	Benzoyllecgonine	Benzoyllecgonine-D3	0.5	44	290>168	20	290>105	32
	Cocaethylene	Cocaethylene-D3	4.5	32	318>196	22	318>82	32
	Levamisole	Cocaine-D3	1.7	48	205>178	22	205>91	36
	THC-OH	THC-OH-D3	5.0	32	331>313	14	331>193	26
	THC-COOH	THC-COOH-D3	2.2	46	345>327	18	345>299	22
	mCPP	mCPP-D8	2.4	44	197>154	18	197>44	22
	Mephedrone	Mephedrone-D3	2.1	30	178>160	12	178>145	20
	Ketamine	Ketamine-D4	2.7	36	238>125	24	238>207	16
	Methadone	Methadone-D3	5.1	50	310>105	32	310>265	16
EDDP	EDDP-D3	5.3	28	278>234	34	278>249	26	
Internal Standards	Alprazolam-D5	-	2.7	60	314>286	34	-	-
	α-hydroxyalprazolam-D5	-	2.5	64	330>302	24	-	-
	Diazepam-D5	-	3.7	60	290>198	32	-	-
	Nordiazepam-D5	-	3.3	60	276>140	30	-	-
	Oxazepam-D5	-	2.6	38	292>246	22	-	-
	Temazepam-D5	-	3.0	42	306>260	20	-	-
	Lorazepam-D4	-	2.7	45	325>279	20	-	-
	Methylphenidate-D9	-	3.5	6	243>93	20	-	-
	Ritalinic Acid-D10	-	0.4	22	230>93	20	-	-
	Citalopram-D6	-	4.7	50	331>109	30	-	-
	N-desmethylocitalopram-D3	-	4.8	44	314>109	26	-	-
	Fluoxetine-D6	-	5.0	26	316>44	14	-	-

Analyte	IS	RT (min)	CV (V)	Q1	CE (V)	Q2	CE (V)
Norfluoxetine-D6	-	5.0	18	302>30	10	-	-
Sertraline-D3	-	5.0	24	309>275	12	-	-
Norsertaline- ¹³ C6	-	5.0	14	298>159	20	-	-
Venlafaxine-D6	-	4.8	38	284>64	20	-	-
O-desmethylvenlafaxine-D6	-	2.9	32	270>64	18	-	-
Duloxetine-D3	-	5.0	16	301>157	6	-	-
Paroxetine-D6	-	4.9	48	336>74	30	-	-
Trazodone-D6	-	4.2	52	378>150	38	-	-
Amphetamine-D6	-	2.2	24	142>125	10	-	-
Methamphetamine-D5	-	3.1	28	155>92	16	-	-
MDMA-D5	-	2.9	28	199>165	12	-	-
Cocaine-D3	-	3.8	30	307>185	22	-	-
Benzoyllecgonine-D3	-	0.5	44	293>171	20	-	-
Cocaethylene-D3	-	4.5	32	321>199	22	-	-
THC-OH-D3	-	5.0	32	334>316	14	-	-
THC-COOH-D3	-	2.2	46	348>46	40	-	-
mCPP-D8	-	2.4	44	205>158	18	-	-
Mephedrone-D3	-	2.1	30	181>148	20	-	-
Ketamine-D4	-	2.7	36	242>129	24	-	-
Methadone-D3	-	5.1	50	313>105	32	-	-
EDDP-D3	-	5.3	28	281>234	34	-	-

Table S2. Comparison of the performance of the proposed method with that of other multi-residue analytical methods for the determination of psychoactive substances in wastewater. Abbreviations: percentage of recovery (%R); method quantification limit (MQL); Glass fiber (GF); Polytetrafluoroethylene (PTFE); Triple quadrupole (QqQ); Quadrupole-time-of-flight (QTOF); Multiple reaction monitoring (MRM); Product ion scan (PIS); Internal standards (IS); Methanol (MeOH); Acetonitrile (ACN); Acetic acid (AA); Formic acid (FA); Positive ionization (PI); Negative ionization (NI).

Analytes	Sample preparation		Separation and detection		%R	MQL (ng/L)	Ref.
	Pretreatment	Extraction	LC-MS	Run time			
65 stimulants, including illicit drugs, benzodiazepines, antidepressants and selected metabolites	100 mL of sample Filtration through GF 2.7 μm filters + GF 0.7 μm filters Acidification to pH 1.8-1.9 Addition of IS	SPE on Oasis MCX 60 mg Filtration of the extract through 0.2 μm PTFE filters	UPLC-(ESI+)-MS/MS on QqQ (MRM) Acquity UPLC BEH C18 column (150 \times 1 mm, 1.7 μm) Mobile phase: 0.3% AA in water:MeOH 80:20 and 0.3% AA in MeOH	34 min	(IS corrected %R) 39-226% > 60% for most analytes	0.5-140	[30]
68 psychoactive substances, including illicit drugs, benzodiazepines, antidepressants and selected metabolites	50 mL of sample Filtration through GF 0.7 μm filters Acidification to pH 2.5 Addition of IS	SPE on Strata XC 200 mg Filtration of the extract through 0.2 μm RC filters	UPLC-(ESI+ and ESI-)-MS/MS on QqQ (MRM) Kinetex PFP column (50 \times 2.1 mm, 1.7 μm) Mobile phase: 0.05% FA in water and 0.05% FA in MeOH	65 min (PI) 35 min (NI)	(Absolute %R) 22-142% 80-120% for > 80% analytes	0.3-558	[6]
23 psychoactive pharmaceuticals, including benzodiazepines, selected metabolites and related pharmaceuticals	100 mL of sample Addition of IS Filtration through GF 0.7 μm filters + membrane 0.45 μm filters	SPE on Oasis MCX 60 mg	HPLC-(ESI+)-MS/MS on QqQ (MRM) Synergi Fusion column (100 \times 2.0 mm, 4 μm) Mobile phase: 0.1% AA in water and 0.1% AA in MeOH	29 min	(IS corrected %R) 84-109%	0.1-18	[26]

51 psychoactive substances including illicit drugs, benzodiazepines, antidepressants and selected metabolites	100 mL of sample Addition of IS	Auto SPE-DEX HLB discs (47 mm i.d.)	UPLC-(ESI+)-MS on QTOF (MS ^e) Acquity UPLC HSS C18 column (150 × 2.1 mm, 1.8 μm) Mobile phase: 5 mM Ammonium Formate in water, 0.1% FA in ACN	15 min	(IS corrected %R) 32-125% > 60% for most analytes	0.4-187	[31]
23 psychoactive pharmaceuticals including illicit drugs, benzodiazepines, antidepressants and selected metabolites	1 mL of sample Addition of IS Filtration through cellulose 0.45 μm syringe filters	In-line SPE Hyperil Gold (20 × 2.1, 12 μm)	HPLC-(ESI+)-MS/MS on QqQ (MRM) HPLC-MS on Q-Orbitrap (Full Scan and Product Ion Scan) Cogen bidentate column (50 × 2.1 mm, 3 μm) Mobile phase: water and ACN	15 min (extraction + LC-MS)	(matrix-matched %R) 34-139% (MRM) 93-146% (PIS)	1.3-15 (MRM) 1.7-11 (PIS)	[7], [32]
27 psychoactive pharmaceuticals including antidepressants and selected metabolites	100 mL of sample Centrifugation Filtration through GF 1 μm filter Addition of IS	SPE on Oasis HLB 200 mg	HPLC-(ESI+)-MS/MS on QqQ (MRM) Hypersil Gold column (150 × 2.1 mm, 3 μm) Mobile phase: 0.1% FA in water and 0.1% FA in MeOH	41 min	(IS corrected %R) 51-130%	0.1-20	[4]
38 psychoactive substances including illicit drugs, benzodiazepines, antidepressants and selected metabolites	100 mL of sample Addition of IS Filtration through GF 0.7 μm filter Wash of the filters with 2 × 5 mL MeOH	SPE on Oasis MCX 150 mg	UPLC-(ESI+)-MS/MS on QqQ (MRM) Kinetex EVO C18 column (50 × 2.1, 1.7 μm) Mobile phase: 5 mM NH ₃ in water and 5 mM NH ₃ in MeOH	8 min	(IS corrected %R) 65-134%	0.3-83	This study

Figure S1. Effect of the injection volume on the peak shape of trazodone, ritalinic acid and benzoylcegonine.

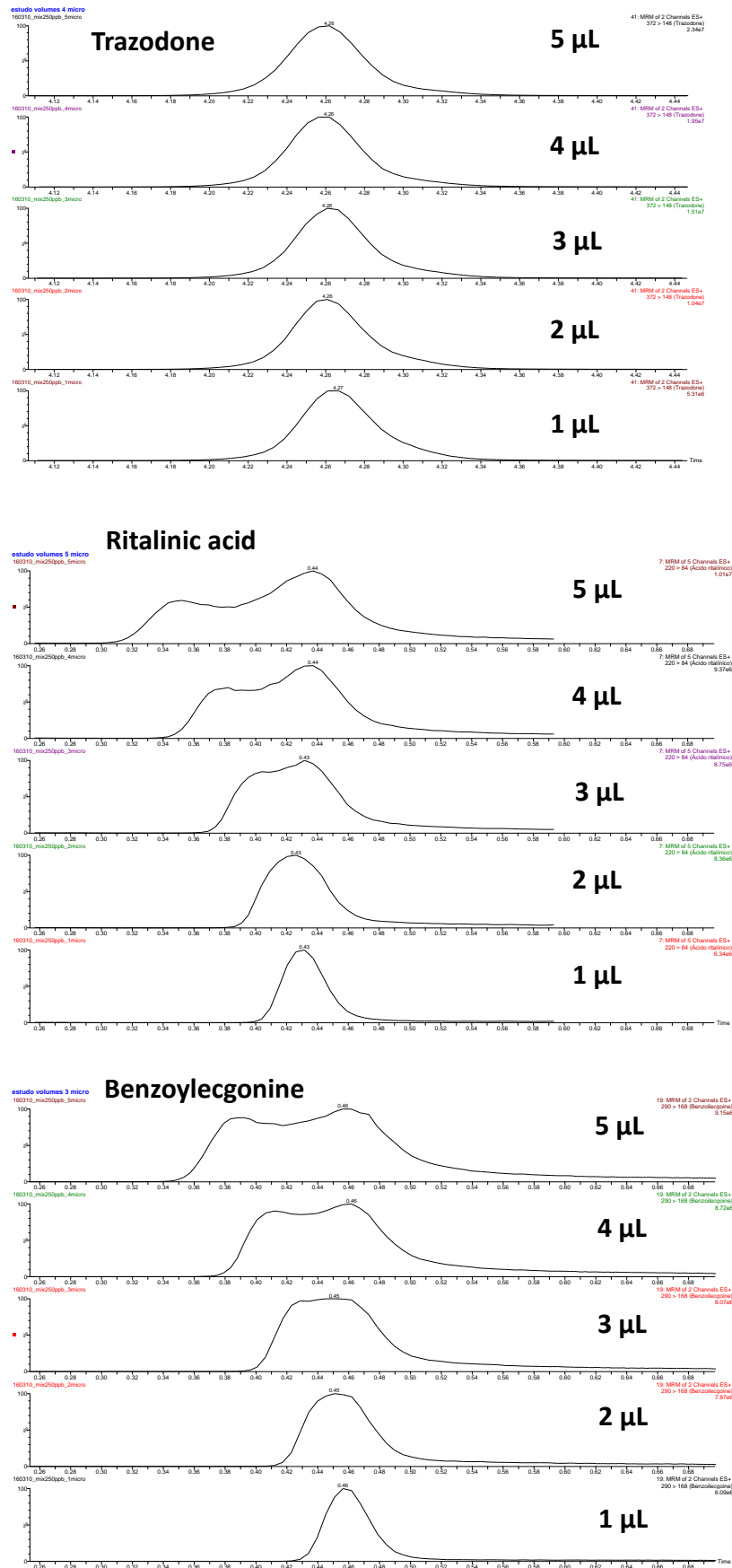


Figure S2. Analyte recovery (% relative to analyte response when spiked after filtration) in wastewater samples and in the methanolic washes of the filters used for their filtration.

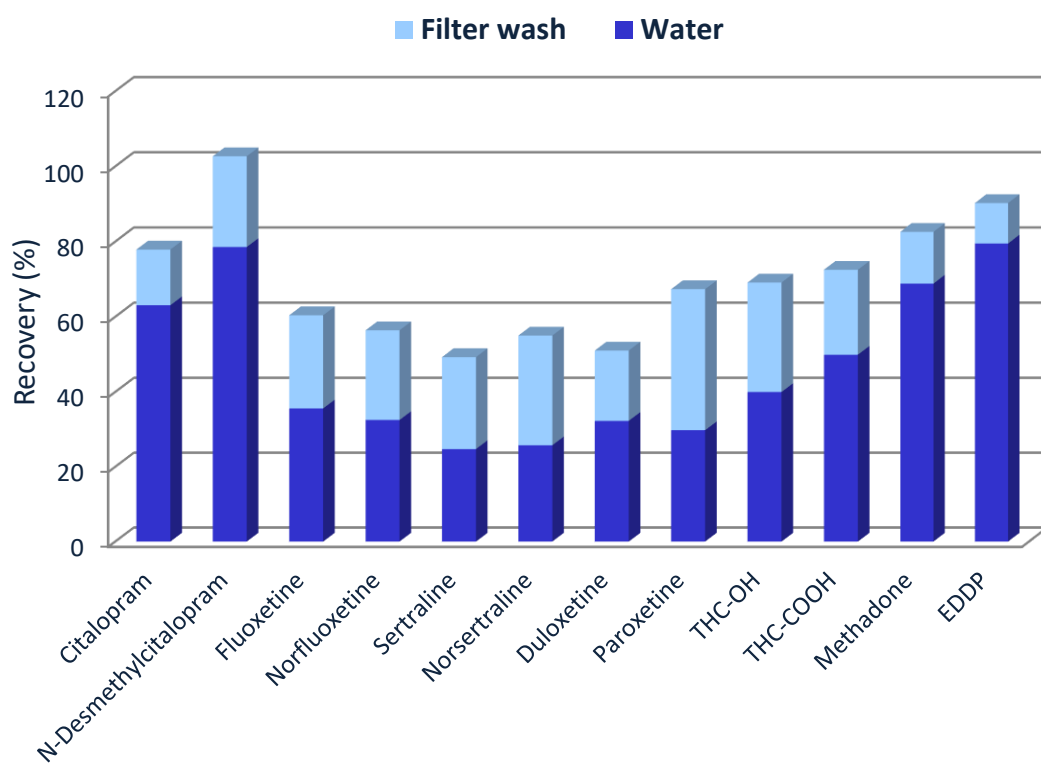


Figure S3. Antidepressant biodegradation profiles in wastewater along 48 h at room temperature.

