



Edinburgh Research Explorer

Detecting the True Extent of Introgression during Anthropogenic Hybridization

Citation for published version:

McFarlane, SE & Pemberton, JM 2019, 'Detecting the True Extent of Introgression during Anthropogenic Hybridization', *Trends in Ecology & Evolution*, vol. 34, no. 4, pp. 315-326. https://doi.org/10.1016/j.tree.2018.12.013

Digital Object Identifier (DOI):

10.1016/j.tree.2018.12.013

Link:

Link to publication record in Edinburgh Research Explorer

Document Version:

Peer reviewed version

Published In:

Trends in Ecology & Evolution

General rights

Copyright for the publications made accessible via the Edinburgh Research Explorer is retained by the author(s) and / or other copyright owners and it is a condition of accessing these publications that users recognise and abide by the legal requirements associated with these rights.

Take down policy
The University of Edinburgh has made every reasonable effort to ensure that Edinburgh Research Explorer content complies with UK legislation. If you believe that the public display of this file breaches copyright please contact openaccess@ed.ac.uk providing details, and we will remove access to the work immediately and investigate your claim.



Download date: 11. May. 2020

- Detecting the true extent of introgression during anthropogenic hybridization
 S. Eryn McFarlane^{1,2*} & Josephine M. Pemberton¹
- 4 1. Institute of Evolutionary Biology, School of Biological Sciences, University of
- 5 Edinburgh, Charlotte Auerbach Road, Edinburgh, United Kingdom, EH9 3FL
- 6 2. Department of Biology, Lund University, Sölvegatan 37, 223 62 Lund, Sweden
- 7 *author for correspondence: eryn.mcfarlane@ed.ac.uk

Abstract

8

9

- 10 Hybridization among naturally separate taxa is increasing due to human impact,
- and can result in taxon loss. Previous classification of anthropogenic
- 12 hybridization has largely ignored the case of bimodal hybrid zones, in which
- 13 hybrids commonly mate with parental species resulting in many backcrossed
- individuals with a small proportion of introgressed genome. Genetic markers can
- be used to detect such hybrids, but until recently too few markers have been
- used to detect the true extent of introgression. Recent studies of wolves and
- 17 trout have used thousands of markers to reveal previously undetectable
- 18 backcrosses. This improved resolution will lead to increased detection of late
- 19 generation backcrosses, shed light on the consequences of anthropogenic
- 20 hybridization, and pose new management issues for conservation scientists.

Anthropogenic hybridization

Anthropogenic hybridization (see Glossary), in which human disturbance leads to range overlap and hybridization of previously reproductively isolated populations or species is a growing conservation concern [1-3]. With increased human-generated movement of species into new ranges, there is an increasing number of cases of hybridization between species that were historically allopatric [4]. Disturbance of habitats can also result in a breakdown of reproductive isolation between previously isolated, sympatric species [1]. Introgression is usually hard to detect from phenotypes and there is growing evidence that backcrossing has often proceeded further than is detectable by low density genetic marker panels. In this article we make the case that genomic approaches are essential and increasingly available to disentangle late generation backcrosses from parental populations after introgression has occurred.

The benefits of anthropogenic hybridization

There are possible benefits of anthropogenic hybridization. Policy makers can use hybridization as a management tool to help endangered populations. In 'genetic rescue' programs (i.e. breeding programs designed to release small populations from inbreeding depression), individuals from a closely related population or subspecies are introduced to an inbred population to manage inbreeding depression. For example, when Florida panthers (*Puma concolor* coryi) were threatened due to inbreeding depression, eight Texas panthers (P. concolor cougaur) were introduced. The hybrid kittens survived better, and the population is now recovering [5]. Approximately 90% of such genetic rescue attempts have been successful, showing that anthropogenic hybridization is a viable conservation method [6]. Adaptive introgression ('evolutionary rescue') in which beneficial alleles from an introduced population are selected for in hybrid individuals is another possible benefit of anthropogenic hybridization. For example, a segment of chromosome 15 that has naturally introgressed from *Populus balsamifera* into *P. trichocarpa* appears to allow *P. trichocarpa* to live in colder, drier areas than *P. trichocarpa* individuals without this haplotype [7]. This suggests that there is potential for adaptive introgression to facilitate

55	evolutionary rescue of populations at risk of extinction due to climate change [8],
56	although such genomic management of at risk populations much enabling
57	research, and should be approached with caution [9, 10].
58	
59	The problems with anthropogenic hybridization
60	Anthropogenic hybridization can cause problems for native species. When no
61	offspring or sterile offspring are produced, reproductive effort is wasted [11].
62	When fertile F1s are formed, introgression between the two previously diverged
63	species is possible. There are two reasons why even low levels of introgression
64	of non-native alleles are of concern from a conservation perspective. First, if all
65	individuals of a species are hybrids then the species as it was is extinct. This has
66	been termed 'extinction by hybridization' [11-15]. Note, however, that there may
67	still be many copies of the native alleles represented in the population, so long as
68	the population itself is large enough, and from a 'gene view point' we may be
69	content with this mode of conservation [16].
70	
71	The second problem with hybridization is that introgression and recombination
72	break up linked gene complexes, and non-native alleles that are favoured (or no
73	longer in linkage with deleterious alleles) can be swept to fixation [17]. While
74	this leads to an initial increase in biodiversity (because alleles from both the
75	native and non-native populations are present) as non-native alleles sweep to
76	fixation, native alleles are lost. If we again take a gene view point of biodiversity,
77	any alleles lost from the native population are a loss in biodiversity from the
78	system. For example, non-native alleles at three out of 68 genetic markers have
79	gone to fixation in some populations of California Tiger Salamanders
80	(Ambystoma californiense) after hybridization with Barred Tiger Salamanders (A.
81	mavortium) [18]. This has occurred in California Tiger Salamander populations
82	that are nearly 100km from the original Barred Tiger Salamander introduction
83	site, suggesting that these alleles have higher fitness than the native, California
84	Tiger Salamander alleles that they have replaced [18].

Goals of studies of anthropogenic hybridization

Studies of anthropogenic hybridization have different goals. A researcher might be interested to know if hybridization has occurred at all in a population to determine whether it should provide the breeding stock for new populations, and or whether it should be guarantined because of hybridization. Relatively few informative markers are needed to detect individuals of hybrid origin in any particular population, as the detection of any non-native allele is a clear indication of hybridization [19]. However, if a researcher wishes to understand more about the underlying process of hybridization and introgression, then many more markers are required. Specific goals might include: to select individuals for breeding programs; to understand the relationship between genotype and phenotype; to understand the type of hybrid system involved (see next section); and to investigate mating patterns and fitness. For any of these goals, it is ideal to quantify individual **admixture** accurately, and to do this this hundreds or thousands of informative markers may be required (see below). **Classifying hybridization** To assist researchers and policy makers in addressing anthropogenic hybridization, Allendorf and colleagues [11] categorized hybridization outcomes. Types 1-3 applied to naturally-occurring hybridization while Types 4-6 applied to anthropogenic hybridization. Type 4 results in few or sterile F1 hybrids, and is characterized by wasted reproductive effort. Type 5 results in a **hybrid swarm** with widespread introgression into particular populations, but some populations do not experience hybridization at all. Finally, Type 6 results in a complete hybrid swarm following break down of reproductive isolation between species across all populations [11]. Three axes of variation determine the outcome of anthropogenic hybridization: differences in hybrid fitness, time since **secondary contact**, and mating patterns of hybrids. Time since secondary contact and mating patterns of hybrids were not explicitly considered in Allendorf et al's original categorization. Type 4 differs from Types 5 and 6 along an axis of hybrid fitness, where intrinsic post

87

88

89

90

91

92

93

94

95

96

97

98

99

100

101

102

103

104

105

106

107

108

109

110

111

112

113

114

115

116

117

118

zygotic isolation affects hybrids in Type 4, but not in Types 5 or 6. This results in little to no backcrossing in Type 4 hybrid zones, as hybrids are extremely unfit compared to parental species. This decrease in hybrid fitness must be extreme, as even with a 90% decrease in fitness, the proportion of hybrids in a hybridizing population is expected to increase [20]. We suggest that the only difference between Allendorf et al's [11] Type 5 and Type 6 is time since secondary contact. When an F1 reproduces, all of its offspring and descendants are admixed to some extent [20]. If Type 5 characterizes a system where only one or few populations have introgression, Type 6 is the logical outcome of this same system, assuming random mating and sufficient time for migration between populations. Thus, we consider Type 5 and Type 6 to be the same, both hybrid swarms with a breakdown of assortative mating, in which hybrids have the same mating success as either of the parental species individuals, and common enough that hybrid x hybrid matings occur. When there is a preference among hybrids for parental species phenotypes, or hybrids are very rare, we expect a different pattern of introgression. Backcrossing into the parental species leads to an increasingly large number of individuals with a small proportion (<10%) of their genome that is from the opposite species. As backcrossing continues, morphological differences between parental species and backcrossed individuals lessen, making it more and more difficult to detect a backcross using only phenotypic traits. This results in many hybrid individuals with very small proportions of another genome, although with a maintained bi-modal distribution of trait values between the two parental species (Figure 1). From a conservation perspective we consider this to be a worst-case scenario as these introgressed individuals are very difficult to detect. This can be contrasted with a general lack of assortative mating, in which hybrid individuals are as likely to breed with other hybrid individuals as with parental species (leading to a hybrid swarm), or, in the unlikely event of true assortative mating, where hybrid individuals preferentially breed with each other, which would lead to the eventual formation of a hybrid species e.g. [21]. The contrast between hybrid zones with unimodal distributions of traits and admixture

120

121

122

123

124

125

126

127

128

129

130

131

132

133

134

135

136

137

138

139

140

141

142

143

144

145

146

147

148

149

150

151

153 scores and those with bimodal distributions has previously been described in the 154 context of naturally occurring hybrid zones [22], but does not yet seem to inform 155 studies of anthropogenic hybridization. 156 157 The distribution of hybrid scores in a system at equilibrium varies depending on 158 ecological factors that can affect hybrid fitness, and hybrid encounter rate. 159 Extrinsic post zygotic isolation can vary according to ecological factors, affecting 160 the ability of hybrids to successfully mate and reproduce [23]. Further, stochastic 161 factors, particularly when hybrids are rare, or management might alter the 162 reproductive success of hybrid individuals in wild systems. However, if hybrids are fertile, the proportion of hybrid individuals in all populations should increase 163 164 [20], leading to the extreme end points of majority hybrid populations which either follow a hybrid swarm or **bimodal hybrid zone** distribution. 165 166 167 Key considerations for genetic analyses of anthropogenic 168 hybridization 169 170 Published studies of anthropogenic hybridization generally follow a similar 171 protocol. Researchers use codominant marker genotypes to estimate divergence 172 between the two species [24] and then use a clustering approach such as 173 STRUCTURE [25-28], or ADMIXTURE [29, 30] to partition individuals into 174 different genetic groups (K). Those individuals with an admixture score (Q) 175 intermediate to the extreme admixture scores associated with parental species 176 individuals are designated hybrids. Many studies then use HYBRIDLAB [31, 32] 177 or similar methods to simulate hybrid genotypes from the sampled genotypes to 178 assess the **efficiency** (i.e. type II error rate, rate of assigning hybrid individuals 179 as parental species), and **accuracy** (i.e. type I error rate, rate of erroneously 180 assigning parental species individuals as hybrids; [33]). The 'overall 181 performance' of an analysis is the product of efficiency and accuracy and this 182 performance can be used to assess the reliability of the study itself [33]. Here we 183 outline some best practices and points to consider in order to avoid 184 underestimation of the extent of hybridization.

Divergence between parental species

It is highly relevant to have an estimate of divergence between the focal species in the absence of hybridization. \mathbf{F}_{st} is often reported in studies of anthropogenic hybridization, but is rarely used to motivate the marker density deployed for estimates of individual admixture, typically because the same markers are used to determine both \mathbf{F}_{st} and individual Q estimates. Simulations have clearly shown that species (or subspecies) with lower divergence will require more markers to accurately estimate admixture, because of shared polymorphisms between them, leading to fewer **diagnostic markers** [33]. While it might not be practical to use markers to estimate \mathbf{F}_{st} and then determine how many markers are needed to estimate individual admixture scores, an initial assessment of \mathbf{F}_{st} will hint at how much power a system has to detect advanced backcrosses.

Historical admixture

Many systems have a history of repeated secondary contact and hybridization. Documenting historical admixture using genomic resources can determine whether the introgression found is due to recent, anthropogenic forces, or to natural causes, which will change the conservation status of the situation [34, 35]. There are techniques for detecting historical admixture. For example, the ABBA-BABA test can be used to determine if there has been historical introgression from a third species or population into each of two closely related sister taxa, to explain variation that is not well explained by a null assumption of bifurcating phylogeny [36]. This technique can be applied to either sequences of single individuals from each population, or to multiple individuals from each population [37], and can be used to indicate historical (hundreds to thousands of generations before present) admixture. Similarly, $\delta a \delta i$ analyses can be used to determine how well different demographic models fit the pattern of variation in the data, where demographic models can include admixture at different time points [38]. For example demographic modeling was used to demonstrate that hybridization between golden-winged (Vermivora chrysoptera) and blue winged warblers (*V. cyanoptera*) has probably been occurring since the original species split, and is not solely due to anthropogenic forces [39]. Finally, researchers can

219 examine the length of haplotype blocks that are identical by descent, as linkage 220 disequilibrium decays over time due to recombination [40, 41]. The distribution 221 of haplotype block lengths should follow a Poisson distribution [41] and 222 deviation from this distribution can be used to infer population admixture over 223 both short (tens of generations) [42] and long time spans [41]. These and other 224 techniques for disentangling historical and contemporary admixture are 225 reviewed in [43]. 226 227 Generations since secondary contact and recombination rates 228 It is important to estimate the number of generations since secondary contact to 229 estimate the potential number of backcross generations in a system. This 230 estimate might have substantial uncertainty, but in many cases of anthropogenic 231 hybridization there are historical records that suggest when a non-native species 232 was first introduced or sighted that can be combined with typical generation 233 times for the taxa involved. The expected proportion of invasive genome in a 234 backcrossed individual halves with each successive generation of backcrossing 235 [44]. 236 237 Recombination each generation leads to less linkage disequilibrium between 238 non-native loci, which means that genotype at a species-specific marker in one 239 position is less informative about surrounding, un-sampled loci. For example, 240 genomic regions with high recombination rates were found to be associated with 241 more introgression of the non-native genome in replicate swordtail (Xiphophorus 242 birchmanni and X. malinche) hybrid zones [17]. Due to obligatory crossing over, 243 which is expected to occur once per chromosome arm [45], at least twice as 244 many markers as there are chromosome arms are needed to cover each 245 independent section of the genome. In some cases, there is a species-specific 246 estimate of recombination (e.g. [46]), or one can refer to taxon-specific patterns. 247 For example, there is as much as 10 times more recombination in avian genomes 248 than in mammalian genomes [47]. Additionally, information on recombination 249 rate can be combined with genomic methods examining haplotype block lengths 250 to date introgression events (as discussed above). We discuss how many 251 markers are needed further in Box 1.

252 253 Assessing the power of markers 254 Many studies of anthropogenic hybridization assess the power of genetic 255 markers used by simulating hybrid genotypes and then determining the power 256 the markers have to detect these hybrid genotypes [48]. When assessing the 257 power of markers in this way, it is important to ensure that the biology of the 258 system is reflected in the simulation. In particular, if the two species of interest 259 have been in contact for many generations and F1s are thought to be fertile 260 (Figure 1), then simulations should account for the possibility of many 261 generations of backcrossing. This is rarely done in conservation genetic studies -262 many studies simulate backcrosses to assess the power of their markers, and 263 find low power to detect even first generation backcrosses, for example finding 264 less than 80% of first generation backcrosses are properly assigned [49, 50]. 265 Further information obtained from laboratory or field studies, such as 266 asymmetry in hybrid fertility (e.g. between sexes, Haldane's Rule [51] or 267 according to the species of the mother of the F1, Darwin's Corollary [52]), should 268 also be included in simulations. For example, if previous laboratory work has 269 established that backcrossing is largely unidirectional because of decreased 270 fitness of hybrid individuals in the opposite direction (as expected by Darwin's 271 Corollary) or due to the relative abundance of the parental species, then 272 mitochondrial markers should be integrated into future analyses to add power to 273 detect hybrids. 274 275 **Defining hybrid individuals** 276 To be defined as a hybrid, a focal individual must be genetically differentiated 277 from both parental species. Parental species are assumed to have an admixture 278 (Q) score of 0 or 1, although because of error (e.g. non-diagnostic markers, 279 genotyping errors), very few individuals will have an estimated score of exactly 0 280 or 1. Any score in between indicates a hybrid [25]. It is typical for a researcher to 281 set a Q score as a cut-off for hybrid individuals, so any individual above (or 282 below) this score is considered parental. Thresholds are determined either by

power, specifically, at what level can the markers differentiate between hybrids

and parental species, or by the number of acceptably mis-matched markers, e.g.

283

285 one allele indicative of the other species might be an error, but two markers 286 suggest hybridization [53]. These thresholds can range widely between studies, 287 from 0.8 [54] to 0.999 [30] in relation to a parental species score of 1.0. 288 Determination of the threshold is a balancing act between Type I and Type II 289 errors, in which the researcher must decide whether it is better to mistakenly 290 assign a parental species individual as a hybrid (Type I; too low 'accuracy'; [33]) 291 or assign hybrid individuals as parental types (Type II; too low 'efficiency'; [33]). 292 If the researcher accepts a higher level of Type II errors, they consider advanced 293 backcrosses as parental species. For example, an admixture score threshold of 294 0.8 would include most second-generation backcrosses (87% of the genome is 295 species A, 13% of the genome is species B on average) as parental species. 296 Similarly, with a Q score of 0.9, third-generation backcrosses (average of 93% 297 species A) would be included as parental species individuals. 298 299 There are two ways to ameliorate error introduced in species assignment using 300 thresholds. One obvious way is to employ more markers (Box 1), which 301 increases the power of a study and allows the setting of thresholds approaching 302 0 and 1. Studies that have used thousands of markers use the most stringent 303 thresholds e.g. [30]. A second solution to the threshold problem is to do away 304 with them entirely. Rather than assigning individuals to species classes based on 305 point estimates, it is more appropriate to use **credible** or **confidence intervals** 306 around point estimates which capture uncertainty in the marker system 307 appropriately (Box 2). In this scenario any individual with a credible interval 308 overlapping 0 or 1 is considered a parental species and all others are considered 309 hybrids. 310 311 An additional problem in separating hybrid individuals from parental species is 312 that some hybrids, particularly later generation of backcrosses, will be 313 homozygous for all sampled diagnostic loci by chance. This is due to increased 314 variation around the proportion of genome inherited from each parental species 315 with each generation of backcrossing ([44]; Box 1). The hybrid nature of these 316 individuals will be undetectable, and they will be classified as parental species, 317 even though unmarked genome regions may be introgressed. Increasing the

number of markers increases the probability of sampling a hybrid individual at loci that are heterozygous or homozygous for alleles representative of both parental species (Box 1).

321

322

323

324

325

326

327

328

329

330

331

332

333

334

335

336

337

338

339

340

341

342

343

344

345

346

347

348

318

319

320

Higher density markers to identify bimodal hybrid zones

When researchers apply higher density marker panels to examples of anthropogenic hybridization, they generally uncover more backcrossed individuals compared to studies using low-density panels, and can draw more accurate conclusions about the system. These newly-detected backcrosses are often genetically very similar to the parental species, with less than 10% introgression, indicative of a bimodal hybrid zone. For example, in a study of Italian wolves that hybridize with domestic dogs, use of 170,000 SNPs found that hybridization had occurred 3 -5 generations prior to sampling [30]. This multigeneration backcrossing was not detectable in the population when 18 microsatellite markers were used [49]. Further, while very few individuals were found to have Q scores between 0.25 and 0.75, as would be expected in a hybrid swarm with a complete breakdown of reproductive isolation, 62% of sampled Eurasian wolves had a small proportion (<5%) of admixture with domestic dogs [55]. The Eurasian wolf – domestic dog system has the distribution of admixture scores and phenotypes that characterizes a bimodal hybrid zone with some degree of mating preference for parental phenotypes, or rare intermediate hybrids. In this system, most individuals are either phenotypically dog-like with extreme Q scores at one end of the distribution, or phenotypically wolf-like with Q scores at the other end of the distribution. There are few individuals with intermediate Q scores and phenotypes. This can be contrasted with the westslope cutthroat (*Oncorhynchus clarki lewisi*) – rainbow trout (*O. mykiss*) system, which has also recently been genotyped using 3180 diagnostic SNPs [56]. While the increase in number of markers did lead to increased detection of advanced backcrosses, there were also many individuals with intermediate Q scores and phenotypes [56, 57]. This suggests that the westslope cutthroatrainbow trout system is a hybrid swarm that has little assortative mating.

349

350

Designing an ideal study of an anthropogenic hybrid zone

351	When embarking on a study of anthropogenic hybridization, there are many
352	considerations in deciding on the genetic resources to be used (Box 1). As whole
353	genome sequencing (WGS) becomes cheaper [58], conservation biologists should
354	consider whether WGS is the best way forward. Firstly, WGS data allows for
355	detection of heterogeneity of introgression across the genome. If conservation
356	biologists truly adopt a 'gene view point' of hybridization [16] then individuals
357	ought to be classified based on whether they carry specific alleles at identified
358	loci, rather than by overall Q scores (but see [10] for a discussion of the difficulty
359	of implementing this approach). Secondly, WGS enables the researcher to
360	distinguish between historical and contemporary introgression. Finally, we
361	anticipate that the use of WGS will result in more diagnostic or ancestry
362	informative markers being detected, and thus make studies more powerful.
363	Researchers will be more confident in their estimates of individual admixture,
364	and will report the power and confidence associated with their analyses (Box 2).
365	While the bioinformatics skills required to assemble a genome and call SNPs may $$
366	seem intimidating, we believe that 1) these are skills are now routinely taught in
367	universities and 2) WGS presents an additional opportunity for conservation
368	biologists to collaborate with speciation geneticists (Box 3). Another
369	consideration is that high quality DNA is needed for the most accurate
370	assemblies, although progress is being made towards high quality sequences
371	from poor quality samples (e.g. [59]). While the use of WGS is more expensive
372	than microsatellite marker studies, when the cost of microsatellite markers,
373	including the cost of labour, was compared to the use of SNP markers in
374	European wolves, SNPs were less expensive if at least 24 samples were
375	genotyped [60]. This suggests that the use of thousands of variable genome wide
376	markers (e.g. from ddRAD [61]) may represent a practical middle ground for
377	conservation biologists, depending on the history and biology of the system.
378	Taken together, we believe that the best way forward to accurately detect
379	backcrossing in studies of anthropogenic hybrid zones is to routinely use higher
380	density markers, including WGS when possible.

Concluding Remarks

Advanced backcrosses are unlikely to have been detected with many of the methods that biologists studying anthropogenic hybridization have used to date. Most studies of anthropogenic hybridization have used fewer than 20 markers [13], too few to reliably detect individuals that are more than two generations backcrossed [33], unless markers are perfectly species diagnostic [44]. For this reason, it is rare for studies to consider backcrossed individuals past the second generation of backcrossing, regardless of the number of generations that have passed since secondary contact. Here, we suggest that studies should attempt to go much further. By accounting for the number of generations since secondary contact and increasing the density of genetic markers accordingly, many more backcrossed individuals will become distinguishable from the parental populations. We echo the call for more genetic markers to be used in these studies to allow for higher accuracy and efficiency [1, 3, 13, 33, 62], particularly since we have now entered the genomics era, making tens or hundreds of thousands of markers obtainable even in non-model systems [58]. It seems likely that anthropogenic hybridization will only increase in frequency and result in increased gene flow between previously isolated species [1]. The increase in number of markers and associated power will also open up the opportunity to ask new questions in these systems, parallel to those speciation biologists explore in natural hybrid zones (Box 3). There are new challenges with increased marker density, but a genomic approach to studying these systems will help researchers to detect backcrosses and make the best policy recommendations. **Acknowledgements**: We would like to thank Konrad Lohse and two anonymous

406

407

408

409

410

383

384

385

386

387

388

389

390

391

392

393

394

395

396

397

398

399

400

401

402

403

404

405

Acknowledgements: We would like to thank Konrad Lohse and two anonymous reviewers for valuable feedback on this manuscript. S.E.M. was funded by the Swedish Research Council (Vetenskapsrådet, International Postdoc 2017-00499).

Additional Elements

413 Glossary

- **Anthropogenic hybridization:** the breakdown of reproductive isolation between two species due to human action, including but not limited to, species introduction, habitat disturbance or escape of domestic species.
- **Accuracy:** the proportion of identified hybrids that are actually of hybrid ancestry [33]. A low accuracy suggests a high rate of type I errors, in which parental species individuals are erroneously assigned as hybrids.
- **Admixture**: the mixing of genomes from structured or diverged populations
- Allopatry: species in non-overlapping ranges
- **Ancestry informative markers:** genetic markers with substantial allele frequency differences across populations, which can be used to assign individuals to each population [63]
- **Bimodal hybrid zone:** a hybridizing population in which preference for parental phenotypes, or scarcity of hybrids with which to mate, results in a population that includes few F1 hybrids, and many backcrossed individuals with a low level of introgression that often resemble the parental species in phenotype. Can be unimodal (if backcrossing is into just one parental species) or bimodal (backcrossing into both parental species) [22]
- **Credible interval:** the range of possible values surrounding a point estimate, representing the uncertainty in the estimate
- Diagnostic markers: markers with fixed allele differences across populations
- Dxy: an absolute measure of genetic differentiation, calculated as the
 proportion of nucleotides that differ between two homologous sequences
 within the same or different population.
- **Efficiency**: Proportion of correctly identified individuals in each group [33]. If the null hypothesis is that an individual is from the parental species rather than a hybrid individual, then low efficiency suggests a

443		high rate of type II errors, in which hybrid individuals are incorrectly
444		assigned as parental species.
445	•	\mathbf{F}_{ST} –A measure of genetic differentiation between populations based on
446		the difference in allele frequencies within and between populations [64]
447	•	Hybridization: mating of individuals from diverged populations
448	•	Hybrid: an individual that has an intermediate genotype between two
449		diverged, parental populations, as the result of interbreeding between
450		these populations
451	•	Hybrid swarm: a hybridizing population that includes F1 hybrids and
452		various backcrosses, due to a total breakdown of assortative mating. Also
453		known as a unimodal hybrid zone [22].
454	•	Introgression: the movement of alleles between genetically
455		differentiated forms (including populations, species, etc), mediated by
456		backcrossing [65]
457	•	Secondary contact: Occurs when two (or more) species that have been in
458		allopatry come back into sympatry
459	•	Sympatry: species in overlapping ranges
460		

Figure 1: Anthropogenic hybridization falls into three main categories. These are 1) systems with inviable or infertile hybrids, 2) bimodal hybrid zones in which there is either mating preference for parental species phenotypes or the relative abundance of parental species means most matings are backcrosses and 3) hybrid swarms in which there is random mating and many hybrid individuals. In this schematic figure we illustrate for each type of anthropogenic hybridization system how many individuals of each admixture (Q) score might be found and typical distributions of mating success across Q scores according to whether there is a high likelihood of hybrid individuals mating with the parental species phenotypes present. While we represent hybrid swarms and bimodal hybrid zones as categorically different, these are probably ends of a continuum and some systems may be intermediate between them. Note that we have represented (2) as a bimodal hybrid zone due to backcrossing into both parental species. Alternatively there can be a single (i.e. unimodal) hybrid zone due to unidirectional backcrossing.

476 Box 1 - How many markers do I need to discover backcrossed individuals 477 in my system? 478 479 Substantial power is needed to detect individuals that are the result of repeated 480 generations of backcrossing. General rules have been suggested, including that 481 for each additional generation twice as many markers are needed [44], and that 482 at least 48 markers would be needed to consistently detect first generation 483 backcrossing in hybrids with parental species that have an F_{st} =0.21 [33]. 484 However, we are now in the age of genomics, when the cost of increasing marker 485 density is dramatically decreasing [58], and thus marker numbers should be less of a barrier than previously. So, how many markers does a study need to reliably 486 487 detect backcrossed individuals? 488 489 To maximize detection of backcrossed individuals, researchers can increase their 490 power in three ways; through increased divergence, the use of diagnostic 491 markers, or with increased numbers of markers. Studies with high divergence 492 between hybridizing species have high power [33]. However, as many 493 conservation biologists choose their study system based on conservation 494 concerns and not to maximize power, this advice is not helpful. Diagnostic 495 markers have fixed allelic differences between parental species and are the most 496 powerful for backcross detection [25]. Ancestry informative markers, those with 497 strong allele frequency divergence between species, are also very powerful [63]. 498 Loci with weak allele frequency divergence between species are least useful. 499 Diagnostic and ancestry informative markers can be determined based on 500 genotyping and contrasting known parental species individuals, although this is 501 not always feasible (e.g. [55]). Additionally, the diagnostic properties of markers 502 are a function of the populations and individuals that have been sampled; more 503 extensive sampling sometimes demonstrates that selected markers are not 504 diagnostic for all populations [66]. Generally speaking, the more markers used, 505 the higher the chance of detection of admixture in an individual [33, 44]. 506 507 Assuming diagnostic markers, it is ideal to know the number of elapsed 508 generations since the initial hybridization, as, for every further generation of

backcrossing, the proportion of introgressed genome halves [44]. The number of generations since hybridization should be interpreted with an eye to policy. After some number of generations of uni-directional backcrossing, policy will dictate that we consider an individual to be parental species (again) [67]. It's best to make this decision prior to marker selection, as it is impossible to apply policy decisions regarding the acceptability of backcrossed individuals without sufficient detection power. If we are interested in all generations of backcrossing, then we can extend the deterministic model developed by Boecklen and Howard ([44]; Equation 2) for the genomics era. We made the same assumptions, specifically that backcrossing is unidirectional, loci are independent and Mendelian, all markers are diagnostic, all backcrossing is between the previous generation of backcrosses and parental species, and all genotypes are equally fecund [44]. We asked what proportion of backcrossed individuals are undetectable because they are homozygous for all diagnostic markers. We modeled 10 generations of backcrossing, and each of 10, 100 and 1000 diagnostic markers (Figure I). When using 10 diagnostic markers, 52% of 4th generation backcrosses are homozygous for one parental species at all loci, and thus undetectable as backcrosses. In contrast, 1000 diagnostic markers allow for powerful (85%) detection of 9th generation backcrosses. Figure I: An extension of the deterministic model presented by Boecklen and Howard [44]. The proportion of hybrid individuals that are homozygous at all the (diagnostic) markers, and are hence indistinguishable from the parental species that is being introgressed, increases with each generation of backcrossing, but decreases with increased marker density. This demonstrates that more markers than are typically used in studies of anthropogenic hybridization are needed to detect advanced backcrosses.

509

510

511

512

513

514

515

516

517

518

519

520

521

522

523

524

525

526

527

528

529

530

531

532

533

534

535

536

537

538

540 **Box 2 - Reporting Error** 541 Credible (or confidence) intervals (CIs) are a powerful, intuitive way to assess 542 confidence in the estimates being presented [68, 69]. Measures of uncertainty 543 are not always presented in estimates of anthropogenic hybridization (although 544 see [53, 70-73] for exceptions), perhaps because the uncertainty is so high where 545 estimated. Credible intervals can be calculated using STRUCTURE [25] and 546 standard errors can be calculated using ADMIXTURE [29], so reporting of error 547 estimates is easily implemented in a routine workflow. 548 549 There are practical implications of the reporting of credible intervals. 550 particularly for individuals with very low or very high admixture values (Q). Cut-551 off thresholds have been used to determine if individuals are members of the 552 parental populations or are admixed, but these thresholds are usually based on 553 the detection power of a study (see main text). Since these are hard cut-offs, 554 individuals with very similar levels of admixture can be assigned to very 555 different populations. For example, with a Q cut-off of 0.80, if individual 1 is 556 assessed as Q=0.79, it is determined to be admixed and, depending on the 557 management of the system, may be culled. In contrast, if individual 2 is estimated 558 to have Q=0.81, it would be considered a parental species individual and be 559 retained for breeding. There may be no substantive difference between these 560 individuals, although this is impossible to tell using only point estimates. 561 562 We recommend that credible intervals should also be included in visual 563 depictions of admixture. Typically, the key figure from a paper on anthropogenic 564 hybridization is the characteristic "STRUCTURE Bar Plot" [25], that uses stacked 565 colours to denote genetic contributions from different source populations. These 566 plots show the point estimates for each individual, and allow the author to 567 determine thresholds for inclusion in each group. While such figures are 568 compelling and easily interpreted, they do not convey the uncertainty around 569 individual point estimates. 570 571 Allendorf and colleagues [11] noted that it is very difficult to make policy 572 decisions when comparing different low point estimates of admixture. We

recommend that researchers should focus on the uncertainty around Q estimates when making decisions about the genetic group each individual belongs to. It has been pointed out that the use of credible intervals demonstrates the high levels of uncertainty researchers are facing [70]. As they should! This problem will of course be substantially alleviated by using more markers (see Box 1).

Box 3: Lessons from Natural Systems

579

611

580 581 Naturally occurring hybrid zones have long been used as 'natural laboratories' to 582 study the speciation process [74]. The field of speciation genomics works to 583 understand how genomic differences build up to cause eventual reproductive 584 isolation [75-78]. Recently, population geneticists have used genome wide 585 markers to ask questions regarding the genomic architecture of reproductive 586 isolation and speciation, and how the genomes of diverged populations change in 587 the face of on-going gene flow [43, 78, 79]. Further, many studies of natural 588 hybrid zones have focused on isolating signals from historical vs. contemporary 589 hybridization (main text 2.1.1, [78]). These questions that speciation biologists 590 ask using hybrid zones could equally be asked in anthropogenic hybrid zones, 591 particularly in studies that used whole genome sequence data. Indeed, studies of 592 anthropogenic hybrid zones may even have more power than those with 593 naturally occurring secondary contact as in some cases of introduced or escaped 594 heterospecifics, phenotypic divergence is more extreme, meaning that fewer 595 individuals would need to be sampled for, for example, admixture mapping [78]. 596 597 Use of genomic data allows speciation geneticists to examine heterogeneity in 598 divergence across the genome. Indeed, the questions we noted above are most 599 interesting when heterogeneity is found. Genome scans look for regions of high 600 divergence between species (F_{st} or \mathbf{d}_{xy}) which may indicate regions that resist 601 introgression, also known as 'speciation islands' [80], or 'islands of 602 differentiation' [79]. While such signals are not without controversy [81], and in 603 some cases may represent phylogenetically derived regions of low 604 recombination, rather than reproductive isolation [82], they represent 605 interesting candidate regions for fixed differences between hybridizing species, 606 and thus could be used diagnostically by conservation biologists. For example, 607 golden-winged (Vermivora chrysoptera) and blue-winged warblers (V. 608 *cyanoptera*), which hybridize in eastern North America are phenotypically 609 distinct but undistinguishable when using low density, microsatellite marker 610 panels [83]. Only with the use of whole genome sequencing were six small

divergent regions of the genome discovered, four of which are associated with

either pigmentation or feather development genes and explain more than 90% of the variation in plumage [39]. This demonstrates that a focus on the use of high density markers to explore heterogeneity across the genome allows for higher power to both distinguish between closely related, hybridizing species genetically, and to associate genomic regions with diverged phenotypes, two possible goals of conservation biologists working on anthropogenic hybrid zones. We echo the call of [1] that conservation biologists can take a cue from speciation biologists that have, in many cases, developed methods that use genomics to ask interesting questions of hybrid zones.

622 **References**

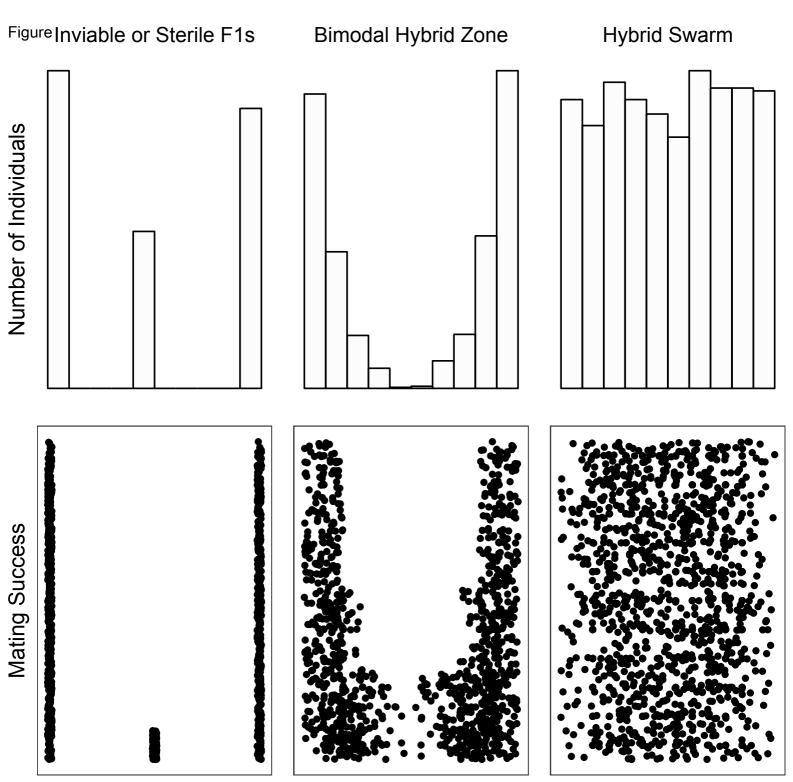
- 623 1. Grabenstein, K.C. and Taylor, S.A. (2018) Breaking Barriers: Causes,
- 624 Consequences, and Experimental Utility of Human-Mediated Hybridization.
- 625 Trends in Ecology & Evolution 33(3): 198-212
- 626 2. Brennan, A.C. et al. (2015) Hybridization due to changing species distributions:
- adding problems or solutions to conservation of biodiversity during global
- change? Evolutionary Ecology Research 16 (6), 475-491.
- 3. Wayne, R.K. and Shaffer, H.B. (2016) Hybridization and endangered species
- protection in the molecular era. Molecular Ecology 25 (11), 2680-2689.
- 4. Crispo, E. et al. (2011) Broken barriers: Human induced changes to gene flow
- and introgression in animals. BioEssays 33 (7), 508-518.
- 5. Pimm, S.L. et al. (2006) The genetic rescue of the Florida panther. Animal
- 634 Conservation 9 (2), 115-122.
- 635 6. Frankham, R. (2015) Genetic rescue of small inbred populations: Meta
- analysis reveals large and consistent benefits of gene flow. Molecular Ecology 24
- 637 (11), 2610-2618.
- 638 7. Suarez Gonzalez, A. et al. (2016) Genomic and functional approaches reveal a
- 639 case of adaptive introgression from *Populus balsamifera* (balsam poplar) in *P.*
- *átrichocarpa* (black cottonwood). Molecular Ecology 25 (11), 2427-2442.
- 8. Hamilton, J.A. and Miller, J.M. (2016) Adaptive introgression as a resource for
- management and genetic conservation in a changing climate. Conservation
- 643 Biology 30 (1), 33-41.
- 9. Kovach, R.P. et al. (2016) Risk and efficacy of human enabled interspecific
- 645 hybridization for climate change adaptation: response to Hamilton and Miller
- 646 (2016). Conservation Biology 30 (2), 428-430.
- 10. Kardos, M. and Shafer, A.B. (2018) The Peril of Gene-Targeted Conservation.
- 648 Trends in Ecology & Evolution 33(11): 827-839
- 11. Allendorf, F.W. et al. (2001) The problems with hybrids: setting conservation
- 650 guidelines. Trends in Ecology & Evolution 16 (11), 613-622.
- 12. Allendorf, F.W. and Luikart, G. (2009) Conservation and the genetics of
- 652 populations, John Wiley & Sons.
- 13. Todesco, M. et al. (2016) Hybridization and extinction. Evolutionary
- 654 Applications 9 (7), 892-908.
- 655 14. Rhymer, J.M. and Simberloff, D. (1996) Extinction by hybridization and
- introgression. Annual Review of Ecology and Systematics, 83-109.
- 15. Rieseberg, L.H. et al. (1989) Hybridization in the island endemic, *Catalina*
- 658 *mahogany*. Conservation Biology 3 (1), 52-58.
- 659 16. Petit, R.J. (2004) Biological invasions at the gene level. Diversity and
- 660 Distributions 10 (3), 159-165.
- 661 17. Schumer, M. et al. (2018) Natural selection interacts with recombination to
- shape the evolution of hybrid genomes. Science DOI: 10.1126/science.aar3684
- 18. Fitzpatrick, B.M. et al. (2010) Rapid spread of invasive genes into a
- threatened native species. Proceedings of the National Academy of Sciences 107
- 665 (8), 3606-3610.
- 19. Amish, S.J. et al. (2012) RAD sequencing yields a high success rate for
- westslope cutthroat and rainbow trout species diagnostic SNP assays.
- Molecular Ecology Resources 12 (4), 653-660.

- 20. Epifanio, J. and Philipp, D. (2000) Simulating the extinction of parental
- lineages from introgressive hybridization: the effects of fitness, initial
- 671 proportions of parental taxa, and mate choice. Reviews in Fish Biology and
- 672 Fisheries 10 (3), 339-354.
- 21. Lamichhaney, S. et al. (2018) Rapid hybrid speciation in Darwin's finches.
- 674 Science 359 (6372), 224-228.
- 675 22. Jiggins, C.D. and Mallet, J. (2000) Bimodal hybrid zones and speciation.
- 676 Trends in Ecology & Evolution 15 (6), 250-255.
- 23. Rundle, H.D. and Nosil, P. (2005) Ecological speciation. Ecology Letters 8 (3),
- 678 336-352.
- 679 24. Allen, B. et al. (2016) Role of genetic background in the introgressive
- 680 hybridization of rainbow trout (*Oncorhynchus mykiss*) with Westslope cutthroat
- trout (*O. clarkii lewisi*). Conservation genetics 17 (3), 521-531.
- 682 25. Pritchard, J.K. et al. (2000) Inference of population structure using multilocus
- 683 genotype data. Genetics 155 (2), 945-959.
- 26. Raj, A. et al. (2014) fastSTRUCTURE: variational inference of population
- structure in large SNP data sets. Genetics 197 (2), 573-589.
- 686 27. Kwan, Y.S. et al. (2014) Genomic replacement of native *Cobitis lutheri* with
- introduced *C. tetralineata* through a hybrid swarm following the artificial
- connection of river systems. Ecology and Evolution 4 (8), 1451-1465.
- 28. Le Roux, J.J. et al. (2015) Genetic analysis shows low levels of hybridization
- 690 between African wildcats (Felis silvestris lybica) and domestic cats (F. s. catus) in
- 691 South Africa. Ecology and Evolution 5 (2), 288-299.
- 692 29. Alexander, D.H. and Lange, K. (2011) Enhancements to the ADMIXTURE
- algorithm for individual ancestry estimation. BMC bioinformatics 12 (1), 246.
- 694 30. Galaverni, M. et al. (2017) Disentangling timing of admixture, patterns of
- introgression and phenotypic indicators in a hybridizing wolf population.
- 696 Molecular Biology and Evolution 34(9):2324-2339
- 31. Nielsen, E.E. et al. (2006) HYBRIDLAB (version 1.0): a program for generating
- 698 simulated hybrids from population samples. Molecular Ecology Resources 6 (4).
- 699 971-973.
- 700 32. van Heugten, R.A. et al. (2017) Sleeping with the 'enemy': hybridization of an
- 701 endangered tree weta. Conservation Genetics 18 (6), 1377-1387.
- 33. Vähä, J.-P. and Primmer, C.R. (2006) Efficiency of model based Bayesian
- methods for detecting hybrid individuals under different hybridization scenarios
- and with different numbers of loci. Molecular Ecology 15 (1), 63-72.
- 34. vonHolt, B.M. et al. (2016) Whole-genome sequence analysis shows that two
- 706 endemic species of North American wolf are admixtures of the coyote and gray
- 707 wolf. Science Advances 2 (7), e1501714.
- 708 35. Hohenlohe, P.A. et al. (2017) Comment on "Whole-genome sequence analysis
- shows two endemic species of North American wolf are admixtures of the coyote
- and gray wolf". Science Advances 3 (6), e1602250.
- 711 36. Durand, E.Y. et al. (2011) Testing for ancient admixture between closely
- related populations. Molecular Biology and Evolution 28 (8), 2239-2252.
- 37. Soraggi, S. et al. (2018) Powerful Inference with the D-Statistic on Low-
- Coverage Whole-Genome Data. G3: Genes, Genomes, Genetics 8 (2), 551-566.
- 38. Gutenkunst, R.N. et al. (2009) Inferring the joint demographic history of
- 716 multiple populations from multidimensional SNP frequency data. PLoS Genetics
- 717 5 (10), e1000695.

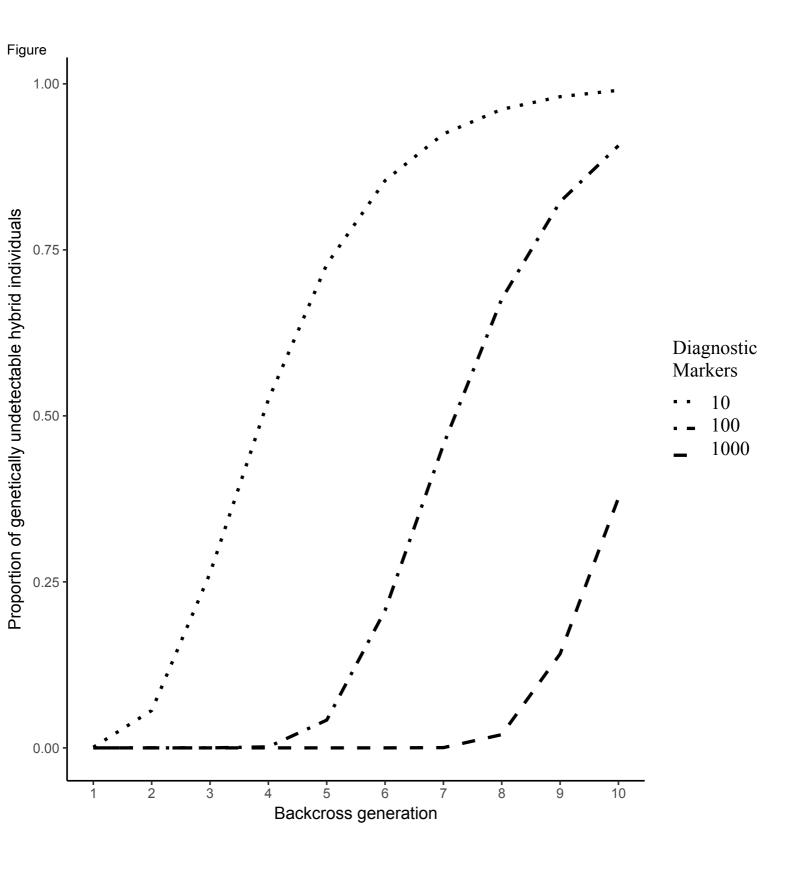
- 718 39. Toews, D.P. et al. (2016) Plumage genes and little else distinguish the
- 719 genomes of hybridizing warblers. Current Biology 26 (17), 2313-2318.
- 40. Gravel, S. (2012) Population genetics models of local ancestry. Genetics
- 721 191(2):607-619
- 41. Harris, K. and Nielsen, R. (2013) Inferring demographic history from a
- spectrum of shared haplotype lengths. PLoS Genetics 9 (6), e1003521.
- 42. Palamara, P.F. et al. (2012) Length distributions of identity by descent reveal
- fine-scale demographic history. The American Journal of Human Genetics 91 (5),
- 726 809-822.
- 43. Payseur, B.A. and Rieseberg, L.H. (2016) A genomic perspective on
- hybridization and speciation. Molecular Ecology 25 (11), 2337-2360.
- 44. Boecklen, W.J. and Howard, D.J. (1997) Genetic analysis of hybrid zones:
- numbers of markers and power of resolution. Ecology 78 (8), 2611-2616.
- 731 45. Dumont, B.L. (2017) Variation and evolution of the meiotic requirement for
- 732 crossing over in mammals. Genetics 205 (1), 155-168.
- 46. Johnston, S.E. et al. (2018) A genomic region containing REC8 and RNF212B
- is associated with individual recombination rate variation in a wild population of
- 735 red deer (*Cervus elaphus*). 8(7):2265-2276
- 47. Ellegren, H. (2010) Evolutionary stasis: the stable chromosomes of birds.
- 737 Trends in Ecology & Evolution 25 (5), 283-291.
- 48. Anderson, E.C. (2009) Statistical methods for identifying hybrids and groups.
- 739 In Population Genetics and Animal Conservation (Bertorelle, G. et al. eds), pp. 25-
- 740 41, Cambridge University Press.
- 49. Randi, E. (2008) Detecting hybridization between wild species and their
- 742 domesticated relatives. Molecular Ecology 17 (1), 285-293.
- 50. Galov, A. et al. (2015) First evidence of hybridization between golden jackal
- 744 (*Canis aureus*) and domestic dog (*Canis familiaris*) as revealed by genetic
- markers. Royal Society Open Science 2 (12), 150450.
- 746 51. Haldane, J.B. (1922) Sex ratio and unisexual sterility in hybrid animals.
- 747 Journal of Genetics 12 (2), 101-109.
- 748 52. Turelli, M. and Moyle, L.C. (2007) Asymmetric postmating isolation: Darwin's
- 749 corollary to Haldane's rule. Genetics 176 (2), 1059-1088.
- 53. Senn, H.V. and Pemberton, J.M. (2009) Variable extent of hybridization
- 751 between invasive sika (*Cervus nippon*) and native red deer (*C. elaphus*) in a small
- 752 geographical area. Molecular Ecology 18 (5), 862-876.
- 753 54. Schulte, U. et al. (2012) Rapid genetic assimilation of native wall lizard
- 754 populations (*Podarcis muralis*) through extensive hybridization with introduced
- 755 lineages. Molecular Ecology 21 (17), 4313-4326.
- 756 55. Pilot, M. et al. (2018) Widespread, long term admixture between grey
- 757 wolves and domestic dogs across Eurasia and its implications for the
- 758 conservation status of hybrids. Evolutionary Applications.
- 759 DOI:10.1111/eva.12595
- 760 56. Hohenlohe, P.A. et al. (2013) Genomic patterns of introgression in rainbow
- and westslope cutthroat trout illuminated by overlapping paired end RAD
- 762 sequencing. Molecular Ecology 22 (11), 3002-3013.
- 57. Boyer, M.C. et al. (2008) Rainbow trout (*Oncorhynchus mykiss*) invasion and
- the spread of hybridization with native westslope cutthroat trout (*Oncorhynchus*
- 765 *clarkii lewisi*). Canadian Journal of Fisheries and Aquatic Sciences 65 (4), 658-
- 766 669.

- 58. Ellegren, H. (2014) Genome sequencing and population genomics in non-
- model organisms. Trends in Ecology & Evolution 29 (1), 51-63.
- 769 59. van der Valk, T. et al. (2017) Whole mitochondrial genome capture from
- 770 faecal samples and museum preserved specimens. Molecular Ecology
- 771 Resources 17 (6), e111-e121.
- 60. de Groot, G.A. et al. (2016) Decades of population genetic research reveal the
- need for harmonization of molecular markers: the grey wolf *Canis lupus* as a case
- 774 study. Mammal Review 46 (1), 44-59.
- 775 61. Andrews, K.R. et al. (2016) Harnessing the power of RADseq for ecological
- and evolutionary genomics. Nature Reviews Genetics 17 (2), 81-92.
- 62. Koskinen, M.T. et al. (2004) The benefits of increasing the number of
- 778 microsatellites utilized in genetic population studies: an empirical perspective.
- 779 Hereditas 141 (1), 61-67.
- 780 63. Shriver, M.D. et al. (2003) Skin pigmentation, biogeographical ancestry and
- 781 admixture mapping. Human Genetics 112 (4), 387-399.
- 782 64. Wright, S. (1943) Isolation by distance. Genetics 28, 114-138.
- 783 65. Avise, J.C. (2012) Molecular markers, natural history and evolution, Springer
- 784 Science & Business Media.
- 785 66. Hand, B.K. et al. (2015) Genomics and introgression: discovery and mapping
- 786 of thousands of species-diagnostic SNPs using RAD sequencing. Current Zoology
- 787 61 (1), 146-154.
- 788 67. Lavretsky, P. et al. (2016) Becoming pure: identifying generational classes of
- admixed individuals within lesser and greater scaup populations. Molecular
- 790 Ecology 25 (3), 661-674.
- 791 68. Altman, D. et al. (2013) Statistics with confidence: confidence intervals and
- 792 statistical guidelines, John Wiley & Sons.
- 793 69. Nakagawa, S. and Cuthill, I.C. (2007) Effect size, confidence interval and
- statistical significance: a practical guide for biologists. Biological Reviews 82 (4),
- 795 591-605.
- 70. Bohling, J.H. and Waits, L.P. (2011) Assessing the prevalence of hybridization
- 797 between sympatric Canis species surrounding the red wolf (*Canis rufus*)
- recovery area in North Carolina. Molecular Ecology 20 (10), 2142-2156.
- 799 71. Trigo, T. et al. (2008) Inter species hybridization among Neotropical cats of
- 800 the genus *Leopardus*, and evidence for an introgressive hybrid zone between *L.*
- 801 geoffroyi and L. tigrinus in southern Brazil. Molecular Ecology 17 (19), 4317-
- 802 4333.
- 72. Beaumont, M. et al. (2001) Genetic diversity and introgression in the Scottish
- 804 wildcat. Molecular Ecology 10 (2), 319-336.
- 73. Yokoyama, R. et al. (2009) Disturbance of the indigenous gene pool of the
- 806 threatened brook lamprey *Lethenteron* sp. S by intraspecific introgression and
- habitat fragmentation. Conservation genetics 10 (1), 29-43.
- 74. Hewitt, G.M. (1988) Hybrid zones-natural laboratories for evolutionary
- studies. Trends in Ecology & Evolution 3 (7), 158-167.
- 75. Nosil, P. and Feder, J. (2012) Genomic divergence during speciation, causes
- and consequences. Philosophical Transations of the Royal Society B 367(1587):
- 76. Seehausen, O. et al. (2014) Genomics and the origin of species. Nature
- 813 Reviews Genetics 15 (3), 176.
- 814 77. Butlin, R.K. (2010) Population genomics and speciation. Genetica 138 (4),
- 815 409-418.

- 78. Nadeau, N.J. and Kawakami, T. (2018) Population Genomics of Speciation and
- 817 Admixture. *in* Population Genomics. Springer, Cham.
- 818 DOI:10.1007/13836_2018_24
- 79. Wolf, J.B. and Ellegren, H. (2017) Making sense of genomic islands of
- differentiation in light of speciation. Nature Reviews Genetics 18 (2), 87.
- 80. Turner, T.L. et al. (2005) Genomic islands of speciation in *Anopheles gambiae*.
- 822 PLoS Biology 3 (9), e285.
- 81. Cruickshank, T.E. and Hahn, M.W. (2014) Reanalysis suggests that genomic
- islands of speciation are due to reduced diversity, not reduced gene flow.
- 825 Molecular Ecology 23 (13), 3133-3157.
- 82. Burri, R. et al. (2015) Linked selection and recombination rate variation drive
- the evolution of the genomic landscape of differentiation across the speciation
- continuum of *Ficedula* flycatchers. Genome Research 25 (11), 1656-1665.
- 83. Vallender, R. et al. (2007) Complex hybridization dynamics between
- 830 golden winged and blue winged warblers (Vermivora chrysoptera and
- 831 *Vermivora pinus*) revealed by AFLP, microsatellite, intron and mtDNA markers.
- 832 Molecular Ecology 16 (10), 2017-2029.



Admixture (Q) scores from 0 - 1



Outstanding questions:

1) Do replicate anthropogenic hybrid zones show similar patterns of introgression?

There are big evolutionary questions that could be answered by the sorts of data that conservation biologists working on anthropogenic hybridization could answer. For example, there are multiple replicate hybrid zones occurring in the wolf/dog, wild cat/domestic cat, red deer/sika deer, westslope cut throat trout/rainbow trout systems. But in many cases, there is limited communication and collaboration between researchers, or different markers are used across studies [60]. Clearly this isn't a problem unique to this field, but it is the case that collaboration between researchers would be made easier with standardized genome wide data aligned to a common genome. Genomic data make cross study comparisons easier, and would allow for easier comparison between studies.

- 2) Once there has been a breakdown of reproductive isolation characterized as hybridization, how common is maintenance of within parental species assortative mating? Is the strength of assortative mating stronger when species are more diverged, or perhaps between closely related species that have recently evolved reproductive isolation?
- 3) What is the relative frequency of hybrid swarms vs bimodal hybrid zones? We expect that the prevalence of bimodal hybrid zones has been underestimated because of the difficulty of detecting highly introgressed backcrosses. Increased use of high-density markers will make these cases easier to detect and would enhance our understanding of the systems that are bimodal hybrid zones.

Highlights:

Anthropogenic hybridization is increasingly common and likely to result in a breakdown of reproductive isolation between 'good' species.

Backcrossed individuals that have only a small proportion of one parental genome are difficult to differentiate from parental individuals using the most common current technologies.

Bimodal hybrid zones are characterized by introgression and backcrossing. The majority of hybrid individuals in these systems have low levels of introgression. The problems posed by bimodal hybrid zones have been largely overlooked in the literature.

Genome wide sampling of genetic markers at high densities allow for increased precision in the estimate of admixture proportions, which makes it feasible to detect multi-generation backcrosses, and will thus make it easier to differentiate bimodal hybrid zones from hybrid swarms or systems without introgression.