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# Status and trends in Arctic vegetation: evidence from experimental warming and long-term monitoring

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**Anne D. Bjorkman** is a postdoctoral scholar at the Senckenberg Biodiversity and Climate Research Centre. Her research interests include biogeography, functional ecology, and community ecology, with a particular focus on the ecological consequences of climate change in tundra ecosystems.

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Status and trends in Arctic vegetation: evidence from experimental warming and long-term
 monitoring

3

#### 4 Abstract

5

6 Changes in Arctic vegetation can have important implications for trophic interactions and ecosystem 7 functioning leading to climate feedbacks. Plot-based vegetation surveys provide detailed insight into 8 vegetation changes at sites around the Arctic and improve our ability to predict the impacts of 9 environmental change on tundra ecosystems. Both long-term monitoring and experimental 10 manipulation of environmental drivers (e.g. temperature) have shown impacts of environmental 11 change, particularly climate warming, on tundra vegetation. Here, we review 30 studies of vegetation 12 change and 12 studies of phenological change from both long-term monitoring and warming 13 experiments in Arctic environments, including vegetation attributes identified in international 14 assessments as monitoring priorities. General responses to natural or experimental warming include 15 an increase in overall plant cover, particularly in the abundance of shrubs and grasses, and a decrease 16 in lichens and mosses, but the most common abundance response was one of no change. Plant 17 phenology often advanced in response to experimental warming, but trends over time varied by site, 18 species, and phenostage. This synthesis demonstrates that Arctic plant communities and species are 19 generally sensitive to warming, but trends over time are heterogeneous and complex and do not 20 always mirror expectations based on responses to experimental manipulations. Our findings highlight 21 the need for more geographically widespread, integrated and comprehensive monitoring efforts that 22 can better resolve the interacting effects of warming and other local and regional ecological factors. 23 24 Keywords 25 26 vegetation change, phenology, Arctic, experimental warming, long-term monitoring 27

#### 28 Introduction

29

30 A major goal of global change ecology is to document and predict the impacts of environmental 31 change on species, communities and ecosystems worldwide. In the Arctic, exceptionally rapid 32 warming (IPCC 2013) has the potential to lead to dramatic changes in vegetation through longer 33 growing seasons, increased thaw depth, and altered snow regimes. High latitudes contain up to 50% 34 of the world's soil carbon stored in permafrost soils; this carbon is vulnerable to loss with warming 35 (Schuur et al. 2015; Crowther et al. 2016; van Gestel et al. 2018). Thus, change in vegetation carbon 36 and nutrient inputs to tundra soils could have potentially global impacts. For example, shifts in 37 species composition could lead to changes in aboveground carbon storage, nutrient cycling,

38 decomposition rates, and albedo (Callaghan et al. 2004), potentially leading to global climate

- 39 feedbacks (Chapin et al. 2005; Pearson et al. 2013). Changing vegetation could also alter trophic
- 40 interactions (Post et al. 2009; Gauthier et al. 2013) and thus influence Arctic wildlife populations and
- 41 the human communities that rely on them for resource provision or cultural purposes (Weller et al.
- 42 2004; Henry et al. 2012; Stern and Gaden 2015).
- 43

44 A key source of information about the consequences of climate warming for Arctic vegetation comes 45 from plot-based research at sites across the Arctic (Henry and Molau 1997). This includes both long-46 term monitoring of species composition, diversity, and phenology over time (up to four decades), as 47 well as experimental manipulation of key abiotic and biotic drivers (e.g., temperature, snow, nutrients, 48 grazing). Community composition, diversity, and phenology have all been identified as "Focal 49 Ecosystem Components" (FECs) by the international Circumpolar Biodiversity Monitoring Program (Christensen et al. 2013), as monitoring of these attributes facilitates a more rapid detection, 50 51 communication, and response to significant biodiversity-related trends and pressures affecting the 52 circumpolar world. In addition, comparing the results of observed trends over time with experimental 53 studies can help to elucidate the drivers of observed trends and inform predictions of future change 54 (Elmendorf et al. 2015).

55

56 Here, we synthesize what is currently known about plot-based changes in vegetation composition 57 (abundance), phenology, diversity, and functional traits. We compiled information from single-site 58 studies of composition and phenological change to document 1) the direction and significance of 59 change over time, and 2) the direction and significance of responses to experimental warming. We 60 compare these results to published syntheses of long-term monitoring and experimental warming. We 61 additionally review studies of plot-based changes in plant functional traits and diversity, for which 62 published observations are relatively scarce. Finally, we discuss the broader implications of observed 63 and predicted Arctic vegetation change and recommend priorities for future monitoring efforts. 64

#### 65 Materials and Methods

- 66
- 67 Literature review of vegetation trends
- 68

69 We conducted a literature review to identify single-site studies of changes in plant community

- 70 composition (abundance) and phenology both over time and in response to experimental warming.
- 71 Our search included combinations of the terms "tundra", "arctic", "vegetation", "plot", "change",
- 72 "ITEX", "cover", "abundance", "phenology", "diversity", "functional trait", "warming", and
- 73 "experiment". These terms encompasses two Focal Ecosystem Components included in the
- 74 Circumpolar Biodiversity Monitoring Program terrestrial monitoring plan: i) diversity, composition

- and abundance and ii) phenology. We do not include the attributes "diversity and spatial structure",
- 76 "productivity", "Rare species, species of concern", or "food species" in this review due to a paucity of
- 77 published plot-based monitoring and/or experimental studies on these topics. The attribute "non-
- native species" is addressed in a separate article in this issue [WASOWICZ ET AL., THIS ISSUE].
  79

80 We included only studies at sites above 63 °N and identified as "Arctic" or "tundra" by the authors. 81 This latitudinal cut-off includes some sub-Arctic sites but is roughly comparable to areas included in 82 the Arctic Biodiversity Assessment (CAFF 2013 2013). For community composition/abundance, we 83 included measured responses in any variable called abundance, biomass, or percent cover. We 84 included studies that analyzed changes in abundance at both the species and functional group level. 85 For studies where abundance trends were identified at the species level, we included all species but 86 grouped them by functional group for visualization purposes. All phenological responses were 87 provided at the species level.

88

89 For phenological studies, we recorded all phenostages provided by the authors, but here we report 90 only the most commonly observed phenostages: leaf emergence, flowering, and leaf senescence. Leaf 91 emergence is the day at which leaf bud-break first occurs or the first day on which overwintered 92 leaves re-green. Flowering encompasses several phases related to the timing of flowering, including 93 inflorescence elongation, first open flower, onset of pollen release, and peak flowering. Leaf 94 senescence is the date on which leaves change color or die, indicating the end of the growing season 95 for most plants. Studies reporting responses of diversity and/or functional traits were scarce; thus, we 96 review the available information but do not attempt to categorize and quantify these responses.

97

98 For all studies we recorded the direction (increase/stable/decrease for abundance change, or 99 earlier/stable/later for phenological change) and significance (yes/no) of responses for all species and 100 functional groups identified. A response could be recorded as directional (increase/decrease or 101 earlier/later) and non-significant if the authors identified it as such, or if the p-value provided was 102 between 0.05 and 0.1. We adopted this approach in order to standardize alpha levels across all studies 103 (e.g. if some studies used an alpha level cut-off of 0.05 to assess significance while others used an 104 alpha level of 0.1). If a response was identified by the authors as directional but no indication of 105 significance was given (either in the text or in a figure/table), the response was categorized as non-106 significant. The difference between significant and non-significant directional changes is shown in the 107 figures and provided in the supplementary data table. We used this "vote-counting" approach, rather 108 than a traditional meta-analysis, in order to include the many studies that do not provide response 109 effect sizes or estimates of error. In addition, this approach allows us to visualize the full distribution 110 of vegetation responses to ambient and experimental warming, as a meta-analysis finding of "no111 change" could in fact be made up of multiple significant changes in different directions (e.g. context 112 dependency).

113

- 114 Experimental warming was generally conducted through the use of clear-sided, open-top chambers
- 115 that passively warm air temperatures by ~1.5-3 °C, with most of the studies following International
- 116 Tundra Experiment (ITEX) protocols (Molau and Mølgaard 1996; Marion et al. 1997), though some
- 117 experiments used greenhouses or other warming methods (Chapin and Shaver 1996; Wang et al.
- 118 2017). The seasonal duration of warming also varies by study; some warming chambers were in place
- 119 only during the summer, while others were present year-round. Both warming chambers and
- 120 greenhouses can influence environmental factors other than temperature (e.g. soil moisture, wind,
- 121 snow accumulation), though the magnitude and significance of these effects are variable among sites
- 122 (Marion et al. 1997).
- 123
- 124 Comparison to tundra-wide syntheses
- 125

126 In order to evaluate the consistency of patterns revealed by the literature review, we compared the results of our review with tundra-wide syntheses of community composition and phenological change 127 128 (Arft et al. 1999; Walker et al. 2006; Elmendorf et al. 2012a; Elmendorf et al. 2012b; Oberbauer et al. 129 2013), both over time and in response to experimental warming. These syntheses used primary data 130 and were not based on published studies, though some data included in the syntheses may be from the 131 same sites as the single-site studies included in our literature review. However, the synthesis and 132 single-site studies likely include different combinations of sites and years, and use different statistical 133 methods to analyze responses. In addition, many of the synthesis studies included both Arctic and 134 alpine tundra sites, while here we focused exclusively on Arctic and sub-Arctic locations. Thus, 135 evidence that synthesis studies found trends consistent with those documented in this literature review 136 can help evaluate the robustness of observed patterns in Arctic vegetation change.

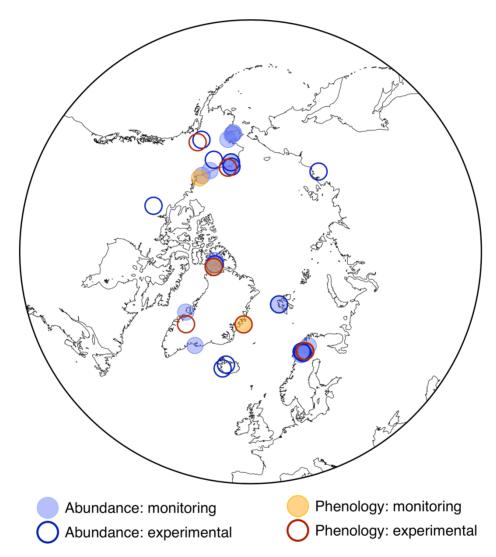
137

#### 138 **Results**

139

- We identified a total of 560 vegetation composition/abundance observations (species or functional
  group) from 19 studies of long-term monitoring and 209 observations from 14 studies of responses to
  experimental warming (Figure 1, Table 1). The duration of monitoring studies ranged from five to 43
- experimental warming (Figure 1, Fable 1). The datation of monitoring studies ranged nois rive to 45
- 143 years, with a median duration of 19 years. We additionally identified long-term monitoring of
- 144 phenology for 17 species in three studies and responses to experimental warming of 52 species from
- 145 nine studies (Figure 1, Table 1). Phenological monitoring studies ranged from nine to 21 years in
- 146 duration, with a median duration of 16 years.
- 147

- 148 Our literature review reveals geographical gaps in both long-term monitoring and experimental
- 149 warming studies. The FEC (Christensen et al. 2013) encompassing composition and abundance is
- 150 better represented than that encompassing phenology, but both lack published records of change from
- 151 Siberia and wide swaths of the Canadian Arctic. Intensive, multivariate monitoring is concentrated
- primarily in Alaska and Scandinavia, with the exception of one site in high-Arctic Canada (Muc et al.
- 153 1989; Freedman and Svoboda 1994; Hudson and Henry 2009; Hill and Henry 2011; Bjorkman et al.
- 154 2015).
- 155

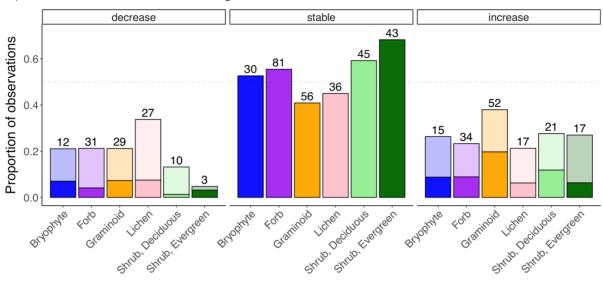


156

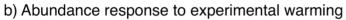
Figure 1. Map of plot-based vegetation change studies identified in a review of the literature. Blue points designate studies of community composition (abundance) change, while orange points designate studies of phenological change. Filled circles denote long-term monitoring studies (change over time) while open circles

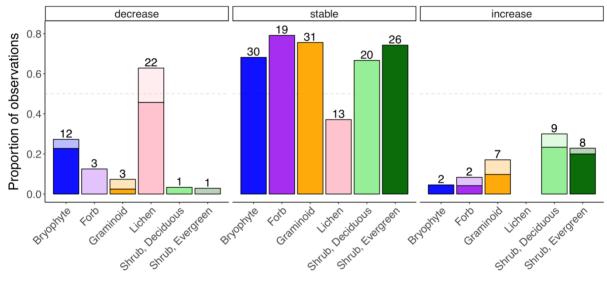
- 160 *indicate experimental studies (responses to experimental warming).*
- 161
- 162 Vegetation composition change
- 163 In all cases, the most common response documented by long-term monitoring of compositional
- 164 change was one of no trend (52-84% of trends did not differ from zero, depending on the significance

- 165 cutoff used; Figure 2). This is likely an underestimate of the proportion of no-change responses, as
- 166 some studies reported results only for species that changed significantly over time (Tømmervik et al.
- 167 2004). Forbs, graminoids and both evergreen and deciduous shrubs were slightly more likely to
- 168 increase in abundance over time than decrease, but were most likely to remain stable. Experimental
- 169 warming led to more dramatic responses, particularly in lichens, which were far more likely to
- decrease in abundance in response to experimental warming (46-63%) than to increase (0%) or
- 171 remain stable (37%). Bryophytes also had a tendency to respond negatively to experimental warming,
- 172 while evergreen and deciduous shrubs were more likely to respond positively.
- 173
- 174 Few studies included both above- and below-ground measurements; of those that did, above- and
- 175 below-ground responses were not always consistent. Above-ground responses to experimental
- 176 warming in northern Alaska were greater than below-ground responses (Chapin and Shaver 1996), but
- below-ground biomass increased more than above-ground biomass over 30 years of monitoring at
- 178 Alexandra Fiord, Ellesmere Island (Hill and Henry 2011). At Daring Lake, Canada, experimental
- 179 warming enhanced both above- and below-ground biomass in evergreen shrubs, but only above-
- 180 ground biomass in deciduous shrubs (Zamin et al. 2014).
- 181



a) Direction of abundance change over time





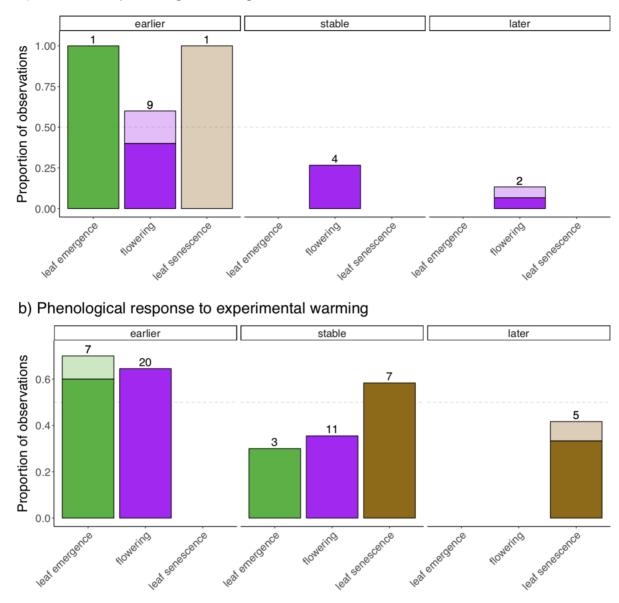


183 Figure 2. Summary of studies investigating abundance change over time (a) and abundance change in response 184 to experimental warming (b) by species or functional group. Panels represent, from left to right, the proportion 185 of observations decreasing in abundance, stable, or increasing in abundance over time (median 20.5 years) or in 186 response to experimental warming. Species-specific trends were grouped into the relevant functional group 187 category. The darker portions of each bar represent "significant" (p < 0.05) change (decrease or increase) or 188 insignificant (stable; p > 0.1) results, while lighter colors represent borderline or marginally significant change 189 (e.g., p-values between 0.05 and 0.1). The numbers above each bar represent a count of the number of 190 observations included in that group. The proportion of "stable" species is underrepresented in this figure, as 191 some studies only reported results for species that changed in abundance.

- 192 These responses are largely in line with those revealed in tundra-wide syntheses of vegetation change
- 193 at the functional group level. In a recent thirty-year study of vegetation change across 46 Arctic,
- alpine and Antarctic tundra locations (Elmendorf et al. 2012b), only evergreen shrubs (but not
- deciduous) increased significantly over time. Bryophytes were more likely to decrease than increase,
- 196 but the response was not significant. Similar to results from our literature review, responses to
- 197 experimental warming were more dramatic. Deciduous but not evergreen shrubs increased

- 198 significantly in abundance in response to experimental warming, while both lichens and bryophytes
- decreased significantly (Elmendorf et al. 2012a). In both monitoring and experimental synthesis
- 200 studies, the quantity of dead material (litter and attached dead) increased over time or with warming
- 201 (Elmendorf et al. 2012a; Elmendorf et al. 2012b).
- 202
- 203 Phenological change
- 204 In general, the timing of leaf emergence and flowering advanced both over time and with
- 205 experimental warming, though a minority of species experienced stable or even delayed flowering
- 206 over time (Figure 3). Interestingly, experimental warming led to later leaf senescence in all cases,
- while the one study that documents long-term trends in leaf senescence (Myers-Smith et al. 2018)
- 208 found a non-significant trend toward earlier leaf senescence over 16 years of monitoring.
- 209

a) Direction of phenological change over time



<sup>210</sup> 

Figure 3. Summary of studies investigating phenological change over time (a) and in response to experimental warming (b). Panels represent, from left to right, the proportion of observations that advanced ("earlier") in a given phenological stage, remained stable, or were delayed ("later") over time or in response to experimental warming. All observations represent species-specific responses. The darker portions of each bar represent "significant" (p < 0.05) change (decrease or increase) or insignificant (stable; p > 0.1) results, while lighter colors represent borderline or marginally significant change (e.g., p-values between 0.05 and 0.1). The numbers above each bar represent a count of the number of observations included in that group.

- 218 In a 17-year synthesis of phenological trends at 12 tundra sites (including two alpine locations), there
- 219 was no significant change in the timing of flowering or leaf senescence, though both events tended to
- advance over time (Oberbauer et al. 2013), as we also found in the literature review. Surprisingly, this
- same synthesis study found that leaf emergence was significantly delayed over time despite increasing
- temperatures over the same period. In a separate synthesis of responses to experimental warming at 10
- 223 Arctic sites, leaf emergence and flowering both occurred significantly earlier when warmed, but

- senescence was not affected (Arft et al. 1999). This is also in agreement with our literature review,
  where most sites reported no difference in leaf senescence or a slight delay.
- 226

#### 227 Discussion

228

229 Studies of plot-based vegetation change reveal that while some sites and species or functional groups 230 have experienced substantial shifts in vegetation and phenology in response to ambient or 231 experimental warming, the most common response overall is one of no change. Furthermore, 232 vegetation changes over time did not always match responses to experimental warming. One pattern 233 emerges: both long-term monitoring and experimental studies suggest that the graminoid and shrub 234 functional groups respond positively to warming and are slightly more likely to increase in abundance 235 over time. This is in line with studies of shrub infilling and expansion across much of the tundra 236 (Sturm et al. 2001; Myers-Smith et al. 2011a; Martin et al. 2017). Even so, the majority of graminoid 237 and shrub abundance responses in our literature review were that of no significant trend over time and 238 no significant response to experimental warming. Further exploration of these trends reveals that 239 increasing shrub abundance primarily occurs in relatively warm tundra regions with mesic or wet 240 soils, while colder and dry tundra sites have not experienced increasing shrub abundance (Elmendorf 241 et al. 2012b) consistent with patterns in the climate sensitivity of shrub growth (Myers-Smith et al. 242 2015). Grazing may also influence shrub responses to summer temperature change (Bråthen et al. 243 2017). The lack of strong trends over time in many sites and for many species suggests that tundra 244 plant communities are remarkably resilient to moderate warming, at least over decadal time spans, 245 and that site-specific factors such as moisture availability and grazing may limit vegetation responses 246 to warming (Elmendorf et al. 2012b; Myers-Smith et al. 2015; Ackerman et al. 2017).

247

Differing responses to experimental and ambient (natural) warming highlight both the benefits and the
 challenges of using experimental approaches to understand tundra vegetation responses to climate
 change. When experimental and monitoring results agree, experimental studies enable us to pinpoint

the likely drivers of change over time (Elmendorf et al. 2015), and improve our confidence in

252 predictions of the impacts of warming on vegetation (e.g., increasing shrub abundance). Diverging

responses can challenge our understanding of tundra vegetation change. For example, while

- experimental warming led to fairly dramatic declines in lichen abundance (Walker et al. 2006;
- Elmendorf et al. 2012a and this study), lichen abundance did not decline over time in long-term
- 256 monitoring studies (Elmendorf et al. 2012b and this study). Similarly, while experimental warming
- 257 generally led to later leaf senescence (Arft et al. 1999 and this study), monitoring studies indicate that
- senescence is in fact advancing over time, though not significantly (Oberbauer et al. 2013 and this
- study). Trends in the timing of flowering and leaf emergence are also varied despite a relatively
- 260 consistent advance in these variables in response to experimental warming.

261

- 262 The reasons underlying these heterogeneous and contrasting trends are not entirely clear, but may 263 have to do with interactions among environmental drivers that are not captured by experimental 264 isolation of a single driver. For example, lichens are sensitive to soil moisture, and may be responding 265 to changes in precipitation, hydrology, or snow regimes over time rather than temperature trends 266 alone (Björk and Molau 2007). Similarly, phenological advance with warming temperatures (Høve et 267 al. 2007) may be limited by concurrent changes in winter snowfall (Bjorkman et al. 2015) and 268 snowmelt date (Cooper et al. 2011). In addition, growing season phenology might be controlled by 269 deterministic leaf age (Starr et al. 2000) or adaptation to photoperiod (Kummerow 1992; Bjorkman et 270 al. 2017) in many Arctic species, thus limiting the impact of temperature change alone. Phenological 271 responses to different drivers may interact or be non-linear, leading to more complex responses than 272 can be easily detected from simple experiments or ecological monitoring studies (Iler et al. 2013). 273 Finally, experimental warming chambers can alter conditions other than temperature alone (Marion et 274 al. 1997), and vegetation could be responding to these unwanted environmental side-effects.
- 275

Improved monitoring of multiple environmental drivers and experimental studies that manipulate several variables simultaneously (e.g. snow depth, moisture availability) could help to elucidate the importance of these interactions. Multi-site syntheses can also help to clarify the context dependency of trends over time. For example, additional syntheses of tundra plant phenology have shown that a plant's sensitivity to temperature varies by the temperature of the site (greater sensitivity at colder sites; Prevéy et al. 2017) as well as the phenological niche of the species (greater sensitivity in lateflowering species; Prevéy et al. 2018).

283

#### 284 Other vegetation trends: functional traits and diversity

285

286 While trends in composition and phenology are perhaps the most studied plot-based responses to 287 global change, a handful of studies document changes in other vegetation parameters. Of these, 288 increasing height is likely the most well-documented. Increases in community height have been 289 documented by synthesis studies of responses to experimental warming (Elmendorf et al. 2012a) and 290 over time (Bjorkman et al. 2018), a change driven primarily by the influx of taller species into the 291 monitoring plots (Bjorkman et al. 2018). Some single-site studies have also documented increasing 292 height over time (Hollister et al. 2015) and in response to experimental warming (Hudson et al. 2011; 293 Hollister et al. 2015; Baruah et al. 2017). Changes in other plant traits have also been documented. 294 Experimental warming at Alexandra Fiord in high Arctic Canada resulted in greater leaf size, lower 295 specific leaf area (the ratio of leaf area to leaf dry mass), and decreased leaf carbon content for at least 296 some species-site combinations (Hudson et al. 2011) but did not affect leaf nitrogen (N) content, leaf 297 dry matter content, or nitrogen isotope signatures. Other studies have documented mixed responses of

298 leaf size to experimental warming in the Swedish subarctic tundra (Graglia et al. 1997; Baruah et al. 299 2017) and one study found trends toward reduced leaf size over time (Barrett et al. 2015). Additional 300 studies of leaf N content responses to experimental warming are also mixed; leaf N content increased 301 in response to winter but not summer warming across six species at Eight Mile Lake, Alaska (Natali 302 et al. 2012) but was either unaffected by temperature or declined in response to warming at Toolik 303 Lake, Alaska (Chapin and Shaver 1996) and Alexandra Fiord, Canada (Tolvanen and Henry 2011). A 304 synthesis of community-weighted mean functional trait change across the tundra biome (including 305 alpine sites) over 27 years found no significant change in leaf area, leaf N content, leaf dry matter 306 content, or specific leaf area (Bjorkman et al. 2018). Overall, species composition has shifted toward 307 more thermophilic (warm-loving) species both over time and in response to experimental warming 308 (Elmendorf et al. 2015).

309

310 Over the long term, climate warming may lead to increased diversity in the Arctic as southern, 311 species-rich floras move northward (Parmesan 2006). However, short-term responses to warming 312 might differ substantially from long-term trends, as immigration is likely to be slow relative to local 313 assembly processes (e.g. competition; Walker et al. 2006). Thus far, evidence of plot-scale diversity 314 change in Arctic ecosystems is mixed. A multi-site synthesis found a significant decline in both 315 Shannon diversity and species richness after three to six years of experimental warming (Walker et al. 316 2006), but a more recent, longer-term synthesis found no response (Elmendorf et al. 2012a). Lichen 317 diversity was found to decline significantly in response to long-term experimental warming at three 318 sites in northern Sweden and Alaska (Lang et al. 2012). Among monitoring studies, a recent synthesis 319 found no change in vascular plant diversity over three decades of monitoring across dozens of tundra 320 sites (Elmendorf et al. 2012b). This is in stark contrast to ongoing changes in European mountaintop 321 plant communities, which have experienced rapid and accelerating increases in richness over the past 322 century (Steinbauer et al. 2018). This difference could indicate that diversity change in non-alpine 323 tundra communities is limited by dispersal rates of southerly, warm-adapted species, or that strong 324 gradients in environmental variables other than temperature (e.g. photoperiod) across latitudes limit 325 the establishment success of warm-adapted species from farther south (Bjorkman et al. 2017).

326

327 *Consequences of Arctic vegetation change* 

328

329 Changes in tundra vegetation could have far-reaching impacts across trophic levels and to human

330 societies (Weller et al. 2004). Shifts in plant phenology and reproductive success influence individual-

and population-level fitness (Berteaux et al. 2004; Cleland et al. 2012) and could lead to trophic

mismatches of resources for pollinators (Høye et al. 2013; Wheeler et al. 2015; Prevéy et al. 2018),

333 breeding birds (McKinnon et al. 2012; Gauthier et al. 2013; Boelman et al. 2015) and mammals

334 (Hertel et al. 2017). For example, one long-term study at Zackenberg, Greenland documented a

shortening of the flowering season with climate warming over time and a concurrent decline in the
abundance of insect visits to flowers (Høye et al. 2013). Berry-producing (Hertel et al. 2017) and
other tundra plants provide forage for hunted or domestic wildlife (Post and Stenseth 1999; Kerby and

- Post 2013) and represent culturally important resources for Arctic peoples (Henry et al. 2012).
- 339

340 The nearly ubiquitous shifts in phenology in response to experimental warming (Arft et al. 1999 and 341 this study) suggest that many Arctic plant species are inherently sensitive to interannual variations in 342 temperature, though concurrent changes in other environmental variables (e.g. precipitation, 343 cloudiness) might limit the degree of advance over time with warming. A meta-analysis of 344 phenological responses to experimental warming in temperate and alpine regions found that the 345 temperature sensitivity of a species' phenology correlates with better growth and/or reproductive 346 performance (Cleland et al. 2012), but it is not known if this pattern holds true in the Arctic. A 347 synthesis of responses to four years of experimental warming at 10 Arctic sites revealed increased 348 reproductive effort (e.g. number of flowers produced) and success (e.g. number of seeds/fruits 349 produced or seed mass) in experimentally warmed plots, though responses were generally not 350 significant (Arft et al. 1999). Single-site studies have also found evidence of increased reproductive 351 effort in experimentally warmed plots (Welker et al. 1997; Klady et al. 2011). Contrasting responses 352 have been documented for seed germination rates, which increased with experimental warming at 353 Alexandra Fiord, Canada (Klady et al. 2011) but not at Toolik Lake, Alaska (Welker et al. 1997).

354

355 Due to the large amount of carbon stored in tundra permafrost soils (Koven et al. 2011; Schuur et al. 356 2015; Crowther et al. 2016) and well-established links between vegetation and carbon storage, 357 vegetation change in the Arctic can influence regional carbon cycling and feedbacks to the global 358 climate (Callaghan et al. 2004; Sturm and Douglas 2005; Petrenko et al. 2016). For example, 359 increasing shrub abundance and/or plant height can lead to increased winter snow trapping, greater 360 insulation of underlying soils, warmer winter soil temperatures (Myers-Smith and Hik 2013), and 361 potentially increased active layer depth and decomposition (Blok et al. 2016). Taller shrubs may also 362 extend above the snowpack, decreasing winter albedo and increasing absorbed solar radiation (Sturm 363 and Douglas 2005). Bryophytes have also been shown to play an important role in soil insulation and 364 energy fluxes; experimental removal of bryophytes leads to increased evapotranspiration and ground 365 heat flux (Blok et al. 2011). Thus, future declines in bryophytes – observed in warming experiments 366 but not yet in monitoring studies – could also lead to deeper summer permafrost thaw and soil carbon 367 release, representing another positive feedback to climate warming. 368

369 Changing vegetation can also impact carbon cycling through changes in the quantity and

decomposability of litter (Callaghan et al. 2004), as litter decomposition contributes nearly 70% of

371 global CO<sub>2</sub> fluxes from soils (Raich and Potter 1995). A long-term increase in shrubs, which have

- 372 relatively recalcitrant litter, could lead to reduced litter decomposability and a negative feedback to
- 373 climate warming (Cornelissen et al. 2007). A change in litter composition can also indirectly
- influence soil carbon storage by driving changes in soil microbial communities (Christiansen et al.
- 375 2018) or altering tundra fuel loads. For example, increased woody litter inputs from shrub expansion
- 376 might also increase flammability, which could lead to positive feedbacks through fire-induced soil
- 377 carbon loss (Cornelissen et al. 2007; van Altena et al. 2012).
- 378

### 379 Conclusions

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381 Rapid warming in the Arctic has the potential to cause substantial shifts in vegetation, potentially 382 driving widespread changes across trophic levels and altering tundra ecosystem functions. While our 383 review identifies significant shifts at some sites and in some species, the large variation in the 384 magnitude and even direction of responses illustrates the high degree of context dependency in tundra 385 vegetation change. This context dependency highlights the importance of maintaining multiple 386 monitoring sites in many different habitat types across the entire Arctic, as well as increasing 387 monitoring of local ecological and environmental conditions that would improve our understanding of 388 how factors other than temperature influence Arctic vegetation change. Thus, we recommend that 389 international bodies such as the Circumpolar Biodiversity Monitoring Program (Christensen et al. 390 2013) prioritize monitoring efforts that i) fill current geographical gaps, particularly in Canada and 391 Siberia, and ii) enable us to better disentangle the relative importance of climate warming and other 392 environmental factors on the diverging responses reported here. 393

#### 394 Table 1. Studies of abundance and phenology included in this review.

					Abundance		Phenology	
Study	Site name	Lat	Lon	Duration (years)	monitoring	experiment	monitoring	experiment
(Alatalo and Totland 1997)	Latnjajaure, Lapland, Sweden	68.21	18.3	1				Х
(Bjorkman et al. 2015)	Alexandra Fiord, Ellesmere Island, Canada	78.53	-75.55	21			X	х
(Boulanger-Lapointe et al. 2014)	Alexandra Fiord, Ellesmere Island, Nunavut	78.86	-75.9	13-15	х			
(Boulanger-Lapointe et al. 2014)	Sverdrup Pass, Ellesmere Island, Nunavut	79.13	-79.73	5-23	х			
(Callaghan et al. 2011)	Disko Island, Greenland	69.15	-53.34	43	х			
(Chapin and Shaver 1996)	Toolik Lake, Alaska	68.38	-149.34	4		х		
(Chapin et al. 1995)	Toolik Lake, Alaska	68.38	-149.34	9		х		
(Daniëls and de Molenaar 2011)	Tasiilaq, Southeast Greenland	65.62	-37.67	41	Х			
(Graglia et al. 2001)	Abisko, Sweden	68.35	18.82	10		х		
(Hill and Henry 2011)	Alexandra Fiord, Ellesmere Island, Canada	78.53	-75.55	25	х			
(Hobbie and Chapin 1998)	Toolik Lake, Alaska	68.38	-149.34	3		х		
(Hollister and Webber 2000)	Barrow, Alaska, USA	71.18	-156.4	1				Х
(Hollister et al. 2015)	Atqasuk, Alaska	70.45	-157.41	16	х	х		
(Hollister et al. 2015)	Barrow, Alaska	71.29	-156.64	17	х	Х		
(Høye et al. 2007)	Zackenberg, Greenland	74.28	-20.34	9			Х	
(Hudson and Henry 2009)	Alexandra Fiord, Ellesmere Island, Nunavut	78.88	-75.92	28	Х			
(Hudson and Henry 2010)	Alexandra Fiord, Ellesmere Island, Nunavut	78.88	-75.92	16		х		
(Jägerbrand et al. 2009)	Latnjajaure, Lapland, Sweden	68.35	18.5	5		Х		
(Jandt et al. 2008)	Northwestern Alaska	65.1	-163.4	10-15	Х			
(Joly et al. 2007)	Seward Peninsula, Alaska	64.85	-163.7	25	Х			
(Jonasson et al. 1999)	Abisko, Sweden	68.35	18.82	5		Х		

(Jones et al. 1997)	Alexandra Fiord, Ellesmere Island, Canada	78.53	-75.55	1				Х
(Jones et al. 1997)	Barrow, Alaska, USA	71.19	-156.37	1				Х
(Jones et al. 1997)	Latnjajaure, Lapland, Sweden	68.21	18.3	1				Х
(Jónsdóttir et al. 2005)	Audkuluheidi, Iceland	65.27	-20.25	5		Х		
(Jónsdóttir et al. 2005)	Thingvellir, Iceland	64.28	-21.08	5		Х		
(Jorgenson et al. 2015)	Arctic National Wildlife Refuge, Alaska	69.8	-144.25	26	Х			
(Marchand et al. 2004)	Zackenberg, Greenland	74.28	-20.34	1				Х
(Molau 2010)	Latnjajaure, Lapland, Sweden	68.35	18.5	12	Х			
(Myers-Smith et al. 2011b)	Qikiqtaruk-Herschel Island, Yukon	69.57	-138.91	11	Х			
(Myers-Smith et al. 2018)	Qikiqtaruk-Hershel Island, Yukon, Canada	69.57	-138.91	16-19	х		х	
(Natali et al. 2012)	Eight Mile Lake, Alaska, USA	63.52	-149.13	2		Х		Х
(Pattison et al. 2015)	Arctic National Wildlife Refuge, Alaska	69.8	-144.25	26	X			
(Post and Pedersen 2008)	Kangerlussuaq, Greenland	67.6	-50.2	2				Х
(Richardson et al. 2002)	Abisko Valley, Sweden	68	19	9		Х		
(Robinson et al. 1998)	Ny Alesund, Svalbard	78.93	11.83	5	Х	Х		
(Rundqvist et al. 2011)	Abisko Valley, Sweden	68.35	18.82	35	X			
(Stenström and Jónsdóttir 1997)	Latnjajaure, Lapland, Sweden	68.22	18.13	1				х
(Tømmervik et al. 2004)	Kautokeino, Norway	69	23.1	38	Х			
(Villarreal et al. 2012)	Barrow, Alaska	71.3	-156.67	39	Х			
(Vowles et al. 2017)	Ritsem, Sweden	67.82 4	17.715	18	х			
(Wang et al. 2017)	Kytalyk, Siberia	70.82	147.48	4		Х		
(Wilson and Nilsson 2009)	Cievrratjäkka, Sweden	68.01	18.81	21	Х			
(Wookey et al. 1993)	Abisko, Sweden	68.21	18.49	1				Х
(Zamin et al. 2014)	Daring Lake, NWT	64.87	-111.57	8		Х		

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