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CONNECTION BETWEEN GLOBAL CONSERVATION STATUS, GEOGRAPHICAL RANGE SIZE, MIDPOINT LATITUDE, FEMALE CARAPACE LENGTH, AND CLUTCH SIZE OF TESTUDINES

by

Ian E. Wick



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FACULTY OF NATURAL RESOURCES MANAGEMENT LAKEHEAD UNIVERSITY THUNDER BAY, ONTARIO

April 21, 2020

CONNECTION BETWEEN GLOBAL CONSERVATION STATUS, GEOGRAPHICAL RANGE SIZE, MIDPOINT LATITUDE, FEMALE CARAPACE LENGTH, AND CLUTCH SIZE OF TESTUDINES

by

Ian E. Wick

An Undergraduate Thesis Submitted in Partial Fulfillment of the Requirements for the Degree of Honours Bachelor of Environmental Management

Faculty of Natural Resources Management

Lakehead University

April 21, 2020

Major Advisor Dr. Stephen J. Hecnar Second Reader Dr. William Wilson

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ABSTRACT

Wick, Ian E. 2020. Connection between global conservation status, geographical range size, midpoint latitude, female carapace length, and clutch size of Testudines.

Keywords: Carapace, clutch, conservation, distribution, family, habitat loss, IUCN, latitude, range, road mortality, status, Testudines, turtle

The need for species conservation is only magnified with each passing day. Testudines are one of the taxonomic orders most at risk of extinction on Earth. Over 70% of Testudines are globally listed on the IUCN Red List and over 60% of those are at risk of extinction. Testudines face many threats including habitat loss and degradation. At time of data collection there were 258 turtles globally listed on the IUCN Red List. Following justified additions there were recognized to be 266 turtle species globally listed on the IUCN Red List for the purpose of this study. I collected data for 357 turtle species and examined the association of conservation status with geographic range size, midpoint latitude, female carapace length, and clutch size to determine if any of these attributes would be useful for determining extinction risk. IUCN status rank for species was most highly correlated with mean female carapace length. The positive association of risk with increasing body size supports concerns about the impact of harvesting or poaching of turtles and tortoises by humans. Testudines are clearly in need of conservation efforts.

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INTRODUCTION

With each passing day, the importance of conservation worldwide becomes increasingly salient. As human population growth continues, pollution and resource consumption increase while the resources and habitat available for other species decreases. The resulting impact on various organisms may differ in connection with their biological attributes. It is estimated the cost of conserving biodiversity globally could be over 75 billion USD annually (McCarthy 2013). According to Rhodin *et al.* (2018), Testudines (turtles and tortoises) are one of the orders of vertebrates most at risk of extinction similar to that of Caudata (salamanders) and Primates. Currently 62.8% of Testudine species on the IUCN Red List are at risk of extinction, while 50.8% of Caudata and 59.8% of Primates listed are at risk of extinction (IUCN 2019a; IUCN 2019b; IUCN 2019c). Without conservation efforts, Testudine diversity could suffer significantly within the next century (Buhlmann *et al.* 2009). Behler (2000) stated, "there is no vertebrate group facing greater survival problems today".

Testudines face numerous threats that vary in severity. The most serious threat to Testudines is habitat loss and fragmentation (Lesbarréres *et al.* 2014). Another major threat is poaching for both meat and medicines, and the exotic pet trade (Rhodin *et al.* 2018). The shell and bones are used for some traditional Chinese medicines and as Rhodin *et al.* (2018) notes "Asia is at the epicentre of the global turtle extinction crisis". Road mortality is another major threat to some species of Testudines (Ashley *et al.* 2007). However, road mortality is not always accidental, as Ashley *et al.* (2007) found that 2.7 out of every 100 drivers will intentionally hit a turtle that is on the road. Other threats include climate change, pollution, infectious diseases, invasive species, and nest predation (Rhodin *et al.* 2018). Plastic pollutants found in the oceans are a major threat

to Sea Turtles as they can ingest them or become entangled (Assuncao Ivar do Sul *et al.* 2010). A turtle found floating near Melbourne Beach in Florida defecated 74 foreign objects, requiring over a month to do so, following the removal of a gastrointestinal tract obstruction (Stamper *et al.* 2009). These threats are all very real and require human attention, as they are something the evolution of the nomadic turtle home more often referred to as the shell cannot protect them against. As Rhodin *et al.* (2011) stated "turtles are in serious trouble".

The unfortunate plight facing Testudines may be further expedited via taxonomic bias. Although in terms of species richness herpetofauna comprise over 40% of terrestrial vertebrates, Christoffel and Lepczyk (2012) found they were given less than 6% of the space in six wildlife journals over the last 30 years. Library holdings of post-secondary educational institutions and reintroduction projects have also displayed taxonomic bias (Seddon *et al.* 2005; Hecnar 2009).

Testudines play important roles in the functioning of the ecosystems they inhabit (Stanford *et al.* 2018). They can act as cleaners by scavenging and eating carrion (Langley 2018). They can also act as important agents of seed dispersal or create homes for other organisms (Braun and Brooks 1987; Langley 2018). Sometimes seeds can be reliant on turtles for germination (Rhodin *et al.* 2018). The Gopher Tortoise (*Gopherus polyphemus*) is considered a keystone species, as the burrows they create are shared with over 350 other species (Florida Fish and Wildlife 2019).

Over 250 of 360 extant species of Testudines are listed on IUCN's Red List (Rhodin *et al.* 2018). The International Union for Conservation of Nature (IUCN) was established in 1964 and provides critical information regarding status of the world's biodiversity (IUCN 2019d). The Species Survival Commission (SSC) is responsible for

completing Red List assessments (Campbell 2012). According to IUCN (2019d) 28,000 of the 105,700 species from all taxa listed on the Red List are threatened with extinction. In order to be considered threatened a species must be listed as "Vulnerable", "Endangered", or "Critically Endangered" (IUCN 2019e). Of the 258 currently listed Testudine species on the global IUCN Red List, 162 are at risk of extinction, 76 are described as decreasing, and only seven have a population trend described as increasing (Figure 1) (IUCN 2019f).

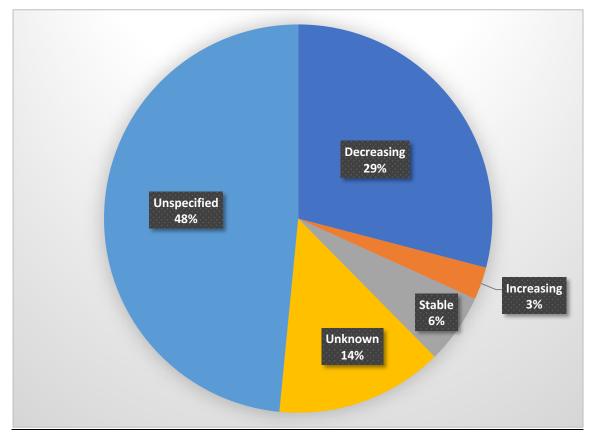


Figure 1. Population trend of the 258 Testudine species listed on the IUCN Red List (Adapted from IUCN 2019f).

LITERATURE REVIEW

The Order Testudines encompasses all the turtles and tortoises on Earth (Rhodin *et al.* 2018). Testudines can be found on land and in water, both fresh and salt, in every continent in the world excluding Antarctica (Zug 2019). All existing Testudines are considered in the suborders Pleurodira or Cryptodira, which are side-neck turtles and hidden neck turtles respectively (Pough *et al.* 2018). Some Cryptodires however lack the ability to fully retract their head into the shell, as seen in some members of the Chelydridae, Cheloniidae, and Dermochelyidae families (Pough *et al.* 2018; Boyer and Innis 2019). Pleurodires often have very long necks and are unable to retract their neck or head into their shell altogether, instead folding them sideways (Pough *et al.* 2018; Boyer and Innis 2019). There are only three families of Pleurodires as they are not as common as Cryptodires (Pough *et al.* 2018).

Understanding the connection between current conservation status and biological attributes may aid in understanding how Testudines become threatened as well as how conservation efforts can be better directed. Literature on conservation status and biological attributes of Testudines individually is abundant, however the connection between them is not often examined. Interestingly, of all extant reptiles, Testudines offer the most complete fossil record, however they are also possibly the most threatened vertebrate on earth (Lovich 2018; Pough 2018). Orentstein (2012) introduced readers to one of the earliest accounts of the devastation of turtles at the hands of humans, as the story of William Dampier and the Galapagos Islands is briefly reviewed. William Dampier is described as an "explorer and some-time pirate" and was the first to describe the tortoises of the Galapagos Islands, noting their abundance and size while also discussing the tortoises of Madagascar, the West Indies, and the Mascarenes

(Orenstein 2012). As Orenstein (2012) noted, unfortunately Dampier was not acknowledging them with admiration, respect, preservation or conservation in mind, instead merely thinking of the sustenance they could provide.

CONSERVATION

As of 2018, 14.9% of Earth's terrestrial area was protected (UNEP-WCMC et al. 2018). Testudines are in need of conservation as over 50% are threatened with extinction (Lovich 2018). They often have, or ideally have, what could be described as odd age structures within populations, and this is an important consideration in turtle conservation (Klemens 2000). As Klemens (2000) noted, the ideal proportion of eggs and juveniles in comparison to adults can vary greatly depending on species, however this proportion should greatly favour the eggs and juveniles. Rhodin et al. (2018) examined the IUCN conservation status for every extant Testudine species, while also including any recently extinct members, and noted, that ongoing evaluation of turtle species status and the efforts of the IUCN are essential components of future conservation efforts. There could be a significant reduction in turtle diversity in the near future (Buhlmann et al. 2009). Potential priority areas for turtle conservation and the need for conservation planning are outlined by Buhlmann et al. (2009) while examining tortoises and freshwater turtles. Iverson (1991) noted the importance of understanding that turtles followed a Type III survivorship curve and its importance in conservation. Rodrigues et al. (2006) discussed the importance and value of the IUCN Red List to conservation. In addition to designations provided by the IUCN there is useful data that can aid in conservation planning (Rodrigues et al. 2006). The IUCN is not without its critics though, as noted by Mrosovsky (1997), as he explored the importance of sound

and open science. Campbell (2011) explored the political side of the IUCN while examining the Hawksbill Sea Turtle (*Eretmochelys imbricata*).

ATTRIBUTES

Turtles date back over 200 million years, with the oldest known fossil being that of a carapace-lacking turtle known as *Odontochelys semitestacea* (Orenstein 2012). Absence of a carapace (dorsal shell) indicates it evolved after the plastron (ventral shell) (Pough *et al.* 2018). Turtles vary in size, as Orenstein (2012) noted, the largest turtle of all time was the Cretaceous sea turtle *Archelon ischyros*, weighing up to approximately 2040 kg and measuring up to approximately 4.5 m from snout to tail. Today, the largest turtle in the world is the last remaining member of the Family Dermochelyidae, the Leatherback Sea Turtle (*Dermochelys coriacea*), with a carapace length of up to two metres (Government of Canada 2019). Likely the smallest Testudine is the endangered Speckled Dwarf Tortoise (*Chersobius signatus*), which is endemic to South Africa and has a maximum carapace length of 110 mm (Orenstein 2012; Hofmeyr *et al.* 2018).

Turtles are well known for having a shell, although the origin of this conspicuous adaptation is somewhat controversial (Scoch *et al.* 2019). The shell is comprised of the carapace and plastron, which are the top and bottom of the shell respectively. The carapace and plastron are connected on each side by what is called the bridge and inside the shell the vertebrae are fused to the carapace (Pough *et al.* 2018). Most Testudines have a bony shell, but there are three families that possess a shell covered by leathery skin (Pough *et al.* 2018). Testudines rely on their shells as a means of protection from predators (Balani *et al.* 2011). Although there were once turtles that had teeth, such as *Odontochelys semitestacea*, all extant turtles lack teeth, instead having a keratinous beak (Orenstein 2012; Pough *et al.* 2018).

Turtles are oviparous (lay eggs), and clutch size refers to the number of eggs laid at one time (Shine 1983). Clutch size is associated with maternal body size and this notion of positive correlation is supported by a substantial amount of evidence (Ford and Seigel 1989; Ashton et al. 2007). Shine and Iverson (1995) explored the connection between maximum body size and age of sexual maturation, finding that much like other reptiles, the majority of turtles reach sexual maturity at approximately 70% of their maximum body size. Sexual dimorphism, reproductive strategies, and the size of male and female turtles were explored by Berry and Shine (1980), finding that terrestrial and aquatic species often have differing comparable sizes between sexes. Berry and Shine (1980) found that males tend to be larger in terrestrial species and females are likely to be larger in aquatic species. Body size of Testudines is significant, not simply age, as size often determines age of maturity (Iverson 1992a). This is evident in Snapping Turtles (*Chelvdra serpentina*) in Ontario, where they hibernate for multiple months annually and do not reach maturity for 15-20 years, however in Florida, where hibernation is not required, they can reach maturity in as little as four to eight years (Government of Canada 2016; Government of Ontario 2019). Therefore, clutch size is also associated with geographical location. A study by Ashton et al. (2007), found that mean clutch size in Gopher Tortoises (Gopherus polyphemus) decreased with increasing latitude. This study also found that clutch size increased with increasing productivity, was positively correlated with temperature, and negatively correlated with seasonality. Ashton *et al.* (2007) also noted that a decrease in clutch size in the largest individuals can be attributed to the senescence hypothesis. Collins and Crump (2009) suggested that low clutch sizes can be associated with population declines in amphibians.

Rhodin *et al.* (2017) provided detailed distribution maps for almost every extant member of the Testudines Order. Life histories of Testudines make populations more susceptible to threats (Gibbs and Shriver 2002). Siliceo and Diaz (2010) explored the connection between conservation status and clutch and range sizes of lacertid lizards, showing there was a connection between range size and conservation status, as well as clutch size and conservation status. Hero *et al.* (2005) explored a similar topic relating to the decline of amphibians in eastern Australia, finding that conservation status was correlated with both clutch size and geographic range and could aid in predicting a species vulnerability to extinction. According to Rapoport's rule range size increases with increasing latitude (Stevens 1989). Hecnar (1999a) provided evidence that Rapoport's rule is a local effect rather than a general rule. According to Harris and Pimm (2008) "small geographical range size is the best predictor of threat of extinction in terrestrial species".

OBJECTIVE

My objective was to determine if the global conservation status of Testudine species was associated with body size, clutch size, geographical range size, and latitude. To do so I compared IUCN conservation status of turtle species with biological attribute data of interest collected from numerous sources.

NULL HYPOTHESIS

Conservation status of Testudines is not correlated with the attribute of interest.

MATERIALS AND METHODS

I gathered information regarding the global conservation status and biological attributes of Testudines from a wide variety of sources including peer-reviewed journal articles, books, and online resources. I used the IUCN Red List to collect all available information regarding global conservation status and population trends for each listed Testudine. Rhodin *et al.* (2017) was used to complete the list of Testudines to be examined in my study. Information regarding Testudines both listed and not listed on the IUCN Red List was collected via the aforementioned resources. A table featuring data for all 357 turtle species within this study can be found in the Appendix.

I calculated midpoint latitude for all 357 turtle species within this study using maps contained within Rhodin *et al.* (2017). Midpoint was determined as the intersection of lines connecting the most northern and southern extent with the most eastern and western extent. I then used a conspicuous geographical feature, border, or recognizable point to determine the precise midpoint latitude on Google (2019).

I calculated geographical range size using maps contained within Rhodin *et al.* (2017), Iverson (1992b), with unpublished data used for maps in Hecnar (1999b). I calculated the area (km²) of each species range using a dot grid applied to the range map (Iverson 1992b, Hecnar 1999b, or Rhodin *et al.* 2017) and determined the scale from geographic features and Google (2019).

I collected mean female carapace length and clutch size data from multiple peerreviewed journal articles, books, and online resources. If mean clutch size or female carapace length could not be located, but a range could be, I used the midpoint of the range as the mean. For example, an estimated mean clutch size of 15.5 corresponds to the range of 6-25 for *Cyclanorbis senegalensis* (IUCN 2020). A clutch size for

Cyclanorbis elegans could not be found, so I used a published count of 27 oviductal eggs (Demaya et al. 2019). In the case of Actinemys pallida the clutch size and mean female carapace length was found using distribution maps in correlation with two journal articles. It is believed a journal article regarding *Clemmys marmorata* which offered clutch size and mean female carapace length was in fact A. pallida. A later journal article referenced this paper and discussed the turtle by the name A. pallida, therefore the information provided for C. marmorata was used for A. pallida (Lovich and Meyer 2002; Rhodin et al. 2017; Cummings et al. 2018). The clutch size and mean female carapace length within Iverson et al. (1993) for Chelodina oblonga was not used. Kennett *et al.* (2014) describes this as "a fairly large freshwater turtle". The data contained within Iverson et al. (1993) for C. oblonga comes from Clay (1981). The map used in Clay (1981) does not match the updated geographical range of C. oblonga provided in Rhodin et al. (2017). Being that Chelodina siebenrocki is a synonym for C. oblonga, the data provided for C. siebenrocki was used for C. oblonga. In Iverson et al. (1993) Chelodina novaeguineae has a clutch size listed as 10 based on Kennett et al. (1992). Based on distribution maps from Rhodin et al. (2017), as well as the distribution map provided in Kennett et al. (1992) and a journal article by McCord and Thomson (2002) it is believed this was in fact the *Chelodina canni*. Therefore, the provided mean female carapace length and clutch was used for C. canni. The carapace lengths obtained from Powell et al. (2016) were not all stated as female. However, it is presumed the book showed differing lengths for males and females when there was a notable difference.

According to Rhodin *et al.* (2017) *Actinemys pallida* is listed under *Actinemys marmorata*, *Amyda ornata* is listed under *Amyda cartilaginea*, *Chrysemys dorsalis* is

listed under *Chrysemys picta*, *Cuora cyclornata* is listed under *Cuora trifasciata*, *Graptemys sabinensis* is listed under *Graptemys ouachitensis*, *Kinosternon steindachneri* is listed under *Kinosternon subrubrum*, and *Pseudemys floridana* is listed under *Pseudemys concinna*. Therefore, these turtles were given the same global IUCN status and population trends for which they were listed. According to IUCN (2020a) *Chelodina colliei* is listed under *Chelodina oblonga*, therefore it was given the same global IUCN status and population trend as well. For the purpose of this study, this brings the total turtle species globally listed on the IUCN Red List from 258 to 266.

I constructed a database and did some initial analyses using Microsoft Excel. I initially tested data for normality and calculated basic descriptive statistics using Microsoft Excel and SYSTAT 13. To determine the relationship between global IUCN status and biological variables I calculated Pearson correlation coefficients using JASP and SYSTAT 13. I then conducted a Spearman's non-parametric test when transformation did not normalize data for comparison. *T*-tests were run using Microsoft Excel and SYSTAT 13. To explain the association between IUCN status and biological attribute variables I constructed complete and stepwise (forward and backward) multivariate regression models. For descriptive models and comparison, I also used Akaike Information Criterion (AIC) analyses. Regressions and AIC models were analysed using SYSTAT 13.

Pearson correlation tests were run for six differing sets of data (Table 2 and 3). The first included all turtle species with the global IUCN status given an ascending rank from one through eight (not listed = 1, not defined or data deficient = 2, least concern = 3, near threatened = 4, vulnerable = 5, endangered = 6, critically endangered = 7, and extinct in the wild or extinct = 8). The second test included the same data and was run using the same ranking system however the seven extant sea turtles were removed. For the third run, "Not Listed", "Not Defined", and "Data Deficient" were removed as only species with a defined conservation status and the remaining turtle species were given an ascending rank (least concern = 1, near threatened = 2, vulnerable = 3, endangered = 4, critically endangered = 5, and extinct in the wild or extinct = 6). The fourth test included the same data and ranking system as the third however the sea turtles were removed. For test five "Not Listed", "Not Defined", "Data Deficient", "Extinct in the Wild", and "Extinct" were removed, with the remaining species given an ascending rank based on conservation status (least concern = 1, near threatened = 2, vulnerable = 3, endangered = 4, and critically endangered = 5). The sixth test had the same parameters as the fifth, however sea turtles were removed.

Two-sample *t*-tests assuming unequal variances were run for four differing sets of data (Table 4). The confidence interval used for all *t*-tests was 95%. Each of the 30 *t*-tests presented within Table 4 compared means of differing attributes for four differing sets of data. The first set of *t*-tests examined species at risk of extinction. For this set "Not Listed", "Extinct in the wild", and "Extinct" were eliminated from the test, while the remaining status' were listed as either "yes" or "no" (vulnerable, endangered, and critically endangered = yes and not defined, data deficient, least concern, and near threatened = no). The second set used every species of turtle and compared the means of listed and not listed species for differing attributes (not listed = no and not defined, data deficient, least concern, near threatened, vulnerable, endangered, critically endangered, extinct in the wild, and extinct = yes). The third set examined the differences between suborders and included their IUCN global status, each given an ascending rank (not listed = 1, not defined or data deficient = 2, least concern = 3, near threatened = 4,

vulnerable = 5, endangered = 6, critically endangered = 7, and extinct in the wild or extinct = 8). The fourth set of *t*-tests were the same as the third however excluded sea turtles.

Additional two sample *t*-tests analyzing suborders (Cryptodires vs. Pleurodires) were completed using SYSTAT 13. An ascending rank for each species status category was assigned (least concern = 1, near threatened = 2, vulnerable = 3, endangered = 4, critically endangered = 5, and extinct in the wild and extinct = 6). For a non-parametric equivalent, a Mann-Whitney *U* tests was run comparing the global IUCN status for Cryptodira and Pleurodira using the same ascending numbers. These will help examine if there is a significant difference between each suborder.

The final set of tests was multiple types of regressions using SYSTAT 13. First a complete model was estimated with status rank as the dependent variable and geographical range size (km^2), midpoint latitude, average clutch size (n), and mean female carapace length (mm) were the independent variables. For each species, negative value latitudes (southern hemisphere midpoint) were made positive by first squaring the value, followed by square rooting the value so that a global-scale geographic assessment was possible. The first regression run was a linear regression. This test can help explain how much the attribute variables contribute to the dependent variable, which in this case was global IUCN status, as well as if this contribution is statistically significant. This test included all turtles listed on the IUCN except "Not Defined" and "Data Deficient" and each were assigned an ascending number (1 = least concern, 2 = near threatened, 3 = vulnerable, 4 = endangered, 5 = critically endangered, and 6 = extinct in the wild and extinct). The next test run was the backward stepwise regression, which examines each variable's contribution and constructs the best model by eliminating the weakest

variable(s) first. The third test run was the forward stepwise regression, which looks for the strongest variable(s) first. Both of these tests aid in explaining what variables contribute to the global IUCN status and whether this contribution is significant. For each of the backward and forward stepwise regression tests the same ascending numbers for global IUCN status were used. Finally, an Akaike Information criterion (AIC) test was run. This test attempts to show the best combination of predictor variables as displayed by the lowest score.

RESULTS

Status ranks were available for 266 species (Figure 2). The category with the greatest number of species was "Not Listed" with 91. Numbers in the global IUCN categories ranged from one in the "Not Defined" and "Extinct in the Wild" to "Vulnerable" with 69.

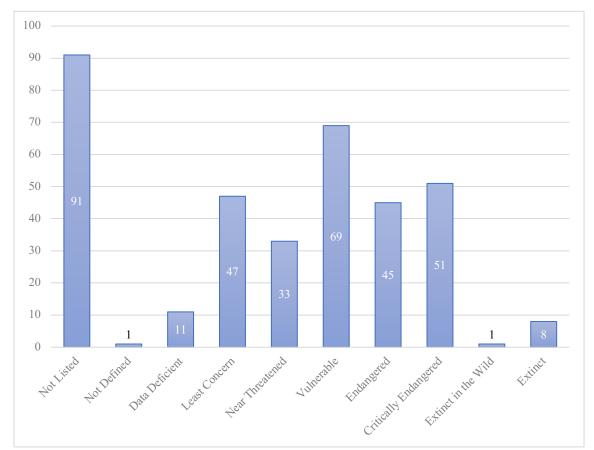


Figure 2. Number of species within each global IUCN Red List category, as well as number of species not listed.

Categories with the highest and lowest mean geographical range (km²) are "Vulnerable" with 4,350,111 +/- 2,052,324.5 km² (SEM) and "Extinct" with 819 +/-362.1 km² (SEM). Categories with the highest and lowest mean midpoint latitude are "Least Concern" with 19.27 (range -32.87 to 43.45°) and "Extinct" with -13.31 (range -21.14 to 0.57°). Categories with the highest and lowest mean clutch size are "Not Defined" with 97.1 and "Near Threatened" with 6.6 +/- 0.89 (SEM). The "Not Defined" category contains one turtle and the category with the second highest mean clutch size is "Extinct in the Wild". "Extinct in the Wild" also contains one species and the category with the third highest mean clutch size is "Vulnerable" with 17.1 +/- 3.16 (SEM). The categories with the highest and lowest mean female carapace length (mm) are "Extinct" with 680 mm and "Near Threatened" with 194.7 mm (Figure 3).

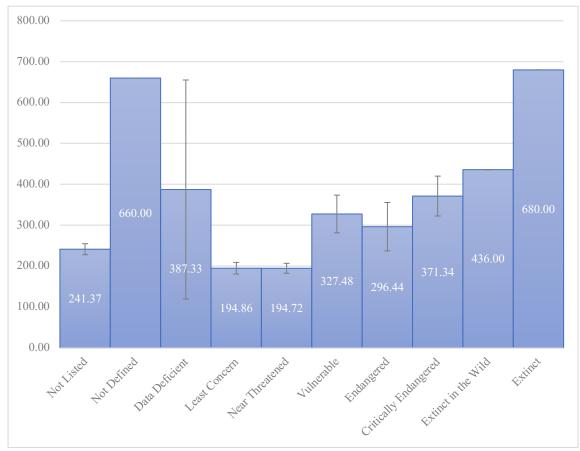


Figure 3. Mean female carapace length (mm) for each category. Bars indicate standard error of the mean.

Excluding sea turtles shows their potential effect on data analysis. The "Data Deficient" category drops drastically from 387.33 mm to 119.5 mm. The "Vulnerable" category drops from 327.48 mm to 263.63 mm, dropping below the "Endangered" category (Figure 4).

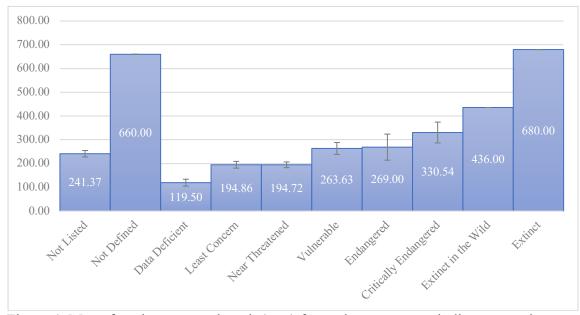


Figure 4. Mean female carapace length (mm) for each category excluding sea turtles. Bars indicate standard error of the mean.

There were 231 turtle species with a midpoint latitude in the northern hemisphere and 126 in the southern hemisphere. Forty-five species in the southern hemisphere were not listed, while 46 in the northern hemisphere were not listed. Seven of the eight extinct species were from the southern hemisphere. Sixty-four percent of the southern hemisphere species were listed on the IUCN, while 80% of the northern hemisphere species were listed. The mean female carapace length of northern hemisphere species was 263 +/- 18.7 mm (SEM) and for southern hemisphere species 296 +/- 26.9 mm (SEM). There was no significant difference in category status (all categories included) between hemispheres (t = -1.58, 222 df, P = 0.115).

There are 94 species within the suborder Pleurodira and 263 species within the Cryptodira suborder. The mean midpoint latitude of Pleurodira was -10.01 (range -32.70 to 16.23°) and for Cryptodira 15.37 (range -33.31 to 45.57°), while the mean geographical range size was 725,993 +/- 104,712.1 km² (SEM) and 2,458,894 +/- 916,720.2 km² (SEM) respectively. The mean geographical range size of Cryptodira

dropped to $631,700 \pm 72,812.8 \text{ km}^2$ (SEM) when sea turtles were removed. Mean clutch sizes of Pleurodira and Cryptodira were 14.22 ± 1.8 (SEM) and 12.04 ± 1.4 (SEM) respectively. The mean clutch size of Cryptodira dropped to 8.9 ± 0.74 (SEM) when sea turtles were removed. The mean female carapace lengths are shown in Figure 5.

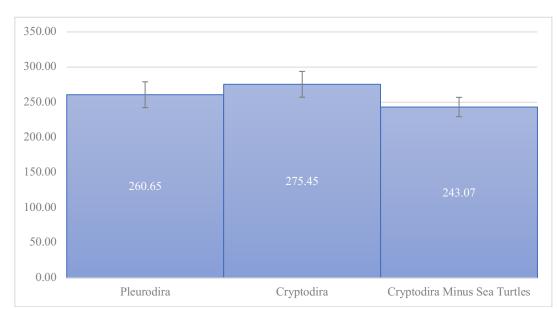


Figure 5. Mean female carapace length (mm) for suborders. Cryptodira is shown both including and excluding sea turtles. Bars indicate standard error of the mean

Geographical range size varied among families (Table 1). The family with the largest mean geographical range size is Dermochelyidae, with one extant member, the Leatherback Sea Turtle (*Dermochelys coriacea*). The family with the second highest mean geographical range is Cheloniidae which contains the remaining sea turtles. The family with the lowest mean geographical range was Staurotypidae. Families with the highest and lowest mean midpoint latitudes are Emydidae and Chelidae respectively. The families with the highest and lowest mean clutch sizes are Cheloniidae and Platysternidae respectively. Families with the highest and lowest mean female carapace length (mm) are Dermochelyidae and Kinosternidae respectively (Table 1).

| Family | Mean Geographical Range Size (+/- SEM km ²) | Mean Midpoint Latitude (°) | Mean Clutch Size (+/- SEM n) | Mean Female Carapace Length (+/- SEM mm) |
|------------------|--|----------------------------------|---------------------------------|--|
| Carettochelyidae | 435329 +/- N/A | -11.254 | 15.0 +/- N/A | 457 +/- N/A |
| Chelidae | 551615 +/- 128712 | -15.022 | 10.4 +/- 0.84 | 229 +/- 13.3 |
| Cheloniidae | 66482194 +/- 28936490 | 3.476 | 104.6 +/- 12.28 | 794 +/- 49.1 |
| Chelydridae | 1197756 +/- 894642 | 25.784 | 29.2 +/- 1.98 | 366 +/- 57.0 |
| Dermatemydidae | 137221 +/- N/A | 17.320 | 17.9 +/- N/A | 470 +/- N/A |
| Dermochelyidae | 86080919 +/- N/A | 7.876 | 79.8 +/- N/A | 1470 +/- N/A |
| Emydidae | 511025 +/- 132055 | 29.212 | 9.4 +/- 0.75 | 199 +/- 10.2 |
| Geoemydidae | 442371 +/- 65366 | 17.922 | 6.0 +/- 0.92 | 242 +/- 24.1 |
| Kinosternidae | 630701 +/- 290664 | 23.342 | 3.1 +/- 0.32 | 114 +/- 4.6 |
| Pelomedusidae | 977539 +/- 194988 | -1.943 | 16.7 +/- 2.91 | 244 +/- 11.7 |
| Platysternidae | 1120779 +/- N/A | 21.501 | 2.3 +/- N/A | N/A |
| Podocnemididae | 1109829 +/- 369015 | -1.906 | 30.9 +/- 10.60 | 418 +/- 69.1 |
| Staurotypidae | 124354 +/- 42794 | 16.900 | 6.7 +/- 2.02 | 190 +/- 52.5 |
| Testudinidae | 718761 +/- 181334 | -1.914 | 5.4 +/- 0.72 | 286 +/- 34.4 |
| Trionychidae | 1040968 +/- 256546 | 17.223 | 24.7 +/- 4.53 | 464 +/- 87.1 |

Table 1. Biological attribute means for each family.

There are 15 families within the Testudines order. The family with the lowest rank based on listing and global IUCN status is Pelomedusidae and the highest is Dermatemydidae (Figure 6).

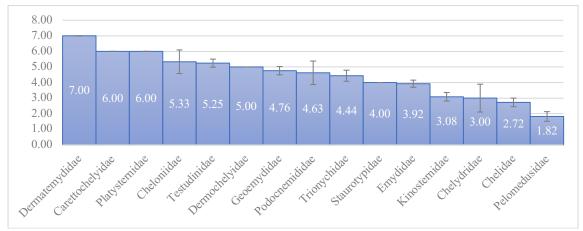


Figure 6. Mean global IUCN status for each family. For this histogram each turtle species was given a rank based on global IUCN status and the mean of these ranks for each family was calculated (not listed = 1, not defined or data deficient = 2, least concern = 3, near threatened = 4, vulnerable = 5, endangered = 6, critically endangered = 7, and extinct in the wild or extinct = 8). Bars indicate standard error of the mean.

The positive correlation between global IUCN status and both mean female carapace length (mm) and LOG mean female carapace length (mm) were highly significant in all six tests, regardless of differing data inclusion parameters. Global IUCN status was negatively correlated and significant with geographical range size in three of six tests, each of which excluded sea turtles. The correlation between global IUCN status and both average clutch size and square root average clutch size was variable in terms of positive and negative and not significant in all but one test, which was marginally significant (Table 2).

Table 2. Pearson correlation test results for Global IUCN Status vs. biological attributes for test one through six. The *r* values are on top of the corresponding shaded *p*-values. Significant results are bolded.

| Biological Attribute | Run One | Run Two | Run Three | Run Four | Run Five | Run Six |
|-------------------------------|---------|---------|--------------|----------|----------|---------|
| Range Size (km ²) | 0.055 | -0.212 | 0.014 | -0.282 | 0.028 | -0.264 |
| Range Size (Rin) | 0.296 | <0.001 | 0.0824 | <0.001 | 0.668 | <0.001 |
| LOG Range | -0.357 | -0.406 | -0.419 | -0.469 | -0.332 | -0.386 |
| LOO Kalige | <0.001 | <0.001 | <0.001 | <0.001 | <0.001 | <0.001 |
| Midpoint Latitude (°) | 0.091 | 0.088 | -0.201 | -0.203 | -0.144 | -0.145 |
| Midpoliti Latitude () | 0.087 | 0.099 | 0.001 | 0.001 | 0.025 | 0.025 |
| Average Clutch Size (n) | 0.079 | -0.018 | 0.111 | 0.042 | 0.108 | 0.032 |
| Average Cluten Size (ii) | 0.2 | 0.772 | 0.108 | 0.548 | 0.117 | 0.645 |
| SQRT Clutch | 0.034 | -0.044 | 0.068 | -0.003 | 0.061 | -0.016 |
| | 0.576 | 0.482 | 0.325 | 0.962 | 0.38 | 0.815 |
| Mean Female CL (mm) | 0.229 | 0.221 | 0.331 | 0.345 | 0.303 | 0.299 |
| | 0.003 | 0.004 | <0.001 | <0.001 | <0.001 | <0.001 |
| LOG CL | 0.17 | 0.132 | 0.318 | 0.29 | 0.284 | 0.244 |
| | 0.026 | 0.09 | <0.001 | <0.001 | <0.001 | <0.001 |

Average clutch size and mean female carapace length were highly correlated in all six tests. The correlation between geographical range size (km²) and midpoint latitude was not significant in any of the Pearson correlation tests (Table 3).

| results are bolded. | | | | | | |
|---|---------|---------|--------------|----------|----------|---------|
| Biological Attributes | Run One | Run Two | Run Three | Run Four | Run Five | Run Six |
| Range Size (km ²) vs. | -0.025 | 0.001 | -0.04 | 0.096 | -0.048 | 0.078 |
| Midpoint Latitude (°) | 0.632 | 0.989 | 0.523 | 0.131 | 0.458 | 0.229 |
| Range Size (km ²) vs. | 0.602 | 0.133 | 0.648 | 0.22 | 0.648 | 0.223 |
| Average Clutch Size (n) | <0.001 | 0.032 | <0.001 | 0.001 | <0.001 | 0.001 |
| Range Size (km ²) vs. | 0.495 | 0.123 | 0.547 | 0.228 | 0.549 | 0.232 |
| SQRT Clutch | <0.001 | 0.049 | <0.001 | 0.001 | <0.001 | <0.001 |
| Range Size (km ²) vs. | 0.5 | 0.094 | 0.532 | 0.147 | 0.545 | 0.171 |
| Mean Female CL (mm) | <0.001 | 0.227 | <0.001 | 0.091 | <0.001 | 0.051 |
| Range Size (km ²) vs. | 0.371 | 0.111 | 0.401 | 0.177 | 0.413 | 0.197 |
| LOG CL | <0.001 | 0.156 | <0.001 | 0.042 | <0.001 | 0.024 |
| LOG Range vs. Midpoint | 0.126 | 0.144 | 0.235 | 0.268 | 0.175 | 0.21 |
| Latitude (°) | 0.017 | 0.007 | <0.001 | <0.001 | 0.006 | 0.001 |
| LOG Range vs. Average | 0.357 | 0.168 | 0.411 | 0.213 | 0.412 | 0.214 |
| Clutch Size (n) | <0.001 | 0.007 | <0.001 | 0.002 | <0.001 | 0.002 |
| LOG Range vs. SQRT | 0.326 | 0.157 | 0.392 | 0.207 | 0.393 | 0.209 |
| Clutch | <0.001 | 0.011 | <0.001 | 0.003 | <0.001 | 0.003 |
| LOG Range vs. Mean | 0.185 | -0.113 | 0.2 | -0.119 | 0.265 | -0.045 |
| Female CL (mm) | 0.015 | 0.148 | 0.018 | 0.172 | 0.002 | 0.612 |
| LOG Range vs. LOG CL | 0.142 | -0.063 | 0.151 | -0.073 | 0.212 | -0.011 |
| Loo Runge vs. Loo CL | 0.062 | 0.425 | 0.077 | 0.401 | 0.013 | 0.9 |
| Midpoint Latitude (°) vs. | -0.042 | -0.025 | -0.028 | 0.024 | -0.03 | 0.022 |
| Average Clutch Size (n) | 0.495 | 0.683 | 0.682 | 0.729 | 0.671 | 0.76 |
| Midpoint Latitude (°) vs. | -0.035 | -0.016 | 0.005 | 0.052 | 0.003 | 0.048 |
| SQRT Clutch | 0.568 | 0.8 | 0.94 | 0.462 | 0.966 | 0.493 |
| Midpoint Latitude (°) vs. | -0.143 | -0.128 | -0.134 | -0.115 | -0.13 | -0.109 |
| Mean Female CL (mm) | 0.061 | 0.102 | 0.116 | 0.187 | 0.131 | 0.214 |
| Midpoint Latitude (°) vs. | -0.172 | -0.156 | -0.133 | -0.111 | -0.131 | -0.108 |
| LOG CL | 0.024 | 0.045 | 0.118 | 0.203 | 0.128 | 0.221 |
| Average Clutch Size (n) vs. Mean Female CL | 0.777 | 0.73 | 0.773 | 0.748 | 0.773 | 0.746 |
| (mm) | <0.001 | <0.001 | <0.001 | <0.001 | <0.001 | <0.001 |
| Average Clutch Size (n) vs. LOG CL | 0.688 | 0.641 | 0.678 | 0.654 | 0.679 | 0.651 |
| | <0.001 | <0.001 | <0.001 | <0.001 | <0.001 | <0.001 |
| SQRT Clutch vs. Mean | 0.81 | 0.745 | 0.812 | 0.748 | 0.811 | 0.745 |
| Female CL (mm) | <0.001 | <0.001 | <0.001 | <0.001 | <0.001 | <0.001 |
| SORT Clutch ve I OC CI | 0.782 | 0.734 | 0.777 | 0.74 | 0.777 | 0.736 |
| SQRT Clutch vs. LOG CL | <0.001 | <0.001 | <0.001 | <0.001 | <0.001 | <0.001 |

Table 3. Pearson correlation test results for biological attributes for test one through six.The r values are on top of the corresponding shaded p-values. Significantresults are bolded.

None of the geographical range size *t*-tests showed a statistically significant difference between means (Table 4). The midpoint latitude *t*-tests showed statistically significant differences in three of four tests. Both *t*-tests that included global IUCN status showed a statistically significant difference. These tests were also run on the suborders, one including all 357 turtles and the other excluding the seven sea turtles (Table 4).

| <i>t</i> -test | At Risk of Extinction | Listed | Suborder | Suborder (no sea turtles) |
|----------------------------|--------------------------|--------|----------|------------------------------|
| Geographical | 1.767 | 1.661 | 1.878 | -0.739 |
| Range Size | 167 | 275 | 269 | 189 |
| (km ²) | 0.079 | 0.098 | 0.061 | 0.461 |
| | -3.474 | -5.051 | -1.360 | -1.909 |
| LOG Range | 254 | 295 | 200 | 192 |
| | <0.001 | <0.001 | 0.175 | 0.058 |
| Milling | -1.486 | 4.511 | 14.806 | 14.889 |
| Midpoint Latitude (°) | 161 | 182 | 241 | 244 |
| Latitude () | 0.139 | <0.001 | <0.001 | <0.001 |
| Assessed Classific | 2.123 | 1.476 | -0.956 | -2.721 |
| Average Clutch Size (n) | 212 | 180 | 136 | 78 |
| 512e (11) | 0.035 | 0.142 | 0.341 | 0.008 |
| | 1.750 | 0.559 | -2.720 | -4.192 |
| SQRT Clutch | 214 | 99 | 127 | 93 |
| | 0.082 | 0.577 | 0.007 | <0.001 |
| Mean Female CL | 3.534 | 1.660 | 0.569 | -0.765 |
| (mm) | 115 | 135 | 102 | 71 |
| (11111) | <0.001 | 0.099 | 0.571 | 0.447 |
| | 3.355 | -0.318 | -1.194 | -2.160 |
| LOG CL | 132 | 83 | 82 | 72 |
| | 0.001 | 0.751 | 0.236 | 0.034 |
| | N/A | N/A | 7.552 | 7.434 |
| Global IUCN Status | N/A | N/A | 161 | 164 |
| Status | N/A | N/A | <0.001 | <0.001 |

 Table 4. Results of various *t*-tests. Statistically significant tests are bolded. For each test the following are listed in descending order: *t*-value, df value, *p*-value.

Global IUCN status did not differ between orders Pleurodira vs. Cryptodira (separate variance t = 1.37, 52.18 df, P = 0.177; pooled variance t = 1.33, 252 df, P =

0.186). A non- parametric equivalent test concurred (Mann-Whitney U = 4,682, 1 df, P = 0.57).

A complete multivariate regression model (geographical range size, midpoint latitude, carapace length, and clutch size) explained 14.3% of the global IUCN status value (F = 5.44; 4, 130 df; P < 0.001; $R^2 = 0.143$).

Both forward and backward stepwise multiple regression produced the same

highly significant model with midpoint latitude and carapace length (F = 12.24; 2, 136

df; P < 0.001; $R^2 = 0.153$).

The best AIC model found included latitude, carapace length, and clutch size,

and explained 13.4% of the IUCN status. This model produced essentially the same

level of description (Δ AIC<2) as the complete model which explained 14.3% (Table 5).

Table 5. Results of the AIC test. The individual or combination of attributes with the lowest AIC score indicates the best option. The *R* squared value, derived from multiple regression, represents the percent the correlating attributes can explain the global IUCN status. Note: Lat = midpoint latitude (°), CL = mean female carapace length (mm), Clutch = clutch size (n), and Range = geographical range size (km²). Schwarz Criterion also shown (BIC).

| enterion dise she wit (Bre | -). | | | |
|----------------------------|-------|-------|-------|-------|
| Attributes | AIC | AICc | BIC | R^2 |
| Lat+CL+Clutch | 469.4 | 469.9 | 483.9 | 0.134 |
| Range+Lat+CL+Clutch | 470 | 470.6 | 487.4 | 0.143 |
| Range+CL+Clutch | 474.3 | 474.8 | 488.8 | 0.102 |
| Range+Lat+CL | 480.4 | 480.9 | 495.1 | 0.167 |
| Lat+CL+Clutch | 480.8 | 481.1 | 492.5 | 0.153 |
| CL | 485.6 | 485.8 | 494.4 | 0.11 |
| Range+CL | 485.7 | 486 | 497.4 | 0.122 |
| Lat+Clutch | 736.9 | 737.1 | 750.3 | 0.066 |
| Range+Lat+Clutch | 737.9 | 738.2 | 754.7 | 0.071 |
| Clutch | 746.7 | 746.8 | 756.8 | 0.012 |
| Range+Clutch | 748.2 | 748.4 | 761.9 | 0.015 |
| Lat | 899.9 | 900 | 910.5 | 0.074 |
| Range+Lat | 901.6 | 901.8 | 915.8 | 0.075 |
| Range | 919.4 | 919.5 | 930 | 0 |

DISCUSSION

The results include 91 turtle species not listed on the global IUCN Red List. This can be misleading, as this is not an indication of assessment but merely an indication of listing. According to Rhodin et al. (2017) some of the species within the "Not Listed" category in this study have been assessed, while others have not. For the purpose of this study, there were considered to be 266 turtle species listed on the global IUCN Red List. This is eight more than were listed on the IUCN Red List at time of data collection and reflects taxonomic revisions separating species that were included within the global IUCN Red List listing of other species. Although unlikely, it is possible that this was the case for other species as well and it was missed during the research process. There has been a species added to the global IUCN Red List since time of data collection. *Elseva rhodini*, which was not listed at time of data collection, has been added to the "Least Concern" category (IUCN 2020b). The addition of this species as well as the eight species included for the purpose of this study brings the total species listed on the IUCN Red List to 267. This study recognized a total of 357 species therefore this would indicate 74.8% of Testudines are listed on the IUCN Red List. As Rhodin et al. (2017) and IUCN (2019f) were used to compile the list of 357 species, the newly discovered *Pelodiscus variegatus* was not included in this study (Farkas et al. 2019).

The results show the "Vulnerable" category having the highest mean geographical range size with a value of 4,350,111 km². This is misleading as the "Vulnerable" category contains three sea turtle species, which undoubtedly inflates the mean geographical range size. The mean geographical range size of non-sea turtle Testudines is 657,024.2 km², but 69,282,011.5 km² for sea turtles. When removing the sea turtles, the "Vulnerable" category saw its mean geographical range drop to 807,694.6 km². The "Not Listed" category then had the highest mean geographical range with a value of 869,328.9 km². This begins to show the impact the inclusion or exclusion of sea turtles can have on the analyses of data for Testudines. This was an issue recognized by Hecnar (1999a) as sea turtles were not included in a study exploring geographic range sizes "because their range sizes are poorly known and they have a different mode of life". The impact of sea turtles is seen in other attributes as well. For example, the "Vulnerable" category has a mean female carapace length of 327 mm including sea turtles but 264 mm excluding sea turtles, dropping below the "Endangered" category which falls from 296 mm to 269 mm, as seen in the results section in Figures 3 and 4.

The impact of sea turtles is further exemplified within the results of the Pearson correlation tests (Table 2). Each run excluding sea turtles shows a statistically significant correlation between geographic range size and global IUCN status, but each run including them shows the opposite. A similar effect is seen within the Pearson correlation tests between midpoint latitude and LOG CL, LOG range and LOG CL, and LOG range and mean female carapace length (mm). In each of these cases, the tests ran including sea turtles showed a significant correlation and the test ran excluding sea turtles showed no significant correlation. This difference is once again seen when examining the mean clutch size of all species between Cryptodira and Pleurodira. A *t*-test showed no significant difference when including sea turtles but showed a significant difference when including sea turtles but showed a significant difference when including sea turtles but showed a significant difference when including sea turtles but showed a significant difference when including sea turtles but showed a significant difference when including sea turtles but showed a significant difference when including sea turtles but showed a

This however was not the case when examining the global IUCN status of suborders using *t*-tests. When including all categories involved in this study there was a

strongly significant difference between suborders whether sea turtles were included or not. However, when excluding "Not Listed", "Not Defined", and "Data Deficient" and examining global IUCN status of suborders the separate and pooled variance tests showed no significant difference with *p*-values of 0.177 and 0.186 respectively. This can also begin to exemplify the impact of including and excluding differing categories during analyses of data.

Mean female carapace length appeared to be the attribute most strongly related to the global IUCN status. This can be observed in the Pearson correlation tests as each of the six runs show a significant correlation between mean female carapace length and global IUCN status. This connection was again evident when a *t*-test found a significant difference between mean female carapace length of species at risk and not at risk of extinction. Both multiple regression models included mean female carapace length and these tests were highly statistically significant (F = 12.24; 2, 136 df; P < 0.001; $R^2 =$ 0.153). The standard coefficient of midpoint latitude and mean female carapace length was -0.231 and 0.23 respectively, meaning about 46% of the *R* squared value can be explained by these two variables. The tolerance values for each was 0.81, meaning these two variables are relatively independent of one another.

This connection was also seen in the AIC test as mean female carapace length was included in each of the top seven AIC scores and zero of the remaining seven options. It also had by far the best individual score with 485, with the remaining attributes scoring 746.7-919.4 individually. The same AIC test indicated that mean female carapace length (mm) could explain 11% of a species global IUCN status, while midpoint latitude, clutch size, and geographical range offered 7.4%, 1.2%, and 0% respectively. If this correlation is correct, it would not be specific to turtles as animals

with large body size are often more susceptible to extinction (Cardillo *et al.* 2005). As expected, there was also a strong connection between mean female carapace length and clutch size.

The connection between global IUCN status and mean female carapace length may be related to anthropogenic activities. There are many places on Earth where humans eat turtles and it is logical to deduce that they would prefer to eat habitually larger turtles. Conway-Gomez (2007) examined two species within the Podocnemis genus and found that turtle abundance was negatively impacted by hunting pressure and this negative impact was positively correlated with proximity to human communities. The *Podocnemis* genus in my study had a mean female carapace length of 442 mm, which is high. Only one of six Podocnemis turtles were not listed on the IUCN Red List, one of these is critically endangered. Conway-Gomez (2007) examined both Podocnemis expansa and P. unifilis, which are both listed on the IUCN Red List, P. *unifilis* being vulnerable and thus considered at risk of extinction. Humans eating turtles is by no means a new phenomenon. Frazier et al. (2018) found evidence of the intense exploitation of sea turtles in Oman up to 6,500 years ago, as well as evidence that some communities of humans worshipped turtles, finding turtle bones and skulls within human graves. This indicates turtles were not only utilized as a nutritional resource but offered cultural significance as well.

According to Harris and Pimm (2008) "small geographical range size is the single best predictor of threat of extinction in terrestrial species". This notion of a correlation between risk of extinction and small geographic range sizes is also offered in Runge *et al.* (2015). My study did not show geographical range size as the most significant factor in determining global IUCN status. The Pearson correlation tests

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showed a significant correlation between geographic range size and IUCN status for the three out of six runs which excluded the sea turtles. The *t*-tests examining species that are and are not at risk of extinction and the listed vs. non listed turtles both showed no significant difference. This attribute was not included in the forward stepwise multiple regression and was eliminated in the backward stepwise multiple regression while also having the lowest individual attribute AIC score and *R* squared value in the AIC test results. As supported by the Pearson correlation tests, the lack of importance in determining IUCN status found within this study may be directly tied to the inclusion of sea turtles in data analysis.

It is somewhat surprising that clutch size was not more important in determining global IUCN status, as mean female carapace length (mm) and clutch size are so highly correlated. All six Pearson correlation tests showed a highly significant correlation between clutch size and mean female carapace length (mm). Only one of six Pearson correlation tests showed a significant correlation between global IUCN status and clutch size, however this was marginally significant. Clutch size even had the second lowest *R* squared value in the AIC test, explaining a mere 1.2% of the global IUCN status. There was however shown to be a significant difference between the clutch sizes of the at risk and not at risk of extinction turtles. This likely stems from its association with mean female carapace length (mm) as the mean female carapace length of species that are at risk of extinction and not at risk of extinction were 330 +/- 29.5 mm (SEM) and 211 +/- 16.3 mm (SEM) respectively.

There are several potential sources of error within this study as well as attributes that were not explored. Because of the process used in calculating geographical range size (km²) and midpoint latitude, it is impossible to state the values for each species are

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100% accurate. Although minimal, this could impact the geographical range size category. As the range extent was used to calculate geographical range, the calculated value could be inflated in comparison to area of occupancy. Ramesh (2017) noted range extent often includes unsuitable habitat resulting in the inflation of a species range. The range extent is found by calculating the area within the boundary whereas area of occupancy seeks to eliminate areas of unsuitable or unoccupied habitat and includes the areas within the range extent a species occupies (Figure 7) (IUCN 2001).

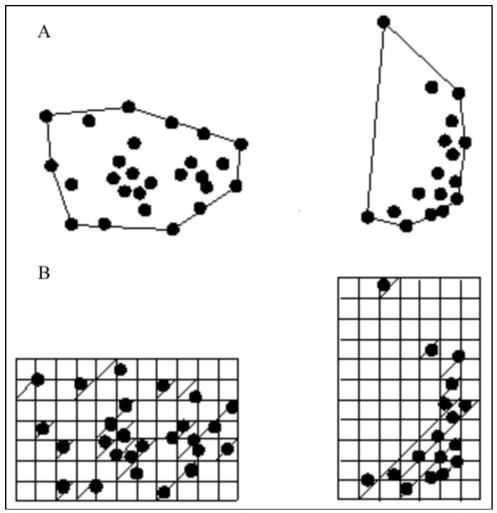


Figure 7. Range extent (A) vs. area of occupancy (B). In range extent the area is calculated as the area within the boundary. For area of occupancy the area is calculated by summing the occupied squares (as marked by the diagonal lines). Notice how different results could be produced using each method (Adapted from IUCN 2001).

The outer boundary of a range also may not always be representative of species abundance. Fortin *et al.* (2005) stated "it is important to identify internal distribution of abundance within range boundaries". The shape of polygons used to outline distributions can also be a point of controversy. The issue can be illustrated using the distribution map of the Common Snapping Turtle (*Chelydra serpentina*) provided within Rhodin *et al.* (2017) and used in this study to calculate geographical range size and midpoint latitude (Figure 8).



Figure 8. Common Snapping Turtle distribution map (Rhodin et al. 2017).

Notice the most northwesterly portion of the distribution has very few distribution dots in comparison to much of the map and the range extends beyond distribution points. A study by Palminteri *et al.* (2011) explored the use of range polygons, acknowledging their shortcomings and their potential for overestimation.

The process of calculating midpoint latitude would not result in 100% accuracy, meaning the midpoint latitude values are not perfect. The analyses may have been impacted by missing data as only 267 clutch sizes and 172 mean female carapace lengths were located during the research process. If there was more available data, it could have impacted the results. For example, only one mean female carapace length was found for a species listed as "Extinct".

The subject of this paper could have been further explored with the inclusion of attributes such as longitude. Examination of continents could have helped further explore the connection between IUCN status and various attributes, as well as exploring the connection between continent located and IUCN status. Analyses of continents may help explore regional anthropogenic effects.

It is interesting to note that four of the top six families in terms of mean global IUCN status contained one turtle. The two that did not were Cheloniidae (mean global IUCN status: 5.33) and Testudinidae (mean global IUCN status: 5.25). Testudinidae is the most terrestrial of the Testudine families (Wyneken *et al.* 2008). Cheloniidae are negatively impacted by fishing practices and plastic pollutants, while also being dependent on beaches for nesting (Assuncao Ivar do Sul *et al.* 2010; WWF 2020). Over 33% of the global human population lives within 100 km of an ocean's coastline (NASA 2020). It is fair to wonder if it is their inherent relationship with humans that has caused high mean global IUCN status.

CONCLUSION

Global biodiversity is currently under attack (Ceballos et al. 2017). Many Testudine species are at risk of extinction and there is a need for increased conservation efforts. Testudines are the subject of taxonomic bias, as they are often underrepresented in wildlife journals, post-secondary library holdings, and reintroduction programs (Seddon et al. 2005; Hecnar 2009; Christoffel and Lepczyk 2012). Although facing many threats, one of the most serious threats to Testudines is habitat loss (Lesbarreres et al. 2014). My study indicates that there is a positive correlation between female carapace length and global IUCN status. This was supported by Pearson correlation tests, forward and backward stepwise multiple regressions, as well as an AIC test. It is possible that poaching and human consumption plays a role in this correlation. The impact of including sea turtles in analyses was evident throughout this study as they often have larger clutch and body sizes as well as significantly larger geographical range sizes than other Testudines. The two families containing greater than one turtle with the highest mean global IUCN status are Cheloniidae and Testudinidae. There are many threats facing Testudines today. Humans can potentially have a positive impact on the current state of Testudines by offering more time and funding to the conservation of Testudines while also minimizing or reversing human caused threats such as habitat loss or fragmentation, road mortality, and poaching.

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APPENDIX

APPENDIX I

| Scientific Name | Suborder | Family | Global IUCN Status | РТ | Range | Midpoint | CS | CL | Sources |
|-------------------------------|------------|------------------|-----------------------|--------------------|-------------|------------|---------|---------|--|
| Acanthochelys macrocephala | Pleurodira | Chelidae | Near Threatened | Decreasing | 448346.8811 | -19.591839 | 6 | | (Iverson 1992b; Hecnar 1999b; Mocelin et al. 2008; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Acanthochelys pallidipectoris | Pleurodira | Chelidae | Endangered | Decreasing | 245148.6108 | -25.97381 | 3.5 | 144 | (Rhodin et al. 2017; Google 2019; IUCN 2019f; IUCN 2020c) |
| Acanthochelys radiolata | Pleurodira | Chelidae | Near Threatened | Unspecified | 218949.2576 | -15.872541 | 3 | 165 | (Mocelin et al. 2008; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Acanthochelys spixii | Pleurodira | Chelidae | Near Threatened | Unspecified | 419440.2731 | -24.473112 | 4 | 139.78 | (Mocelin et al. 2008; Neto et al. 2011; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Actinemys marmorata | Cryptodira | Emydidae | Vulnerable | Unspecified | 322550.9905 | 42.706168 | 4.65 | 145.75 | (Iverson 1992b; Iverson et al. 1993; Hecnar 1999b; Germano and Riedle 2015; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Actinemys pallida | Cryptodira | Emydidae | Vulnerable | Unspecified | 85408.86535 | 33.781446 | 4.46 | 144 | (Lovich and Meyer 2002; Rhodin et al. 2017; Cummings et al. 2018; Google 2019) |
| Amyda cartilaginea | Cryptodira | Trionychidae | Vulnerable | Unspecified | 1015790.054 | 0.263773 | 19.875 | | (Kusrini 2015; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Amyda ornata | Cryptodira | Trionychidae | Vulnerable | Unspecified | 921208.9425 | 17.994226 | 19.075 | | (Rhodin <i>et al.</i> 2017; Google 2019) |
| Apalone ferox | Cryptodira | Trionychidae | Least Concern | Unknown | 296126.3952 | 29.24447 | 20.04 | 339.3 | (Iverson 1992b; Iverson et al. 1993; Hecnar 1999b; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Apalone mutica | Cryptodira | Trionychidae | Least Concern | Unknown | 1230399.945 | 38.099586 | 13.57 | 211 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| | | Trionychidae | Least Concern | Stable | 2919573.137 | 38.121094 | 25.315 | 372.5 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Apalone spinifera | Cryptodira | | | | | | 8.93 | | (Iverson 1992b; Iverson et al. 1993; Hecnar 1999b; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Astrochelys radiata | Cryptodira | Testudinidae | Critically Endangered | Decreasing | 17302.94103 | -24.406753 | | 336.6 | |
| Astrochelys yniphora | Cryptodira | Testudinidae | Critically Endangered | Decreasing | 1005.568986 | -16.106108 | 3.555 | 370.1 | (Iverson 1992b; Hecnar 1999b; Pedrono et al. 2001; Pedrono and Markwell 2001; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019 |
| Batagur affinis | Cryptodira | Geoemydidae | Critically Endangered | Decreasing | 150319.8329 | 3.844432 | 26 | 525 | (Rhodin et al. 2017; Google 2019; IUCN 2019f; IUCN 2020d) |
| Batagur baska | Cryptodira | Geoemydidae | Critically Endangered | Decreasing | 158561.326 | 17.03779 | 24.39 | 488 | (Iverson et al. 1993; Jenkins 1995; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Batagur borneoensis | Cryptodira | Geoemydidae | Critically Endangered | Unspecified | 320769.5722 | 1.902391 | 12.49 | 466 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Batagur dhongoka | Cryptodira | Geoemydidae | Critically Endangered | Decreasing | 513488.3981 | 25.777057 | 26.165 | 420 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Batagur kachuga | Cryptodira | Geoemydidae | Critically Endangered | Decreasing | 283040.5921 | 25.618661 | 17 | 560 | (Iverson et al. 1993; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Batagur trivittata | Cryptodira | Geoemydidae | Critically Endangered | Decreasing | 123367.2168 | 21.014693 | 25 | | (Furtado 1988; Jenkins 1995; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Caretta caretta | Cryptodira | Cheloniidae | Vulnerable | Decreasing | 63472091.39 | 4.252266 | 117 | 859 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Carettochelys insculpta | Cryptodira | Carettochelyidae | Endangered | Decreasing | 435328.8801 | -11.253778 | 15 | 457 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Pough et al. 2018; Google 2019; IUCN 2019f) |
| Centrochelys sulcata | Cryptodira | Testudinidae | Vulnerable | Unspecified | 3150200.929 | 15.631379 | 17 | 510 | (Iverson et al. 1993; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Chelodina burrungandjii | Pleurodira | Chelidae | Not Listed | N/A | 39296.25574 | -13.358267 | | | (Rhodin et al. 2017; Google 2019) |
| Chelodina canni | Pleurodira | Chelidae | Not Listed | N/A | 402361.9655 | -17.609859 | 10 | 193 | (Kennett et al. 1992; Iverson et al. 1993; McCord and Thomson 2002; Rhodin et al. 2017; Google 2019) |
| Chelodina colliei | Pleurodira | Chelidae | Near Threatened | Unspecified | 96890.30002 | -32.695482 | 6 | 239.3 | (Hosgson and Bencini 2016; Rhodin et al. 2017; Google 2019; Government of Western Australia 2020; IUCN 2020a) |
| Chelodina expansa | Pleurodira | Chelidae | Not Listed | N/A | 365025.6856 | -29.812686 | 20.1667 | 362 | (Iverson 1992b; Iverson et al. 1993; Booth 1998; Hecnar 1999b; Bower and Hodges 2014; Rhodin et al. 2017; Google 2019) |
| Chelodina gunaleni | Pleurodira | Chelidae | Not Listed | N/A | 22743.99972 | -5.047031 | | | (Rhodin et al. 2017; Google 2019) |
| Chelodina kuchlingi | Pleurodira | Chelidae | Not Listed | N/A | 13938.1636 | -14.828699 | | | (Rhodin et al. 2017; Google 2019) |
| Chelodina longicollis | Pleurodira | Chelidae | Not Listed | N/A | 1547543.85 | -29.488503 | 13.9 | 233 | (Iverson et al. 1993; Rhodin et al. 2017; Google 2019) |
| Chelodina mccordi | Pleurodira | Chelidae | Critically Endangered | Decreasing | 1863.042104 | -9.664251 | 9.5333 | | (Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f; IUCN 2020e) |
| Chelodina novaeguineae | Pleurodira | Chelidae | Least Concern | Unspecified | 82879.07209 | -7.469683 | 14.75 | | (Ernst and Barbour 1989; Rhodin 1994; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Chelodina oblonga | Pleurodira | Chelidae | Near Threatened | Unspecified | 85025.06039 | -13.229964 | 10.5 | 230 | (Iverson 1992b; Iverson et al. 1993; Hecnar 1999b; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Chelodina parkeri | Pleurodira | Chelidae | Vulnerable | Unspecified | 41245.12994 | -7.379536 | 10.5 | 220 | (Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Chelodina pritchardi | Pleurodira | Chelidae | Endangered | Unspecified | 7105.455106 | -9.509652 | | 220 | (Rhodin <i>et al.</i> 2017; Google 2019; IUCN 2019f) |
| Chelodina reimanni | Pleurodira | Chelidae | Near Threatened | | 50402.42747 | -7.722873 | 12.99 | | (Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Chelodina steindachneri | Pleurodira | Chelidae | Not Listed | Unspecified N/A | 561642.2713 | -24.777398 | 12.99 | | (Rhodin et al. 2017; Google 2019) |
| | | | | | | | | | (Rhotin et al. 2017; Google 2019) |
| Chelodina walloyarrina | Pleurodira | Chelidae | Not Listed | N/A | 99431.74758 | -16.141519 | 101.78 | 0.88.48 | (Iverson et al. 1993; Chen and Chang 1995; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Chelonia mydas | Cryptodira | Cheloniidae | Endangered | Decreasing | 190288146.5 | 4.241584 | 101.65 | 872.67 | |
| Chelonoidis abingdonii | Cryptodira | Testudinidae | Extinct | Unspecified | 31.57 | 0.56874 | | 680 | (Zug 2013; Rhodin <i>et al.</i> 2017; Google 2019; IUCN 2019f) |
| Chelonoidis becki | Cryptodira | Testudinidae | Vulnerable | Unknown | 225.49 | 0.028956 | | | (Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Chelonoidis carbonarius | Cryptodira | Testudinidae | Not Listed | N/A | 4410390.715 | -6.967653 | 4 | 289 | (Iverson <i>et al.</i> 1993; Rhodin <i>et al.</i> 2017; Google 2019) |
| Chelonoidis chathamensis | Cryptodira | Testudinidae | Endangered | Increasing | 216.3 | -0.804992 | 5 | | (Zug 2013; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Chelonoidis chilensis | Cryptodira | Testudinidae | Vulnerable | Unspecified | 930928.0412 | -31.460861 | 4.07 | 283 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Chelonoidis darwini | Cryptodira | Testudinidae | Critically Endangered | Increasing | 174.07 | -0.272162 | 7 | | (Zug 2013; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Chelonoidis denticulata | Cryptodira | Testudinidae | Vulnerable | Unspecified | 6072026.292 | -5.477871 | 6 | 317.3 | (Ernst and Barbour 1989; Iverson 1992; Hecnar 1999b; Bohm 2011; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Chelonoidis donfaustoi | Cryptodira | Testudinidae | Critically Endangered | Increasing | 115.92 | -0.653057 | | | (Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Chelonoidis duncanensis | Cryptodira | Testudinidae | Vulnerable | Increasing | 13.09 | -0.624612 | 5 | 570 | (Zug 2013; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Chelonoidis guntheri | Cryptodira | Testudinidae | Critically Endangered | Unknown | 465.1918416 | -0.876101 | 9 | 670 | (Zug 2013; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Chelonoidis hoodensis | Cryptodira | Testudinidae | Critically Endangered | Increasing | 38.36 | -1.384919 | 5 | | (Zug 2013; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Chelonoidis microphyes | Cryptodira | Testudinidae | Endangered | Unknown | 115.69 | -0.229839 | 9 | 640 | (Zug 2013; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Chelonoidis niger | Cryptodira | Testudinidae | Extinct | Unspecified | 48.44 | -1.253821 | | | (Rhodin et al. 2017; Google 2019; IUCN 2019f) |

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|--------------------------|----------------|-------------------|-----------------------|-------------|-------------|------------|---------|------------------|--|
| Scientific Name | Suborder | Family | Global IUCN Status | PT | Range | Midpoint | CS | CL | Sources |
| Chelonoidis phantasticus | Cryptodira | Testudinidae | Critically Endangered | Unknown | 76.65471074 | -0.417862 | | | (Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Chelonoidis porteri | Cryptodira | Testudinidae | Critically Endangered | Increasing | 250.4739606 | -0.698153 | 9.5 | | (Ernst and Barbour 1989; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Chelonoidis vandenburghi | Cryptodira | Testudinidae | Vulnerable | Increasing | 483.8236072 | -0.438676 | | | (Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Chelonoidis vicina | Cryptodira | Testudinidae | Endangered | Unknown | 341.7947669 | -0.947902 | 9 | | (Ernst and Barbour 1989; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Chelus fimbriata | Pleurodira | Chelidae | Not Listed | N/A | 2904438.402 | -3.158025 | 12 | 275 | (Iverson et al. 1993; Rhodin et al. 2017; Google 2019) |
| Chelydra acutirostris | Cryptodira | Chelydridae | Not Listed | N/A | 464115.2138 | 6.94193 | 30 | | (Rhodin et al. 2017; Google 2019; Leenders 2019) |
| Chelydra rossignonii | Cryptodira | Chelydridae | Vulnerable | Unknown | 167173.5135 | 16.945526 | 25 | | (Rhodin et al. 2017; Google 2019; IUCN 2019f; IUCN 2020f) |
| Chelydra serpentina | Cryptodira | Chelydridae | Least Concern | Stable | 4753001.029 | 39.996143 | 29.4225 | 255.37 | (Iverson et al. 1993; Rhodin 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Chersina angulata | Cryptodira | Testudinidae | Least Concern | Stable | 382999.0562 | -30.715079 | 1.345 | 158 | (Iverson 1992b; Iverson et al. 1993; Hecnar 1999b; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Chersobius boulengeri | Cryptodira | Testudinidae | Endangered | Decreasing | 144403.8143 | -32.274488 | 1 | 100 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Chersobius signatus | Cryptodira | Testudinidae | Endangered | Decreasing | 70640.83096 | -30.617205 | 1 | 86.8 | (Loehr et al. 2004; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f; IUCN 2020g) |
| Chersobius solus | Cryptodira | Testudinidae | Vulnerable | Unspecified | 26359.88129 | -27.123214 | | | (Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Chitra chitra | Cryptodira | Trionychidae | Critically Endangered | Decreasing | 308025.1309 | 3.892459 | | | (Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Chitra indica | Cryptodira | Trionychidae | Endangered | Unspecified | 1185555.306 | 21.939712 | 101.715 | 1110 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Chitra vandijki | Cryptodira | Trionychidae | Not Listed | N/A | 125547.0826 | 20.973012 | | | (Rhodin et al. 2017; Google 2019) |
| Chrysemys dorsalis | Cryptodira | Emydidae | Least Concern | Stable | 332469.3756 | 33.27097 | 4.5 | 112.5 | (Ernst and Barbour 1989; Powell et al. 2016; Rhodin et al. 2017; Google 2019) |
| Chrysemys picta | Cryptodira | Emydidae | Least Concern | Stable | 4672592.596 | 43.451549 | 7.6625 | 149.125 | (Iverson et al. 1993; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Claudius angustatus | Cryptodira | Staurotypidae | Near Threatened | Unspecified | 150235.0628 | 17.820183 | 4 | 149.125 | (Iverson et al. 1993; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Clemmys guttata | Cryptodira | Emydidae | Endangered | Decreasing | 1149992.932 | 38.041667 | 4.19 | 104 | (Iverson et al. 1993; Powell et al. 2016; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Cuora amboinensis | Cryptodira | Geoemydidae | Vulnerable | Unspecified | 3563737.757 | 9.338949 | 1.5 | 175 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| | Cryptodira | | Critically Endangered | Unspecified | 34223.76687 | 30,929003 | 3 | 1/5 | (Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019) (Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Cuora aurocapitata | | Geoemydidae | | | 41377.52074 | 15.237455 | 1.67 | | (Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 20191) (Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 20191) |
| Cuora bourreti | Cryptodira | Geoemydidae | Critically Endangered | Decreasing | | | 1.07 | | |
| Cuora cyclornata | Cryptodira | Geoemydidae | Critically Endangered | Unspecified | 162254.3951 | 18.816265 | | | (Rhodin <i>et al.</i> 2017; Google 2019) |
| Cuora flavomarginata | Cryptodira | Geoemydidae | Endangered | Unspecified | 498587.1674 | 27.652792 | 1.8 | 152 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Cuora galbinifrons | Cryptodira | Geoemydidae | Critically Endangered | Decreasing | 130555.8832 | 20.105884 | 2 | | (Rhodin <i>et al.</i> 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Cuora mccordi | Cryptodira | Geoemydidae | Critically Endangered | Unspecified | 25500.86549 | 23.330565 | 1.5 | 137.1 | (Iverson <i>et al.</i> 1993; Rhodin <i>et al.</i> 2017; Google 2019; IUCN 2019f) |
| Cuora mouhotii | Cryptodira | Geoemydidae | Endangered | Unspecified | 720387.7556 | 20.229636 | 2.5 | | (Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Cuora pani | Cryptodira | Geoemydidae | Critically Endangered | Unspecified | 93868.93795 | 32.262434 | | | (Rhodin <i>et al.</i> 2017; Google 2019; IUCN 2019f) |
| Cuora picturata | Cryptodira | Geoemydidae | Critically Endangered | Decreasing | 7610.906759 | 12.59039 | 2 | | (Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Cuora trifasciata | Cryptodira | Geoemydidae | Critically Endangered | Unspecified | 255004.6304 | 21.946521 | 2 | | (Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Cuora yunnanensis | Cryptodira | Geoemydidae | Critically Endangered | Decreasing | 35254.26867 | 26.360778 | 8 | | (Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Cuora zhoui | Cryptodira | Geoemydidae | Critically Endangered | Unspecified | 16087.16225 | 22.926434 | | | (Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Cyclanorbis elegans | Cryptodira | Trionychidae | Critically Endangered | Decreasing | 621347.1744 | 9.960505 | 27 | | (Rhodin et al. 2017; Demaya et al. 2019; Google 2019; IUCN 2019f) |
| Cyclanorbis senegalensis | Cryptodira | Trionychidae | Vulnerable | Decreasing | 3055567.965 | 10.760232 | 15.5 | | (Rhodin et al. 2017; Google 2019; IUCN 2019f; IUCN 2020h) |
| Cyclemys atripons | Cryptodira | Geoemydidae | Not Listed | N/A | 28315.5264 | 11.833386 | | | (Rhodin et al. 2017; Google 2019) |
| Cyclemys dentata | Cryptodira | Geoemydidae | Near Threatened | Unspecified | 1452889.416 | 1.810697 | 2.89 | 200 | (Furtado 1988; Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Cyclemys enigmatica | Cryptodira | Geoemydidae | Not Listed | N/A | 895497.2068 | 0.206994 | | | (Rhodin et al. 2017; Google 2019) |
| Cyclemys fusca | Cryptodira | Geoemydidae | Not Listed | N/A | 332930.1705 | 22.784701 | | | (Rhodin et al. 2017; Google 2019) |
| Cyclemys gemeli | Cryptodira | Geoemydidae | Not Listed | N/A | 280127.438 | 24.430557 | | 223.5 | (Praschag et al. 2009; Rhodin et al. 2017; Google 2019) |
| Cyclemys oldhamii | Cryptodira | Geoemydidae | Not Listed | N/A | 877102.7048 | 15.234805 | 3 | 182 | (Aryal et al. 2010; Rhodin et al. 2017; Google 2019; Seateun et al. 2019) |
| Cyclemys pulchristriata | Cryptodira | Geoemydidae | Not Listed | N/A | 79960.17286 | 13.759032 | | | (Rhodin et al. 2017; Google 2019) |
| Cycloderma aubryi | Cryptodira | Trionychidae | Vulnerable | Decreasing | 757029.6439 | -0.811129 | 24.5 | | (Maran 2002; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Cycloderma frenatum | Cryptodira | Trionychidae | Endangered | Decreasing | 463969.5487 | -14.445755 | 18.5 | 560 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Cylindraspis indica | Cryptodira | Testudinidae | Extinct | Unspecified | 2448.38 | -21.135312 | | | (Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Cylindraspis inepta | Cryptodira | Testudinidae | Extinct | Unspecified | 1825.03 | -20.289053 | | | (Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Cylindraspis peltastes | Cryptodira | Testudinidae | Extinct | Unspecified | 108.09 | -19.713362 | | | (Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Cylindraspis triserrata | Cryptodira | Testudinidae | Extinct | Unspecified | 1825.03 | -20.289053 | | | (Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Cylindraspis vosmaeri | Cryptodira | Testudinidae | Extinct | Unspecified | 108.09 | -19.713362 | | | (Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Deirochelys reticularia | Cryptodira | Emydidae | Not Listed | N/A | 892781.4137 | 31.266511 | 8.75 | 185.8 | (Iverson 1992b; Iverson et al. 1993; Hecnar 1999b; Rhodin et al. 2017; Google 2019) |
| Dermatemys mawii | Cryptodira | Dermatemydidae | Critically Endangered | Decreasing | 137220.8014 | 17.319978 | 17.855 | 470 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Dermochelys coriacea | Cryptodira | Dermochelyidae | Vulnerable | Decreasing | 86080918.59 | 7.875652 | 79.8 | 1470 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |

Table 6 (continued). Attribute data for 357 species of turtles (Note: PT = Population Trend, Range = Geographical Range Size (km^2), Midpoint = Midpoint Latitude (•), CS = Mean Clutch Size, CL = Mean Female Carapace Length (mm)).

| Scientific Name | Suborder | Family | Global IUCN Status | PT | Range | Midpoint | CS | CL | Sources |
|-------------------------------------|------------|----------------|-----------------------|-------------|-------------|------------|--------|---------|---|
| Dogania subplana | Cryptodira | Trionychidae | Least Concern | Unspecified | 1332399.318 | 2.971593 | 5 | | (CITES 2013; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Elseya albagula | Pleurodira | Chelidae | Not Listed | N/A | 141438.4254 | -24.697889 | 14 | 305.152 | (Thomson et al. 2006; Thomson and Georges 2016; Rhodin et al. 2017; Google 2019) |
| Elseya bellii | Pleurodira | Chelidae | Endangered | Unspecified | 20421.8008 | -29.921514 | 18.3 | | (Fielder et al. 2015; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Elseya branderhorsti | Pleurodira | Chelidae | Vulnerable | Unspecified | 169667.5272 | -7.488476 | 23.5 | | (Georges et al. 2006; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Elseya dentata | Pleurodira | Chelidae | Not Listed | N/A | 175553.0351 | -15.866946 | 5 | 343 | (Legler and Cann 1980; Iverson et al. 1993; Rhodin et al. 2017; Google 2019) |
| Elseya flaviventralis | Pleurodira | Chelidae | Not Listed | N/A | 160321.8328 | -13.945441 | | | (Rhodin et al. 2017; Google 2019) |
| Elseya georgesi | Pleurodira | Chelidae | Data Deficient | Unspecified | 1143.594997 | -30.510118 | 12.5 | | (Cann 1997a; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Elseya irwini | Pleurodira | Chelidae | Not Listed | N/A | 11507.54811 | -19.108534 | 12 | | (Cann 1997b; Rhodin et al. 2017; Google 2019) |
| Elseya lavarackorum | Pleurodira | Chelidae | Not Listed | N/A | 40759.788 | -17.96667 | 7.5 | 302.1 | (Freeman et al. 2014; Rhodin et al. 2017; Google 2019) |
| Elseya novaeguineae | Pleurodira | Chelidae | Least Concern | Unspecified | 780905.913 | -2.084928 | 12 | | (Iverson 1992b; Hecnar 1999b; Georges et al. 2006; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Elseya purvisi | Pleurodira | Chelidae | Data Deficient | Unspecified | 2961.389178 | -31.583117 | 12 | | (Rhodin <i>et al.</i> 2017; Google 2019; IUCN 2019f) |
| Elseya rhodini | Pleurodira | Chelidae | Not Listed | N/A | 288244.4047 | -6.765692 | | | (Rhodin <i>et al.</i> 2017; Google 2019) |
| Elseya schultzei | Pleurodira | Chelidae | Not Listed | N/A | 191554.8259 | -5.170434 | | | (Rhodin et al. 2017; Google 2019) |
| Eliseya schuizei Elusor macrurus | Pleurodira | Chelidae | | | 11066.02155 | -26.011024 | 14.5 | | (Cann and Legler 1994; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| | | | Endangered | Unspecified | | | 14.5 | 201.0 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Emydoidea blandingii | Cryptodira | Emydidae | Endangered | Decreasing | 822599.3505 | 43.119686 | | 201.9 | (Iverson et al. 1995; Rhodin et al. 2017; C11ES 2018; Google 2019; IUCN 20191) (Iverson et al. 1993; Rhodin et al. 2017; Google 2019) |
| Emydura macquarii | Pleurodira | Chelidae | Not Listed | N/A | 919313.0282 | -26.259173 | 13.59 | 237 | |
| Emydura subglobosa | Pleurodira | Chelidae | Least Concern | Unspecified | 443808.4148 | -12.665685 | - | 210 | (Iverson et al. 1993; Georges et al. 2006; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Emydura tanybaraga | Pleurodira | Chelidae | Not Listed | N/A | 177068.5376 | -14.282904 | 16 | | (Green 2014; Rhodin <i>et al.</i> 2017; Google 2019) |
| Emydura victoriae | Pleurodira | Chelidae | Not Listed | N/A | 192050.609 | -15.732669 | 10 | | (Gaikhorst et al. 2011; Rhodin et al. 2017; Google 2019) |
| Emys orbicularis | Cryptodira | Emydidae | Near Threatened | Unspecified | 4159281.65 | 45.567738 | 9 | | (Zuffi et al. 1999; Speybroeck et al. 2016; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Emys trinacris | Cryptodira | Emydidae | Data Deficient | Unknown | 17170.38613 | 37.455303 | | | (Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Eretmochelys imbricata | Cryptodira | Cheloniidae | Critically Endangered | Decreasing | 43884542.71 | -4.535115 | 142.93 | 820 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| rymnochelys madagascariensis | Pleurodira | Podocnemididae | Critically Endangered | Decreasing | 154212.9886 | -18.768993 | 19.8 | | (Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Geochelone elegans | Cryptodira | Testudinidae | Vulnerable | Decreasing | 915868.1916 | 16.750981 | 5.39 | | (Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Geochelone gigantea | Cryptodira | Testudinidae | Vulnerable | Unspecified | 350.24 | -9.408341 | 9.65 | | (Iverson et al. 1993; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Geochelone platynota | Cryptodira | Testudinidae | Critically Endangered | Unspecified | 88243.25193 | 21.779138 | 5.39 | | (Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Geoclemys hamiltonii | Cryptodira | Geoemydidae | Endangered | Decreasing | 1079456.342 | 27.155367 | 15.5 | | (Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Geoemyda japonica | Cryptodira | Geoemydidae | Endangered | Unspecified | 2880.270126 | 26.496052 | 1.67 | | (Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Geoemyda spengleri | Cryptodira | Geoemydidae | Endangered | Unspecified | 345333.2805 | 20.924741 | 1.335 | 101 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Glyptemys insculpta | Cryptodira | Emydidae | Endangered | Decreasing | 1046695.49 | 43.143739 | 10.265 | 198.3 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Glyptemys muhlenbergii | Cryptodira | Emydidae | Critically Endangered | Unknown | 86382.79076 | 39.226568 | 3.65 | 91.82 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Gopherus agassizii | Cryptodira | Testudinidae | Vulnerable | Unspecified | 129429.2795 | 36.977928 | 5.665 | 220 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Gopherus berlandieri | Cryptodira | Testudinidae | Least Concern | Unspecified | 213898.2507 | 25.845486 | 3.16 | 169 | (Iverson 1992b; Iverson et al. 1993; Hecnar 1999b; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019 |
| Gopherus evgoodei | Cryptodira | Testudinidae | Vulnerable | Decreasing | 42855.89729 | 27.280223 | | | (Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Gopherus flavomarginatus | Cryptodira | Testudinidae | Critically Endangered | Decreasing | 13131.22157 | 27.117308 | 5.57 | | (Iverson 1992b; Hecnar 1999; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Gopherus morafkai | Cryptodira | Testudinidae | Not Listed | N/A | 209975.5548 | 31.46279 | | | (Rhodin et al. 2017; Google 2019) |
| Gopherus polyphemus | Cryptodira | Testudinidae | Vulnerable | Unspecified | 190834.9299 | 29.216133 | 6 | 268.2 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Graptemys barbouri | Cryptodira | Emydidae | Vulnerable | Decreasing | 37823.63592 | 31.647536 | 8.555 | 220 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Graptemys caglei | Cryptodira | Emydidae | Endangered | Decreasing | 10159.18853 | 29.465651 | 3.5 | | (Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Graptemys ernsti | Cryptodira | Emydidae | Near Threatened | Decreasing | 12978.6986 | 31.207754 | 8.67 | | (Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Graptemys flavimaculata | Cryptodira | Emydidae | Vulnerable | Decreasing | 9250.577022 | 31.310554 | 7.81 | 140 | (Powell <i>et al.</i> 2016; Rhodin <i>et al.</i> 2017; CITES 2018; Google 2019; IUCN 2019f) |
| | | | | Stable | 1214525.73 | 40.38447 | 11.42 | 226 | (Iverson <i>et al.</i> 1993; Rhodin <i>et al.</i> 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Graptemys geographica | Cryptodira | Emydidae | Least Concern | | | 40.38447 | 5.66 | 220 | (Iverson et al. 2017; CITES 2018; Google 2019; IUCN 20191) (Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 20191) |
| Graptemys gibbonsi | Cryptodira | Emydidae | Endangered | Decreasing | 12278.27517 | | | 166 | (Riodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Graptemys nigrinoda | Cryptodira | Emydidae | Least Concern | Stable | 60072.83353 | 32.729285 | 6.275 | 155 | |
| Graptemys oculifera | Cryptodira | Emydidae | Vulnerable | Unknown | 18835.02294 | 31.697544 | 5.885 | 148 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Graptemys ouachitensis | Cryptodira | Emydidae | Least Concern | Stable | 686495.653 | 37.759628 | 10.52 | 205 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Graptemys pearlensis | Cryptodira | Emydidae | Endangered | Decreasing | 12127.30608 | 31.547863 | 6.4 | | (Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Graptemys pseudogeographica | Cryptodira | Emydidae | Least Concern | Unknown | 1036721.25 | 38.948502 | 12.305 | 225 | (Iverson 1992b; Iverson et al. 1993; Hecnar 1999b; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019 |
| Graptemys pulchra | Cryptodira | Emydidae | Near Threatened | Unknown | 55079.43756 | 32.844736 | 6.295 | 247 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Graptemys sabinensis | Cryptodira | Emydidae | Least Concern | Stable | 36238.06947 | 31.428414 | | | (Rhodin et al. 2017; Google 2019) |
| Graptemys versa | Cryptodira | Emydidae | Least Concern | Stable | 16815.26751 | 30.498133 | 9.52 | 157 | (Iverson 1992b; Hecnar 1999b; Powell et al. 2016; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f |

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Table 6 (continued). Attribute data for 357 species of turtles (Note: PT = Population Trend, Range = Geographical Range Size (km^2), Midpoint = Midpoint Latitude (°), CS = Mean Clutch Size, CL = Mean Female Carapace Length (mm)).

| Scientific Name | Suborder | Family | Global IUCN Status | PT | Range | Midpoint | CS | CL | Sources |
|---------------------------|------------|---------------|-----------------------|-------------|-------------|------------|--------|--------|--|
| Hardella thurjii | Cryptodira | Geoemydidae | Vulnerable | Unspecified | 531644.2694 | 26.597598 | 14 | | (Aryal et al. 2010; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Heosemys annandalii | Cryptodira | Geoemydidae | Endangered | Unspecified | 353915.4404 | 11.61644 | | | (Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Heosemys depressa | Cryptodira | Geoemydidae | Critically Endangered | Unspecified | 52011.91138 | 19.786588 | | | (Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Heosemys grandis | Cryptodira | Geoemydidae | Vulnerable | Unspecified | 607561.3016 | 10.341112 | 6 | | (Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Heosemys spinosa | Cryptodira | Geoemydidae | Endangered | Unspecified | 999063.276 | 4.247088 | 1.165 | 186 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Homopus areolatus | Cryptodira | Testudinidae | Least Concern | Decreasing | 137044.7414 | -32.865133 | 2.815 | 110.5 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Homopus femoralis | Cryptodira | Testudinidae | Least Concern | Unknown | 226354.7975 | -30.086834 | 1.725 | 140 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Hydromedusa maximiliani | Pleurodira | Chelidae | Vulnerable | Unspecified | 118411.5023 | -22.134069 | | | (Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Hydromedusa tectifera | Pleurodira | Chelidae | Not Listed | N/A | 993673.1138 | -27.788139 | | | (Rhodin et al. 2017; Google 2019) |
| Indotestudo elongata | Cryptodira | Testudinidae | Critically Endangered | Decreasing | 2034295.048 | 18.216561 | 3.905 | 260 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Indotestudo forstenii | Cryptodira | Testudinidae | Endangered | Unspecified | 46107.7129 | -0.190635 | 3.16 | 138.9 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Indotestudo travancorica | Cryptodira | Testudinidae | Vulnerable | Unspecified | 46637.74953 | 11.960558 | 3.54 | 156.9 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Kinixys belliana | Cryptodira | Testudinidae | Not Listed | N/A | 3511150.08 | 0.203455 | 2.535 | 180 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019) |
| Kinixys erosa | Cryptodira | Testudinidae | Data Deficient | Unspecified | 3235985.526 | 0.336668 | 2.5 | 180 | (Rhodin <i>et al.</i> 2017; CITES 2018; Google 2019; IUCN 2019f) |
| | | | | | | | 3 | | (Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019) (Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Kinixys homeana | Cryptodira | Testudinidae | Vulnerable | Decreasing | 810990.1839 | 3.188019 | 2.77 | | |
| Kinixys lobatsiana | Cryptodira | Testudinidae | Vulnerable | Decreasing | 78575.62722 | -25.019486 | | | (Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Kinixys natalensis | Cryptodira | Testudinidae | Vulnerable | Decreasing | 74817.11858 | -26.814129 | 2.77 | | (Rhodin <i>et al.</i> 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Kinixys nogueyi | Cryptodira | Testudinidae | Not Listed | N/A | 2034681.068 | 9.445458 | 2.77 | | (Rhodin <i>et al.</i> 2017; CITES 2018; Google 2019) |
| Kinixys spekii | Cryptodira | Testudinidae | Not Listed | N/A | 2067674.296 | -13.963863 | 3.5 | | (Rhodin et al. 2017; CITES 2018; Google 2019) |
| Kinixys zombensis | Cryptodira | Testudinidae | Not Listed | N/A | 608327.6559 | -16.233215 | 2.77 | | (Rhodin et al. 2017; CITES 2018; Google 2019) |
| Kinosternon abaxillare | Cryptodira | Kinosternidae | Not Listed | N/A | 20850.02137 | 16.102395 | | | (Rhodin et al. 2017; Google 2019) |
| Kinosternon acutum | Cryptodira | Kinosternidae | Near Threatened | Unspecified | 124678.1816 | 17.690669 | | | (Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Kinosternon alamosae | Cryptodira | Kinosternidae | Data Deficient | Unknown | 31731.68074 | 27.307838 | 4 | 105 | (Iverson 1992b; Iverson et al. 1993; Hecnar 1999b; Rhodin et al. 2017; Google 2019; IUCN 2019 |
| Kinosternon angustipons | Cryptodira | Kinosternidae | Vulnerable | Unspecified | 12237.61851 | 10.073042 | 1.5 | 112 | (Iverson et al. 1993; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Kinosternon arizonense | Cryptodira | Kinosternidae | Least Concern | Stable | 47713.99789 | 30.503674 | 4.7 | | (Iverson 1989; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Kinosternon baurii | Cryptodira | Kinosternidae | Least Concern | Unknown | 341350.5769 | 31.002572 | 2.55 | 91.6 | (Iverson et al. 1993; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Kinosternon chimalhuaca | Cryptodira | Kinosternidae | Least Concern | Unknown | 16292.82456 | 19.46086 | | | (Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Kinosternon creaseri | Cryptodira | Kinosternidae | Least Concern | Stable | 88334.88209 | 19.938 | 1 | 116 | (Iverson et al. 1993; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Kinosternon dunni | Cryptodira | Kinosternidae | Vulnerable | Unspecified | 9047.765427 | 6.278786 | 2 | 150 | (Iverson 1992b; Iverson et al. 1993; Hecnar 1999b; Rhodin et al. 2017; Google 2019; IUCN 2019 |
| Kinosternon durangoense | Cryptodira | Kinosternidae | Data Deficient | Unknown | 27278.18981 | 26.097544 | | | (Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Kinosternon flavescens | Cryptodira | Kinosternidae | Least Concern | Unknown | 1382481.008 | 33.512755 | 5.3533 | 114.43 | (Iverson et al. 1993; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Kinosternon herrerai | Cryptodira | Kinosternidae | Near Threatened | Decreasing | 86478.19358 | 21.181463 | 2 | 143.1 | (Iverson et al. 1993; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Kinosternon hirtipes | Cryptodira | Kinosternidae | Least Concern | Decreasing | 207499.5186 | 24.94192 | 3 | 107.4 | (Iverson et al. 1993; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Kinosternon integrum | Cryptodira | Kinosternidae | Least Concern | Stable | 458802.8519 | 22.679446 | 5 | 149.3 | (Iverson et al. 1993; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Kinosternon leucostomum | Cryptodira | Kinosternidae | Not Listed | N/A | 791169.3987 | 9.069126 | 1 | 137 | (Iverson et al. 1993; Rhodin et al. 2017; Google 2019) |
| Kinosternon oaxacae | Cryptodira | Kinosternidae | Data Deficient | Unknown | 22672.62785 | 16.426269 | | | (Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Kinosternon scorpioides | Cryptodira | Kinosternidae | Not Listed | N/A | 6781733.539 | -1.627003 | 3 | 115 | (Iverson et al. 1993; Rhodin et al. 2017; Google 2019) |
| Kinosternon sonoriense | Cryptodira | Kinosternidae | Near Threatened | Unknown | 184756.5976 | 31.308447 | 4.3133 | 124.67 | (Iverson et al. 1993; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Kinosternon steindachneri | Cryptodira | Kinosternidae | Least Concern | Unknown | 89807.76192 | 27.905451 | 10100 | 87.5 | (Powell et al. 2016; Rhodin et al. 2017; Google 2019) |
| Kinosternon subrubrum | Cryptodira | Kinosternidae | Least Concern | Unknown | 1543698.384 | 34.981114 | 3.25 | 93.85 | (Iverson 1992b; Iverson et al. 1993; Hecnar 1999b; Rhodin et al. 2017; Google 2019; IUCN 2019 |
| Lepidochelys kempii | Cryptodira | Cheloniidae | Critically Endangered | Unknown | 1596717.949 | 30.669352 | 106.09 | 657 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019) |
| | | | Vulnerable | | 97296825.04 | 2.732609 | 108.93 | 633 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 20191) (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 20191) |
| Lepidochelys olivacea | Cryptodira | Cheloniidae | | Decreasing | | | 108.93 | 000 | (Iverson et al. 2017; CITES 2018; Google 2019; IUCN 20191) (Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 20191) |
| Leucocephalon yuwonoi | Cryptodira | Geoemydidae | Critically Endangered | Unspecified | 31386.34656 | 0.219 | 1 | | |
| Lissemys ceylonensis | Cryptodira | Trionychidae | Not Listed | N/A | 60970.68028 | 7.880438 | 0.055 | 205 | (Rhodin <i>et al.</i> 2017; Google 2019) |
| Lissemys punctata | Cryptodira | Trionychidae | Least Concern | Unspecified | 3211517.642 | 21.851272 | 8.855 | 205 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Lissemys scutata | Cryptodira | Trionychidae | Data Deficient | Unspecified | 75742.02174 | 20.503622 | 11.88 | | (Iverson 1992b; Hecnar 1999b; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Macrochelys suwanniensis | Cryptodira | Chelydridae | Not Listed | N/A | 16151.35741 | 30.070884 | 36 | 444 | (Enge et al. 2014; Powell et al. 2016; Rhodin et al. 2017; Google 2019) |
| Macrochelys temminckii | Cryptodira | Chelydridae | Vulnerable | Unspecified | 588338.1921 | 34.964587 | 25.52 | 400 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Malaclemys terrapin | Cryptodira | Emydidae | Vulnerable | Decreasing | 151434.0443 | 33.527398 | 8.52 | 195 | (Iverson et al. 1993; Powell et al. 2016; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019 |
| Malacochersus tornieri | Cryptodira | Testudinidae | Critically Endangered | Decreasing | 142927.3352 | -3.619516 | 1.915 | 145 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Malayemys khoratensis | Cryptodira | Geoemydidae | Not Listed | N/A | 34877.39687 | 16.414313 | | | (Rhodin et al. 2017; Google 2019) |

Table 6 (continued). Attribute data for 357 species of turtles (Note: PT = Population Trend, Range = Geographical Range Size (km^2), Midpoint = Midpoint Latitude (*), CS = Mean Clutch Size, CL = Mean Female Carapace Length (mm)).

| Scientific Name | Suborder | Family | Global IUCN Status | PT | Range | Midpoint | CS | CL | Sources |
|--------------------------------------|------------|---------------|-----------------------|------------------------|-------------|------------|---------|-------|---|
| Malayemys macrocephala | Cryptodira | Geoemydidae | Not Listed | N/A | 170446.2498 | 12.272206 | | | (Rhodin et al. 2017; Google 2019) |
| Malayemys subtrijuga | Cryptodira | Geoemydidae | Vulnerable | Unspecified | 189767.8487 | 12.923553 | | | (Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Manouria emys | Cryptodira | Testudinidae | Critically Endangered | Decreasing | 925871.4824 | 10.480069 | 32.76 | | (Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Manouria impressa | Cryptodira | Testudinidae | Vulnerable | Unspecified | 601187.6141 | 15.693891 | 13 | | (Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Mauremys annamensis | Cryptodira | Geoemydidae | Critically Endangered | Unspecified | 13826.30422 | 14.566932 | | | (Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Mauremys caspica | Cryptodira | Geoemydidae | Not Listed | N/A | 870978.0382 | 35.114822 | 5 | | (Ernst and Barbour 1989; Rhodin et al. 2017; Google 2019) |
| Mauremys japonica | Cryptodira | Geoemydidae | Near Threatened | Unspecified | 162089.717 | 34.424437 | 6.165 | 150 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Mauremys leprosa | Cryptodira | Geoemydidae | Not Listed | N/A | 949636.7553 | 36.025814 | 5.8 | 168.8 | (Iverson et al. 1993; Rhodin et al. 2017; Google 2019) |
| Mauremys mutica | Cryptodira | Geoemydidae | Endangered | Unspecified | 990483.6563 | 24.171152 | 1.5 | 130 | (Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Mauremys nigricans | Cryptodira | Geoemydidae | Endangered | Unspecified | 205738.2995 | 23.870105 | 3.5 | | (Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Mauremys reevesii | Cryptodira | Geoemydidae | Endangered | Unknown | 1119558.423 | 30.405418 | 5.02 | 174.5 | (Iverson et al. 1993; Zug 2013; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Mauremys rivulata | Cryptodira | Geoemydidae | Not Listed | N/A | 388098.6523 | 38.32816 | 8 | | (Speybroeck et al. 2016; Rhodin et al. 2017; Google 2019) |
| Mauremys sinensis | Cryptodira | Geoemydidae | Endangered | Unspecified | 494385.5967 | 23.276006 | 5.25 | 214 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Melanochelys tricarinata | Cryptodira | Geoemydidae | Vulnerable | Unspecified | 315712.9037 | 25.883284 | 2 | 138 | (Kumar et al. 2009; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Melanochelys trijuga | Cryptodira | Geoemydidae | Near Threatened | Unspecified | 1611867.25 | 18.523231 | 3.835 | 175 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Mesoclemmys dahli | Pleurodira | Chelidae | Critically Endangered | Unspecified | 29150.77622 | 9.480531 | 4 | 191 | (Iverson et al. 1993; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Mesoclemmys gibba | Pleurodira | Chelidae | Not Listed | N/A | 2920384.434 | -1.105113 | 3 | 184.5 | (Iverson et al. 1993; Rhodin et al. 2017; Google 2019) |
| Mesoclemmys heliostemma | Pleurodira | Chelidae | Not Listed | N/A | 1240495.102 | -4.352941 | | | (Rhodin et al. 2017; Google 2019) |
| Mesoclemmys hogei | Pleurodira | Chelidae | Critically Endangered | Decreasing | 12825.32925 | -21.709454 | 8 | | (Iverson 1992b; Hecnar 1999b; Rhodin et al. 2017; Google 2019; IUCN 2019f; IUCN 2020i) |
| Mesoclemmys nasuta | Pleurodira | Chelidae | Not Listed | N/A | 184666.5216 | 4.04227 | 7 | | (Vogt 2008; Rhodin et al. 2017; Google 2019) |
| Mesoclemmys perplexa | Pleurodira | Chelidae | Not Listed | N/A | 253274.8397 | -8.915007 | | | (Rhodin et al. 2017; Google 2019) |
| Mesoclemmys raniceps | Pleurodira | Chelidae | Not Listed | N/A | 2271303.401 | -7.566802 | 5.6 | | (Rhodin <i>et al.</i> 2017; Cunha <i>et al.</i> 2019; Google 2019) |
| Mesoclemmys tuberculata | Pleurodira | Chelidae | Not Listed | N/A | 520162.5844 | -10.096053 | 5.0 | | (Santana et al. 2016; Rhodin et al. 2017; Google 2019) |
| Mesoclemmys vanderhaegei | Pleurodira | Chelidae | Near Threatened | Unspecified | 955253.5688 | -18.331146 | 6.4 | | (Iverson 1992b; Hecnar 1999b; Marques et al. 2014; Rhodin et al. 2017; Google 2019; IUCN 2019 |
| Mesoclemmys zuliae | Pleurodira | Chelidae | Vulnerable | Unspecified | 1498.110276 | 8.784567 | 7 | 263 | (Iverson 1992b; Iverson et al. 1993; Hecnar 1999b; Rhodin et al. 2017; Google 2019; IUCN 2019 |
| Morenia ocellata | Cryptodira | Geoemydidae | Vulnerable | Unspecified | 259329.0745 | 18.60912 | , | 205 | (Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Morenia petersi | Cryptodira | Geoemydidae | Vulnerable | Unspecified | 471004.9386 | 25.878341 | 8 | | (Aryal <i>et al.</i> 2010; Rhodin <i>et al.</i> 2017; Google 2019; IUCN 2019f) |
| Myuchelys latisternum | Pleurodira | Chelidae | Not Listed | N/A | 462876.6009 | -20.272396 | 17 | 232 | (Legler and Cann 1980; Rhodin <i>et al.</i> 2017; Google 2019) |
| Natator depressus | Cryptodira | Cheloniidae | Data Deficient | Unspecified | 2354838.211 | -16.502395 | 51.045 | 923 | (Iverson <i>et al.</i> 1993; Rhodin <i>et al.</i> 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Nilssonia formosa | Cryptodira | Trionychidae | Endangered | Unspecified | 125568.478 | 21.043185 | 26.3333 | 923 | (Rodin et al. 2017; Platt et al. 2018; Google 2019; IUCN 20191) (Rhodin et al. 2017; Platt et al. 2018; Google 2019; IUCN 20191) |
| | | Trionychidae | Vulnerable | - | 1350885.502 | 27.182781 | 30.135 | 675 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Nilssonia gangetica | Cryptodira | Trionychidae | Vulnerable | Unspecified | 845388.9172 | 25.991177 | 22.6667 | 675 | (Das et al. 2010; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Nilssonia hurum Nilssonia leithii | Cryptodira | | Vulnerable | Unspecified Unknown | 570791.8731 | 15.938402 | 22.0007 | | (Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| | Cryptodira | Trionychidae | | | 96583.77974 | 25.165232 | 20.2 | 436 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Nilssonia nigricans | Cryptodira | Trionychidae | Extinct in the Wild | Unspecified | 804325.4274 | | 20.2 | 430 | (Rodin et al. 2017; Google 2019; IUCN 2019f) |
| Notochelys platynota | Cryptodira | Geoemydidae | Vulnerable | Unspecified | | 2.272848 | 10 | 175 | |
| Orlitia borneensis | Cryptodira | Geoemydidae | Endangered | Unspecified | 626133.3885 | 0.350834 | 12 | 475 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) (Bhadin et al. 2017; CITES 2018; Coogle 2019; IUCN 20106) |
| Palea steindachneri | Cryptodira | Trionychidae | Endangered | Unspecified | 257916.567 | 20.263077 | 15.5 | 100 | (Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Pangshura smithii | Cryptodira | Geoemydidae | Near Threatened | Unspecified | 708343.6271 | 28.463224 | 6.635 | 190 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Pangshura sylhetensis | Cryptodira | Geoemydidae | Endangered | Unspecified | 154805.7366 | 24.840922 | 9 | | (Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Pangshura tecta | Cryptodira | Geoemydidae | Least Concern | Unspecified | 963573.3823 | 26.602614 | 7.5 | | (Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Pangshura tentoria | Cryptodira | Geoemydidae | Least Concern | Unspecified | 469840.4959 | 23.472658 | 6.36 | 212.5 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Pelochelys bibroni | Cryptodira | Trionychidae | Vulnerable | Unspecified | 268069.3786 | -6.787621 | 26 | | (Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Pelochelys cantorii | Cryptodira | Trionychidae | Endangered | Unspecified | 1653989.372 | 10.649015 | | | (Rhodin <i>et al.</i> 2017; Google 2019; IUCN 2019f) |
| Pelochelys signifera | Cryptodira | Trionychidae | Vulnerable | Decreasing | 109957.5213 | -3.669564 | | | (Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Pelodiscus axenaria | Cryptodira | Trionychidae | Not Listed | N/A | 90290.66909 | 27.574465 | | | (Rhodin <i>et al.</i> 2017; Google 2019) |
| Pelodiscus maackii | Cryptodira | Trionychidae | Not Listed | N/A | 657694.9906 | 43.763193 | | | (Rhodin <i>et al.</i> 2017; Google 2019) |
| Pelodiscus parviformis | Cryptodira | Trionychidae | Not Listed | N/A | 114825.3709 | 19.649973 | | | (Rhodin et al. 2017; Google 2019) |
| Pelodiscus sinensis | Cryptodira | Trionychidae | Vulnerable | Decreasing | 1786617.216 | 30.224365 | 14.5 | 230 | (Iverson et al. 1993; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Pelomedusa barbata | Pleurodira | Pelomedusidae | Not Listed | N/A | 79978.18893 | 15.611715 | | | (Rhodin et al. 2017; Google 2019) |
| Pelomedusa galeata | Pleurodira | Pelomedusidae | Least Concern | Unknown | 722364.7032 | -30.075026 | | | (Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Pelomedusa gehafie | Pleurodira | Pelomedusidae | Not Listed | N/A | 157663.3343 | 16.232838 | | | (Rhodin et al. 2017; Google 2019) |

| Table 6 (continued). Attribute data for 357 species of turtles | (Note: PT = Population Trend, Range | = Geographical Range Size (km ²), Midpoint = Midpoint Latitud | de (°), CS = Mean Clutch Size, CL = Mean Female Carapace Length (mm)). |
|--|-------------------------------------|---|--|
| | | | |

| Scientific Name | Suborder | Family | Global IUCN Status | PT | Range | Midpoint | CS | CL | Sources |
|--|------------|----------------|-----------------------|-------------|-------------|------------|--------|-------|--|
| Pelomedusa kobe | Pleurodira | Pelomedusidae | Not Listed | N/A | 296814.8738 | -5.807681 | | | (Rhodin et al. 2017; Google 2019) |
| Pelomedusa neumanni | Pleurodira | Pelomedusidae | Not Listed | N/A | 460363.7126 | 0.679782 | | | (Rhodin et al. 2017; Google 2019) |
| Pelomedusa olivacea | Pleurodira | Pelomedusidae | Not Listed | N/A | 1683940.853 | 13.025863 | | | (Rhodin et al. 2017; Google 2019) |
| Pelomedusa schweinfurthi | Pleurodira | Pelomedusidae | Not Listed | N/A | 418966.2811 | 6.091247 | | | (Rhodin et al. 2017; Google 2019) |
| Pelomedusa somalica | Pleurodira | Pelomedusidae | Not Listed | N/A | 306884.3672 | 9.429055 | | | (Rhodin et al. 2017; Google 2019) |
| Pelomedusa subrufa | Pleurodira | Pelomedusidae | Not Listed | N/A | 2154348.29 | -14.224446 | 25.5 | 270 | (Ernst and Barbour 1989; Iverson et al. 1993; Rhodin et al. 2017; Google 2019) |
| Pelomedusa variabilis | Pleurodira | Pelomedusidae | Not Listed | N/A | 389636.3736 | 10.614321 | | | (Rhodin et al. 2017; Google 2019) |
| Peltocephalus dumerilianus | Pleurodira | Podocnemididae | Vulnerable | Unspecified | 1845748.45 | 1.808225 | 11.8 | 324 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Pelusios adansonii | Pleurodira | Pelomedusidae | Not Listed | N/A | 2079142.748 | 10.881575 | 7 | | (Ernst and barbour 1989; Rhodin et al. 2017; Google 2019) |
| Pelusios bechuanicus | Pleurodira | Pelomedusidae | Not Listed | N/A | 355037.7173 | -15.997654 | 32.25 | | (Broadley 1981; Iverson 1992b; Hecnar 1999b; Rhodin et al. 2017; Google 2019) |
| Pelusios broadleyi | Pleurodira | Pelomedusidae | Vulnerable | Unspecified | 23679.31035 | 3.464762 | | | (Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Pelusios carinatus | Pleurodira | Pelomedusidae | Not Listed | N/A | 388473.5032 | -1.531417 | | | (Iverson 1992b; Hecnar 1999b; Rhodin et al. 2017; Google 2019) |
| Pelusios castaneus | Pleurodira | Pelomedusidae | Not Listed | N/A | 1748253.312 | 1.762215 | 9 | | (Ernst and Barbour 1989; Bour et al. 2016; Rhodin et al. 2017; Google 2019) |
| Pelusios castanoides | Pleurodira | Pelomedusidae | Least Concern | Unspecified | 1547270.235 | -15.667035 | 25 | 220 | (Ernst and Barbour 1989; Iverson 1992b; Iverson et al. 1993; Hecnar 1999b; Rhodin et al. 2017; Google 2019; IUCN 2019 |
| Pelusios chapini | Pleurodira | Pelomedusidae | Not Listed | N/A | 1028560.788 | 1.124205 | | | (Rhodin et al. 2017; Google 2019) |
| Pelusios cupulatta | Pleurodira | Pelomedusidae | Not Listed | N/A | 344097.8921 | 6.302687 | | | (Rhodin et al. 2017; Google 2019) |
| Pelusios gabonensis | Pleurodira | Pelomedusidae | Not Listed | N/A | 2981682.854 | -0.768546 | 12 | | (Iverson 1992b; Hecnar 1999b; Rhodin <i>et al.</i> 2017; Spawls <i>et al.</i> 2018; Google 2019) |
| Pelusios marani | Pleurodira | Pelomedusidae | Not Listed | N/A N/A | 50325.92485 | -1.901661 | 12 | | (Rhodin et al. 2017; Google 2019) |
| Pelusios nanus | Pleurodira | Pelomedusidae | Not Listed | N/A N/A | 1047338.516 | -10.457611 | | | (Rhodin et al. 2017; Google 2019) |
| | Pleurodira | Pelomedusidae | Near Threatened | | 278178.2951 | 2.177852 | 25 | 257.7 | (Akani et al. 2015; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Pelusios niger | | | | Decreasing | | -14.251753 | 12.625 | 230 | (Broadley 1981; Iverson 1992b; Iverson <i>et al.</i> 1993; Hecnar 1999b; Rhodin <i>et al.</i> 2017; Google 2019; IUCN 2019f) |
| Pelusios rhodesianus | Pleurodira | Pelomedusidae | Least Concern | Unspecified | 3066734.129 | | 12.625 | 230 | (Bloadley 1961, Iverson 1992), Iverson et al. 1993, Flechar 19930, Ribdin et al. 2017, Google 2019, IUCN 20191) (Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Pelusios seychellensis | Pleurodira | Pelomedusidae | Extinct | Unspecified | 155.64 | -4.648031 | | | |
| Pelusios sinuatus | Pleurodira | Pelomedusidae | Not Listed | N/A | 3284755.202 | -12.445972 | 10 | | (Iverson 1992b; Hecnar 1999b; Branch 2016; Rhodin <i>et al.</i> 2017; Google 2019) |
| Pelusios subniger | Pleurodira | Pelomedusidae | Least Concern | Unspecified | 2286246.326 | -15.18629 | 8.5 | | (Broadley 1981; Iverson 1992b; Hecnar 1999b; Rhodin <i>et al.</i> 2017; Gerlach 2018; Google 2019; IUCN 2019f) |
| Pelusios upembae | Pleurodira | Pelomedusidae | Data Deficient | Unspecified | 10709.16556 | -8.702999 | | | (Iverson 1992b; Hecnar 1999b; Rhodin <i>et al.</i> 2017; Google 2019; IUCN 2019f) |
| Pelusios williamsi | Pleurodira | Pelomedusidae | Not Listed | N/A | 179493.9328 | -0.137333 | | | (Iverson 1992b; Hecnar 1999b; Rhodin et al. 2017; Google 2019) |
| Phrynops geoffroanus | Pleurodira | Chelidae | Not Listed | N/A | 3742007.503 | -11.773345 | 15 | | (Ernst and Barbour 1989; Rhodin et al. 2017; Google 2019) |
| Phrynops hilarii | Pleurodira | Chelidae | Not Listed | N/A | 797730.5523 | -28.890451 | 18 | 280 | (Iverson et al. 1993; Rhodin et al. 2017; Google 2019) |
| Phrynops tuberosus | Pleurodira | Chelidae | Not Listed | N/A | 215256.1048 | 5.170188 | | | (Rhodin et al. 2017; Google 2019) |
| Phrynops williamsi | Pleurodira | Chelidae | Vulnerable | Decreasing | 328922.4904 | -28.967375 | 9 | | (Ernst and Barbour 1989; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Platemys platycephala | Pleurodira | Chelidae | Not Listed | N/A | 5074985.226 | -3.126892 | 1 | 150 | (Ernst and Barbour 1989; Iverson et al. 1993; Rhodin et al. 2017; Google 2019) |
| Platysternon megacephalum | Cryptodira | Platysternidae | Endangered | Unspecified | 1120778.994 | 21.501034 | 2.25 | | (Jenkins 1995; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Podocnemis erythrocephala | Pleurodira | Podocnemididae | Vulnerable | Unspecified | 799031.5258 | -0.503688 | 8.43 | | (Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Podocnemis expansa | Pleurodira | Podocnemididae | Not Defined | Unspecified | 2065033.881 | -3.236399 | 97.115 | 660 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Podocnemis lewyana | Pleurodira | Podocnemididae | Critically Endangered | Decreasing | 74213.19968 | 6.182906 | 18.5 | 475 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Podocnemis sextuberculata | Pleurodira | Podocnemididae | Vulnerable | Unspecified | 792136.4667 | -3.109082 | 51.5 | | (Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Podocnemis unifilis | Pleurodira | Podocnemididae | Vulnerable | Unspecified | 2907785.183 | -3.075128 | 26.815 | 360.2 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Podocnemis vogli | Pleurodira | Podocnemididae | Not Listed | N/A | 240468.3595 | 5.455609 | 13.165 | 271 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019) |
| Psammobates geometricus | Cryptodira | Testudinidae | Critically Endangered | Decreasing | 13067.49954 | -33.312041 | 3.19 | 124.1 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Psammobates oculifer | Cryptodira | Testudinidae | Not Listed | N/A | 1077342.903 | -23.900913 | 1.5 | | (Branch 2016; Rhodin et al. 2017; Google 2019) |
| Psammobates tentorius | Cryptodira | Testudinidae | Near Threatened | Decreasing | 694526.5958 | -29.191217 | 2.115 | 125 | (Iverson 1992b; Iverson et al. 1993; Hecnar 1999b; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Pseudemydura umbrina | Pleurodira | Chelidae | Critically Endangered | Unspecified | 95.49337049 | -31.742085 | 4.335 | 125 | (Iverson 1992b; Iverson et al. 1993; Hecnar 1999b; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Pseudemys alabamensis | Cryptodira | Emydidae | Endangered | Unspecified | 6130.122821 | 30.734503 | 13 | 254.5 | (Nelson et al. 2009; Powell et al. 2016; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Pseudemys concinna | Cryptodira | Emydidae | Least Concern | Unknown | 1240884.796 | 33.548019 | 17 | 289 | (Iverson et al. 1993; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Pseudemys floridana | Cryptodira | Emydidae | Least Concern | Unknown | 284711.8497 | 33.0982 | 14.15 | 280 | (Iverson <i>et al.</i> 1993; Powell <i>et al.</i> 2016; Rhodin et al. 2017; Google 2019) |
| Pseudemys gorzugi | Cryptodira | Emydidae | Near Threatened | Unknown | 71115.25499 | 29.656776 | 9 | 254.5 | (Powell <i>et al.</i> 2016; Rhodin <i>et al.</i> 2017; Google 2019; IUCN 2019f; IUCN 2020j) |
| Pseudemys gorzugi Pseudemys nelsoni | Cryptodira | Emydidae | Least Concern | Stable | 113134.7207 | 29.030770 | 14.3 | 298 | (Iverson et al. 1993; Rhodin et al. 2017; Google 2019; IUCN 2019) |
| Pseudemys peninsularis | Cryptodira | Emydidae | Least Concern | Unknown | 85391.23156 | 27.855189 | 14.5 | 298 | (Powell <i>et al.</i> 2016; Rhodin <i>et al.</i> 2017; Google 2019; IUCN 2019f; IUCN 2020k) |
| | | | Near Threatened | | | 37.902241 | 15 | 304 | (Fowen et al. 2010, Kilodin et al. 2017, Google 2019, TOCN 2020K) (Iverson et al. 1993; Rhodin et al. 2017; Google 2019; IUCN 2020K) |
| Pseudemys rubriventris | Cryptodira | Emydidae | | Unknown | 86121.82427 | | 8.4 | | (Iversion et al. 1995; Khodin et al. 2017; Google 2019; IUCN 2019f) (Lindeman 2007; Powell et al. 2016; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Pseudemys texana | Cryptodira | Emydidae | Least Concern | Unknown | 161765.7189 | 31.044119 | 8.4 | 217.5 | (Lindeman 2007; Powell et al. 2010; Knodin et al. 2017; Google 2019; IUCN 2019f) |

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Table 6 (continued). Attribute data for 357 species of turtles (Note: PT = Population Trend, Range = Geographical Range Size (km^2), Midpoint = Midpoint Latitude (+), CS = Mean Clutch Size, CL = Mean Female Carapace Length (mm)).

| Scientific Name | Suborder | Family | Global IUCN Status | PT | Range | Midpoint | CS | CL | Sources |
|------------------------------|--------------------------|----------------------|------------------------------|-------------|-------------|------------|--------|--------|--|
| Pyxis planic auda | Cryptodira | Testudinidae | Critically Endangered | Decreasing | 4685.723564 | -19.86912 | 1.17 | | (Iverson 1992b; Hecnar 1999b; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Rafetus euphraticus | Cryptodira | Trionychidae | Endangered | Decreasing | 351708.871 | 34.200547 | | | (Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Rafetus swinhoei | Cryptodira | Trionychidae | Critically Endangered | Unspecified | 82393.39106 | 26.484672 | | | (Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Rheodytes leukops | Pleurodira | Chelidae | Vulnerable | Unspecified | 8690.103673 | -23.961051 | 18.8 | 249 | (Iverson 1992b; Iverson et al. 1993; Hecnar 1999b; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Rhinemys rufipes | Pleurodira | Chelidae | Near Threatened | Unspecified | 480481.3961 | -0.173123 | 7.5 | | (Ernst and Barbour 1989; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Rhinoclemmys annulata | Cryptodira | Geoemydidae | Near Threatened | Unspecified | 328269.268 | 6.920305 | 1.5 | | (Rhodin et al. 2017; Google 2019; IUCN 2019f; Leenders 2019) |
| Rhinoclemmys areolata | Cryptodira | Geoemydidae | Near Threatened | Decreasing | 214663.5814 | 18.328362 | 1 | 166.6 | (Iverson et al. 1993; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Rhinoclemmys diademata | Cryptodira | Geoemydidae | Not Listed | N/A | 43947.65369 | 9.571282 | 2 | 203 | (Iverson et al. 1993; Rhodin et al. 2017; Google 2019) |
| Rhinoclemmys funerea | Cryptodira | Geoemydidae | Near Threatened | Unspecified | 76431.0208 | 11.897212 | 3.2 | 273 | (Iverson et al. 1993; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Rhinoclemmys melanosterna | Cryptodira | Geoemydidae | Not Listed | N/A | 211474.3223 | 6.049135 | 5 | 243 | (Iverson et al. 1993; Rhodin et al. 2017; Google 2019) |
| Rhinoclemmys nasuta | Cryptodira | Geoemydidae | Near Threatened | Unspecified | 90584.23622 | 3.443446 | 1 | 218 | (Iverson et al. 1993; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Rhinoclemmys pulcherrima | Cryptodira | Geoemydidae | Not Listed | N/A | 181670.4314 | 18.604778 | 1.5 | | (Rhodin et al. 2017; Google 2019; Leenders 2019) |
| Rhinoclemmys punctularia | Cryptodira | Geoemydidae | Not Listed | N/A | 1639313.468 | -1.49323 | 1.5 | | (Ernst and Barbour 1989; Rhodin et al. 2017; Google 2019) |
| Rhinoclemmys rubida | Cryptodira | Geoemydidae | Near Threatened | Decreasing | 54458.90825 | 17.986547 | | | (Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Sacalia bealei | Cryptodira | Geoemydidae | Endangered | Unspecified | 199968.1772 | 24.65135 | 3.5 | | (Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Sacalia quadriocellata | Cryptodira | Geoemydidae | Endangered | Unspecified | 348409.8176 | 19.481445 | 2 | | (Jenkins 1995; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Siebenrockiella crassicollis | Cryptodira | Geoemydidae | Vulnerable | Unspecified | 928349.8018 | 4.07725 | 1.365 | 186 | (Iverson 1992b; Iverson et al. 1993; Hecnar 1999b; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019 |
| Siebenrockiella leytensis | Cryptodira | Geoemydidae | Critically Endangered | Unspecified | 4924.42 | 10.561359 | 1.5 | | (Diesmos et al. 2012; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Staurotypus salvinii | Cryptodira | Staurotypidae | Near Threatened | Unspecified | 40761.9242 | 15.002467 | 5.36 | 180 | (Iverson et al. 1993; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Staurotypus triporcatus | Cryptodira | Staurotypidae | Near Threatened | Unspecified | 182065.7947 | 17.877418 | 10.61 | 285 | (Iverson et al. 1993; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Sternotherus carinatus | Cryptodira | Kinosternidae | Least Concern | Unknown | 352096.744 | 32.741645 | 4.2 | 116.5 | (Iverson et al. 1993; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Sternotherus depressus | Cryptodira | Kinosternidae | Critically Endangered | Decreasing | 10093.10909 | 33.725957 | 2.2 | 95 | (Iverson et al. 1993; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Sternotherus minor | Cryptodira | Kinosternidae | Least Concern | Unknown | 370377.1442 | 32.806647 | 2.73 | 102.55 | (Iverson 1992b; Iverson <i>et al.</i> 1993; Hecnar 1999b; Rhodin <i>et al.</i> 2017; Google 2019; IUCN 2019f) |
| Sternotherus odoratus | Cryptodira | Kinosternidae | Least Concern | Stable | 2135642.3 | 36.106362 | 3.775 | 97.2 | (Iverson et al. 1993; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Stigmochelys pardalis | Cryptodira | Testudinidae | Least Concern | Unknown | 7569425.193 | -11.861399 | 10.9 | 484 | (Iverson 1992b; Iverson et al. 1993; Hecnar 1999b; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019 |
| Terrapene carolina | Cryptodira | Emydidae | Vulnerable | Decreasing | 2663071.095 | 31.92525 | 3.855 | 134 | (Iverson 1992b; Iverson et al. 1993; Heenar 1999b; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019) (Iverson 1992b; Iverson et al. 1993; Heenar 1999b; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019) |
| Terrapene coahuila | | | | Decreasing | 2228.232533 | 26.92915 | 3.855 | 101.6 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019) (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Terrapene nelsoni | Cryptodira Cryptodira | Emydidae Emydidae | Endangered Data Deficient | Unspecified | 54118.62002 | 25.276717 | 2.69 | 134 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| | | | Near Threatened | | 1694862.879 | 35.716161 | 4.39 | 134 | (Iverson et al. 1993; Riodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Terrapene ornata | Cryptodira | Emydidae | | Decreasing | | | | 117 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Testudo graeca | Cryptodira | Testudinidae | Vulnerable | Unspecified | 1833795.542 | 36.745387 | 4.955 | 189 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Testudo hermanni | Cryptodira | Testudinidae | Near Threatened | Decreasing | 352313.6986 | 41.603166 | 4.8075 | 190 | (Redin et al. 2017; CITES 2018; Google 2019; IUCN 20191) (Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Testudo horsfieldii | Cryptodira | Testudinidae | Vulnerable | Unspecified | 1395626.829 | 39.304597 | 4 | | (Rhodin <i>et al.</i> 2017; CITES 2018; Google 2019; IUCN 2019f) (Rhodin <i>et al.</i> 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Testudo kleinmanni | Cryptodira | Testudinidae | Critically Endangered | Decreasing | 90369.62785 | 31.513048 | 2.93 | | |
| Testudo marginata | Cryptodira | Testudinidae | Least Concern | Stable | 65530.06297 | 38.98508 | 6.07 | 253 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) (Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Trachemys adiutrix | Cryptodira | Emydidae | Endangered | Unspecified | 31137.11821 | -2.689964 | | | |
| Trachemys callirostris | Cryptodira | Emydidae | Not Listed | N/A | 135258.8427 | 9.026841 | 11 | 260 | (Ernst 2003a; Bock <i>et al.</i> 2010; Rhodin et al. 2017; Google 2019) |
| Trachemys decorata | Cryptodira | Emydidae | Vulnerable | Unspecified | 11215.99915 | 18.446403 | 12 | 261 | (Iverson et al. 1993; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Trachemys decussata | Cryptodira | Emydidae | Not Listed | N/A | 124117.7561 | 20.731081 | 5 | 171 | (Iverson et al. 1993; Rhodin et al. 2017; Google 2019) |
| Trachemys dorbigni | Cryptodira | Emydidae | Not Listed | N/A | 456245.9743 | -31.526073 | 13.05 | 260 | (Iverson et al. 1993; Bager et al. 2007; Rhodin et al. 2017; Google 2019) |
| Trachemys gaigeae | Cryptodira | Emydidae | Vulnerable | Unknown | 69481.63036 | 29.770134 | 19 | 164 | (Morjan and Stuart 2001; Powell et al. 2016; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Trachemys grayi | Cryptodira | Emydidae | Not Listed | N/A | 228065.2056 | 12.131748 | | | (Rhodin <i>et al.</i> 2017; Google 2019) |
| Trachemys nebulosa | Cryptodira | Emydidae | Not Listed | N/A | 32955.39204 | 25.272894 | | | (Rhodin <i>et al.</i> 2017; Google 2019) |
| Trachemys ornata | Cryptodira | Emydidae | Vulnerable | Decreasing | 36236.74741 | 22.769475 | 20 | | (Rhodin et al. 2017; Google 2019; IUCN 2019f; IUCN 2020l) |
| Trachemys scripta | Cryptodira | Emydidae | Least Concern | Stable | 1808510.103 | 34.717577 | 9.007 | 164 | (Iverson et al. 1993; Powell et al. 2016; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |
| Trachemys stejnegeri | Cryptodira | Emydidae | Near Threatened | Unspecified | 53578.9163 | 19.589167 | 8.5 | | (Rhodin et al. 2017; USGS 2018; Google 2019; IUCN 2019f) |
| Trachemys taylori | Cryptodira | Emydidae | Endangered | Decreasing | 2820.901206 | 27.174885 | | | (Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Trachemys terrapen | Cryptodira | Emydidae | Vulnerable | Unspecified | 10469.16815 | 18.04401 | 5.5 | | (Currie et al. 2019; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Trachemys venusta | Cryptodira | Emydidae | Not Listed | N/A | 604963.6894 | 16.507278 | 26.5 | 300 | (Ernst and Seidel 2006; Rhodin et al. 2017; Google 2019; Leenders 2019) |
| Trachemys yaquia | Cryptodira | Emydidae | Vulnerable | Decreasing | 56996.33319 | 28.902469 | | 275 | (Ernst 2003b; Rhodin et al. 2017; Google 2019; IUCN 2019f) |
| Trionyx triunguis | Cryptodira | Trionychidae | Vulnerable | Decreasing | 7367511.81 | 4.26208 | 47.5 | 506 | (Iverson 1992b; Iverson et al. 1993; Hecnar 1999b; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019 |
| Vijavachelys silvatica | Cryptodira | Geoemydidae | Endangered | Unknown | 26944.72106 | 11.325848 | 2 | 121 | (Iverson et al. 1993; Rhodin et al. 2017; CITES 2018; Google 2019; IUCN 2019f) |